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Emerging investigator series: use of behavioural endpoints in the regulation of chemicals

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Interest in behavioural ecotoxicology is growing, partly due to technological and computational advances in recording behaviours but also because of improvements of detection capacity facilitating reporting effects at environmentally relevant concentrations. The peer-reviewed literature now contains studies investigating the effects of chemicals, including pesticides and pharmaceuticals, on migration, dispersal, aggression, sociability, reproduction, feeding and anti-predator behaviours in vertebrates and invertebrates. To understand how behavioural studies could be used in regulatory decision-making we: (1) assessed the legal obstacles to using behavioural endpoints in EU chemicals regulation; (2) analysed the known cases of use of behavioural endpoints in EU chemicals regulation; and (3) provided examples of behavioural endpoints of relevance for population level effects. We conclude that the only legal obstacle to the use of behavioural endpoints in EU chemicals regulation is whether an endpoint is considered to be relevant at the population level or not. We also conclude that ecotoxicity studies investigating behavioural endpoints are occasionally used in the EU chemicals regulation, and underscore that behavioural endpoints can be relevant at the population level. To improve the current use of behavioural studies in regulatory decision-making contribution from all relevant stakeholders is required. We have the following recommendations: (1) researchers should conduct robust, well-designed and transparent studies that emphasize the relevance of the study for regulation of chemicals; (2) editors and scientific journals should promote detailed, reliable and clearly reported studies; (3) regulatory agencies and the chemical industry need to embrace new behavioural endpoints of relevance at the population level.

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Environmental significance

The peer-reviewed literature contains ecotoxicity studies investigating behavioural effects such as migration, dispersal, aggression, grouping, reproduction and feeding in vertebrates and invertebrates. However, little is known about the studies' contribution to regulatory decision-making. This study concludes that it is possible to use behavioural endpoints in EU chemicals regulation if the endpoint is assessed as relevant at the population level, but that there are few examples of such use. Recommendations for researchers, regulators, risk assessors and scientific journals who strive to improve the use of behavioural endpoints in environmental risk assessment of chemicals are provided.

Introduction

In recent years there has been growing interest surrounding behavioural ecotoxicology. For example, there has been a steady increase in behavioural effect measurements in the US EPA ECOTOX database,¹ currently adding up to 17 324 measurements for the aquatic environment and 13 809 for the terrestrial environment. An area of particular rapid expansion is the use of fish models for drug discovery and design.^{2,3} This interest in behavioural endpoints is partly driven by technological and computational advances in recording behaviours, but also due to an increasing number of laboratory studies recording behavioural effects at environmentally relevant concentrations.⁴⁻⁶ What was in the past quite a laborious process of

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careful observations and potentially watching hours of video footage is now high-throughput computer recorded endpoints without human subjectivity. Such advances are important because behavioural ecotoxicology has been previously criticized for concerns over observational errors. Technological advances, for example in electronic tagging, have also allowed the monitoring of behaviour *in situ* during field studies, which is helping scientists and regulators gain confidence in laboratory derived endpoints. In fact, changes in behaviour are increasingly considered valuable for advancing the next generation of ecotoxicology research.⁷

Traditionally the regulatory evaluation of chemicals, including pesticides and pharmaceuticals, has not included behavioural endpoints, even though such endpoints have several advantages. These advantages include that behaviour: (1) provides a connection from molecular and physiological processes to population level processes; (2) is a sensitive 'early warning signal' of chemical contamination since behavioural responses can occur at lower levels of contamination than more traditional endpoints; (3) improves the ecological relevance of environmental risk assessments due to the well-established theoretical framework and suite of fitness-related endpoints underpinning behavioural studies; and (4) provides a high throughput identification of potential underlying chemical mode of actions.⁸⁻¹⁴ In addition, it can be argued that from a resource and animal welfare perspective, behavioural endpoints of sufficient reliability (*i.e.* inherent quality and repeatability) and relevance (*i.e.* appropriateness for a particular hazard identification¹⁵), should be used in chemicals regulation, particularly considering that public health and environment research is routinely funded by public funds.

Hazard and risk assessments to decide on management measures for chemicals are common for national and international regulation agencies. Ecotoxicity studies are used to determine the potentially environmentally hazardous properties of a chemical and to support establishment of acceptable exposure levels in environmental matrices. Traditionally, regulatory assessments have been based on ecotoxicity studies measuring mortality, growth, reproduction, and development. These individual-level endpoints have been described as having a clear connection to the persistence of a population.¹⁶ To facilitate the assessment process and promote use of studies across jurisdictions, standard studies investigating these population relevant endpoints have been developed by international and national organizations, such as OECD and US Environmental Protection Agency.^{17,18} Inclusion of other endpoints, such as behaviour, are often seen in higher tier assessments, and/or in what is called a "weight of evidence" assessment. Ecotoxicity studies are evaluated for their reliability and relevance when deciding which studies to include in the regulatory assessment.¹⁹ Following standard procedures in ecotoxicity studies has led to reliable results applicable for chemicals regulation due to robust test protocols and detailed reporting requirements. In addition, partly due to the use of endpoints considered to be important on a population level, standard studies have been assessed as relevant for chemicals regulation. Ecotoxicity studies that are not performed according

to a standard procedure and include alternative endpoints, such as behaviour, run the risk of being disregarded or seen as evidence of lower weight. Often, standard studies are performed by, or on behalf of, the regulated party, while non-standard studies are performed by researchers working in academia.²⁰

Aim and methodology

Given the abundance of information available in the peer-reviewed literature on the impact of chemicals on the behaviour of non-target organisms, this study aimed to:

(1) Assess the legal obstacles to using behavioural endpoints in EU chemicals regulation.

Regulatory guidance documents that set the scope for assessment of chemicals were analysed with the purpose to understand if, how and where ecotoxicity studies investigating behavioural endpoints could be used in the assessment of chemicals. Guidance documents for the following six EU policy and regulatory areas were analysed: the Registration, Evaluation, Authorisation and Restriction of Chemicals (REACH) regulation; Classification, Labelling and Packaging of chemical substances and mixtures (*i.e.* the CLP regulation); the Biocidal Products Regulation; the Plant Protection Products regulation; Environmental assessment for human and veterinary products; and Derivation of Environmental Quality Standards (within the Water Framework Directive). The following aspects were investigated:

- If the guidance document prohibits use of behavioural endpoints.
- If the guidance document recommends or require tests using behavioural endpoints.
- What weight behavioural endpoints are given in chemical assessments.
- How population level effect is defined.
- Other aspects relevant for use of behavioural endpoints in chemicals regulation.

(2) Analyse the known cases of behavioural endpoints used in EU chemicals regulation.

Due to lack of transparency in EU chemicals regulation and lack of searchable regulatory databases, it is not possible to, in an accessible way, get an overview of current use of behavioural endpoints. Instead we chose to analyse the six known, to us, cases in EU chemicals regulation. These six cases were known to us because of public discussions or due to our personal discussions with regulators and risk assessors at research institutes and national regulatory agencies. The regulatory use was examined to clarify the role of ecotoxicity studies investigating behavioural endpoints for decision-making in these specific cases. The following aspects were investigated:

- Type of behaviour endpoint used.
- Which regulatory framework it was used in.
- The weight the study was given in the chemical assessment by the risk assessors (as key, supportive or low) and how this rank was justified. Key studies are studies used when setting guideline values such as an environmental quality standard (EQS), supportive studies are considered important evidence but are not considered to be key study (*e.g.* due to issues related



to reliability and relevance of the study, or other studies showing toxicity at lower concentrations), and studies with low weight are the ones that are discussed but do not contribute to the overall conclusion of the assessment.

A comparison with the use in the US, and a comparison with human health assessment in EU, was made.

(3) Provide examples of behavioural endpoints with relevance for population level effects.

From a regulatory perspective, contaminant exposure only matters if it influences survival, development and reproduction because these measures can cause population decline. However, there are a number of fitness related behaviours that result in population change. Examples from both invertebrates and vertebrates were provided for different behavioural categories where there are clear links to population dynamics. These behavioural categories include migration, dispersal, aggression, sociability, reproduction, feeding, and anti-predator behaviours, and were selected because of their strong link to population growth and health. We describe how the behavioural effects of chemical contaminants, which are observed at the individual level, can have important cascading impacts and ultimately detrimental effects on population level processes.

Results

Regulatory guidance on the use of behavioural studies in EU chemical assessments

None of the analysed EU guidance documents prohibited use of behavioural endpoints when assessing environmental effects (Table 1). In four of the policy areas, behaviour endpoints were not mentioned as examples of endpoints of interest. Instead, there was a general focus on traditional endpoints such as survival, reproduction, and development. Two of the guidance documents for the REACH regulation stated that behavioural studies could be used as supportive evidence, but not without backing from studies using more traditional endpoints. Examples of such behavioural endpoints included sediment avoidance or burrowing activity.

The guidance document for deriving EQS states that behavioural endpoints are unsuitable as the basis for EQS derivation since the endpoints “do not include direct measurements of survival, development or reproduction”. However, the guidance document makes an exception for changes in behaviour resulting in impaired competitive fitness and avoidance reactions that would make individuals avoid contaminated habitats where otherwise they normally would be present. In contrast to the guidance document for deriving EQS, the guidance documents for Plant Protection Products regarded behavioural endpoints as key evidence. Examples of this are the assessment of field data for bees, and acknowledgement of avoidance and behaviour related to reproduction as important endpoints when assessing risks for birds and mammals.

For the four policy areas where the regulated party is obliged to perform new ecotoxicity studies if existing studies are not sufficient to fulfil the data demands, it is recommended to use standard studies, primarily from OECD. Currently, observed changes in behaviour need to be reported according to several

OECD standard studies. For example, the OECD TG 222 Earthworm Reproduction Test states that an inability to dig into the soil should be reported. In the OECD TG 246 Acute Contact Toxicity Test for bumblebees, signs of reduced coordination should be reported. However, very few standard studies are designed to directly investigate effects on specific behaviours; the behavioural observations made are instead an additional measure for the primary endpoints studied, *i.e.* effects on mortality, growth, reproduction and development.¹ The clearest examples of situations where behavioural endpoints are considered important in OECD standard studies, together with other endpoints, are when assessing effects on bees using OECD TG 213, 214 or 245. In contrary, there are two standard studies from ASTM International specifically addressing behaviour: E1604 for testing aquatic organisms, and E1768 for testing freshwater fish. For human health assessment, the OECD TG 426 Developmental Neurotoxicity Study on rats is used to investigate behavioural endpoints such as motor activity, learning and memory, anxiety, and social behaviour.

Several of the EU regulatory documents that we analysed mentioned that there is a need to develop new standard studies with behavioural endpoints. For Plant Protection Products, standard studies for birds that include endpoints such as avoidance behaviour, parental care and nesting behaviour, are requested. For bees, a standard study investigating chronic toxicity for larvae is missing, the guidance document recommends using an extended version of OECD TG 213 where behavioural endpoints are included. Currently, the OECD is developing the standard study “Homing flight test on honeybee after single exposure to sublethal doses”.²¹

A principle within regulatory assessments of chemicals has been to base decisions on toxicity and ecotoxicity studies investigating adverse effects. WHO defines adverse effects as “Change in the morphology, physiology, growth, development, reproduction or lifespan of an organism, system or (sub)population that results in an impairment of functional capacity to compensate for additional stress or an increase in susceptibility to other influences”.²² What separates human health assessments from environmental assessments is that population relevant effects and not individual effects are considered in the latter. The only known exception to this level of a protection goal in EU is a recommendation from EFSA to protect aquatic vertebrates (fish and amphibians) at the individual level in acute risk assessments to avoid visible mortality.²³

In the outdated and no longer used Guidance Document on Risk Assessment for Birds and Mammals,²⁴ the focus on population relevant effects was interpreted as direct effects on survival, development or reproduction, and therefore guidance like the following can be found: “One aim of the ecological risk assessment is to predict effects on the population level, although this is difficult or impossible to measure directly. The usual approach is based on the consideration that effects on populations will not occur if the survival rate, reproduction rate and development of individuals are not affected. Therefore, in principle, only endpoints in toxicity tests which are related to these key factors of population dynamics are ecotoxicologically relevant”. However, in more recently developed guidance documents populations have





Table 1 Overview of EU guidance documents for six different policy and regulatory areas analysed in terms of their inclusion of behavioural endpoints

Responsible agency and policy/regulatory area	Guidance document(s)	Recommendations from the guidance documents	Example text from the guidance document (the number represent guidance document in column 2)
European Chemicals Agency. REACH regulation	(1) Part B hazard assessment (2) Chapter R.2 information requirements (3) Chapter R.4 evaluation of available information (4) Chapter R.7b + R.7c Endpoint specific guidance	Behavioural endpoints mentioned in two documents, where behavioural studies can be used as supportive evidence. Standard studies are recommended when performing new studies. All available studies are recommended for evaluation	(3) "Behavioural endpoints like sediment avoidance or burrowing activity have not been standardised. Such endpoints can give indications on toxic effects but should not be interpreted in isolation" (4) "Reproduction tests include parental and reproductive endpoints. An endpoint relating to overall reproductive success should normally be selected to define the long-term NOEC. Depending on the individual case and the availability of data, this could be the reproduction rate, the survival or growth rate of the offspring, or behavioural parameters in adults or young". "Screening endpoints such as behavioural responses, i.e. avoidance testing should not be interpreted in isolation"
European Chemicals Agency. Classification, Labelling and Packaging (CLP) regulation	Guidance on the application of the CLP criteria	Behavioural endpoints not mentioned. All available studies are recommended for evaluation	—
European Chemicals Agency. Biocidal Products Regulation	(1) Volume IV: environment. Part A: information requirements (2) Volume IV: environment – assessment and evaluation (parts B + C)	Behavioural endpoints not mentioned. Standard studies are recommended when performing new studies. All available studies are recommended for evaluation	—
European Food Safety Authority, the European Parliament and Council. Plant Protection Products Regulation	(1) Guidance on tiered risk assessment for plant protection products for aquatic organisms in edge-of-field surface waters (2) Guidance document on terrestrial ecotoxicology, SANCO 2002 (3) Guidance document on risk assessment for birds and mammals, EFSA journal 2009 (5) Draft EFSA guidance document on the risk assessment of plant protection products on bees	Behavioural endpoints mentioned for bees, birds and mammals. Standard studies are recommended when performing new studies. All available studies are recommended for evaluation	(2) "Key parameters which may be considered in a field trial include: mortality (assessed via the use of dead bee traps), behaviour (including foraging behaviour in the crop and around the hive), honey crop (assessed via weighing the hive at appropriate intervals) and state of colony (including an assessment of brood)" (3) "Granivorous birds and mammals may be able to distinguish treated seeds from non-treated seeds and may show a preference for either treated or untreated seeds in their diet. This may be influenced by various factors including appearance, taste or surface texture of the treated seed, and aversive reactions to the active



Table 1 (Contd.)

Responsible agency and policy/regulatory area	Guidance document(s)	Recommendations from the guidance documents	Example text from the guidance document (the number represent guidance document in column 2)
European Medicines Agency. Medicinal Products Regulations	(1) Guideline on the environmental risk assessment of medicinal products for human use (2) Guideline on environmental impact assessment (EIAS) for veterinary medicinal products – phase I and II (3) Guideline on environmental impact assessment for veterinary medicinal products in support of the VICH guidelines GL6 and GL38	Behavioural endpoints not mentioned. Standard studies are recommended when performing new studies. All available studies are recommended for evaluation	<i>substance. Information on such preferences/avoidance behaviour can, in combination with data on the availability of treated and non-treated seeds on the soil surface, be used to refine the risk assessment. No standard guideline for testing avoidance is as yet available”. “...the one-generation avian reproduction study does not include exposure during all relevant stages of the bird's development or the measurement of other relevant endocrine-sensitive endpoints such as behaviour (e.g. parental care, nesting behaviour, territoriality and mounting behaviour)”</i> —
European Commission. Deriving Environmental Quality Standards	Technical Guidance for Deriving Environmental Quality Standards. Guidance document no. 27	Behavioural endpoints mentioned, some can be used. All available studies are recommended for evaluation	<i>“...the assessor may be faced with data from studies describing endpoints that do not include direct measurements of survival, development or reproduction but, rather, describe e.g. behavioural effects, anatomical differences between control and treatment groups, effects at the tissue or sub-cellular level, such as changes in enzyme induction or gene expression. Generally these are unsuitable as the basis for EQS derivation. However, some other endpoints are relevant. For example, anatomical changes to gonad development that would prevent successful reproduction, or changes in behaviour if the effect described would impair competitive fitness may be relevant. Avoidance reactions may also be relevant if populations are likely to avoid a contaminated habitat where they would normally be present”</i>

been linked to abundance/biomass and individuals have been linked to behaviour/survival/growth.²³

There are two guidance documents explicitly stating that behavioural endpoints are relevant to the protection goal. The EFSA's guidance document on the risk assessment of plant protection products on bees list behaviour as key factor, and it does so already in the regulation (EC 1107/2009).²⁵ The new EFSA guidance for the identification of endocrine disruptors stresses that behavioural endpoints are implicitly covered by the WHO definition of adversity since these type of effects will have implications on reproduction and development.²⁶

In comparison to the EU, the Canadian Council of Ministers of Environment and some provincial jurisdictions, such as the British Columbia Ministry of Environment and Climate Change Strategy (BCENV) have established water quality guideline-development protocols that allow for behavioural endpoints if ecological relevance can be established. Such endpoints include predator avoidance, swimming ability, or certain olfactory-mediated behaviours that can be linked to ecological relevance. Ecological relevance, then, is defined as follows: "Ecological relevance pertains to whether physical abilities (e.g., swimming speed, orientation ability, and migratory fitness), physical traits (e.g., fin size/shape), physiological traits (e.g., production of a certain enzyme), and/or behavioural tendencies (e.g., swimming in groups) of organisms are important enough to influence a species' ecological competitiveness. Characteristics that are of high ecological relevance are those that have a strong positive or negative influence on survival, reproductive ability, and growth (e.g., stunting, high fertility, and organ failure)".²⁷ In addition, BCENV continues to monitor the literature for contaminant-induced olfactory impairment (including sensitive olfactory-

mediated behaviours) to ensure that existing water quality guidelines remain protective against these sublethal effects (A. Azizishirazi, *Pers. Comm.*, 2019). Thus, there are protocols and precedents that could be exported to other jurisdictions in order to facilitate the integration of behavioural endpoints into both established and newly developed regulatory frameworks for chemicals.

Known cases of behavioural endpoints used in EU chemicals regulation

The total number of cases where behavioural endpoint has been used in EU chemicals regulation is not known, and due to lack of transparency and searchable databases in the EU chemicals regulation it is not possible to make a quantitative analysis in an assessable way. Based on public and personal discussions with regulators and risk assessors we conclude that the overall regulatory use of behaviour endpoints seems to be rare. The six cases presented below are the only known publicly available cases to us (Table 2). From the analysis below we could conclude that there is spatial, temporal and product-specific variation in how these behavioural studies have been used in chemical registration and pre-market assessments in the different EU regulations.

Assessed and used as key study. A study investigating avoidance behaviour in eel²⁸ was used in the European Union Risk Assessment Report (RAR) from 2002 for the solvent and fuel component methyl tertiary-butyl ether (MTBE).^{29,30} This assessment was performed under the Existing Substances Regulation, the forerunner of the REACH regulation. The avoidance study was performed since, according to the RAR, there was a need for further information and/or testing. The

Table 2 Examples of current use of behavioural endpoints in environmental assessments of chemicals in EU

Substance & regulation	Reference	Behaviour endpoint & test animal	Importance for the regulatory decision
Methyl tertiary-butyl ether (MTBE) in existing substances regulation	28	Avoidance (attraction) in fish (<i>A. anguilla</i>)	Key study
Methyl tertiary-butyl ether (MTBE) in Danish national EQS-derivation	28		Key study
4-Nonylphenol in REACH (substance evaluation, SVHC)	32	Spawning in fish (<i>P. promelas</i>)	Supportive evidence but not key endpoints for concluding that the substance is an endocrine disrupting chemical
	33	Feeding, social and aggression in fish (<i>O. mykiss</i>)	
	34	Sexual in fish (<i>P. reticulata</i>).	
Lambda-cyhalothrin in the Plant Protection Product Regulation	36	Drifting in insect species (<i>B. rhodani</i> & <i>L. cf. Fusca</i>)	Supportive evidence but not key endpoints
	37	Pre-copulatory behaviour in amphipod (<i>G. pulex</i>)	
	38	Drifting in invertebrates (<i>L. nigra</i> , <i>G. pulex</i> , <i>H. sulphurea</i>)	
Alpha-cypermethrin	Field study from chemical company	Foraging activity in honeybee (<i>A. mellifera</i>)	Low weight, other endpoints showing limited or low effects were considered more important
Di-2-ethylhexyl phthalate (DEHP) in the Existing Substances Regulation	41	Predation efficiency in dragon fly larvae (<i>Aeshna</i>)	Low weight, due to low reliability
Zinc pyrithione in the Biocidal Products Directive	43	Avoidance in amphipod (<i>M. affinis</i>)	Low weight, due to insufficient reporting of results. However, used to justify selection of assessment factor



same study was, in 2009, used by the Danish Environmental Protection Agency to derive an EQS under the Water Framework Directive.³¹ In both cases, the behavioural study was considered to be the key study.

Assessed and used as supporting evidence. Three non-standard studies investigating different behavioural endpoints in fish^{32–34} were included in the assessment report from the European Chemicals Agency (ECHA) for the industrial chemical 4-nonylphenol.³⁵ A so-called “substance evaluation” was performed under the REACH-regulation with the purpose to investigate if 4-nonylphenol was to be considered a “Substance of Very High Concern” (SVHC) due to its endocrine disrupting properties. The behavioural studies acted as supportive evidence but were not considered key studies in the assessment.

In the assessment for the insecticide lambda-cyhalothrin in the Plant Protection Product Regulation three non-standard studies investigating behaviour endpoints were used.^{36–38} The behavioural studies were considered to have relevant results but due to a number of factors the studies were only used as supporting evidence. Factors mentioned in the assessment were: the tested formulation was not identical to the assessed formulation; effects were observed at all treatments and hence no NOEC value could be obtained; unclear reporting of methodology; concerns regarding the study reliability; the study was a non-standard study.³⁹

Assessed as low weight due too low effect. The Swedish Chemical Agency assessed a product containing the insecticide alpha-cypermethrin. A short-term reduction in foraging activity in bees was reported in a field study submitted by the chemical company.⁴⁰ However, since studies investigating mortality only showed a limited and short-term increase, and since no effects occurred on the brood nest or bee brood development, the behavioural effects were not considered sufficient to alter the conclusion from the other studies. The behavioural endpoints were given low weight in the assessment due to the limited effects shown in the study.

Assessed as low weight due to low reliability. The Swedish Chemicals Agency used a study that measured the predation efficiency of dragonfly larvae exposed to the plasticizer di-2-ethylhexyl phthalate (DEHP)⁴¹ under the Existing Substances Regulation.⁴² However, the study was considered to be of low reliability due to the use of only one test concentration and the high amounts of ethanol used when spiking the sediment, and therefore the study was given low weight in the risk assessment.

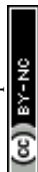
Assessed as “not assignable” due to insufficient reporting. A study investigating avoidance to zinc pyrithione in a sediment living amphipod⁴³ was evaluated by the Swedish Chemicals Agency in an assessment of copper pyrithione (zinc and copper pyrithione are considered to be the same substances at low concentrations, dominated by the free ion of pyrithione). This was done under the Biocidal Products Directive (BPD), the forerunner of the Biocidal Products Regulation (BPR).⁴⁴ The study was assessed as “not assignable”, using the reliability evaluation categories first developed by Klimisch *et al.*,⁴⁵ due to insufficient reporting of results. The study only presented an EC₅₀-value and the raw data needed to calculate a NOEC could not be located by the author of the study. In the risk assessment

document, it was also mentioned that the tested concentrations had not been analytically confirmed in the study. Despite this, the study was used as supporting evidence to justify the choice of assessment factor when deriving the Predicted No Effect Concentration (PNEC): “*The assessment factor 100 seems to be most applicable, given that there is only one test (the Hyalella) where concentrations were measured over the whole exposure period, and given that there are only two organisms types tested (for mortality), and finally, also considering that the non-guidance research test with avoidance (Monoporeia) as endpoint indicate effects are expected at ~100 times lower*”. To clarify, in this assessment the behavioural study was both disregarded and used as justification. This raises concern regarding the consistency of the assessment methodology.

Use of behavioural endpoints in the US regulation. An example from the US shows use of behavioural endpoints in another type of chemical assessment. The US Environmental Protection Agency included behavioural endpoints in pilot “Biological Evaluations” for the data rich pesticides chlorpyrifos, malathion, and diazinon. These assessments were performed to analyse impacts on species listed under the Endangered Species Act (ESA). As introduced above, behavioural endpoints for fish, amphibians, invertebrates, birds, and mammals are described as appropriate evidence for individual level and not population level, *e.g.* from the assessment for diazinon: “*This endpoint is considered relevant to the fitness of an individual because limited locomotion would potentially increase the likelihood that an individual would be susceptible to predation as well as an inability to fly and thus migrate*”. Still, behavioural endpoints are included as one line of evidence in a weight of evidence approach, and thereby acting as supporting evidence in these ESA assessments.⁴⁶

Use in human health risk assessments. Use of behavioural endpoints (in animals) in human health risk assessments is more common, but still, studies investigating behavioural effects are not part of the core dataset for neither pesticides nor biocides. Instead, they are requested if triggered by results from studies on lower tiers in the assessment structure.^{47,48} The use of behavioural endpoints in human health risk assessments are primarily facilitated through two OECD standard studies, the two-generation reproduction toxicity study (OECD TG 416) and the developmental neurotoxicity study (OECD TG 426). In addition to other endpoints, these standard studies measure reproductive behaviour and neuro-behavioural endpoints, respectively. Their regulatory use has not been straight forward. For example, in the assessments for the insecticides deltamethrin and lambda-cyhalothrin the use of OECD TG 426, or a similar standard study, raised concerns regarding the sensitivity of the strain of rat (another strain of rat had shown to be more sensitive in other studies) and the exposure level in offspring was not clear (exposure through lactation). An additional assessment factor was suggested to account for the uncertainties, but this was ultimately not accepted. In the end, the studies were used as supporting evidence in both assessments.^{49,50}

The difficulties interpreting the results from OECD TG 426 was acknowledged in a study aimed at identifying areas of



improvement in this standard study. The OECD TG 426 offers several possible options in the test design when assessing “learning and memory”. For example, active avoidance of an unpleasant action can be tested, or the ability to find a way through a maze. These two test designs assess different types of behaviour. Experienced evaluators and thorough guidance are needed to be able to detect study designs that likely will result in negative results. The flexibility in the test design was considered the main reason for the interpretation difficulties, however, the flexibility was also considered necessary due to the large variation in tested substances.⁵¹ An analysis of neurotoxicity studies for bisphenol A showed that effects were more often observed in the behavioural endpoints not required (*e.g.* social and sexual behaviour) according to OECD TG 426, compared to the endpoints required (*e.g.* motor activity).⁵² This was seen especially at low doses, and exemplifies the importance of non-standard behavioural endpoints in regulation of chemicals.

Non-standard studies investigating behavioural endpoints have also contributed to regulatory decisions. When the restriction proposal for bisphenol A in thermal paper (used for cash register receipts) was adopted by the European Commission effects on behaviour (alteration of spatial memory and learning functions) was one of four identified key risks for the unborn child, together with effects on the female reproductive system, effects on the mammary gland, and risk for obesity. However, since the available studies, including the behavioural studies, did not allow for a quantification of the dose–response relationship, the studies could not be used directly for the derived no-effect level (DNEL, *i.e.* the level of exposure to a substance above which humans should not be exposed). Instead, a study investigating kidney effects in mice was used as key study together with the standard assessment factor recommended in the regulation and an additional assessment factor (of six) to take into consideration the effects seen in the behavioural studies.⁵³

Relevance of behavioural endpoints for population level effects

The behaviour of an organism is intrinsically linked to its ‘fitness’ in a variety of ways through its capacity to communicate, find food, evade predation or catch prey, find mates, defend territories, or undergo large scale migrations. Ecologists have been linking behaviour with population level effects for decades⁵⁴ but within the field of ecotoxicology the need for incorporating ecologically important behavioural measures has only been articulated more frequently in recent years.⁶ Behavioural ecologists have studied how behavioural variation among animals leads to differential fitness, measured as effects on life-history characters (*e.g.* growth rate, development rate, survival). For example, reduced mobility (*i.e.* activity) has repeatedly been linked with reduced growth and development of a wide range of organisms.^{55–58} Another behaviour where variation has been tightly linked to fitness is sociality, an individual’s tendency to associate spatially with conspecifics. More social individuals generally experience increased likelihood of survival in the presence of lethal predators.^{59–62} These are just two types of

behaviours (out of several) where trait-variation has been tightly linked to changes in life-history characters and as such translates into fitness (*i.e.* individual-level) and population effects. However, identifying which specific behaviours within or among various species are concretely linked to changes at population level due to chemical exposure has not received extensive attention. Below are examples where behavioural effects on individual level in ecotoxicology have been linked with effects and responses of population level importance.

Migration and dispersal. Although changes in animal activity rates are one of the most commonly measured behavioural responses to chemical contamination in the laboratory setting,^{63,64} very few studies have connected these measures of small scale activity to dispersal or migration in the wild. Recently, Hellström *et al.*⁶⁵ found that the anxiolytic pharmaceutical, oxazepam, caused Atlantic salmon (*Salmo salar*) smolts to migrate faster both in laboratory migration pools and downstream in a river. This anxiolytic-induced boldness, however, was not beneficial in the wild because exposed salmon were predated upon to a greater extent.⁶⁶ Similarly, Woodman and colleagues⁶⁷ found that virile crayfish were strongly attracted to sertraline, and that sertraline-exposed crayfish showed increased aggression towards control animals. This result also suggests a maladaptive response that could increase the risk of predation in the wild. Marentette *et al.*⁶⁸ found that invasive round goby (*Neogobius melanostomus*) collected from environments heavily contaminated with metals, polychlorinated biphenyls (PCBs), and polyaromatic hydrocarbons (PAHs) dispersed slower through a maze in the laboratory compared to gobies collected from a relatively cleaner site. However, by using mark-recapture, the authors also found that fish from the same contaminated and clean sites showed similar dispersal rates in the field, underscoring the importance of also measuring behaviour in the wild when possible. Animals migrate or disperse for many reasons, including to avoid unfavourable environmental conditions, to access more productive habitats, or to reproduce. Migration and dispersal success are therefore important behaviours that can shape the structure and viability of populations.^{69–71}

Aggression, sociability and reproductive behaviours. Chemical contaminants are known to impact complex social interactions such as sociability (association of two or more conspecifics in space), aggression, and reproductive behaviour, and these behaviours have direct implications for fitness and population dynamics.^{6,54,72} For example, killifish (*Fundulus diaphanus*) exposed to environmentally relevant doses of 4-nonylphenol form looser shoals, and this was linked to disrupted chemical signals emitted by the signalling fish.⁷³ The antidepressant fluoxetine has been shown to disrupt the integration of pheromone cues to control sexual behaviours in male goldfish (*Carassius auratus*).⁷⁴ Further, contaminants that impact conspecific detection can also change aggression and dominance hierarchies among individuals. Sloman and colleagues^{75–78} found that rainbow trout (*Oncorhynchus mykiss*) exposed to cadmium, which damages the olfactory epithelium, were less aggressive towards an unexposed rival and this reduction in aggression allowed groups of 10 fish to form stable



dominance hierarchies faster when exposed. In contrast, round goby (*Neogobius melanostomus*) collected from sites with heavy sediment contamination (PCBs, PAHs and the metals arsenic, cadmium, iron, lead, zinc) took longer to establish dominance with more aggression in a resource contest, and this dominance relationship was less stable when compared to pairs collected from cleaner areas.⁷⁹ Finally, environmentally relevant levels of 17-alpha ethinylestradiol not only have direct effects on reproduction, it also alters courtship behaviour of fish and amphibians. Exposure reduced nest building and courtship behaviours, but most importantly, altered sexual selection of a marine fish sand goby (*Pomatoschistus minutus*)^{80,81} and modified the mating calls of the amphibian *Xenopus laevis*.⁸²

In the wild, disruption of reproductive behaviours as a result of exposure to endocrine disrupting compounds has been documented in a range of species. For example, exposures to organochlorine pesticides reduced herring gull (*Larus smithsonianus*) and merlin (*Falco columbarius*) offspring incubation time, nest attendance, and defence behaviours as a result of the complex interactions of these pesticides with acetylcholine and estrogen signalling.^{83–86} More recently, male fish (*Pimephales promelas*; *Gasterosteus aculeatus*) exposed to municipal wastewater treatment plant effluents with high concentrations of estrogen-active compounds were less likely to secure a breeding site and performed fewer courtship behaviours.^{87–89} Sources of contamination in the wild, such as wastewater effluents, are often complex mixtures of compounds that may have many mechanisms of action beyond the endocrine system. As such, researchers have also documented reduced aggression⁹⁰ and increased courtship behaviours⁹¹ following exposure to wastewater.

Feeding and anti-predator behaviours. Chemical contaminants can also affect interactions between organisms and therefore impact foraging for prey and anti-predator responses. Predator-prey interactions have important implications for animal fitness and provide a link between individual-level and population-level responses to contamination.^{92–94} Increases or decreases in predation can have cascading effects on animals at other trophic levels, leading to possible indirect effects. Contaminants may directly impair movement abilities, making animals less adept at capturing prey and/or escaping predators, and this has been noted in birds, fish, and reptiles exposed to acetylcholinesterase-inhibiting pesticides.^{83,95} For example, the antidepressants fluoxetine⁹⁶ and sertraline⁹⁷ and the antihistamine diphenhydramine⁹⁸ significantly altered feeding rates of juvenile fathead minnows (*Pimephales promelas*) at levels lower than growth response thresholds were observed.

Furthermore, neurologically active contaminants, such as the antidepressant fluoxetine, which modulate adaptive stress or fear responses to environmental threats also have great potential to impact anti-predator responses. Indeed, several studies have found that guppies (*Poecilia reticulata*) exposed to environmentally relevant concentrations of fluoxetine show delayed escape responses to a predator, froze for longer and spent more time under cover.^{99–101} Similarly, Brodin and colleagues¹⁰² found that European perch (*Perca fluviatilis*) exposed to the psychoactive pharmaceutical oxazepam were

more active and bold in a novel environment, potentially increasing their susceptibility to predation. Interestingly, using a multi-stressor approach, Saaristo and colleagues⁶ found the initial movements of European perch to be significantly affected, and fish became bolder (*i.e.* entered the white background) in the oxazepam, high temperature, and predation treatments. When internal plasma doses of the antidepressant sertraline exceeded human therapeutic plasma levels, serotonin reuptake transporter binding, an anchor 1 molecular initiation event, and anxiety behaviour, and anchor 2 responses with an adverse outcome pathway context, of adult male fathead minnows were significantly altered.¹⁰³

Contaminants that interfere with the ability to detect chemical cues will also alter predator-prey behaviours by inhibiting the ability to sense prey, detect predators, or group with conspecifics to avoid threat. For example, fathead minnows (*Pimephales promelas*) exposed to copper and goldfish (*Carassius auratus*) exposed to the herbicide atrazine both failed to detect and react to conspecific skin-based alarm cues.^{104,105} Also pharmaceuticals have been found to disrupt infochemicals, for example the pharmaceutical propranolol lowering the anti-predator response of amphipods to predator cues, albeit at rather high concentrations.¹⁰⁶ In addition, the antidepressant fluoxetine was shown to cause elevated alarm responses in Arabian killifish (*Aphanius dispar*)¹⁰⁷ and slower predator avoidance response in larval fathead minnows (*Pimephales promelas*).¹⁰⁸ McPherson and colleagues¹⁰⁹ demonstrated that Iowa darter (*Etheostoma exile*) could detect and avoid traps treated with a conspecific alarm cue in a clean lake, but failed to avoid similarly treated traps in a contaminated lake. This result demonstrates that some behavioural endpoints that are observed under the controlled conditions of the laboratory also occur under natural conditions. Finally, a recent study showed that the chemosensory perception of predators by the gray tree frog (*Dryophytes versicolor*) was reduced by 50% when tadpoles were housed in polluted stream water and wastewater effluent compared to clean tap water.¹¹⁰

Automated response to stimuli (high-throughput behavioural analysis). An increasing number of studies are making use of automated high throughput devices that stimulate experimental subjects to evoke behaviours such as startle responses.³ These startle responses can be used a proxy to determine the efficiency of predatory escape response, however, more work needs to be done to link these 'in a box' experiments with survival responses in wild. There are an increasing number of platforms on the market which allow for high-throughput analysis of behaviours using video cameras (*e.g.* ToxTrac,¹¹¹ Daniovision™, ZebraBox™ and Zantiks™) whilst others make use of an organism's movement through electrical conductivity.¹¹² Some platforms can stimulate organisms through lighting, vibration/noise and electricity and can measure an enormous amount of different behaviours from basic speed/distance to turn angles and rates of acceleration. Longer term experiments can be conducted so that more complex behaviours can be studied such as learning and memory.¹¹³ Others can be adapted to multiple ways to measure attraction or repulsion from various stimuli. Within the fields of clock gene



biology¹¹⁴ much has been made of devices that utilize beam-breakers (TriKinetics™) of infrared light in apparatus which can be adapted to the size of the organisms. Other labs are generating lab-on-a-chip technologies¹¹⁵ which can also be adapted in various ways for the endpoints of interest. These technologies have the capacity to increase the throughput of behavioural research with not only increasing speed, productivity and ultimately lower research costs, but also by reducing the subjectivity of behavioural research.¹¹⁶

Discussion

Chemicals regulation is based on a combination of scientific knowledge, pragmatic considerations and policy decisions. With advancements in science and changes in community values comes new regulations, albeit with the delay that can be expected in a democratic society where legal certainty is important.

Some might argue that the precautionary principle, a cornerstone in the EU chemicals regulation, could be used to justify the use of behavioural endpoints in environmental risk assessments. According to the European Commission, the precautionary principle should be used when there is evidence of potential harm, but there is not enough information to determine the risk with sufficient certainty.¹¹⁷ However, according to a recent review of the operation of the REACH regulation, risk management actions evoking the precautionary principle have been limited. This is because scientific uncertainties are not assessed to the extent needed for decisions based on the precautionary principle.^{118,119}

This study concludes that the only legal obstacle to the use of behavioural endpoints in EU chemicals regulation is whether an endpoint could be considered to be relevant at the population level or not. Since there is no agreed method for establishing this connection, this decision must currently be taken case-by-case. Here, the individual assessor and/or the responsible institution can play a large role. A broader interpretation of what is relevant to populations may result in more behavioural studies being considered for chemicals regulation. A consequence of allowing for additional studies is that more resources are needed for literature search, evaluation and assessment. Since the majority of studies investigating behavioural effects are non-standard studies with possibly new test designs, methodologies and test species, increased complexity in the study evaluations is expected. Absence of standardized methods likely prevents assessors/institutions from including behavioural studies in chemical assessments.

This study also concludes that the current use of ecotoxicity studies investigating behavioural endpoints in regulation of chemicals is low. We do not claim to provide a complete overview of the current use of behavioural endpoints but since many regulators and risk assessors we discussed the matter with had difficulties remembering if behavioural studies had been used in chemicals regulation, stating that they are rarely used is appropriate. The available examples show that behavioural studies have been evaluated for use in several different regulatory frameworks, and that studies have been used as both key

and supporting evidence. The available examples also show that studies have been disregarded due to low reliability as well as insufficient reporting. This is a problem that applies to peer-reviewed studies in general, and not behavioural studies in particular, and it can be improved by scientific journals introducing reporting requirements and by increased awareness of the regulatory system among academic researchers.^{20,120,121} What cannot be captured by this study are the studies that were disregarded in the early steps of the assessment process, e.g. because they were not searched for, or they were disregarded because the endpoint was not considered relevant.

Interestingly, behavioural studies are used in the regulation of pesticides as evidence of the efficiency of the substance. For example, behavioural endpoints are used when assessing repellents for ticks, mosquitoes, and birds.¹²² This appears important because pesticides are designed to be biologically active compounds that elicit effects through specific mechanisms of action. For plant production products, the most common use of animal behaviour studies is as evidence that the natural behaviour of animals will prevent them from being exposed to chemicals. For example, studies showing that certain birds and rodents prefer to avoid short grass due to lack of food and hiding places are used by chemical companies as evidence that the animals will not be exposed to pesticides used on greens at golf courses.¹²³ This use of behavioural studies shows that these endpoints are accepted as evidence in pesticide assessments, both by regulators and by the chemical industry.

Finally, this study concludes that there is evidence that behavioural endpoints are relevant for population level effects. This conclusion is supported by a recent theme issue of the *Philosophical Transactions of the Royal Society B* where the connection between animal behaviour and dynamics of populations and communities was highlighted. The editorial stated that “*Behavioural ecology has accumulated a rich toolbox for quantifying how the main behaviours of animals relating to foraging, predation, mating, parental care, communication and sociality are affected by the current threats to biodiversity, notably habitat loss and fragmentation, overexploitation, climate change, pollution, disease, and invasive species. This provides a firm foundation for a bottom-up approach to understanding human impacts on the natural world*”.¹²⁴

In summary, there are three principal obstacles to improved use of behavioural endpoints in environmental risk assessments of chemicals. Below these obstacles are discussed together with possible ways forward (Fig. 1).

(1) Lack of promotion of behavioural endpoints in chemicals legislation

In EU chemicals regulation, the chemical industry is obliged or recommended by legislators to submit particular types of ecotoxicological studies. The regulatory guidance documents mention either specific standard studies, which currently lack behavioural endpoints (with a few exceptions), or specific effects, focusing on survival rate, reproduction rate and developmental effects. Consequently, the chemical industry is not encouraged to submit studies investigating behavioural endpoints, even though such endpoints might be equally or more relevant for a particular chemical. A way forward is to





Fig. 1 Schematic overview of the risk management process and how the use of behavioral endpoints can improve in decision-making. Environmental legislation and guidance documents stipulate which studies that need to be included in the risk assessment and provide guidance for the study evaluation process. The environmental risk assessment consists of standard and non-standards studies of sufficient reliability and relevance. A risk management decision is taken, partly based on the conclusions from the risk assessment.

adjust the guidance regarding the recommended effects of relevance, whenever justified (see obstacle 3).

Future efforts are warranted to develop standard behavioural studies for regulatory applications, or to add behavioural endpoints to already established standard studies. To accomplish this, it will be important to catalogue historical and more contemporary laboratory behavioural responses to contaminants. From this, information on specific behaviours of representative taxa should be identified that are predictive of ecologically important adverse outcomes in the field. Research will be needed to accomplish this goal. These efforts should initially employ specifically acting chemicals that elicit, or are anticipated to cause, behaviour changes through specific pathways because such observations could provide diagnostic positive controls for future studies of other chemicals. Laboratory studies aimed at uncovering the degree and nature of variability should be conducted to ensure that experimental design considerations (*e.g.*, statistical power), repeatability and practices are adequate for regulatory adoption.

(2) Low use of non-standard studies in chemicals regulation

Since very few standard studies measure behavioural effects, an immediate increased use of behavioural endpoints in environmental risk assessments would depend on the use of non-standard studies. To increase the use of non-standard studies legislation must encourage or demand inclusion of such approaches, when available. One example of this practice is reflected in the draft of the new guideline on the environmental risk assessment of medicinal products for human use: “*Behaviour is an example of an ecotoxicological endpoint not yet established as a reliable and standardised endpoint. It may however be very relevant for neuro-active substances and when standardised guidelines become available, be taken up in a tailored risk assessment scheme for neuro-active substances*”.¹²⁵ However, it should be noted that adding non-standard studies to environmental risk assessments puts additional demand on those evaluating studies at chemicals

companies or regulatory agencies due to the variety of test designs and test organisms used. Here regulatory guidance that promotes thorough and robust evaluations is needed.¹²⁶

Non-standard studies are primarily performed by academic researchers, during curiosity driven or exploratory research, as opposed to standard studies performed by, or on behalf of, the chemical industry to comply with test demands in chemicals regulations. Recent debated cases show that the use of non-standard studies in chemicals regulation have not been without obstacles, often resulting in disqualification of studies.¹²⁷ Common reasons why non-standard studies have been assessed as “not reliable” are due to shortages in the test design or in information about how it was performed. For example, some studies are considered unreliable because of a lack of measured tested concentrations, too few replicates, or too few tested concentrations. But other studies have also been assessed as “not assignable” due to a lack of transparency in the description of the methodology and results. However, these problems have a workable solution that does not necessary request that the original research idea is compromised. In fact, the scientific quality of the study can even benefit from it. Researchers can adjust the methodology, performance and presentation of the study in minor ways thereby ensuring that the regulatory requirements are fulfilled, *e.g.* by increasing the statistical power by adjusting the number of replicates and by improving the transparency of the reporting.¹²⁸

(3) Lack of clarification of the importance of behavioural endpoints at the population level

When a behavioural study is disqualified for regulatory use because the endpoint used is not considered relevant at the population level, the possibility for the individual researcher to influence the potential regulatory impact of the study is low. Theoretically, a detailed scientific explanation of how a behavioural change in an individual can result in effects at the population and/or community level would be sufficient. A



conceptual framework for this has been developed, intended for researchers that would like to expand their investigation from individual level to community level.⁶ Tools like Systematic Reviews and Systematic Maps has been suggested to facilitate use of behavioural studies in conservation management and policy.¹²⁹ In EU and the U.S. there is a growing interest for such methodology in chemicals regulation.¹³⁰

Currently, to claim that behavioural endpoints are relevant at the population level (according to the current EU regulation), their connection to survival rate, reproduction rate and developmental effects needs to be established. Consequently, there may be a lower regulatory threshold for behavioural endpoints closely related to these effects, *e.g.* as shown in the use of behavioural endpoints in two standard toxicity studies used in human health risk assessments to assess reproductive and developmental effects (OECD TG 416 and 426). The results in the present study show that what is considered relevant at the population level may change between jurisdictions. For example, some legislation has opened up for use of some behavioural endpoints while others only recommend the more traditional endpoints. What is considered population relevant may also change over time. As one example, adverse histopathological effects on the kidney was recently used as key evidence when an EQS for the pharmaceutical diclofenac was set.¹³¹ Harmonisation among regulatory frameworks represents a useful goal.

Conclusions and recommendations

There are three main conclusions from this study:

- (1) Behavioural endpoints representing a wide range of behaviours are relevant at the population level and should therefore be used in chemicals regulation.
- (2) There is currently a low use of ecotoxicity studies investigating behavioural endpoints in chemicals regulation.
- (3) There are no apparent legal obstacles to be able to use behavioural studies in EU chemicals regulation, except that the endpoints need to be considered relevant at the population level.

In order to improve the current use of behavioural studies in regulation of chemicals, all relevant stakeholders are recommended to contribute:

- Researchers need to write robust and reliable papers, and also carefully explain the relevance of the study for regulation of chemicals.
- Editors and scientific journals need to promote and demand detailed study reports and the highest level of study reliability and transparency, including responsible data sharing.
- Regulatory agencies and the chemical industry need to embrace new behavioural endpoints and develop standardized methods for behaviours when relevant at the population level.

Disclaimer

This paper is the work of the authors and the opinions herein are those of the authors alone and do not necessarily reflect the view or opinions of their institutions.

Authors contributions

Ågerstrand devised the study and drafted the manuscript. All authors contributed to the manuscript and approved the final manuscript. All data is available in the manuscript.

Conflicts of interests

The authors have no competing financial interests to declare.

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References

- 1 J. F. Postma and C. M. Keijzers, *Behavior as response parameter a literature review on the relevance for population sustainability*, Ecofide, 2014, http://www.ecofide.nl/uploads/files/2014-behavior_as_response_parameter_final_report.pdf.
- 2 A. V. Kalueff, A. M. Stewart and R. Gerlai, Zebrafish as an emerging model for studying complex brain disorders, *Trends Pharmacol. Sci.*, 2014, **35**(2), 63–75.
- 3 G. Pyle and A. T. Ford, Behaviour revised: contaminant effects on aquatic animal behaviour, *Aquat. Toxicol.*, 2017, **182**, 226–228.
- 4 A. Gerhardt, Aquatic Behavioral Ecotoxicology—Prospects and Limitations, *Hum. Ecol. Risk Assess. Int. J.*, 2007, **13**(3), 481–491.
- 5 G. Pyle and A. T. Ford, Behaviour revised: contaminant effects on aquatic animal behaviour, *Aquat. Toxicol.*, 2017, **182**, 226–228.
- 6 M. Saaristo, B. Tomas, B. Sigal, G. Bertram Michael, W. Brooks Bryan, M. Ehlman Sean, *et al.*, Direct and indirect effects of chemical contaminants on the behaviour, ecology and evolution of wildlife, *Proc. R. Soc. B*, 2018, **285**(1885), 20181297.
- 7 X. Zhang, Environmental DNA Shaping a New Era of Ecotoxicological Research, *Environ. Sci. Technol.*, 2019, **53**(10), 5605–5612.
- 8 C. Amiard-Triquet, Behavioral Disturbances: The Missing Link between Sub-Organismal and Supra-Organismal Responses to Stress? Prospects Based on Aquatic Research, *Hum. Ecol. Risk Assess. Int. J.*, 2009, **15**(1), 87–110.
- 9 S. Brander, S. Hecht and K. Kuivila, The challenge: “Bridging the gap” with fish: advances in assessing exposure and effects across biological scales, *Environ. Toxicol. Chem.*, 2015, **34**(3), 459.
- 10 S. D. Melvin and S. P. Wilson, The utility of behavioral studies for aquatic toxicology testing: a meta-analysis, *Chemosphere*, 2013, **93**(10), 2217–2223.



- 11 B. B. M. Wong and U. Candolin, Behavioral responses to changing environments, *Behav. Ecol.*, 2015, **26**(3), 665–673.
- 12 P. D. Robinson, Behavioural toxicity of organic chemical contaminants in fish: application to ecological risk assessments (ERAs), *Can. J. Fish. Aquat. Sci.*, 2009, **66**(7), 1179–1188.
- 13 J. S. Weis, G. Smith, T. Zhou, C. Santiago-Bass and P. Weis, Effects of Contaminants on Behavior: Biochemical Mechanisms and Ecological Consequences, *BioScience*, 2001, **51**(3), 209–217.
- 14 W. B. Steele, L. A. Kristofco, J. Corrales, G. N. Saari, S. P. Haddad, E. P. Gallagher, *et al.*, Comparative behavioral toxicology with two common larval fish models: exploring relationships among modes of action and locomotor responses, *Sci. Total Environ.*, 2018, **640–641**, 1587–1600.
- 15 European Chemicals Agency, *REACH Guidance documents*, Helsinki, Finland, 2011, <https://echa.europa.eu/guidance-documents/guidance-on-reach>.
- 16 European Chemicals Agency, *Guidance on information requirements and chemical safety assessment. Part B: Hazard assessment, Version 2.1*, 2011, http://echa.europa.eu/documents/10162/13643/information_requirements_part_b_en.pdf.
- 17 OECD, *OECD Guideline for testing of chemicals. Daphnia sp., Acute immobilisation test*, OECD 202, Paris, France, 2004.
- 18 US EPA, *Ecological Effects Test Guidelines OPPTS 850.1400, Fish Early-Life Stage Toxicity Test*, 1996.
- 19 European Chemicals Agency, *REACH Guidance documents*, Helsinki, Finland, 2011, <https://echa.europa.eu/guidance-documents/guidance-on-reach>.
- 20 M. Ågerstrand, A. Sobek, K. Lilja, M. Linderoth, L. Wendt-Rasch, A.-S. Wernersson, *et al.*, An academic researcher's guide to increased impact on regulatory assessment of chemicals, *Environ. Sci.: Processes Impacts*, 2017, **19**(5), 644–655.
- 21 OECD, *Work plan for the Test Guidelines Programme (TGP)*, 2016, https://www.oecd.org/env/ehs/testing/TGP%20workplan_July%202016-for%20publication.pdf.
- 22 WHO, *Principles and methods for the risk assessment of chemicals in food*, WHO. 2009, <http://www.who.int/foodsafety/publications/chemical-food/en/>.
- 23 European Food Safety Authority, *Guidance on tiered risk assessment for plant protection products for aquatic organisms in edge-of-field surface waters*, 2013, <http://www.efsa.europa.eu/en/efsajournal/doc/3290.pdf>.
- 24 European Commission, *Guidance Document on Risk Assessment for Birds and Mammals Under Council Directive 91/414/EEC*, 2002, http://ec.europa.eu/food/plant/pesticides/guidance_documents/docs/wrkd0c19_en.pdf.
- 25 European Food Safety Authority, Guidance on the risk assessment of plant protection products on bees (*Apis mellifera*, *Bombus* spp. and solitary bees), *EFSA J.*, 2013, **11**(7), 3295.
- 26 European Food Safety Authority, Guidance for the identification of endocrine disruptors in the context of Regulations (EU) No 528/2012 and (EC) No 1107/2009, *EFSA J.*, 2018, **16**(6), 5311.
- 27 Canadian Council of Ministers of Environment, *A protocol for the derivation of water quality guidelines for the protection of aquatic life*, 2007, p. 37, [Guidance Document]. Retrieved from Canadian Council of Ministers of the Environment website: http://www.ccme.ca/files/Resources/supporting_scientific_documents/protocol_aql_2007e.pdf.
- 28 G. I. Petersen, Eels (*Anguilla anguilla*) avoidance test with MTBE, *Report to EFOA – European Fuel Oxygenates Association*, Brussels, Belgium, 2003.
- 29 European Chemicals Bureau, *European Union Risk Assessment Report. Tert-butyl methyl ether*, Joint Research Center, European Communities, 2002.
- 30 European Chemicals Bureau, *Addendum to the Environmental Risk Assessment for tert-butyl methyl ether (MTBE)*.
- 31 Danish Environmental Protection Agency, Determining quality standards for the aquatic environment, in *Methyl tertiary-Butyl Ether (MTBE), (CAS no. 1634-04-4)*, Danish, 2009.
- 32 H. L. Schoenfuss, S. E. Bartell, T. B. Bistodeau, R. A. Cediell, K. J. Grove, L. Zintek, *et al.*, Impairment of the reproductive potential of male fathead minnows by environmentally relevant exposures to 4-nonylphenol, *Aquat. Toxicol.*, 2008, **86**(1), 91–98.
- 33 A. J. Ward, A. J. Duff and S. Currie, The effects of the endocrine disrupter 4-nonylphenol on the behaviour of juvenile rainbow trout (*Oncorhynchus mykiss*), *Can. J. Fish. Aquat. Sci.*, 2006, **63**(2), 377–382.
- 34 M. Cardinali, F. Maradonna, I. Olivotto, G. Bortoluzzi, G. Mosconi, A. M. Polzonetti-Magni, *et al.*, Temporary impairment of reproduction in freshwater teleost exposed to nonylphenol, *Reprod. Toxicol.*, 2004, **18**(4), 597–604.
- 35 European Chemicals Agency, *Support document for identification of 4-nonylphenol, branched and linear, as substances of very high concern because due to their endocrine disrupting properties they cause probable serious effects to the environment which give rise to an equivalent level of concern to those of CMRs and PBT/vPvBs*, 2012.
- 36 R. B. Lauridsen and N. Friberg, Stream macroinvertebrate drift response to pulsed exposure of the synthetic pyrethroid lambda-cyhalothrin, *Environ. Toxicol.*, 2005, **20**(5), 513–521.
- 37 L.-H. Heckmann, N. Friberg and H. W. Ravn, Relationship between biochemical biomarkers and pre-copulatory behaviour and mortality in *Gammarus pulex* following pulse-exposure to lambda-cyhalothrin, *Pest Manage. Sci.*, 2005, **61**(7), 627–635.
- 38 U. Nørnum, N. Friberg, M. R. Jensen, J. M. Pedersen and P. Bjerregaard, Behavioural changes in three species of freshwater macroinvertebrates exposed to the pyrethroid lambda-cyhalothrin: laboratory and stream microcosm studies, *Aquat. Toxicol.*, 2010, **98**(4), 328–335.



- 39 Swedish Chemicals Agency, *Renewal Assessment Report under Regulation (EC) 1107/2009. lambda-cyhalothrin. Active substance data*, 2013.
- 40 Swedish Chemicals Agency, *Fastac 50, product report*, 2009.
- 41 P. Woin and P. Larsson, Phthalate esters reduce predation efficiency of dragonfly larvae, Odonata; Aeshna, *Bull. Environ. Contam. Toxicol.*, 1987, **38**(2), 220–225.
- 42 European Chemicals Bureau, *European Union Risk Assessment Report bis(2-ethylhexyl) phthalate (DEHP)*, 2008, <https://echa.europa.eu/documents/10162/e614617d-58e7-42d9-b7fb-d7bab8f26feb>.
- 43 A.-K. Eriksson Wiklund, T. Börjesson and S. J. Wiklund, Avoidance response of sediment living amphipods to zinc pyrethrin as a measure of sediment toxicity, *Mar. Pollut. Bull.*, 2006, **52**(1), 96–99.
- 44 Swedish Chemicals Agency, *Competent Authority Report for Copper Pyrethrin (PT 21); Document II A: Effect assessment for the active substance*, Final CAR, 2014.
- 45 H.-J. Klimisch, M. Andreae and U. Tillmann, A Systematic Approach for Evaluating the Quality of Experimental Toxicological and Ecotoxicological Data, *Regul. Toxicol. Pharmacol.*, 1997, **25**(1), 1–5.
- 46 US EPA, *Biological Evaluation under the Endangered Species Act for chlorpyrifos, malathion, and diazinon*, 2015, <https://www.epa.gov/endangered-species/implementing-nas-report-recommendations-ecological-risk-assessment-endangered-and>.
- 47 European Commission, *Commission regulation (EU) No 544/2011 of 10 June 2011 implementing Regulation (EC) No 1107/2009 of the European Parliament and of the Council as regards the data requirements for active substances*, 2011, <https://eur-lex.europa.eu/LexUriServ/LexUriServ.do?uri=OJ:L:2011:155:0001:0066:EN:PDF>.
- 48 European Commission, *Regulation (EU) No 528/2012 of the European Parliament and of the Council of 22 May 2012 concerning the making available on the market and use of biocidal products*, 2012, <https://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:02012R0528-20140425&from=SV>.
- 49 European Chemicals Agency, *Directive 98/8/EC concerning the placing biocidal products on the market Inclusion of active substances in Annex I or IA to Directive 98/8/EC. Assessment Report lambda-cyhalothrin*, 2011, <https://echa.europa.eu/documents/10162/0c9351c8-a352-6a0e-0ae9-99db17768b70f>.
- 50 European Chemicals Agency, *Directive 98/8/EC concerning the placing biocidal products on the market Inclusion of active substances in Annex I or IA to Directive 98/8/EC Assessment Report Deltamethrin*, 2011, <https://echa.europa.eu/documents/10162/56abada9-e4bd-ee0e-9426-134d23c6eab1>.
- 51 A. Beronius, A. Hanberg, R. Heimier and H. Håkansson, *Risk assessment of developmental neurotoxicity: Evaluation of the OECD TG 426 test guideline and guidance documents*, Karolinska Institutet, 2013, <https://ki.se/sites/default/files/2013-1.pdf>.
- 52 A. Beronius, N. Johansson, C. Rudén and A. Hanberg, The influence of study design and sex-differences on results from developmental neurotoxicity studies of bisphenol A: implications for toxicity testing, *Toxicology*, 2013, **311**(1–2), 13–26.
- 53 European Chemicals Agency, *Opinion on an Annex XV dossier proposing restrictions on bisphenol A*, 2015, <https://echa.europa.eu/documents/10162/209030fc-ca4b-4745-97b6-98bfc4d6bdd3>.
- 54 *Behavioural Ecology: An Evolutionary Approach*, ed. J. R. Krebs, and N. B. Davies, 4th edn, 1997, <https://www.wiley.com/en-us/Behavioural+Ecology%3A+An+Evolutionary+Approach%2C+4th+Edition-p-9780865427310>.
- 55 B. R. Anholt and E. E. Werner, Predictable changes in predation mortality as a consequence of changes in food availability and predation risk, *Evol. Ecol.*, 1998, **12**(6), 729–738.
- 56 S. L. Ball and R. L. Baker, Predator-Induced Life History Changes: Antipredator Behavior Costs or Facultative Life History Shifts?, *Ecology*, 1996, **77**(4), 1116–1124.
- 57 T. A. Cowl and A. P. Covich, Predator-Induced Life-History Shifts in a Freshwater Snail, *Science*, 1990, **247**(4945), 949–951.
- 58 B. L. Peckarsky, C. A. Cowan, M. A. Penton and C. Anderson, Sublethal Consequences of Stream-Dwelling Predatory Stoneflies on Mayfly Growth and Fecundity, *Ecology*, 1993, **74**(6), 1836–1846.
- 59 R. C. Miller, The Significance of the Gregarious Habit, *Ecology*, 1922, **3**(2), 122–126.
- 60 D. C. Krakauer, Groups confuse predators by exploiting perceptual bottlenecks: a connectionist model of the confusion effect, *Behav. Ecol. Sociobiol.*, 1995, **36**(6), 421–429.
- 61 J. Krause and G. D. Ruxton, *Living in Groups*, Oxford University Press, Oxford, New York, USA, 2002, p. 224.
- 62 T. Brodin, S. Fogarty, A. Sih and J. Cote, Personality-dependent survival of the invasive mosquitofish: being social can be deadly, *Aquatic Invasions*, 2019, **14**(3), 465–477.
- 63 E. E. Little and S. E. Finger, Swimming behavior as an indicator of sublethal toxicity in fish, *Environ. Toxicol. Chem.*, 1990, **9**(1), 13–19.
- 64 P. D. Robinson, Behavioural toxicity of organic chemical contaminants in fish: application to ecological risk assessments (ERAs), *Can. J. Fish. Aquat. Sci.*, 2009, **66**(7), 1179–1188.
- 65 G. Hellström, J. Klaminder, M. Jonsson, J. Fick and T. Brodin, Upscaling behavioural studies to the field using acoustic telemetry, *Aquat. Toxicol.*, 2016, **170**, 384–389.
- 66 J. Klaminder, M. Jonsson, J. Leander, J. Fahlman, T. Brodin, J. Fick, *et al.*, Less anxious salmon smolt become easy prey during downstream migration, *Sci. Total Environ.*, 2019, **687**, 488–493.
- 67 S. G. Woodman, D. Steinkey, W. A. Dew, S. R. Burket, B. W. Brooks and G. G. Pyle, Effects of sertraline on behavioral indices of crayfish *Orconectes virilis*, *Ecotoxicol. Environ. Saf.*, 2016, **134**, 31–37.



- 68 J. R. Marentette, S. Tong, G. Wang, N. M. Sopinka, M. D. Taves, M. A. Koops, *et al.*, Behavior as biomarker? Laboratory *versus* field movement in round goby (*Neogobius melanostomus*) from highly contaminated habitats, *Ecotoxicology*, 2012, **21**(4), 1003–1012.
- 69 B. B. Chapman, C. Brönmark, J.-Å. Nilsson and L.-A. Hansson, The ecology and evolution of partial migration, *Oikos*, 2011, **120**(12), 1764–1775.
- 70 T. Avgar, G. Street and J. M. Fryxell, On the adaptive benefits of mammal migration, *Can. J. Zool.*, 2013, **92**(6), 481–490.
- 71 F. Pulido, Evolutionary genetics of partial migration – the threshold model of migration revis(it)ed, *Oikos*, 2011, **120**(12), 1776–1783.
- 72 L. Ellis, Dominance and reproductive success among nonhuman animals: a cross-species comparison, *Ethol. Sociobiol.*, 1995, **16**(4), 257–333.
- 73 A. J. W. Ward, J. Duff Alison, S. Horsfall Jennifer and S. Currie, Scents and scents-ability: pollution disrupts chemical social recognition and shoaling in fish, *Proc. R. Soc. B*, 2008, **275**(1630), 101–105.
- 74 J. A. Mennigen, W. E. Lado, J. M. Zamora, P. Duarte-Guterman, V. S. Langlois, C. D. Metcalfe, *et al.*, Waterborne fluoxetine disrupts the reproductive axis in sexually mature male goldfish, *Carassius auratus*, *Aquat. Toxicol.*, 2010, **100**(4), 354–364.
- 75 K. A. Sloman, Effects of trace metals on salmonid fish: the role of social hierarchies, *Appl. Anim. Behav. Sci.*, 2007, **104**(3), 326–345.
- 76 K. A. Sloman, D. W. Baker, C. G. Ho, D. G. McDonald and C. M. Wood, The effects of trace metal exposure on agonistic encounters in juvenile rainbow trout, *Oncorhynchus mykiss*, *Aquat. Toxicol.*, 2003, **63**(2), 187–196.
- 77 K. A. Sloman, G. R. Scott, Z. Diao, C. Rouleau, C. M. Wood and D. G. McDonald, Cadmium affects the social behaviour of rainbow trout, *Oncorhynchus mykiss*, *Aquat. Toxicol.*, 2003, **65**(2), 171–185.
- 78 K. A. Sloman, O. Lepage, J. T. Rogers, C. M. Wood and S. Winberg, Socially-mediated differences in brain monoamines in rainbow trout: effects of trace metal contaminants, *Aquat. Toxicol.*, 2005, **71**(3), 237–247.
- 79 N. M. Sopinka, J. R. Marentette and S. Balshine, Impact of contaminant exposure on resource contests in an invasive fish, *Behav. Ecol. Sociobiol.*, 2010, **64**(12), 1947–1958.
- 80 M. Saaristo, J. A. Craft, K. K. Lehtonen and K. Lindström, Sand goby (*Pomatoschistus minutus*) males exposed to an endocrine disrupting chemical fail in nest and mate competition, *Horm. Behav.*, 2009, **56**(3), 315–321.
- 81 M. Saaristo, J. A. Craft, K. K. Lehtonen and K. Lindström, Exposure to 17 α -ethinyl estradiol impairs courtship and aggressive behaviour of male sand gobies (*Pomatoschistus minutus*), *Chemosphere*, 2010, **79**(5), 541–546.
- 82 F. Hoffmann and W. Kloas, Estrogens Can Disrupt Amphibian Mating Behavior, *PLoS One*, 2012, **7**(2), e32097.
- 83 C. E. Grue, P. L. Gibert and M. E. Seeley, Neurophysiological and Behavioral Changes in Non-Target Wildlife Exposed to Organophosphate and Carbamate Pesticides: Thermoregulation, Food Consumption, and Reproduction, *Am. Zool.*, 1997, **37**(4), 369–388.
- 84 D. M. Fry, Reproductive Effects in Birds Exposed to Pesticides and Industrial Chemicals, *Environ. Health Perspect.*, 1995, **103**, 165–171.
- 85 D. B. Peakall, Behavioral responses of birds to pesticides and other contaminants, in *Residue Reviews*, ed. F. A. Gunther, Springer, New York, 1985, pp. 45–77.
- 86 G. A. Fox and T. Donald, Organochlorine Pollutants, Nest-Defense Behavior and Reproductive Success in Merlins, *Condor*, 1980, **82**, 81–84.
- 87 N. Garcia-Reyero, C. M. Lavelle, B. L. Escalon, D. Martinović, K. J. Kroll, P. W. Sorensen, *et al.*, Behavioral and genomic impacts of a wastewater effluent on the fathead minnow, *Aquat. Toxicol.*, 2011, **101**(1), 38–48.
- 88 D. Martinović, W. T. Hogarth, R. E. Jones and P. W. Sorensen, Environmental estrogens suppress hormones, behavior, and reproductive fitness in male fathead minnows, *Environ. Toxicol. Chem.*, 2007, **26**(2), 271–278.
- 89 M. Sebire, I. Katsiadaki, N. G. H. Taylor, G. Maack and C. R. Tyler, Short-term exposure to a treated sewage effluent alters reproductive behaviour in the three-spined stickleback (*Gasterosteus aculeatus*), *Aquat. Toxicol.*, 2011, **105**(1), 78–88.
- 90 E. S. McCallum, E. Krutzmann, T. Brodin, J. Fick, A. Sundelin and S. Balshine, Exposure to wastewater effluent affects fish behaviour and tissue-specific uptake of pharmaceuticals, *Sci. Total Environ.*, 2017, **605–606**, 578–588.
- 91 M. Saaristo, J. Myers, R. Jacques-Hamilton, M. Allinson, A. Yamamoto, G. Allinson, *et al.*, Altered reproductive behaviours in male mosquito fish living downstream from a sewage treatment plant, *Aquat. Toxicol.*, 2014, **149**, 58–64.
- 92 C. Amiard-Triquet, Behavioral Disturbances: The Missing Link between Sub-Organismal and Supra-Organismal Responses to Stress? Prospects Based on Aquatic Research, *Hum. Ecol. Risk Assess. Int. J.*, 2009, **15**(1), 87–110.
- 93 T. Brodin, S. Piovano, J. Fick, J. Klaminder, M. Heynen and M. Jonsson, Ecological effects of pharmaceuticals in aquatic systems—impacts through behavioural alterations, *Philos. Trans. R. Soc., B*, 2014, **369**, 20130580, <https://www.ncbi.nlm.nih.gov/pmc/articles/PMC4213591/>.
- 94 *Behavioural Ecology: An Evolutionary Approach*, ed. J. R. Krebs and N. B. Davies, 4th edn, 1997, <https://www.wiley.com/en-us/Behavioural+Ecology%3A+An+Evolutionary+Approach%2C+4th+Edition-p-9780865427310>.
- 95 S. E. DuRant, W. A. Hopkins and L. G. Talent, Impaired terrestrial and arboreal locomotor performance in the western fence lizard (*Sceloporus occidentalis*) after exposure to an AChE-inhibiting pesticide, *Environ. Pollut.*, 2007, **149**(1), 18–24.
- 96 J. K. Stanley, A. J. Ramirez, C. K. Chambliss and B. W. Brooks, Enantiospecific sublethal effects of the antidepressant fluoxetine to a model aquatic vertebrate and invertebrate, *Chemosphere*, 2007, **69**(1), 9–16.



- 97 T. W. Valenti, P. Perez-Hurtado, C. K. Chambliss and B. W. Brooks, Aquatic toxicity of sertraline to Pimephales promelas at environmentally relevant surface water pH, *Environ. Toxicol. Chem.*, 2009, **28**(12), 2685–2694.
- 98 J. P. Berninger, B. Du, K. A. Connors, S. A. Eytcheson, M. A. Kolkmeier, K. N. Prosser, *et al.*, Effects of the antihistamine diphenhydramine on selected aquatic organisms, *Environ. Toxicol. Chem.*, 2011, **30**(9), 2065–2072.
- 99 M. Pelli and V. P. Connaughton, Chronic exposure to environmentally-relevant concentrations of fluoxetine (Prozac) decreases survival, increases abnormal behaviors, and delays predator escape responses in guppies, *Chemosphere*, 2015, **139**, 202–209.
- 100 M. Saaristo, A. McLennan, C. P. Johnstone, B. O. Clarke and B. B. M. Wong, Impacts of the antidepressant fluoxetine on the anti-predator behaviours of wild guppies (*Poecilia reticulata*), *Aquat. Toxicol.*, 2017, **183**, 38–45.
- 101 J. M. Martin, M. Saaristo, M. G. Bertram, P. J. Lewis, T. L. Coggan, B. O. Clarke, *et al.*, The psychoactive pollutant fluoxetine compromises antipredator behaviour in fish, *Environ. Pollut.*, 2017, **222**, 592–599.
- 102 T. Brodin, J. Fick, M. Jonsson and J. Klaminder, Dilute Concentrations of a Psychiatric Drug Alter Behavior of Fish from Natural Populations, *Science*, 2013, **339**(6121), 814–815.
- 103 T. W. Valenti, G. G. Gould, J. P. Berninger, K. A. Connors, N. B. Keele, K. N. Prosser, *et al.*, Human Therapeutic Plasma Levels of the Selective Serotonin Reuptake Inhibitor (SSRI) Sertraline Decrease Serotonin Reuptake Transporter Binding and Shelter-Seeking Behavior in Adult Male Fathead Minnows, *Environ. Sci. Technol.*, 2012, **46**(4), 2427–2435.
- 104 W. A. Dew, A. Azizishirazi and G. G. Pyle, Contaminant-specific targeting of olfactory sensory neuron classes: Connecting neuron class impairment with behavioural deficits, *Chemosphere*, 2014, **112**, 519–525.
- 105 P. Saglio and S. Trijasse, Behavioral responses to atrazine and diuron in goldfish, *Arch. Environ. Contam. Toxicol.*, 1998, **35**(3), 484–491.
- 106 A.-K. E. Wiklund, H. Oskarsson, G. Thorsén and L. Kumblad, Behavioural and physiological responses to -pharmaceutical exposure in macroalgae and -grazers from a Baltic Sea littoral community, *Aquat. Biol.*, 2011, **14**(1), 29–39.
- 107 M. J. Barry, Effects of fluoxetine on the swimming and behavioural responses of the Arabian killifish, *Ecotoxicology*, 2013, **22**(2), 425–432.
- 108 M. M. Painter, M. A. Buerkley, M. L. Julius, A. M. Vajda, D. O. Norris, L. B. Barber, *et al.*, Antidepressants at environmentally relevant concentrations affect predator avoidance behavior of larval fathead minnows (*Pimephales promelas*), *Environ. Toxicol. Chem.*, 2009, **28**(12), 2677–2684.
- 109 T. D. McPherson, R. S. Mirza and G. G. Pyle, Responses of wild fishes to alarm chemicals in pristine and metal-contaminated lakes, *Can. J. Zool.*, 2004, **82**(5), 694–700.
- 110 R. R. Troyer and A. M. Turner, Chemosensory Perception of Predators by Larval Amphibians Depends on Water Quality, *PLoS One*, 2015, **10**(6), e0131516.
- 111 A. Rodriguez, H. Zhang, J. Klaminder, T. Brodin, P. L. Andersson and M. Andersson, ToxTrac: A fast and robust software for tracking organisms, *Methods Ecol. Evol.*, 2018, **9**(3), 460–464.
- 112 Z. Ren, J. Zha, M. Ma, Z. Wang and A. Gerhardt, The early warning of aquatic organophosphorus pesticide contamination by on-line monitoring behavioral changes of *Daphnia magna*, *Environ. Monit. Assess.*, 2007, **134**(1), 373.
- 113 M. O. Parker, A. J. Brock, A. Sudwatts and C. H. Brennan, Atomoxetine reduces anticipatory responding in a 5-choice serial reaction time task for adult zebrafish, *Psychopharmacology*, 2014, **231**(13), 2671–2679.
- 114 J. C. Chiu, K. H. Low, D. H. Pike, E. Yildirim and I. Edery, Assaying Locomotor Activity to Study Circadian Rhythms and Sleep Parameters in *Drosophila*, *J. Visualized Exp.*, 2010, (43), 2157, <https://www.ncbi.nlm.nih.gov/pmc/articles/PMC3229366/>.
- 115 R. Cartlidge, D. Nugegoda and D. Wlodkovic, Millifluidic Lab-on-a-Chip technology for automated toxicity tests using the marine amphipod *Allorchestes compressa*, *Sens. Actuators, B*, 2017, **239**, 660–670.
- 116 B. W. Brooks and W. B. Steele, Ecotoxicological Perspectives on Healthcare and the Environment, in *Healthcare and Environmental Contaminants*, ed. A. B. A. Boxall and R. S. Kookana, Elsevier, London, UK; Amsterdam, 2018, ch. 4, pp. 41–67, available from: <http://www.sciencedirect.com/science/article/pii/B9780444638571000048>.
- 117 European Commission, *Communication from the Commission of 2 February 2000 on the precautionary principle*, 2000, http://ec.europa.eu/dgs/health_consumer/library/pub/pub07_en.pdf.
- 118 European Commission, Communication from the Commission to the European Parliament, the council and the European economic and social committee, *Commission General Report on the operation of REACH and review of certain elements*, 2018, <https://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:52018DC0116&from=EN>.
- 119 European Commission, Commission staff document. Accompanying the document: Communication from the Commission to the European Parliament, the council and the European economic and social committee, *Commission General Report on the operation of REACH and review of certain elements*, 2018, https://eur-lex.europa.eu/resource.html?uri=cellar:2834985c-2083-11e8-ac73-01aa75ed71a1.0001.02/DOC_1&format=PDF.
- 120 M. Ågerstrand, S. Christiansen, A. Hanberg, C. Rudén, L. Andersson, S. Andersen, *et al.*, A call for action: Improve reporting of research studies to increase the scientific basis for regulatory decision-making, *J. Appl. Toxicol.*, 2018, **38**(5), 783–785.



- 121 M. Ågerstrand, L. Edvardsson and C. Rudén, Bad Reporting or Bad Science? Systematic Data Evaluation as a Means to Improve the Use of Peer-Reviewed Studies in Risk Assessments of Chemicals, *Hum. Ecol. Risk Assess. Int. J.*, 2013, **20**(6), 1427–1445.
- 122 European Chemicals Regulation, *Guidance on the Biocidal Products Regulation. Volume II Efficacy – Assessment and Evaluation, (Parts B+C)*, 2018.
- 123 European Food Safety Authority, *Risk Assessment for Birds and Mammals*, 2009, <http://www.efsa.europa.eu/en/efsajournal/doc/1438.pdf>.
- 124 J. Bro-Jørgensen, D. W. Franks and K. Meise, Linking behaviour to dynamics of populations and communities: application of novel approaches in behavioural ecology to conservation, *Philos. Trans. R. Soc., B*, 2019, **374**(1781), 20190008.
- 125 European Medicine Agency, *EMA/CHMP/SWP/4447/00 Rev. 1 2 Committee for Medicinal Products for Human Use (CHMP) 3 Guideline on the environmental risk assessment of medicinal products for human use – Draft*, 2019.
- 126 R. Kase, M. Korkaric, I. Werner and M. Ågerstrand, Criteria for Reporting and Evaluating ecotoxicity Data (CRED): comparison and perception of the Klimisch and CRED methods for evaluating reliability and relevance of ecotoxicity studies, *Environ. Sci. Eur.*, 2016, **28**(1), 7, <http://www.enveurope.com/content/28/1/7>.
- 127 R. E. Alcock, B. H. MacGillivray and J. S. Busby, Understanding the mismatch between the demands of risk assessment and practice of scientists – the case of Deca-BDE, *Environ. Int.*, 2011, **37**(1), 216–225.
- 128 C. T. A. Moermond, R. Kase, M. Korkaric and M. Ågerstrand, CRED: Criteria for reporting and evaluating ecotoxicity data, *Environ. Toxicol. Chem.*, 2016, **35**(5), 1297–1309.
- 129 O. Berger-Tal, A. L. Greggor, B. Macura, C. A. Adams, A. Blumenthal, A. Bouskila, *et al.*, Systematic reviews and maps as tools for applying behavioral ecology to management and policy, *Behav. Ecol.*, 2019, **30**(1), 1–8.
- 130 A. A. Rooney, A. L. Boyles, M. S. Wolfe, J. R. Bucher and K. A. Thayer, Systematic Review and Evidence Integration for Literature-Based Environmental Health Science Assessments, *Environ. Health Perspect.*, 2014, 711–718, <http://ehp.niehs.nih.gov/1307972>.
- 131 C. T. A. Moermond and C. E. Smit, Derivation of water quality standards for carbamazepine, metoprolol, and metformin and comparison with monitoring data, *Environ. Toxicol. Chem.*, 2016, **35**(4), 882–888.

