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1 **Antimony measurements in environmental matrices: Seven considerations**

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3 20 **Abstract**
4

5 21 The development of robust methods for determining the concentration and speciation of
6 22 antimony (Sb) in natural samples is essential to understanding its distribution and cycling in
7 23 nature. Here we discuss our experiences with a variety of approaches for measuring the
8 24 content and speciation of Sb in environmental matrices. Total Sb concentration measurements
9 25 in waters require digestion with HNO₃-HCl to release Sb from particulate material and may
10 26 require a preconcentration step to remove Sb from saline matrices or to obtain the required
11 27 sensitivity. Plant analyses require the use of HNO₃-HBF₄ or HNO₃-HF while sediments
12 28 require the use of HNO₃-HCl to solubilise Sb and prevent adsorption to silicates. Methods for
13 29 Sb speciation should be fit for purpose. Volatile Sb species can be measured successfully
14 30 using SPME-GCMS, waters via hydride generation-trapping ICPMS and sediment extracts
15 31 using HPLC-ICPMS. Extraction of Sb from sediments and plants presents a challenge;
16 32 however, the use of citrate is adequate for extraction of Sb from sediments predominately
17 33 containing Sb associated with Fe-Mn oxyhydroxide phases. We have been unable to
18 34 successfully quantify organic Sb species in plants because of the oxidation of Sb(III) to Sb
19 35 (V) and the degradation and transformation of organic Sb species. The analysis of solid
20 36 samples using x-ray absorption spectroscopy should be considered as it has been shown to
21 37 discriminate between Sb(III) and Sb(V) as well as Sb minerals, oxides and adsorbed Sb
22 38 species.
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40 Introduction

41 Understanding the environmental cycling and toxicology of antimony (Sb) has gained
42 increased attention in recent years. Antimony used to be found mainly in elevated
43 concentrations in soils near mines and systems receiving mine runoff.¹⁻³ In addition, Sb is
44 used in flame retardants, brake linings, batteries, PET bottles, pharmaceuticals and paints,
45 which are also potential sources to the environment.⁴⁻¹¹ Environmental Sb analysis has been
46 extended to roadside dust, atmospheric particles, plants and animals.¹²⁻¹⁶ As well as total Sb
47 concentrations, the speciation of Sb is required to understand antimony's mobility and
48 toxicology. Antimony is considered to be non-essential for plants and animals⁴ and there is
49 concern about its potential toxicity and other biological effects to plants and animals.

50 Antimony occurs mainly in inorganic forms as Sb (III) and Sb(V)^{17, 18} while methyl Sb
51 species (CH₃SbX, (CH₃)₂SbX, (CH₃)₃Sb) have been reported in seawater,^{19, 20} freshwater,³
52 geothermal waters,²¹ sediments and sediment pore waters,²² bacterial cultures,^{23, 24} landfill
53 gases²⁵ and plants.^{3, 26, 27}

54 We have been analysing Sb in environmental matrices for over 20 years and in this paper, we
55 discuss our experiences in measuring the content and speciation of Sb in bacterial cultures,
56 waters, sediments, plants and animal tissues. We have structured the discussion around seven
57 questions that should be considered before undertaking Sb analyses.

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59 Total Sb measurements

60 *Question 1: Can you measure total Sb concentrations?*

61 Total Sb concentrations are required to assess the distribution and movement of Sb in the
62 environment. As well, accurate Sb analyses are required for speciation purposes to measure
63 mass balances (extraction efficiencies, column recoveries etc.). Different procedures are
64 required for different matrices (Figure 1)

65 Water

66 Analysis of total Sb concentrations in water and waste waters may require a pre-digestion
67 step to release Sb from particulate material. In our laboratory, microwave digestion is used
68 extensively with HNO₃-HCl (2;1 v/v); 1 mL of acid mixture to 10 mL of sample at 150 °C for
69 30 min with analyses by ETAAS or ICPMS. If pre-concentration or separation from saline
70 matrices is required, we use pre-concentration by co-precipitation, chelation-solid phase
71 extraction and/or hydride generation.^{12, 20, 28} Digestion, however, is not always necessary or
72 desired; for example, the measurement of Sb in bottled water²⁹. As well, in toxicity studies,
73 the aim may be to distinguish between particulate and dissolved Sb thus filtration not
74 digestion is required.

75 Biota

76 For the analysis of animal tissues, most digestion procedures employ HNO₃ or HNO₃/H₂O₂
77 and microwave heating to quantitatively recover Sb.³⁰ Plants, however, require a different
78 digestion procedure if samples contain significant amounts of silica. In 2002,³⁰ we digested
79 all the available certified reference materials) and found a bias of ~25% resulting in lower
80 measured Sb concentrations ($y = 0.745x + 0.0163$, $r^2 = 0.900$, $n = 90$). At that time, only a
81 few certified reference materials were available and hence our confidence in the Sb values
82 reported in the literature was low. Subsequently we made a comparison of the efficiency of
83 digesting samples using HNO₃ and HNO₃-HBF₄ (with the results indicating that
84 underestimation of Sb concentrations occurs in some plant samples containing silica).¹³ This
85 has also been reported by several other studies.^{31, 32} Reanalysis of available certified
86 reference materials including more recently produced ones has shown the use of HNO₃-HBF₄
87 for digestion removes this bias ($y = 1.009x + 0.0002$, $r^2 = 0.940$, $n = 80$). The use of HBF₄,
88 however, may cause minor problems during ICPMS analysis; these problems include the
89 etching of glass nebulisers and alumina injectors. Traditionally, excess HF is complexed
90 using H₃BO₃³³. We prefer not to use boric acid as it can cause nebulisation and plasma
91 effects during ICPMS analyses. In practice, the small bias may not be an issue but needs to be
92 considered before choosing the appropriate digestion acids. We have investigated the Sb
93 species in digests and found in all cases that Sb was present as Sb(V).¹³ This was expected as
94 HNO₃ would convert inorganic Sb to Sb(V). Hydride generation can be used for total Sb
95 measurements as all Sb species have been found to degrade to inorganic Sb. More recently,
96 we have evaluated the use of HNO₃-H₂O₂-HCl and closed vessel microwave digestion at 180
97 °C for 15 min for digestion of plant samples and obtained 102-119% recovery of Sb from

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3 98 NIST citrus leaves.³⁴ Thus, Sb is not adsorbing to silicates and provides a suitable alternative
4 99 to the use of HBF₄

6 100 Sediments and soils

8 101 Evaluation of sediment extraction methods have clearly shown that using HNO₃ or
9 102 HNO₃/HClO₄ results in incomplete recoveries of Sb; probably due to the formation of
10 103 insoluble oxides such Sb₂O₅, SbO₄(OH)₂(NO₃) etc. and/or the adsorption of Sb-oxyhydroxy
11 104 compounds onto silicate minerals.³⁵ The use of HCl as part of the digestion mixture is
12 105 mandatory to solubilise Sb(V) as SbCl₆⁻ and prevent Sb adsorption to silica material.³⁶ The
13 106 use of HCl has the added advantage of dissolving common iron and silica minerals and
14 107 apatite.³⁷ We have evaluated the use of HNO₃-HCl mixtures for releasing Sb from mine
15 108 contaminated sediments³⁸ and found Sb is nearly quantitatively extracted (using a 2:1 v/v
16 109 HNO₃-HCl mixture (92-97%; 95±2%, n = 6). The use of elevated temperature (150 °C)
17 110 enhances the dissolution of Sb containing phases.³⁷ In these sediments, Sb is adsorbed to
18 111 Fe/Mn oxyhydroxides or silicates.³⁸⁻⁴⁰ More recently, the use of a using a 3:1 v/v HNO₃-HCl
19 112 mixture has given good recoveries (>93%) from a certified stream sediment.³⁴ In soils, not
20 113 from shooting ranges where Sb is likely to be in spent bullets, Sb is also likely to be
21 114 substantially adsorbed to Fe/Mn oxyhydroxides. At shooting ranges Sb is likely to be a
22 115 bronze alloy. Mine mineral associated Sb will be mainly Sb sulphides⁴¹⁻⁴³ and conventional
23 116 aqua regia (3:1 v/v HNO₃-HCl) probably will be required to release and retain Sb into
24 117 solution.⁴⁴⁻⁴⁷ Other soils and sediments may require the use of HF for the complete recovery
25 118 of Sb.⁴⁸

31 119 **Antimony speciation measurements**

33 120 ***Question 2: Have you chosen the most appropriate speciation method?***

35 121 The aim of speciation procedures is to maintain the integrity of Sb species and minimise
36 122 sample preparation procedures that may alter Sb speciation. There is a tendency for
37 123 laboratories to choose methods they are familiar with rather than the most appropriate
38 124 procedures likely to obtain accurate and unambiguous speciation data. Similar to total Sb
39 125 concentration measurements, appropriate procedures need to be used to undertake Sb
40 126 speciation measurements (Figure 2). Speciation results are dependent on the separation
41 127 mechanisms selected i.e. size exclusion, ion exchange, reverse phase etc. For example, ion
42 128 exchange is typically used to measure Sb(III) and Sb(V) redox states rather than single Sb
43 129 entities.

47 130 Volatile Sb species

49 131 There have been numerous studies investigating volatile methylated Sb species produced by
50 132 bacterial cultures and landfills.²³⁻²⁵ The most appropriate procedures are those that trap Sb
51 133 species and undertake measurements without derivatisation. Most published procedures
52 134 cryogenically trap Sb species and analyse individual Sb species by ICPMS or other
53 135 spectroscopic methods.^{3, 21, 49} The boiling points of volatile species are SbH₃ (-17 °C),
54 136 CH₃SbH₂ (41 °C), (CH₃)₂SbH (61 °C) and (CH₃)₃Sb (81 °C), respectively, and can be

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3 137 immobilised in a liquid N₂ trap packed with a GC chromatographic phase.²⁰ Programmed
4 138 heating is used to release individual Sb species for analysis. Alternatively, solid-phase micro
5 139 extraction (SPME) procedures utilising polydimethylsiloxane phases can be used to trap head
6 140 space gases. We have used SPME to trap volatile Sb species in the head space of bacterial
7 141 cultures and used GC-MS to analyse Sb species.⁴⁹ A cryogenic (CO₂ cooled) injection port is
8 142 required and again programmed heating is used to separate and measure Sb species. The
9 143 advantages of these procedures are that no derivatisation is required prior to sample analysis.
10 144 The use of mass spectrometry has the added advantage of unequivocal identification of Sb
11 145 species.

14 146 Waters

17 147 Normally river and seawaters not receiving mine wastes have low Sb concentrations and
18 148 require preconcentration of Sb species before analysis. The formation of volatile Sb species
19 149 (SbH₃, CH₃SbH₂, (CH₃)₂SbH and (CH₃)₃Sb) via the use of sodium tetrahydroborate (III) has
20 150 been used successfully for open ocean seawater samples.²⁰ Similar to that described for
21 151 volatile Sb species, a cryogenic trapping system is used prior to measurement by ICPMS or
22 152 other spectrometric methods.^{20, 26, 27, 50-52} The efficiency of hydride generation is critically
23 153 dependent on sample pH. The formation of SbH₃ from Sb(III) occurs over a wide pH range
24 154 (1-7) but Sb(V) reduction markedly decreases as pH increases.^{19, 20} The use of a reducing
25 155 agent such as cysteine or potassium iodide allows quantitative reduction of Sb(V) to Sb(III) at
26 156 higher pH.^{12, 20} The problem with these procedures is minimising the disproportionation of
27 157 (CH₃)₃Sb into other Sb species.^{26, 50, 53} The use of surface passivation, low acid strength (pH
28 158 > 1), exclusion of oxygen and the use of a chelating agent such as cysteine to promote the
29 159 reduction of Sb(V) to Sb (III) for hydride generation, along with the suppression of
30 160 interferences from Fe, Ni, Co, Cu etc. are all required.^{20, 26, 27, 54} We have described a fully
31 161 automated hydride generation-trapping ICPMS system that satisfies these conditions.²⁰
32 162 Published data using batch hydride generation systems indicates that they suffer from
33 163 disproportionation problems.^{26, 27} We have found that the fully automated system, that has no
34 164 glass components, also reduces contamination, i.e. lower blanks, and has much better
35 165 reproducibility compared to batch systems. The trapping of Sb species critically depends on
36 166 the chromatographic packing used in the trapping system so particular attention should be
37 167 given to this aspect of the system.

43 168 Sediments and plants

46 169 HPLC-ICPMS has been used extensively to analyse sediment and plant extracts.^{38, 55, 56} The
47 170 use of a PRP-X100 column with an EDTA-phthalate buffer is commonly used.^{50, 57} We have
48 171 found these approaches to be suitable for the analysis of citrate extracts of river sediments
49 172 and soils where Sb has been mobilised from a gold-antimony mine.³⁸ Extracts predominately
50 173 contain Sb (V) that has been released by the oxidative or reductive dissolution of Sb minerals
51 174 and reabsorbed by Fe/Mn oxyhydroxides and Al minerals.^{2, 39, 40, 58, 59} Citrate is commonly
52 175 used in sequential extraction schemes to extract Fe and Mn and associated metals.⁶⁰ Extracts
53 176 generally have low organic carbon content and thus artefacts are not typically formed. We
54 177 have not confidently used HPLC-ICPMS to analyse Sb species in plant extracts containing

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3 178 appreciable quantities of carbon because of artefact formation (see Question 6). Others have
4 179 reported similar difficulties in analysing organic rich extracts such as orange juice and
5 180 yoghurt.^{29, 61}

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8 181 **Question 3: Do you have confidence in Sb speciation standards?**

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10 182 There are various ways of preparing inorganic Sb(III) and Sb(V) standards. We prefer to use
11 183 KSb(III) tartrate ($\text{KSb}_2(\text{C}_4\text{O}_6\text{H}_2)_2^{2-}$) and KSb(V)(OH)₆ as forms of Sb(OH)₃ and Sb(OH)₆⁻,
12 184 respectively, as these are readily soluble in water to at least 1 mg Sb L⁻¹.⁶² Other available
13 185 compounds such as SbCl₃ and Sb₂O₅ are not sufficiently soluble. Antimony (V) standards are
14 186 stable for long periods, however, Sb(III) standards oxidise over short periods i.e. days.
15 187 Antimony (III) standards can be stabilised by adding citrate to form stable Sb(III)-citrate
16 188 complexes.⁶³ However, it should be noted that the ICPMS signals of Sb complexed with
17 189 EDTA, tartrate or citrate may give different responses to aqueous standards,²⁸ so care must be
18 190 taken to match the Sb species used for calibration curves.

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21 191 Trimethyl Sb halides (X= chloride or bromide) are readily synthesised but CH₃SbX,
22 192 (CH₃)₂SbX have not been synthesised in sufficient purity to be used as standards.²⁷ We
23 193 prepare methylated Sb hydride standards as a mixture by disproportionation of (CH₃)₃SbCl in
24 194 glacial acetic acid or by heating under less acidic solutions (pH <1) with sodium
25 195 tetrahydroborate (III). (CH₃)₃SbX readily oxidises to ((CH₃)₃Sb)₂O₂ and (CH₃)₃SbX(O₂H)₂⁶⁴
26 196 that are not likely to be soluble in water, although it has been suggested it can occur in
27 197 solution as ((CH₃)₃Sb(OH))⁺ and ((CH₃)₃SbOH_x(H₂O))⁺.⁵⁰

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31 198 **Question 4: Can you extract Sb species of interest?**

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33 199 Any speciation technique that relies on extracting Sb species requires knowledge of
34 200 extraction efficiencies and species integrity. Hence, trapping of volatile Sb species and water
35 201 analyses by formation of Sb hydrides is normally preferred because individual species are
36 202 preserved. We have successfully used microwave extraction with 25-50 mM citric acid to
37 203 extract large quantities of Sb (50-80%) from sediments containing Sb mobilised from mining
38 204 operations, which are associated with Fe-Mn oxyhydroxides and humic acids.^{2, 39, 40} Recovery
39 205 spikes of Sb(III) and Sb(V) indicate that Sb species are stable. This was expected as Sb
40 206 species are complexed with citric acid as they are released. We have had less success in
41 207 extracting Sb from other terrestrial sediments (<10%) and air particulates (<30%) as, unlike
42 208 the mine derived sediments, the Sb in these matrices are probably not associated with Fe/Mn
43 209 oxohydroxides.¹³ Others have also reported the low extraction of Sb from soils using citrate
44 210 extraction⁶⁵ and other extractants.⁶⁰

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48 211 We have had little success with extracting Sb species from plant tissues using a variety of
49 212 extractants (e.g. 11% and 29% extraction efficiency, respectively, for ferns and algae).¹³ The
50 213 trend is for lower extraction efficiencies as the total Sb concentration increases. A sequential
51 214 extraction scheme was designed to remove Sb associated with lipids (CHCl₃-CH₃OH),
52 215 cytosol (CH₃OH-H₂O) and proteins (HNO₃) but only 1-6 % of Sb was extracted, suggesting
53 216 that Sb must be associated with other plant constituents. Wang *et al*⁶⁶ investigated the
54 217 subcellular distribution of Sb in *Fucus tikova* and found that 73-88% Sb was associated with

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3 218 cell walls, with only 8-19% extracted in the cytoplasm. Feng *et al*⁶⁷ found that 43-89% of Sb
4 219 in four ferns was associated cell walls. Thus, these low Sb recoveries are probably a general
5 220 phenomenon in plants with high Sb concentrations. Wang *et al*⁶⁶ postulated that the cell wall
6 221 has many functional groups that can bind with Sb and restricts its transport into cells
7
8 222 protecting the protoplasm from Sb toxicity. The use of enzymes that destroy cell walls may
9 223 prove successful in releasing Sb from plant material. It is also possible, however, that Sb is
10 224 present as nano-particles as reported for other plants that accumulate metals and metalloids.⁶⁸⁻
11 225 ⁷⁰ If so, the use of enzymatic hydrolysis coupled with single particle analysis by ICPMS or
12 226 scanning electron microscopy is required to gain a greater understanding of the nature of Sb
13 227 inclusions or particles. We have had some success in extracting Sb from animal tissues. For
14 228 example, 52% of Sb could be extracted from DOLT-1 Dogfish tissue using CHCl₃-CH₃OH
15 229 (< 0.5%), CH₃OH-H₂O (42%) and HNO₃ (10%). We, however, have not information on the
16 229 (< 0.5%), CH₃OH-H₂O (42%) and HNO₃ (10%). We, however, have not information on the
17 230 stability of Sb species extracted.
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232 ***Question 5: Do you understand the HPLC chromatography for separating Sb species***

233 As indicated previously, the most common HPLC technique used to separate Sb species is
234 based on the use of a PRP-X100 anion exchange column with an EDTA-phthalate mobile
235 phase buffer in the pH 4.5-5 range. Only Sb(III) forms an EDTA complex and Sb(V) is
236 chromatographed as Sb(OH)₆⁻ and appears near the void volume. As the log K of the EDTA-
237 Sb complex is much greater than the log K of the tartrate-Sb complex (24.8 compared to
238 9.4),⁷¹ EDTA also substitutes for Sb(III) standards prepared as the tartrate complex.
239 Optimising buffer strength, pH and temperature only improves the chromatography of the
240 Sb(III) EDTA complex. If citrate is added to standards before chromatography is undertaken,
241 the Sb(III) EDTA complex is still formed as Sb(III)-citrate log K is lower than the Sb(III)-EDTA
242 log K (1.8 and 24.8 respectively at pH 6) and the retention time is similar to using EDTA
243 alone (Fig 10). Sb(OH)₆⁻ forms a citrate complex and is retained longer on the column.
244 Sufficient citric acid needs to be added to fully complex Sb(V). In the literature, it has been
245 reported that two Sb (V) citrate complexes (1:1 and 1:2) can be formed.^{63, 72} We have used 25
246 mM citric acid for extraction purposes and only found one peak.

247 ***Question 6: Are you forming speciation artefacts***

248 As previously mentioned, we have not had success in determining the Sb species in plant
249 tissue extracts. We conducted a series of experiments in which Sb (III) and Sb(V) were added
250 to NIST CRM 1572 citrus leaves pre- and post-extraction. The results revealed considerable
251 change in the oxidation state of Sb and the creation of chromatographic artefacts.¹³ Sb (V)
252 complexed with organic matter in extracts producing a number of Sb peaks in addition to the
253 inorganic Sb(V) peak. Addition of Sb(III) pre-extraction revealed considerable oxidation of
254 Sb(III) to Sb(V) while the addition of Sb(V) pre-extraction resulted in chromatographic
255 artefact formation. The addition of Sb(III) post-extraction resulted in little oxidation of
256 Sb(III) but addition of Sb(V) again resulted in artefact formation. Formation of artefacts by
257 Sb(V) was not surprising as Sb(V) co-elutes with organic material.¹³ Antimony(V) is known

258 to readily complex with organic material in solution.^{73, 74} It may be possible to prevent the
259 oxidation of Sb(III) during extraction by the incorporation of citric acid in the extraction
260 mixture. Amereih *et al*⁷² examined the effect of temperature on conversion of Sb (III) to
261 Sb(V) when citrate was used to extract Sb from soil. As the temperature was raised from 30
262 to 60 °C, a fraction of the Sb-citrate complex was converted to Sb(OH)₆⁺ indicating that
263 although raising the temperature increased extraction efficiencies, Sb oxidation and
264 degradation of Sb-citrate complexes occurred. Gregori *et al*⁵⁴ also reported the complete
265 oxidation of Sb(III) to Sb(V) at 90 °C. As well, many publications have reported poor
266 chromatographic recoveries when analyses are undertaken of Sb species in juices (30%),^{61, 63}
267 urine (51-57%, K and E), yoghurt (73%)⁷⁵ and algal extracts (60%).⁵⁴ As Sb(V) can be
268 chelated to α -hydroxy carboxylic acids^{76, 77} and to vicinal hydroxyl groups,^{73, 75} low
269 chromatographic recoveries are attributed to incorporation of Sb into polymers.⁷³ The
270 prevention of Sb (V) artefact formation may not be possible.

271 For animal tissues, the stability of spiked extracts of DORM-2 during, before and after
272 extraction have been investigated.⁷⁸ After 3 days, there was a significant loss of Sb(III)(40%)
273 and Sb(V)(40-70%) and several new Sb species appeared. The new species formed did not
274 account for the loss of inorganic Sb. Again, artefacts are probably formed with organic
275 constituents in extracts.

276 ***Question 7: Have you considered the use of X-ray absorption spectroscopy?***

277 The analysis of solid samples by synchrotron radiation X-ray absorption near edge structure
278 (XANES) spectroscopy has been shown to be capable of discriminating Sb(III) and Sb(V)
279 oxidation states^{8, 42, 79} and has been used to characterise Sb incorporated into and adsorbed
280 onto various minerals.⁷⁹ Antimony speciation in plants exposed to Sb(III) and Sb(V) has also
281 been reported.⁸⁰ Sample analysis is performed on cryofixed samples at low temperatures,
282 which is claimed to maintain sample integrity.⁸¹ Unfortunately, Sb K-edge (30491 eV)
283 XANES spectra are characterised by absorption edge positions separated by only a few eV
284 due to the core-hole lifetime broadening at this edge. While the first derivative of the
285 normalised energy spectra has been shown to reliably discriminate Sb(III) and Sb(V)
286 coordinated to oxygen,⁸² differentiating Sb(III) coordinated to oxygen and sulphur, for
287 example, requires the use of extended X-ray absorption fine structure spectroscopy (EXAFS),
288 which has poorer detection limits and much longer analysis times than XANES.⁸³ These
289 methods are also unsuitable for measuring methylated Sb species due to their presence at very
290 low concentrations (< 5%). Overall, we recommend that care be taken when using Sb K-edge
291 XAS for environmental samples due to the risk for erroneous interpretation of spectra that
292 only exhibit subtle changes between oxidation states and local coordination environments.
293 Robust statistical assessment of linear combination fit quality (e.g. combinatorial fitting with
294 Hamilton testing, or principal component analysis with target transformation) is critical for
295 ensuring accurate determination of Sb speciation.^{83, 84} Ultimately, we expect that the routine
296 application of high energy resolution fluorescence detected XANES (HERFD-XANES) in the
297 near future will overcome the limitations of conventional Sb XANES by providing far lower
298 detection limits and superior spectral resolution.^{85, 86}

299 Concluding remarks

300 Total Sb concentration measurements in plants, animal tissues and sediment require digestion
301 procedures incorporating HBF₄, HF or HCl to prevent Sb adsorption to silicates. Methods for
302 Sb speciation should be fit for purpose and appropriate methods are available for volatile Sb
303 species (SPME-GCMS), waters (hydride generation-trapping ICPMS) and some sediment
304 extracts (HPLC-ICPMS). Measurement of Sb species in other sediments, soils and plants is a
305 challenge due to low extraction efficiencies, Sb oxidation and formation of chromatographic
306 artefacts. Antimony species can be extracted from sediments with citrate when Sb is
307 associated with Fe-Mn oxohydroxide phases. In our view, it is not possible to measure
308 inorganic Sb species in plants and most sediments and soils by HPLC-ICPMS and solid
309 sample measurement techniques such as X-ray absorption spectroscopy should be considered.
310 These techniques, however, will not be able to quantify methylated Sb species at the low
311 concentrations reported in plants and sediments.

312 Note: Papers cited not freely available can be obtained on request from authors

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3 A variety of approaches for measuring the content and speciation of Sb in environmental
4 matrices are discussed.
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Fig 1 Digestion of environmental samples with various acid mixtures for measuring antimony

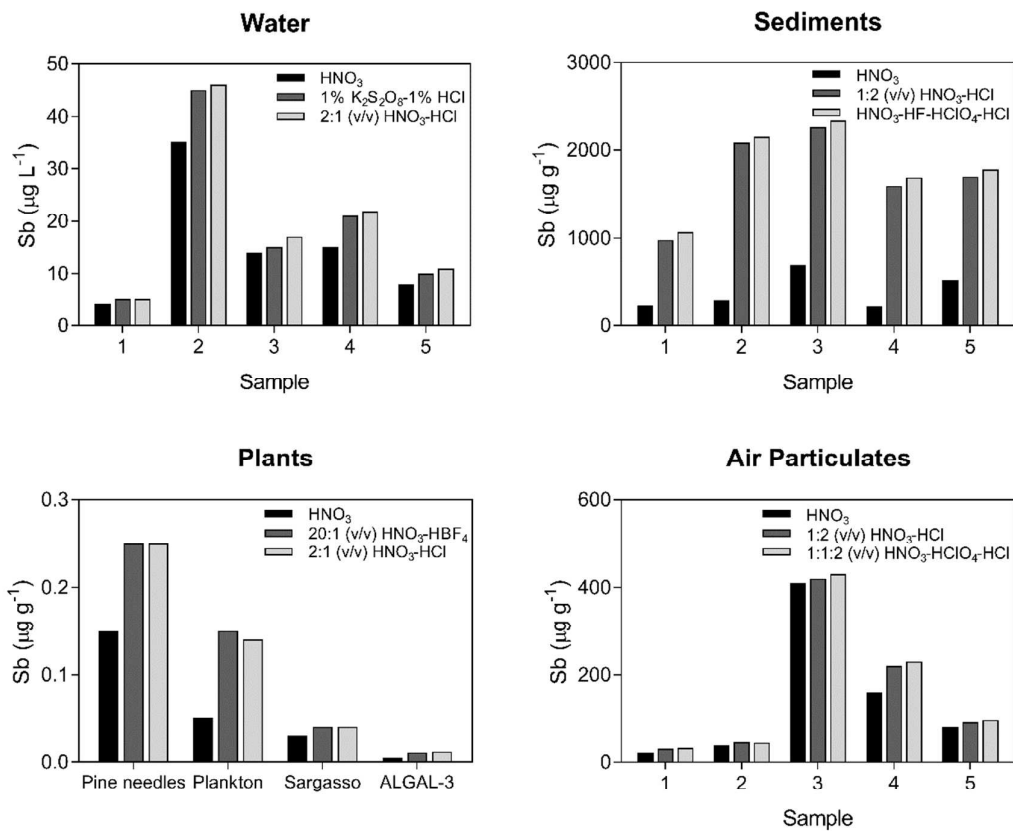


Fig 2 Instrumental techniques for measuring antimony in environmental samples

