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Unrecognized Volatile and Semi-Volatile Organic Compounds from Brake Wear

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Electronic Supplementary Information includes

Materials and Methods

Figs. S1-S13

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References

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34 **Environmental Significance**

35 To mitigate climate change and improve air quality, conventional combustion-powered vehicles
36 are being replaced by zero tailpipe emissions vehicles. We show that in addition to well-
37 documented emissions of particles, automotive braking also emits a complex mixture of volatile
38 organic gases that include Hazardous Air Pollutants, climate active species and gases known to
39 participate in the degradation of air quality. Thus, vehicles will continue to contribute to air
40 quality and climate problems once tailpipe emissions are eliminated.

41

42 **Abstract**

43 Motor vehicles are among the major sources of pollutants and greenhouse gases in urban areas
44 and a transition to “zero emission vehicles” is underway worldwide. However, emissions
45 associated with brake and tire wear will remain. We show here that a variety of previously
46 unrecognized volatile and semi-volatile organic compounds, some of which are greenhouse gases
47 or classified as Hazardous Air Pollutants, as well as nitrogen-containing organics, nitrogen
48 oxides and ammonia, are emitted during braking. The distribution and reactivity of these gaseous
49 emissions are such that they can react in air to form ozone and other secondary pollutants with
50 adverse health and climate consequences. Some of the compounds may prove to be unique
51 markers of brake emissions. At higher temperatures, nucleation and growth of nanoparticles is
52 also observed. Regions with high traffic, which are often disadvantaged communities, as well as
53 commuters can be impacted by these emissions even after combustion-powered vehicles are
54 phased out.

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3 **57 Introduction:**

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5 58 The inextricably intertwined issues¹ of air quality and climate change present major threats and
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7 59 challenges globally, especially to vulnerable communities. Since the start of the industrial
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9 60 revolution and the expanding use of fossil fuels, atmospheric concentrations of carbon dioxide
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11 61 and other greenhouse gases (GHG) such as nitrous oxide and methane have increased
12
13 62 dramatically.^{2,3} As global temperatures rise, extreme weather events such as wildfires are
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15 63 increasing.⁴ This has resulted in large episodic emissions of both gases and particles that impact
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17 64 air quality, health and climate.⁵⁻⁷

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19 66 Over the past five decades, tailpipe emissions of particles, volatile organic compounds (VOC)
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21 67 and nitrogen oxides (NO_x) from vehicles have declined dramatically in response to regulations.⁸⁻
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23 68 ¹² With the urgent need to address climate change, there is a transition underway from the use of
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25 69 fossil fuels in vehicles to what are termed zero emission vehicles (ZEV). For example, both the
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27 70 European Union and California have banned the sale of new gasoline-fueled vehicles starting in
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29 71 2035. However, these will not truly be ZEV since there will be continued emissions of both
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31 72 particles and gases from brakes and tires, as well as resuspension of road dust; it has been
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33 73 suggested that these vehicles actually be designated *ZEEV* for “*zero exhaust emission*
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35 74 *vehicles*”.¹³

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37 76 Indeed, with the dramatic reduction in tailpipe emissions over the years, particle emissions from
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39 77 brakes and tires¹³⁻²¹ are now thought to be about equal by mass to those from tailpipes¹⁶ in
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41 78 developed regions of the world. These have the potential to impact visibility and climate through
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43 79 scattering incoming solar radiation and altering cloud formation and properties, as well as having
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45 80 deleterious effects on humans and ecosystems.^{15,22-29} The impacts of particle emissions fall
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47 81 disproportionately on socioeconomically disadvantaged communities,^{30,31} often described as
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49 82 environmental justice (EJ) communities, many of which are in heavily trafficked areas either
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51 83 close to major roadways or large distribution centers that have heavy truck traffic. In addition,
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53 84 commuters will continue to be exposed to these emissions.^{32,33} In recent Los Angeles area
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55 85 studies, 21% of the oxidative potential of PM_{2.5} was attributed to brake and tire emissions, and
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57 86 both PM_{2.5} mass and the oxidative potential exposures increased in socioeconomically
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59 87 disadvantaged communities.^{24,34}

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89 While there is an accelerating transition to EVs, there is no agreement on the magnitude of the
90 changes in non-exhaust emissions that will result. For example, the use of regenerative braking
91 decreases brake emissions but the heavier weight due to the batteries leads to increased
92 emissions.^{13-16,34,35} It is noteworthy that the use of autonomous vehicles is expected to increase
93 these emissions since their algorithms require more frequent braking.¹⁴

94
95 While particles associated with non-tailpipe sources have been characterized in a number of
96 studies,^{13-16,36-39} relatively little is known about the gases. The most relevant study is that of
97 Placha et al.⁴⁰ who measured gas phase benzene, toluene, ethylbenzene, xylenes, polycyclic
98 aromatic hydrocarbons and total organic carbon along with particle composition during brake
99 wear. These results were also reported for a non-commercial brake pad formulated for their
100 experiments rather than commonly used commercial brake pads.

101
102 In this study, we focus on C1-C21 volatile and semi-volatile gases emitted during braking,
103 including probing the relationship between these gas emissions and the associated generation and
104 composition of particles from either ceramic or semi-metallic brake pads used widely in the
105 United States. Given the previous extensive work on particle emissions,^{13-16,36-39} we do not
106 include a comprehensive treatment of particles but only those aspects that are correlated with the
107 gas emissions. In addition, a search for potential specific gas markers for brake wear was carried
108 out.

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110 This appears to be the first report of this wide suite of VOC as well as NO_x, which individually
111 or through well-known secondary chemistry, are classified as Hazardous Air Pollutants, are
112 climate-active, or generate a host of pollutants⁴¹ that have deleterious impacts on human health
113 and welfare.²² These harmful secondary pollutants include ozone (also a GHG), nitric acid and
114 particles.

115 116 **Materials and Methods:**

117 A custom-built brake dynamometer (Fig. S1A)⁴² was used for these experiments. The facility
118 employs a heavy-duty metal working 22" lathe (Lodge & Shipley) to rotate a disc brake system

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3 119 (rotor), spanning the torques and temperatures found in normal driving conditions. The disc
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5 120 brake caliper is a common model (Kodiak model 225) for which a large variety of brake pads are
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7 121 commercially available. Different compositions of brake pad linings exist and are favored in
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9 122 different international markets. Two representative classes of linings were chosen for this study
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11 123 that were readily available: a ceramic (Kodiak model DBC-225) and a semi-metallic (BrakeBest
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13 124 model MKD289) brake pad. The exact formulation of the linings investigated is confidential and
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15 125 proprietary, but brake pads typically include five crucial common components: friction material,
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17 126 mixed in with binders, fillers, lubricants and reinforcement fibers.⁴³⁻⁴⁶ This results in a complex
18
19 127 mixture of organic and inorganic components. Semi-metallic brake pads typically have larger
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21 128 amounts of steel fibers and other metals compared to ceramic brakes^{14,47} while the ceramic
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23 129 brakes (i.e. aftermarket non-asbestos organic brakes)⁴⁸ are composed of mostly organic materials
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25 130 reinforced with aramid, glass or ceramic fibers.⁴³ These two brake types are common in the US
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27 131 market.⁴⁷
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31 133 Braking force was applied using an electric over hydraulic brake actuator (Hydrastar, model
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33 134 HBA-12) and a brake controller (Tekonsha, model PowerTrac), the latter of which was modified
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35 135 to accept computer control of braking force and time. For the braking system, DOT3 brake fluid
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37 136 (O'Reilly, 72120) was used. The brake caliper and rotor were enclosed in an 87 L aluminum
38
39 137 chamber to allow for clean purge air to be delivered, isolating emissions from surrounding
40
41 138 ambient air. The clean air was provided by a purge air generator (Parker-Balston, model 75-62)
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43 139 and was continuously introduced into the chamber via a side-port at 35 L min⁻¹. Sensors housed
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45 140 inside the chamber included a relative humidity (RH) and temperature sensor (Vaisala, model
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47 141 HPM-44), an infrared non-contact temperature sensor for measuring the temperature the rotor
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49 142 surface (Omega model OS301-HT), a pressure sensor for monitoring the pneumatic fluid
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51 143 pressure (AiM, model MC-327), and a torque sensor for monitoring the torque applied to the
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53 144 brake caliper (Ato, model ATO-TQS-S01). All instruments sampling VOC and particle
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55 145 emissions were connected along a single axis across the bottom section of the chamber front
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57 146 panel as seen in Fig. S1B. Sampling lines were ¼ in. aluminum or copper tubing, except for the
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59 147 whole air sampling (WAS) canister and semi-volatile organic compounds (SVOC) sorbent tubes,
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148 which were Teflon. The chamber was thoroughly cleaned to remove all particulate residues prior
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to each experiment.

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5 151 The lathe was operated at 173 RPM, corresponding to a driving speed of 18 miles hr⁻¹ for a
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7 152 typical 35" passenger vehicle wheel. Rotor rotation occurred without braking for several minutes
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9 153 before the first braking regime was applied (hereafter referred as to '*spinning only*' conditions).

10 154 The first regime (regime 1) was representative of light braking conditions and the second (regime
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12 155 2) was representative of heavy braking conditions. In regime 1, brake pressure was maintained at
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14 156 20-22 psi with a torque of ~120-190 N m, while in regime 2, brake pressure was maintained at
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16 157 ~30-32 psi with a torque of ~200-270 N m. In regime 2, the brake torque tends to decrease as the
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18 158 brake heats up, suggesting the brake pads undergo a loss of friction, known as brake fade.^{49,50} In
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20 159 regime 1, rotor temperatures typically increased from room temperature to a maximum
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22 160 temperature of 86-177°C depending on the experiment, while in regime 2 the temperatures were
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24 161 typically higher (ranging from 164-358°C). Light braking (regime 1) is characteristic of urban
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26 162 driving, while heavy braking (regime 2) is characteristic of rural and highway driving.⁵¹ The
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28 163 brake torque and rotor temperatures we achieve in the braking regimes adopted in the present
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30 164 study in general compare well with those reported as typical for some common make and model
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32 165 vehicles.⁵² The maximum temperatures reached in each experiment are shown in Table S1. A
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34 166 typical set of these parameters for a given experiment for each brake type is presented in Fig. S2.
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36 167 A separate set of experiments focused on collecting particles for toxicology studies was also
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38 168 performed with the same brakes. For these, it was necessary to run the dynamometer under
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40 169 regime 1 conditions for extended times (3-5 hr) which wore the surface of the brake pads
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42 170 heavily. We report here results from six experiments that were chosen to represent a range of
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44 171 conditions from new to heavily worn brake pads.

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43 173 Although some test cycles have been developed for brakes and tires,^{36,51,53} there is no universally
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45 174 accepted protocol. There are a number of factors that impact emissions, including the velocity
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47 175 when the brakes are first applied, the frequency and duration of braking, and the braking power
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49 176 and deceleration rate. Other factors such as the composition of the brake pads and the use history
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51 177 are important as well. As seen in Figs. 1A and 1C, gas emission is closely related to the rotor
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53 178 temperature, the range of which is covered by both braking regimes studied here. The intent of
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55 179 the present studies was not to mimic the overall emissions from proposed test cycles, but rather
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57 180 to determine whether there are significant gas emissions that have been overlooked, how they

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3 181 depend on braking conditions, and how they may impact air quality. Once universally accepted
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5 182 braking protocols have been established, the emissions as a function of rotor temperature can be
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7 183 used to estimate gas emissions under more realistic driving habits.

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10 185 Offline analysis of the C1-C10 gas phase VOC was performed using whole air sampling (WAS)
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12 186 with evacuated canisters followed by multicolumn, multidetector gas chromatography.⁵⁴
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14 187 Multiple samples were collected at discrete regular time intervals during each regime. In
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16 188 addition, proton-transfer reaction mass spectrometry (PTR-MS; Ionicon Analytik, model 8000)
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18 189 was used to capture the real-time evolution of VOC throughout the experiments. Further,
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20 190 Tenax/Carbograph 5 sorbent tubes were also collected at discrete regular time intervals and
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22 191 analyzed for volatile and semi-volatile compounds with volatility ranging from pentane (C5) to
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24 192 heneicosane (C21). Additional gas measurements included carbon monoxide (CO), nitric oxide
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26 193 (NO) and nitrogen dioxide (NO₂) that were performed using two commercial analyzers
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28 194 (ThermoFisher model 48i and 42C respectively). Particle size distributions were recorded using a
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30 195 customized scanning mobility particle sizer described in the Electronic Supplementary Information
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32 196 (ESI) and an aerodynamic particle sizer (TSI Inc., model 3936). For composition, particles were
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34 197 collected on carbon coated copper grids for scanning transmission electron microscopy (STEM)
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36 198 and electron dispersion spectroscopy (EDS). Real-time measurement of submicron particle
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38 199 composition was also achieved using a high-resolution aerosol mass spectrometer (AMS,
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40 200 Aerodyne, Inc.). More details on the measurements can be found in the ESI.

41 201

42 202 **Results and Discussion**

43 203 Figure 1A shows, for a ceramic brake pad, the rotor temperature, equivalent total VOC, CO, NO
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45 204 and NO₂ for regime 1, characterized in this experiment by brake rotor temperatures of up to
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47 205 115°C, followed by regime 2 braking conditions. Regime 2 is characterized by a new particle
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49 206 formation event (Fig. 1B) when rotor temperatures spanned from 115-177°C. Similar data for
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51 207 semi-metallic brake pads show that even higher temperatures are reached in regime 2 (Fig. 1C
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53 208 and 1D). Emissions of VOC, CO and NO_x increase dramatically in regime 2 for both brake
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55 209 types. The actual onset of the VOC emissions reported by the PTR-ToF-MS in regime 2 was
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57 210 similar to that of CO, NO and NO₂ as illustrated in Fig. S5. The emissions recorded for the
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59 211 semi-metallic brake pads were systematically delayed compared to those of the ceramic brake
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3 212 pads. As noted in the experimental section, ceramic brake pad linings are expected to have a
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5 213 higher content of organic material that may be more prone to thermal degradation than that of the
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7 214 semi-metallic brake linings. At the end of each experiment, VOC and other trace gases decline
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9 215 quickly, as soon as braking stops. This results from a combination of the rotor temperature
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11 216 decreasing and a dilution effect as the chamber is being continuously flushed with air. Note that
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13 217 the NO_x profiles were temperature dependent, and the semi-metallic brake produces less NO and
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15 218 NO₂ than the ceramic brake (Fig. S6). Emissions during regime 1 (lower rotor temperature)
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17 219 were much smaller than those encountered during regime 2, and dominated by small carbonyls
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19 220 (e.g. acetaldehyde, acetone). However, in order to fully characterize the suite of emissions, we
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21 221 focus hereafter on regime 2 results as the concentrations allowed for a variety of simultaneous
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23 222 accurate measurements of both gases and the associated particles.

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24 224 **Particles.** Size distributions in regimes 1 and 2 for ceramic brake pads are shown in Fig. 1B and
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26 225 for semi-metallic pads in Fig. 1D. Relatively few, large (micron-sized) particles are generated
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28 226 under the light braking (regime 1), with their metal content and morphology (Figs. S3 and S4,
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30 227 Table S3) similar to those reported previously.^{13-16,34,55,56} As the braking intensity and rotor
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32 228 temperature rise (regime 2), large numbers of small particles ranging in size from a few to
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34 229 several hundred nanometers in mobility diameter are observed. The appearance of these
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36 230 nanometer sized particles corresponds to an increase in organic gas emission in regime 2 (Figs.
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38 231 1A and 1C). This suggests nucleation is occurring that involves higher molecular mass/lower
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40 232 volatility organics, perhaps with inorganic seeds such as sulfuric acid⁶⁵ that are known to initiate
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42 233 new particle formation in ambient air. The resulting nucleated ultrafine particles appeared in
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44 234 general at critical temperatures ranging from 147°C – 277°C for each brake type which is in line
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46 235 with a previous reported range of 140-240°C.^{39,57,58} The only exception was the experiment
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48 236 performed with a fresh ceramic brake (exp #77) which had much higher VOC emissions, and
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50 237 hence the nucleation event happened at a lower temperature (T = 128°C). Nucleation from
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52 238 vapors emitted from the brake pads was also observed during a separate heating-only experiment
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54 239 where a piece of commercial brake pad was heated step-wise up to 250°C in a closed cell (data
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56 240 not shown). This supports the hypothesis that gaseous organics emitted during braking are
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58 241 responsible for forming new particles in regime 2. The detailed processes governing these
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60 242 processes are the subject of current studies.

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5 244 ***Volatile organic compounds distribution.*** A total of 85 individual VOC was identified and
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7 245 quantified in the WAS samples, with an additional ~ 20 compounds tentatively identified for
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9 246 which standards were not available. One striking observation is the relatively high amount of CO
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11 247 measured from brakes (orders of magnitude above background, Fig. 1) but surprisingly little
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13 248 carbon dioxide (CO₂, only double the background, Fig. S7). This ratio of CO/CO₂ is often
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15 249 associated with a low-combustion efficiency processes such as smoldering fires and was the case
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17 250 for both regimes 1 and 2. Figure 2 shows the overall emission ratio (ER; $\Delta\text{VOC}/\Delta\text{CO}$)
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19 251 distribution of the VOC measured in regime 2 classified by chemical groups. Quantitative ER
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21 252 values and standard deviations for each individual VOC for both types of brakes are given in
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23 253 Table S4 and S5 respectively. Common components include alkanes, alkenes, alkynes and
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25 254 aromatics compounds as well as alcohols, and carbonyls, with a notable contribution from
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27 255 nitriles. Note that a significant amount of methane was observed in both brake type experiments,
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29 256 with ER values of 81 and 76 pptv/ppbv CO for the ceramic and semi-metallic brakes respectively
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31 257 (Fig. S8; Table S1). Methane emission ratios observed in this study for brakes are the same order
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33 258 of magnitude as that from biomass burning (Fig. 3D). This potentially has important climate
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35 259 implications, as methane is an important greenhouse gas.

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37 260
38 261 The nature of the individual VOC is similar for both types of brake pads, but there is some
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40 262 variation in the relative contributions of different classes of compounds. For example, the semi-
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42 263 metallic brakes emit more alkanes and less aromatic compounds (Fig. 2). Figure 3A presents the
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44 264 average emissions of the 25 most abundant VOC measured in regime 2 as a ratio to CO for the
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46 265 six experiments performed with ceramic brake pads; similar ER data for the semi-metallic brakes
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48 266 are shown in Fig. 3B, with a direct comparison between the two brakes in Fig. 3C. It is
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50 267 noteworthy that a similar set of VOC have been identified from combustion of 18 different
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52 268 biomass fuels characteristic of three different regions of the U.S. (north, southeast, southwest)
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54 269 under controlled laboratory conditions⁵⁹ as well as in wildfire plumes (Figs. 3D and S9A). More
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56 270 than half of the top 25 compounds from both brakes are also in the top 25 from biomass burning.
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58 271 In addition, smaller contributors to the total VOC pool included furans and halogenated
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60 272 compounds (dominated by CH₃Cl), that are characteristic of biomass burning plumes,⁵⁹⁻⁶² and
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62 273 whose emissions are commonly not found in typical urban settings. Note that biomass burning

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3 274 emissions vary by fuel and conditions such as temperature and stage of the fire,^{60,61,63} and a 1-to-
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5 275 1 correlation with biomass burning is not expected; it is nonetheless intriguing that brake
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7 276 emissions share such similarities with biomass burning plumes (Fig. 3D).

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10 278 Significant amounts of NO_x are generated during heavy braking (Figs. 1 and S6) along with
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12 279 VOC, and their reactions in air generate ozone (O₃) and other oxidants, as well as particles.⁴¹
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14 280 Both O₃ and particles have climate and health impacts.²² Since a major loss process for VOC in
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16 281 the atmosphere that initiates this chemistry is reaction with the hydroxyl (OH) radical, the
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18 282 relative contributions to reactivity for individual VOC compared to that of CO can be estimated
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20 283 by $k_{OH} \times [\text{VOC}_i] / k'_{OH} \times [\text{CO}]$, where k_{OH} is the second-order rate constant for the reaction of the
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22 284 individual organic (VOC_i) with OH (Table S6) and k'_{OH} is the second-order rate constant for OH
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24 285 with CO ($2.4 \times 10^{-13} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$). Comparing the relative reactivities of brake emissions,
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26 286 illustrated in Figs. 4A and 4B (with a direct comparison in Fig. 4C), with those reported for
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28 287 biomass burning (Figs. 4D and S9B),⁵⁹ it is evident that a significant overlap exists between the
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30 288 most reactive compounds for brake and biomass burning emissions, with alkenes and carbonyls
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32 289 being major contributors in both cases. Note that the total reactivity relative to CO in Figs. 4A
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34 290 and 4B is somewhat higher for the brakes compared to that of biomass burning (Fig. 4D).

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36 291
37 292 Real-time analysis of the emitted gases was also carried out using PTR-MS which detects and
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39 293 measures some species that were not possible to measure by WAS. It also provides elemental
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41 294 composition for individual compounds that are sufficiently volatile to remain in the gas phase.
42
43 295 Figure S10 shows typical unit-resolution mass spectra for the two types of brake pads measured
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45 296 during regime 2, and the exact masses and elemental formulae from the high-resolution mass
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47 297 spectra are summarized in Table S7. A total of 93 individual ions were identified, with most
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49 298 having also been previously identified in biomass burning plumes also by PTR-MS.^{61,63-68} Of
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51 299 specific interest is the detection of formaldehyde, phenol, hydrogen cyanide (HCN) and
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53 300 isocyanic acid (HNCO) and a variety of nitrogen-containing organics including pyrrole, pyridine
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55 301 and a series of nitriles. Emission ratios for these VOC were estimated using the integrated VOC
56
57 302 concentrations measured by the PTR-MS at the same time as WAS canisters were collected.
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59 303 These VOC were ratioed to CO concentrations determined by WAS and averaged across the six
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61 304 experiments performed for each brake pad type. Note that VOCs associated with brake fluid

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3 305 vapors, which are glycol ethers (Table S7), were often observed at the highest temperatures, but
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5 306 typically not in large quantity for these experiments. Results are presented in Fig. 5 (stippled
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7 307 bars for ceramic brakes; solid colors for semi-metallic brakes), grouped by functional classes.
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9 308 Ammonia (NH₃), known to be present in car exhaust,⁶⁹ was also a significant contributor. A
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11 309 noticeable difference between the two types of brakes is the relatively larger concentrations and
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13 310 numbers of nitrogen-containing compounds for the ceramic brake compared to the semi-metallic
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15 311 brake pad emissions, likely linked to the differences in their proprietary formulations.
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17 312
18 313 Our measurements (Figs. S11A and S11B) also show small concentrations of ethylbenzene and
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20 314 xylenes compared to that of benzene for both brake types, in good agreement with Placha et al.⁴⁰
21
22 315 It is evident from Fig. S11 that the ER values of ethylbenzene and xylenes are much smaller in
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24 316 brakes compared to tailpipe emissions (Fig. S11C). Thus, these ratios may serve to differentiate
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26 317 brake from tailpipe emissions in air. Notably, the ratios for BTEX from brakes, as well as
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28 318 acetonitrile, match fairly well those of biomass burning samples from both laboratory burns and
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30 319 wildfires (Fig. S11D).

31 320
32 321 Nitrogen compounds detected from brakes (Fig. 5) have also been reported in biomass burning
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34 322 plumes.^{64,66,70,71} Some components such as nitrogen-containing organics and phenol will react in
35
36 323 air to form light-absorbing aerosol particles known as brown carbon,⁷² also potentially impacting
37
38 324 climate. Nitriles, especially acetonitrile, have been suggested as markers for biomass
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40 325 burning,^{59,64} and the higher concentrations from brakes may also make them useful for tracking
41
42 326 brake emissions.

43 327
44 328 Brake pads contain a complex mixture of organic and inorganic components, with phenolic
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46 329 resins as a common and significant constituent.^{48,73,74} The synthesis of these resins often involves
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48 330 diisocyanates and a tertiary amine, incorporating nitrogen and cyano groups into the brake
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50 331 pads.^{43,45,75} Additives such as silicone, epoxy- and rubber resins are frequently included.^{43,45,73}
51
52 332 The synthetic rubbers are often derived from acrylonitrile, and other nitrogen-containing resins
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54 333 based on cyanate esters and aramid pulp, and polybenzoxazines are common components.^{45,73,76}
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56 334 Upon decomposition, these materials are expected to release nitriles, such as acetonitrile,
57
58 335 acrylonitrile and benzonitrile as major products.⁷⁷⁻⁷⁹ Lastly, “green” components and natural
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3 336 fibers derived from plant-based material (e. g. cashew nut shell liquid or cashew dust, cellulose
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5 337 and lignin ...etc) have been introduced in the formulation of brake pads.^{43-46,80} The strong
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7 338 frictional and thermal forces in the tribological interactions that occur at high brake pad-rotor
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9 339 interface temperatures during braking^{81,82} generate a variety of gases that contain carbon, oxygen
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11 340 and nitrogen. The significant contribution of organonitrogen compounds to the gases and
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13 341 particles observed here is consistent with the presence of nitrogenous components in brake pads.
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15 342 It is therefore not surprising that the emissions from degradation of the organic components in
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17 343 the brake pads have similarities to those of biomass burning.
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19 345 Furthermore, studies of the thermal decomposition of some phenolic resins report generation of a
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21 346 suite of gases that were also observed here during braking. These include phenol, methane,
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23 347 ethane, ethene, propene, 1-butene, 1-pentene, acetonitrile, propanenitrile, benzonitrile, HCN,
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25 348 NH₃, hydrogen (H₂) and aromatics such as benzene, toluene and the xylenes.^{75,83,84}
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27 349 Decomposition of a phenol resin was observed to start around 200°C,⁷⁵ similar to regime 2 in the
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29 350 present studies.

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31 352 ***Semi-volatile organic compounds.*** To understand the nature of the observed new particle
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33 353 formation, volatile and semi-volatile organic components (SVOC) with boiling points ranging
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35 354 from 174°C to 356°C (corresponding to C10 (decane) to C21 (heneicosane)) were measured by
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37 355 sampling onto Tenax sorbent cartridges and performing thermal desorption GC-MS analysis. In
38
39 356 the SVOC analysis there are more than 300 distinct peaks in the chromatograms, illustrating the
40
41 357 complex mixture of gases that are emitted. Figures 6A and 6C summarize the average SVOC
42
43 358 organics (C10 and larger) measured for both brake types under the heavy braking conditions of
44
45 359 regime 2 grouped by class (variability across experiments is presented in Figs. S13A and S13C).
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47 360 Oxygen and nitrogen-containing organics contribute about 48% of the total SVOC mass for the
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49 361 ceramic brake and 41% for the semi-metallic brake, with hydrocarbons responsible for most of
50
51 362 the remainder.

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53 364 If the SVOC contribute significantly to new particle formation, one might expect some
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55 365 relationship between their distribution and those found in the newly formed particles. Thus,
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57 366 particle composition was also measured simultaneously using a high-resolution time-of-flight
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3 367 aerosol mass spectrometer (AMS) and is shown grouped by compound class in Fig. 6B for the
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5 368 ceramic brake and Fig. 6D for the semi-metallic brake. These represent averages across three
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7 369 experiments for each brake type with the variation across experiments given in Figs. S13B and
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9 370 S13D. The two types of measurements (SVOC and AMS) show somewhat similar composition
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11 371 for both type of brakes. The largest difference is the nitrogenated species whose contribution is
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13 372 larger in the particles measured by the AMS compared to the SVOC distribution. This could
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15 373 indicate that smaller more volatile species may be trapped in the particles as they grow quickly to
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17 374 > 50-100 nm or may represent species that are not detected by the sorbent tube. These results
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19 375 also align with previous reports of complex mixtures of higher molecular mass organic gases and
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21 376 particles from brakes.^{40,85}

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23 378 ***Environmental Impacts.*** There are potential direct impacts of the gas emissions on health and
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25 379 climate. For example, acetonitrile, acetaldehyde, benzene, toluene, the xylenes, methanol,
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27 380 acrolein and phenol are designated as hazardous air pollutants (HAP) by the U.S. Environmental
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29 381 Protection Agency.⁸⁶ Given that the brake emissions are at street level, communities in high
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31 382 traffic areas and commuters will continue to be exposed to toxic gases even when the tailpipe
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33 383 emissions have dropped to zero. The major emissions of VOC and SVOC from brake usage are
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35 384 also common to those from biomass burning which have known impacts on health and
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37 385 climate.^{6,7,59,87,88} Given the similarity to biomass burning and the potential increasing usage of
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39 386 eco-friendly natural fibers, fillers and binders in brake manufacturing, emissions from “ZEEV”
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41 387 vehicles might be described as due to a *slow-burning, continuous wildfire*, that would persist
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43 388 even when tailpipe emissions are eliminated. Although brake emissions from individual brake
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45 389 pads may not be as large on an absolute scale as those from individual wildfires, brake emissions
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47 390 are not episodic, but are ongoing and cumulative in urban areas.

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50 393 Some of the gases associated with brake emissions, such as directly emitted methane and ozone
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52 394 formed from secondary VOC-NO_x reactions, are greenhouse gases. Others such as H₂ measured
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54 395 in the present studies (Fig. S12) and reported by others^{61,89} have an indirect effect by reacting
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56 396 with OH radicals in air, reducing the OH concentration and hence increasing the lifetimes of
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58 397 other species with which it reacts such as methane.⁹⁰ The H₂ emissions measured in these studies

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3 398 are sufficiently small that they are unlikely to be important compared to other sources,^{61,89} for
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5 399 example, leakage from pipelines and storage facilities. However, as the use of H₂ as a
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7 400 replacement fuel increases, emissions associated with brakes should be included in the hydrogen
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9 401 budget.

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12 403 Quantifying the contribution of brake emissions to air quality is challenging due to the lack of
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14 404 unique markers. In a number of studies,^{13,15,16,24,26,34,91,92} source apportionment techniques have
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16 405 been applied using different combinations of trace metals in particles to estimate the contribution
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18 406 of brakes to airborne particulate matter. While this approach is useful, it is complicated by the
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20 407 number of different sources of airborne metals, and in addition, some markers used for brake
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22 408 pads such as copper are being phased out. In terms of potential gas phase markers, acetonitrile
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24 409 has been suggested as a marker for biomass burning,^{59,64} and the higher concentrations from
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26 410 brakes compared to exhaust also support its use as a marker of brake emissions in the absence of
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28 411 wildfires. Thus, the ratio of acetonitrile to CO in biomass burning plumes has been reported as
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30 412 2.01 ± 0.16 pptv/ppbv, an order of magnitude larger than in urban air, 0.26 ± 0.16 pptv/ppbv.⁹³ In
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32 413 the present studies, this ratio was 1.8 ± 1.3 pptv/ppbv CO for ceramic brakes and 5.6 ± 1.7
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34 414 pptv/ppbv CO for semi-metallic brakes in regime 1. In regime 2, the ratio was 15.7 ± 3.9
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36 415 pptv/ppbv CO and 7.7 ± 1.0 pptv/ppbv CO for ceramic and semi-metallic brakes respectively
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38 416 (Table S4 and S5). Other nitriles which are emitted at lower levels such as propanenitrile,
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40 417 acrylonitrile and benzonitrile may also make them useful for estimating brake emissions. Nitriles
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42 418 are relatively slow to react in air (Table S6) and lifetimes with respect to OH radicals are
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44 419 estimated to be on the order of 3 to 526 days at typical OH radical concentration of 10^6 cm⁻³.

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47 421 Another potential marker of brake emissions is phenol.⁴⁰ The phenol-to-CO ratio from brakes
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49 422 measured using PTR-MS for phenol is as high as 17 pptv/ppbv CO for the semi-metallic brake
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51 423 (regime 2; the value for the ceramic brakes was 3.6 pptv/ppbv CO), significantly larger than that
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53 424 reported from biomass burning, 0.5-2.1 pptv/ppbv CO.⁵⁹ With much lower temperatures
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55 425 recorded in regime 1, the phenol-to-CO ratio was 0.12 pptv/ppbv CO for the semi-metallic brake
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57 426 (there was no correlation with CO for the ceramic brake). The rate constant for OH reaction with
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59 427 phenol is 2.7×10^{-11} cm³ molecule⁻¹ s⁻¹ at 296 K.^{94,95} Using an average OH radical
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3 428 concentration of 10^6 cm^{-3} , the lifetime of phenol with respect to OH corresponds to
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5 429 approximately 10 hours.

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8 431 With continuous brake emissions that may be found near heavily trafficked roads and the low
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10 432 reactivity with OH, phenol and nitriles may be detectable. For example, previous studies^{96,97}
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12 433 reported acetonitrile near roadsides and attributed its source to traffic, however acetonitrile is low
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14 434 in tailpipe measurements of common light duty vehicles.^{98,99} Our findings suggest that emissions
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16 435 from brakes may have contributed to their observations. The key attribute of such markers is
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18 436 that they are not found in significant concentrations from modern tailpipe emissions.¹⁰⁰ Sampling
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20 437 of the exhaust of several passenger vehicle showed that phenol was present only at very low
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22 438 concentrations (Fig. S11C). Thus acetonitrile and phenol may be good markers for brake
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24 440 source apportionment capabilities in heavily trafficked urban settings going forward. Future
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26 441 studies are needed to confirm unique markers for brake emissions and to measure them in urban
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28 442 environments so that their impacts on air quality and climate can be quantified.

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31 444 While technology such as electrostatic precipitators have been in development to remove
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33 445 particulate matter emitted from brakes, no solution has yet been proposed to reduce emissions of
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35 446 gases. Elucidation and speciation of the VOC emitted will allow mitigation strategies to be
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37 447 developed and implemented. This could include, for example, solid sorbent traps that capture the
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39 448 VOC before their release into the atmosphere, and/or designing heat-resistant brake pad
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41 449 materials⁸¹ that will produce fewer toxic VOC.

42 450

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54 457 Science Foundation through the UC Irvine Materials Research Science and Engineering Center
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56 458 (DMR-2011967).

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3 459 **Figure Captions:**

4 460
5 461 **Figure 1. Time profile of gas phase species and particles recorded during a typical brake**
6 462 **dynamometer experiment using the ceramic brake pad (A-B) and the semi-metallic brake**
7 463 **pad (C-D).** Panels A and C show the time profiles of gas phase species including total equivalent
8 464 VOC mixing ratio measured by PTR-MS (green trace; 1 min average; the dark green symbols
9 465 highlight when the WAS canisters were collected), NO (red trace) and NO₂ (blue trace) mixing
10 466 ratios measured using a chemiluminescence analyzer, CO (yellow trace) mixing ratio measured
11 467 using a CO monitor as well as measured in the WAS canisters (dark yellow triangles). The rotor
12 468 temperature measured during the experiment is indicated by the grey trace. Panels (B and D)
13 469 show the particle size distributions measured using a scanning mobility particle sizer (SMPS)
14 470 and aerodynamic particle sizer (APS). Rotation of the rotor occurred without braking for several
15 471 minutes before the first braking regime was applied (*‘spinning only’* condition). Regime 1
16 472 corresponds to light braking conditions while regime 2 corresponds to heavy braking conditions.
17 473 The rapid decrease of VOC and other trace gases at the end of the experiment is observed as
18 474 soon as braking stops, which is associated with a cooling of the rotor temperature and dilution as
19 475 the chamber is continuously flushed with air.
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21 477 **Figure 2. Volatile organic compound distribution.** Distribution of emission ratios (relative to
22 478 CO) from VOC measured simultaneously using whole air sampling (WAS) collected in regime 2
23 479 for (A) ceramic brakes and (B) semi-metallic brakes, averaged over six brake dynamometer
24 480 experiments per brake type. Note that the alkane category does not include CH₄. Additionally,
25 481 the C1-C2 halogens category, dominated by CH₃Cl, was omitted because its contribution was
26 482 less than 0.2% for both brake types. Labels include MTs for monoterpenes and S-cpds for sulfur-
27 483 containing compounds. Table S1 and Fig. S8 provide information on the experimental
28 484 variability across experiments
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30 486 **Figure 3. Average emission ratios of individual VOC (top 25) relative to CO.** Average
31 487 emission ratios from (A) ceramic brakes and (B) semi-metallic brakes measured using whole air
32 488 sampling (WAS) collected in regime 2, averaged over six brake dynamometer experiments per
33 489 brake type and ranked from the most abundant to the least. Error bars for emission ratios
34 490 represent one standard deviation and are used to determine the corresponding error in relative
35 491 reactivity. Panel C is a direct comparison between the two brake types. Biomass burning data in
36 492 panel D were taken from Gilman et al. (2015) and represent laboratory biomass burning studies
37 493 for combustion of some fuels characteristic of the north (N), southeast (SE) and southwest (SW)
38 494 regions of the United States.⁵⁹ In all panels, TMB stands for trimethylbenzene.
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40 496 **Figure 4. Relative reactivity of individual VOC compounds (top 25) compared to that of**
41 497 **CO.** Relative reactivity of individual VOC from (A) ceramic brakes and (B) semi-metallic
42 498 brakes based on emission ratios determined in Fig. 3. Errors bars represent 1 standard deviation
43 499 calculated from the error in the ER values from Fig. 3, without taking into account the error in
44 500 the rate constants. Panel C is a direct comparison between the two brake types. TMB stands for
45 501 trimethylbenzene. For each brake, the total reactivity was calculated similarly to ref. 59 using
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3 502 $total\ OH\ reactivity = \sum(ER \times k_{OH} \times 2.46 \times 10^{10}\ molecules\ cm^{-3}\ ppbv^{-1})$ taking into
4 503 account the top 25 VOCs listed.
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7 505 **Figure 5. Average emission ratios (ER) of VOC measured with the PTR-MS grouped by**
8 506 **functional and structural group class for the ceramic brake (stippled bars) and the semi-**
9 507 **metallic brake (solid colored bars).** These ER were determined during regime 2 conditions and
10 508 averaged over six dynamometer experiments per brake type. The C_xH_y category corresponds
11 509 exclusively to unsubstituted compounds, and substituted VOC with multiple functional groups
12 510 were counted once in each appropriate functional group category. The contributions from several
13 511 important VOC with high ER are indicated in separate colors, while the rest of the compounds of
14 512 a given family are represented in red. NH_3 almost systematically saturated the instrument during
15 513 regime 2, even with the dilution on, and was excluded from this plot. The contribution from
16 514 specific brake fluid VOCs (i. e. glycol ethers, see Table S7) was small during these experiments,
17 515 and to keep the focus strictly on brake emission, these were not included here. Note that some
18 516 contribution of brake fluid to $m/z\ 45\ (C_2H_7O^+)$ cannot be excluded.
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23 518 **Figure 6. Average semi-volatile organic compound ($\geq C_{10}$) distribution and particle**
24 519 **chemical composition measured simultaneously for ceramic or semi-metallic brake pads**
25 520 **under heavy braking conditions (regime 2).** Mass distributions of (A) SVOC with 10 carbons
26 521 and more and, (B) particle components for ceramic brakes. Mass distributions of (C) SVOC with
27 522 10 carbons and more and (D) particle components for semi-metallic brakes. The SVOC analyses
28 523 represent an average of 6 experiments per brake type, while the particle analyses represent an
29 524 average of 3 experiments. No sulfur-containing compounds identified in the SVOCs category,
30 525 while in panels (B) and (D), sulfur compounds include $C_xH_yS_z^+$ as well as $HS_xO_y^+$ fragments.
31 526 No halocarbons were identified in the SVOCs category. Figure S13 shows data for each
32 527 experiment and the corresponding average.
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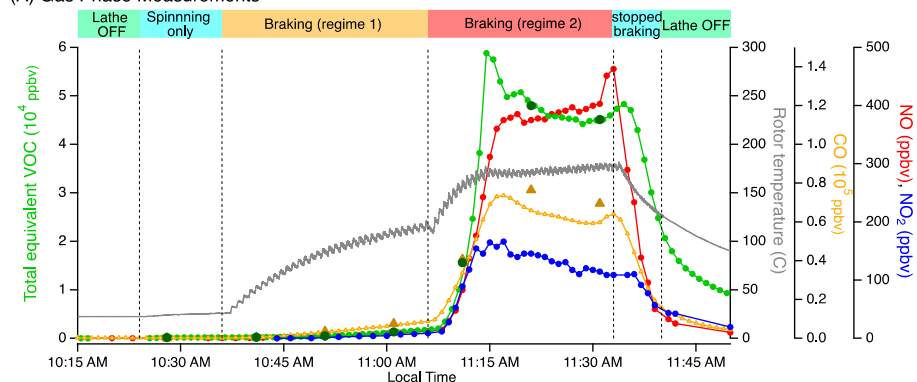
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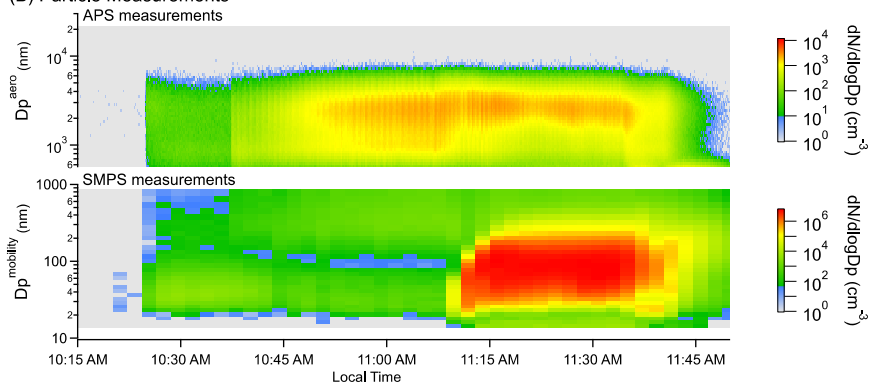
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Ceramic brakes

(A) Gas Phase Measurements

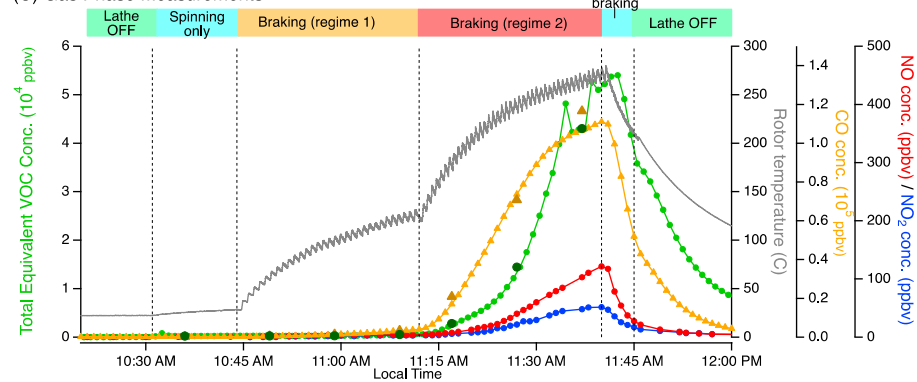


(B) Particle Measurements



Semi-metallic brakes

(C) Gas Phase Measurements



(D) Particle Measurements

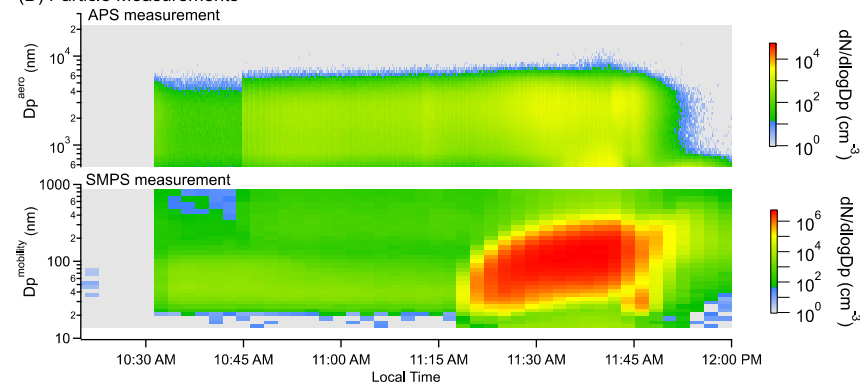


Figure 1. Time profile of gas phase species and particles recorded during a typical brake dynamometer experiment using the ceramic brake pad (A-B) and the semi-metallic brake pad (C-D). Panels A and C show the time profiles of gas phase species including total equivalent VOC mixing ratio measured by PTR-MS (green trace; 1 min average; the dark green symbols highlight when the WAS canisters were collected), NO (red trace) and NO₂ (blue trace) mixing ratios measured using a chemiluminescence analyzer, CO (yellow trace) mixing ratio measured using a CO monitor as well as measured in the WAS canisters (dark yellow triangles). The rotor temperature measured during the experiment is indicated by the grey trace. Panels (B and D) show the particle size distributions measured using a scanning mobility particle sizer (SMPS) and aerodynamic particle sizer (APS). Rotation of the

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3 rotor occurred without braking for several minutes before the first braking regime was applied (*‘spinning only’* condition). Regime 1
4 corresponds to light braking conditions while regime 2 corresponds to heavy braking conditions. The rapid decrease of VOC and
5 other trace gases at the end of the experiment is observed as soon as braking stops, which is associated with a cooling of the rotor
6 temperature and dilution as the chamber is continuously flushed with air.
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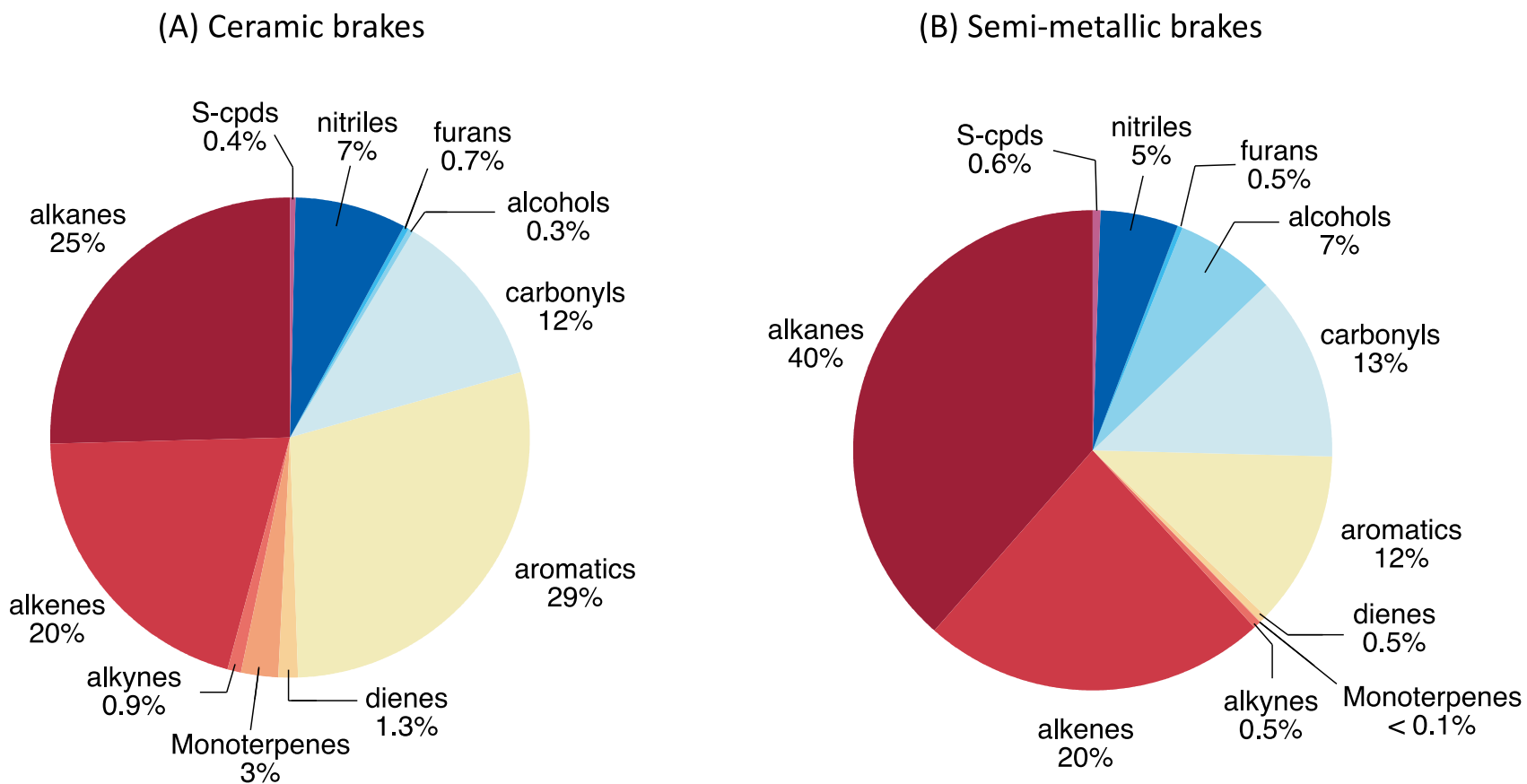
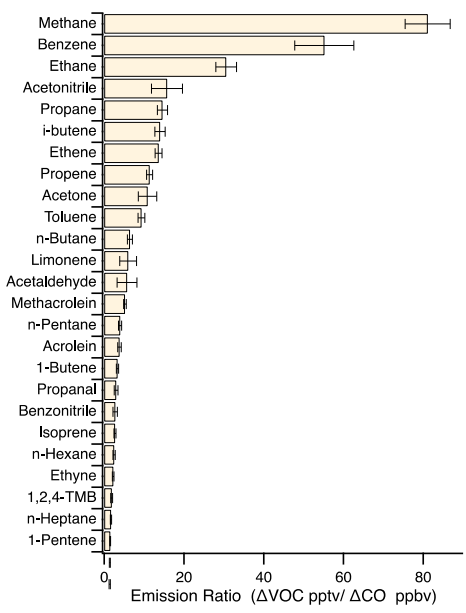


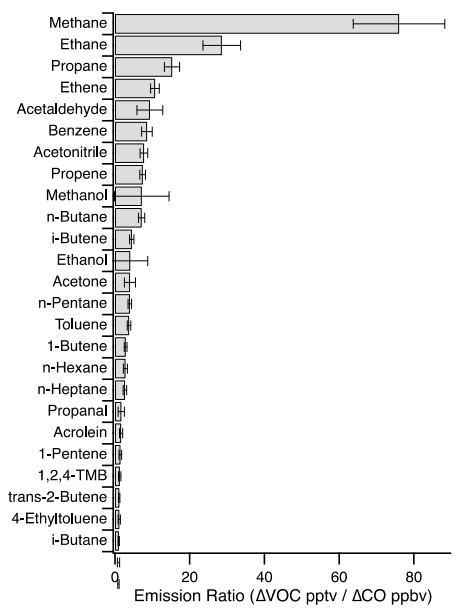
Figure 2. Volatile organic compound distribution. Distribution of emission ratios (relative to CO) from VOC measured simultaneously using whole air sampling (WAS) collected in regime 2 for (A) ceramic brakes and (B) semi-metallic brakes, averaged over six brake dynamometer experiments per brake type. Note that the alkane category does not include CH₄. Additionally, the C1-C2 halogens category, dominated by CH₃Cl, was omitted because its contribution was less than 0.2% for both brake types. Labels include MTs for monoterpenes and S-cpds for sulfur-containing compounds. Table S1 and Fig. S8 provide information on the experimental variability across experiments.

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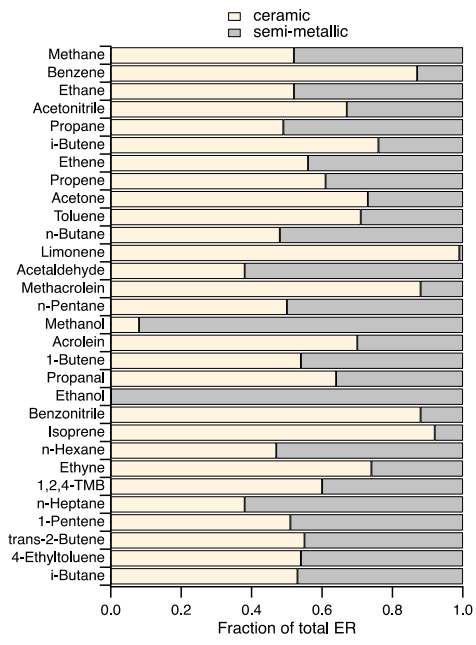
(A) Ceramic brakes



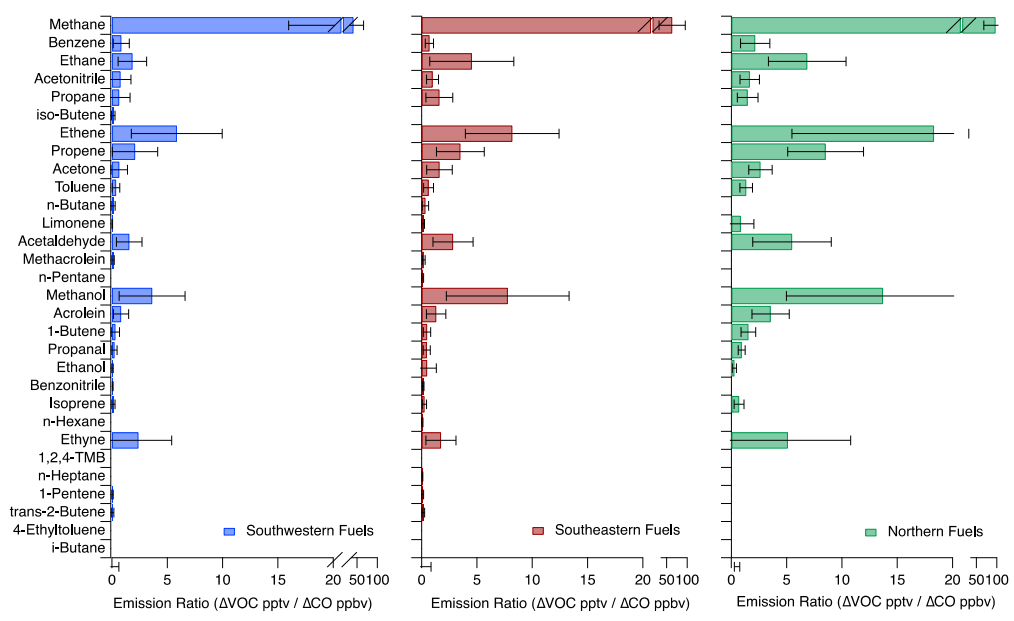
(B) Semi-metallic brakes



(C) Comparison between brake type

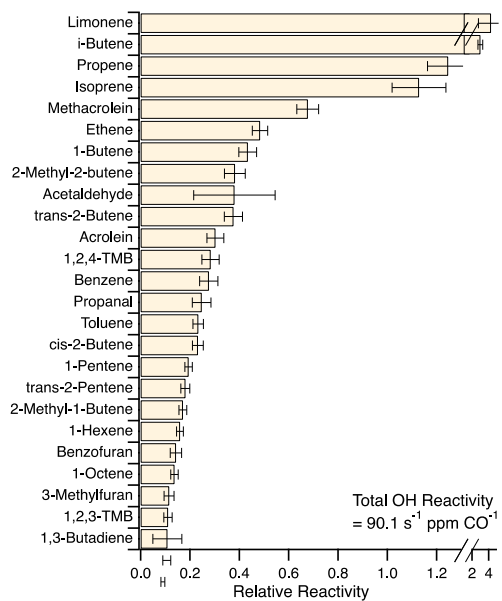


(D) Comparison with biomass burning

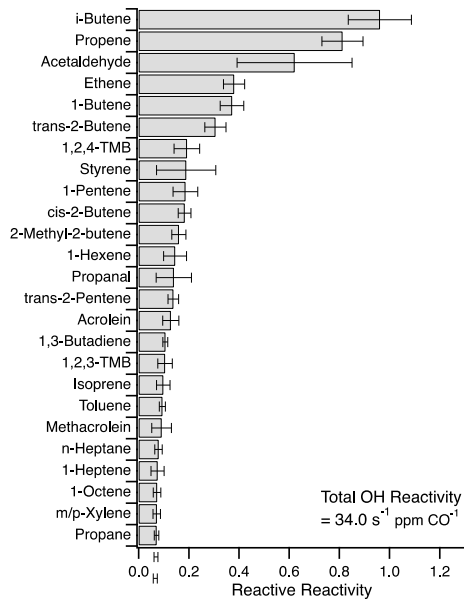


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3 **Figure 3. Average emission ratios of individual VOC (top 25) relative to CO.** Average emission ratios from (A) ceramic brakes and
4 (B) semi-metallic brakes measured using whole air sampling (WAS) collected in regime 2, averaged over six brake dynamometer
5 experiments per brake type and ranked from the most abundant to the least. Error bars for emission ratios represent one standard
6 deviation and are used to determine the corresponding error in relative reactivity. Panel C is a direct comparison between the two
7 brake types. Biomass burning data in panel D were taken from Gilman et al. (2015) and represent laboratory biomass burning studies
8 for combustion of some fuels characteristic of the north (N), southeast (SE) and southwest (SW) regions of the United States.⁵⁹ In all
9 panels, TMB stands for trimethylbenzene.
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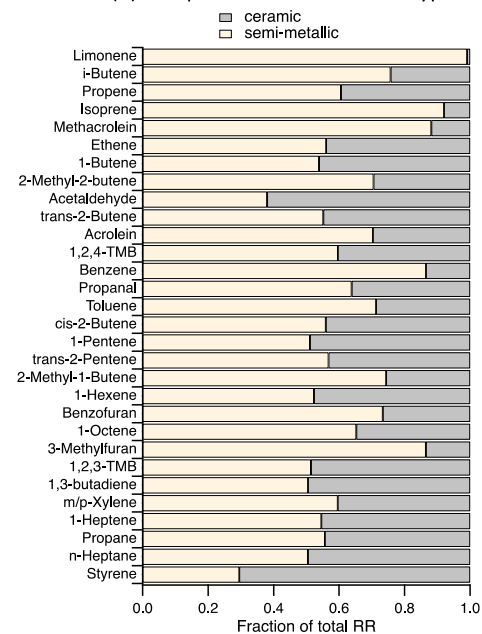
(A) Ceramic brakes



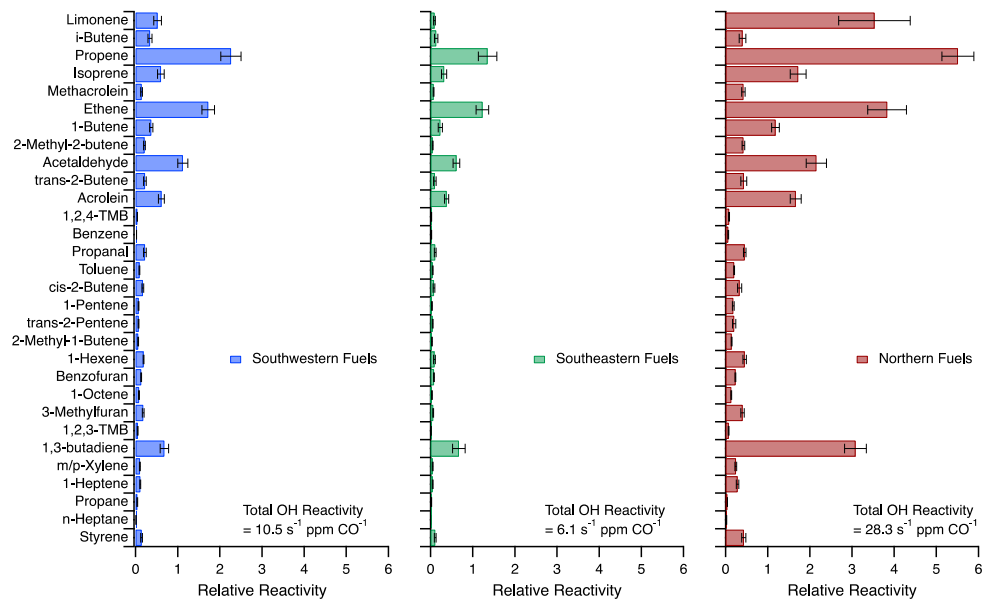
(B) Semi-metallic brakes



(C) Comparison between brake type



(D) Comparison with biomass burning



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3 **Figure 4. Relative reactivity of individual VOC compounds (top 25) compared to that of CO.** Relative reactivity of individual
4 VOC from (A) ceramic brakes and (B) semi-metallic brakes based on emission ratios determined in Fig. 3. Errors bars represent 1
5 standard deviation calculated from the error in the ER values from Fig. 3, without taking into account the error in the rate constants.
6 Panel C is a direct comparison between the two brake types. TMB stands for trimethylbenzene. For each brake, the total reactivity
7 was calculated similarly to ref. 59 using *total OH reactivity* = $\sum(ER \times k_{OH} \times 2.46 \times 10^{10} \text{ molecules cm}^{-3} \text{ ppbv}^{-1})$ taking into
8 account the top 25 VOCs listed.
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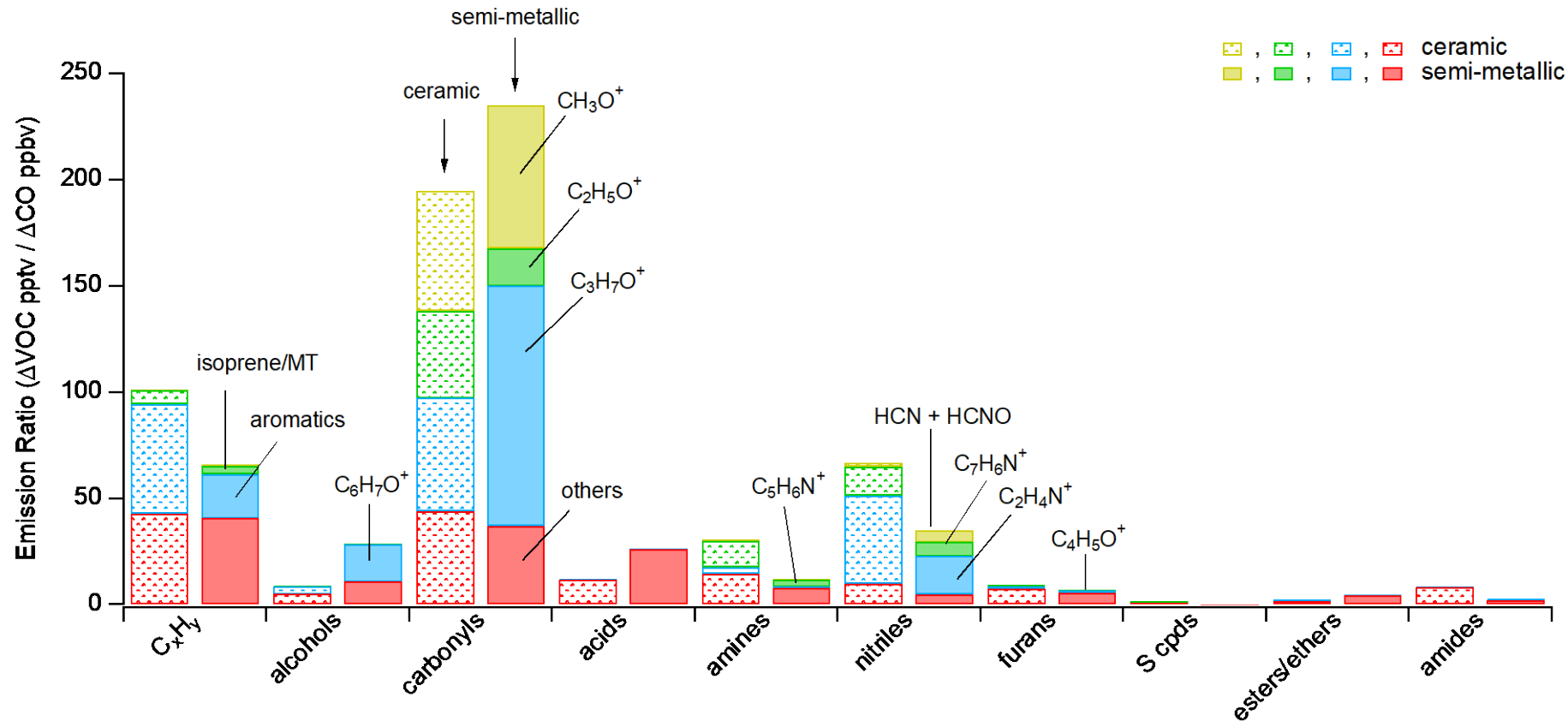


Figure 5. Average emission ratios (ER) of VOC measured with the PTR-MS grouped by functional and structural group class for the ceramic brake (stippled bars) and the semi-metallic brake (solid colored bars). These ER were determined during regime 2 conditions and averaged over six dynamometer experiments per brake type. The C_xH_y category corresponds exclusively to unsubstituted compounds, and substituted VOC with multiple functional groups were counted once in each appropriate functional group category. The contributions from several important VOC with high ER are indicated in separate colors, while the rest of the compounds of a given family are represented in red. NH_3 almost systematically saturated the instrument during regime 2, even with the dilution on, and was excluded from this plot. The contribution from specific brake fluid VOCs (i. e. glycol ethers, see Table S7) was small during these experiments, and to keep the focus strictly on brake emission, these were not included here. Note that some contribution of brake fluid to m/z 45 ($\text{C}_2\text{H}_7\text{O}^+$) cannot be excluded.

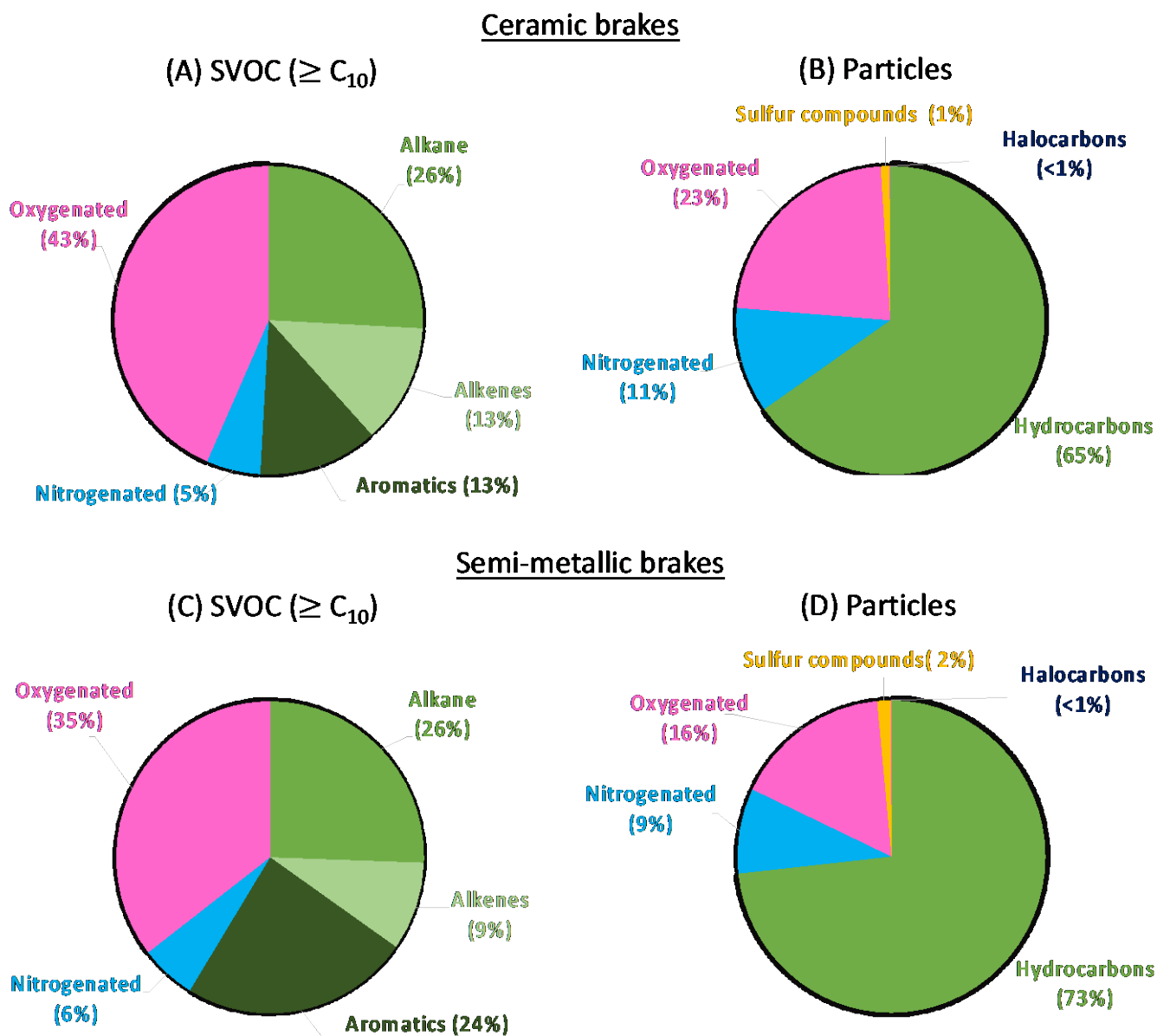


Figure 6. Average semi-volatile organic compound ($\geq C_{10}$) distribution and particle chemical composition measured simultaneously for ceramic or semi-metallic brake pads under heavy braking conditions (regime 2). Mass distributions of (A) SVOC with 10 carbons and more and, (B) particle components for ceramic brakes. Mass distributions of (C) SVOC with 10 carbons and more and (D) particle components for semi-metallic brakes. The SVOC analyses represent an average of 6 experiments per brake type, while the particle analyses represent an average of 3 experiments. No sulfur-containing compounds identified in the SVOCs category, while in panels (B) and (D), sulfur compounds include $C_xH_yS_z^+$ as well as $HS_xO_y^+$ fragments. No halocarbons were identified in the SVOCs category. Figure S13 shows data for each experiment and the corresponding average.