

## PAPER

[View Article Online](#)  
[View Journal](#) | [View Issue](#)Cite this: *RSC Sustainability*, 2024, 2, 1508Fast hydrolysis for chemical recycling of polyethylene terephthalate (PET)<sup>†</sup>Patrícia Pereira, <sup>a</sup> Willem Slear, <sup>a</sup> Angelo Testa, <sup>a</sup> Kevin Reasons, <sup>a</sup> Peter Guirguis, <sup>a</sup> Phillip E. Savage <sup>\*a</sup> and Christian W. Pester <sup>\*ab</sup>

Chemical recycling of post-consumer polyethylene terephthalate (PET) provides a path toward circular production and use of this material. This article reports the effect of reaction severity and PET/water mass ratio on fast hydrolysis of PET, which can lead to complete depolymerization to monomers in tens of seconds. Rapid heating ( $5\text{--}10\text{ }^{\circ}\text{C s}^{-1}$ ) during fast hydrolysis can promote hydrolytic PET depolymerization. This non-isothermal, non-catalytic reaction provided recovery of terephthalic acid (TPA), one of PET's monomers, exceeding 90% for reaction times as short as 75–95 seconds. More severe conditions (higher temperature and longer time) increased TPA decomposition while lower severities reduced PET depolymerization. Initial PET/water mass ratios of 1/10 or 1/8 gave the highest TPA yields and lowest values for the environmental energy impact metric associated with the depolymerization reaction. We determined an environmental energy impact factor of  $436 \pm 66\text{ }^{\circ}\text{C min}$ , which is the lowest to date and surpasses the performance of isothermal PET hydrolysis, whether performed in neutral water (as is done here) or under acid catalysis. Byproducts, including phthalic acid, isophthalic acid, bis(2-hydroxyethyl)terephthalate, and benzoic acid, were also observed during the process.

Received 23rd January 2024  
Accepted 6th April 2024

DOI: 10.1039/d4su00034j

[rsc.li/rscsus](https://rsc.li/rscsus)

## Sustainability spotlight

This study delves into the chemical recycling of post-consumer polyethylene terephthalate (PET). The fast hydrolysis process described herein, uses a rapid heating rate of  $5\text{--}10\text{ }^{\circ}\text{C s}^{-1}$ , and produces over 90% yield of terephthalic acid (TPA) in just 75–95 seconds. This non-isothermal, non-catalytic methodology provides the lowest environmental energy impact reported to date for PET hydrolysis. This work aligns with UN Sustainable Development Goal 12 (Responsible Consumption and Production), by working toward a circular solution for sustainable PET use.

## Introduction

Polyethylene terephthalate (PET) is a widely used thermoplastic, prevalent in consumer products such as water bottles, food containers, and textiles. The recycling rate of postconsumer PET in the United States was 29% in 2021, with most of the waste plastic going to landfill due to costly collecting and sorting processes, and PET's susceptibility to degradation during mechanical recycling.<sup>1,2</sup>

Chemical recycling, *i.e.*, depolymerizing PET into its monomers, has emerged as a potential alternative. PET is a condensation polymer with water being a synthetic byproduct. Hence,

the reverse reaction pathway, hydrolysis of ester bonds, can depolymerize the plastic into its monomers: terephthalic acid (TPA) and ethylene glycol (EG). This is usually performed under isothermal conditions, being neutral, alkaline, or acidic.<sup>3–12</sup>

In preliminary work,<sup>12</sup> we reported that fast hydrolysis of PET in neutral water produced nearly 80% TPA yield in less than one minute. This approach uses rapid heating at about  $5\text{--}10\text{ }^{\circ}\text{C s}^{-1}$  and total reaction times of tens of seconds. The non-isothermal process is akin to fast pyrolysis and fast hydrothermal liquefaction,<sup>13–18</sup> which use rapid heating to decompose the biopolymers in biomass. Conventional neutral hydrolysis or catalyzed hydrolysis under isothermal conditions has an environmental energy impact metric of at least  $10^4\text{ }^{\circ}\text{C min}$ .<sup>9,19</sup> This metric is widely used in this field, and it accounts for environmental impacts related to energy demand, waste production, and product yield. Using microwave heating during base-catalyzed hydrolysis can decrease the environmental energy impact metric by one order of magnitude.<sup>19</sup> However, for catalyst-free, rapid-heating, fast hydrolysis, the environmental energy impact metric can be reduced by two orders of

<sup>a</sup>Department of Chemical Engineering, The Pennsylvania State University, University Park, PA 16802, USA. E-mail: [psavage@psu.edu](mailto:psavage@psu.edu); [pester@psu.edu](mailto:pester@psu.edu)

<sup>b</sup>Department of Chemistry, Department of Materials Science and Engineering, The Pennsylvania State University, University Park, Pennsylvania 16802, USA

<sup>†</sup> Electronic supplementary information (ESI) available: Water and PET loadings. Summary of experimental results for each fast hydrolysis condition examined. Contour plots of yields of byproducts. Environmental energy impact vs. holding time and set point temperature. See DOI: <https://doi.org/10.1039/d4su00034j>

magnitude, indicating it is an even more environmentally friendly approach.<sup>12</sup> Previously, PET hydrolysis has been achieved in under three minutes, however, it has required more complex processes, including the use of microwave heating combined with acid or base catalysts, or a sequential solvolysis approach.<sup>20–22</sup>

This article explores fast hydrolysis systematically over a range of batch holding times from 15 seconds to 3 minutes, a range of heat source set point temperatures (which control the heating rate experienced by the reactor) from 300 °C to 570 °C, and a range of PET/water mass feed ratios from 1/10 to 1/2. The overall goal is to map out the relevant parameter space and identify reaction conditions that give high yields of TPA with low environmental energy impacts.

## Experimental section

### Materials

Post-consumer PET was obtained from green 16.9 oz. Perrier® sparkling water bottles purchased locally. The labels and caps of the bottles were removed, and the bottles were subsequently cut into small chips with average dimensions of  $6 \pm 2$  mm  $\times$   $8 \pm 2$  mm. The bottle body had a thickness of 0.5 mm, while the bottom was slightly thicker, measuring 2 mm. Information about the plastic bottle chips is provided elsewhere.<sup>12</sup> The melting point of the post-consumer PET was determined to be 250 °C using differential scanning calorimetry (DSC). Dimethyl sulfoxide (DMSO), which was used to dissolve reaction products, was purchased from Millipore Sigma. Terephthalic acid (TPA) and isophthalic acid (both 99% purity, TCI), phthalic acid ( $\geq 99.5\%$ , Sigma Aldrich), bis(2-hydroxyethyl)terephthalate (BHET, Sigma Aldrich), and benzoic acid (99%, Thermo Scientific Chemicals) were employed for high performance liquid chromatography (HPLC) calibration purposes. Deionized water, obtained from an in-house water purification system comprising ion exchange, reverse osmosis, high-capacity ion exchange, UV sterilization, and submicron filtration units, served as the reaction medium.

### Experimental procedure for PET fast hydrolysis

Swagelok reactors were assembled from a stainless-steel port connector and two caps, resulting in an internal volume of about 4 mL. Some reactors were equipped with thermocouples that measured the internal temperature. The reactors were loaded with PET and deionized water, sealed, and placed in a Techné fluidized sand bath, which had been preheated to set point temperatures ranging from 300 °C to 570 °C, for batch holding times from 15 to 175 seconds. Higher set point temperatures provided faster heating. Additional runs were done with PET/water mass ratios ( $m_{\text{PET}}/m_{\text{w}}$ ) of 1/8, 1/6, 1/4, 1/3, and 1/2. Table S1† provides information on all reactor loadings. To terminate the reaction, reactors were withdrawn from the sand bath and immediately submerged in room-temperature water. The cooled reactors were then opened. Small amounts of deionized water (10 mL in total) were pipetted into the reactor, withdrawn, and filtered through grade 1 cellulose filter

paper (from Whatman) using a reusable 25 mm filter holder (from Cole-Parmer) to isolate solids. These solids, comprising unconverted PET and the reaction product (as a white solid based on visual observation), were recovered by drying the filters, syringes, and reactors overnight at 80 °C. The aqueous filtrate was dried in an oven at 45 °C overnight to recover any dissolved solids. We performed at least three independent hydrolysis runs at each experimental condition and report the mean values as the best estimates for product yields. Standard deviations were employed to assess run-to-run variability.

### Analysis of hydrolysis products

To dissolve any TPA in the solid-phase, 5 mL of DMSO was added. The resulting solution was filtered using a 25 mm diameter PTFE membrane filter with a pore size of 1  $\mu$ m. The solids remaining after filtration comprised unconverted PET and DMSO-insoluble byproducts, which we refer to as undissolved solids (US). Solids recovered by evaporating the aqueous filtrate were also dissolved in DMSO. Both DMSO solutions were analyzed using a HPLC equipped with a Waters reverse-phase symmetry C18 column (5  $\mu$ m particle size, 150 mm  $\times$  4.6 mm) operating at 40 °C. Detection was performed with a photodiode array detector (SPD-M20A) set at 240 nm. The mobile phase consisted of HPLC-grade acetonitrile at 0.1 mL min<sup>−1</sup> and a 0.1 v% H<sub>3</sub>PO<sub>4</sub> aqueous solution at 0.3 mL min<sup>−1</sup>. A sample injection volume of 1  $\mu$ L was used. Calibration curves were established by analyzing standard solutions with known TPA, phthalic acid, isophthalic acid, BHET, and benzoic acid concentrations in DMSO. These calibrations and HPLC analysis of the reaction products provided the number of moles of each molecular reaction product. Molar yields were calculated as the moles of a given product divided by the moles of PET repeat units initially loaded into the reactor as we reported in a previous publication.<sup>12</sup>

### Machine learning analysis

We used the RandomForestRegressor function in the sklearn module in Python to determine whether the severity index, which combines both temperature and time effects, or the PET/water mass ratio had the larger influence on TPA yield. To enhance the predictive accuracy of the model and to minimize errors, we varied the number of trees from 1 to 1000. We reserved 10% of the dataset to test predictions and another 10% of the remaining data for model optimization. The balance of the dataset was allocated for training the model.

## Results and discussion

### PET hydrolysis

The objective of fast hydrolysis is to accomplish PET depolymerization while minimizing overreaction and formation of undesirable byproducts. Reaction conditions need to be severe enough, with respect to temperature and time, for effective depolymerization, but not so severe that the TPA product decomposes and/or unwanted side reactions occur. Fig. 1 shows that the reactor temperature profiles for the different set point



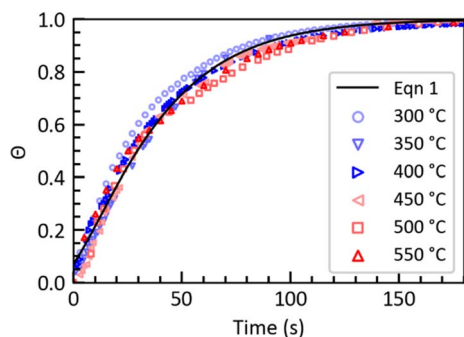


Fig. 1 Measured reactor temperature profiles (discrete points) at different  $T_{SP}$  and calculated via eqn (1) (solid line).

temperatures ( $T_{SP}$ ) collapse onto a single curve when represented by a dimensionless temperature ( $\theta$ ), defined as

$$\theta = \frac{T - T_0}{T_{SP} - T_0} = (1 - e^{-0.031(t*60+9.844)})^2, \quad (1)$$

where  $T$  is temperature within the reactor in °C,  $T_0$  is the initial reactor temperature at time zero (25 °C), and  $t$  is time in minutes. It takes approximately 150 seconds for the internal reactor temperature to reach the sand bath set point temperature.

Table S2† presents the product yields from all the fast hydrolysis experiments. As a check on the completeness of product recovery, we calculated the “ring balance” for each experimental condition. The ring balance is the ratio of the moles of aromatic rings in the quantified products to the moles of aromatic rings loaded into the reactor in the initial PET. The hydrothermal reaction conditions are expected to leave aromatic rings intact. Results from experimental conditions that gave a ring balance that differed by more than  $\pm 15\%$  from 100% or that had a mean value that differed from 100% by a statistically significant margin ( $p < 0.05$ ) were excluded from figures displayed herein.

Table S2† shows that negligible TPA yield (<2%) was observed even after 175 seconds at a set point temperature of 300 °C. This is likely due to the relatively low temperature inside the reactor and the short reaction time, which prevented the depolymerization reaction from proceeding effectively. At

$T_{SP} = 350$  °C,  $Y_{TPA}$  increased to  $11 \pm 10\%$  after 115 seconds and further to  $35 \pm 13\%$  after 175 seconds. At the 350 °C set point, the internal temperature of the reactor exceeds the melting point of PET in about 55 seconds. Hydrolysis of PET in neutral water is much faster above its melting point.

Fig. 2a presents the temporal variation of the TPA yield ( $Y_{TPA}$ ) and yield of undissolved solids ( $Y_{US}$ , wt%) from fast hydrolysis at  $T_{SP} = 400$  °C.  $Y_{TPA}$  increased steadily with time (achieving  $87 \pm 9\%$  at 175 s) and  $Y_{US}$  steadily decreased. At set point temperature of 510 °C (Fig. 2b),  $Y_{TPA}$  reached a maximum and then declined, likely due to TPA degradation. This behavior appeared at  $T_{SP} = 480$  °C to 570 °C.

At set point temperatures above 480 °C, increasing the set point temperature decreases  $Y_{TPA}$  at 175 seconds. This indicates more TPA decomposition under the more severe reaction conditions. In addition, the higher the set point temperature, the less time is required for  $Y_{TPA}$  to achieve a maximum value. For example, at  $T_{SP} = 480$  °C  $Y_{TPA}$  starts decreasing after 95 seconds and at  $T_{SP} = 510$  °C it decreases after 75 seconds. The highest  $Y_{TPA}$  values obtained were  $94 \pm 3\%$  at  $T_{SP} = 510$  °C and 75 seconds and  $93 \pm 4\%$  at  $T_{SP} = 540$  °C and 75 seconds.

The contour plot in Fig. 3a summarizes how the combination of batch holding time and sand bath set point temperature influence the TPA yield from PET fast hydrolysis. Times less than 50 seconds give low  $Y_{TPA}$ , regardless of the  $T_{SP}$ , within the range studied. Times greater than 130 seconds also gave low  $Y_{TPA}$ , provided the  $T_{SP}$  is below 350 °C (where hydrolysis is slow) or above 510 °C (where TPA decomposes). At intermediate temperatures ( $T_{SP} = 400$  to 480 °C),  $Y_{TPA}$  values around 80% are accessible. This range further extends to higher  $T_{SP}$  where yields surpassing 90% are attainable within shorter intervals (e.g., 500 °C, 75 s). It appears that high TPA yields from fast hydrolysis are not confined to a specific set of conditions but are attainable over a wide range of set point temperatures and holding times. This flexibility in operating conditions is an important feature as it can facilitate process optimization at scale to minimize cost and environmental impact.

Fig. 3b presents the contour plot for the yields of undissolved solids from the fast hydrolysis experiments. There is a continuous decrease in the yield of undissolved solids as the reaction time increases for each set point temperature. Higher set point temperatures result in faster conversion of the undissolved

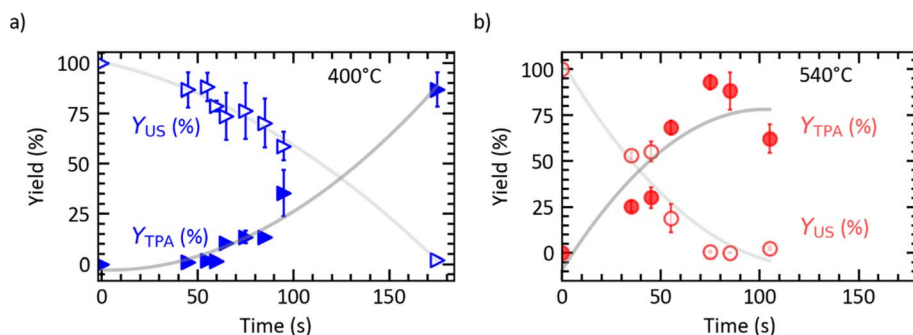


Fig. 2 Effects of batch holding time on yields of TPA and undissolved solids from PET fast hydrolysis at  $T_{SP}$  of (a) 400 °C and (b) 540 °C ( $m_{PET}/m_w = 1/10$ ). The grey lines are a guide to the eyes.



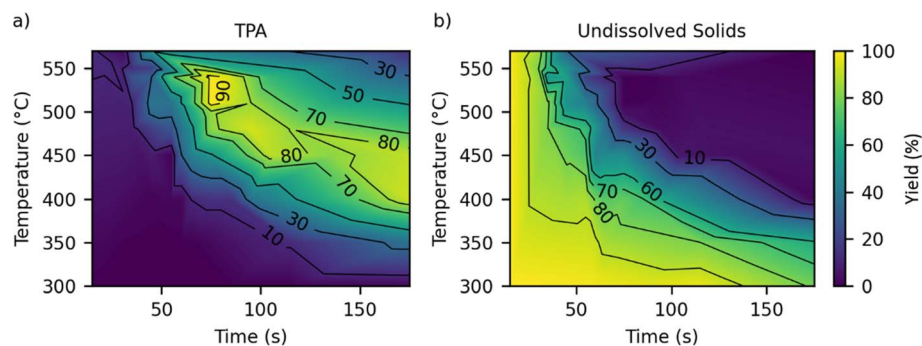


Fig. 3 Contour plot illustrating effects of set point temperature and batch holding time on (a) TPA yield and (b) yield of undissolved solids from fast hydrolysis of PET ( $m_{\text{PET}}/m_{\text{w}} = 1/10$ ).

solids. For example, 175 s, 95 s, and 75 s are required for the yield of undissolved solids to become <5% for  $T_{\text{SP}} = 400$  °C, 480 °C, and 510 °C, respectively.

A typical HPLC chromatogram can be found in Fig. S3†. Fig. S4 and S5† provide contour plots describing the yields of isophthalic acid and benzoic acid from PET hydrolysis at  $m_{\text{PET}}/m_{\text{w}} = 1/10$ . The isophthalic acid yields reach up to  $21 \pm 3\%$  and depend on the reaction conditions. However, because the isophthalic acid content of commercial PET carbonated drink bottles is 1–3 wt%,<sup>23</sup> the isophthalic acid byproduct cannot solely originate from release in commercial bottles during depolymerization. Čolnik *et al.*<sup>24</sup> also observed the production of isophthalic acid during PET hydrolysis, and its yield was also dependent on temperature and reaction time. Further investigations are required to elucidate the pathway(s) for isophthalic acid formation during fast hydrolysis of PET. Table S2† shows that  $Y_{\text{BA}}$  increased with temperature and time, particularly under conditions where the yields of TPA begins decreasing.<sup>25</sup> This behavior is consistent with benzoic acid being a decomposition product of TPA and perhaps isophthalic acid. The presence of these compounds indicates that the TPA produced is not pure. Some separations and purification of TPA would be needed before repolymerization to make PET. Methods for TPA purification include hydrogenation, use of activated carbon, recrystallization, and filtration.<sup>26,27</sup>

Table S2† provides data for other byproducts. Phthalic acid, another TPA isomer, was also present, but in quantities too small to draw meaningful conclusions regarding variations in yield with reaction conditions. Bis(2-hydroxyethyl)terephthalate (BHET) is an additional byproduct.  $Y_{\text{BHET}}$  was <10% at  $m_{\text{PET}}/m_{\text{w}} = 1/10$  and showed no clear correlation with reaction conditions.

### Severity index

Fast hydrolysis is a non-isothermal process wherein the reactor temperature continually increases with time. This complicates the ability to compare results from runs at different set point temperatures. Therefore, we used an empirical severity index (SI),<sup>28</sup> which combines the effects of temperature and time to facilitate the comparison of results over the wide range of reaction conditions used in these experiments. The severity index can be calculated as<sup>9</sup>

$$\text{SI} = \int_0^t A e^{-\frac{E_a}{R} \left( \frac{1}{T+273} - \frac{1}{T_{\text{ref}}+273} \right)} dt. \quad (2)$$

where  $E_a$  is the activation energy of  $1.19 \times 10^5 \text{ J mol}^{-1}$  and  $A$  is the pre-exponential factor of  $1.66 \times 10^8 \text{ min}^{-1}$ . These values were obtained by modelling PET depolymerization into TPA *via* neutral hydrolysis as a pseudo first-order reaction.<sup>12</sup>  $R$  is the gas constant in  $\text{J (mol}^{-1} \text{ K}^{-1})$ ,  $T$  the temperature within the reactor in °C (from eqn (1)),  $T_{\text{ref}}$  is an arbitrary reference temperature (427 °C), and  $t$  is batch holding time in minutes.

Fig. 4a illustrates TPA yield variations as a function of SI.  $Y_{\text{TPA}}$  is essentially zero up to  $\log(\text{SI}) = 9$ . The yield then increases with increasing severities to reach over 90% at  $\log(\text{SI}) \approx 18$ . At more severe conditions,  $Y_{\text{TPA}}$  decreases due to increased side reactions. Fig. 4a shows that TPA yields published previously for isothermal hydrolysis of PET in neutral water<sup>12</sup> follow the same trends with SI as the present TPA yields from fast hydrolysis.

Fig. 4b shows the yields of undissolved solids as a function of  $\log(\text{SI})$ . For  $\log(\text{SI}) < 12$ ,  $Y_{\text{US}}$  was above 80% (low PET conversion). The yield decreases sharply around  $\log(\text{SI}) = 14$  to 16, and it ultimately levels out at 0–20% for  $\log(\text{SI}) > 16$ . The decline corresponds to the depolymerization of the solids. As was the case for TPA yields, the yields of undissolved solids from both fast and isothermal hydrolysis of PET follow the same correlation with  $\log(\text{SI})$ .

### Effect of $m_{\text{PET}}/m_{\text{w}}$ on fast hydrolysis of PET

For fast hydrolysis to be conducted at scale, reducing the amount of water (relative to PET) would reduce material and energy inputs and environmental impacts (*e.g.*, less wastewater). While the experiments discussed above were performed at a fixed PET/water mass ratio of 1/10, this section describes results from experiments with PET/water mass ratios of 1/8, 1/6, 1/4, and 1/2.

Fig. 5a shows there is no statistically significant difference in TPA yield whether using a PET/water mass ratio ( $m_{\text{PET}}/m_{\text{w}}$ ) of 1/10 or 1/8 across different set point temperatures and reaction times. However, Fig. 5b shows that further decreasing the amount of water decreases  $Y_{\text{TPA}}$  for hydrolysis at a  $T_{\text{SP}} = 510$  °C. Notably, this decrease is not related to insufficient water being present for hydrolysis. For PET hydrolysis, stoichiometry would require  $m_{\text{PET}}/m_{\text{w}} = 1/0.166$ , and even at  $m_{\text{PET}}/m_{\text{w}} = 1/2$ , water is present in large stoichiometric excess.





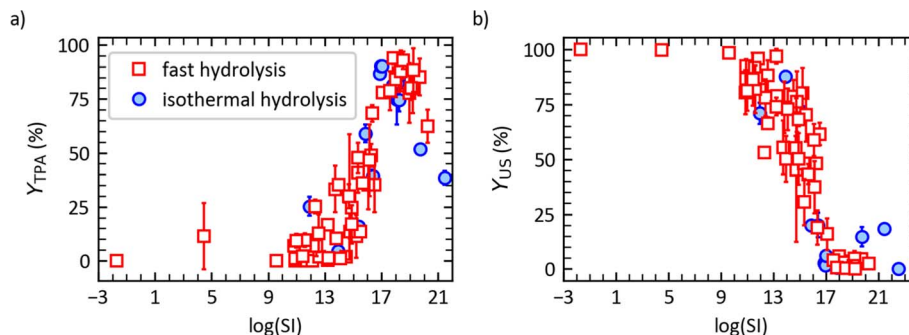


Fig. 4 (a) TPA molar yield and (b) yield of undissolved solids vs. log of the severity index for isothermal hydrolysis<sup>12</sup> ("filled circles") and fast hydrolysis ("empty squares") of post-consumer PET.

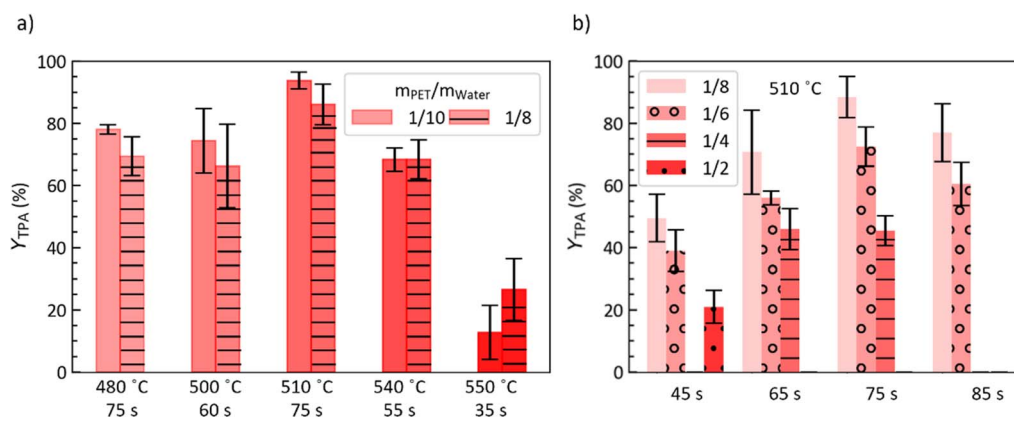


Fig. 5 TPA yield from fast hydrolysis of post-consumer PET for (a)  $m_{\text{PET}}/m_w = 1/10$  and  $1/8$  at  $T_{\text{SP}}$  and  $t$  indicated, and (b) different  $m_{\text{PET}}/m_w$  and  $t$  at  $T_{\text{SP}} = 510$  °C.

Table S2† shows that unlike  $Y_{\text{TPA}}$ ,  $Y_{\text{isophthalic acid i-TPA}}$ ,  $Y_{\text{US}}$ , and  $Y_{\text{BHET}}$  show little systematic variation with  $m_{\text{PET}}/m_w$  for a given batch holding time. This suggests the possibility of additional reactions occurring at higher PET/water mass ratios, potentially involving product interconversion and/or an equilibrium between byproducts and intermediates. Such reactions would include interconversion from TPA to isophthalic acid and TPA decarboxylation to benzoic acid.<sup>25</sup> Overall, these results suggest a practical upper bound of  $m_{\text{PET}}/m_w$  of  $1/8$  for PET fast hydrolysis.

We performed additional analysis to determine which variable predominantly affects the  $Y_{\text{TPA}}$ : the severity index, reflecting the temperature–time interplay, or the PET/water mass ratio. Insights derived from a machine learning random forest regression analysis reveal that the normalized logarithm of severity index exerts a significantly greater influence on the TPA yield than does the normalized PET/water mass ratio. The respective contributions are 94.7% and 5.3%. Fig. S1† shows the model did a good job fitting the experimental data, and Fig. S2† shows the predictive capability.

### Environmental energy impact

The environmental energy impact metric,  $\xi$ , quantifies the environmental impact of the process.<sup>19,29</sup>

$$\xi = \frac{0.1m_w \int_0^t T(t)dt}{Y_{\text{TPA}}/100 \times m_{\text{TPA}}} \quad (3)$$

Temperature ( $T$ ) is in units of celsius (°C) and time ( $t$ ) is in minutes. Lower  $\xi$  values indicate a more environmentally friendly process (higher TPA yields, lower temperature, shorter time, and less waste). This metric has found wide use in prior studies as it incorporates energy requirements, waste generation, and product yield.<sup>19,30</sup> In accordance with established practice, 10% of the reaction medium ( $m_w$ ) was assumed to be replenished due to losses.

Table S2† details  $\xi$  for each batch holding time,  $T_{\text{SP}}$ , and  $m_{\text{PET}}/m_w$  ratio. The lowest  $\xi$  values at  $m_{\text{PET}}/m_w = 1/10$  appeared for hydrolysis for 75 seconds at  $T_{\text{SP}} = 510$  °C ( $\xi = 455 \pm 31$  °C min) and 540 °C ( $\xi = 481 \pm 43$  °C min). Fig. S6† shows that within the vicinity of these conditions, there is a larger area where the  $\xi$  values remain relatively low and are less than 1000. However, the environmental energy impact equation shows that, all other conditions being equal,  $\xi$  can be reduced by using less water. Fig. 6 shows  $\xi$  is statistically indistinguishable for the experiments at  $T_{\text{SP}} = 510$  °C with  $m_{\text{PET}}/m_w = 1/10$ ,  $1/8$ , and  $1/6$  at 75 seconds.

The environmental energy impact metrics for fast hydrolysis are lower than those for isothermal hydrolysis, whether in



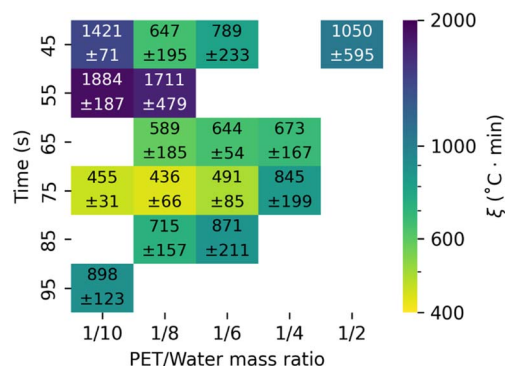


Fig. 6 Effect of batch holding time and  $m_{\text{PET}}/m_{\text{w}}$  on the environmental energy impact metric from fast hydrolysis of post-consumer PET at  $T_{\text{SP}} = 510$  °C.

neutral water or acid/base-catalyzed. Isothermal processes typically have  $\xi$  around  $10^4$  °C min.<sup>9,19</sup> The previous lowest  $\xi$  ( $235 \pm 44$  °C min) was achieved from fast hydrolysis at a set point temperature of 500 °C with a reaction time of 1 minute and a PET/water mass ratio of 1/3.<sup>12</sup> This value was based on three independent experimental runs. As we did additional replicates at this condition, however, the average  $\xi$  increased to  $630 \pm 420$  °C min, with the uncertainty largely due to variability in TPA yield ( $51 \pm 17\%$ , Table S2†). Thus, the combinations of  $T_{\text{SP}} = 510$  °C, 75 seconds, and a  $m_{\text{PET}}/m_{\text{w}} = 1/10$ , 1/8, or 1/6 give the lowest  $\xi$  among the conditions tested to date for fast hydrolysis of post-consumer PET.

### Remarks on fast hydrolysis at scale

Given the promise that fast hydrolysis shows for effective depolymerization of PET (e.g., potential technical feasibility), it is important to consider economic feasibility, especially relative to conventional isothermal hydrolysis. Fast hydrolysis experiments that led to high yields of TPA involved the reaction temperature reaching  $T = 443$  °C (510 °C set point after 75 seconds, Table S2†) just before the reaction was quenched. In contrast, isothermal hydrolysis can be conducted effectively at a lower temperature (e.g., 310 °C). We assume, for simplicity, that the reactor energy demand is the heat duty required to increase the enthalpy of the water in the reactor, as it has the greatest thermal mass in the system. The heat duty, assuming no heat loss for the reacting system, would be  $2.8 \text{ MJ kg}^{-1}$  (enthalpies obtained from the steam tables)<sup>31</sup> for fast hydrolysis. This is approximately twice the heat duty needed for isothermal hydrolysis ( $1.3 \text{ MJ kg}^{-1}$ ). This greater amount of energy would need to be supplied over a shorter time scale, meaning the power requirement for fast hydrolysis would be greater than that for isothermal hydrolysis. Of course, a mitigating factor with the energy demand is heat integration, which would be part of any well-designed chemical process. That is, at steady state, the actual energy input would be less than  $\Delta H$  for the reactor, as thermal energy would be recovered from the hot effluent stream and used to heat the feed stream. A second mitigating factor is the short residence time. Reaction severity is determined by the combination of the temperature profile and

the reaction time. The severity index for fast hydrolysis at a set point temperature of 510 °C for 75 s ( $\log(\text{SI}) = 17.8$ ) is less than that for isothermal hydrolysis at 310 °C for 30 min ( $\log(\text{SI}) = 18.2$ ). Thus, when assessing the severity of reaction conditions, one needs to consider the reaction time and not reaction temperature alone. Further, a fast hydrolysis reactor would only require about 5% of the volume of an isothermal hydrolysis reactor because fast hydrolysis required 75 seconds of residence time, while isothermal hydrolysis requires 1800 seconds. For a given flow rate, reactor volume is proportional to the residence time required. Thus, one expects the capital cost for the reactor section in a fast hydrolysis process to be lower than that for isothermal hydrolysis.<sup>32</sup>

If fast hydrolysis were to be done at scale, continuous tubular reactors would likely be used. These offer superior heat and mass transfer rates, and they have been used in pilot plants for continuous hydrothermal liquefaction processes.<sup>33</sup> Naphtha cracking is a current industrial process that requires high temperatures and short reactor residence times. In a typical cracker, the reactor coils are at 750–900 °C, the hydrocarbon feedstock flows through the reactor in 0.1–0.6 seconds, and the flowing fluid absorbs heat at a rate of 50–80 kW m<sup>-2</sup>.<sup>34</sup> These conditions are even more demanding than those anticipated for fast hydrolysis at scale.

Adaptability is another issue that is important when considering scalability. That is, how effective is the fast hydrolysis approach in processing other condensation polymers and mixed plastics, beyond just post-consumer PET, which was the focus of this study? Previous work<sup>35</sup> demonstrated that TPA can be produced in high yields from PET *via* isothermal hydrolysis of a binary mixture of plastics. In fact, the second plastic enhanced the yield of TPA and reduced the yields of byproducts. Experiments are underway in our lab to address these important questions regarding fast hydrolysis.

An issue related to feedstock flexibility is separation of TPA from other hydrolysis products and production of pure TPA at scale. As noted earlier in this article, methods have been developed for TPA purification.<sup>26,27</sup> These would need to be assessed and perhaps adjusted as part of the scale-up research and development portfolio.

This brief discussion highlights just a few key process differences for fast and isothermal hydrolysis. Additional differences could be expected as these processes receive additional scrutiny and further details emerge (e.g., design of separation systems). Of course, complete technoeconomic assessments beyond these considerations would be needed to holistically identify all relative advantages and disadvantages of a fast hydrolysis process at scale.

## Conclusions

We investigated the potential of fast hydrolysis for chemical recycling of post-consumer polyethylene terephthalate (PET). Herein we show the lowest environmental energy impact metrics to date for neutral hydrolysis of PET and even for acid/base-catalyzed hydrolysis. TPA yield can be correlated and predicted by employing an empirical severity index that accounts



for temperature and time. Higher severities led to TPA decomposition, while lower severities result in insufficient TPA production. TPA yields from isothermal neutral hydrolysis of PET followed the same correlation with SI as did yields from fast hydrolysis. Given this general ability of SI to correlate TPA yields, temperature, time, and heating rate can be manipulated to optimize TPA production from PET hydrolysis. Severity index was shown to have a higher impact on the TPA yield than did the PET/water mass ratio. We observed a practical upper bound of  $m_{\text{PET}}/m_{\text{w}}$  of 1/8 for PET fast hydrolysis, above which TPA yield decreased. As is often the case in engineered systems, there are tradeoffs when comparing fast hydrolysis and isothermal hydrolysis. A fast hydrolysis reactor would be smaller than an isothermal reactor, due to the holding time being about one minute instead of tens of minutes. This feature means reduced capital costs. Fast hydrolysis involves rapid heating to temperatures higher than those used in isothermal hydrolysis, so greater thermal energy would be required. Of course, heat recovery would be part of any well-engineered process, and this could offset some of the higher energy demand.

## Conflicts of interest

There are no conflicts to declare.

## Acknowledgements

This material is based upon work supported by the National Science Foundation under Grant EFRI E3P number 2029397.

## References

- 1 K. Ghosal and C. Nayak, *Mater. Adv.*, 2022, **3**, 1974–1992.
- 2 NAPCOR, *PET Recycling Reports*, <https://napcor.com/reports-resources/>, accessed 7 April 2023.
- 3 C. N. Onwucha, C. O. Ehi-Eromosele, S. O. Ajayi, M. Schaefer, S. Indris and H. Ehrenberg, *Ind. Eng. Chem. Res.*, 2023, **62**, 6378–6385.
- 4 W. Yang, J. Wang, L. Jiao, Y. Song, C. Li and C. Hu, *Green Chem.*, 2022, **24**, 1362–1372.
- 5 G. P. Karayannidis, A. P. Chatziavgoustis and D. S. Achilias, *Adv. Polym. Technol.*, 2002, **21**, 250–259.
- 6 S. Ügdüler, K. M. Van Geem, R. Denolf, M. Roosen, N. Mys, K. Ragaert and S. De Meester, *Green Chem.*, 2020, **22**, 5376–5394.
- 7 J. V. Valh, B. Vončina, A. Lobnik, L. F. Zemljič, L. Škodič and S. Vajnhandl, *Text. Res. J.*, 2020, **90**, 1446–1461.
- 8 W. Yang, R. Liu, C. Li, Y. Song and C. Hu, *Waste Manage.*, 2021, **135**, 267–274.
- 9 P. Pereira, P. E. Savage and C. W. Pester, *Green Chem.*, 2024, **26**, 1964–1974.
- 10 S. D. Mancini and M. Zanin, *Prog. Rubber, Plast. Recycl. Technol.*, 2004, **20**, 117–132.
- 11 D. Stanica-Ezeanu and D. Matei, *Sci. Rep.*, 2021, **11**, 4431.
- 12 P. Pereira, C. W. Pester and P. E. Savage, *ACS Sustain. Chem. Eng.*, 2023, **11**, 7203–7209.
- 13 B. Motavaf and P. E. Savage, *ACS ES&T Eng.*, 2021, **1**, 363–374.
- 14 D. S. Scott, S. R. Czernik, J. Piskorz and D. A. St G Radlein, *Energy Fuels*, 1990, **4**, 407–411.
- 15 J. Ni, L. Qian, Y. Wang, B. Zhang, H. Gu, Y. Hu and Q. Wang, *Fuel*, 2022, **327**, 125135.
- 16 Q.-V. Bach, M. V. Sillero, K.-Q. Tran and J. Skjermo, *Algal Res.*, 2014, **6**, 271–276.
- 17 B. Hu, Z. Zhang, W. Xie, J. Liu, Y. Li, W. Zhang, H. Fu and Q. Lu, *Fuel Process. Technol.*, 2022, **237**, 107465.
- 18 L. Zhu, M. Huang, Z. Ma, B. Cai, W. Zhang, H. Peng and D. Chen, *Fuel*, 2023, **349**, 128734.
- 19 E. Barnard, J. J. R. Arias and W. Thielemans, *Green Chem.*, 2021, **23**, 3765–3789.
- 20 A. Căta, M. N. Ștefănuț, I. M. C. Ienașcu, C. Tănăsie and M. Miclău, *Rev. Roum. Chim.*, 2017, **62**, 531–538.
- 21 J. J. R. Arias and W. Thielemans, *Green Chem.*, 2021, **23**, 9945–9956.
- 22 M. Azeem, M. B. Fournet and O. A. Attallah, *Arabian J. Chem.*, 2022, **15**, 103903.
- 23 M. Atkins, N. Curry and S. Evans, *Polymer Recycling*, WO2022171874A1, 2021, <https://patents.google.com/patent/US5869543A/en>.
- 24 M. Čolnik, Ž. Knez and M. Škerget, *Chem. Eng. Sci.*, 2020, **223**, 116389.
- 25 J. B. Dunn, M. L. Burns, S. E. Hunter and P. E. Savage, *J. Supercrit. Fluids*, 2003, **27**, 263–274.
- 26 H. L. Lee, C. W. Chiu and T. Lee, *Chem. Eng. J. Adv.*, 2021, **5**, 100079.
- 27 Y. Peng, J. Yang, C. Deng, Y. Fu and J. Deng, *Nat. Commun.*, 2023, **14**, 3249.
- 28 L. Qian, S. Wang and P. E. Savage, *Appl. Energy*, 2020, **260**, 114312.
- 29 A. M. Al-Sabagh, F. Z. Yehia, G. Eshaq, A. M. Rabie and A. E. ElMetwally, *Egypt. J. Pet.*, 2016, **25**, 53–64.
- 30 J. J. R. Arias and W. Thielemans, *Green Chem.*, 2021, **23**, 9945–9956.
- 31 A. H. Harvey, *Thermodynamic Properties of Water: Tabulation from the IAPWS Formulation 1995 for the Thermodynamic Properties of Ordinary Water Substance for General and Scientific Use*, <https://www.nist.gov/srd/nistir5078>, accessed 12 September 2021.
- 32 J. R. Couper, D. W. Hertz and F. L. Smith, in *Perry's Chemical Engineers' Handbook*, ed. D. W. Green and R. H. Perry, The McGraw-Hill Companies, Inc., New York, 8th edn, 2008, pp. 9.
- 33 I. Johannsen, B. Kilsgaard, V. Milkevych and D. Moore, *Processes*, 2021, **9**, 1–18.
- 34 *Kirk-Othmer Concise Encyclopedia of Chemical Technology*, ed. Kirk-Othmer, Wiley-Interscience, New York, 4th edn, 1985.
- 35 S. M. Subramanya, Y. Mu and P. E. Savage, *ACS Eng. Au*, 2022, **2**, 507–514.

