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Capture of toxic gases in MOFs: SO_2 , H_2S , NH_3 and NO_x

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MOFs are promising candidates for the capture of toxic gases since their adsorption properties can be tuned as a function of the topology and chemical composition of the pores. Although the main drawback of MOFs is their vulnerability to these highly corrosive gases which can compromise their chemical stability, remarkable examples have demonstrated high chemical stability to SO_2 , H_2S , NH_3 and NO_x . Understanding the role of different chemical functionalities, within the pores of MOFs, is the key for accomplishing superior captures of these toxic gases. Thus, the interactions of such functional groups (coordinatively unsaturated metal sites, μ -OH groups, defective sites and halogen groups) with these toxic molecules, not only determines the capture properties of MOFs, but also can provide a guideline for the design of new multi-functionalised MOF materials. Thus, this perspective aims to provide valuable information on the significant progress on this environmental-remediation field, which could inspire more investigators to provide more and novel research on such challenging task.

1. Introduction

The accelerated growth of our modern society demands huge amounts of energy. Unwittingly, in order to provide these high energetic levels, an indiscriminate combustion of large volumes of fossil fuels occurs, leading to an incommensurable release of toxic pollutants to the atmosphere. Emissions of anthropogenic air pollutants generate a vast range of health complications (*e.g.*, premature death and morbidity). Additionally, these air pollutants are also responsible for the reduction of the biodiversity, crop damages and acidification of soils and waters.¹ As an example of the adverse impacts to humans, the World Health Organization (WHO) recently announced that air pollution was responsible for the premature mortality of approximately 4.2 million people in 2016 alone.² In fact, air pollution is now the single largest environmental health risk worldwide since it is responsible for one in eight premature global deaths.³ For example, $\text{PM}_{2.5}$ (fine inhalable particles, with diameters that are

generally 2.5 micrometres and smaller) are responsible for approximately half of the deaths related to air pollution.⁴ Thus, $\text{PM}_{2.5}$ have been classified as the air pollutant with the highest impact in premature mortality,⁵ and a reduction of their emissions is crucial. Such reduction can be achieved in two steps: (i) mitigation of primary $\text{PM}_{2.5}$ emissions and (ii) extenuation of secondary inorganic aerosols (SIA); which are oxidised from precursor emissions such as sulphur dioxide (SO_2), nitrogen oxides (NO_x), ammonia (NH_3) and volatile organic compounds (VOCs).⁶ Thus, in order to extenuate these emissions, different actions have been taken, such as the shutdown of some coal-fired power plants and their replacement by thermoelectric power plants, importing electricity from other countries with strict restrictions on the electricity production and the use of more environmentally friendly energy sources.⁷ In addition to these actions, the development of efficient technologies for the capture of toxic gases (*e.g.*, NO_x , SO_2 , NH_3 and H_2S) from static and mobile sources is necessary, in order to achieve a cleaner environment.⁸

Porous metal-organic frameworks (MOFs) or porous coordination polymers (PCPs) are amongst the most promising candidates for the capture of these toxic gases since their sorption selectivity is directly tuneable as a function of the topology and chemical composition of the pores.⁹ Undoubtedly, for NO_x , SO_2 , NH_3 and H_2S capture, there is a substantial emphasis on optimising the interactions between MOF materials and these toxic molecules, leading to the discovery of new functional porous materials with enhanced gas adsorption properties.¹⁰ Although MOF materials have shown very promising capabilities for the capture of these toxic gases, their main

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drawback is their vulnerability to these highly corrosive molecules capable of compromising the chemical stability of the MOFs. NO_x , SO_2 , NH_3 and H_2S can disrupt the coordination bonds between the organic ligands (e.g., carboxylate) and the metal centres, occasioning the breakdown of the MOF structure. Therefore, chemical stability of the MOFs is a fundamental requirement for the capture of these toxic gases. The present contribution aims to provide a useful reference in the field of capturing toxic gases in MOFs, emphasising on their chemical stability and the role of the functional groups to enhance such captures. We hope to encourage more research groups to explore the exciting frontiers of science in environmental-remediation applications for MOFs.

2. Sulphur dioxide

The anthropogenic release of SO_2 into the atmosphere is mainly due to the combustion of fossil fuels (e.g., electric power generation). One of the first environmental problems related to SO_2 was observed during the last century in the acidic deposition.¹¹ To avoid this, since 1990 many European countries have considerably reduced their SO_2 emissions.¹² Nevertheless, the rapid industrialisation of some developing countries has caused a constant presence of SO_2 in the atmosphere,¹³ risking the health of millions of people (*vide supra*).

SO_2 is a colourless, non-flammable and corrosive gas with high solubility (120 g l^{-1}) in water. Among the technologies for SO_2 capture, wet and dry flue-gas desulfurization (FGD) processes are commonly used,¹⁴ where the cost and recyclability depend on the used technology. Among them, alkaline solutions, activated carbons, zeolites and silica have typically showed low SO_2 adsorption capacities, accompanied by the corrosion of pipelines and high energy regeneration overheads.¹⁵ Thus, the design of new materials is extremely necessary in order to meet all the challenges that the capture of SO_2 signifies. MOF materials are an exciting alternative for the capture of SO_2 as they have demonstrated promising results for the SO_2 capture assignment, mainly due to their specific chemical functionality and pore dimensions.¹⁶ The following section focuses on the highlights of remarkable MOF materials stable towards SO_2 , with the emphasis not only on their capture performance, but also on showing the role of different chemical functionalities which are the key for the SO_2 capture.

2.1 MOFs for SO_2 capture

The chemical nature of the SO_2 molecule and its main binding modes: oxygen, an electron-rich atom that can act as a Lewis base; and sulphur that can be a Lewis acid site, have been exploited to improve its adsorption on the porous surface of the MOFs. Due to the specific functionalisation of these MOFs, interesting SO_2 capture results have been achieved. Such functionalisation can be classified in four main categories: (i) coordinatively unsaturated metal sites (open metal sites), (ii) $\mu\text{-OH}$ groups, (iii) defective sites and (iv) halogen functionalisation.

(i) **Coordinatively unsaturated metal sites.** Although some research has proven the capture of SO_2 by different MOFs, the majority of these examples have shown to be chemically unstable upon SO_2 exposure.¹⁷ For example, MOF-177 (an open Zn^{2+} sites system reported by Janiak and co-workers¹⁸) holds the SO_2 capture record (25.7 mmol g^{-1} at 298 K and 1 bar) but suffers a partial structural degradation, after the exposure to SO_2 , as demonstrated by PXRD and a significant BET surface area reduction.¹⁸ Conversely, MFM-170 an open Cu^{2+} sites MOF; $[\text{Cu}_2(\text{L})]$ ($\text{H}_4\text{L} = 4',4'''-(\text{pyridine}-3,5\text{-diyl})\text{bis}([\text{1},1'\text{-biphenyl}]-3,5\text{-dicarboxylic acid})$), not only demonstrated to be highly stable to SO_2 (17.5 mmol g^{-1} at 298 K and 1 bar), but also to humid SO_2 .¹⁹ Yang and Schröder elegantly demonstrated the reversible coordination of SO_2 to open Cu^{2+} sites in MFM-170 by *in situ* synchrotron single-crystal X-ray diffraction, *in situ* FTIR spectroscopy and *in situ* inelastic neutron scattering (INS).¹⁹ In brief, all of this comprehensive experimental evidence corroborated that the Cu^{2+} site is the thermodynamically strongest SO_2 binding site (see Fig. 1), but is sufficiently weak to be almost entirely desorbed upon reducing the pressure. The material can also be fully regenerated by heating to 400 K without any loss of crystallinity even after completing fifty SO_2 adsorption-desorption cycles at 298 K .¹⁹

Our group reported a partially fluorinated version of MIL-101(Cr) named MIL-101(Cr)-4F(1%).²⁰ The particularity of MIL-101(Cr)-4F(1%) is the different chemical character of the Cr^{3+} metal centres, in comparison to MIL-101(Cr). The incorporation of fluorine in MIL-101(Cr) promoted a higher acidity of some of the Cr^{3+} metal centres (open metal sites), due to the capability of fluorine to attract electrons.²⁰ Such difference afforded a total

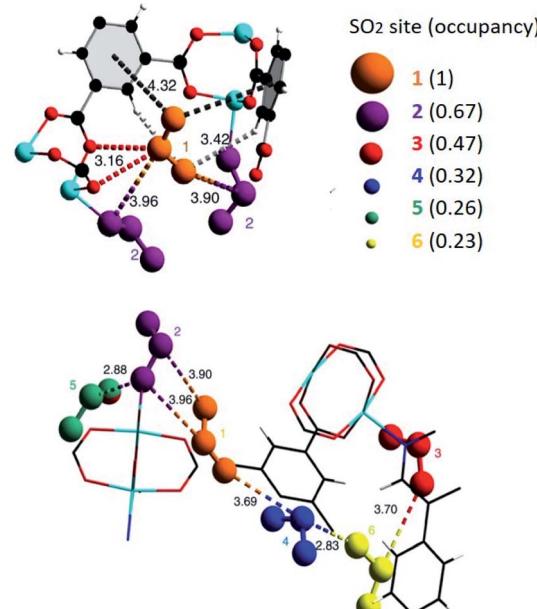


Fig. 1 SO_2 positions within MFM-170 pores obtained from *in situ* single-crystal X-ray diffraction. Purple SO_2 molecule is located at the thermodynamically strongest binding site. Cu, cyan; C, black; O, red; N, blue. Reprinted (adapted) with permission from ref. 19. Copyright (2019) Springer Nature.



SO₂ capture of 18.4 mmol g⁻¹ at 298 K and up to 1 bar, chemical stability towards dry and humid SO₂ and an exceptional cycling performance with facile regeneration.²⁰

This uptake represents the highest SO₂ capture for a structurally stable MOF material. In addition, *in situ* DRIFT spectroscopy upon the adsorption of CO demonstrated the efficient packing of SO₂ molecules within MIL-101(Cr)-4F(1%). Thus, the SO₂ uptake mechanism takes place in three stages: (i) adsorption at acid (Lewis) Cr³⁺ sites with a relatively high heat of adsorption for SO₂; (ii) adsorption at both acid (Lewis and Brønsted) sites of MIL-101(Cr)-4F(1%) and (iii) adsorption within the cavities of this MOF material. Up to this point, we have presented three examples (MOF-177, MFM-170 and MIL-101(Cr)-4F(1%)) of MOF materials with coordinatively unsaturated metal sites that demonstrated the highest SO₂ captures for any MOF. One material (MOF-177) demonstrated poor structural stability towards SO₂ *versus* the high structural stability of MFM-170 and MIL-101(Cr)-4F(1%) in the presence of SO₂ and humid SO₂. Unwittingly, the fundamental question arises: why is MOF-177 unstable while MFM-170 and MIL-101(Cr)-4F(1%) are stable to SO₂? In order to answer this question, their comparison is presented in Table 1, allowing to propose a stability hypothesis for these materials.

The design of stable frameworks for capture, confinement and release of corrosive gases uses tools centred in the robustness of the metal-linker bond, the use of inert or higher-valent metals and, the use of poly-nuclear secondary building units (SBUs).²¹ In the case of MOF-177, MFM-170 and MIL-101(Cr)-4F(1%), these exhibit at least one of such design tools. These three MOFs are composed of carboxylate ligands and robust SBUs, this multinuclearity gives, in theory, a better thermal and chemical stability. Nonetheless, in the case of MOF-177,¹⁸ that possess the highest BET surface area, pore volume and the

record of SO₂ capture, the use of robust polynuclear cluster [Zn₄(μ₄-O)(O₂CR)₆] (see comparative Table 1), is not enough to achieve a reversible SO₂ adsorption and the framework suffers a partial decomposition.

This result is most likely due to the propensity of the Zn–O bond to bind SO₂ and form SO₃²⁻. Similar behaviour was observed on a ZnO surface exposed to SO₂ gas, that reacts with the oxygen atoms from the Zn–O bonds forming the SO₃²⁻ and SO₄²⁻ species.^{22,23} In fact, other Zn-based MOFs that have been studied for SO₂ adsorption have shown similar stability problems. This is also the case of Zn-MOF-74 (ref. 24) with a rod-like SBU and coordinatively unsaturated zinc sites, where a Zn–O=S=O interaction leads to a chemisorption. Additionally, materials with paddlewheel SBUs, such as Zn²⁺-BDC (MOF-2) where the catalytic oxidation of SO₂ under humid conditions was confirmed by the presence of sulphate species.²⁵ Conversely, when the coordinatively unsaturated zinc site is blocked (*i.e.*, pillared MOFs), direct interaction of adsorbate with the metal centre is not observed, as for example in Zn(bdc)(ted)_{0.5} (ref. 26) and Zn-DMOF.²⁷ These comparative results suggest that in the case of Zn-based materials, the robustness of the SBU plays a secondary role in chemical stability against SO₂, and the fact that they are constructed from a divalent metal that forms labile bonds,²⁸ makes them more susceptible to collapse.

Although MFM-170 contains the Cu²⁺ paddlewheel SBU, a cluster frequently unstable under harsh conditions such as corrosive or acidic gases,^{29,30} it is stable even after 50 adsorption–desorption cycles of SO₂.¹⁹ This extraordinary chemical stability is mainly due to two important factors: (i) the unusual paddlewheel SBU formed in the framework and (ii) additional intermolecular interactions formed between SO₂ and the pore-walls of the framework. The di-cooper paddlewheel SBU shows the particularity of only one of the Cu²⁺ centres axially

Table 1 Comparison of structural characteristics in metal–organic frameworks with the highest SO₂ capture

Material	MOF-177	MFM-170	MIL-101(Cr)-4F(1%)
Ligand	H ₃ BTB	H ₄ L = 4',4''-(pyridine-3,5-diyl)bis([1,1'-biphenyl]-3,5-dicarboxylic acid)	H ₂ BDC/H ₂ BDC-4F
SBU	[Zn ₄ (μ ₄ -O)(O ₂ CR) ₆]	[Cu ₂ (O ₂ CR) ₄]	[Cr ₃ (μ ₃ -O)(O ₂ CR) ₆]
BET surface area [m ² g ⁻¹]	4100	2408	2176
Pore volume [cm ³ g ⁻¹]	1.51	0.88	1.19
SO ₂ adsorption [mmol g ⁻¹]	25.7	17.5	18.4
Packing density [cm ³ g ⁻¹]	1.09	1.27	1.00
Stability under SO ₂	Partial destruction after SO ₂ exposure	Stable after 50 cycles	Stable after 50 cycles



coordinated to a nitrogen atom from the pyridyl-based ligand, leaving the other Cu^{2+} metal centre coordinatively unsaturated and pointing toward the centre of the pore (Table 1). The effect of this structural arrangement is evident on the high SO_2 uptake, where the Cu^{2+} sites are able to fix SO_2 by forming a strong Cu–O bond without ligand displacement. Such behaviour was previously observed in a Cu^{2+} -BDC MOF material.²⁵ It is significant that this interaction does not represent the main SO_2 binding site due to the steric hindrance generated by a neighbouring SO_2 molecule. This latter molecule has several interactions: (i) with the C–H portion of the ligand, (ii) with the oxygen atoms in the paddlewheel SBU, and (iii) with other SO_2 molecules. All these interactions promote an efficient packing within the pore surface, maximising the adsorbate–adsorbent bonds and presumably avoids the frame destruction. Both materials reviewed above, MOF-170 and MFM-177, are formed by first-row transition metals Zn^{2+} and Cu^{2+} , respectively, and follow the stability order of the Irving–Williams series for divalent metals: $\text{Mn} < \text{Fe} < \text{Co} < \text{Ni} < \text{Cu} > \text{Zn}$.³¹ Thus, the high SO_2 capture can be attributed to the large BET surface area and pore size rather than to the local pore environment, while structural stability is associated with the type of metal, and the interactions adsorbate–adsorbent.

Additionally, the use of higher-valent metal ions such as Al^{3+} , Sc^{3+} , Cr^{3+} or Zr^{4+} allows the formation of stronger M–O bonds with carboxylate ligands compared to divalent ions.²¹ Thus, the lability of metal–ligand bonding is higher in Cu^{2+} or Zn^{2+} -based MOFs than Cr^{3+} MOFs. As a consequence, MOFs constructed from divalent metals such Cu^{2+} and Zn^{2+} , typically show partial or total degradation under harsh conditions.³² Due to the inertness of the Cr–O bond, Cr^{3+} -based MOFs highly resistant to the attack of acid and base guests were reported.^{33,34} As described above, the third material, MIL-101(Cr)-4F(1%), has the highest SO_2 adsorption among SO_2 -stable MOFs.²⁰ This framework poses a robust trinuclear metal cluster with coordinatively unsaturated chromium sites after the activation process (Table 1). When comparing MIL-101(Cr)-4F(1%) to MFM-170, both materials stable towards SO_2 , we can observe that the Cr^{3+} is harder than Cu^{2+} ions according to the Pearson acid–base concept³⁵ and therefore the coordinated bond with carboxylic ligands is stronger for the trivalent cation than for the divalent cation. However, the Cu^{2+} paddlewheel increases its stability when the N-containing ligand coordinates to one of the Cu^{2+} ions.³⁶

(ii) μ -OH groups. This particular hydroxo-functionalisation in MOF materials is well exemplified by the MFM-300(M) (MFM = Manchester Framework Material, M = Al^{3+} , In^{3+} , Ga^{3+} and Sc^{3+}) family. These outstanding MOF materials are chemically stable towards SO_2 and have been reported by Schröder and Yang.³⁷ The general chemical formula of this group of MOFs is $[\text{M}_2(\text{OH})_2(\text{L})] (\text{H}_4\text{L} = \text{biphenyl-3,3',5,5'-tetra-carboxylic acid} = \text{C}_{16}\text{O}_8\text{H}_6)$. MFM-300(M) shows a 3D open framework which contains $[\text{MO}_4(\text{OH})_2]$ octahedra connected via the *cis*- μ -OH groups into infinite chains, and further coordinated by the tetradentate ligand (L^{4-}). This particular organisation produces highly porous materials (see Table 2), well defined by one-dimensional pore channels organized in a “wine rack” display.

The first MFM-300 material investigated for the capture of SO_2 was MFM-300(Al).^{37a} Although the SO_2 capture was not remarkably high (see Table 2), the identification of the preferential SO_2 adsorption sites was achieved by sophisticated *in situ* INS and PXRD experiments, revealing that the μ -OH groups bind SO_2 molecules through the formation of $\text{O}=\text{S}=\text{O}(\delta^-)\cdots\text{H}(\delta^+)-\text{O}$ hydrogen bonds, reinforced by weak supramolecular interactions with C–H atoms from the aromatic rings of the framework (Fig. 2).^{37a} In the case of MFM-300(Sc) (also known as InOF-1 (ref. 38)) the SO_2 uptake was slightly higher than for MFM-300(Al) (see Table 2), and *in situ* INS and PXRD experiments corroborated the same preferential adsorption site for SO_2 (see Fig. 3) with exceptional chemical stability for SO_2 capture under both dry and humid conditions.^{37b} In fact, Eddaoudi and Salama fabricated an advanced chemical capacitive sensor, using MFM-300(Sc), for the detection of very low concentrations of SO_2 (≈ 5 ppb) at room temperature.³⁹

MFM-300(Sc) showed the highest SO_2 uptake of the MFM-300(M) materials (see Table 2), Grand Canonical Monte Carlo (GCMC) simulations demonstrated the same preferential adsorption sites (μ -OH) and by confining small amounts of EtOH (2.6 wt%) the SO_2 capture increased by 40% (see Fig. 4).^{37c} The only MFM-300 remaining to be investigated in the capture of SO_2 is MFM-300(Ga),^{37d} where we anticipate, based on the pore volume, performance similar to MFM-300(Sc).

Thus, the MFM-300(M) MOF materials have shown the relevance of the hydroxo-functionalisation to the capture and detection of SO_2 with high chemical stability and excellent cyclability involving a remarkably facile regeneration. The key for all of such outstanding properties is the strength of the coordination bonds between the oxygen atoms form the

Table 2 Comparison of MFM-300(M) family

Metal center (M)	BET Surface area [$\text{m}^2 \text{ g}^{-1}$]	Pore volume [$\text{cm}^3 \text{ g}^{-1}$]	SO_2 adsorption [mmol g^{-1}]	Packing density [g cm^{-3}]	Q_{st} [kJ mol^{-1}]	Stability
Al	1370	0.375	7.1	1.213	31.25 ^a ^{37e}	Stable over 4 years of exposure
In	1071	0.419	8.28	1.266	34.5	Stable in dry and humid conditions
Ga	1045	0.62	—	—	—	—
Sc	1360	0.56	9.4	1.075	36.2	Stable in dry and humid conditions/over 10 cycles

^a Adsorption near the zero-coverage calculated by GCMC.



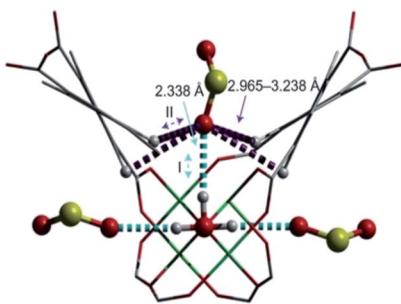


Fig. 2 Preferred binding sites for SO_2 molecules inside MFM-300(Al) determined by *in situ* PXRD. (I) Moderated hydrogen bond. (II) Weak hydrogen bond. Reprinted (adapted) with permission from ref. 37a. Copyright (2012) Springer Nature.

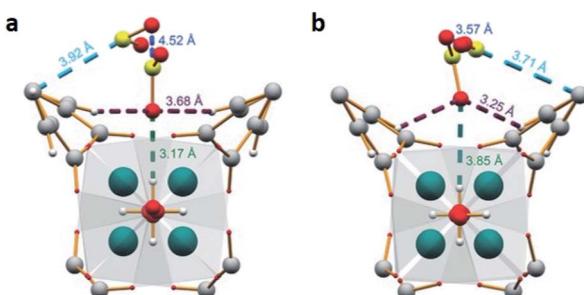


Fig. 3 Binding sites for SO_2 molecules within MFM-300(Al) determined by single-crystal X-ray diffraction (a), and DFT calculations (b). Reprinted (adapted) with permission from ref. 37b. Copyright (2016) John Wiley & Sons.

carboxylic ligand (biphenyl-3,3',5,5'-tetracarboxylic acid) and the M^{3+} metal centres. The ionic radii of Al^{3+} , In^{3+} and Sc^{3+} cations, in octahedral coordination, are 0.675, 0.940 and 0.885 Å, respectively.^{40a}

This leads to a higher surface charge density for the Al^{3+} metal centre and, therefore, the $\text{M}-\text{O}$ bond strength decreases

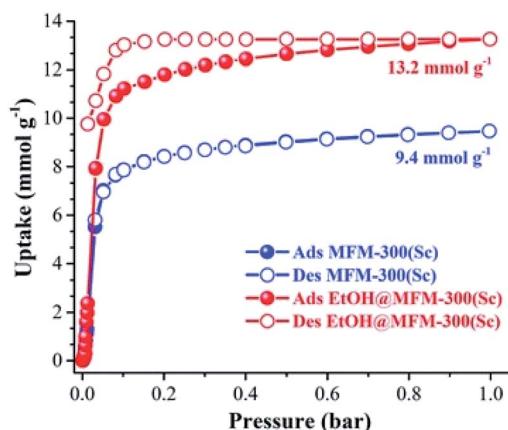


Fig. 4 SO_2 adsorption isotherm of scandium-based MFM-300 (blue isotherm) and after confinement of 2.8% EtOH (red isotherm). Reprinted with permission from ref. 37c. Published (2019) by The Royal Society of Chemistry.

in this order: $\text{Al}^{3+} > \text{Sc}^{3+} > \text{In}^{3+}$. Thus, a higher ligand exchange can occur in the In-based complex.^{40b} However, MFM-300(Al) does not suffer apparent displacement of carboxylate ligands after SO_2 exposure. In the MFM-300(M) family all the metal cations are nonmagnetic because of all paired electrons,⁴¹ and the comparison between octahedral complexes formed by trivalent cations of group 13 such as Al^{3+} , Ga^{3+} and In^{3+} , with the transition-metal analogue Sc^{3+} , has shown that in the group 13 metal complexes the d orbitals are not involved in the metal-linker bonding due to their high energy ($3d^{10}$ and $4d^{10}$ configuration). In the case of Sc^{3+} , a complex with $12e^-$, d orbitals form part of the $\text{M}-\text{O}$ bonds (Fig. 5a). On the other hand, these types of complexes are considered as electron-rich hypervalent species with 7-center-12-electron bonding pattern (Fig. 5b).^{40c} This type of binding confers to the group 13 complexes thermal stability and in this particular case, the kinetic stability is comparable to that of the transition metal analogues. At this point, we can corroborate that frameworks constructed with trivalent cations are more stable than MOFs based on divalent cations (considering the stability of Irving-Williams series discussed above), as well that carboxylate ligands form with M^{3+} cations stronger bonds than divalent cations with N-based ligands.³²

(iii) **Defective sites.** Crystal irregularities (composition inhomogeneities or defects) are fundamental characteristics of some solid-state materials and provide very particular physical and chemical properties.⁴² Such defects do not necessarily mean adverse effects. Due to their extraordinary modularity and tunability, MOFs allow the introduction of different kinds of defects while maintaining the overall structure integrity. Typically, “defective MOFs” have showed extraordinary superior catalysis performances,⁴³ while Navarro and co-workers have taken a step forward and elegantly demonstrated the self-detoxification of defective UiO-66 examples to filter chemical-

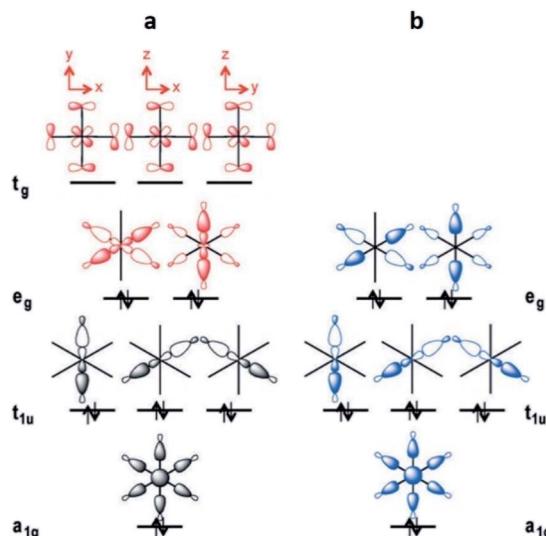


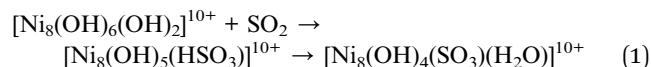
Fig. 5 Schematic MO diagrams for an O_h -symmetric $12e^-$ complex with d orbitals (a, red) and the MO diagram without d orbitals in 7c- $12e^-$ pattern of MO (b, blue). Reprinted (adapted) with permission from ref. 40c. Copyright (2015) John Wiley & Sons.



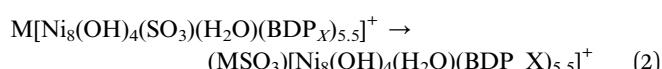
warfare agents.⁴⁴ In the case of SO_2 , the same research group reported on the increase of the SO_2 adsorption capacities, energies and SO_2/CO_2 selectivity for defective nickel pyrazolate MOF materials, prepared by introducing extra-framework Ba^{2+} ions into the porous structure and ligand functionalisation.⁴⁵

Thus, the post-synthetic treatment of the $[\text{Ni}_8(\text{OH})_4(\text{H}_2\text{O})_2(\text{BDP}_X)_6]$ ($\text{H}_2\text{BDP}_X = 1,4\text{-bis(pyrazol-4-yl) benzene-4-X}$ with $X = \text{H}$ (1), OH (2), NH_2 (3)) systems with ethanolic solutions of potassium hydroxide leads to the formation of defective K $[\text{Ni}_8(\text{OH})_3(\text{EtO})_3(\text{BDP}_X)_{5.5}]$ (1@KOH, 3@KOH) and $\text{K}_3[\text{Ni}_8(\text{OH})_3(\text{EtO})(\text{BDP}_O)_5]$ (2@KOH).⁴⁶ Exposing the 1@KOH-3@KOH materials to aqueous $\text{Ba}(\text{NO}_3)_2$ solutions incorporates barium cations into the solution, yielding the defective ion exchanged $\text{Ba}_{0.5}[\text{Ni}_8(\text{OH})_3(\text{EtO})_3(\text{BDP}_X)_{5.5}]$ (1@Ba(OH)₂, $X = \text{H}$; 3@Ba(OH)₂, $X = \text{NH}_2$), and $\text{Ba}_{1.5}[\text{Ni}_8(\text{OH})_3(\text{EtO})(\text{BDP}_O)_5]$ (2@Ba(OH)₂) systems.⁴⁵ Interestingly, density functional theory (DFT) calculations located the extra-framework cations close to the crystal defect sites (Fig. 6a). In addition, DFT results demonstrated that the average 3D structure of the MOF framework is preserved.⁴⁵ SO_2 dynamic adsorption experiments confirmed the beneficial influence of the deliberate introduction of defects by a subsequent K^+ to Ba^{2+} ion exchange process on the SO_2 capture.⁴⁵ The authors demonstrated that the pre-synthetic introduction of amino and hydroxyl functional groups on the organic linkers and the post-synthetic modifications, synergistically increase the SO_2 capture capacity of these materials. In addition, chemisorption of SO_2 was observed in all the systems: after all available sites for SO_2 chemisorption were occupied, during the first exposure, reversible adsorption takes place in a steady way

evidencing the stability of the materials upon continuous SO_2 exposure. This high affinity of these MOF materials for SO_2 was attributed to the increased basicity of the nickel hydroxide clusters after the introduction of defects. Lewis acid-base interactions with SO_2 molecules can afford the formation of HSO_3^- and SO_3^{2-} species according to eqn (1):



Furthermore, the existence of extra-framework cations located close to defects sites promotes the formation of more stable MSO_3 ($\text{M} = \text{Ba}, 2$ K) sulphite species according to eqn (2):



These chemisorption mechanisms were supported by the DFT simulations (Fig. 6), showing that 1@Ba(OH)₂ establishes stronger interactions with SO_2 molecules than 1@KOH due to the specific interactions with extra-framework barium cations.

Thus, this sophisticated and comprehensive study, by Navarro *et al.*,⁴⁵ carefully incorporated essential variables to improve the SO_2 capture performance of these MOF materials. Such variables play together to increase the interaction with SO_2 and can be summarised as (i) enhanced basicity of metal hydroxide clusters as a result of additional hydroxide anions replacing the missing linkers defects; (ii) affinity of extra-framework Ba^{2+} ions for SO_2 sequestration; (iii) higher pore accessibility of the 3-D structure due to missing linker defects and (iv) the fine-tuning of the pore surface polarity by the benzene functional groups.

The post-synthetic inclusion of Ba^{2+} cations to improve the SO_2 adsorption was later validated by Navarro and co-workers⁴⁷ in the Ni-based MOF material $\text{Ni}_2^{2+}[\text{Ni}_4^{2+}[\text{Cu}_4^{2+}-(\text{Me}_3\text{mpba})_2]_3] \cdot 54\text{H}_2\text{O}$ (1). In this case, the exchange of the Ni^{2+} cations for Ba^{2+} was accomplished by soaking 1 in an aqueous solution of $\text{Ba}(\text{NO}_3)_2$ for two days to finally obtain a novel crystalline phase of heterotrimetallic oxamato-based MOF with the formula $[\text{Ba}^{2+}(\text{H}_2\text{O})_4]_{1.5}[\text{Ba}^{2+}(\text{H}_2\text{O})_5]_{0.5} \cdot \{\text{Ni}_4^{2+}[\text{Cu}_2^{2+}-(\text{Me}_3\text{mpba})_2]_3\} \cdot 57.5\text{H}_2\text{O}$ (2). On this occasion, the presence of the hydrated Ba^{2+} counterions within the MOF's pores allowed the SO_2 molecules to interact with the network reversibly *via* $\text{Cu}^{2+} \cdots \text{O}_{\text{SO}_2}$ and $\text{O}_{\text{H}_2\text{O}} \cdots \text{O}_{\text{SO}_2}$ interactions. The amount of SO_2 adsorbed in 2 increased from 2.0 to 2.5 mmol g⁻¹ in comparison to 1, respectively. This reversibility was conserved over 10 adsorption-desorption cycles. Theoretical calculations demonstrated that the principal binding site of SO_2 molecules is with coordinated water molecules, and it showed moderate adsorption energy of $-65.5 \text{ kJ mol}^{-1}$, otherwise, the direct interaction of SO_2 with Ba^{2+} cations would induce a higher adsorption energy ($>100 \text{ kJ mol}^{-1}$) due to the formation of an irreversible bond. Thus, this research group has elegantly proved that the formation of defective sites and the incorporation of large cations such as Ba^{2+} is a good and feasible post-synthetic method for modulating the adsorptive capacities of MOF materials.

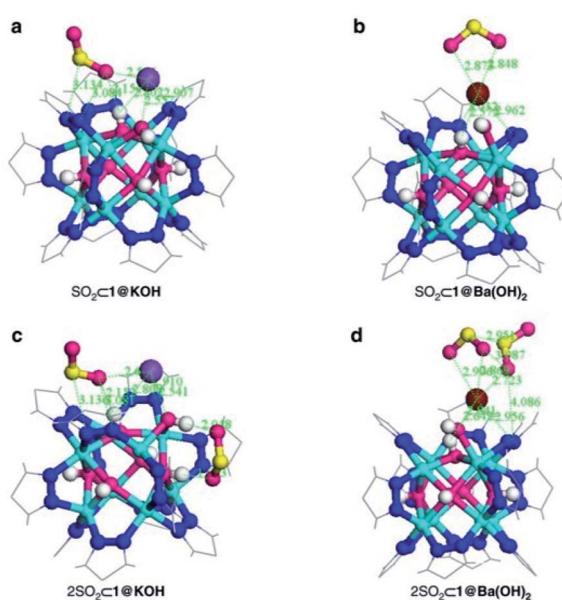


Fig. 6 DFT structure minimization of molecular configuration of one (a and b) and two (c and d) adsorbed SO_2 molecules on 1@KOH (left) and 1@Ba(OH)₂ (right) materials. For sake of clarity, only the region around the metal cluster is shown. Ni (cyan); K (purple); Ba (wine); C (grey); N (blue); O (magenta); H (white); S (yellow). Reprinted (adapted) with permission from ref. 45. Copyright (2017) Springer Nature.

(iv) Halogen functionalisation. Halogen functionalised materials are developing as promising technological platforms for innovative applications in different research fields such as catalysis, environmental remediation, sensing, and energy transformation.⁴⁸ On the other hand, cationic or anionic frameworks have shown potential applications in the capture of gases as the ions present inside the pores of the MOFs can be a benefit for the interaction with guest molecules.⁴⁹ The presence of halogen counterions in porous materials has demonstrated an enhancement of the SO_2 capture performance due to the involvement of ionic bonding.⁵⁰ The following section is devoted to MOFs with halogen ions in their structure that participate in the adsorption of gaseous SO_2 .

Yang and Xing incorporated inorganic hexafluorosilicate (SiF_6^{2-} , SIFSIX) anions (as pillars) into a series of MOF materials (SIFSIX-1-Cu, SIFSIX-2-Cu (2 = 4,4'-dipyridylacetylene), SIFSIX-2-Cu-i, SIFSIX-3-Zn (3 = pyrazine), and SIFSIX-3-Ni) and investigated the SO_2 adsorption properties of these materials.⁵¹ The remarkable highly efficient removal of SO_2 from other gases, particularly at a very low SO_2 concentrations, and excellent SO_2/CO_2 and SO_2/N_2 selectivities were attributed to the strong electrostatic interactions between the SO_2 molecules and the SiF_6^{2-} anions ($\text{S}^{\delta+}\cdots\text{F}^{\delta-}$), assisted by dipole–dipole interactions with the ligand ($\text{O}^{\delta-}\cdots\text{H}^{\delta+}$).⁵¹ The authors identified the interactions between the SO_2 molecule and SIFSIX materials by modelling studies using first-principles DFT-D (dispersion-corrected density functional theory) calculations (see Fig. 7).⁵¹ These computational calculations were experimentally corroborated by Rietveld refinement of the powder X-ray diffraction patterns of SO_2 -loaded samples to locate the adsorbed positions of the SO_2 molecules in the crystal structure of SIFSIX materials.⁵¹ Thus, this strategy highlighted the relevance on creating the multiple binding sites (anionic and aromatic linkers) which can originate the specific recognition of SO_2 to optimise its capture.

Similarly, Salama and Eddaoudi demonstrated how fluorinated MOFs (KAUST-7 and KAUST-8) can be formidable candidates for sensing SO_2 from flue gas and air (250 ppm to 7% of SO_2).⁵² SCXRD data collected on the SO_2 -loaded KAUST-7 and KAUST-8 identified the preferential adsorption sites for SO_2

revealing, comparable to SIFSIX materials, strong electrostatic interactions between the SO_2 molecules and the fluorinated pillars ($\text{S}^{\delta+}\cdots\text{F}^{\delta-}$), supported by dipole–dipole interactions with the pyrazine ligand ($\text{O}^{\delta-}\cdots\text{H}^{\delta+}$), see Fig. 8.⁵² These interactions were also corroborated by DFT calculations.

Another remarkable halogen functionalisation was presented by Janiak and co-workers,⁵³ where MOF-801 was modified by reacting zirconium halides (ZrX_4 ; X = Cl, Br, I) in water with acetylenedicarboxylic acid. The HX addition and material construction occurred in a one-pot reaction yielding three microporous HHU-2-X MOFs (X = Cl, Br, I). The material HHU-2-Cl showed in increased SO_2 uptake (by 21%) in comparison to the nonhalogenated MOF-801.⁵³ Therefore, as previously demonstrated by the SIFSIX materials, KAUST-7, and KAUST-8, the electrostatic interactions between the SO_2 molecules and the Cl ($\text{S}^{\delta+}\cdots\text{Cl}^{\delta-}$), might be responsible for such SO_2 capture enhancement.

3. Hydrogen sulphide

H_2S is released to the environment *via* natural events such as volcanic eruptions, gas streams, hot springs, decomposition of organic matter and bacterial reduction.⁵⁴ Additionally, H_2S is emitted by some chemical industries, *e.g.*, oil desulphurisation processes at oil refineries, burning fossil fuels and mass transportation.⁵⁵ H_2S is a main air pollutant which negative impacts the environmental as it is one of the main sources of acid rain,⁵⁵ and it is highly toxic to humans (concentrations above 700 ppm in the air can cause death).⁵⁶

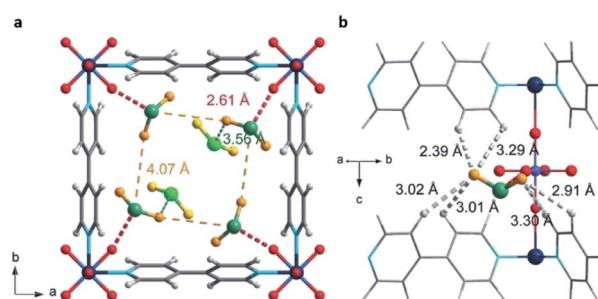


Fig. 7 Binding sites for SO_2 molecules adsorbed within SIFSIX-1-Cu determined by DFT-D calculations. Interaction $\text{S}^{\delta+}\cdots\text{F}^{\delta-}$ (a), and dipole–dipole interactions with the ligand $\text{O}^{\delta-}\cdots\text{H}^{\delta+}$ (b). Cu (dark teal); F (red); Si (light blue); C (grey); H (light grey); N (sky blue); O (orange); S (sea green). Reprinted (adapted) with permission from ref. 51. Copyright (2017) John Wiley & Sons.

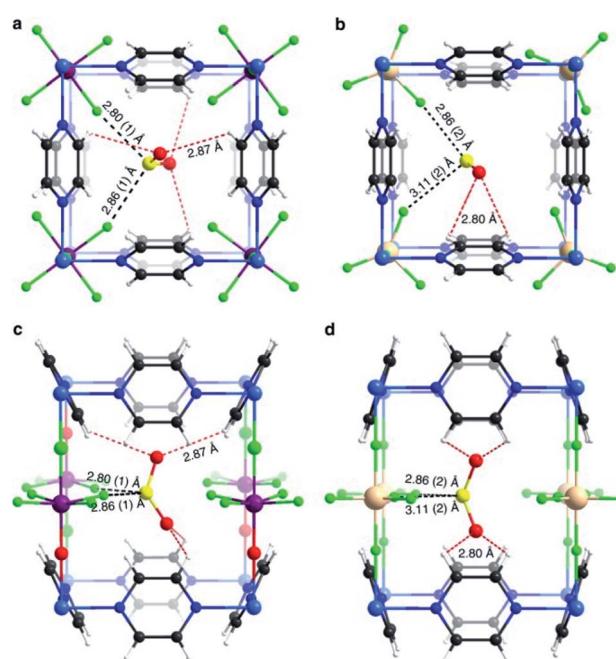


Fig. 8 SO_2 interactions within KAUST-7 (a and c), and KAUST-8 (b and d). Ni (blue light); Nb (purple); Al (beige); F (green); C (black); N (blue); O (red); H (white); S (yellow). Reprinted (adapted) with permission from ref. 52. Copyright (2019) Springer Nature.



H_2S is a colourless, flammable gas with a characteristic rotten egg odour. Typical strategies to capture H_2S comprise alkanolamines and ionic liquids, adsorption in solid materials (*i.e.*, zeolites, activated carbons and metal oxides), membrane separation and cryogenic distillation.⁵⁷ However, these techniques have major drawbacks, such as low H_2S capture, corrosion of pipelines and large cost of expenditure and recovery.⁵⁸ Thus, the development of new technologies for a dry adsorption process (avoiding solvents/water consumption is essential to reduce waste generation), for removal and capture of H_2S signifies a technological challenge which could allow not only the capture of H_2S from the main source, but also its use as feedstock, similarly to the CO_2 technologies (CCS).⁵⁹ MOFs have been visualised for the capture of H_2S ; however, most of them have shown poor chemical stability in its presence.⁶⁰ Nonetheless, chemically stable MOF materials have demonstrated promising results in the reversible capture of H_2S .⁶¹ This section highlights outstanding MOF materials that have shown high H_2S capture performances and *in situ* H_2S transformations, as well as biomedical applications.

3.1 MOFs for H_2S capture

Distinctive investigations reported by De Weireld,⁶² Zou⁶³ and Eddaoudi⁶⁴ reveal that the majority of MOF materials experience structural decomposition upon adsorption of H_2S or the desorption of it resulted complicated, due to relatively strong host–guest binding in the pores (strong physisorption or chemisorption). In such cases, the reactivation of the materials involves a large energy penalty or is not feasible. Thus, the identification of new MOFs capable of capturing H_2S under industrially practical pressure-swing desorption conditions,⁶⁵ represents a very important challenge to solve. This perspective focused on (i) high and reversible H_2S capture in MOFs and (ii) chemical transformation of H_2S within the pores of MOFs: formation of polysulphides.

(i) High and reversible H_2S capture in MOFs. The capture of H_2S by MOFs has confirmed some crucial difficulties such as the formation of strong bonds, typically irreversible, (*e.g.*, a metal–sulphur bond), which can compromise the chemical stability and cyclability of MOFs. Therefore, it is required to modulate host–guest interactions between the MOFs and H_2S to avoid structure collapse and afford a feasible cyclability. In order to achieve these goals, such interactions should arise through noncovalent bonding between the functionalisation of MOFs and H_2S guest molecules. For example, hydrogen bonding has been postulated, by density functional theory (DFT) methods and grand canonical Monte Carlo (GCMC) simulations,⁶⁶ as a promising moderate interaction between MOFs and H_2S molecules. Experimentally, a remarkable work by Hamon and co-workers showed the high chemical stability and cyclability of MIL-47 ($\text{M} = \text{V}^{4+}$) and MIL-53(M) ($\text{M} = \text{Al}^{3+}$, Cr^{3+}) when exposed to H_2S .⁶⁷ These MOF materials are constructed with $\mu\text{-OH}$ functional groups which presumably established hydrogen bonds with H_2S .

Humphrey and Maurin showed the H_2S capture on Mg-CUK-1.⁶⁸ This MOF is assembled from a 2,4-pyridinecarboxylate

ligand and Mg^{2+} octahedral centres, connected into infinite chains of $[\text{Mg}_3(\mu_3\text{-OH})]^{5+}$ clusters by $\mu_3\text{-OH}$ groups. Although the total H_2S capture of Mg-CUK-1 was relatively low (3.1 mmol g⁻¹), this material demonstrated to be chemically stable to H_2S and a remarkably easy regeneration after the H_2S capture using temperature swing re-activation (TSR). Most importantly, GCMC simulations showed how H_2S molecules interact with the hydroxo functional groups (by moderate hydrogen bonds $\mu_3\text{-O-H}\cdots\text{SH}_2$), see Fig. 9.⁶⁸ Such moderate $\text{H}_2\text{S}/\text{Mg-CUK-1}$ interaction was found to be consistent with easy regeneration of the material after H_2S exposure.

This work demonstrated experimentally and in good correlation with advanced computational calculations, the chemical structure integrity of a hydroxo-functionalised MOF material to H_2S , and the critical role of a moderate interaction (hydrogen bonding) to facilitate its cyclability. The study of more $\mu\text{-OH}$ functionalised MOFs for the efficient and cyclable capture of H_2S initiated.

Maurin and Gutiérrez-Alejandre investigated another $\mu\text{-OH}$ functionalised MOF (MIL-53(Al)-TDC) for the capture of H_2S .⁶⁹ This material, an Al^{3+} -based constructed with a carboxylate ligand (TDC = 2,5-thiophenedicarboxylate) which contains $[\text{AlO}_4\text{-trans-(}\mu\text{-OH)}_2]$ octahedra where the Al^{3+} centre coordinates to two $\mu\text{-OH}$ groups and six oxygen atoms from the TDC ligands, has established the highest H_2S adsorption (18.1 mmol g⁻¹ at room temperature) reported for a microporous material.⁶⁹ Structural stability of MIL-53(Al)-TDC, after the H_2S capture experiment was corroborated by PXRD and SEM analyses. TGA experiments demonstrated the complete desorption of H_2S molecules at 65 °C. H_2S sorption–desorption cycles (five cycles with a value of 18.5 ± 0.7 mmol g⁻¹) demonstrated high H_2S regeneration capacity. This cyclability confirmed the full regeneration of the material by only increasing the temperature to 65 °C, showing the low energy requirement to fully desorb H_2S . This suitable cyclability indicated weak interactions between H_2S molecules and the pores of the material. *In situ* DRIFT experiments investigated the interactions between H_2S and MIL-53(Al)-TDC, showing (i) the formation of hydrogen bonds between H_2S molecules themselves confined in the pores of MIL-53(Al)-TDC; (ii) relatively weak interaction between H_2S and the $\mu\text{-OH}$ groups and (iii) H_2S molecules interact with the

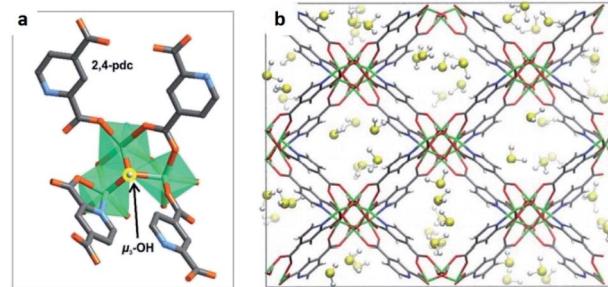


Fig. 9 H_2S adsorption preferred site of magnesium-based CUK-1 material. And after confinement of 2.8% EtOH (red isotherm). Reprinted (adapted) with permission from ref. 68. Published (2018) by The Royal Society of Chemistry.



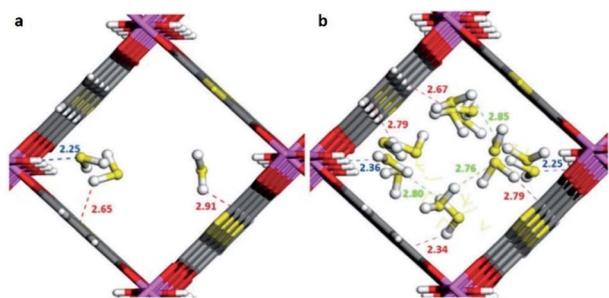


Fig. 10 Illustrative arrangements of H_2S in the pores of MIL-53-TDC generated from the GCMC simulations at (a) 0.5 mmol g^{-1} and (b) 18.5 mmol g^{-1} . The distances are reported in Å. (Al, pink; O, red; S, yellow; C, grey; H, white). Reprinted (adapted) with permission from ref. 69. Published (2019) by The Royal Society of Chemistry.

other functionality of the MOF: the thiophene ring from the TDC ligand.⁶⁹ To corroborate these interpretations from the DRIFT experiments, Monte Carlo simulations were carried out for different loadings corresponding to the experimental discoveries. These calculations corroborated that at low loading, H_2S interacts *via* its S-atom with the H-atom of the μ -OH group (see Fig. 10), representing the preferential adsorption site. In addition, it was shown that H_2S also interacts with the thiophene ligand, and, upon increasing the H_2S loading, hydrogen bonds between the H_2S molecules were also identified (Fig. 10).⁶⁹

These theoretical findings revealed the fundamental role of the thiophene ligand. Experimentally speaking, this M-OH hydroxo group containing functional material shows an extraordinarily facile re-activation leading to an easy H_2S cyclability, which should not be anticipated due to the dominant interaction between the μ -OH group and H_2S reinforced further by the interaction of the H_2S molecule with the thiophene unit as demonstrated by *in situ* DRIFT measurements. However, computational calculations postulate that the interaction of H_2S with the thiophene ring weakens the interaction with the preferential adsorption site (μ -OH group), favouring the easy displacement of the H_2S molecule when the material is re-activated.

As previously discussed, (*vide supra*) for the MFM-300(M) family, the extraordinary chemical stability arises from the strength of the coordination bonds between the oxygen atoms from the carboxylic ligand and the M^{3+} metal centres. Thus, MIL-53(Al)-TDC is very similar to MFM-300(Al) since it is also constructed with Al^{3+} . Therefore, the chemical stability of MIL-53(Al)-TDC towards H_2S is envisaged based on the same arguments used for MFM-300(Al).

(ii) Chemical transformation of H_2S within the pores of MOFs: formation of polysulphides. As described in the previous section, the study of more hydroxo-functionalised MOF materials attracted high interest since the promising results on the effective and efficient cyclable capture of H_2S . Thus, a couple of materials from the MFM-300(M) family were investigated by Maurin and Gutiérrez-Alejandre,⁷⁰ for the capture of H_2S : MFM-300(Sc) and MFM-300(In).

First, MFM-300(Sc) exhibited a H_2S uptake of 16.5 mmol g^{-1} (at 25°C), which is comparable to the H_2S uptake of MIL-53(Al)-TDC (18.1 mmol g^{-1}).⁶⁹ Upon an inspection of the structural integrity of the material after the H_2S experiment, PXRD experiments revealed the retention of the crystalline structure. Interestingly, when investigating the porosity of MFM-300(Sc) after the H_2S uptake experiment, the pore volume was reduced by 34% from $0.56 \text{ cm}^3 \text{ g}^{-1}$ to $0.37 \text{ cm}^3 \text{ g}^{-1}$ indicating that some remaining species are still present in the pores of the material. After confirming the retention of the crystallinity and a reduction of the intrinsic porosity of MFM-300(Sc) after the first H_2S cycle, additional H_2S cycling experiments were performed. Over the second cycle, the H_2S adsorption capacity decreased by 39% to $10.08 \text{ mmol g}^{-1}$, which is in a good agreement with the reduction of the pore volume by about 34%. The average H_2S capture in cycles 2 to 5 was $10.22 \text{ mmol g}^{-1}$. After the fifth cycle, PXRD experiments corroborated the retention of the crystalline structure of MFM-300(Sc), while a N_2 adsorption experiment showed that the pore retained its reduced volume of $0.37 \text{ cm}^3 \text{ g}^{-1}$. In an attempt to remove the remaining sulphur species formed inside the pores of MFM-300(Sc), the sample after the first H_2S cycle was activated at 250°C . After this thermal treatment, the pore volume did not change ($0.37 \text{ cm}^3 \text{ g}^{-1}$) confirming that the sulphur species were irreversibly adsorbed within the MFM-300(Sc) framework.⁷⁰

The identification of these sulphur species was first approached by TGA and diffuse-reflectance infrared Fourier-transform spectroscopy (DRIFT) experiments which corroborated strong interactions between the guest species and the pore-walls of MFM-300(Sc). Raman spectroscopy, complemented with elemental analysis (EDX) and conventional elemental analysis, postulated the first indication of the possible nature of the remaining (irreversibly adsorbed) sulphur species within the pores, *i.e.*, polysulphides (see Fig. 11).⁷⁰ Even though the redox properties of polysulphides are highly complex, electrochemical experiments were the key to fully identify them. An electrochemical cell (MFM-300(Sc))-CSP/PVDF/1 M LiTFSI in triglyme/LiO (PVDF = poly(vinylidene difluoride), LiTFSI = $\text{LiN}[\text{SO}_2\text{CF}_3]_2$, triglyme = $\text{MeO}[\text{CH}_2\text{CH}_2\text{O}]_3\text{Me}$) was examined (for 20 h) through open-circuit potential (OCP) measurements and analysed by cyclic voltammetry. Then, this electrochemical cell showed an initial potential of 2.29 V (vs. Li^0/Li^+) for MFM-300(Sc) and after reaching equilibrium, this potential was equal to 2.15 V (vs. Li^0/Li^+).

Later, a H_2S exposed sample of MFM-300(Sc) ($\text{H}_2\text{S}@\text{MFM-300(Sc)}$) was used to construct a different electrochemical cell showing a different behaviour on the variation of the potentials (the potential increased from 1.90 V to 2.33 V (vs. Li^0/Li^+)). According to Mikhaylik and Akridge,⁷¹ electrochemical potentials lower than 2.10 V correspond to low-order polysulphides (S_n^{2-}), ($n = 2$). Thus, the initial potential for $\text{H}_2\text{S}@\text{MFM-300(Sc)}$ of 1.90 V suggested the formation of such low-order S_2^{2-} polysulphides. The formation of these low-order species arises from the strong mutual $\text{H}_2\text{S}\cdots\text{H}_2\text{S}$ hydrogen bond interactions with a characteristic $\text{H}(\text{H}_2\text{S})\cdots\text{S}(\text{H}_2\text{S})$ distance of 2.91 \AA , as demonstrated in the corresponding RDF plot from the Monte Carlo simulations.



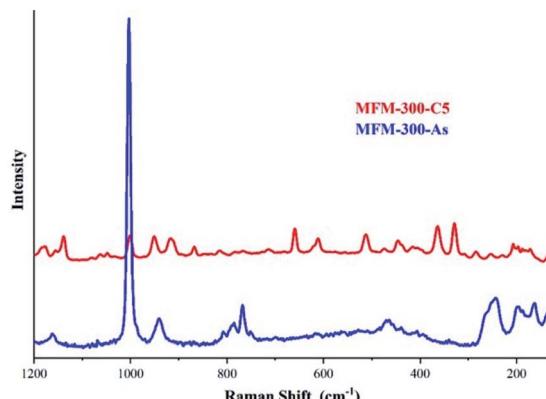


Fig. 11 Raman spectra of MFM-300(Sc) after 5 cycles of H_2S (red) and as synthesised (blue). The peaks observed in the $350\text{--}520\text{ cm}^{-1}$ region are associated to the S–S stretching vibrations modes of several polysulphides with various chain lengths. Reprinted with permission from ref. 70. Copyright (2020) American Chemical Society.

In the case of MFM-300(In) by constructing the equivalent electrochemical cell (MFM-300(In))-CSP/PVDF/1 M LiTFSI in triglyme/ Li^0 and same operation conditions (OCP measurements and analysed by cyclic voltammetry for 20 h), the electrochemical cell exhibited an initial potential of 2.65 V (vs. Li^0/Li^+) for MFM-300(In) and after reaching equilibrium, this potential was equal to 2.15 V (vs. Li^0/Li^+). The difference of 0.36 V, from the initial potentials for MFM-300(Sc) and MFM-300(In), was attributed to the difference in the electronegativity of the metal centres (Sc^{3+} and In^{3+}) in both materials. After the adsorption of H_2S , $\text{H}_2\text{S}@\text{MFM-300}(\text{In})$, MFM-300(In) showed a different trend than MFM-300(Sc): the potential decreased from 2.50 to 2.15 V (vs. Li^0/Li^+). Such a trend difference for both stabilisation potentials and the potential values indicated that the chemical composition of the sulphur species in both materials (MFM-300(Sc) and MFM-300(In)) were different. Again, Mikhaylik and Akridge indicate,⁷¹ that electrochemical potentials higher than 2.10 V correspond to high-order polysulphides (S_n^{2-}), ($n = 6$). Such large polysulphides block completely the pores of MFM-300(In), as corroborated by the loss of the pore volume in $\text{H}_2\text{S}@\text{MFM-300}(\text{In})$ ($0.02\text{ cm}^3\text{ g}^{-1}$) and Raman spectroscopy. Interestingly, although these two materials are isostructural, a small difference in the pore size (8.1 and 7.5 Å for MFM-300(Sc) and MFM-300(In), respectively) contributes along with the different electrochemical potential of both materials (different electron densities) to the formation of distinct polysulphide species.

Although there is a hypothesis that justifies the polysulphide formation in both materials (based on the “disproportionation” type reaction,⁷² were the protons of H_2S play the role of an oxidant and the sulphide plays the role of a reducing agent), there is a great opportunity to deeply investigate this phenomenon. For example, can the polysulphide formation be a consequence of reversible metal–ligand bonding upon the adsorption of H_2S ? Recently Brozek and co-workers⁷³ elegantly demonstrated metal–ligand dynamics for carboxylate benchmark MOFs, *via* variable-temperature diffuse reflectance infrared

Fourier transform spectroscopy (VT-DRIFTS) coupled with *ab initio* plane wave density functional theory. Thus, new exciting horizons can be discovered by taking a different approach when H_2S is adsorbed in MFM-300(M), which could explain such fascinating polysulphide formation.

The last MOF material, to date, that demonstrated the chemical transformation of H_2S to polysulphides is SU-101.⁷⁴

This bioinspired material was synthesised (by Inge *et al.*⁷⁴), using ellagic acid, a common natural antioxidant, and bismuth ($\text{Bi}_2\text{O}(\text{H}_2\text{O})_2(\text{C}_{14}\text{H}_{20}\text{O}_8)\cdot n\text{H}_2\text{O}$), see Fig. 12. Then, the resultant breakthrough H_2S experiment led to a gas uptake of 15.95 mmol g^{-1} , representing one of the highest H_2S uptakes reported and even more interesting since the BET surface area of SU-101 ($412\text{ m}^2\text{ g}^{-1}$) is considerably lower than top H_2S capture materials (e.g., MIL-53(Al)-TDC; BET = $1150\text{ m}^2\text{ g}^{-1}$). On a second H_2S adsorption cycle, the initially observed capacity and surface area were reduced to only 0.2 mmol g^{-1} and $15\text{ m}^2\text{ g}^{-1}$, respectively, even though the crystallinity was retained as confirmed by PXRD. Raman spectra of SU-101 (before and after H_2S adsorption) confirmed the presence of low-order polysulphides ($n = 2$), S_4^{2-} . Thus, the electrochemical potential of SU-101 should be similar to the one for MFM-300(Sc) (2.29 V),⁷⁰ considering that S_4^{2-} species were only found for MFM-300(Sc). Future investigations are anticipated to continue in order to verify the electrochemical potential of SU-101 and any possible reversible metal–ligand bonding upon the adsorption of H_2S .

A variety of MOFs, as well as classic materials such as zeolites, activated carbon and metal oxides, have been studied for the adsorption of these toxic gases, where some of these are summarized in Table 3. Noteworthy, the comparison of materials for H_2S capture is sometimes challenging due to the difference in the experimental conditions (e.g., flue gas concentration).

3.2 Bio-compatible MOFs for H_2S detection and controlled delivery

Although H_2S is catalogued as highly toxic, paradoxically, it can also be crucial as an endogenous biological mediator.⁸²

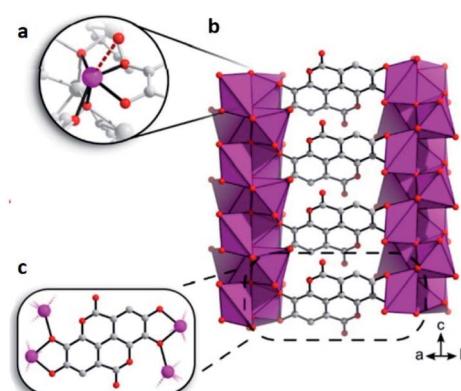


Fig. 12 (a) The coordination environment around Bi^{3+} . The bond to a coordinated water molecule is represented as a dashed line. (b and c) Chelation of ellagate ligand toward the bismuth oxo rods. Reprinted (adapted) with permission from ref. 74. Copyright (2020) American Chemical Society.



Table 3 Summary of the adsorption capacity of H₂S for selected metal–organic frameworks and benchmark materials

MOF	BET Surface area [m ² g ⁻¹]	H ₂ S adsorption [mmol g ⁻¹]	Ref.
MIL-47(V)	930	14.6	67
MIL-53(Cr)	1946	12.0	67
MIL-53(Al)-BDC	1103	11.7	67
Ni-CPO-27	1547	12.0	85
Mg-CUK-1	604	3.1 ^a	68
MIL-53(Al)-TDC	1150	18.5	69
MIL-100(Fe)	2000	16.7	67
MIL-101(Cr)	2916	30.1	34
MIL-101(Cr)-4F(2%)	2176	36.9	34
MFPM-300(Sc)	1350	16.5	70
Cu-BTC	1590	1.1	63
SU-101	412	15.95	74
PAF-302	5600	51.94	75
COF-102	3621	35.57 ^b	75
HKUST-1 ^c	—	2.7 ^d	60
10Mn–45Zn–45Ti–O ^e	—	0.0005	76
4A molecular sieve zeolite	49.5	0.0002 ^f	77
13X zeolite	440	0.007 ^g	78
SBA-15	950	1.98 ^h	79
PEI(50)/SBA-15(MBS-2)	80	0.0002 ^h	80
Wood-based AC	1400	0.008	81

^a 15% vol H₂S. ^b GCMC simulations. ^c Graphite-oxide composite. ^d 1000 ppm H₂S in moist air. ^e H₂S/H₂/CO₂/H₂O (600 ppm : 25% : 7.5% : 1%). ^f 10% vol H₂S. ^g 80 ppmv H₂S. ^h 4000 ppmv of H₂S and 20 vol% of H₂ in N₂.

Endogenous pathways for H₂S have been found to have wide extending activities *in vivo*, such as in the cardiovascular system.⁸³ For example, H₂S executes a crucial activity in the regulation of blood pressure, neurotransmission, anti-inflammatory mechanism, anti-oxidation, angiogenesis and apoptosis.⁸⁴ H₂S can be biosynthesised by enzymatic reactions *via* a sequence of endogenous processes,⁸⁵ and in the biological medium, the H₂S concentration can fluctuate from nano-to micromolar levels.⁸⁶ However, at higher concentrations of H₂S in the bloodstream can cause severe physiological disorders such as diabetes, Alzheimer's disease, cirrhosis, different types of cancer and Down's syndrome.⁸⁷

Therefore, in order to recognise the specific role of H₂S in these processes, the detection (spatial and temporal information) of H₂S levels in living cells and organisms is crucial. However, due to the high reactivity, volatility and diffusible properties of H₂S, traditional detection technologies (chromatography, high performance liquid chromatography, electrochemical analysis and colorimetric)⁸⁸ are unsuitable for the identification of H₂S in living cells because they result in destruction of the cells and tissues and difficulties associated to real-time detection.⁸⁹ Fluorometric detection techniques have emerged as highly significant alternatives since their non-destructive characteristics, fast response, high selectivity and sensitivity, fast response, high spatial sampling capability, real-time monitoring and easy sample preparation.⁹⁰ Turn-on type fluorescence probes for H₂S detection have been recently investigated and these are commonly based on the reactivity characteristics of H₂S (e.g., copper sulphide precipitation, dual nucleophilic addition, H₂S-mediated hydroxyl amide and nitro/azide reduction).⁹¹ Such characteristic reactivity of H₂S can

efficiently distinguish it from other biological species; however, many of these detectors do not fulfil basic requirements such as rapid detection and selectivity. Thus, the investigation of new H₂S probes is still very attractive due of the complexity of different molecular events involved in signalling transduction, and other complications related to the probes. Therefore, biocompatible MOFs have been postulated as promising probes for the detection of H₂S.

The first concept of a MOF material in *in vitro* H₂S delivery was demonstrated by Morris and co-workers who used CPO-27 for this purpose.⁹² H₂S can be released by exposing the material to H₂O since CPO-27 is an open metal site system thus, the water molecule replaces the hydrogen sulphide molecule at the metal site. The authors analysed the H₂S delivery in two isostructural materials (Zn-CPO-27 and Ni-CPO-27) finding that the Ni²⁺ material exhibited higher delivery capacity than the Zn²⁺ material: after 30 min 1.8 mmol g⁻¹ of H₂S was released from Ni-CPO and only 0.5 mmol g⁻¹ from the Zn-CPO. After 1 h, the H₂S release for both materials was completed. The amount of H₂S released by each material corresponds to approximately 30% of the chemisorbed gas. Later, the release of H₂S (stored in Zn-CPO-27 and Ni-CPO-27) was investigated under physiological conditions (endothelium-intact ring of pig coronary artery), finding a substantial vasodilatory action (relaxation) with only a short induction period of approximately 5 min.⁹² These results represented a very significant progress in the field of MOFs and biological applications, since such H₂S release mechanisms offered new protocols to better understand the effects of H₂S on this particular vascular system.

Later, the first fluorescence probe example based on a MOF for the detection of H₂S was presented by Wang and co-



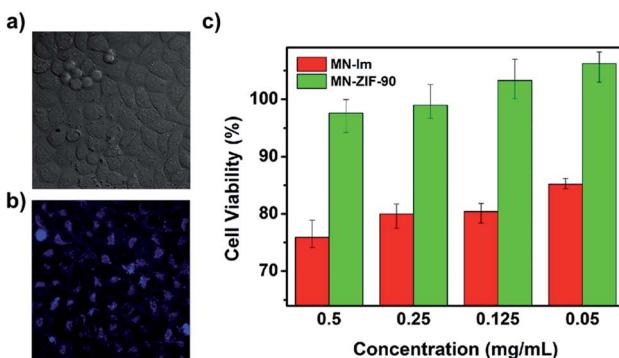


Fig. 13 (a, b) Bright field image and fluorescence image after treating the HeLa cells with MN-ZIF-90 (0.05 mg mL^{-1}) after 48 h of incubation. (c) *In vitro* HeLa cell viabilities after 48 h of incubation with MN-IM and MN-ZIF-90 at different concentrations. Reprinted with permission from ref. 93. Copyright (2014) Springer Nature.

workers.⁹³ Through a post-synthetic modification of ZIF-90 with malononitrile ($\text{N}\equiv\text{C}-\text{CH}_2-\text{C}\equiv\text{N}$), they obtained MN-ZIF-90 which undergoes a specific reaction with H_2S , achieving an enhancement of photoluminescence, constituting the base for the detection of H_2S (see Fig. 13). Then, the malononitrile moieties conjugated to the host (ZIF-90) through double bonds aided as quenchers of the host fluorescence through intramolecular photoinduced electron transfer. The detection mechanism was based on the α, β -unsaturated bond of the malononitrile which is susceptible to thiol compounds, leading to the breaking of the double bond and thus, the fluorescence of MN-ZIF-90 was recovered once H_2S was introduced into the system. In addition, the probe exhibited favourable biocompatibility.⁹³

Taking a different approach but keeping the fundamental idea of introducing a reactive site for H_2S , Tang *et al.*,⁹⁴ demonstrated the use of a porphyrin-based MOF for the fluorescent detection of H_2S . By introducing reactive Cu^{2+} metal centres into a Al^{3+} MOF material $\{\text{CuL}[\text{AlOH}]_2\}_n$ (H_6L = mesotetrakis(4-carboxylphenyl) porphyrin) the detection of H_2S was selectively followed by fluorescence under physiological pH (see Fig. 14). Additionally, they successfully demonstrated the capability of the probe to detect exogenous H_2S in living cells, while the probe showed low toxicity and high biocompatibility.⁹⁴

Later, Biswas and co-workers demonstrated, on a dinitro-functionalised UiO-66 ($\text{UiO-66-(NO}_2\text{)}_2$), fluorescence turn-on behaviour towards H_2S in simulated biological medium (HEPES buffer, pH = 7.4).⁹⁵ Remarkably, $\text{UiO-66-(NO}_2\text{)}_2$ exhibited highly sensitive fluorometric H_2S sensing while also showing a visually detectable colorimetric change to H_2S in daylight (see Fig. 15). In addition, the high selectivity of this functionalised MOF material to H_2S was preserved even when several other biological species were present in the detecting medium. Finally, fluorescence microscopy studies on J774A.1 cells demonstrated the effectiveness of $\text{UiO-66-(NO}_2\text{)}_2$ for H_2S imaging in living cells, detection of H_2S in human blood plasma (HBP) and monitoring of the sulphide concentration in real water samples.⁹⁵ These results emphasised the biocompatibility

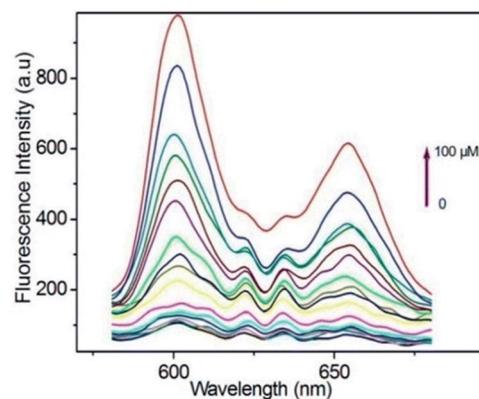


Fig. 14 Fluorescence spectra of porphyrin ligand ($10 \mu\text{M}$) in BBS buffer ($\text{pH} = 7.40, 20 \text{ mM}$) after titration with HS^- from 0 – $100 \mu\text{M}$, $\lambda_{\text{ex}} = 419$. Reprinted with permission from ref. 94. Copyright (2014) American Chemical Society.

of $\text{UiO-66-(NO}_2\text{)}_2$. The reaction mechanism of $\text{UiO-66-(NO}_2\text{)}_2$ and H_2S is caused by the reduction of the nitro groups to electron donating amine groups causing an increase in the fluorescence intensity that can be registered.

$\text{Eu}^{3+}/\text{Ag}^+@\text{UiO-66-(COOH)}_2$ (EAUC) composites were reported by Li and Qian as biomarkers for the for potential diagnosis of asthma.⁹⁶ The decrease of the H_2S production in the lung has been identified as an early detection biomarker for asthma. Thus, fluorescent experiments showed that EAUC exhibited high selectivity and sensitivity with a limit of a real-time H_2S detection of $23.53 \mu\text{M}$. The detection of H_2S by EAUC was based on the introduction of active metal centre (Ag^+ , H_2S -responding site) to the $\text{Eu}^{3+}@\text{UiO-66-(COOH)}_2$ as the lanthanide-luminescence sensitisier.⁹⁶ MTT assay and cell viability analysis in PC12 cells demonstrated relatively low cytotoxicity for EAUC and biocompatibility. Finally, the determination of H_2S was tested in diluted fetal bovine and human serum samples demonstrating that EAUC can detect H_2S in real biological samples. Zhang and co-workers⁹⁷ covalently modified (*via* click chemistry) PCN-58 with target-responsive two-photon (TP) organic moieties to fluorescently detect H_2S . These TP-MOF probes showed good photostability, high selectivity, minor cytotoxicity, and excellent H_2S sensing performance in live cells. These modified PCN-58 materials also showed intracellular sensing and depth imaging capabilities (penetration depths up to $130 \mu\text{m}$).⁹⁷



Fig. 15 Colorimetric detection behaviour of the HEPES suspension of $\text{UiO-66-(NO}_2\text{)}_2$ with increasing concentration of Na_2S in daylight. Reprinted with permission from ref. 95. Published (2018) by The Royal Society of Chemistry.



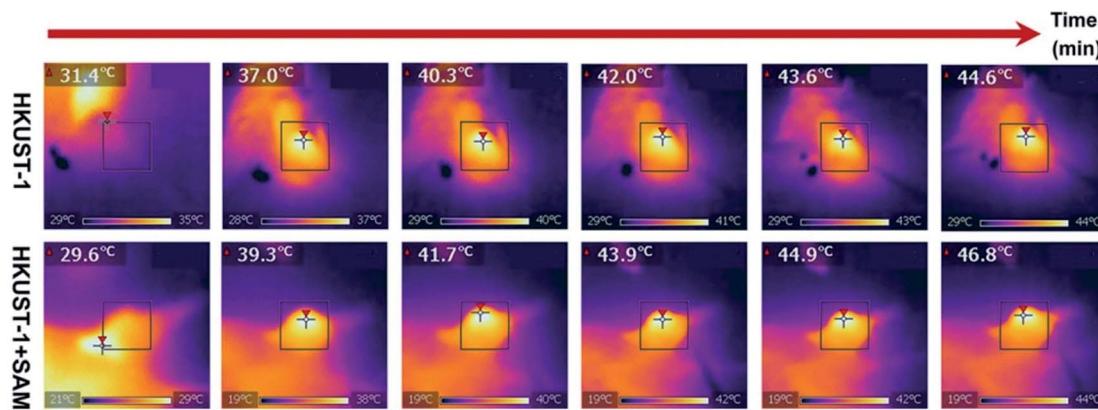


Fig. 16 *In vivo* antitumour efficacy of HKUST-1. Photothermal images of the tumour site after 4 h of intratumoural injection of HKUST-1 and HKUST-1 + SAM exposure to laser irradiation (808 nm, 1 W cm^{-2}) at the 2 min interval. Reprinted with permission from ref. 98. Copyright (2020) American Chemical Society.

Up to this point we have presented remarkable examples of MOF materials for the detection of H_2S (biomarkers). All of these have been constructed with the fundamental principle of incorporating into these MOFs “the right chemical functionality” which can react with H_2S . Remarkably, Wang and Xie⁹⁸ took a step forward and very recently showed how an endogenous MOF-biomarker-triggered “turn-on” strategy was capable to produce therapeutic agents *in situ*, as a promising example in nanomedicine for the precise treatment of colon cancer. HKUST-1 was used as an endogenous H_2S -activated copper MOF which demonstrated to synergistically mediate H_2S -activated near-infrared photothermal therapy and chemodynamic therapy. As a proof of principle, the protocol worked as follows: in normal tissues, the photoactivity of as-synthesised HKUST-1 nanoparticles was in the “OFF” state, and no obvious adsorption within the NIR region was observed. On the other hand, when HKUST-1 nanoparticles with high overexpression of H_2S reached the colon tumour tissues, these nanoparticles were activated to the “ON” state by reacting with endogenous H_2S to produce *in situ* photoactive copper sulphide with stronger NIR absorption, which is viable for the subsequent photothermal therapy and corresponding thermal imaging (see Fig. 16). In addition, besides the *in situ* formation of the photothermal agent, HKUST-1 nanoparticles also showed a horseradish peroxidase (HRP)-mimicking activity to efficiently convert overexpressed H_2O_2 within cancer cells into more toxic $\cdot\text{OH}$ radicals for chemodynamic therapy. Thus, this promising H_2S -triggered “turn-on” strategy based on the endogenous biomarker from the tumour microenvironment can provide a precise diagnosis, marginal invasion, and possibly clinical translation.⁹⁸

4. Ammonia

NH_3 is one of the most important chemicals in the world since it is an irreplaceable feedstock for global agriculture and industry (e.g., fertilizers, drug production, heat pumps and fuel cells).⁹⁹ Although fundamental to our diet supply and economy, NH_3 is a highly toxic (even in small concentrations) and corrosive gas (*i.e.*, difficult to handle and store). Thus, considerably large

emissions of NH_3 from livestock breeding, industrial production, fertiliser application and public transportation can have serious consequences on the environment (highly unfavourable to air quality and aquatic life), and human health.¹⁰⁰ In addition, NH_3 reacts with NO_x and SO_2 to form $\text{PM}_{2.5}$, seriously contributing to the increase of air pollution.

NH_3 is a colourless gas with a characteristic pungent smell which is typically liquefied and confined in metal tanks in order to be efficiently stored and transported.¹⁰¹ This implies high pressures (approximately 18 bar) and constant corrosion of pipelines and containers. Thus, new sorbent technologies capable of efficiently removing or storing NH_3 are highly desirable for air remediation and separation of NH_3 from N_2 and H_2 since NH_3 is considered an interesting energy intermediate.¹⁰² It is clear that capture and recycling of NH_3 implies a dual connotation in the fields of environment and energy. Classic materials such as activated carbons, zeolites, silicas and even organic polymers have been investigated for the capture and separation of NH_3 showing low uptakes, poor selectivity and, in some cases, irreversible storages.^{103,104} MOFs have been postulated as a promising option for NH_3 capture, due to the access to a wide range of chemical functionality (*e.g.*, Lewis or Brønsted acid sites which provide higher affinity to the basic NH_3 molecule).¹⁰⁴ However, the main problem that most MOF materials exhibit poor chemical stability towards NH_3 .¹⁰⁵ The following section describes chemically stable MOFs with interesting NH_3 capture performances.

4.1 MOFs for NH_3 capture

The NH_3 molecule is a Lewis and Brønsted base which is the key to its adsorption on the porous surface of chemically stable MOFs. Thus, the incorporation of acidic functional groups within the MOFs is required to achieve relevant NH_3 capture results. These functionalisations can be organised in three main types: (i) coordinatively unsaturated metal sites (open metal sites), (ii) $\mu\text{-OH}$ groups and (iii) defective sites.

(i) Coordinatively unsaturated metal sites. The majority of MOF materials evaluated for the capture of NH_3 have demonstrated to be chemically unstable.¹⁰⁶ However, really interesting



examples have been reported by Dincă and co-workers.¹⁰⁷ They reported a series of new mesoporous MOFs constructed from extended bisbenzenetriazolate ligands and coordinatively unsaturated metal sites (Mn^{2+} , Co^{2+} , and Ni^{2+}), which showed high and reversible NH_3 uptakes (15.47, 12.00, and 12.02 mmol g⁻¹, respectively).¹⁰⁷ During desorption at different temperatures, all three materials showed pronounced hysteresis which emphasised the strong interaction between the open metal sites and NH_3 . Such bound NH_3 molecules can be fully removed upon heating the materials up to 200 °C under dynamic vacuum. Interestingly, none of the three materials showed a decrease in the NH_3 uptake capacity upon cycling. Thus, the high chemical stability of azolate MOFs with open metal sites, towards NH_3 postulate them as promising alternatives in environmental applications such as the capture of corrosive gases from power plant flue gas streams. This chemical stability can be attributed to the strength of the N atom coordinated to divalent metal centres as described previously.²¹ In this case, the use of linear bistriazolate ligand, in combination with late transition metals such as Ni^{2+} , increases the heterolytic metal-ligand bond strength due to its higher basicity (compared to a carboxylate ligand) which allows better MOF stability to acidic gas exposure. This trend is also explained by Pearson's acid and base principle due to Mn^{2+} , Cu^{2+} and Ni^{2+} are considered soft cations due to their charge/radius ratio, and nitrogen-based linkers are softer than carboxylate linkers.³⁵

Later, the Dincă's group reported another series of microporous triazolate MOFs (containing open metal sites; Cu^{2+} , Co^{2+} , and Ni^{2+} , see Fig. 17), that exhibited remarkable static and dynamic NH_3 capacities (up to 19.79 mmol g⁻¹, at 1 bar and 298 K).¹⁰⁸ These isoreticular analogues to the bisbenzenetriazolate examples,¹⁰⁷ are constructed from smaller triazolate ligands and therefore, they exhibited smaller pore windows (see Fig. 17). Interestingly, these microporous MOFs showed higher NH_3 captures than their mesoporous analogues.¹⁰⁷ This remarkable property was not only accounted by the increase in the density

of open metal sites, in addition cooperative proximity effects resulted very important. NH_3 breakthrough experiments on these microporous materials showed the potential applicability for both personal protection and gas separations (NH_3 capacities of 8.56 mmol g⁻¹). Finally, once again, the superior chemical stability of these triazolate MOFs arises, as previously described, from the coordination of the N atom to metal centres (*vide supra*).

Very recently, Hong and co-workers recently presented the NH_3 adsorption properties of $M_2(dobpdc)$ MOFs ($M = Mg^{2+}$, Mn^{2+} , Co^{2+} , Ni^{2+} , and Zn^{2+} ; $dobpdc^{4-} = 4,4$ -dioxidobiphenyl-3,3-dicarboxylate), which after activation exhibited open metal sites (see Fig. 18).¹⁰⁹ The NH_3 uptake of $Mg_2(dobpdc)$ at 298 K and 1 bar was 23.9 mmol g⁻¹ at 1 bar and 8.25 mmol g⁻¹ at 570 ppm, representing the NH_3 record high capacities at both pressures among existing porous adsorbents.¹⁰⁹ $Mg_2(dobpdc)$ remarkably demonstrated chemical stability to not only dry NH_3 , but also to wet NH_3 ($NH_3 \cdot H_2O$ vapour). Three consecutive NH_3 sorption isotherms (activation included only application of vacuum) demonstrated the cyclability of $Mg_2(dobpdc)$. To evaluate the affinity of the NH_3 molecule to the open Mg^{2+} sites in $Mg_2(dobpdc)$, NH_3 -TPD curves demonstrated (based on the Redhead analysis) the activation energy for NH_3 desorption of 146.7 kJ mol⁻¹. The NH_3 adsorption mechanism of $Mg_2(dobpdc)$ was investigated by *in situ* FTIR experiments, indicating that NH_3 was preferentially adsorbed, by coordination bonds, to the open metal sites of $Mg_2(dobpdc)$. As the MOF sample was exposed to atmosphere the adsorbed NH_3 in $Mg_2(dobpdc)$ was converted to NH_4^+ , due to its reaction with atmospheric moisture. Dynamic NH_3 breakthrough curves under dry conditions (0% relative humidity (RH) and 298 K) showed a NH_3 capacity of 8.37 mmol g⁻¹ for $Mg_2(dobpdc)$. When performing these NH_3 breakthrough experiments under 80% RH conditions, the NH_3 capacity of $Mg_2(dobpdc)$ decreased to 6.14 mmol g⁻¹, due to the competitive adsorption between NH_3 and H_2O . Additionally, this NH_3 capacity remained

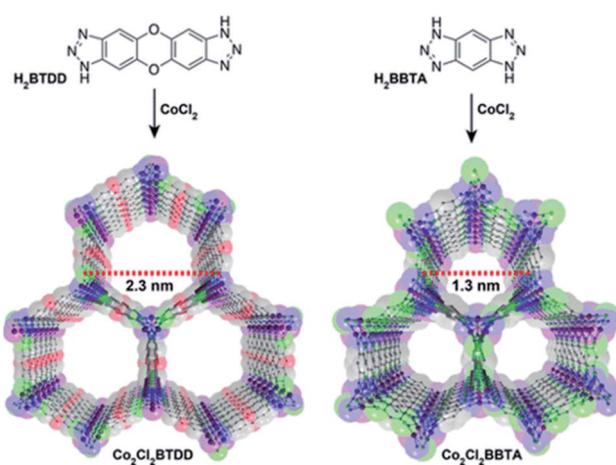


Fig. 17 Structure of Co_2Cl_2 BTDD (left) and Co_2Cl_2 BBTA (right). C, gray; O, red; N, blue; Cl, green; Co, purple. Hydrogen atoms have been omitted for clarity. Reprinted with permission from ref. 108. Copyright (2018) American Chemical Society.

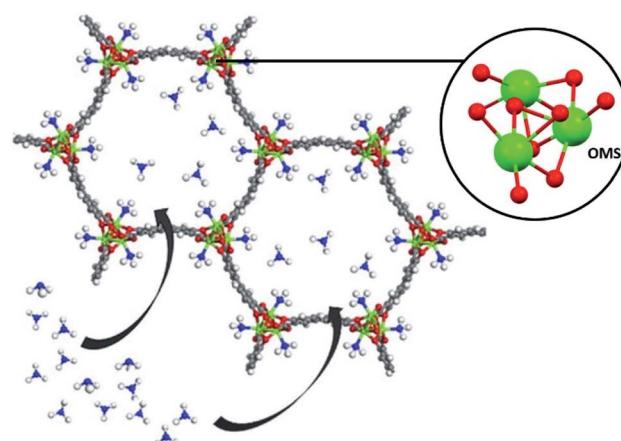


Fig. 18 Atomic structure of $M_2(dobpdc)$ complexes with one-dimensional hexagonal channels. It is possible to observe that each coordinatively unsaturated metal site points towards the channel. Reprinted (adapted) with permission from ref. 109. Copyright (2020) John Wiley & Sons.

constant over five breakthrough wet cycles.¹⁰⁹ Thus, Mg₂(-dobpdc) is an exceptional MOF material that can be easily synthesised (*i.e.*, microwave-assisted), highly recyclable (even after exposure to a humid NH₃) without structural degradation or NH₃ capacity loss and holds the record for the highest NH₃ uptake at room temperature and atmospheric pressure.

All of these relevant properties are possible due to the exceptional chemical stability of this Mg²⁺-based MOF material. But why is it so stable towards NH₃? The authors proposed that such chemical stability was due to the higher affinity of Mg²⁺ to oxygen atoms than nitrogen atoms, as confirmed by van der Waals (vdW)-corrected density functional theory (DFT) calculations. They further investigated the origin of such remarkable chemical stability finding that the oxygen adjacent to the carboxylate participates in the coordination of the ligand to the Mg²⁺ cation acting as a tetratopic linker, similar to a chelate effect, increasing the stability of the cluster. This strategy is widely used to prevent pore collapse due to the activation energy barrier to ligand reorganisation or removal, which is increased by the higher linker connectivity.²¹

(ii) μ -OH groups. The MFM-300(M) family (see the section of SO₂) demonstrated remarkable NH₃ adsorption properties with high uptakes and attractive cyclabilities. First, MFM-300(Al), reported by Yang and Schröder, showed a total NH₃ capture of 15.7 mmol g⁻¹ at 273 K and 1.0 bar.¹¹⁰ Although the NH₃ capture was not particularly high (even at a lower temperature than typically 298 K), the identification of the preferential NH₃ adsorption sites was achieved by sophisticated *in situ* neutron powder diffraction (NPD) and synchrotron FTIR micro-spectroscopy. Thus, a structural analysis *via* Rietveld refinement of NPD data for 1.5 ND₃ (deuterated ammonia)/Al-loaded MFM-300(Al) identified three distinct binding sites (I, II and III) in [Al₂(OH)₂(L)](ND₃)₃, (H₄L = 1,1'-biphenyl-3,3',5,5'-tetracarboxylic acid = C₁₆O₈H₆) (see Fig. 19). This cooperative network of ND₃ molecules spreads down the length of the 1D channel, fixed in place by site I. Bond distances for sites II···III and a slightly lengthened site I···II are similar to a characteristic intermolecular bond between ND₃ molecules in the solid state at very low temperature (*i.e.*, 2 K (N···D = 2.357(2) Å)), while the

bond between the framework μ_2 -OH and site I is considerably shorter. Upon increasing the ND₃ loading (from 0.5 ND₃/Al to 1.5 ND₃/Al), a general shortening of the framework μ_2 -OH···site I and sites I···II and an increase in the site II···III was observed.

Refinement of the NPD data for ND₃-loaded MFM-300(Al) demonstrated that on increasing the loading and thus the ND₃/ μ_2 -OH ratio, the hydrogen atoms of the hydroxo functional groups experienced a reversible site exchange with the deuterium from the guest ND₃ molecules residing at site I in the pore to resulting in the formation of μ_2 -OD moieties.¹¹⁰ Interestingly, this H-D reversible exchange did not lead to any detectable structural degradation of the long-range order of the MOF material. Thus, the adsorption of ND₃ in MFM-300(Al) showed a new type of adsorption mechanism where adsorbent and adsorbate experienced a rapid site-exchange *via* reversible formation and cleavage of O-H and O-D chemical bonds, in other words a pseudo-chemisorption binding mechanism.

Very recently, Yang and Schröder expanded the investigation on more MFM-300(M) (M = Fe²⁺, V⁴⁺, Cr³⁺, In³⁺) materials and NH₃.¹¹¹ MFM-300(M) (M = Fe²⁺, V⁴⁺, Cr³⁺) demonstrated fully reversible NH₃ capture for over 20 adsorption–desorption cycles, under pressure-swing conditions, reaching capacities of 16.1, 15.6, and 14.0 mmol g⁻¹, respectively, at 273 K and 1 bar. In the case of MFM-300(In), a significant loss of NH₃ capture capacity over repeated NH₃ cycles was shown which corroborated its chemical instability. On the other hand, MFM-300(V⁴⁺) exhibited the highest NH₃ uptake (17.3 mmol g⁻¹) among the MFM-300(M) family. Interestingly, the NH₃ desorption phase for MFM-300(V⁴⁺), with pore dimensions of approximately 6.7 × 6.7 Å², showed a hysteresis loop. Although this could indicate a characteristic capillary NH₃ condensation (*e.g.*, in mesopores and/or due to a broad distribution of pore size and shape), the authors demonstrated, taking into account the pore dimensions of this MOF, a specific and potentially strong host–guest charge transfer upon the adsorption of NH₃. In addition, MFM-300(V⁴⁺) showed an increase of both NH₃ capacity and residue within the first 18 cycles, which was not observed for the rest of the MFM-300(M) materials. Such residual amount of NH₃ left within MFM-300(V⁴⁺) upon regeneration (pressure-swing) gradually increased from 8 to 20% along these cycles, indicating an increase of strongly bound NH-derived species in MFM-300(V⁴⁺). This NH-derived residue, which was not desorbed by only reducing the pressure, was completely removed by increasing the temperature under dynamic vacuum, although some structural degradation of MFM-300(V⁴⁺) was observed. The chemical stability of these materials to NH₃ under humid conditions was investigated by PXRD, confirming the retention of their crystallinity for both MFM-300(M) (M = Cr³⁺, V³⁺), while MFM-300(M) (M = Fe³⁺, V⁴⁺) showed some structural degradation.¹¹¹

Neutron powder diffraction data for ND₃-loaded MFM-300(M) combined with Rietveld refinements, showed the preferential binding sites for ND₃ (Fig. 20). Two binding sites for ND₃ were clearly identified for MFM-300(M) (M = In³⁺, V³⁺), while MFM-300(Fe) has an additional binding site for ND₃. Site I showed the highest occupancy, with hydrogen bonding between the hydroxo functional group and the ND₃ molecule (μ_2 -OH···

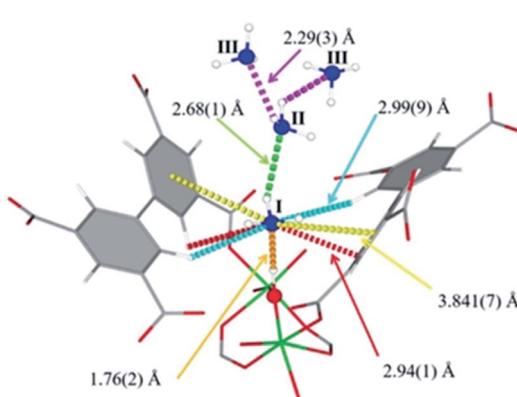


Fig. 19 Binding sites for ND₃ molecules adsorbed within MFM-300(Al) determined by *in situ* NPD studies. Reprinted with permission from ref. 110. Copyright (2018) John Wiley & Sons.



ND_3) (1.411–1.978 Å), complemented by additional hydrogen bonding from the organic ligand ($\text{H}_{\text{aromatic}} \cdots \text{ND}_3 = 2.738$ –3.174 Å; $\text{ND} \cdots \text{O}_{\text{ligand}} = 3.078$ –3.179 Å) and electrostatic interactions ($\text{ND}_3 \cdots \text{aromatic rings} = 2.946$ –3.132 Å) (Fig. 20). Similar to MFM-300(Al),¹¹⁰ hydrogen/deuterium site exchange was also observed between the adsorbed ND_3 and the $\mu_2\text{-OH}$ group on for MFM-300(M) ($\text{M} = \text{In}^{3+}$, Fe^{3+} , V^{3+}). Site II is located toward the centre of the pore (fixed by hydrogen-bonding interactions $\text{ND}_3 \cdots \text{O}_{\text{ligand}} = 2.284$ –3.065 Å). Site III in MFM-300(Fe) is constructed by electrostatic interactions ($\text{ND}_3 \cdots \text{aromatic rings} = 3.146$ Å). Additionally, in MFM-300(Fe), intermolecular hydrogen bonds between ND_3 molecules (2.327 Å) were identified, propagating along the 1D channel to form a cooperative $\{\text{ND}_3\}_\infty$ network.

The preferential ND_3 binding sites for MFM-300(V^{4+}) were identified and these were different to the previous examples. MFM-300(V^{4+}) is not constructed with $\mu_2\text{-OH}$ groups. It shows bridging oxo centres to balance the charge of the oxidised V centre. Thus, ND_3 molecules are located in the centre of the pore. This particular arrangement implicates a very interesting situation: chemical reaction between the ND_3 molecules: a N_2D_4 molecule was formed and located at site II with a ND_3 molecule at site I (being partially protonated to ND_4^+). Then, both sites are stabilised *via* hydrogen bonding ($\text{ND}_3 \cdots \text{O}_{\text{ligand}} = 2.529$ –3.092 Å), and the amount of N_2D_4 at site II was estimated as 0.5 N_2H_4 per V centre.

The formation of N_2H_4 was explained in terms of the redox activity of MFM-300(V^{4+}) which promotes a host–guest charge transfer between the material and the adsorbed NH_3 molecules, promoting the reduction of the V^{4+} centres and oxidation of NH_3 to N_2H_4 .¹¹¹ These host–guest charge transfer results were

experimentally determined by EPR and corroborated by bond valence sum (BVS) calculations, finding as well that the charge transfer between adsorbed NH_3 molecules and the V^{4+} centre can only occur when the loading of NH_3 is sufficiently high so that a predominant occupancy of the N site, which is located close to the metal chain, is reached to initiate the redox reaction. Finally, such mechanisms could also be investigated by the approach of metal–ligand dynamics as recently proposed by Brozek and co-workers,⁷³ to fully understand the interaction of NH_3 and the hydroxo functional group in the MFM-300(M) family.

(iii) Defective sites. As previously described in the SO_2 capture section (*vide supra*), crystal irregularities can provide very particular physical and chemical properties. Then, defects in MOF materials also demonstrated remarkable NH_3 adsorption properties. Very recently, Wu and Tsang presented a comprehensive study on the responsive adsorption behaviours of defect-rich Zr-based MOFs upon the progressive incorporation of NH_3 .¹¹² They investigated UiO-67 and UiO-bpydc containing 4,4'-biphenyl dicarboxylate and 4,4'-(2,2'-bipyridine) dicarboxylate ligands, that despite their structural similarities demonstrated a drastic difference in the NH_3 adsorption properties when the biphenyl groups in the organic ligand were replaced by the bipyridine moieties. Such replacement can confer flexibility to the framework in the context of mainly ligand “flipping” but without significant pore volume alteration.

Then, defective UiO-67 (non-monodisperse pore structure created by missing ligand defects) was synthesised thanks to the high connectivity of the Zr_6 oxoclusters, which help to retain the overall crystal structures even with some ligands are missing. UiO-67 involves uniform trigonal windows with a diameter of 11 Å that lead to interconnected tetrahedral and octahedral pores inside the structures. In the presence of missing ligand defects, the trigonal windows surrounding the defects in UiO-bpydc are fused into lozenge windows with a dynamic size larger than 14 Å. NH_3 adsorption–desorption isotherms at 298 K and up to 1 bar were performed for both materials: UiO-67 and UiO-bpydc. In the case of UiO-67, the adsorption isotherm showed “step-like shape” with two events (see Fig. 21, bottom). This characteristic isotherm shape is caused in MOFs by a gate-opening phenomenon due to the interaction between guest NH_3 molecules and pore walls, or a pore filling process.¹¹³ Thus, at the beginning of the adsorption, the rapid and sudden increase in NH_3 adsorption at approximately 30 mbar (1.70 mmol g^{−1}) (position I) suggests the presence of strong adsorption sites inside the framework, *e.g.*, strong interaction of the NH_3 molecules with the hydroxo ($\mu_2\text{-OH}$) functional groups. Later, the NH_3 uptake increased from 2.40 to 4.40 mmol g^{−1} at 250 mbar (position II) and from 5.60 to 8.40 mmol g^{−1} at 650 mbar (position III, Fig. 21). The desorption phase exhibited large, opened hysteresis, indicating that the NH_3 molecules strongly interact with the pore walls of the material.¹¹² These NH_3 molecules were fully desorbed by heating up to 423 K for 1 h under dynamic vacuum. Later, the chemical stability of UiO-67 towards NH_3 was investigated by conducting three NH_3 adsorption–desorption cycles and finally exposing the material to NH_3 vapour for 1 week. PXRD experiments corroborated the

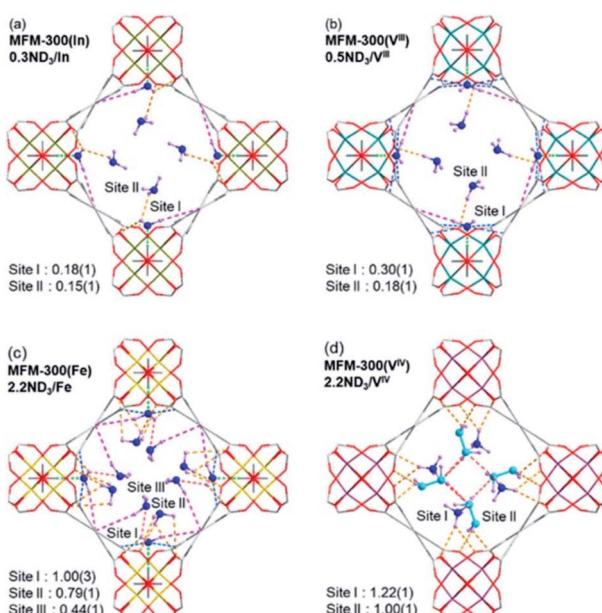


Fig. 20 Preferential ND_3 binding sites in MFM-300(M) family ($\text{M} = \text{Cr}$, Fe , V^{3+} , V^{4+}) determined by NPD at 10 K. Views along c axis. Reprinted with permission from ref. 111. Copyright (2021) American Chemical Society.



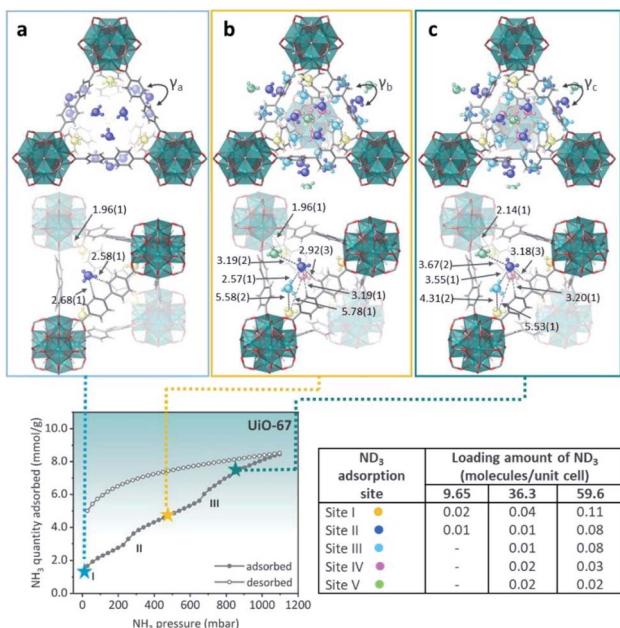


Fig. 21 (Top) Refined ND₃ positions in the UiO-67 structure at ND₃ loadings of (a) 1.23, (b) 4.63, and (c) 7.60 mmol g⁻¹ obtained from NPD data at 300 K. ND₃ molecules at different binding sites: site I: yellow, site II: blue, site III: light blue, site IV: pink, and site V: green. (Bottom) NH₃ sorption isotherm at 298 K and up to 1100 mbar; adsorption: closed circles; desorption: open circles reprinted with permission from ref. 112. Copyright (2021) American Chemical Society.

stability of UiO-67 to dry and humid NH₃. Then, in order to identify the preferential NH₃ adsorption sites in UiO-67, *in situ* high-resolution neutron powder diffraction (NPD) and synchrotron powder X-ray diffraction (SPXRD) experiments were carried out. The refined structures of UiO-67 at different ND₃ pressures, revealed five such adsorption preferential sites (Fig. 21).¹¹² At site I, the ND₃ was found close to the μ_3 -OH with a OH \cdots N_I distance of 1.96(1) Å, and O \cdots N_I distance of 2.80(1) Å (see Fig. 21). This distance suggested the formation of a relatively strong H-bonding interaction between the μ_3 -OH and ND₃, similar to the one reported by Yang and Schröder in the MFM-300(M) family.^{110,111} ND₃ molecules at site II were situated close to the walls of the trigonal windows suggesting interactions with the MOF organic ligands (H_{ligand} \cdots N_{II} = 2.58(1) and 2.68(1) Å), Fig. 21. On increasing the amount of ND₃, these molecules were located near to the tetragonal pore (sites III and IV). At site III, the ND₃ molecule filled the shallow pore positions with N_{II} \cdots N_{III} bond distances among these ND₃ sites in the range of 2.33(1) and 2.57(1) Å, creating a H-bonded network of these ND₃ molecules (Fig. 21).

In the case of UiO-bpydc, the NH_3 adsorption isotherm exhibited an uptake of 8.4 mmol g^{-1} at 298 K and 1 bar (see Fig. 22). Conversely to UiO-67, UiO-bpydc showed only one large and sharper transition step (position II') for the NH_3 adsorption phase (Fig. 22). Then, *in situ* NPD experiments and followed by Rietveld refinements corroborated the existence of missing ligand defects. Similarly to UiO-67, the preferential NH_3 adsorption sites within UiO-bpydc were also identified by *in situ*

NPD experiments (replacing NH_3 by ND_3). Thus, two independent preferential sites of ND_3 were found close to the $\mu_3\text{-OH}$ functional group (site I') and to the organic ligands (site II', bipyridine ligand), Fig. 22. This strongly suggested that the adsorption of ND_3 mainly takes place at site II'. For site I', the $\text{OH}\cdots\text{N}_\text{I}$ bond distance was estimated to be $2.10(2)$ Å, which is similar to that found in the isostructural UiO-67 . Interestingly, the ND_3 -bipyridine ligand interaction was estimated to be stronger than the ND_3 -biphenyl ligand interaction with longer distances ($\text{H}_\text{ligand}\cdots\text{N}_\text{II} = 2.68(1)$, $3.19(1)$, and $3.20(1)$ Å, respectively) (see Fig. 22). Increasing the dosing of ND_3 molecules, increased the H-bonding between the ND_3 molecules giving rise to a network of six ND_3 molecules with the three bipyridine ligand located around the trigonal window. This particular situation demonstrated the gate-opening/closing behaviour *via* ligand flipping of UiO-bpydc upon ND_3 adsorption.¹¹²

This phenomenon was further investigated by DFT calculations to understand the role of the bipyridine ligands of UiO-bpydc in forming the H-bonding network with NH_3 .

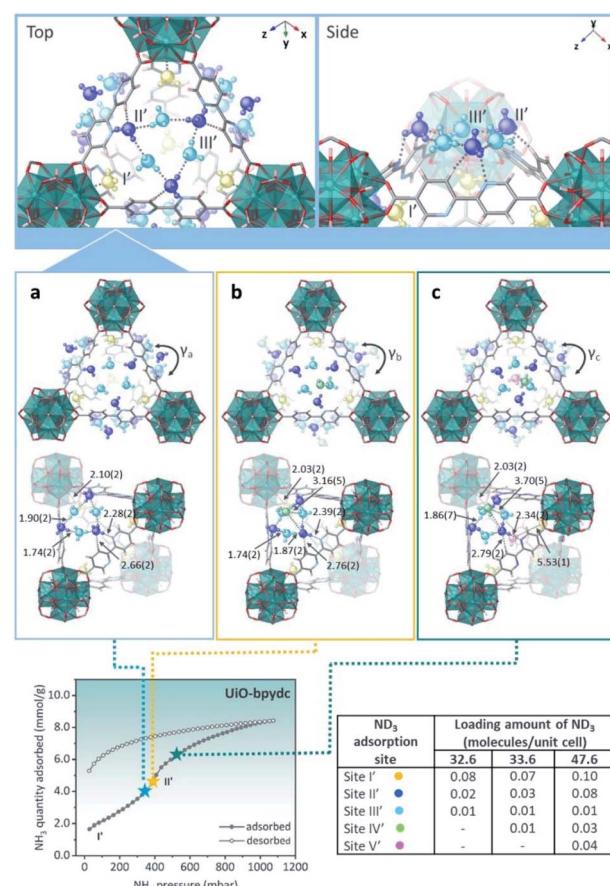


Fig. 22 (Top) Refined ND₃ positions in the UiO-bpydc structure at ND₃ loadings of (a) 4.15, (b) 4.27, and (c) 6.06 mmol g⁻¹ obtained from NPD data at 300 K. (Bottom) ND₃ molecules at different binding sites: site I: yellow, site II: blue, site III: light blue, site IV: green, and site V: pink. NH₃ sorption isotherm at 298 K and up to 1100 mbar; adsorption: closed circles; desorption: open circles reprinted with permission from ref. 112. Copyright (2021) American Chemical Society.

molecules, in comparison with the biphenyl ligands of UiO-67. Thus, the DFT-optimised structures demonstrated the progressive distortion of the aligned bipyridine ligands with γ changed from 20.34° (in good agreement with the NPD refined desolvated structure of UiO-bpydc) to 19.47° , 10.83° , 9.49° , and 6.33° by increasing the number of NH_3 molecules at the trigonal window. Finally, these remarkable results demonstrated that the different pore openings (windows) induced by missing ligands can introduce stepped NH_3 sorption with a strong hysteresis into the UiO type MOFs (*i.e.*, biphenyl ligands are replaced by pyridine ligands).¹¹²

5. NO_x

Nitrogen oxides ($\text{NO}_x = \text{N}_2\text{O}$, NO , N_2O_3 , NO_2 , N_2O_4 and N_2O_5) are considered one of the major air pollutants generated by anthropogenic activities, particularly those that involve fuel combustion from stationary and mobile sources such as thermal power plants and vehicles.¹¹⁴ It has been estimated that fossil fuel combustion generates around 5% of NO_2 , and 95% NO .¹¹⁴ Nevertheless, once NO is released into the atmosphere, it rapidly reacts with O_2 to form NO_2 . The latter is a poisonous red-orange gas responsible for the reddish-brown colour of the smog, although at lower temperatures it can dimerize into a colourless N_2O_4 dimer. Due to the highly reactive nature of NO_x , the uncontrolled emissions of such gases into the atmosphere are associated with a series of health and environmental issues.^{114–116} For instance, the presence of high concentrations of NO_2 has been directly related to the formation of tropospheric ozone O_3 ,^{109,116} which not only is considered an important greenhouse gas; but is also associated with pulmonary and chronic respiratory diseases.^{115,116} The adverse effects caused by atmospheric NO_2 have motivated an intense research for the development of NO_2 abatement technologies.^{114,117}

One of the most studied methods to mitigate the atmospheric NO_2 is the development of porous materials for the selective capture of this pollutant. The traditional adsorbent materials used for NO_2 removal include zeolites,¹¹⁸ calcium-based adsorbents,¹¹⁹ and activated carbon.¹²⁰ However, such materials are typically affected by the reactive oxidative nature

of NO_2 , which hampers the fully reversible desorption of the guest molecules and limits the material regeneration.¹¹⁷

5.1 MOFs for NO_x capture

Current investigations in NO_x sorption have pointed out that MOFs represent an attractive alternative for storage, selective separation, and/or catalytic transformation of NO_2 .¹²¹ Such materials not only display a large surface area, but they also exhibit tuneable pore functionalities¹²² allowing for the stabilisation of guest molecules through the formation of supramolecular host-guest interactions. For instance, a study reported by Peterson *et al.*¹²³ points out the key role of the organic linker for the capture of NO_2 . In this work, the authors conducted a comparative study about the NO_2 sorption performance of UiO-66 and UiO-66- NH_2 .

The gas sorption experiments were performed under dry and controlled humid conditions. The results obtained from the microbreakthrough experiments reveal that UiO-66- NH_2 exhibits a higher capacity for the NO_2 sorption than UiO-66 (20.3 mmol g⁻¹ vs. 8.8 mmol g⁻¹; respectively). Under humid conditions (80% RH), UiO-66- NH_2 exhibits higher uptake of NO_2 than under dry conditions (31.2 mmol g⁻¹ vs. 20.3 mmol g⁻¹; respectively), and it produces a significantly lower amount of NO as a by-product than its analogous UiO-66 (4.5%, 9.5%; respectively).

Such differences were explained in terms of the higher capability of UiO-66- NH_2 to adsorb H_2O vapours. As the H_2O co-adsorbed within the pore network might enhance the stabilization of NO_2 molecules through the formation of supramolecular interactions and it facilitates the preferential formation of nitrous acid as a by-product. The complementary characterization of UiO-66- NH_2 upon NO_2 adsorption revealed that although the crystallinity of UiO-66- NH_2 remains intact, the high reactivity of the adsorbate leads to a series of reactions with the organic linker, which ends up in the post functionalization of the phenyl group (Fig. 23).¹²³

The main transformation that suffered the organic linker upon NO_2 uptake were the nitration of the aromatic ring and the formation of the diazonium ion at the amino group, while in the

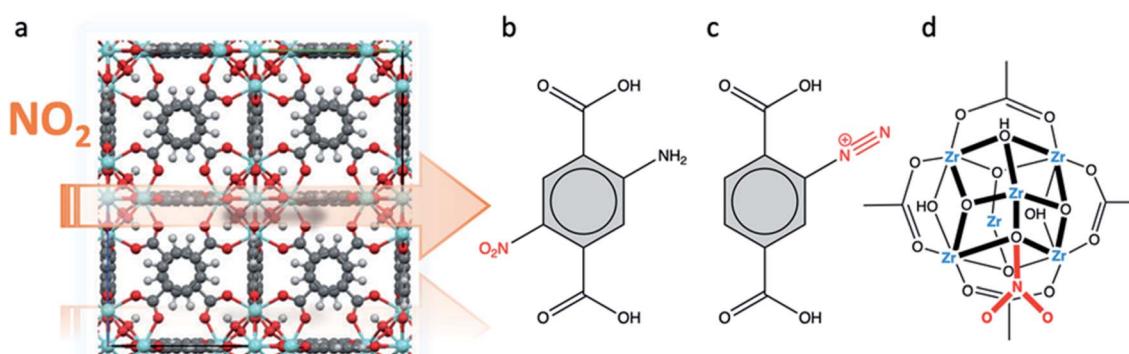


Fig. 23 (a) Structure of UiO-66 analogues. Post functionalization of phenyl ring modifications upon NO_2 sorption: (b) nitration of the phenyl ring, (c) diazonium ion formation. (d) Nitration of the bridging hydroxyl group at the SBU. Reprinted (adapted) with permission from ref. 123. Copyright (2016) John Wiley & Sons.



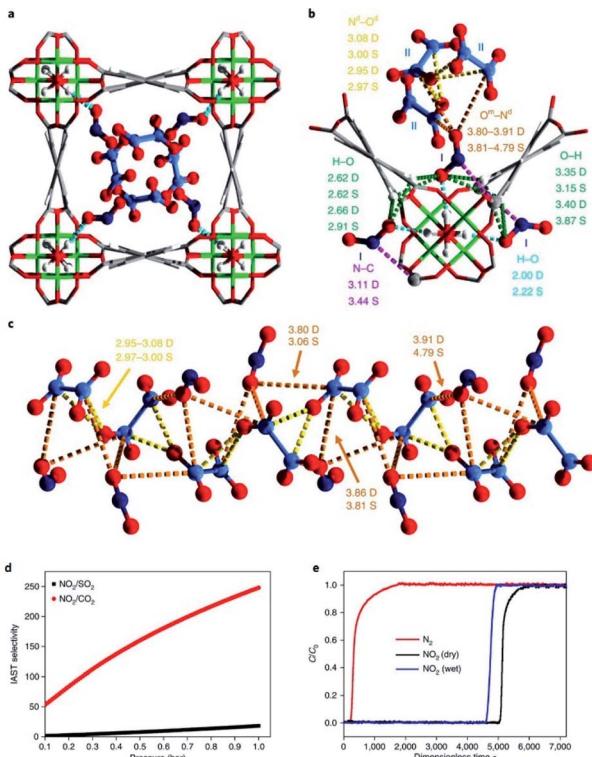


Fig. 24 Views of the structural model of MFM-300(Al)·(NO₂)(N₂O₄)₂. Al-green, C-grey, O-red, H-white, N-blue. (a) and (b) host-guest supramolecular interactions. (c) Structural view of 1D helical chain (NO₂, N₂O₄)_∞ within the channel of MFM-300(Al). (d) Comparison of IAST selectivities for equimolar mixtures of NO₂/SO₂ and NO₂/CO₂ at 0.1–1.0 bar for MFM-300(Al) at 298 K. (e) Dimensionless breakthrough curve of 0.5% NO₂ (5000 ppm) diluted in He/N₂ under both dry and wet conditions through a fixed bed packed with MFM-300(Al) at 298 K and 1 bar. Reprinted (adapted) with permission from ref. 124. Copyright (2018) Springer Nature.

inorganic SBU the terminal -OH group is replaced by the NO³⁻ ion. More recently, a study conducted by Schröder and Yang,¹²⁴ demonstrated that the synergistic effect between the organic ligand and the inorganic building block in MFM-300(Al) allows for the stabilisation of highly reactive NO₂ species within the pore network (Fig. 24a-c). The authors reported that under ambient conditions MFM-300(Al) exhibits a fully reversible NO₂ isotherm uptake of 14.1 mmol g⁻¹ at 298 K. Moreover, the host material retains its crystallinity and sorption capacity after five cycles of NO₂ adsorption/desorption. Remarkably, MFM-300(Al) displays outstanding performance for the selective removal of low-concentration of NO₂ (5000 to <1 ppm) from gas mixtures (Fig. 24d). The optimal uptake and selectivity of MFM-300(Al) for NO₂ was attributed to the existence of both host-guest and guest-guest interactions. The former involves mainly the hydrogen bonding interaction between NO_x and the -OH pendant group of the inorganic node, while the latter refers to the supramolecular interactions between the guest molecules in their monomeric (NO₂) and dimeric (N₂O₄) form. Such interactions give rise to a one-dimensional helical chain arrangement comprised of alternating monomer-dimer molecules

(NO₂·N₂O₄)[∞] running along the channel of MFM-300(Al) (Fig. 24d).

The cooperative supramolecular interactions between NO₂ and N₂O₄, confined within the pore network, allows for the stabilization of the highly reactive NO₂ molecules and inhibits the guest-host electron transfer. This affords an unusual, fully reversible desorption of NO₂ without altering the framework structure of the host material. The structural versatility of MOFs not only permits the stabilization of highly reactive NO_x species within the pore channels but also opens the possibility for the catalytic conversion of NO_x into more valuable and/or less abrasive compounds. In this regard, Schröder and Yang reported the adsorption and catalytic transformation of NO₂ by using a redox-active MOF,¹²⁵ termed MFM-300(V) [V₂(OH)₂(C₁₆H₆O₈)]. This system not only exhibits high adsorption capacity for NO₂ (13 mmol g⁻¹ at 298 K and 1.0 bar), but it also allows for the catalytic reduction of NO₂ into NO. The structural analysis reveals that upon gas sorption, NO₂ molecules get primarily anchored to the pore walls through a hydrogen-bonding interaction with the bridging hydroxyl groups of the inorganic nodes. Further stabilization is reached by forming of 8-fold supramolecular interactions with the aromatic ligand (Fig. 25). Then, those interactions give rise to the host-guest charge transfer process and the formation of NO (through the oxidation of the metal centre from V³⁺ to V⁴⁺), and water, which is produced by the deprotonation of the bridging -OH groups anchored to the SBU.

Recent research conducted by Dincă and co-workers highlights the capacity of M-MFU-4*l* (M = Cu, Zn) MOFs for NO capture and on-demand desorption of guest molecules by a thermal treatment.¹²⁶ In this study, the authors demonstrated that the gas-sorption properties of the host material can be easily adjusted by modifying the cation identity and the oxidation state of the inorganic building block. To this aim, three different systems were selected as host materials (i) Zn₅Cl₄(-BTDD)₃; (ii) Zn₃Cu^{II}₂Cl₄(BTDD)₃; and (iii)

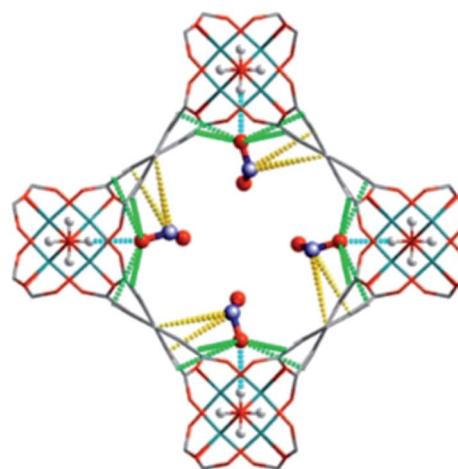


Fig. 25 Rapid scan RS model showing the packing of guest molecules within MFM-300(V³⁺). Reprinted with permission from ref. 125. Copyright (2020) American Chemical Society.



$\text{Zn}_3\text{Cu}^1\text{Cl}_2(\text{BTDD})_3$ ($\text{H}_2\text{BTDD} = \text{bis}(1\text{H}-1,2,3\text{-triazolate}[4,5-\text{b};4',5'-\text{i}]\text{dibenzo}[1,4]\text{dioxin})$). The first material $\text{Zn}_3\text{Cl}_4(\text{BTDD})_3$ (Zn-MFU-4l) displayed the lowest capacity for NO sorption (0.51 mmol g^{-1} at 750 torr). The second system, $\text{Zn}_3\text{Cu}^2\text{Cl}_2(\text{BTDD})_3$ ($\text{Cu}^{\text{II}}\text{-MFU-4l}$), was obtained by substituting two Zn^{2+} atoms from Zn-MFU-4l by Cu^{2+} ions. This subtle modification resulted in a significant increase in the NO sorption capacity (1.43 mmol g^{-1} at 750 torr).

These findings were explained in terms of the differences in the tetrahedral coordination environment around the metal centre $\text{N}_3\text{M}-\text{Cl}$ ($\text{M} = \text{Zn}^{2+}, \text{Cu}^{2+}$). In the first case, the saturated coordination environment around Zn^{2+} prevents the formation of stronger metal-NO interactions, whereas in the second system, the $\text{N}_3\text{Cu}-\text{Cl}$ geometry is sufficiently distorted to allow the NO molecules to approach closer to the metal centre leading to stronger Cu-NO interactions. The formation of Cu^{2+} -nitrosyl species upon NO adsorption was corroborated by diffuse-reflectance infrared Fourier-transform spectroscopy (DRIFT) study. Finally, the third system ($\text{Cu}^1\text{-MFU-4l}$) was obtained by the reduction of Cu^{2+} to Cu^1 accompanied by the concomitant loss of chloride. This system displays a significant improvement in the low-pressure NO uptake (1.24 mmol g^{-1} below 1.4 torr). This value is three times the amount of NO adsorbed by $\text{Cu}^{2+}\text{-MFU-4l}$ ($0.0008 \text{ mmol g}^{-1}$) and $\text{Zn}^{2+}\text{-MFU-4l}$ ($0.0002 \text{ mmol g}^{-1}$) under the same conditions. The DRIFT analysis suggests that the metal centre in $\text{Cu}^1\text{-MFU-4l}$ allows for the formation of $\text{Cu}^1\text{-NO}$ complex even when the material is exposed to low concentrations of NO in Ar (10 ppm). This interaction favours the NO disproportionation $3 \text{ NO} \rightarrow \text{NO}_2 + \text{N}_2\text{O}$ and the oxidation of the metal centre to Cu^{2+} which strongly binds to the NO_2 product (Fig. 26). Then, the Cu^{2+} -nitrite bond can be cleaved by exposing $\text{Cu}^{2+}\text{-MFU-4l-(NO}_2\text{)}$ to high temperatures, releasing

the anchored NO_2 and thereby regenerating the adsorbent $\text{Cu}^1\text{-MFU-4l}$. According to the authors, this catalytic system represents a potentially attractive scheme for cold-start NO capture. The design of porous materials for storage and controlled release of NO have gained relevance in biomedicine, as it has been shown that the exposure to controlled doses of NO induces wound healing and prevents the formation of blood clots. Moreover, NO is a potent antimicrobial agent; therefore, the development of materials for the local release of NO could be highly advantageous to reduce the risk of infections.

One of the first reports about the use of MOFs for the on-demand delivery of NO was published in 2007 by Morris and co-workers.¹²⁷ In this work, the authors used HKUST-1 as a reservoir for NO. The gravimetric adsorption experiments performed at 196 K and 298 K display significant isotherm hysteresis with an adsorption capacity of *ca.* 9 mmol g^{-1} and 3 mmol g^{-1} at one bar, respectively. Upon gas desorption processes, both isotherms exhibit a remaining amount of NO trapped within the pore network (2 mmol g^{-1}). The IR analysis of NO-loaded material reveals that the presence of open copper sites in the walls of HKUST-1 framework allows for the formation of $\text{NO}\cdots\text{Cu}^{2+}$ coordination adduct; thereby leading to the irreversible adsorption of NO. However, NO molecules trapped within the pore network can be released on demand upon exposure of the NO-loaded HKUST-1 to water vapours, as H_2O acts as a nucleophile replacing the NO molecules coordinated to the open copper sites.

More recently, the same research group reported the development of MOF-based composites for the controlled delivery of NO.¹²⁸ The selected host material was CPO-27-Ni, as this MOF affords highly efficient adsorption, storage, and release cyclable profile for NO.¹²⁹ Moreover, this material exhibits high stability towards adsorbed water, which allows for the controlled release of NO when exposed to moisture, without altering the structural arrangement of the host material.

The optimal MOF films were obtained by varying the amount of CPO-27-Ni (wt%) integrated within a polyurethane matrix. The composite films exhibited homogeneous distribution of the MOF within the polymer matrix. The release of NO was triggered by exposing the composite films to 11% relative humidity. Such conditions allow for the controlled release of the NO adsorbed without affecting the polymer properties. Finally, the antimicrobial properties of the resultant MOF-based composite demonstrated that the controlled release of biologically active levels of NO provides a bactericidal effect against *Escherichia coli* (*E. coli*) and *Staphylococcus aureus* (*S. aureus*).

6. Selectivity of trace gases

One important aspect of evaluating MOF materials is their selectivity of adsorbing one gas preferentially over other molecules that are present in the industrial flue gas streams, as these are mainly composed of CO_2 (10–15% [v]), N_2 (70–75% [v]), and H_2O (~10–18% [v]),¹³⁰ whereas gases such as SO_2 (0.00325% [v], 500–3000 ppm), H_2S (13.5% [v], 1000–4000 ppm), NH_3 (300–800 ppm) and NO_x (200–800 ppm) are present in trace concentrations.¹³¹ For this reason, it is highly desirable that the study of the capture of trace

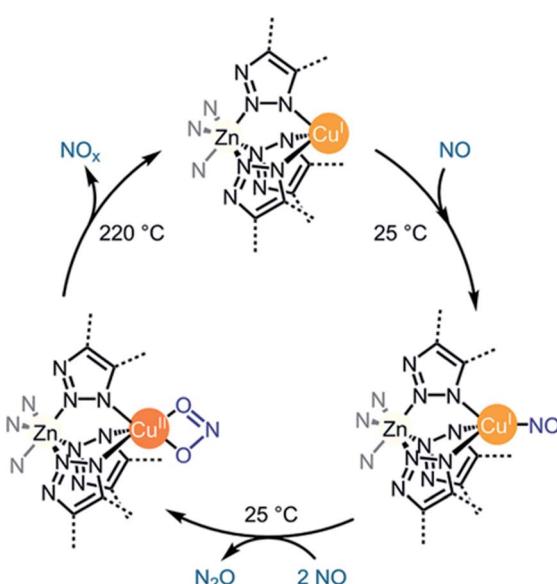


Fig. 26 Proposed NO Disproportionation Cycle Using $\text{Cu}^1\text{-MFU-4l}$. During the thermal decomposition of $\text{Cu-MFU-4l-(NO}_2\text{)}$, NO is the major NO_x species detected. Reprinted with permission from ref. 126. Copyright (2021) American Chemical Society.



gases by MOFs should be accompanied by a selectivity study in the presence of other gases in order to have a complete profile of their potential applications at industrial levels.

Table 4 lists the most promising MOF materials for the separation of SO_2 and H_2S gases, with a good selectivity performance.

7. Stability and reusability of MOFs

As previously shown, there is a wide variety of MOFs with high structural stability towards corrosive gases (SO_2 , H_2S , NH_3 and NO_x). The stability, as well as the reusability of the MOFs, is required for tangible industrial applications on the adsorption field. Considering that the CSD currently contains near to 70 000 MOF structures,¹⁴³ a selection criterion for MOFs that could be potentially useful in the adsorption of toxic gases is essential, as are the guidelines for generating new materials focused on this target.¹⁴⁴ Although the nature of MOFs is very

diverse, MOFs stable in the presence of corrosive gases feature some common characteristics: (i) the strength of the ligand–metal bond; (ii) oxidation state of the metal centre; (iii) robustness of the metallic cluster; and (iv) the formation of supramolecular host–guest interactions. The presence of at least one of these characteristics is responsible for the stabilization of the framework under such harsh conditions. For example, when the strongest binding site corresponds to a free metal site, the metal–ligand bond should be strong enough to avoid linker displacement and subsequent structure degradation. The combination of ligands with strong donor groups/atoms and late transition metals, or the use of metals with a higher oxidation state together with carboxylate ligands are good synthetic strategies to improve the strength of the metal–ligand bonds and, therefore, increase the overall stability of the framework.^{21,40a,145} For example, MOFs formed by triazolate ligands such as the $\text{M}_2\text{Cl}_2\text{BTDD}$ family ($\text{M} = \text{Ni}^{2+}$, Co^{2+} , Mn^{2+} , and Cu^{2+})¹⁰⁷ with uncoordinated metal sites are stable even after

Table 4 Summary of the adsorption captures of SO_2 and H_2S and IAST selectivity of each one in various MOFs materials

Gas target	MOF	Gas adsorption [mmol g ⁻¹]	Flue gas concentration	Selectivity					Ref.
				CO_2	N_2	CH_4	Other gases	Ref.	
SO_2	MIL-160	7.2	SO_2/CO_2 (10 : 90–50 : 50 v/v)	124–128 (IAST)	31.25 ^{37e}	—	—	—	18
	MIL-125(Ti)-NH ₂	10.8	SO_2/CO_2 (10 : 90–50 : 50 v/v)	42–55 (IAST)	—	—	—	—	18
	MFM-170	17.5	Equimolar SO_2/N_2	35	944	260	203 (CO)	—	19
	MFM-300(Al)	7.1	50–350 mbar ^b	—	6522	3620	10^5 (H_2)	3105(CO) 4974 (O_2)	37a
	MFM-300(In)	8.28	At 50 : 50 mixture of each gas	60	5000	425	—	—	37b
	MFM-601	12.3	SO_2/CO_2 , SO_2/N_2 (50 : 50 to 10 : 90)	32	255	—	—	—	132
	ELM-12	2.73	10 : 90 mixture at 298 K and 1 bar	30	4064	871	—	—	133
	1@Ba(OH)2	4.0 ^a	$\text{N}_2/\text{CO}_2/\text{SO}_2$ (83.5 : 14 : 2.5 v/v)	990	—	—	—	—	45
	SIFSIX-1-Cu	11.01	At 10 : 90 mixture of each gas	70.7	3145.7	1241.4	—	—	51
	SIFSIX-2-Cu-i	6.9	At 10 : 90 mixture of each gas	87.1	3103.2	1017.1	—	—	51
	SIFSIX-3-Zn	2.1	At 10 : 90 mixture of each gas	—	506.7	276	—	—	51
	SIFSIX-3-Ni	2.74	At 10 : 90 mixture of each gas	—	701.8	371.6	—	—	51
	KAUST-7	1.4 ^a	$\text{SO}_2/\text{CO}_2/\text{N}_2$ (4%:4% : balance) and (500 ppm : 10% : balance)	1	—	—	—	—	52
	KAUST-8	1.6 ^a	$\text{SO}_2/\text{CO}_2/\text{N}_2$: 0.05/10/89.95	66	—	—	—	—	52
	ECUT-77	8.0	SO_2/CO_2 (1 : 99 v/v; 2000 ppm SO_2)	44–36	—	—	—	—	134
	ECUT-100	4.95	At 1 : 99 mixture of each gas	27.5–26.9	—	2302	—	—	135
	ECUT-111	11.56	SO_2/CO_2 , SO_2/N_2 (1 : 99 v/v)	22.2–25.2	860.9	—	—	—	136
H_2S	MIL-101(Cr)	0.419 ^a	$\text{H}_2\text{S}/\text{CO}_2/\text{He}$ (1 : 10 : 89 v/v)	5	—	—	—	—	63
	HKUST-1	~1 ^a	$\text{H}_2\text{S}/\text{CO}_2/\text{He}$ (1 : 10 : 89 v/v)	~40	—	—	—	—	63
	KAUST-7	—	$\text{H}_2\text{S}/\text{CO}_2/\text{CH}_4$ (20 : 20 : 60)	—	—	20.7 ^d	—	—	139
	KAUST-8	—	$\text{H}_2\text{S}/\text{CO}_2/\text{CH}_4$ (20 : 20 : 60)	—	—	18.6 ^d	—	—	139
	KAUST-8	—	$\text{H}_2\text{S}/\text{CO}_2/\text{CH}_4$ (5 : 5 : 90)	1	—	—	—	—	137
	SIFSIX-2-Ni-i	—	$\text{H}_2\text{S}/\text{CO}_2/\text{CH}_4$ (5 : 5 : 90)	3.3	—	—	—	—	137
	MIL-125(Ti)-NH ₂	—	H_2S 0.001	—	—	70 ^c	—	—	138
	Ga-soc-MOF-1a	4.4	$\text{CO}_2/\text{H}_2\text{S}/\text{CH}_4$ (5% : 5% : 90%)	7	—	—	—	—	64
	Y-fum-fcu-MOF	1.1 ^a	$\text{CO}_2/\text{H}_2\text{S}/\text{CH}_4$ (5% : 5% : 90%)	6.4	—	—	—	—	139
	[BMIM][Cl]/Cu-TDPAT ^e	—	H_2S 1000 ppm	—	—	1302	—	—	140
	Cu-TDPAT ^e	—	H_2S 1000 ppm	—	—	141	—	—	140
	MIL-53(Al) ^f	—	$\text{CH}_4/\text{CO}_2/\text{H}_2\text{S}$ (60% : 39.9% : 0.1%)	—	—	23–34	—	—	141
	MIL-47(V)-Br ^g	—	$\text{H}_2\text{S}/\text{CH}_4$ (5 : 95) 0.1 MPa	—	278	68	—	—	142
	MIL-47(V) ^g	—	$\text{H}_2\text{S}/\text{CH}_4$ (5 : 95) 0.1 MPa	—	179	223	—	—	142
			$\text{H}_2\text{S}/\text{CH}_4$ (1 : 99) 0.1 MPa			51			
			$\text{H}_2\text{S}/\text{N}_2$ (50 : 50)						

^a Breakthrough experimental result. ^b Obtained from ratio of slopes of initial adsorption isotherm plot at a low pressure region. ^c GCMC simulations at 303 K and 10 bar. ^d Membrane composite measured at 35 °C. ^e Composite materials. ^f Pellets. ^g GCMC simulations at 273 K.



several sorption and desorption cycles of NH_3 . Such stability increases according to the Irving–Williams series, being the Ni^{2+} framework more stable than the Cu^{2+} . In this case, the presence of heterolytic metal–ligand bonds and a less labile metal centre, enhances the MOF's stability. In MFM-170,¹⁹ a MOF formed by the Cu^{2+} paddlewheel SBU, the coordination of one of the Cu^{2+} centres to a N-donating ligand, allowed reversible SO_2 adsorption, even if one of the Cu^{2+} atoms presented weak contacts with SO_2 . Additionally, the interactions between SO_2 and the surface of the framework contributed to the reversibility of the adsorption process. This research represents an example of enhancing the chemical stability of a MOF by increasing the metal–ligand bond strength, a highly effective strategy to increase the stability of MOFs with labile metals.

On the other hand, the use of small metal cations confers stability against chemical degradation, as shown for several MOFs constructed with trivalent and tetravalent cations such as Al^{3+} , In^{3+} , Sc^{3+} , and Cr^{3+} and Zr^{4+} , V^{4+} , and Ti^{4+} , respectively. These metal centres tend to form robust clusters, increasing the stability of the framework towards corrosive gases. For example, MIL-101(Cr) and MIL-101(Cr)-4F(1%) formed by $[\text{Cr}_3(\mu_3\text{-O})(\text{O}_2\text{CR})_6]$ clusters, are stable to SO_2 and H_2S ;^{20,34} Mg-CUK-1 with $[\text{Mg}_3(\mu_3\text{-OH})]^{5+}$ clusters is stable to H_2S ;⁶⁸ MIL-53(Al)-TDC formed by *trans*-corner-sharing $[\text{Al}(\mu_2\text{-OH})(p\text{-BDC})]$ polyhedra is stable to H_2S .⁶⁹ Additionally, the MFM-300 family formed by infinite chains of $[\text{M}_2(\mu_2\text{-OH})_2(\text{C}_{16}\text{O}_8\text{H}_6)]$ ($\text{M} = \text{Al}^{3+}$, In^{3+} , Sc^{3+} , V^{3+}), has demonstrated to be stable to SO_2 , H_2S , NH_3 , and NO_2 .^{37,70,110,111,125} This family represents an excellent example of how the non-covalent interactions formed between the framework surface and the host molecules provide to the additional structural stability and adsorption reversibility in almost all the studied cases. The principal interaction involves the $\mu_2\text{-OH}$ motif as hydrogen bond donor and the gas molecule as an acceptor, followed by guest–ligand and guest–guest interactions. This last aspect has a direct impact on the recyclability of the MOF, where most of the materials with good cyclability exhibit similar host–guest interactions such as hydrogen bonds (strong and weak), electrostatic interactions inside the pore framework, and metal–guest interactions. For example: MFM-170,¹⁹ MIL-101(Cr)-4F(1%),²⁰ are stable even after 50 SO_2 adsorption–desorption cycles while the following MOFs are stable for different numbers of SO_2 cycles $[\text{Ni}_8(\text{OH})_4(\text{H}_2\text{O})_2(\text{-BDP-X}_6)]$ (10 cycles),⁴⁵ SIFSIX-1-Cu (4–6 cycles),⁵¹ and MFM-300(Sc) (10 cycles).^{37c} Furthermore, Mg-CUK-1,⁶⁸ MIL-53(Al)-TDC,⁶⁹ and MFM-300(Sc)⁷⁰ are stable for at least five adsorption–desorption cycles of H_2S while $\text{Mg}_2(\text{dobpdc})$,¹⁰⁹ $\text{Co}_2\text{Cl}_2\text{BBTA}$ and $\text{Co}_2\text{Cl}_2\text{BTDD}$,¹⁰⁸ MFM-300(Al) (50 cycles),¹¹¹ and UiO-67 .¹¹² Finally, MFM-300(Al) is not affected even by five cycles of NO_2 .¹²⁴

8. Conclusions and future perspectives

Air pollution due to the emissions of toxic gases is the single largest environmental health risk around the world since it is responsible for one in eight premature global deaths, reduction

of the biodiversity, crop damages and acidification of soils and waters. Thus, the development of efficient technologies for the capture of toxic gases (*i.e.*, NO_x , SO_2 , NH_3 and H_2S) from static and mobile sources is essential, in order to achieve a cleaner environment. MOFs are amongst the most promising candidates for the capture of these toxic gases since their sorption selectivity is directly tuneable as a function of the topology and chemical composition of the pores, which can optimise the interactions between MOFs and these toxic molecules, affording enhanced gas adsorption properties. Although the main drawback of MOFs is their vulnerability to these highly corrosive gases which can compromise their chemical stability, remarkable examples have demonstrated high chemical stability to NO_x , SO_2 , NH_3 and H_2S .

Chemically stable MOFs for the capture of SO_2 have provided promising advances in the field and the understanding of the role of different chemical functionalities, is the key for accomplishing superior SO_2 captures. Coordinatively unsaturated metal sites (open metal sites) incorporated into chemical stable MOFs (*e.g.*, MFM-170 and MIL-101(Cr)-4F(1%)) demonstrated the relevance of the interaction of these open metal sites and SO_2 molecules. The design of new functionalised MOF materials with coordinatively unsaturated metal sites signifies an interesting alternative to obtain superior SO_2 captures and therefore, the challenge is to be able to also incorporate chemical stability to these MOFs. Thus, high SO_2 uptakes are related to large BET surface area, pore size and the open metal sites incorporation into MOFs, while structural (chemical) stability is associated with the type of metal, and the interactions adsorbate–adsorbent inside the pore.

Hydroxo-functionalisation ($\mu\text{-OH}$ groups) in structural stable MOFs (*e.g.*, MFM-300(M)) have shown excellent capabilities for the capture and detection of SO_2 with high chemical stability and excellent cyclability involving a remarkably facile regeneration. The key for all of such outstanding properties is the preferential interaction between the functional group ($\mu\text{-OH}$) and SO_2 , in which the $\mu\text{-OH}$ groups binds SO_2 molecules through the formation of $\text{O}=\text{S}=\text{O}(\delta^-)\cdots\text{H}(\delta^+)-\text{O}$ hydrogen bonds, reinforced by weak supramolecular interactions with C–H atoms from the aromatic rings of the framework. The chemical stability of the MFM-300(M) family relies on the strength of the coordination bonds between the oxygen atoms from the carboxylic ligand (biphenyl-3,3',5,5'-tetracarboxylic acid) and the M^{3+} metal centres.

Defective sites *a.k.a.*, crystal irregularities, composition inhomogeneities or defects in MOFs have exhibited increased SO_2 adsorption capacities, energies, and SO_2/CO_2 selectivity. For example, in nickel pyrazolate MOFs it was demonstrated that the pre-synthetic introduction of amino and hydroxyl functional groups on the organic linkers and the post-synthetic modifications, synergistically increase the SO_2 capture capacity of these MOF materials.

Halogen functionalised MOF materials (*e.g.*, hexafluorosilicate (SiF_6^{2-}) incorporation in SIFSIX MOFs) showed highly efficient removal of SO_2 from other gases, at a very low SO_2 concentrations, and excellent SO_2/CO_2 and SO_2/N_2 selectivities. These properties were attributed to the strong

electrostatic interactions between the SO_2 molecules and the SiF_6^{2-} anions ($\text{S}^{\delta+}\cdots\text{F}^{\delta-}$), assisted by dipole–dipole interactions with the ligand ($\text{O}^{\delta-}\cdots\text{H}^{\delta+}$).

The identification of MOFs capable of capturing H_2S under industrially practical pressure-swing desorption conditions, symbolises a very promising application which is based on the chemical stability of these materials. For example, stable MOF materials constructed with μ -OH functional groups (e.g., MIL-53(Al)-TDC) exhibited high and fully reversible H_2S capture where the formation of hydrogen bonds between H_2S and the μ -OH functional group (a relatively weak interaction) facilitates the H_2S cyclability.

The chemical transformation of H_2S within the pores of MOFs to produce, *in situ*, polysulphides is a new and exciting discovery in MFM-300(M) and SU-101 MOF materials. In addition to the promising application as novel MOF-lithium/sulphur batteries, the chemical investigation behind them, provided addition information on the electrochemical potential of MOFs which could guide new interesting properties to be discovered. Then, although the polysulphide formation was explained in terms of a “disproportionation” type reaction (protons of H_2S play the role of an oxidant and the sulphide plays the role of a reducing agent), other hypotheses need to be explored. What if the polysulphide formation is a consequence of reversible metal–ligand bonding upon the adsorption of H_2S ? We anticipate new perspectives will be shortly investigated.

Paradoxically, the role of H_2S as an endogenous biological mediator in the human body is crucial. For example, H_2S executes a vital activity in the regulation of blood pressure, neurotransmission, anti-inflammatory mechanism, anti-oxidation, angiogenesis and apoptosis. Thus, the delivery and detection of H_2S levels in living cells and organisms is an outstanding application for MOF materials. For example, Ni-CPO-27 showed a promising release of H_2S (under physiological conditions) with a short induction period of approximately 5 min. In the field of H_2S detection, for example, a post-synthetic modification of ZIF-90 with malononitrile ($\text{N}\equiv\text{C}-\text{CH}_2-\text{C}\equiv\text{N}$), exhibited a specific reaction with H_2S completing an enhancement of photoluminescence, which constitutes the base for the detection of H_2S . Thus, the development of this research is expanding by the fundamental principle of introducing a reactive site for H_2S within MOF material (monitored by fluorescence), which can work under physiological pH, to detect exogenous H_2S in living cells, while the MOF-probe shows low toxicity and high biocompatibility.

Although most MOF materials have exhibited poor chemical stability towards NH_3 , chemically stable MOFs constructed with different chemical functionalities have shown interesting NH_3 capture performances.

Coordinatively unsaturated metal sites incorporated, for example, in bisbenzenetriazolate and triazolate MOFs, showed high and reversible NH_3 uptakes with strong interactions between the open metal sites and NH_3 . The chemical stability of these MOF materials was attributed to the strength of the N atom coordinated to divalent metal centres. $\text{Mg}_2(\text{dobpdc})$ demonstrated high NH_3 capture and chemical stability to not only dry NH_3 , but also to wet NH_3 . Interestingly, it was

experimentally and computationally demonstrated that NH_3 was preferentially adsorbed, by coordination bonds, to the coordinatively unsaturated metal sites of $\text{Mg}_2(\text{dobpdc})$. The exceptional chemical stability of this Mg^{2+} -based MOF material credited to the higher affinity of Mg^{2+} to oxygen atoms than nitrogen atoms, as confirmed by van der Waals (vdW)-corrected density functional theory (DFT) calculations.

The μ -OH functionalised MFM-300(M) family demonstrated remarkable NH_3 adsorption properties, high chemical stability, high uptakes and attractive cyclabilities. Not surprisingly, as in the case of SO_2 and H_2S , the preferential adsorption sites of the MFM-300(M) materials were found at the hydroxo functional group. For example, in the case of MFM-300(Al) the adsorption of NH_3 showed an atypical adsorption mechanism where adsorbent and adsorbate experienced a rapid site-exchange *via* reversible formation and cleavage of $\text{O}-\text{H}$ and $\text{O}-\text{H}_{(\text{from NH}_3)}$ chemical bonds, in other words a pseudo-chemisorption binding mechanism.

As previously described in the SO_2 capture section, crystal irregularities can provide very particular physical and chemical properties. Then, defects in MOF materials also demonstrated remarkable NH_3 adsorption properties. For example, UiO-67 and UiO-bpydc containing 4,4'-biphenyl dicarboxylate and 4,4'-(2,2'-bipyridine) dicarboxylate ligands, despite their structural similarities demonstrated a drastic difference in the NH_3 adsorption properties when the biphenyl groups in the organic ligand were replaced by the bipyridine moieties. Such replacement can confer flexibility to the framework in the context of mainly ligand “flipping” but without significant pore volume alteration.

Current investigations in NO_x capture have pointed out that chemically stable MOFs represent an attractive alternative for storage, selective separation, and/or catalytic transformation of NO_2 . For example, the synergistic effect between the organic ligand and the inorganic building block in MFM-300(Al) showed the stabilisation of highly reactive NO_2 species within the micropores. The optimal uptake and selectivity of MFM-300(Al) for NO_2 was attributed to the existence of both host–guest and guest–guest interactions. The former involves mainly the hydrogen bonding interaction between NO_x and the hydroxo functional group, while the latter refers to the supramolecular interactions between the guest molecules in their monomeric (NO_2) and dimeric (N_2O_4) form.

Interestingly, the design of MOF materials for storage and controlled release of NO have gained relevance in biomedicine, as it has been shown that the exposure to controlled doses of NO induces wound healing and prevents the formation of blood clots. For example, CPO-27-Ni exhibited highly efficient adsorption, storage, and release cyclable profile for NO with a high stability to water, allowing a controlled release of NO when exposed to moisture.

Despite the significant progress in the capture of NO_x , SO_2 , NH_3 and H_2S pollutants by MOF-based technology, further research in this field is required to overcome some of the main challenges: chemical stability, reusability, and suitable functionalisation. Considering the large number of MOF materials synthesised to date (over 70 000 structures reported in the



CCD), it is crucial to understand what makes the above-reported MOFs so chemically stable. Thus, this perspective summarises their most important characteristics (metal centre, strength of the metal-ligand bond and functional groups) that should be further investigated in order to be able to design tailor-made MOFs for the capture of corrosive gases. Additionally, an important aspect to consider is the study of breakthrough experiments and selectivity calculations, where more realistic conditions for industrial applications are investigated, providing a direct performance comparison of MOFs to classic materials such as activated carbon, zeolites and metal oxides. On the other hand, the synthesis scalability of MOFs is one of the weak points to contemplate for corrosive gas adsorption applications, as only a few materials can be synthesised on an industrial scale. We desire that this perspective can provide useful information on the significant progress of this field and inspire new investigations to be carried out.

Author contributions

E. M.-A., M. L. D.-R., and M. J. V.-H. conducted the literature research and drafted the manuscript. V. J. and I. A. I. conceived the topic and structure of the article. Outlined, drafted and supervised the completion of the manuscript. All authors reviewed and contributed to this manuscript.

Conflicts of interest

There are no conflicts to declare.

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Notes and references

- R. A. Silva, J. J. West, J.-F. Lamarque, D. T. Shindell, W. J. Collins, G. Faluvegi, G. A. Folberth, L. W. Horowitz, T. Nagashima, V. Naik, S. T. Rumbold, K. Sudo, T. Takemura, D. Bergmann, P. Cameron-Smith, R. M. Doherty, B. Josse, I. A. MacKenzie, D. S. Stevenson and G. Zeng, *Nat. Clim. Change*, 2017, **7**, 647–651.
- World Health Organization, [https://www.who.int/en/news-room/fact-sheets/detail/ambient-\(outdoor\)-air-quality-and-health](https://www.who.int/en/news-room/fact-sheets/detail/ambient-(outdoor)-air-quality-and-health), accessed December 2020.
- K. Balakrishnan, M. Brauer, N. Bruce, F. Forastiere, N. C. Gouveia, R. M. Harrison, F. J. Kelly, P. L. Kinney, M. Lippmann, S. Mehta, B. Ostro, A. Peters, P. H. N. Saldiva, J. M. Samet, M. J. Utell, M. Zuk, *World Health Organization, Air quality guidelines. Global update 2005. Particulate matter, ozone, nitrogen dioxide and sulfur dioxide*, WHO Regional Office for Europe, Copenhagen, 2006.
- G. Shaddick, M. L. Thomas, P. Mudu, G. Ruggeri and S. Gumy, *npj Clim. Atmos. Sci.*, 2020, **3**, 1–5.
- J. Lelieveld, J. S. Evans, M. Fnais, D. Giannadaki and A. Pozzer, *Nature*, 2015, **525**, 367–371.
- B. Chu, X. Zhang, Y. Liu, H. He, Y. Sun, J. Jiang, J. Li and J. Hao, *Atmos. Chem. Phys.*, 2016, **16**, 14219–14230.
- (a) Y. Wang, P. K. Hopke, D. C. Chalupa and M. J. Utell, *Aerosol Sci. Technol.*, 2019, **45**, 1245–1249; (b) R. Figueiredo, P. Nunes, M. Meireles, M. Madaleno and M. C. Brito, *J. Cleaner Prod.*, 2019, **222**, 129–142; (c) X. Kan, F. Hedenus and L. Reichenberg, *Energy*, 2020, **195**, 117434.
- (a) M. Sharifzadeh, G. Triulzi and C. L. Magee, *Energy Environ. Sci.*, 2019, **12**, 2789–2805; (b) J. Guo, W. Jiao, G. Qi, Z. Yuan and Y. Liu, *Chin. J. Chem. Eng.*, 2019, **27**, 1361–1373.
- (a) E. Martínez-Ahumada, A. López-Olvera, V. Jancik, J. E. Sánchez-Bautista, E. González-Zamora, V. Martis, D. R. Williams and I. A. Ibarra, *Organometallics*, 2020, **39**, 883–915; (b) T. Islamoglu, Z. Chen, M. C. Wasson, C. T. Buru, K. O. Kirlikovali, U. Afrin, M. R. Mian and O. K. Farha, *Chem. Rev.*, 2020, **120**(16), 8130–8160.
- (a) K. Dedecker, E. Dumas, B. Lavédrine, N. Steunou and C. Serre, in *Metal-Organic Frameworks (MOFs) for Environmental Applications*, ed. S. K. Ghosh, Elsevier, Amsterdam, 2019, ch. 5, pp. 141–178; (b) A. J. Rieth, A. M. Wright and M. Dincă, *Nat. Rev. Mater.*, 2019, **4**, 708–725.
- F. C. Men and H. M. Seip, *Environ. Sci. Policy*, 2004, **7**, 253–265.
- (a) M. R. Taylor, E. S. Rubin and D. A. Hounshell, *Technol. Forecast. Soc. Change*, 2005, **72**, 697–718; (b) S. Reis, P. Grennfelt, Z. Klimont, M. Amann, H. ApSimon, J.-P. Hettelingh, M. Holland, A.-C. LeGall, R. Mass, M. Posch, T. Spranger, M. A. Sutton and M. Williams, *Science*, 2012, **338**, 1153–1154.
- T. Larssen, E. Lydersen, D. Tang, Y. He, J. Gao, H. Liu, L. Dual, H. M. Seip, R. D. Vogt, J. Mulder, M. Shao, Y. Wang, H. Shang, X. Zhang, S. Solberg, W. Aas, T. Oakland, O. Eilertsen, V. Angell, Q. Li, D. Zhao, R. Xiang, J. Xiao and J. Luo, *Environ. Sci. Technol.*, 2006, **40**, 418–425.
- R. K. Srivastava, W. Jozewicz and C. Singer, *Environ. Prog.*, 2001, **20**, 219–228.
- (a) F. Rezaei, A. A. Rownaghi, S. Monjezi, R. P. Lively and C. W. Jones, *Energy Fuels*, 2015, **29**(9), 5467–5486; (b) Y. Sun, E. Zwolińska and A. G. Chmielewski, *Crit. Rev. Environ. Sci. Technol.*, 2016, **46**, 119–142.
- X. Han, S. Yang and M. Schröder, *Nat. Rev. Chem.*, 2019, **3**, 108–118.
- (a) W. P. Mounfield III, C. Han, S. H. Pang, U. Tumuluri, Y. Jiao, S. Bhattacharyya, M. R. Dutzer, S. Nair, Z. Wu, R. P. Lively, D. S. Sholl and K. S. Walton, *J. Phys. Chem. C*, 2016, **120**(48), 27230–27240; (b) S. Bhattacharyya, S. H. Pang, M. R. Dutzer, R. P. Lively, K. S. Walton, D. S. Sholl and S. Nair, *J. Phys. Chem. C*, 2016, **120**(48),



27221–27229; (c) S. Glomb, D. Woschko, G. Makhloufi and C. Janiak, *ACS Appl. Mater. Interfaces*, 2017, **9**, 37419–37434.

18 P. Brandt, A. Nuhnen, M. Lange, J. Möllmer, O. Weingart and C. Janiak, *ACS Appl. Mater. Interfaces*, 2019, **11**, 17350–17358.

19 G. L. Smith, J. E. Eyley, X. Han, X. Zhang, J. Li, N. M. Jacques, H. G. W. Godfrey, S. P. Argent, L. J. M. McPherson, S. J. Teat, Y. Cheng, M. D. Frogley, G. Cinque, S. J. Day, C. C. Tang, T. L. Easun, S. Rudić, A. J. Ramirez-Cuesta, S. Yang and M. Schröder, *Nat. Mater.*, 2019, **18**, 1358–1365.

20 E. Martínez-Ahumada, M. L. Díaz-Ramírez, H. A. Lara-García, D. R. Williams, V. Martis, V. Jancik, E. Lima and I. A. Ibarra, *J. Mater. Chem. A*, 2020, **8**, 11515–11520.

21 A. J. Rieth, A. M. Wright and M. Dincă, *Nat. Rev. Mater.*, 2019, **4**, 708–725.

22 S. Luo, J. Liu and Z. Wu, *J. Phys. Chem. C*, 2019, **123**, 11772–11780.

23 S. Chaturvedi, J. A. Rodríguez, T. Jirsak and J. Hrbek, *J. Phys. Chem. B*, 1998, **102**(36), 7033–7043.

24 (a) D. Britt, D. Tranchemontagne and O. M. Yaghi, *Proc. Natl. Acad. Sci. U. S. A.*, 2008, **105**, 11623–11627; (b) T. G. Glover, G. W. Peterson, B. J. Schindler, D. Britt and O. Yaghi, *Chem. Eng. Sci.*, 2011, **66**, 163–170; (c) K. Tan, S. Zuluaga, H. Wang, P. Canepa, K. Soliman, J. Cure, J. Li, T. Thonhauser and Y. J. Chabal, *Chem. Mater.*, 2017, **29**, 4227–4235.

25 A. C. Elder, S. Bhattacharyya, S. Nair and T. M. Orlando, *J. Phys. Chem. C*, 2018, **122**, 10413–10422.

26 K. Tan, P. Canepa, Q. Gong, J. Liu, D. H. Johnson, A. Dyvoich, P. K. Thallapally, T. Thonhauser, J. Li and Y. J. Chabal, *Chem. Mater.*, 2013, **25**, 4653–4662.

27 J. Hungerford, S. Bhattacharyya, U. Tumuluri, S. Nair, Z. Wu and K. S. Walton, *J. Phys. Chem. C*, 2018, **122**, 23493–23500.

28 M. Eigen, *Pure Appl. Chem.*, 1963, **6**, 97–115.

29 C. Petit, B. Levasseur, B. Mendoza and T. J. Bandosz, *Microporous Mesoporous Mater.*, 2012, **154**, 107–112.

30 C. Petit, B. Mendoza and T. J. Bandosz, *ChemPhysChem*, 2010, **11**(17), 3678–3684.

31 H. Irving and J. P. Williams, *J. Chem. Soc.*, 1953, 3192–3210.

32 L. Feng, K.-Y. Wang, G. S. Day, M. R. Ryder and H.-C. Zhou, *Chem. Rev.*, 2020, **120**(23), 13087–13133.

33 L. Hamon, C. Serre, T. Devic, T. Loiseau, F. Millange, G. Férey and G. De Weireld, *J. Am. Chem. Soc.*, 2009, **131**, 8775–8777.

34 M. L. Díaz-Ramírez, E. Sanchez-González, J. R. Álvarez, G. A. González-Martínez, S. Horike, K. Kadota, K. Sumida, E. González-Zamora, M. A. Springuel-Huet, A. Gutiérrez-Alejandre, V. Jancik, S. Furukawa, S. Kitagawa, I. A. Ibarra and E. Lima, *J. Mater. Chem. A*, 2019, **7**, 15101–15112.

35 R. G. Pearson, *J. Am. Chem. Soc.*, 1963, **85**(22), 3533–3539.

36 M. Bosch, M. Zhang and H.-C. Zhou, *Adv. Chem.*, 2014, **2014**, 1155.

37 (a) S. Yang, J. Sun, A. J. Ramirez-Cuesta, S. K. Callear, W. I. F. David, D. P. Anderson, R. Newby, A. J. Blake, J. E. Parker, C. C. Tang and M. Schröder, *Nat. Chem.*, 2012, **4**, 887–894; (b) M. Savage, Y. Cheng, T. L. Easun, J. E. Eyley, S. P. Argent, M. R. Warren, W. Lewis, C. Murray, C. C. Tang, M. D. Frogley, G. Cinque, J. Sun, S. Rudić, R. T. Murden, M. J. Benham, A. N. Fitch, A. J. Blake, A. J. Ramirez-Cuesta, S. Yang and M. Schröder, *Adv. Mater.*, 2016, **28**, 8705–8711; (c) J. A. Zárate, E. Sánchez-González, D. R. Williams, E. González-Zamora, V. Martis, A. Martínez, J. Balmaseda, G. Maurin and I. A. Ibarra, *J. Mater. Chem. A*, 2019, **7**, 15580–15584; (d) C. P. Krap, R. Newby, A. Dhakshinamoorthy, H. García, I. Cebula, T. E. Easun, M. Savage, J. E. Eyley, S. Gao, A. J. Blake, W. Lewis, P. H. Beton, M. R. Warren, D. R. Allan, M. D. Frogley, C. C. Tang, G. Cinque, S. Yang and M. Schröder, *Inorg. Chem.*, 2016, **55**, 1076; (e) J.-X. Liu, J. Li, W.-Q. Tao and Z. Li, *Fluid Phase Equilib.*, 2021, **536**, 112963.

38 I. A. Ibarra, S. Yang, X. Lin, A. J. Blake, P. J. Rizkallah, H. Nowell, D. R. Allan, N. R. Champness, P. Hubberstey and M. Schröder, *Chem. Commun.*, 2011, **47**, 8304–8306.

39 V. Chernikova, O. Yassine, O. Shekhah, M. Eddaoudi and K. N. Salama, *J. Mater. Chem. A*, 2018, **6**, 5550–5554.

40 (a) R. D. Shannon, *Acta Crystallogr. A*, 1976, **32**, 751–767; (b) L.-H. Schilling, H. Reinsch and N. Stock, in *The Chemistry of Metal-Organic Frameworks. Synthesis, Characterization, and Applications*, ed. S. Kaskel, WILEY-VCH Verlag GmbH & Co. KGaA, Germany, 2016, vol. 1, ch. 5, pp. 106–108; (c) M. G. Goesten, C. F. Guerra, F. Kapteijn, J. Gascon and F. M. Bickelhaupt, *Angew. Chem., Int. Ed.*, 2015, **54**, 12034–12038.

41 S. Kancharlapalli, S. Natarajan and T. K. Ghanty, *J. Phys. Chem. C*, 2019, **123**(45), 27531–27541.

42 (a) S. Dissegna, K. Epp, W. R. Heinz, G. Kieslich and R. A. Fischer, *Adv. Mater.*, 2018, **30**, 1704501–1717045; (b) W. Xiang, Y. Zhang, Y. Chen, C.-J. Liu and X. Tu, *J. Mater. Chem. A*, 2020, **8**, 21526–21546.

43 (a) O. Kozachuk, I. Luz, F. X. Llabrés i Xamena, H. Noei, M. Kauer, H. B. Albadia, E. D. Bloch, B. Marley, Y. Wang, M. Muhler and R. A. Fischer, *Angew. Chem., Int. Ed.*, 2014, **53**, 7058–7062; (b) J. Canivet, M. Vandichel and D. Farrusseng, *Dalton Trans.*, 2014, 4090–4099; (c) J. Wang, L. Liu, C. Chen, X. Dong, Q. Wang, L. Alfilfil, M. R. AlAlouni, K. Yao, J. Huang, D. Zhang and Y. Han, *J. Mater. Chem. A*, 2020, **8**, 4464–4472; (d) R. S. Forgan, *Chem. Sci.*, 2020, **11**, 4546–4562.

44 E. López-Maya, C. Montoro, L. M. Rodríguez-Albelo, S. D. A. Cervantes, A. A. Lozano-Pérez, J. L. Cenís, E. Barea and J. A. R. Navarro, *Angew. Chem., Int. Ed.*, 2015, **54**, 6790–6794.

45 L. M. Rodriguez-Albelo, E. Lopez-Maya, S. Hamad, A. R. Ruiz-Salvador, S. Calero and J. A. R. Navarro, *Nat. Commun.*, 2017, **8**, 14457.

46 E. López-Maya, C. Montoro, V. Colombo, E. Barea and J. A. R. Navarro, *Adv. Funct. Mater.*, 2014, **24**, 6130–6135.

47 M. Mon, E. Tiburcio, J. Ferrando-Soria, R. G. San Millan, J. A. R. Navarro, D. Armentano and E. Pardo, *Chem. Commun.*, 2018, **54**, 9063–9066.

48 (a) Y. Yuan, J. Li, X. Sun, G. Li, Y. Liu, G. Verma and S. Ma, *Chem. Mater.*, 2019, **31**, 1084–1091; (b) J. Hou,



M. L. R. Gómez, A. Krajnc, A. McCaul, S. Li, A. M. Bumstead, A. F. Sapnik, Z. Deng, R. Lin, P. A. Chater, D. S. Keeble, D. A. Keen, D. Appadoo, B. Chan, V. Chen, G. Mali and T. D. Bennett, *J. Am. Chem. Soc.*, 2020, **142**(8), 3880–3890; (c) L.-M. Yang, G.-Y. Fang, J. Ma, R. Pushpa and E. Ganz, *Phys. Chem. Chem. Phys.*, 2016, **18**, 32319–32330; (d) M. Kalaj, M. R. Momeni, K. C. Bentz, K. S. Barcus, J. M. Palomba, F. Paesani and S. M. Cohen, *Chem. Commun.*, 2019, **55**, 3481–3484.

49 A. Karmakar, A. V. Desai and S. K. Ghosh, *Coord. Chem. Rev.*, 2016, **307**, 313–341.

50 X. Suo, Y. Yu, S. Qian, L. Zhou, X. Cui and H. Xing, *Angew. Chem., Int. Ed.*, 2021, **60**, 1–7.

51 X. Cui, Q. Yang, L. Yang, R. Krishna, Z. Zhang, Z. Bao, H. Wu, Q. Ren, W. Zhou, B. Chen and H. Xing, *Adv. Mater.*, 2017, **29**, 1606929.

52 M. R. Tchalala, P. M. Bhatt, K. N. Chappanda, S. R. Tavares, K. Adil, Y. Belmabkhout, A. Shkurenko, A. Cadiou, N. Heymans, G. De Weireld, G. Maurin, K. N. Salama and M. Eddaoudi, *Nat. Commun.*, 2019, **10**, 1328.

53 T. M. M. Ntep, H. Breitzke, L. Schmolke, C. Schlüsener, B. Moll, S. Millan, N. Tannert, I. E. Aita, G. Buntkowsky and C. Janiak, *Chem. Mater.*, 2019, **31**, 8629–8638.

54 W. Jaeschke, H. Claude and J. Herrmann, *J. Geophys. Res.*, 1980, **85**, 5639–5644.

55 (a) O. A. Habeeb, R. Kanthasamy, G. A. M. Ali, S. Sethupathi and R. B. M. Yunus, *Rev. Chem. Eng.*, 2018, **34**, 837–854; (b) G. E. Likens, R. F. Wright, J. N. Galloway and T. J. Butler, *Sci. Am.*, 1979, **241**, 43–51.

56 R. O. Beauchamp, J. S. Bus, J. A. Popp, C. J. Boreiko, D. A. Andjelkovich and P. Leber, *CRC Crit. Rev. Toxicol.*, 1984, **13**, 25–97.

57 M. S. Shah, M. Tsapatsis and J. I. Siepmann, *Chem. Rev.*, 2017, **117**, 9755–9803.

58 A. Ros, M. A. Montes-Moran, E. Fuente, D. M. Nevskaia and M. J. Martin, *Environ. Sci. Technol.*, 2006, **40**, 302–309.

59 M. Bui, C. S. Adjiman, A. Bardow, E. J. Anthony, A. Boston, S. Brown, P. S. Fennell, S. Fuss, A. Galindo, L. A. Hackett, J. P. Hallett, H. J. Herzog, G. Jackson, J. Kemper, S. Krevor, G. C. Maitland, M. Matuszewski, I. S. Metcalfe, C. Petit, G. Puxty, J. Reimer, D. M. Reiner, E. S. Rubin, S. A. Scott, N. Shah, B. Smit, J. P. M. Trusler, P. Webley, J. Wilcox and N. Mac Dowell, *Energy Environ. Sci.*, 2018, **11**, 1062–1176.

60 C. Petit, B. Mendoza and T. J. Bandosz, *ChemPhysChem*, 2010, **11**, 3678–3684.

61 M. Khabazipour and M. Anbia, *Ind. Eng. Chem. Res.*, 2019, **58**, 22133–22164.

62 L. Hamon, C. Serre, T. Devic, T. Loiseau, F. Millange, G. Férey and G. De Weireld, *J. Am. Chem. Soc.*, 2009, **131**, 8775–8777.

63 J. Liu, Y. Wei, P. Li, Y. Zhao and R. Zou, *J. Phys. Chem. C*, 2017, **121**, 13249–13255.

64 Y. Belmabkhout, R. S. Pillai, D. Alezi, O. Shekhah, P. M. Bhatt, Z. Chen, K. Adil, S. Vaesen, G. De Weireld, M. Pan, M. Suetin, A. J. Cairns, V. Solovyeva, A. Shkurenko, O. E. Tall, G. Maurin and M. Eddaoudi, *J. Mater. Chem. A*, 2017, **5**, 3293–3303.

65 (a) J. Izumi, T. Morimoto, H. Tsutaya and K. Araki, *Stud. Surf. Sci. Catal.*, 1993, **80**, 293–299; (b) P. B. S. Rallapalli, K. Cho, S. H. Kim, J.-N. Kim and H. C. Yoon, *Fuel*, 2020, **281**, 118985.

66 P. Lyn and G. Maurin, *ACS Appl. Mater. Interfaces*, 2021, **13**, 4813–4822.

67 L. Hamon, C. Serre, T. Devic, T. Loiseau, F. Millange, G. Férey and G. De Weireld, *J. Am. Chem. Soc.*, 2009, **131**, 8775–8777.

68 E. Sánchez-González, P. G. M. Mileo, M. Sagastuy-Breña, J. R. Álvarez, J. E. Reynolds III, A. Villareal, A. Gutiérrez-Alejandre, J. Ramírez, J. Balmaseda, E. González-Zamora, G. Maurin, S. M. Humphrey and I. A. Ibarra, *J. Mater. Chem. A*, 2018, **6**, 16900–16909.

69 J. A. Zárate, E. Sánchez-González, T. Jurado-Vázquez, A. Gutiérrez-Alejandre, E. González-Zamora, I. Castillo, G. Maurin and I. A. Ibarra, *Chem. Commun.*, 2019, **55**, 3049–3052.

70 J. G. Flores, J. A. Zárate-Colín, E. Sánchez-González, J. R. Valenzuela, J. A. Gutiérrez-Alejandre, J. Ramírez, V. Jancik, J. Aguilar-pliego, M. C. Zorrilla, H. A. Lara-García, E. González-Zamora, G. Guzmán-González, I. González, G. Maurin and I. A. Ibarra, *ACS Appl. Mater. Interfaces*, 2020, **12**, 18885–18892.

71 Y. V. Mikhaylik and J. R. Akridge, *J. Electrochem. Soc.*, 2004, **151**, A1969–A1976.

72 D. Moy, A. Manivannan and S. R. Narayanan, *J. Electrochem. Soc.*, 2015, **162**, A1–A7.

73 A. B. Andreeva, K. N. Le, L. Chen, M. E. Kellman, C. H. Hendon and C. K. Brozek, *J. Am. Chem. Soc.*, 2020, **142**, 19291–19299.

74 E. S. Grape, J. G. Flores, T. Hidalgo, E. Martínez-Ahumada, A. Hautier, D. R. Williams, M. O. Kee, O. Lars, T. Willhammar, P. Horcajada, I. A. Ibarra and A. K. Inge, *J. Am. Chem. Soc.*, 2020, **142**, 16795–16804.

75 H. Wang, X. Zeng, W. Wang and D. Cao, *Chem. Eng. Sci.*, 2015, **135**, 373–380.

76 K. Polychronopoulou, J. L. G. Fierro and A. M. Efstathiou, *Appl. Catal., B*, 2005, **57**, 125–137.

77 X. Liu and R. Wang, *J. Hazard. Mater.*, 2017, **326**, 157–164.

78 L. Sigot, G. Ducom and P. Germain, *Chem. Eng. J.*, 2016, **287**, 47–53.

79 X. Wang, X. Ma, X. Xu, L. Sun and C. Song, *Top. Catal.*, 2008, **49**, 108–117.

80 X. Ma, X. Wang and C. Song, *J. Am. Chem. Soc.*, 2009, **131**, 5777–5783.

81 F. Adib, A. Bagreev and T. J. Bandosz, *Langmuir*, 2000, **16**(4), 1980–1986.

82 H. Dong, Q. Zhou, L. Zhang and Y. Tian, *Angew. Chem., Int. Ed.*, 2019, **58**, 13948–13953.

83 R. Motterlini and L. E. Otterbein, *Nat. Rev. Drug Discovery*, 2010, **9**, 728–743.

84 G. Yang, L. Wu, B. Jiang, W. Yang, J. Qi, K. Cao, Q. Meng, A. K. Mustafa, W. Mu, S. Zhang, S. H. Snyder and R. Wang, *Science*, 2008, **322**, 587–590.



85 D. J. Polhemus and D. J. Lefer, *Circ. Res.*, 2014, **114**, 730–737.

86 M. N. Hughes, M. N. Centelles and K. P. Moore, *Free Radical Biol. Med.*, 2009, **47**, 1346–1353.

87 C. Szabo and A. Papapetropoulos, *Pharmacol. Rev.*, 2017, **69**, 497–564.

88 H. A. Henthorn and M. D. Pluth, *J. Am. Chem. Soc.*, 2015, **137**(48), 15330–15336.

89 D.-W. Li, L.-L. Qu, K. Hu, Y.-T. Long and H. Tian, *Angew. Chem.*, 2015, **127**, 12949–12952.

90 C. Yang, X. Yin, S.-Y. Huan, L. L. Chen, X.-X. Hu, M.-Y. Xiong, K. Chen and X.-B. Zhang, *Anal. Chem.*, 2018, **90**, 3118–3123.

91 J. Yang and Y. W. Yang, *Small*, 2020, **16**, 1906846.

92 P. K. Allan, P. S. Wheatley, D. Aldous, M. Infas, C. Tang, J. A. Hriljac, I. L. Megson, K. W. Chapman, G. D. Weireld, S. Vaesen and R. E. Morris, *Dalton Trans.*, 2012, **41**, 4060–4066.

93 H. Li, X. Feng, Y. Guo, D. Chen, R. Li, X. Ren, X. Jiang, Y. Dong and B. Wang, *Sci. Rep.*, 2014, **4**, 4366.

94 Y. Ma, H. Su, X. Kuang, X. Li, T. Zhang and B. Tang, *Anal. Chem.*, 2014, **86**, 11459–11463.

95 S. Nandi, S. Banesh, V. Trivedi and S. Biswas, *Analyst*, 2018, **143**, 1482–1491.

96 X. Zhang, L. Fang, K. Jiang, H. He, Y. Yang, Y. Cui, B. Li and G. Qian, *Biosens. Bioelectron.*, 2019, **130**, 65–72.

97 C. Yang, K. Chen, M. Chen, X. Hu, S. Y. Huan, L. Chen, G. Song and X. B. Zhang, *Anal. Chem.*, 2019, **91**, 2727–2733.

98 Y. Li, J. Zhou, L. Wang and Z. Xie, *ACS Appl. Mater. Interfaces*, 2020, **12**, 30213–30220.

99 C. D. Demirhan, W. W. Tso, J. B. Powell and E. N. Pistikopoulos, *AIChE J.*, 2019, **65**, e16498.

100 (a) A. Fangmaier, A. Hadwiger-Fangmaier, L. Van der Eerden and H. J. Jäger, *Environ. Pollut.*, 1994, **86**, 43–82; (b) S. N. Behera, M. Sharma, V. P. Aneja and R. Balasubramanian, *Environ. Sci. Pollut. Res.*, 2013, **20**, 8092–8131.

101 R. Metkemeijer and P. Achard, *J. Power Sources*, 1994, **49**, 271–282.

102 Z. Tamainot-Telto, S. J. Metcalf, R. E. Critoph, Y. Zhong and R. Thorpe, *Int. J. Refrig.*, 2009, **32**, 1212–1229.

103 (a) C. Petit, B. Mendoza and T. J. Bandosz, *Langmuir*, 2010, **26**, 15302–15309; (b) M. J. Katz, A. J. Howarth, P. Z. Moghadam, J. B. DeCoste, R. Q. Snurr, J. T. Hupp and O. K. Farha, *Dalton Trans.*, 2016, **45**, 4150–4153.

104 D. W. Kang, S. E. Ju, D. W. Kim, M. Kang, H. Kim and C. S. Hong, *Adv. Sci.*, 2020, **16**, 2002142.

105 J. Helminen, J. Helenius, E. Paatero and I. Turunen, *J. Chem. Eng. Data*, 2001, **46**, 391–399.

106 (a) C. Petit and T. J. Bandosz, *Adv. Funct. Mater.*, 2010, **20**, 111–118; (b) D. Britt, D. Tranchemontagne and O. M. Yagui, *Proc. Natl. Acad. Sci. U. S. A.*, 2008, **105**, 11623–11627.

107 A. J. Rieth, Y. Tulchinsky and M. Dincă, *J. Am. Chem. Soc.*, 2016, **138**, 9401–9404.

108 A. J. Rieth and M. Dincă, *J. Am. Chem. Soc.*, 2018, **140**, 3461–3466.

109 D. W. Kim, D. W. Kang, M. Kang, J.-H. Lee, J. H. Choe, Y. S. Chae, D. S. Choi, H. Yun and C. S. Hong, *Angew. Chem.*, 2020, **59**, 22531–22536.

110 H. G. W. Godfrey, I. da Silva, L. Briggs, J. H. Carter, C. G. Morris, M. Savage, T. L. Easun, P. Manuel, C. A. Murray, C. C. Tang, M. D. Frogley, G. Cinque, S. Yang and M. Schröder, *Angew. Chem., Int. Ed.*, 2018, **57**, 14778–14781.

111 X. Han, W. Lu, Y. Chen, I. Silva, J. Li, L. Lin, W. Li, A. M. Sheveleva, H. G. W. Godfrey, Z. Lu, F. Tuna, E. J. L. McInnes, Y. Cheng, L. L. Daemen, L. J. M. McPherson, S. J. Teat, M. D. Frogley, S. Rudic, P. Manuel, A. J. Ramirez-cuesta, S. Yang and M. Schröder, *J. Am. Chem. Soc.*, 2021, **143**(8), 3153–3161.

112 T. Yoskamtorn, P. Zhao, X.-P. Wu, K. Purchase, F. Orlandi, P. Manuel, J. Taylor, Y. Li, S. Day, L. Ye, C. C. Tang, Y. Zhao and S. C. E. Tsang, *J. Am. Chem. Soc.*, 2021, **143**(8), 3205–3218.

113 P. Zhao, H. Fang, S. Mukhopadhyay, A. Li, S. Rudic, I. J. McPherson, C. C. Tang, D. Fairen-Jimenez, S. C. E. Tsang and S. A. T. Redfern, *Nat. Commun.*, 2019, **10**, 999.

114 F. Gholami, M. Tomas, Z. Gholami and M. Vakili, *Sci. Total Environ.*, 2020, **714**, 136712.

115 D. Krewski, M. Jerrett, R. T. Burnett, R. Ma, E. Hughes, Y. Shi, M. C. Turner, C. A. Pope, G. Thurston, E. E. Calle, M. J. Thun, B. Beckerman, P. DeLuca, N. Finkelstein, K. Ito, D. K. Moore, K. B. Newbold, T. Ramsay, Z. Ross, H. Shin and B. Tempalski, *Res. Rep. Health Eff. Inst.*, 2009, 5–114; discussion 115–136.

116 S. Safieddine, C. Clerbaux, M. George, J. Hadji-Lazaro, D. Hurtmans, P.-F. Coheur, C. Wespes, D. Loyola, P. Valks and N. Hao, *J. Geophys. Res. Atmos.*, 2013, **118**, 10 555–10 566.

117 F. Rezaei, A. A. Rownaghi, S. Monjezi, R. P. Lively and C. W. Jones, *Energy Fuels*, 2015, **29**, 5467–5486.

118 J. Despres, M. Koebel, O. Kröcher, M. Elsener and A. Wokaun, *Microporous Mesoporous Mater.*, 2003, **58**, 175–183.

119 X. Chang, G. Lu, Y. Guo, Y. Wang and Y. Guo, *Microporous Mesoporous Mater.*, 2013, **165**, 113–120.

120 W. J. Zhang, S. Rabiei, A. Bagreev, M. S. Zhuang and F. Rasouli, *Appl. Catal. B*, 2008, **83**, 63–71.

121 N. Gargiulo, A. Peluso and D. Caputo, *Processes*, 2020, **8**, 613.

122 J.-R. Li, R. J. Kuppler and H.-C. Zhou, *Chem. Soc. Rev.*, 2009, **38**, 1477–1504.

123 G. W. Peterson, J. J. Mahle, J. B. DeCoste, W. O. Gordon and J. A. Rossin, *Angew. Chem., Int. Ed.*, 2016, **55**, 6235–6238.

124 X. Han, H. G. W. Godfrey, L. Briggs, A. J. Davies, Y. Cheng, L. L. Daemen, A. M. Sheveleva, F. Tuna, E. J. L. McInnes, J. Sun, C. Drathen, M. W. George, A. J. Ramirez-Cuesta, K. M. Thomas, S. Yang and M. Schröder, *Nat. Mater.*, 2018, **17**, 691–696.

125 X. Han, Y. Hong, Y. Ma, W. Lu, J. Li, L. Lin, A. M. Sheveleva, F. Tuna, E. J. L. McInnes, C. Dejoie, J. Sun, S. Yang and M. Schröder, *J. Am. Chem. Soc.*, 2020, **142**, 15235–15239.



126 A. M. Wright, C. Sun and M. Dincă, *J. Am. Chem. Soc.*, 2021, **143**, 681–686.

127 B. Xiao, P. S. Wheatley, X. Zhao, A. J. Fletcher, S. Fox, A. G. Rossi, I. L. Megson, S. Bordiga, L. Regli, K. M. Thomas and R. E. Morris, *J. Am. Chem. Soc.*, 2007, **129**, 1203–1209.

128 S. M. Vornholt, M. J. Duncan, S. J. Warrender, R. Semino, N. A. Ramsahye, G. Maurin, M. W. Smith, J.-C. Tan, D. N. Miller and R. E. Morris, *ACS Appl. Mater. Interfaces*, 2020, **12**, 58263–58276.

129 A. C. McKinlay, B. Xiao, D. S. Wragg, P. S. Wheatley, I. L. Megson and R. E. Morris, *J. Am. Chem. Soc.*, 2008, **130**, 10440–10444.

130 (a) M. O. Andreae and P. Merlet, *Global Biogeochem. Cycles*, 2001, **15**, 955–966; (b) J.-Y. Lee, T. C. Keener and Y. J. Yang, *J. Air Waste Manage. Assoc.*, 2009, **59**, 725–732.

131 (a) G. V. Last and M. T. Schmick, *Identification and Selection of Major Carbon Dioxide Stream Compositions*, US Department of Energy, Pacific Northwest National Laboratory, Washington U. S. A., 2011; (b) R. Zevenhoven and P. Kilpinen, in *Control of Pollutants in Flue Gases and Fuel Gases*, Helsinki Univ. of Technology, Otaniemi (Finland). Energy Engineering and Environmental Protection, Finland, 2005, ch. 2, pp. 2–4.

132 J. H. Carter, X. Han, F. Y. Moreau, I. da Silva, A. Nevin, H. G. W. Godfrey, C. C. Tang, S. Yang and M. Schröder, *J. Am. Chem. Soc.*, 2018, **140**, 15564–15567.

133 Y. Zhang, P. Zhang, W. Yu, J. Zhang, J. Huang, J. Wang, M. Xu, Q. Deng, Z. Zeng and S. Deng, *ACS Appl. Mater. Interfaces*, 2019, **11**(11), 10680–10688.

134 Y. L. Fan, H. P. Zhang, M. J. Yin, R. Krishna, X. F. Feng, L. Wang, M. B. Luo and F. Luo, *Inorg. Chem.*, 2021, **60**(1), 4–8.

135 L. J. Gou, X. F. Feng, Z. Gao, R. Krishna and F. Luo, *Inorg. Chem.*, 2021, **60**(3), 1310–1314.

136 M. J. Yin, X. H. Xiong, X. F. Feng, W. Y. Xu, R. Krishna and F. Luo, *Inorg. Chem.*, 2021, **60**(5), 3447–3451.

137 Y. Belmabkhout, P. M. Bhatt, K. Adil, R. S. Pillai, A. Cadiau, A. Shkurenko, G. Maurin, G. Liu, W. J. Koros and M. Eddaoudi, *Nat. Energy*, 2018, **3**, 1059–1066.

138 S. Vaesen, V. Guillerm, Q. Yang, A. D. Wiersum, B. Marszalek, B. Gil, A. Vimont, M. Daturi, T. Devic, P. L. Llewellyn, C. Serre, G. Maurin and G. De Weireld, *Chem. Commun.*, 2013, **49**, 10082–10084.

139 P. M. Bhatt, Y. Belmabkhout, A. H. Assen, L. J. Weselinski, H. Jiang, A. Cadiau, D.-X. Xue and M. Eddaoudi, *Chem. Eng. J.*, 2017, **324**, 392–396.

140 Z. Li, Y. Xiao, W. Xue, Q. Yang and C. Zhong, *J. Phys. Chem. C*, 2015, **119**(7), 3674–3683.

141 E. N. H. S. Vaese and G. De Weireld, *Microporous Mesoporous Mater.*, 2012, **154**, 93–99.

142 J. Xu, W. Xing, H. Wang, W. Xu, Q. Ding, L. Zhao, W. Guo and Z. Yan, *J. Mater. Sci.*, 2016, **51**, 2307–2319.

143 A. G. P. Maloney, P. A. Wood, S. C. Ward and D. Fairen-Jimenez, *Chem. Mater.*, 2017, **29**, 2618–2625.

144 D. Ongari, L. Talirz and B. Smit, *ACS Cent. Sci.*, 2020, **6**(11), 1890–1900.

145 A. J. Howarth, Y. Liu, P. Li, Z. Li, T. C. Wang, T. T. Hupp and O. K. Farha, *Nat. Rev. Mater.*, 2016, **1**, 15018.

