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Magnetic Fe–N–C nanoparticles as a dual nanozyme for label-free colorimetric detection of antibiotics†

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In this study, we report a simple pyrolytic strategy for N–C co-doped Fe-based nanoparticles (Fe–N–C) and exhibit their oxidase-mimicking and peroxidase-mimicking activity in a chromogenic reaction with 3,3',5,5'-tetramethylbenzidine dihydrochloride (TMB) as a substrate. Structural analysis revealed that Fe–N_x and Fe₃C structures were formed with Fe-Imace coordination compounds as precursors and melamine as the nitrogen source, which provided abundant active sites for the Fe–N–C nanozyme. Interestingly, owing to the blocking effect triggered by π – π stacking between the tetracycline antibiotics (TCs) tetraphenyl skeleton and Fe–N–C, the substrate affinity of the Fe–N–C was significantly blocked, resulting in the solution fading. Under the optimum conditions, the UV absorption intensity *versus* the concentration of TCs was found to be linear over the range of 0.08–90 μ M for tetracycline (TC), 0.1–80 μ M for chlortetracycline (CTC) and 0.09–100 μ M for oxytetracycline (OTC). The detection limits were 62 nM for TC, 85 nM for CTC and 88 nM for OTC. The Fe–N–C displayed good storage and long-term stability when stored at 4 °C for 30 days. In addition, based on its superparamagnetism, Fe–N–C can be effectively recycled and reused through an external magnetic field, avoiding secondary pollution.

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Environmental significance

Today, the presence of tetracycline antibiotics (TCs) in aquatic environments has drawn great attention from environmental scientists. The establishment of fast and convenient detection methods is one of the first problems to be solved in environmental analysis. Therefore, nanozymes are deemed ideal materials for constructing colorimetric sensing platforms for detection and analysis of antibiotics and have important environmental significance.

1 Introduction

Tetracycline antibiotics (TCs) are a class of broad-spectrum antibiotics that include tetracycline (TC), chlortetracycline (CTC), oxytetracycline (OTC), *etc.* (Fig. 1). Due to their good bactericidal effect, they are typically used to treat human diseases or in livestock husbandry. However, improper or excessive use of TCs can lead to high levels of drug residues, posing a serious threat to the environment and human health. To date, various convenient and accurate techniques have been applied to TCs monitoring. Instrumental analysis such as liquid

chromatography-mass spectrometry (LC/MS),¹ high performance liquid chromatography (HPLC)^{2,3} and capillary electrophoresis (CE)⁴ showed a high degree of automation, as well as satisfactory sensitivity and accuracy. However, the above techniques require long analysis time and expensive instruments, which make it difficult to achieve real-time analysis of large-scale samples. In addition, electrochemical,⁵ fluorescent⁶ and chemiluminescent (CL)⁷ sensors have been developed for the



Fig. 1 Chemical structure of (A) tetracycline, (B) oxytetracycline, and (C) chlortetracycline.

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2.2 Synthesis of Fe–N–C

Fe–N–C was synthesized based on Li's work with minor modifications.²² Briefly, 3.36 g Imace and 2.32 g (CH₃COO)₂Fe were added to 40 mL deionised water and stirred continuously for 6 h at 70 °C. Subsequently, 67 mL deionised water and 5.33 g melamine were added for ultrasonic treatment. After being stirred at room temperature for 6 h, the above solution was dried at 70 °C for 15 h, and a brown powder was obtained. To synthesize Fe–N–C, the product was first placed in a ceramic vessel and the tube furnace was heated to 550 °C at a programmed rate of 3 °C min⁻¹, and kept for 2 h at 550 °C. Then, the temperature was raised to 700, 800 and 900 °C and maintained for 2 h with a ramp rate of 3 °C min⁻¹, respectively. All the above pyrolysis processes were carried out in a nitrogen atmosphere. After cooling to room temperature, the black products were leached in 1 M HCl solution and deionised water three times to remove the unstable iron nanoparticles, then dried overnight at 70 °C. The obtained black materials were denoted as Fe–N–C-700, Fe–N–C-800 and Fe–N–C-900 respectively to represent pyrolysis temperature. The synthesis process of N–C was similar to that of the Fe–N–C, except that no (CH₃COO)₂Fe was added. The black powders obtained were named Fe–N–C-*X*, where “*X*” indicates the pyrolysis temperature.

2.3 Dual-response enzyme mimicking activity of Fe–N–C

The OXD-mimicking and POD-mimicking activities of Fe–N–C were assessed by OPD, ABTS and TMB as reaction substrates in a colorimetric assay. In a typical experiment, 80 μL of 10 mM TMB, 80 μL of 0.1 M H₂O₂ (present or absent) and 9 μL of 1 mg mL⁻¹ Fe–N–C nanozyme aqueous dispersion were mixed with 1.5 mL HAc–NaAc buffer solution (pH = 4). The color variation of the mixed solution and the absorbance change at 652 nm in a UV-vis spectrophotometer was monitored at room temperature over 17 minutes.

2.4 Steady-state kinetic analysis

The steady-state kinetic measurements were carried out in time-course mode, and all measurements were based on the absorbance change of ox-TMB at 652 nm. The absorbance coefficient of the product (ox-TMB) at 652 nm is 39 000 M⁻¹ cm⁻¹. To determine the initial reaction rate, different concentrations of H₂O₂ were reacted with Fe–N–C at a certain concentration of TMB, and *vice versa*. The rate of reaction is the slope of the absorbance changes with time. The kinetic parameters of enzyme activity were calculated using the Lineweaver–Burk plot of the double reciprocal of the Michaelis–Menten equation: $1/v = K_m/V_{max}(1/[S] + 1/K_m)$, where V , V_{max} , K_m and $[S]$ represent the initial velocity of the reaction, the maximal reaction velocity, the Michaelis–Menten constant and the substrate concentration, respectively.

2.5 Mechanism analysis

Different active species, including O²⁻, ·OH and ¹O₂, are captured through various scavengers. Fe–N–C, H₂O₂, TMB, p-BQ, isopropanol and L-histidine were put into HAc–NaAc

buffer (pH = 4) with a total volume of 1.5 mL under the optimum experimental conditions. The absorbance (652 nm) of the system was recorded at ambient temperature after 17 min.

All simulations were performed using first principles calculations based on density functional theory (DFT) within the generalized gradient approximation (GGA) using the Perdew–Burke–Ernzerhof (PBE) formulation with Grimme's energy method in the Dmol3 module. All the geometry optimizations were performed using the Broyden–Fletcher–Goldfarb–Shanno (BFGS) algorithm. The *k*-points were set at gamma (1 × 1 × 1) and the convergence tolerance was set at 1.0 × 10⁻⁵ Ha, 2.0 × 10⁻³ Ha Å, and 5.0 × 10⁻³ Å for energy, maximum force, and maximum displacement, respectively.

2.6 TCs detection

In a typical assay of antibiotics, 9 μL of 1 mg mL⁻¹ Fe–N–C aqueous dispersion, 80 μL of 10 mM TMB, 80 μL of 0.1 M H₂O₂ and gradient concentrations of antibiotics were mixed in 1.5 mL HAc–NaAc buffer solution (pH = 4) for 17 minutes at room temperature, and the absorbance of the mixed solution was recorded at 652 nm.

2.7 Detection of antibiotics in actual samples

Lake water samples were collected from Shanxi University and filtered through a 0.22 μm microporous filter. The subsequent detection procedures were similar to those described previously.

3 Results and discussion

3.1 Characterization of Fe–N–C

The micromorphology and microstructure of the prepared Fe–N–C were characterized by SEM and TEM. As depicted in Fig. 2A, the SEM micrograph clearly shows that the Fe–N–C has uniform polyhedral morphology and is stacked up to some extent. Fig. S1† shows the micromorphology of Fe–N–C-*X* at different calcination temperatures. It can be seen that the particle size of Fe–N–C-*X* decreased with the increase of calcination temperature, and the smaller particle size indicates a higher catalytic efficiency to some extent. More importantly, higher calcination temperatures lead to a higher degree of graphitization. As seen from Fig. 2B–D, the TEM images show that lots of tiny Fe based nanoparticles in the form of hyperfine Fe₃C nanoparticles are evenly distributed on or in the carbon layers, which is due to the fact that Fe₃C nanoparticles can effectively promote the graphitization of the surrounding C atoms. The Fe₃C nanoparticles coated by graphitic carbon layers can not only enrich the electron density on the carbon surface, but also avoid the dissolution and agglomeration caused by direct contact between Fe₃C and the external medium, ensuring favorable catalytic and stability activity of Fe–N–C. Furthermore, the elemental mapping images validate the uniform dispersion of C, Fe, N and O elements in the carbon matrix (Fig. 2E–I). These results indicate that Fe–N co-doped carbon structure catalysts can be easily synthesized by pyrolysis.



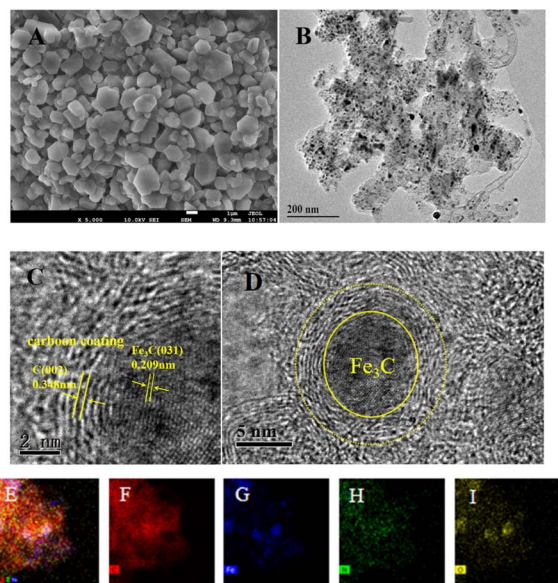


Fig. 2 (A) SEM image of Fe-N-C-900; (B) TEM and (C and D) HR-TEM images of Fe-N-C-900; (E-I) the TEM-EDX mapping images of C, Fe, N, and O elements.

The chemical composition and elemental binding states of Fe-N-C-X were further characterized by XPS. As illustrated in Fig. 3A, Fe, C, N and O can be clearly identified from XPS spectra. This result matched well with the mapping result. According to the integrated peak areas, the atomic ratios of Fe, C, N and O on the surfaces of Fe-N-C-X are listed in Table S1.† The results showed that the atomic percentage of Fe increased gradually with increasing calcination temperatures. However, the N atomic percentage showed the opposite trend. Among them, the low Fe content may be due to the fact that Fe₃C nanoparticles are wrapped in the graphitized carbon layer and cannot be detected. The N 1s emission spectra (Fig. 3B) of Fe-N-

C can be fitted as four peaks at 398.4 eV, 399.7 eV, 401.2 eV and 402.5 eV, corresponding to pyridinic N, pyrrolic N, graphitic N and oxidized N, respectively. Moreover, although the N atomic percentage decreased with the increase of calcination temperature, the ratio of graphitic N varied from 14.12% to 23.30%, indicating that the graphitization degree of Fe-N-C catalysts was enhanced by increasing the temperature. The results are in good agreement with the TEM results. It has been proved that the pyridinic N and graphitic N are rich in active sites and thus accelerate the ORR process by facilitating the adsorption of O₂ and the breaking of O-O bonds.²³ Besides, pyridinic N could link with Fe atoms to form Fe-N_x active centers, which further improves oxygen reduction performance because they are conducive to the initial adsorption of oxygen.²⁴ Furthermore, the high resolution Fe 2p spectra of Fe-N-C-X are shown in Fig. 3C, which can be fitted to two pairs of peaks (Fe 2p_{3/2} and Fe 2p_{1/2}). Among them, the peaks at the binding energies of 711.3 eV and 713.6 eV are attributed to the Fe 2p_{3/2} orbitals of Fe(II) and Fe(III), and the peaks at the binding energies of 720.5 eV and 724.6 eV are attributed to the Fe 2p_{1/2} orbitals of Fe(II) and Fe(III), indicating the co-existence of Fe(II) and Fe(III). Abundant Fe(II) species are beneficial to the generation of the Fe-N_x configuration with high catalytic activity. Meanwhile, the peak of Fe₃C is observed at the binding energy of 706.8 eV, which is consistent with TEM results. According to the above XPS analysis, the prepared Fe-N-C is rich in active sites due to the coexistence of Fe₃C, pyridinic N (Fe-N_x) and graphitic N, which plays a positive role in improving the catalytic activity of the catalyst.

To determine the structure of the Fe-N-C, the XRD spectrum was measured and is presented in Fig. 3D. In the XRD pattern, the strong peak near 26.0° and the weak peak around 75.0° both correspond to the typical (002) lattice plane of graphitic carbon, while the diffraction peaks at 43.7° and 51.0° are attributed to Fe₃C, and the remaining diffraction peaks are attributed to metallic iron. The XRD results show that the Fe-N-C is crystalline, and demonstrate the co-existence of graphitic carbon and iron carbide in the Fe-N-C, which confirmed the TEM results again. Previous studies have suggested that the most significant parameters affecting the catalytic efficiency and stability are crystallinity and structure.²⁵ SEM micrographs have revealed that Fe-N-C has a clear boundary, neat arrangement and sharp edges. At the same time, the good crystal structure of Fe-N-C is conducive to the improvement of catalytic activity.

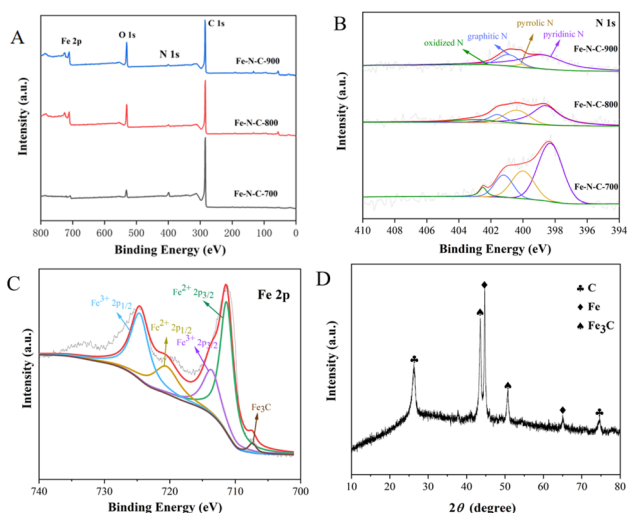


Fig. 3 (A) XPS survey spectra of Fe-N-C-X; (B and C) XPS spectra of N 1s and Fe 2p; (D) XRD pattern of Fe-N-C.

3.2 Dual response enzyme-mimicking activity

OPD, ABTS and TMB are commonly used as chromogenic substrates to demonstrate oxidase and peroxidase activities in nanozyme. To assess the catalytic activities of Fe-N-C, the experiments of catalytic substrate oxidation were carried out in the presence and absence of H₂O₂, respectively. We found that the Fe-N-C can catalyze the oxidation of OPD, ABTS and TMB by dissolved oxygen in the absence of H₂O₂, and the above solutions eventually changed from colorless to typical yellow, green and blue (Fig. 4A). Fig. S2† displays the UV absorption spectrum of TMB catalyzed oxidation, with the maximum





Fig. 4 (A) Dual response enzyme-mimicking activity; (B) effect of materials on the catalytic efficiency of Fe-N-C; (C) time-dependent absorbance changes of the reaction system for different Fe-N-C concentrations; (D) effect of pH on the catalytic efficiency of Fe-N-C.

absorption peak at approximately 652 nm. The experimental results support our initial presumption that the Fe-N-C has OXD-mimicking activity. In contrast, with the participation of H₂O₂, the color variations were much faster and showed darker colors at the same time intervals. The results confirmed that the Fe-N-C exhibits intrinsic POD-mimicking properties. An additional control experiment without Fe-N-C showed no color variation, which further proved the bifunctional enzyme-like catalytic activities of Fe-N-C.

The effect of materials on the catalytic activity of TMB was investigated. Fig. 4B shows that the Fe-N-C/TMB/H₂O₂ system exhibits stronger absorption at 652 nm than the N-C/TMB/H₂O₂ system, indicating the excellent enzyme-mimicking activity of Fe-N-C and the importance of Fe active sites. By comparison, we also evaluated the catalytic activity of Fe-N-C synthesized at different temperatures for TMB, respectively. It can be observed that with the increase of pyrolysis temperature from 700 °C to 900 °C, the catalytic activity of Fe-N-C for TMB is enhanced. This is because high temperature increases the value of the specific surface area and more active sites are exposed after pickling, which is favorable for efficient oxygen transport and rapid catalytic reaction. In addition, it has been confirmed that high temperature is conducive to nucleation of nanoparticles, while low temperature is conducive to nuclei growth. An increase of synthesis temperature results in small-sized nanoparticles. Taking this into account as well as the results of both TEM and XPS, Fe-N-C-900 was selected as an optimal material for subsequent analyses.

Similar to the natural enzyme, the catalytic activity of Fe-N-C is influenced by many external factors in practical measurements. Therefore, we optimized the experimental conditions, including the reaction time, the amount of Fe-N-C added and the pH of the HAC-NaAc buffer solution. The reaction time ranged from 0 to 20 minutes, and the dark blue solution appeared at 17 minutes (Fig. 4C). After 17 minutes, the intensity of the characteristic peak remained almost constant, demonstrating that the TMB molecules were basically completely

oxidized. Considering the practicability and operability, the addition amount of Fe-N-C was set at 9 μg for subsequent analysis. Similar to natural enzymes, Fe-N-C exhibited highly pH-dependent properties. For the tests to determine the pH effect, the pH value of HAC-NaAc buffer was adjusted from 2 to 6, and the results showed that pH = 4 was the optimal pH for this catalytic reaction, which was similar to the results of previous studies (Fig. 4D).²⁶ Some publications have confirmed that a slightly acidic solution is conducive to the decomposition of H₂O₂ that generates the ·OH radical, but H₂O₂ becomes unstable at high pH and easily decomposes into H₂O and O₂.²⁷

3.3 Kinetic analysis of Fe-N-C

Steady-state kinetic analysis of Fe-N-C with enzyme-like activity was performed by adjusting one substrate concentration while keeping the other substrate concentration fixed. The experimental data were obtained by detecting absorbance at 652 nm, and the Michaelis-Menten curves (Fig. 5) were obtained. The molar extinction coefficient of ox-TMB in the Beer-Lambert law equation was used to calculate the variation of absorbance per unit time.

$$A_{652} = \varepsilon_{\text{ox-TMB}} \times c \times l$$

where A_{652} represents the absorbance of ox-TMB at 652 nm, $\varepsilon_{\text{ox-TMB}}$ represents the molar extinction coefficient ($39\,000\text{ M}^{-1}\text{ cm}^{-1}$) and l represents the path length of the sample in the cuvette (1 cm). Subsequently, the data were fitted to the Lineweaver-Burk plot (insets in Fig. 5A and B) to obtain the kinetic parameters such as Michaelis-Menten constant (K_m) and maximum initial velocity (V_{max}). Among them, K_m is an indicator of enzyme affinity to the substrate, and the smaller the K_m value is, the higher the affinity of the enzyme to the substrate is. Horseradish peroxidase (HRP) with high peroxidase-mimicking activity was used as the reference. As shown in Table S2,† the K_m values of Fe-N-C with H₂O₂ and TMB as the substrate are 0.2 and 0.124 respectively, both lower than that of HRP, illustrating Fe-N-C had a higher affinity and activity for H₂O₂ and TMB. That is to say, H₂O₂ is easily adsorbed on the surface of Fe-N-C and then rapidly decomposed by Fe-N-C. Meanwhile, the V_{max} value of Fe-N-C was significantly higher than that of HRP. These results suggest that the high enzyme-like activity of Fe-N-C is derived from the superior affinity for the substrates.

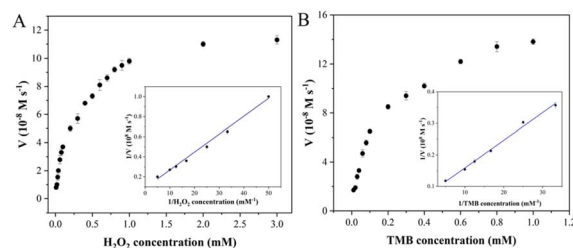


Fig. 5 Steady-state kinetic analysis utilizing the Michaelis-Menten model (A and B) and Lineweaver-Burk double-reciprocal model for Fe-N-C at pH 4.0.



3.4 Colorimetric detection of TCs with Fe–N–C

On the basis of the excellent enzyme-mimicking activity of the Fe–N–C, a colorimetric method for TCs detection was established. As illustrated in Scheme 1, the Fe–N–C is first added to an HAc–NaAc buffer containing TMB and H₂O₂, and the solution gradually changed from colorless to a typical blue. Next, in the presence of TCs, π – π stacking was formed between the TCs tetraphenyl skeleton and Fe–N–C, resulting in the catalytic blocking effect of the material on the substrate. Eventually, the solution faded and the absorption intensity of ox-TMB decreased at 652 nm.

To evaluate the performance of the colorimetric method for TCs detection, the UV-vis absorbance of ox-TMB at 652 nm was measured with different concentrations of TCs. As shown in Fig. 6 and S3,[†] the colorimetric biosensor had similar responses to the three TCs due to their similar chemical structures, and the UV absorbance at 652 nm decreased gradually as TCs concentration increased. There was a linear relationship between the ΔA value ($\Delta A = A_0 - A$, where A_0 and A represent the absorbance intensities of ox-TMB at 652 nm in the absence and presence of TC, respectively) and TC concentrations in the range of 0.08–10 μM and 10–90 μM (Fig. 6A). Color variation with the increase of TC dosage could also be visualized by the naked eye, gradually going from blue to colorless (Fig. S4[†]). And the limit of detection (LOD) for TC calculated as $3\sigma/\text{slope}$ was 62 nM. The linear range of UV absorbance measured by CTC was found to be 0.1–20 μM and 20–80 μM while the LOD was 85 nM. And the OTC detection as a function of concentration was 0.09–10 μM and 10–100 μM while the LOD was 88 nM. The above results strongly proved that the colorimetric biosensor has a satisfactory and wide linear range in the case of low TCs concentration, which provides a possibility for its practical application. Compared with other detection methods, the performance of the Fe–N–C biosensor has comparable or better features. Although some reported methods showed higher sensitivity or a wider linear range, our proposed method has a simpler analytical process, lower cost and reliable analytical results (Table S3[†]).

In actual sample detection, the high selectivity is a key factor for assessing the feasibility of the analytical detection platform. The selectivity of TCs colorimetric sensors was evaluated using cations (Zn²⁺, Mg²⁺, K⁺ and Na⁺), anions (SO₄²⁻, Cl⁻), common antibiotics (DOX, AML, NFZ, KAN, STR and CPL) and TCs (TC,

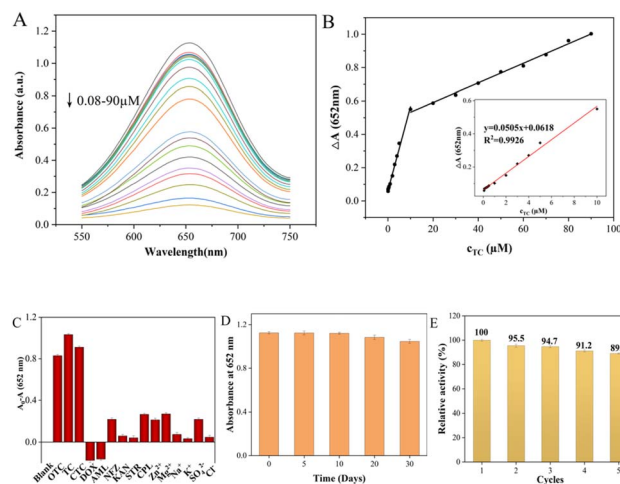
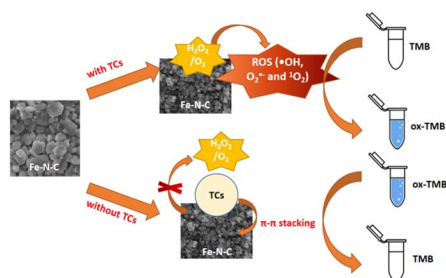


Fig. 6 (A) UV-vis absorption spectra of this Fe–N–C nanozyme sensing system following incubation with TC in the range from 0.08 to 90.0 μM . (B) Calibration curve of absorbance variations versus quantities of TC added. (C) Selectivity of the Fe–N–C nanozyme sensing system for TC over ions and other antibiotics. (D) Long-time stability of the Fe–N–C nanozyme sensing system. (E) Five consecutive catalysis cycles of Fe–N–C.

CTC and OTC) under the same conditions. The error bar represents the standard deviation of three measurements. As shown in Fig. 6C, the $A_0 - A$ values of the TCs/Fe–N–C/TMB/H₂O₂ system were all as high as expected due to the similar structure of TCs. However, they are rarely added to the same drug simultaneously, so that they do not interfere with each other when they are tested separately. In addition, the absorbance variation of the Fe–N–C/TMB/H₂O₂ system containing other competitive antibiotics was much lower than that of the sensor solution containing TCs, indicating that the proposed system can detect the three TCs with high specificity. Furthermore, the long-term performance stability of Fe–N–C nanomaterials for the catalytic oxidation of TMB substrates was also tested to evaluate the efficiency of the sensor system for a long duration. Based on their absorbance at 652 nm, the stability of the nanoparticles did not significantly decrease within 20 days of storage at 4 $^{\circ}\text{C}$ (Fig. 6D). On the 30th day, the catalytic activity of the Fe–N–C decreased slightly, possibly due to its increased hydrodynamic size, but its catalytic activity remained at 92.98%. In addition, even after 631 days had passed since we synthesized this material, its activity was still considerable. Based on colorimetric experiments and UV measurements, we demonstrate that Fe–N–C remains enzyme active after 631 days at 85.07% of its original activity (Fig. S5[†]). It is worth mentioning that the nanomaterials prepared in this paper exhibited superparamagnetic features, which creates a great possibility for the regeneration and reutilization of the catalyst. As can be seen from Fig. S6,[†] a magnet is placed outside a glass container containing the Fe–N–C dispersion, and the catalyst is magnetically drawn to one side of the container's inner wall. In the enzyme-mimicking cyclic experiment, the Fe–N–C was collected after each measurement and repeatedly cleaned with deionized water and ethanol, then dried in a vacuum oven at 60 $^{\circ}\text{C}$ for later



Scheme 1 Schematic illustration of the colorimetric biosensor for TCs detection using Fe–N–C as a nanozyme.



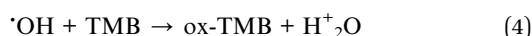
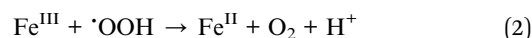
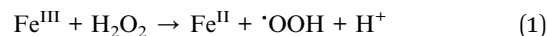
use. The Fe–N–C was then stored in a sealed dark bottle in a refrigerator at 4 °C and the catalytic oxidation to TMB substrates was monitored at 5 day intervals. As shown in Fig. 6E, the catalytic properties of Fe–N–C remained at 93.2% after 5 cycles under the same reaction conditions, confirming the satisfactory recycling stability of the catalyst. Furthermore, it is easy to collect in the actual application of TCs detection due to Fe–N–C's superparamagnetism, thus minimizing the secondary pollution caused by the detection process.

3.5 Investigation of the catalysis mechanism of Fe–N–C

To further determine the active sites of Fe–N–C, a 10 mM KSCN poison treatment blocking experiment was performed. As shown in Fig. 7A, the absorbance of the solution after KSCN poison treatment was significantly decreased, indicating that Fe sites played the key role of the active site in the catalytic process. In addition, we further evaluated the role of Fe(II) and Fe(III) on the catalytic activity of Fe–N–C. Fig. S7† shows the XPS spectra of Fe 2p before and after the catalytic reaction. By calculating the peak area, the ratio of Fe(II)/Fe(III) dropped from 1.25 to 0.50, which confirmed the dominant activity of Fe(II). Meanwhile, we observed that the catalytic activity of Fe–N–C was severely inhibited in N₂-saturated solution but enhanced obviously in O₂-saturated solution, confirming that O₂ is indispensable in this research (Fig. 7B).

Previous reports have indicated that the OXD-mimicking or POD-mimicking activity of metal-based nanomaterials was mainly attributed to an electron transfer mechanism caused by ROS production or the valence variations of metal ions.²⁸ To verify the above conjecture, we firstly conducted radical scavenging experiments without H₂O₂ participation. As shown in Fig. 7C, the contribution of the ·OH radical was explored through the inhibition of ·OH by isopropanol. The results

showed that the absorbance of ox-TMB was almost unchanged with the increase of isopropanol concentration, indicating that almost no ·OH was produced during the catalytic reaction. In addition, the absorbance at 652 nm decreased obviously when the concentration of O₂·⁻ radical scavenger p-BQ increased from 1 mM to 10 mM, which proved that Fe–N–C effectively catalyzed the formation of O₂·⁻ from O₂ in solution. This hypothesis was further investigated by adding L-histidine as an ¹O₂ radical scavenger. Fig. 7C shows a strong inhibitory effect in the presence of L-histidine, and it can be deduced that ¹O₂ was involved in a significant inhibitory effect. To sum up, the main active substance in the reaction system was ¹O₂ and the O₂·⁻ radical rather than the ·OH radical. Apart from mimicking oxidase, the prepared Fe–N–C can also act simultaneously as a POD-mimicking nanozyme in the presence of H₂O₂. It has been reported that a POD-mimicking nanozyme can decompose the O–O bond of H₂O₂ into strong oxidizing double hydroxyl radicals (·OH).²⁹ Simultaneously, reactions involving ·OH are triggered to perform one or two electron oxidations of substrates. Fe–N–C acts as the catalyst, organic substrates act as electron donors, molecular oxygen or H₂O₂ acts as the electron acceptor and is reduced according to the nature of enzyme catalysis.³⁰ In this system, Fe(II) on the surface of Fe–N–C plays a leading role in reacting with H₂O₂ to produce ·OH. The process can be described in the following reactions:³¹



To confirm this assumption, we chose terephthalic acid (TA) as a fluorescent probe to evaluate the formation of ·OH. TA hardly produces fluorescence on its own, but easily reacts with the ·OH radical to form highly fluorescent 2-hydroxyterephthalic acid (TAOH). As exhibited in Fig. 7D, TA is catalyzed by Fe–N–C with a higher fluorescence signal, indicating that more ·OH was generated. The above results proved that the prepared Fe–N–C possessed the ability to catalyze the decomposition of H₂O₂ to generate the ·OH radical, which could oxidize the peroxidase substrate TMB to ox-TMB. This result is consistent with some early reports.³²

DFT calculations were performed to elucidate possible inhibitory catalytic pathways. The configuration of Fe–N–C (Fig. 8A) and structures of H₂O₂ and TC adsorption configurations on the Fe–N–C surface are presented in Fig. 8. In all the configurations, in accordance with what is observed in Fig. 8B, H₂O₂ is chemically adsorbed on the top surface of Fe–N–C with an O atom of H₂O₂ being bonded with an Fe atom. And the DFT calculation signifies that H₂O₂ adsorbs on Fe–N–C with an adsorption energy of –0.97 eV. In comparison, the interaction between TC and Fe–N–C is more intense, with the two O atoms of TC being bonded by one Fe atom through a similar coordination. More importantly, the unsaturated (poly)cyclic

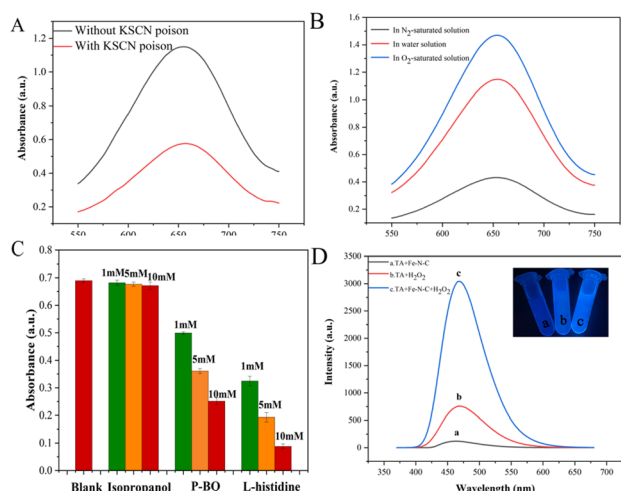


Fig. 7 (A) The absorbance changes of ox-TMB poisoned with 10 mM KSCN in HAC–NaAc buffer (pH = 4). (B) UV absorbance spectra of the colorimetric reaction in a saturated N₂/O₂ atmosphere. (C) Absorbances after inhibition of the reaction by different scavengers. (D) The effect of the Fe–N–C on the formation of the hydroxyl radical with TA as a fluorescence probe.



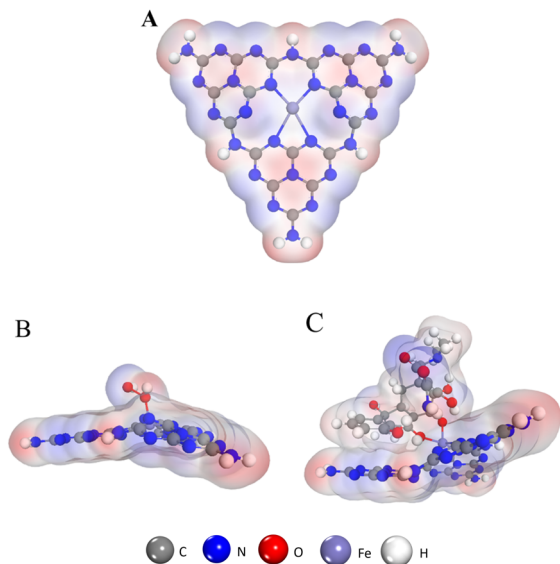


Fig. 8 (A) The structure of Fe–N–C. (B) The structure of the H₂O₂ absorption process on the Fe–N–C. (C) The structure of the TC absorption process on the Fe–N–C.

structures of TC and Fe–N–C are tightly adsorbed by π – π stacking (Fig. 8C). In the TC_{Fe–N–C} configuration, the adsorption energy of TC on Fe–N–C is -2.24 eV, which inhibits the decomposition of H₂O₂ into \cdot OH. Hence, according to the above calculations, the binding interaction between TC and Fe–N–C is strong enough to compete with H₂O₂ for Fe–N–C and fade the solution.

3.6 Application of the Fe–N–C in the determination of TCs in water samples

Due to their high hydrophilicity and low volatility characteristics, TCs can exhibit continuous persistence in aqueous environments.³³ Therefore, we have evaluated the practical application of the colorimetric biosensor for TCs in lake water samples. The recoveries of TCs ranged between 90.2% and 112.8%, and the RSDs ranged from 2.94% to 4.3% (Table S4[†]), indicating that the colorimetric sensor based on Fe–N–C is reliable for TCs detection in lake water.

4 Conclusions

In summary, dual enzyme mimicking Fe–N–C was synthesized through a one-step pyrolytic process. The established colorimetric platform showed excellent catalytic performance in terms of high selectivity to TCs, and is a promising sensor in environmental analysis. In addition, this assay displayed a recoverability and stable sensing performance, and its practical application in water samples has been verified.

Conflicts of interest

There are no conflicts to declare.

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Notes and references

- 1 K. Bousova, H. Senyuva and K. Mittendorf, *J. Chromatogr. A*, 2013, **1274**, 19–27.
- 2 J. Zhou, X. Xue, Y. Li, J. Zhang, F. Chen, L. Wu, L. Chen and J. Zhao, *Food Chem.*, 2009, **115**, 1074–1080.
- 3 K. Phomai, S. A. Supharoek, J. Vichapong, K. Grudpan and K. Ponghong, *Talanta*, 2023, **252**, 123852.
- 4 C. Zhou, J. Deng, G. Shi and T. Zhou, *Electrophoresis*, 2017, **38**, 1060–1067.
- 5 X. Liu, S. Zheng, Y. Hu, Z. Li, F. Luo and Z. He, *Food Anal. Methods*, 2016, **9**, 2972–2978.
- 6 J. Zhang, Z. Bao, J. Qian, H. Zhou and K. Zhang, *Anal. Chim. Acta*, 2022, **1216**, 339991.
- 7 M. Kaczmarek, K. Staninski and M. Stodolny, *J. Lumin.*, 2021, **36**, 1945–1952.
- 8 W. Kong, X. Guo, M. Jing, F. Qu and L. Lu, *Biosens. Bioelectron.*, 2020, **150**, 111875.
- 9 W. Xia, P. Zhang, W. Fu, L. Hu and Y. Wang, *Chem. Commun.*, 2019, **55**, 2039–2042.
- 10 C. Song, W. Ding, H. Liu, W. Zhao, Y. Yao and C. Yao, *New J. Chem.*, 2019, **43**, 12776–12784.
- 11 D. Singh, D. Sillu, A. Kumar and S. Agnihotri, *Environ. Sci.: Nano*, 2021, **8**, 1308–1325.
- 12 W. Li, G. C. Fan, F. Gao, Y. Cui, W. Wang and X. Luo, *Biosens. Bioelectron.*, 2019, **127**, 64–71.
- 13 B. Xu, Y. Cui, W. Wang, S. Li, C. Lyu, S. Wang, W. Bao, H. Wang, M. Qin, Z. Liu, W. Wei and H. Liu, *Adv. Mater.*, 2020, **32**, e2003563.
- 14 H. Wei and E. Wang, *Chem. Soc. Rev.*, 2013, **42**, 6060–6093.
- 15 J. Wu, X. Wang, Q. Wang, Z. Lou, S. Li, Y. Zhu, L. Qin and H. Wei, *Chem. Soc. Rev.*, 2019, **48**, 1004–1076.
- 16 P. Zhang, D. Sun, A. Cho, S. Weon, S. Lee, J. Lee, J. W. Han, D. P. Kim and W. Choi, *Nat. Commun.*, 2019, **10**, 940.
- 17 K. Fan, J. Xi, L. Fan, P. Wang, C. Zhu, Y. Tang, X. Xu, M. Liang, B. Jiang, X. Yan and L. Gao, *Nat. Commun.*, 2018, **9**, 1440.
- 18 N. Ramaswamy, U. Tylus, Q. Jia and S. Mukerjee, *J. Am. Chem. Soc.*, 2013, **135**, 15443–15449.
- 19 T. Zhang, H. Li, M. Liu, H. Zhou, Z. Zhang, C. Yu, C. Wang and G. Wang, *Sens. Actuators, B*, 2021, **341**, 129987.
- 20 X. Zhou, M. Wang and X. Su, *Sens. Actuators, B*, 2021, **338**, 129777.
- 21 Y. Wang, Z. Zhang, G. Jia, L. Zheng, J. Zhao and X. Cui, *Chem. Commun.*, 2019, **55**, 5271–5274.
- 22 D. Li, Y. Qu, X. Liu, C. Zhai and Y. Liu, *Int. J. Hydrogen Energy*, 2021, **46**, 18364–18375.
- 23 M. Liu, L. Wang, K. Zhao, S. Shi, Q. Shao, L. Zhang, X. Sun, Y. Zhao and J. Zhang, *Energy Environ. Sci.*, 2019, **12**, 2890–2923.



- 24 L. Wang, K. Liang, L. Deng and Y.-N. Liu, *Appl. Catal., B*, 2019, **246**, 89–99.
- 25 Y. Xiong, S. Chen, F. Ye, L. Su, C. Zhang, S. Shen and S. Zhao, *Chem. Commun.*, 2015, **51**, 4635–4638.
- 26 A. C. Razavi, T. N. Kelly, M. J. Budoff, L. A. Bazzano, J. He, C. Fernandez, J. Lima, K. Nasir, R. S. Blumenthal, M. J. Blaha and S. P. Whelton, *Atherosclerosis*, 2021, **326**, 56–62.
- 27 H. Q. Zheng, Y. N. Zeng, J. Chen, R. G. Lin, W. E. Zhuang, R. Cao and Z. J. Lin, *Inorg. Chem.*, 2019, **58**, 6983–6992.
- 28 H. Wang, K. Wan and X. Shi, *Adv. Mater.*, 2019, **31**, e1805368.
- 29 P. Ni, H. Dai, Y. Wang, Y. Sun, Y. Shi, J. Hu and Z. Li, *Biosens. Bioelectron.*, 2014, **60**, 286–291.
- 30 Y. Zhang, C. Dai, W. Liu, Y. Wang, F. Ding, P. Zou, X. Wang, Q. Zhao and H. Rao, *Mikrochim. Acta*, 2019, **186**, 340.
- 31 B. Liu, H. Zhu, R. Feng, M. Wang, P. Hu, J. Pan and X. Niu, *Sens. Actuators, B*, 2022, **370**, 132451.
- 32 P. Chen, H. Zhong, X.-r. Li, M. Li and S. Zhou, *Appl. Clay Sci.*, 2021, **209**, 106109.
- 33 R. Daghrir and P. Drogui, *Environ. Chem. Lett.*, 2013, **11**, 209–227.

