



Cite this: RSC Adv., 2017, 7, 37065

# Fate and transformation of nanoparticles (NPs) in municipal wastewater treatment systems and effects of NPs on the biological treatment of wastewater: a review

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Many types of nanoparticles (NPs) have been reported to be toxic to organisms. Wastewater treatment plants are among the important pathways for the migration and transformation of NPs in the environment. Based on a summary of the toxicity of NPs to microorganisms, this review discusses the advances in the fate and transformation of NPs in wastewater treatment plants (WWTPs) and the impacts of NPs on the biological treatment of wastewater. These NPs primarily include Ag NPs, Cu NPs, CuO NPs, ZnO NPs, TiO<sub>2</sub> NPs, Al<sub>2</sub>O<sub>3</sub> NPs, SiO<sub>2</sub> NPs, carbon nanotubes (CNTs) and magnetic NPs. This review observes that the different properties of these NPs result in significant differences in the fate, transformation, antimicrobial activities and effects on biological wastewater treatment. Although many studies have been undertaken in this field, more studies are needed to address the adverse effects of NPs on organic and nutrient removal, as well as their toxic mechanisms.

Received 20th May 2017

Accepted 5th July 2017

DOI: 10.1039/c7ra05690g

rsc.li/rsc-advances

## 1. Introduction

Nanomaterials and nanotechnology have affected human life in many ways. Many types of nanoparticles (NPs), such as carbon nanotubes (CNTs), zinc oxide (ZnO NPs), titanium dioxide (TiO<sub>2</sub> NPs), manganese oxide (MnO<sub>2</sub> NPs), copper oxide (CuO NPs) and copper (Cu NPs), and silver (Ag NPs), have been widely used in such applications as medicines, pesticides, alternative energy, catalysts, and consumer products.<sup>1</sup> For example,<sup>2</sup> Ag

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power than Ag NPs against the oral pathogenic species *Streptococcus mutans*.<sup>32</sup>

Similar to magnetic nanoparticles (MNPs), nano zero-valent iron (nZVI) and magnetite ( $\text{Fe}_3\text{O}_4$ ) are widely used in water treatment. These applications make MNPs likely to enter the environment, and then get into WWTPs.<sup>33</sup> nZVI can penetrate cellular membranes and get into living cells,<sup>34</sup> which results in the accumulation of NPs in living organisms. Two other toxicological studies<sup>35,36</sup> showed that nZVI can rapidly inactivate *Escherichia coli*, *Pseudomonas fluorescens* and *Bacillus subtilis* var. *niger*, but *Aspergillus versicolor* is not affected by them. Some microorganisms can be positively affected by nZVI. For example, methanogens are significantly stimulated by nZVI, and methane production is increased while dechlorinating organisms are inhibited.<sup>37</sup>

There are two types of carbon nanotubes (CNTs), namely, multi-walled carbon nanotubes (MWCNTs) and single-walled carbon nanotubes (SWCNTs). The microbial toxicity of CNTs has been demonstrated under different conditions. The toxicity of CNTs to microorganisms is believed to be related to physical properties (length and diameter), concentration, and types of functionalization.<sup>38–41</sup> It was reported that *Staphylococcus warneri* and *Staphylococcus aureus* are not able to grow over MWCNTs films.<sup>42</sup> Both SWCNTs and MWCNTs at concentrations of  $5 \text{ mg L}^{-1}$  are demonstrated to have antimicrobial activity towards *Escherichia coli*.<sup>43</sup> Long SWCNTs ( $5\text{--}30 \mu\text{m}$ ) can cause significantly greater reductions in the viability of entrapped cells of *Escherichia coli* than short SWCNTs at concentrations of  $5\text{--}200 \text{ mg L}^{-1}$ .<sup>44</sup> When a phenol wastewater treatment system is exposed to SWCNTs ( $2.5 \text{ g L}^{-1}$ ) over an extended period, the structure, richness and diversity of microbial communities are shifted.<sup>45</sup> The effect of MWCNTs on upflow anaerobic sludge blanket (UASB) microflora shows that the reduction in a colony-forming unit (CFU) reaches 29% and 58% at MWCNTs concentrations of 1 and  $100 \text{ mg L}^{-1}$ , respectively.<sup>46</sup>

In summary, most of the NPs mentioned above are toxic to microorganisms at certain concentrations, which makes it necessary to identify the possible adverse effects of them on the microbial activities of activated sludge in wastewater treatment plants.

### 3. Fate and transformation of NPs during the biological treatment of municipal wastewater

NPs released from different nanomaterials are finding their way to wastewater treatment tanks and end up in wastewater sludge.<sup>47–49</sup> It is essential to investigate the fate and transformation of NPs to evaluate the potential environmental impacts on wastewater treatment systems. Many research findings indicate that for the removal of NPs, such as Ag NPs, Cu NPs, ZnO NPs,  $\text{TiO}_2$  NPs, CuO NPs and  $\text{CeO}_2$  NPs, from wastewater, multiple mechanisms are involved, including the physical entrapment of NPs into the floc, active cellular uptake, and binding to extracellular polymers substances (EPS) or

bacteria surfaces.<sup>50,51</sup> Regardless of the mechanisms of removal, most NPs can be removed from suspension with the primary and/or secondary solids that are settled, wasted, and dewatered to form biosolids. Namely, only a very small fraction of NPs are discharged with the treated effluent.<sup>52</sup> For example, it is estimated that only 6% of  $\text{CeO}_2$  NPs are transmitted through a WWTP,<sup>53</sup> and 95% of Cu NPs are removed from wastewater.<sup>54</sup> The predominant mechanisms of copper removal appear to be aggregation and settling (Cu NPs), rather than biosorption. In addition, ZnO NPs are also efficiently removed from wastewater, with an insignificant portion being released into the environment through the treated effluent.<sup>55</sup> A large percentage of ZnO NPs settle out into the sludge. Under wastewater conditions, most NPs aggregate and are attached to the biomass.

Silver can bind strongly to sulphur (both organic and inorganic) in WWTPs, and most Ag NPs are concentrated and then removed.<sup>56</sup> Additionally, most of the Ag NPs in both effluent and sludge are present in the form of  $\text{Ag}_2\text{S}$ .<sup>57</sup> The almost complete removal of citrate-coated Ag NPs is achieved from a simulated sequencing batch reactor (SBR).<sup>58</sup> Ag NPs are efficiently transported into sewer systems without substantial losses to the sewer biofilm.<sup>59</sup> Therefore, Ag NPs can be efficiently removed from wastewater treatment systems.

Raw sewage titanium concentrations range from 181 to  $1233 \mu\text{g L}^{-1}$  in 10 representative WWTPs in the United States. The WWTPs can remove more than 96% of the influent  $\text{TiO}_2$  NPs, and the effluent titanium concentrations are less than  $25 \mu\text{g L}^{-1}$  in all WWTPs.<sup>60</sup>

In many cases, >90% of NPs from sewage appear to be removed, but poor removal efficiencies of  $\text{SiO}_2$  NPs have been reported. For example, the flocculation and sedimentation processes used in primary wastewater treatment are shown to be ineffective in removing  $56 \text{ nm SiO}_2$  NPs at  $2470 \text{ mg L}^{-1}$  due to their stability and slow settling rate.<sup>61</sup> In another study,<sup>62</sup> 96% of  $\text{SiO}_2$  NPs was effectively removed from the wastewater during the first 6 d, while the concentration of  $\text{SiO}_2$  NPs in the effluent gradually increased afterwards, and the NPs discharge was as high as 65% of the input after 30 d of  $\text{SiO}_2$  NPs dosing. The poor removal of the  $\text{SiO}_2$  NPs is related to the high colloidal stability of the  $\text{SiO}_2$  NPs in the wastewater and their limited propensity to biosorption.

There are many environmental factors affecting the migration and transformation of NPs. Dissolution, aggregation and agglomeration are the main factors affecting the state of metal NPs in suspensions.<sup>63</sup> Additionally, the salinity also affects the activated sludge and its microorganisms.<sup>64</sup> These processes will consequently impact the bioavailability, uptake, and toxicity of NPs.<sup>65,66</sup> Various properties of the exposure media can influence the dissolution and aggregation of metal NPs, e.g. ionic strength, pH and the presence of natural organic matters (NOM).<sup>67</sup>

The release of  $\text{Zn}^{2+}$  from ZnO NPs in wastewater is more significant under acidic conditions and low ionic strength, but under alkaline conditions, most of the ZnO NPs show a strong affinity for the sewage sludge rather than dissolved or dispersed in the filtrate.<sup>68,69</sup> It was found that the dissolution or ion release of Cu NPs occurs in wastewater treatment systems.<sup>70</sup> Ag NPs can



**Table 1** Summary of fate and transformation of some NPs during the biological treatment of municipal wastewater

| NPs                  | Transformation  | Removal efficiency                       | References                       |
|----------------------|---|--|----------------------------------|
| Ag NPs               | Release $\text{Ag}^+$ , $\text{Cu}^{2+}$ , $\text{Zn}^{2+}$ | High removal efficiency for NPs          | 50, 51, 54–59, 63, 64, 68, 70–72 |
| Cu NPs               |   | Low removal efficiency for released ions |                                  |
| ZnO NPs              |   |  |                                  |
| TiO <sub>2</sub> NPs | No transformation found                                     | High removal efficiency                  | 50, 51–54, 60, 65, 67            |
| CuO NPs              |   |  |                                  |
| CeO <sub>2</sub> NPs |   |  |                                  |
| SiO <sub>2</sub> NPs | No transformation found                                     | Low removal efficiency                   | 61, 62                           |

be transformed into  $\text{Ag}^+$  under certain conditions, which were found to be more toxic than Ag NPs to the microbial communities in activated sludge and the nitrification process. However, DOM can significantly stabilize Ag NPs.<sup>71</sup> Sulphidation can lead to Ag NPs and ZnO NPs aggregation.<sup>72</sup> ZnO NPs are slightly soluble and can dissolve to form  $\text{Zn}^{2+}$  under aqueous conditions. This finding suggests that the released  $\text{Zn}^{2+}$  is responsible for the toxicity of ZnO NPs to *Mycobacterium smegmatis*<sup>73</sup> and *Pseudokirchneriella subcapitata*.<sup>74</sup> However, *Shewanella oneidensis* MR-1 and *Escherichia coli* are insensitive to the released  $\text{Zn}^{2+}$  even though the concentrations of ZnO NPs are higher than  $40 \text{ mg L}^{-1}$ .<sup>73</sup>

Overall, the data collected indicate that biological WWTPs operated using activated sludge can effectively remove most types of NPs from wastewaters. Such NPs as Ag NPs, ZnO NPs, and TiO<sub>2</sub> NPs can be efficiently removed through their interaction with biomass. Ag NPs, Cu NPs and ZnO NPs would be partially transformed into  $\text{Ag}^+$ ,  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$ , respectively, under certain conditions. The presence of both EPS and ionic compounds can effectively enhance the NPs removal. The instability of the NPs resulting from the ionic strength in sewage and the entrapment of NPs by EPS play important roles in NPs removal by activated sludge. Table 1 is a summary of the fate and transformation of NPs during the biological treatment of municipal wastewater.

## 4. Effects of NPs on performances of biological treatment of municipal wastewater

### 4.1. Effects of NPs on nitrification and denitrification

Many studies have been conducted on the effects of Ag NPs on functional microorganisms and microbial communities in biological wastewater treatment processes. For example, among nitrogen-cycling bacteria, the nitrifier *Nitrosomonas europaea* is more sensitive to Ag NPs when the concentration is higher than  $1.44 \text{ mg L}^{-1}$  compared to the denitrifier *Pseudomonas stutzeri* and the nitrogen fixer *Azotobacter vinelandii*.<sup>75</sup> Ag NPs at concentrations  $>2 \text{ mg L}^{-1}$  can induce damage of the cell wall and membrane of *Nitrosomonas europaea*, and disorganize the nucleoids.<sup>76</sup> In addition, Ag NPs can inhibit important protein functions, including biosynthesis, energy production, gene expression and nitrification in the test media at  $5 \text{ mg L}^{-1}$ .<sup>77</sup> García *et al.*<sup>11</sup> reported the inhibitory effects of Ag NPs on the

activity of microbial communities in activated sludge systems through respiration tests and biogas-production analysis. Ag NPs at the concentration of  $0.1 \text{ mg mL}^{-1}$  cause an intermediate inhibition in biogas production (within 33–50%). In biofilm systems, the original wastewater biofilms are highly tolerant to Ag NPs. With the application of  $200 \text{ mg L}^{-1}$  Ag NPs, the reduction in biofilm bacteria measured by heterotrophic plate counts is insignificant after 24 h. However, after the removal of loosely bound EPS, the viability of wastewater biofilms is reduced when treated under the same conditions.<sup>78</sup>

Most of the results indicate that low concentrations of Ag NPs (*e.g.*,  $<2 \text{ mg L}^{-1}$ ) do not have significant effects on nitrogen removal in activated sludge systems. However, Ag NPs only have a mild impact on  $\text{NH}_4^+$ -N removal, but the denitrification-related species are inhibited by  $0.5 \text{ mg L}^{-1}$  Ag NPs, including *Diaphorobacter* species, *Thauera* species and those in the *Sphaerotilus-Leptothrix* group.<sup>79</sup> Ag NPs at concentrations  $>2 \text{ mg L}^{-1}$  disrupt ammonia removal in activated sludge in an SBR, while Ag-ligand formation and microbial adaptation (chloride or sulphide groups) can reduce their overall impacts.<sup>80</sup> When the concentration reaches  $2 \text{ mg L}^{-1}$  of total Ag, biological nutrient removal can be disrupted, whereas Ag NPs do not exhibit a significant effect on ammonia removal during a  $2 \text{ mg L}^{-1}$  spike. However, nitrification recovers quickly, especially at lower concentrations of Ag NPs. In addition, Doolette *et al.*<sup>81</sup> also suggested that the nitrification rate is not affected by Ag NPs at the concentration of  $2.5 \text{ mg L}^{-1}$  in an SBR. Ag NPs are found to be transformed into  $\text{Ag}_2\text{S}$  in activated sludge, and  $\text{Ag}_2\text{S}$  has no adverse effects on nitrification. However, a change in niche populations occur in aerobic sludge, indicating Ag NPs can negatively affect sub-dominant wastewater microbial communities.

Additionally,<sup>82</sup> the continuous loading of Ag NPs ( $0.1 \text{ mg L}^{-1}$ ) for more than 60 d does not show adverse impacts on bacterial activity in a membrane bioreactor (MBR) activated sludge system. The nitrifying bacterial community structure is stable after long-term exposure to Ag NPs, while an abundance of the silver resistance gene *silE* in the MBR is increased after 41 d exposure, and then decreased with continuous exposure. These two results above indicate that activated sludge can effectively reduce the toxicity of Ag NPs by microbial adaptation and adsorbing or precipitating Ag NPs.

In summary, low concentrations of Ag NPs ( $<2 \text{ mg L}^{-1}$ ) mostly do not have significant effects on nitrification and denitrification in conventional activated sludge systems, while





Ag NPs at high concentrations ( $>2 \text{ mg L}^{-1}$ ) can mostly inhibit nitrogen removal with long-term exposure.

Compared with Ag NPs, only a small number of studies have been conducted on the impact of Cu NPs on activated sludge. At concentrations from  $0.1$  to  $10 \text{ mg L}^{-1}$ , TN removal is promoted (improved from  $60.6\%$  to  $>70\%$ ) and  $\text{N}_2\text{O}$  generation is reduced, but ammonia removal is not affected.<sup>83</sup> It was found that most Cu NPs are absorbed onto activated sludge, but the activated sludge surface is not damaged. The released  $\text{Cu}^{2+}$  is the main cause of TN removal improvement and  $\text{N}_2\text{O}$  reduction.

It was also found that the activities of ammonia mono-oxygenase and nitrite oxidoreductase are not affected by Cu NPs, whereas the activities of denitrification enzymes are increased, and Cu NPs increase the number of denitrifiers but decrease nitrite accumulation. All these results are in correspondence with the enhancement of TN removal and the decrease in  $\text{N}_2\text{O}$  production. Overall, Cu NPs do not show much of a negative impact on biological nitrogen removal. Short-term exposure to Cu NPs causes slight toxicity to ammonia oxidizing bacteria, but TN removal is promoted and  $\text{N}_2\text{O}$  generation is reduced by long-term exposure at certain concentrations of Cu NPs.

Several studies reveal the impacts of ZnO NPs on nitrification and denitrification. For example, short-term exposure to ZnO NPs can decrease TN removal from  $81.5\%$  to  $75.6\%$  and  $70.8\%$ , respectively, at the concentrations of  $10$  and  $50 \text{ mg L}^{-1}$ .<sup>84</sup> It was suggested that the inhibition of nitrogen removal mainly results from the release of  $\text{Zn}^{2+}$  from ZnO NPs dissolution and the increase of ROS production, which decreases nitrate reductase (NAR) activity. Similarly, the denitrification-related species were inhibited by  $5 \text{ mg L}^{-1}$  ZnO NPs, including *Diaphorobacter* species, *Thauera* species and those in the *Sphaerotilus-Leptothrix* group, although the  $\text{NH}_4^+\text{-N}$  removal efficiency was either slightly reduced or not reduced at all.<sup>85</sup> In addition, ZnO NPs slightly decrease TN removal efficiency and reduce  $\text{NH}_4^+\text{-N}$  removal by inhibiting the respiration of nitrifying microorganisms.<sup>86</sup> Overall, ZnO NPs affect TN removal efficiency by inhibiting the activity of NAR and the denitrification related to bacteria.

Several reports have described the impact of  $\text{TiO}_2$  NPs on nitrification and denitrification. The concentrations of  $1$  and  $50 \text{ mg L}^{-1}$   $\text{TiO}_2$  NPs do not have acute effects on wastewater nitrogen removal after short-term exposure ( $1 \text{ d}$ ) in an anaerobic-low dissolved oxygen SBR, whereas  $50 \text{ mg L}^{-1}$   $\text{TiO}_2$  NPs significantly decreases TN removal efficiency by  $56\%$  after  $70 \text{ d}$  exposure. Mechanism analysis reveals that  $\text{TiO}_2$  NPs at  $50 \text{ mg L}^{-1}$  obviously reduces the abundance of nitrifying bacteria and the diversity of microbial communities in activated sludge. In particular, ammonia-oxidizing bacteria are greatly decreased after long-term exposure to  $50 \text{ mg L}^{-1}$   $\text{TiO}_2$  NPs, resulting from the serious deterioration of ammonia oxidation. In addition,  $\text{TiO}_2$  NPs at  $50 \text{ mg L}^{-1}$  can inhibit the activities of ammonia mono-oxygenase (AMO) and nitrite oxidoreductase (NOR) after long-term exposure.<sup>87</sup>  $\text{TiO}_2$  NPs reduce the microbial diversity in activated sludge at concentrations from  $100 \text{ mg L}^{-1}$  to  $200 \text{ mg L}^{-1}$ , and the abundance of denitrifying bacteria is obviously decreased, which explains the inhibition.

Summarily, exposure time is one of the important factors of the adverse effects of  $\text{TiO}_2$  NPs. Low concentrations (e.g.  $<50 \text{ mg L}^{-1}$ ) of  $\text{TiO}_2$  NPs do not have acute effects on wastewater nitrogen removal after short-term exposure, but high concentrations (e.g.  $100\text{--}200 \text{ mg L}^{-1}$ ) of  $\text{TiO}_2$  NPs can markedly decrease TN removal efficiency. Different from short-term exposure,  $50 \text{ mg L}^{-1}$   $\text{TiO}_2$  NPs can significantly decrease TN removal efficiency after long-term exposure.  $\text{TiO}_2$  NPs inhibit both nitrifying bacteria and denitrifying bacteria, as well as AMO and NOR.

Only a small number of reports are found on the impact of  $\text{Al}_2\text{O}_3$  NPs on wastewater nitrogen removal. Most of the  $\text{Al}_2\text{O}_3$  NPs can adsorb onto activated sludge, but  $\text{Al}_2\text{O}_3$  NPs have no adverse effects on the surface integrity and viability of activated sludge.<sup>88</sup> Additionally, short-term exposure to  $\text{Al}_2\text{O}_3$  NPs does not present adverse impacts on wastewater nitrification and denitrification. However, the long-term exposure to  $50 \text{ mg L}^{-1}$   $\text{Al}_2\text{O}_3$  NPs can decrease the TN removal efficiency by  $17.9\%$ , owing to a suppressed denitrification process. The abundance of denitrifying bacteria is also decreased, and the activities of key denitrifying enzymes, such as nitrite reductase (NIR), are inhibited. Therefore, only long-term exposure to  $\text{Al}_2\text{O}_3$  NPs has adverse impacts on wastewater nitrification and denitrification.

The acute and chronic effect of  $\text{SiO}_2$  NPs on the nutrient removal performance of activated sludge was reported. Zheng *et al.*<sup>89</sup> noted that the presence of  $\text{SiO}_2$  NPs at an environmentally relevant concentration (e.g.  $1 \text{ mg L}^{-1}$ ) caused no adverse acute and chronic effects on sludge viability and wastewater nitrogen removal. Nevertheless, long-term exposure to  $50 \text{ mg L}^{-1}$   $\text{SiO}_2$  NPs can depress the TN removal efficiency by  $18\%$  after  $70 \text{ d}$  exposure, resulting from the declined activities of denitrifying enzymes and NIR. Further research revealed that the microbial community structure is changed, and the abundance of denitrifying bacteria is reduced after chronic exposure. In summary,  $\text{SiO}_2$  NPs at low concentrations cause no significant adverse effects on nitrogen removal, but high concentrations of  $\text{SiO}_2$  NPs affect TN removal efficiency.

We identified two reports regarding the most widely used magnetic NPs (nZVI, magnetite ( $\text{Fe}_3\text{O}_4$ ) and maghemite ( $\gamma\text{-Fe}_2\text{O}_3$ ) NPs). Wu *et al.*<sup>90</sup> examined the impacts of nZVI on nitrogen removal under continuous aerobic/anaerobic conditions in an activated sludge system and found no measurable effect of nZVI on nitrogen removal at the concentrations of  $50 \text{ mg L}^{-1}$  and below. However,  $200 \text{ mg L}^{-1}$  of nZVI inhibited  $\text{NH}_4^+\text{-N}$  removal. The microbial activities were inhibited on exposure to nZVI; the adenosine triphosphate (ATP) content was decreased by approximately  $13\%$ ,  $31\%$  and  $43\%$  at the nZVI concentrations of  $20$ ,  $50$ , and  $200 \text{ mg L}^{-1}$ , respectively; and ROS production was increased. Lower concentrations of nZVI ( $20$  and  $50 \text{ mg L}^{-1}$ ) boosted the microbial activity, whereas  $200 \text{ mg L}^{-1}$  nZVI caused an approximate  $19\%$  depression in dehydrogenase activity. In addition, the acute exposure to  $50\text{--}200 \text{ mg L}^{-1}$   $\text{Fe}_3\text{O}_4$  NPs can decrease TN removal efficiencies, resulting from the inhibition of nitrifying bacteria and the secretion of EPS.<sup>91</sup> In contrast, chronic exposure to  $50 \text{ mg L}^{-1}$  magnetic NPs can clearly improve TN removal, resulting from the increased abundance of nitrifying bacteria and the improved activities of



| Types of effects  | NPs   | Comments   | Concentrations  | References           |
|---|---|--|---|----------------------|
| Effects of NPs on nitrification and denitrification                   | Ag NPs  | <i>Nitrosomonas europaea</i> is more sensitive than denitrifier <i>Pseudomonas stutzeri</i> and the nitrogen fixer <i>Azotobacter vinelandii</i> | >1.44 mg L <sup>-1</sup>  | 73                   |
|   |   | Cause damage of cell wall and the membrane of <i>Nitrosomonas europaea</i> and make the nucleoids disorganized; disrupt ammonia removal          | >2 mg L <sup>-1</sup>   | 74, 78               |
|   |   | Inhibit important protein functions, including biosynthesis, energy production, gene expression and nitrification                                | 5 mg L <sup>-1</sup>  | 75                   |
|   |   | Cause an intermediate inhibition in biogas production (within 33–50%)  | 0.1 mg mL <sup>-1</sup>   | 10                   |
|   |   | Activated sludge can reduce Ag NPs toxicity  | >2 mg L <sup>-1</sup>   | 78, 79               |
|   | Cu NPs  | TN removal is promoted and N <sub>2</sub> O generation is reduced; ammonia removal is not affected   | 0.1–10 mg L <sup>-1</sup>   | 81                   |
|   | ZnO NPs   | Decreases TN removal   | 10 and 50 mg L <sup>-1</sup>  | 82                   |
|   |   | Denitrification species are inhibited; slightly decrease TN removal  | 5 mg L <sup>-1</sup>  | 83, 84               |
|   | TiO <sub>2</sub> NPs  | Reduce the diversity of microbial community in activated sludge the abundance of nitrifying bacteria   | >50 mg L <sup>-1</sup> (long-term); 100–200 mg L <sup>-1</sup> (short-term) | 21, 85               |
|   | Al <sub>2</sub> O <sub>3</sub> NPs                                    | No adverse effects on TN removal efficiency on (short-term exposure); decrease TN removal efficiency (long-term exposure)                        | 50 mg L <sup>-1</sup> ; 50 mg L <sup>-1</sup>                               | 86                   |
| Effects of NPs on biological phosphorus removal                       | SiO <sub>2</sub> NPs  | No adverse acute and chronic effects on nitrogen removal; depresses TN removal (long-term exposure)  | 1 mg L <sup>-1</sup> ; 50 mg L <sup>-1</sup>                                | 87                   |
|   | nZVI  | No measurable effect on nitrogen removal; inhibit NH <sub>4</sub> <sup>+</sup> -N removal  | ≤50 mg L <sup>-1</sup> ; 200 mg L <sup>-1</sup>                             | 88                   |
|   | Fe <sub>3</sub> O <sub>4</sub> and Fe <sub>2</sub> O <sub>3</sub> NPs | Decreases TN removal (short-term exposure)   | 50–200 mg L <sup>-1</sup>   | 89                   |
|   | Ag NPs  | No significant effect on biological phosphorus removal; Ag <sup>+</sup> shows stronger toxicity (>1 mg L <sup>-1</sup> )                         | <5 mg L <sup>-1</sup>   | 85, 90, 91           |
| Effects of NPs on organic-degradation microbial community             | ZnO NPs   | Inhibition of phosphorus removal is due to the release of zinc ions  | <50 mg L <sup>-1</sup>  | 82                   |
|   | TiO <sub>2</sub> NPs  | No acute effects on phosphorus removal   | <50 mg L <sup>-1</sup>  | 22                   |
|   | Al <sub>2</sub> O <sub>3</sub> NPs                                    | Induce marginal influences on wastewater phosphorus removal  | <50 mg L <sup>-1</sup>  | 86                   |
|   | SiO <sub>2</sub> NPs  | Phosphorus removal is insensitive  | <50 mg L <sup>-1</sup>  | 87                   |
|   | Cu NPs  | Phosphorus removal is unaffected   | 0.1–10 mg L <sup>-1</sup>   | 81                   |
|   | nZVI  | Improves phosphorous removal by boosting the microbial activity  | 20 mg L <sup>-1</sup>   | 88                   |
|   | Effects of NPs on organic-degradation microbial community             | Ag NPs   | Impact the activated sludge microbial community and cell culturability      | 1 mg L <sup>-1</sup> |
| Methane production are not affected                                   |   |  | 2.9 g per (kg TS)   | 79                   |
| COD removal is moderately affected and can be recovered significantly |   |  | 0.2 and 2 mg L <sup>-1</sup>  | 77                   |
| COD removal will diminish as the increased addition of Ag NPs         |   |  | 50 mg L <sup>-1</sup>   | 93                   |
| Negligible impact on anaerobic digestion and methanogenic assemblages |   |  | ≤40 mg L <sup>-1</sup>  | 49                   |
| ZnO NPs   |   | Highly toxic and rapidly disrupt the methanogenic activity   | 34.5 mg Zn per L  | 94                   |
|   |   | Inhibit the microbial activities in the outer layer of biofilms  | 50 mg L <sup>-1</sup>   | 95                   |
|   |   | EPS of AGS and the methane production are not significantly affected; EPS of AGS and the methane production are decreased                        | <50 mg g <sup>-1</sup> -TSS; 200 mg g <sup>-1</sup> -TSS                    | 96                   |



Table 2 (Contd.)

| Types of effects | NPs                  | Comments   | Concentrations         | References |
|------------------|----------------------|--|------------------------|------------|
|                  | CuO NPs              | Marked negative effect on the performance of methanogenesis in UASB reactor                      | 1.4 mg Cu per L        | 99         |
|                  | Cu NPs               | Cause inhibition of acetoclastic methanogens but not H <sub>2</sub> -utilizing methanogens       | 62 mg L <sup>-1</sup>  | 97         |
|                  |                      | No significant effect on sludge fermentative volatile fatty acids (VFA)                          | 50 mg L <sup>-1</sup>  | 100        |
|                  | SiO <sub>2</sub> NPs | Affect the biological activity in activated sludge through the inhibition of total oxygen uptake | >50 mg L <sup>-1</sup> | 101        |

NAR and NIR. The release of iron ions from the dissolution of magnetic NPs may also be responsible for the increase in bacteria activities. In summary, magnetic NPs have little adverse effects on nitrification and denitrification at  $\leq 50$  mg L<sup>-1</sup> of magnetic NPs, whereas high concentrations of NPs can cause a depression in dehydrogenase activity.

Overall, exposure time (short-term or long-term) is one of the important factors affecting the adverse effects of many types of NPs. The adverse effects of many types of NPs are of concern for the nitrification and denitrification in activated sludge. Most NPs at certain concentrations exhibit some adverse effects on nitrogen removal by inhibiting the key enzymes of nitrifying and denitrifying bacteria. Some adverse effects result from the release of ions from dissolved NPs.

#### 4.2. Effects of NPs on biological phosphorus removal

Some studies have been reported on the effects of NPs on biological phosphorus removal. It was reported<sup>92</sup> that Ag NPs ( $<5$  mg L<sup>-1</sup>) do not present obvious effects on biological phosphorus removal in both acute and chronic exposure. However, the removal rate of phosphorus was decreased to 48.8% at 1 mg L<sup>-1</sup> of Ag<sup>+</sup> at short-term exposure, and almost zero phosphorus removal was achieved when the concentration of Ag<sup>+</sup> was greater than 2 mg L<sup>-1</sup>. Ag<sup>+</sup> shows stronger toxicity to polyphosphate accumulating organisms (PAOs) than glycogen accumulating organisms (GAOs). In addition, the decrease in anaerobic phosphorus release by Ag<sup>+</sup> is caused by the inhibition of the activities of adenylate kinase and exopolyphosphatase (PPX), whereas the decrease in aerobic phosphorus uptake is due to the suppression of energy generation for phosphorus uptake. A comparison was conducted on the effects of Ag NPs and Ag<sup>+</sup> at concentrations from 1 to 5 mg L<sup>-1</sup> on the performance of enhanced biological phosphorus removal (EBPR). Ag NPs showed no obvious effect on EBPR at both acute and chronic exposure,<sup>93</sup> whereas phosphorus removal was decreased by a sudden increase in Ag<sup>+</sup> concentration and was gradually recovered after long-term culture. The mechanism study showed that the transformation of polyhydroxyalkanoates (PHA) and glycogen, and the activity of PPX decline when Ag<sup>+</sup>

was suddenly increased, but they were recovered after long-term culture.<sup>82</sup>

The effects of NPs such as ZnO, TiO<sub>2</sub>, Al<sub>2</sub>O<sub>3</sub>, and SiO<sub>2</sub> NPs on biological phosphorus removal are of concern. ZnO NPs at concentrations of 10 and 50 mg L<sup>-1</sup> can inhibit phosphorus removal and the inhibition result from the release of zinc ions from ZnO NPs dissolution, as well as ROS production, causing an inhibitory effect on PAOs and decreased activities of PPX and polyphosphate kinase (PPK).<sup>84</sup> It was reported that 1 and 50 mg L<sup>-1</sup> TiO<sub>2</sub> NPs do not exhibit significant impacts on phosphorus removal at both acute exposures (1 d) and chronic exposure (70 d).<sup>22</sup> Short-term exposure to 1 and 50 mg L<sup>-1</sup> Al<sub>2</sub>O<sub>3</sub> NPs can induce marginal influence on wastewater phosphorus removal.<sup>88</sup> In addition, transformations of intracellular PHA and glycogen were not significantly changed in the presence of 1 and 50 mg L<sup>-1</sup> Al<sub>2</sub>O<sub>3</sub> NPs.

Zheng *et al.* reported that wastewater phosphorus removal is insensitive to 1 and 50 mg L<sup>-1</sup> SiO<sub>2</sub> NPs at both acute and chronic exposures because the important factors related to phosphorus removal, such as the activities of PPX and PPK and the intracellular transformations of PHA and glycogen, were not obviously changed.<sup>89</sup> As illustrated above,<sup>83</sup> the presence of Cu NPs in activated sludge affects the nitrogen removal by increasing the activity of NAR and NIR while the effects of Cu NPs are insignificant on the activities of AMO, NOR, PPX, and PPK. This means that phosphorus removal is unaffected by Cu NPs.<sup>83</sup> As illustrated above,<sup>90</sup> nZVI does not exhibit any measurable influence on nitrogen removal at 50 mg L<sup>-1</sup> and below, whereas the addition of nZVI at 10–20 mg L<sup>-1</sup> significantly improves phosphorus removal by boosting microbial activity.<sup>90</sup>

In summary, the inhibition of phosphorus removal induced by ZnO NPs is due to the release of zinc ions from ZnO NPs dissolution, as well as ROS production. Most other NPs, such as Cu NPs, TiO<sub>2</sub> NPs, Al<sub>2</sub>O<sub>3</sub> NPs and SiO<sub>2</sub> NPs, do not present significantly adverse effects on biological phosphorus removal at the concentrations of 50 mg L<sup>-1</sup> and below, because they have no significant impacts on the activities of PPX and PPK, and the transformations of intracellular PHA and glycogen. Low



concentrations of nZVI ( $10\text{--}20\text{ mg L}^{-1}$ ) can improve phosphorous removal.

#### 4.3. Effects of NPs on organic-degradation microbial communities

Many studies have commenced investigations into the effects of NPs on activated sludge microbial communities. Ag NPs ( $1\text{ mg L}^{-1}$ ) can affect activated sludge microbial communities and cell culturability depending on the physical structure of the activated sludge flocs, the spatial distribution of microorganisms in activated sludge flocs, and the community structures in the activated sludge.<sup>94</sup> Doolette *et al.*<sup>81</sup> found methane production was not affected by Ag NPs ( $\sim 3\text{ g kg}^{-1}\text{-TS}$ ) in an anaerobic batch test. Ag NPs were found to be transformed into  $\text{Ag}_2\text{S}$  in activated sludge, and transformed Ag NPs have no adverse effects on methanogenesis. However, a change in niche populations was found, suggesting that Ag NPs may adversely impact sub-dominant wastewater microbial communities. In addition, Chen *et al.*<sup>79</sup> suggested that COD removal is moderately affected by low concentrations of Ag NPs. When Ag NPs were added at  $0.2$  and  $2\text{ mg L}^{-1}$ , the removal percentages were decreased from 99% to approximately 80%, but COD removal was remarkably recovered 3 d later, and a COD removal efficiency of approximately 93% was achieved. Additionally, a remarkable change occurred in the microbial community structure, which can be interpreted as an improvement in recovery. However, it was reported that the COD removal rate will decrease with the increased addition of Ag NPs and that only 20% of COD is removed at an Ag NPs concentration of  $50\text{ mg L}^{-1}$ .<sup>95</sup> Yang *et al.* suggested that Ag NPs at moderate concentrations ( $\leq 40\text{ mg L}^{-1}$ ) have negligible effects on anaerobic digestion and methanogenic assemblages,<sup>49</sup> whereas  $50\text{ mg L}^{-1}$  Ag NPs significantly decreased the COD removal efficiency.

ZnO NPs at  $34.5\text{ mg L}^{-1}$  are highly toxic and rapidly disrupt the methanogenic activity during UASB operation, and the toxicity was more severe for acetoclastic methanogens compared with hydrogenotrophic methanogens.<sup>96</sup> In addition, sub-ppm levels of ZnO NPs present a remarkable effect over the long-term, as evidenced by the gradual decrease in the methanogenic activity of the biomass. This finding may be attributable to two factors: the enhanced toxicity with exposure time, and the accumulation of ZnO NPs during wastewater treatment that may increase the effective concentration of ZnO NPs in direct contact with the microorganisms in sludge. Also,  $50\text{ mg L}^{-1}$  ZnO NPs can inhibit the microbial activities only in the outer layer of biofilms, and bacteria present in the deeper parts of biofilms become even more active.<sup>97</sup> Methane production was not obviously affected by ZnO NPs at the concentrations of  $10$  and  $50\text{ mg g}^{-1}\text{-TSS}$ , but it was decreased when ZnO NPs concentration was higher than  $100\text{ mg g}^{-1}\text{-TSS}$ , because ZnO NPs caused a decrease of proteins in EPS by 69.6%.<sup>98</sup> Additionally, the C–O–C group of polysaccharides and carboxyl group of proteins in EPS were also changed. The average methane production was decreased by 43.5% at a ZnO NPs concentration of  $200\text{ mg g}^{-1}\text{-TSS}$ . ZnO NPs can also cause severe inhibition of methanization.

Cu NPs can cause inhibition of acetoclastic methanogens, but not of  $\text{H}_2$ -utilizing methanogens at a Cu NPs concentration of  $62\text{ mg L}^{-1}$ .<sup>99</sup> Therefore, ZnO NPs, CuO NPs and Cu NPs can induce inhibition of methane production during anaerobic biological wastewater processes. The production of ROS and the release of LDH showed that exposing the activated sludge to CuO NPs caused oxidative stress and damaged cell membranes in the sludge flocs and the production of EPS; in particular, polysaccharides were enhanced by 89.7% with exposure to  $50\text{ mg L}^{-1}$  CuO NPs.<sup>100</sup> Long-term exposure of the microorganisms retained in the sludge bed to CuO NPs ( $1.4\text{ mg Cu per L}$ ) can cause significant inhibition of *Acetoclastic methanogenesis*, which eventually leads to the complete disruption of the bioreactor performance.<sup>101</sup> The inhibition is likely induced by both CuO NPs and released copper ions.

The addition of Cu NPs to biological wastewater treatment systems presented no significant effects on sludge fermentative volatile fatty acids (VFA), the preferred carbon source for wastewater biological nutrient removal production, as the sludge solubilization was increased, and the acidification was decreased.<sup>102</sup> When Cu NPs are directly added to the anaerobic fermentation reactor, both the hydrolysis and acidification are inhibited at a Cu NPs concentration of  $50\text{ mg L}^{-1}$ , resulting in an ultimate VFA production much less than that of the control by approximately 45% after 10 d exposure.

The adverse impacts of  $\text{SiO}_2$  NPs on activated sludge was studied using the activated sludge respiration inhibition test and it was suggested that  $\text{SiO}_2$  NPs affect the biological activity in activated sludge through the inhibition of total oxygen uptake at concentrations higher than  $50\text{ mg L}^{-1}$ .<sup>103</sup> Transmission Electron Microscopy (TEM) image analysis suggests some of the  $\text{SiO}_2$  NPs are adsorbed on and/or apparently embedded in the microbial cell membrane.

Summarily, most NPs at certain concentrations have adverse effects on microorganisms for organic degradation in activated sludge (Table 2). After exposure, a degree of inhibition might occur to microbial communities and to the activities of certain key enzymes in activated sludge, subsequently resulting in the abnormal performance of activated sludge systems.

## 5. Conclusions and future directions

Many studies have been conducted on the fate of NPs, as well as their effects on biological wastewater treatment, and many achievements have been reported. It can be concluded from this review as follows:

(1) Most NPs, such as Ag, ZnO, CuO,  $\text{Al}_2\text{O}_3$ ,  $\text{SiO}_2$ , CNTs and magnetic NPs, can cause varying degrees of damage to microorganisms at certain concentrations. Ag, Cu and ZnO NPs present relatively significant toxic effects compared with other NPs at similar exposure concentrations. In particular,  $\text{TiO}_2$  NPs do not show strong toxicity to microorganisms at both short-term (even at  $500\text{ mg L}^{-1}$ ) and long-term exposures ( $<50\text{ mg L}^{-1}$ ).

(2) WWTPs operated using activated sludge have the potential to remove most types of NPs, such as Ag, Cu, ZnO, CuO, and  $\text{TiO}_2$ , but not  $\text{SiO}_2$ , by interaction with biomass. Ag NPs, Cu NPs





and ZnO NPs can be partially transformed into  $\text{Ag}^+$ ,  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$ , respectively, under certain conditions. The poor removal of  $\text{SiO}_2$  NPs is due to the high colloidal stability of  $\text{SiO}_2$  NPs in the wastewater and their limited propensity to biosorption.

(3) Most NPs have varying degrees of impacts on the performance of the biological treatment of wastewater, including nitrogen, phosphorus and organic pollutants removal. The impacts are dose-dependent under most conditions, and the exposure time (short-term or long-term) is also an important factor of adverse effects. (a) For nitrogen removal, Ag, CuO and ZnO NPs have slight or moderate effects on TN removal at relatively low concentrations;  $\text{TiO}_2$  NPs show some inhibitory effects on nitrifying and denitrifying bacteria, and AMO and NOR; and  $\text{Al}_2\text{O}_3$  and  $\text{SiO}_2$  NPs have adverse impacts on nitrification and denitrification. (b) For phosphorus removal,  $\text{Al}_2\text{O}_3$ ,  $\text{TiO}_2$  and  $\text{SiO}_2$  NPs do not have significant adverse effects, but ZnO NPs can result in net phosphorus removal failure at certain concentrations. (c) For organics removal, Ag NPs can decrease COD removal. ZnO NPs and CuO NPs can inhibit methane production during anaerobic biological wastewater processes. On the whole, the impacts are dose-dependent under most conditions. Some of the impacts are due to the NPs themselves, and some of impacts result from released ions such as  $\text{Ag}^+$ ,  $\text{Cu}^{2+}$ , and  $\text{Zn}^{2+}$ .

Although many studies have been conducted in this field, the achievements are far from complete. For instance, as to Au NPs and CNTs, no reports are found on their effects on biological nitrogen and phosphorus removal. There are only a few studies found that report the effects of NPs on organics removal. NPs removed from sewage will accumulate in biosolids, and more research is required to understand the fate of NPs upon the ultimate disposal of biosolids. The effects of NPs on biosolids disposal process also need more research. In addition, long-term effects of NPs at low concentrations on the functional microorganisms need more concern.

## Acknowledgements

The Project is sponsored by the Scientific Research Foundation for the Returned Overseas Chinese Scholars, State Education Ministry, China. This work is also supported by the Key Laboratory of Urban Water Resource and Environment of Harbin institute of technology, China (ES201608).

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