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Elimination of transforming activity and gene degradation during UV and UV/H₂O₂ treatment of plasmid-encoded antibiotic resistance genes

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Water Impact Statement

The efficiency and mode of actions for deactivating and degrading antibiotic resistance genes (ARGs) during water treatment with UV (254 nm) and UV/H₂O₂ have been poorly understood. Here, we show that efficiency of elimination of the transforming activity for a plasmid-encoded ARG during the UV-based treatments depends on the rate of formation of cyclobutane-pyrimidine dimers (CPDs) in the plasmid and the repair of such DNA damage during the transformation process in host cells. This work has important contributions to optimizing the monitoring and operation of UV-based water disinfection and oxidation processes for removing ARGs.

1	Elimination of transforming activity and gene degradation during
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20 Abstract

21 To better understand the elimination of transforming activity of antibiotic resistance genes 22 (ARGs), this study determined deactivation of transforming activity of an ARG (in *Escherichia* 23 *coli* as a host) and the ARG degradation (according to quantitative PCR [qPCR] with different 24 amplicon sizes) during UV (254 nm) and UV/ H_2O_2 treatment of plasmid pUC19 containing an ampicillin resistance gene (amp^R) . The required UV fluence for each \log_{10} reduction of the 25 transforming activity during UV treatment was ~37 mJ/cm² for both extra- and intra-cellular 26 pUC19 (the latter within *E. coli*). The resulting fluence-based rate constant (k) of ~ 6.2×10^{-2} 27 28 cm^2/mJ was comparable to the k determined previously for transforming activity loss of 29 plasmids using host cells capable of DNA repair, but much lower (~10-fold) than that for DNA 30 repair-deficient cells. The k value for pUC19 transforming activity loss was similarly much 31 lower than the k calculated for cyclobutane-pyrimidine dimer (CPD) formation in the entire 32 plasmid. These results indicate a significant role of CPD repair in the host cells. The degradation rate constants (k) of amp^{R} measured by qPCR increased with the increasing target 33 34 amplicon size (192–851 bp) and were close to the k calculated for the CPD formation in the 35 given amplicons. Further analysis of the degradation kinetics of plasmid-encoded genes from 36 this study and from the literature revealed that qPCR detected most UV-induced DNA damage. 37 In the extracellular plasmid, DNA damage mechanisms other than CPD formation (e.g., base 38 oxidation) were detectable by qPCR and gel electrophoresis, especially during UV/H_2O_2 39 treatment. Nevertheless, the enhanced DNA damage for the extracellular plasmids did not result 40 in faster elimination of the transforming activity. Our results indicate that calculated CPD 41 formation rates and qPCR analyses are useful for predicting and/or measuring the rate of DNA 42 damage and predicting the efficiency of transforming activity elimination for plasmid-encoded 43 ARGs during UV-based water disinfection and oxidation processes.

44

45 **Keywords:** antibiotic resistance, ampicillin resistance gene, qPCR, transforming activity, UV,

- 46 cyclobutane-pyrimidine dimer
- 47

48 Introduction

49 Increasing antibiotic resistance is a major threat to human and animal health, as it can lower 50 the therapeutic potential of antibiotics against bacterial infections (WHO, 2014). Although antibiotic resistance can occur naturally, overuse or misuse of antibiotics in modern society is 51 52 associated with increased antibiotic resistance (Allen et al., 2010). Antibiotics can select 53 antibiotic-resistant bacteria (ARB) that carry genes (ARGs) responsible for antibiotic resistance 54 mechanisms. The presence of ARB and ARGs in aquatic environments is a concern because it 55 can promote the spread of antibiotic resistance through natural and anthropogenic water cycles 56 (Berendonk et al., 2015; Pruden, 2014). In addition, antibiotic resistance can be disseminated 57 among bacterial populations by sharing (mobile) ARGs through horizontal gene transfer (HGT) 58 processes (Thomas and Nielsen, 2005). To minimize dissemination of environmental sources of 59 antibiotic resistance, the necessity of coordinated national and international strategies has been 60 advised for monitoring, risk assessment, and mitigation of antibiotic resistance (Vikesland et al., 61 2017).

62 Municipal wastewater has been identified as one of the hotspots that release ARB and ARGs 63 into aquatic environments (Rizzo et al., 2013; Tan and Shuai, 2015). Conventional wastewater 64 treatment does not fully eliminate ARB and ARGs (Chen and Zhang, 2013; Rizzo et al., 2013). 65 Biological treatment processes (e.g., activation of sludge) can significantly reduce the load of 66 ARB but may select highly (or multi-) resistant bacterial species (Czekalski et al., 2012). 67 Disinfection of wastewater effluents with chlorine or ultraviolet irradiation (UV) has been 68 widely practiced for water resource protection (Jacangelo and Trussell, 2002). Ozonation has 69 recently received renewed attention as an option for treating municipal wastewater effluents to

eliminate organic micropollutants (Lee and von Gunten, 2016). Advanced oxidation processes
such as UV/H₂O₂ treatment have also been tested to achieve the same goal (Gerrity et al., 2016;
Miklos et al., 2018). There has been growing interest in the efficiency of wastewater
disinfection and oxidation processes to lower the levels of ARB and ARGs, in addition to
micropollutant elimination (Alexander et al., 2016; Czekalski et al., 2016; Dodd, 2012; Ferro et
al., 2017; Lüddeke et al., 2015; McKinney and Pruden, 2012; Pak et al., 2016; Sousaa et al.,
2017; Yoon et al., 2017).

77 ARGs in wastewaters exist in different forms such as intracellular (within bacteria) and 78 extracellular, as free DNA and viruses (Colomer-Lluch et al., 2011; Zhang et al., 2018). ARGs 79 can transfer resistance by HGT mechanisms such as conjugation, transduction, and 80 transformation. Among these HGT mechanisms, transformation requires only intact ARGs for 81 the resistance transfer, as extracellular ARGs can be taken up and incorporated into the genomes 82 of competent bacteria even in the absence of the original donor ARB cell (Thomas and Nielsen, 83 2005). This mechanism is therefore different from conjugation or transduction in which viable 84 donor cells or infective viruses containing ARGs are needed. Considering the potential for ARG 85 transfer via transformation, it is necessary to assess the efficiency of disinfectants at destroying 86 ARGs and eliminating their associated transforming activities (Dodd, 2012).

87 Molecular mechanisms of DNA damage induced by UV or by hydroxyl radicals (OH) are well established. UV (particularly UVC) mainly generates DNA base lesions such as 88 89 cyclobutane-pyrimidine dimers (CPDs) and pyrimidine (6-4) pyrimidone adducts [(6-4) 90 photoproducts] (Görner, 1994; Sinha and Häder, 2002). Good correlations have been found 91 between the UV-induced degradation rate of ARGs and the number of adjacent pyrimidine 92 dimer sites (Chang et al., 2017; Destiani et al., 2017; McKinney and Pruden, 2012). Genomic modeling, an approach to predict the UV sensitivity of microorganisms based on their DNA 93 94 sequence characteristics, has been tested to predict the degradation efficiency of ARGs 95 (Destiani et al., 2017). For 'OH, the DNA damage can range from base oxidation to sugar 96 backbone breakages, where the latter can lead to single strand (ss) and double strand (ds) breaks 97 (von Sonntag, 2006). Despite this knowledge, it is unclear how different types of DNA damage 98 resulting from reactions with disinfectants (e.g., UV, 'OH) are related to the loss of ARG 99 transforming activity.

100 Quantitative polymerase chain reaction (qPCR) has been widely used to detect and quantify 101 ARGs present in aquatic environments (Luby et al., 2016). The qPCR method has also been 102 employed to assess the efficacy of disinfection processes for ARG elimination by quantifying 103 target qPCR amplicons (Alexander et al., 2016; Czekalski et al., 2016; Destiani et al., 2017; 104 Ferro et al., 2017; Lüddeke et al., 2015; McKinney and Pruden, 2012; Pak et al., 2016; Sousaa 105 et al., 2017; Yoon et al., 2017). Most qPCR methods involve amplicons covering only portions 106 of ARGs (e.g., 100–200 bp) and rarely cover the entire genes that are necessary for gene 107 transfer pathways (Chang et al., 2017; McKinney and Pruden, 2012; Yoon et al., 2017). In 108 addition, the sensitivity of qPCR and bacterial gene transformation to DNA damage can differ 109 due to the different degree of DNA repair or DNA polymerase fidelity rates of qPCR vs. the 110 bacterial cell system (Chang et al., 2017). As an alternative approach, the transforming activity 111 of ARGs can be directly measured by transformation assays in which the target ARG-containing 112 DNA (e.g., plasmid) is taken up by and incorporated into the genomes of nonresistant, 113 competent bacterial cells (Dodd, 2012; Luby et al., 2016). Nevertheless, few studies have 114 applied such an ARG transformation assay to assess the efficacy of disinfection processes for 115 the elimination of antibiotic resistance (Chang et al., 2017). Furthermore, the relationship 116 between the qPCR method and the transformation assay for determining a biologically active 117 ARG concentration is still poorly understood.

118 To elucidate the efficiency of deactivation and degradation of ARGs during water 119 disinfection and oxidation processes, in this study, we determined and compared the changes in

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transforming activity and ARG concentrations during UV_{254nm} (hereafter UV) or UV/H₂O₂ 120 121 treatment of plasmid-encoded ARGs in bench scale disinfection experiments. A quantitative 122 transformation assay employing *Escherichia coli* as the recipient was conducted to determine 123 the transforming activity of the target plasmid. ARG concentrations were determined by qPCR 124 with different amplicon sizes to determine the DNA damage in different parts of the target 125 plasmid. In addition, agarose gel electrophoresis was carried out to determine structural changes 126 of the plasmid. Both extracellular and intracellular forms of plasmids were treated to test the 127 effects of cellular components on the efficiency of ARG elimination. The results were evaluated with respect to factors affecting the efficiency of elimination of ARGs' transforming activities; 128 129 for example, DNA repair, plasmid characteristics, the type of DNA damage, and sensitivity of 130 DNA polymerase.

131

132 Materials and Methods

133 Standards and reagents

All chemicals and solvents (mostly of $\geq 95\%$ purity) were purchased from various commercial suppliers and used as received (SI-Text-1). Chemical solutions were prepared with ultrapure water ($\geq 18.2 \text{ M}\Omega \cdot \text{cm}$) that was obtained by means of a Barnstead purification system (Thermo Fisher Scientific, USA). Glassware was washed with ultrapure water and autoclaved at 121°C for 15 min prior to use.

139 Model bacterial strains and ARG-containing plasmid

140 *E. coli* DH5α hosting plasmid pUC19 served as the model ARB in this study (Casali and 141 Preston, 2003). Nonresistant *E. coli* DH5α was used as the recipient strain for the ARG 142 transformation assay. The *E. coli* stocks were prepared at concentrations of $\sim 10^9$ colony-143 forming units (CFU)/ml according to the method described elsewhere (Yoon et al., 2017). Cells from the midexponential growth phase were used. Plasmid pUC19 (2686 bp) is an *E. coli* vector that carries an ampicillin resistance gene (amp^R ; see Figure S1 and Table S1 for gene information). Plasmids were extracted from *E. coli* stocks with the AccuPrep Nano-Plus Plasmid Extraction Kit (Bioneer, 2016). Extracted plasmids were analyzed and quantified on a NanoDrop ND-2000 spectrophotometer (NanoDrop Products, Wilmington, USA). Plasmid concentrations in the extracted stock solutions (50 µL) were ~10¹¹ copies/µL (or 0.5–1 µg/µL). See SI-Text-2 for further details.

151 Determination of transforming activity of plasmid-encoded ARG

152 To quantify the ability of the pUC19 plasmid to transfer its antibiotic resistance, a 153 transformation assay was conducted with nonresistant E. coli DH5 α as the recipient strain 154 (Hanahan, 1983). Competent cells were prepared by treating E. coli DH5 α with calcium 155 chloride and glycerol as the chemical treatment method (Shanehbandi et al., 2013) and stored at -80°C until use (see SI-Text-3 for details). One-hundred μ L of the competent cells (~7 × 10⁸ 156 157 CFU/ml) was thawed and prepared in 1.5 mL tubes, and mixed with 5-10 µL of the plasmid 158 samples. The resulting mixtures were placed on ice for 30 min incubation, then quickly 159 transferred into a water bath at 42°C for 45 s and placed back on ice for 2 min. After heat shock, 160 the samples were mixed with 900 µL of Luria-Bertani (LB) broth and cultured in a shaking incubator (200 rpm) at 37°C for 45 min. The incubated samples were serially diluted with LB 161 broth and plated onto LB agar plates containing 50 mg/L ampicillin. The concentration of the 162 163 transformants was determined by enumerating the ARB colonies on the plates after 24 hours of 164 incubation in the dark at 37°C. Finally, the transforming activity of the samples was calculated 165 as the concentration of transformants from colonies in selective plates (with ampicillin) 166 normalized to the concentration of E. coli cells from colonies in nonselective plates (without

- 167 ampicillin) as presented in Eq. 1. From the nonselective plates, typical concentrations of $\sim 7 \times$
- 168 10^8 CFU/mL *E. coli* cells were determined under the tested conditions.
- 169 Transforming activity = $[Transformants]_{selective plate} \div [E. coli cells]_{nonselective plate}$ [1]
- 170 **qPCR**

Amplicons spanning variable-length segments of the amp^{R} gene (192, 400, 603, and 851 bp 171 172 of the overall 861 bp length of amp^{R}) and *ori* region (190, 390, and 530 bp of the overall 589 bp length of ori) in the pUC19 plasmid (2686 bp) were quantified by means of qPCR (Figure S1, 173 174 Table S1). Amplicon and primer sequences were determined from the pUC19 sequence 175 retrieved from the NCBI GenBank database. Primers were designed using the NCBI Primer-176 BLAST tool (Table S2). qPCR measurements were performed on a CFX96 Real-Time PCR detection system (Bio-Rad, Hercules, CA, USA) using SsoFastTM EvaGreen[®] Supermixes (Bio-177 Rad). Standard curves were generated from 10-fold serial dilutions covering 5 orders of 178 179 magnitude. Each 20 μ L qPCR reaction consisted of 1 μ L of 0.5 pmol forward and reverse primers (0.5 pmol/µl), 1 µL of a DNA sample, 10 µL of an EvaGreen[®] Supermix, and 8 µL of 180 181 autoclaved DNase-free water. The temperature profile of the PCR protocol included one cycle 182 at 95°C for 2 min; 30 cycles at 95°C for 5 s, an annealing step at 55°C for 60 s, and an extension 183 step at 72°C for 20 s; followed by a melt curve analysis from 65°C to 95°C. The same PCR 184 protocol was used for all qPCR assays with different amplicons. Calibration curves for the target amplicons exhibited r^2 values of ≥ 0.98 for all cases (Figure S2 shows a representative 185 example). The average amplification efficiencies for the amp^{R} amplicons were 0.90(±0.04) for 186 187 192 bp, $0.89(\pm 0.06)$ for 400 bp, $0.84(\pm 0.09)$ for 603 bp, and $0.88(\pm 0.08)$ for the 851 bp 188 amplicon. The average amplification efficiencies for the *ori* amplicons were $0.84(\pm 0.05)$ for 190 189 bp, $0.95(\pm 0.03)$ for 390 bp, and $0.93(\pm 0.07)$ for the 530 bp amplicon. The limit of detection 190 (LOD) and quantification (LOQ) were determined as 15 copies and 40 copies per reaction for

191 most qPCR runs. The end products of qPCR analyses were also analyzed by agarose gel 192 electrophoresis (Figure S3) to confirm successful amplification of the target genes. The samples 193 from UV treatment were directly analyzed by the qPCR protocol described above for 194 extracellular plasmids. For intracellular plasmids, 10 mL of a given sample containing 195 disinfectant-treated or untreated E. coli DH5 α cells was centrifuged, and the pellet was 196 resuspended in 100 μ L of Tris-HCl buffer (10 mM, pH 8.5). The resulting concentrated samples 197 were processed with the AccuPrep Nano-Plus Plasmid Extraction Kit (Bioneer, 2016) and 198 subsequently analyzed by the above qPCR protocol.

199 Gel electrophoresis

200 Samples were prepared by treating the extracted pUC19 ($\sim 5 \mu g/mL$) at pH 7 (2 mM phosphate buffered solutions) with UV and UV/H₂O₂ at different UV fluence levels (0-312 201 mJ/cm²). Linearized pUC19 plasmids were prepared by incubating the extracted pUC19 with 202 203 type II restriction enzyme *EcoRI* (NEB, USA) at 37°C for 1 h, followed by enzyme inactivation at 65°C for 20 min. Standards of the amp^{R} amplicons with different sizes (192, 400, 603, and 204 205 851 bp) were prepared by qPCR reactions above. These prepared plasmid samples and a 1 kb 206 DNA ladder (Enzynomics, KOREA) were loaded on 0.8% agarose gels containing 0.5 µg/ml 207 ethidium bromide in 1× TAE (Tris-Acetate-EDTA) buffer and were separated at 4 V/cm for 35 208 min. Gel images were captured on a bench-top UV Transilluminator (Universal mutation 209 detection system, UVP, LLC, USA). The density of each band on the gels was calculated by quantitative band analysis in the ImageJ software (Schneider et al., 2012). Isolated amp^{R} 210 211 amplicons for the transforming activity determination were prepared by cutting the gels loaded with the amp^{R} qPCR amplicon mixtures and by subsequent purification with the AccuPrep[®] Gel 212 213 Purification Kit (Bioneer).

214 UV and UV/H₂O₂ treatments

9

215 Bench scale UV irradiation experiments were conducted in a quasi-collimated beam system 216 (Bolton and Linden, 2003) equipped with a low-pressure mercury lamp emitting 254 nm light (Sankyo Denki Ltd., Tokyo, Japan). The applied photon fluence rate was ~0.3 mW/cm² as 217 218 determined by a UVX digital radiometer (Ultra-Violet Products Ltd., Upland, USA) or by atrazine chemical actinometry (Lee et al., 2016). Solutions of the pUC19 plasmid or E. coli 219 were separately prepared in 2 mM phosphate buffer (pH 7) by adding the corresponding stock 220 solutions at the concentration of ~10¹¹ copies/mL for the plasmid or ~5 \times 10⁶ CFU/mL for E. 221 *coli*, respectively. These samples (120 mL) were placed in a Petri dish reactor with a sample 222 223 depth of 2.5 cm and were irradiated with UV light under gentle stirring using a magnetic stir bar. 224 For UV/H₂O₂ experiments, 10 mg/L H₂O₂ was added to the samples before the UV irradiation. The absorbance of the samples was ≤ 0.02 cm⁻¹ at 254 nm and the light screening (attenuation) 225 226 coefficient in the Petri dish reactor was calculated to be ≥ 0.94 (Bolton and Linden, 2003). The 227 presence of 10 mg/L of H₂O₂ alone as the control experiment did not induce the gene degradation within 40 min of contact time (Yoon et al., 2017). The decrease of H_2O_2 by the UV 228 229 photolysis was less than 10% of its initial concentration within the applied UV fluence range (0 -300 mJ/cm^2). Thus, the degradation of pUC19 by OH radicals (formed from the UV) 230 231 photolysis of H_2O_2) was in a first-order condition throughout the UV irradiation. The reaction 232 solution was sampled (and in the latter case, also supplemented with bovine catalase (40 μ g/mL) 233 to quench residual H_2O_2) and was then stored at $-20^{\circ}C$ for 10 days. Triplicate experiments were conducted for each condition, and average sample concentrations (amplicon copies, 234 235 transforming activity, or viable cell counts) reported with one standard deviation.

236 Statistical analysis

Statistical analyses were conducted in GraphPad Prism 7 (<u>http://www.graphpad.com/</u>). UV
 fluence-based first-order rate constants determined from different sets of experiments were

compared by multiple linear regression analyses. The null hypothesis in these analyses was that the first-order rate constants were not significantly different, with p < 0.05 as the significance cutoff.

242

243 **Results**

244 Quantitative determination of the transforming activity of plasmid-encoded amp^{R}

245 Figure 1 shows that the concentration of transformants after the plasmid transformation increases in proportion to the pUC19 concentration, with a slope of 1.0 in a log-log scale plot. 246 The measured transforming activity ranged from 3.1×10^{-8} to 1.6×10^{-4} for the pUC19 247 concentrations ranging from 10^{-5} to 5.0×10^{-2} µg/mL. This result indicates good capacity of the 248 assay for quantitative determination of the transforming activity. A total of $\sim 2 \times 10^8$ 249 250 transformants were formed per µg of the plasmid. The observed transformation efficiency is comparable to that reported in the literature (i.e., 2×10^8 to 10^9 transformants/µg of plasmid 251 252 DNA; Hanahan et al., 1991).

The transforming activity of pUC19 ($2.5 \times 10^{-2} \,\mu\text{g/mL}$) after the digestion with the *EcoRI* 253 restriction enzyme was found to be $\sim 10^{-6}$, which was lower than that of intact pUC19 by a 254 255 factor of 80. *EcoRI* can linearize pUC19 by DNA cutting (Pingound and Jeltsch, 2001). Thus, dsDNA breaks created by *EcoRI*, even outside the amp^{R} gene (especially at restriction site 284, 256 see Figure S1), can significantly reduce the ampicillin resistance transforming activity of 257 pUC19. Our data are consistent with other studies showing that the transformation efficiency of 258 259 plasmids can decrease by two or three orders of magnitude after digestion of the plasmids with 260 various restriction enzymes (Chang et al., 2017; Palmen et al., 1993; Schulte-Frohlinde, 1987). The transforming activities of amp^{R} gene fragments in a range of amplicon sizes (192, 400, 603, 261 and 851 bp, see Figure S1) were also determined. These amp^{R} gene fragments (prepared at 10¹¹ 262

263 copies/mL) showed negligible transforming activity (lower by more than four orders of 264 magnitude) as compared to pUC19 at the same molar concentrations. This finding indicates that 265 not only amp^{R} but also the whole plasmid is required for the transformation of amp^{R} .

266 Elimination of transforming activity of plasmid-encoded amp^{R}

Figure 2 shows decreases in the *amp*^{*R*} transforming activity (stars) during UV or UV/H₂O₂ treatment of extracellular (i.e., extracted plasmid) or intracellular (i.e., plasmid within *E. coli*) pUC19 at pH 7. The initial transforming activities in these experiments were $\sim 10^{-4}$ and the transformation activity of $\sim 10^{-8}$ was the limit of quantification of the method; thus, a $\sim 4 \log$ reduction could be detected. Without UV and UV/H₂O₂ treatment, *E. coli* and pUC19 were stable for several hours in the phosphate buffer according to the control assays.

273 The elimination of transforming activity followed first-order kinetics with respect to UV fluence in all cases ($r^2 \ge 0.99$). The fluence-based first-order rate constants (k) for the 274 275 transforming activity loss could be determined from the slopes of the linear plots (i.e., k = 2.303) × slope). The k values for intracellular pUC19 (i-ARG) were $6.2(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV 276 treatment (Figure 2a) and $6.4(\pm0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV/H₂O₂ treatment (Figure 2b), and for 277 extracellular pUC19 (e-ARG) were $6.1(\pm 0.3) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and $7.3(\pm 0.4) \times 10^{-2} \text{ cm}^2/\text{mJ}$ for UV (Figure 2c) and 278 10^{-2} cm²/mJ for UV/H₂O₂ treatment (Figure 2d), respectively. The data showed that the rates of 279 elimination of the transforming activity were almost the same for all cases, with k of $\sim 6.2 \times 10^{-2}$ 280 cm^2/mJ (p < 0.001), except for the e-ARG treatment with UV/H₂O₂ yielding a slightly higher 281 (by a factor of 1.2) deactivation rate. At a UV fluence of 40 mJ/cm² (typical for water 282 283 disinfection), the achieved elimination of the transforming activity was a 1.0 log reduction (1.3 log for UV₂₅₄/H₂O₂ treatment of e-ARG). To achieve more extensive elimination of the 284 285 transforming activity such as a 4-log reduction, the required UV fluence was found to be 150 mJ/cm^2 (125 mJ/cm^2 for UV₂₅₄/H₂O₂ treatment of e-ARG). 286

287 Degradation of pUC19 fragments (amp^{R} and *ori* portions) determined by qPCR

288 Figure 2 also presents changes in the logarithmic relative concentrations of the suite of variable-length amp^{R} amplicons measured by qPCR (i.e., 192, 400, 603, and 851 bp) during UV 289 and UV/H₂O₂ treatment of *E. coli* (i-*amp*^{*R*}) and pUC19 plasmid (e-*amp*^{*R*}), respectively. In all 290 291 cases, the depletion of amp^{R} genes followed first-order kinetics with respect to the UV fluence $(r^2 \ge 0.99)$. The k values determined for the gene damage are summarized in Table 1. The 292 following points from these kinetic data are noteworthy. First, the gene damage rates of amp^{R} 293 294 increase with the increasing qPCR amplicon size. This result can be explained by the increasing 295 number of target sites for UV (e.g., adjacent pyrimidine dimer sites) with the increasing gene 296 size. Second, the gene damage was slower or faster than the transforming activity loss 297 depending on the qPCR amplicon size. For example, the 192 bp gene fragment underwent much 298 slower degradation (by a factor of 2.9-4.0 based on relative k) as compared to the loss of 299 transforming activity. The degradation rates of the 400 and 603 bp gene fragments were higher 300 and closer to the rates of ARG deactivation (i.e., transforming activity elimination). The 301 degradation rates of the 851 bp gene fragment were higher by a factor of 1.1-2.5 as compared to 302 those of the transforming activity loss. These results highlight the need to better understand the 303 relation between the gene damage determined by the qPCR method and the elimination of 304 transforming activity of plasmid-encoded ARGs (further details will be discussed later). Third, the gene damage rates of e-amp^R were higher than those of i-amp^R in both UV and UV/H₂O₂ 305 treatments (p < 0.05), and the difference was greater for the UV/H_2O_2 treatment. The gene 306 damage rates of i-amp^R were quite similar between UV and UV/H₂O₂ treatment (p = 0.47 for 307 192 bp, p = 0.13 for 400 bp, p = 0.50 for 604 bp, and p = 0.82 for 851 bp), whereas the gene 308 damage rates of e-*amp*^{*R*} were considerably higher in UV/H₂O₂ than in UV treatment (p < 0.05). 309 310 These results indicated that 'OH from UV photolysis of H₂O₂ can contribute to the degradation 311 of extracellular genes, but the effect of 'OH on intracellular genes is effectively negligible 312

compared to the UV-induced direct gene damage. A similar result was reported in our previous 313 work, and was explained by complete scavenging of 'OH by cell membrane or cytoplasmic 314 components before reaching i-ARG (Yoon et al., 2017). 315 In this study, the qPCR method was also applied to measure damage in the ori region (Figure S1), for assessing DNA reactivity outside the amp^{R} gene in pUC19. Figure S4 shows 316 decreases in the logarithmic relative concentration of ori as measured by the qPCR method with 317 318 amplicon sizes of 190, 390, and 530 bp during UV and UV/H₂O₂ treatment of extracellular pUC19. The reactivity and degradation patterns of e-ori were similar to those of e- amp^{R} when 319

320 considering their gene sizes (see the k values in Table 1).

321 Structural damage to pUC19 measured by gel electrophoresis

322 Figure 3a & 3b presents the images of agarose gel electrophoresis analyses of pUC19 323 (prepared at 5 μ g/mL) before and after UV and UV/H₂O₂ treatment (UV fluence range of 0–312 mJ/cm²). Intact pUC19 yielded a band corresponding to a size slightly larger than the 2-kb 324 325 molecular weight (MW) marker, and pUC19 treated within *EcoRI* showed a band corresponding 326 to a size slightly larger than the 3-kb MW marker. The observed band positions are consistent 327 with the size of pUC19 (i.e., 2686 bp) and the fact that plasmids typically exist in a supercoiled 328 form (i.e., intact pUC19), which migrates faster than the linear form (i.e., *EcoRI*-treated pUC19) 329 in an agarose gel because of their conformational difference. After the UV treatment, the band 330 position of pUC19 changed negligibly. In contrast, the band gradually moved upward with 331 increasing UV fluence during UV/H₂O₂ treatment, indicating the conformational change of 332 pUC19 from the supercoiled to linear form. On the basis of quantitative agarose gel image 333 analyses (Schneider et al., 2012), the degree of linearization of pUC19 was found to be less than 334 10% during the UV treatment but increased up to 93% during the UV/H_2O_2 treatment with an increase in UV fluence up to 312 mJ/cm² (Figure S5). These results can be explained by the fact 335

336 that UV mainly induces damage to the DNA bases such as CPDs (Görner, 1994; Sinha and 337 Häder, 2002) that is not detected in the agarose gel analysis. During UV/H₂O₂ treatment, 'OH is 338 formed and reacts with not only DNA bases but also the sugar phosphate backbone; the latter 339 reaction can cause significant DNA damage such as dsDNA breaks (von Sonntag, 2006) that are 340 detectable by agarose gel analysis. Finally, the rates of the plasmid structural (conformational) 341 change were compared to the rates of transforming activity loss during UV and UV/H_2O_2 342 treatment. Figure 3c shows that the transforming activity loss occurred much faster than the 343 structural degradation of the plasmid during UV and UV/H₂O₂ treatment. This finding indicates 344 that the UV-induced damage to DNA bases is mainly responsible for the elimination of the 345 transforming activity of plasmid-encoded ARGs. A similar conclusion has been reached in other 346 studies (Chang et al., 2017; Yoon et al., 2017).

347

348 **Discussion**

349 Effects of DNA repair and plasmid characteristics on the efficiency of elimination of 350 transforming activity

351 The efficiency of elimination of pUC19's transforming activity during UV irradiation, 352 determined in this study, can be compared with that in other studies, in which several different 353 plasmids were treated with UV (254 nm) and transformed into several strains of E. coli mutants 354 (Gurzadyan et al., 1993) or Acinetobacter baylvi (Chang et al., 2017) as host cells. Table 2 355 summarizes the k values for the elimination of transforming activity and corresponding 356 information on the plasmids and host cells used in these studies. Note that these plasmids 357 (pUC19, pTZ18R, pBR322, and pWH1266) all contain a β -lactamase resistance gene (either amp^{R} or bla_{TEM-1}). pBR322 and pWH1266 also contain a tetracycline resistance gene (*tetA*). 358

359 Data from previous work reveals that the efficiency of elimination of transforming activity 360 is greatly influenced by the type of *E. coli* recipient strain. When *E. coli* AB2480 (a double361 mutant strain lacking both *uvrA* and *recA* genes) was used as a recipient strain, the k value for UV-induced elimination of transforming activity was 5.3×10^{-1} cm²/mJ for pTZ18R and $6.7 \times$ 362 10^{-1} cm²/mJ for pBR322. These values are 22- and 15-fold larger than the k value for pTZ18R 363 $(2.4 \times 10^{-2} \text{ cm}^2/\text{mJ})$ and pBR322 $(4.4 \times 10^{-2} \text{ cm}^2/\text{mJ})$ when E. coli AB1157 (wild-type strain) 364 was used. Relatively low-to-intermediate levels of k were observed for pTZ18R when single-365 mutant E. coli AB1886 (deficient in uvrA, $k = 7.1 \times 10^{-2} \text{ cm}^2/\text{mJ}$) and E. coli AB2463 (deficient 366 in recA, $k = 6.7 \times 10^{-2}$ cm²/mJ) strains were tested (Gurzadyan et al., 1993). These results can 367 368 be explained by the different levels of DNA repair ability of the host cells. *uvrA* is one of the 369 *uvr* genes responsible for DNA repair through the nucleotide excision repair (NER) pathway 370 (Kisker et al., 2013). The recA gene is involved in various types of homologous recombination 371 and is essential for the repair and maintenance of DNA in prokaryotes (Lusetti and Cox, 2002). 372 Thus, the DNA repair in double-mutant *E. coli* AB2480 was negligible and resulted in a highly 373 efficient loss of plasmid transforming activity. In contrast, DNA repair was significant in the 374 other E. coli strains (i.e., single-mutant and wild-type strains), resulting in much slower 375 elimination of the transforming activity.

376 The k value determined here for UV-induced deactivation of pUC19 in E. coli DH5 α (= 6.1 $\times 10^{-2}$ cm²/mJ) is close to that for pTZ18R in the *E. coli* AB2463 strain (= 6.7 $\times 10^{-2}$ cm²/mJ). 377 378 Of note, both E. coli strains are deficient in the recA gene, and the sizes of the two plasmids are 379 comparable (2686 bp for pUC19 and 2861 bp for pTZ18R, Table 2). The k value for pWH1266 and wild-type A. baylyi is 1.1×10^{-1} cm²/mJ (Chang et al., 2017), which is larger than that for 380 381 pTZ18R and pBR322 with wild-type E. coli by a factor of 4.6 and 2.5, respectively. Notably, 382 the size of pWH1266 is larger than that of pTZ18R and pBR322 by a factor of 3.1 and 2.0, 383 respectively. Larger plasmids will typically contain a greater number of potential DNA damage 384 sites and would accordingly be expected to show a higher rate of UV-induced elimination of 385 transforming activity.

386 The rate of pyrimidine dimer formation vs. elimination of transforming activity of the 387 plasmid

388 CPDs and (6-4) photoproducts are known as the major types of UV-induced DNA damage 389 (Cadet and Douki, 2018; Görner, 1994; Sinha and Häder, 2002). These DNA damages are 390 readily formed at adjacent pyrimidine sites such as intrastrand thymine-thymine (TT), thymine-391 cytosine (TC), cytosine-thymine (CT), and cytosine-cytosine (CC) doublets. The formation of 392 CPDs usually predominates over formation of (6-4) photoproducts. In addition, the TT site is 393 the most photoreactive in terms of CPD formation among the bipyrimidine doublets (Douki and 394 Cadet et al., 2001; Douki, 2006). For instance, the following quantum yields of CPD formation 395 (in number of CPDs formed per number of photons absorbed by the entire target DNA) have been reported during UV₂₅₄ irradiation of dsDNA under typical solution conditions; 0.66×10^{-3} : 396 0.23×10^{-3} : 0.10×10^{-3} : 0.02×10^{-3} for TT : TC : CT : CC (Douki, 2006). The TC site also shows 397 high yield for (6-4) photoproduct with a quantum yield of 0.19×10^{-3} (Douki, 2006). In light of 398 399 this, we attempted to analyze the relationship between the formation rate of CPDs as the major 400 UV-induced DNA damage and consequent elimination of the transforming activity of plasmid 401 pUC19.

402 The UV fluence-based formation rate constant of CPDs (k_{CPDs} , cm²/mJ) in dsDNA can be 403 written as Eq. 2:

404
$$k_{\text{CPDs}} = (2.303 \times \varepsilon \times \Phi_{\text{CPDs}}) \div \text{U}$$
 [2]

405 where ε (M⁻¹ cm⁻¹) is the molar absorption coefficient, Φ_{CPDs} (mol/Einstein) is the quantum 406 efficiency of CPD formation, and U (= 4.72 × 10⁵ J/Einstein) is the molar photon energy at 254 407 nm [Lee et al. (2016) and references therein]. The Φ_{CPDs} values reported for dsDNA show 408 variations due to different ways of defining the quantum yield. A Φ_{CPDs} value of ~2.4 × 10⁻³ 409 (Görner, 1994) or ~1.0 × 10⁻³ (Douki, 2006) was obtained by considering the number of CPDs 410 formed per photons absorbed by the entire target DNA (Douki, 2006). In this case (Approach I, 411 $k_{\text{CPDs-I}}$), the ε value for the entire target DNA should be used. The ε value for all base pairs in a strand of dsDNA can be conveniently estimated as $\varepsilon_{bp} = \varepsilon_{sbp} \times (\# \text{ of base pairs})$ in which ε_{sbp} is 412 the average molar absorption coefficient of a single base pair ($\varepsilon_{sbp} = 15,000 \text{ M}^{-1} \text{ cm}^{-1}$, Tataurov 413 414 et al., 2008), and "# of base pairs" is the total number of base pairs in a given DNA. 415 Alternatively, the quantum yield could be defined based on the photons absorbed specifically by 416 two adjacent pyrimidines. In the current study, we focused on photon absorbance by TT sites, as 417 they are the most photoreactive for CPD formation amongst the four pyrimidine doublets TT, 418 TC, CT, and CC, and also reported to be the most slowly repaired in human skin cells (Cadet and Douki, 2018). In the case of TT sites (Approach II, $k_{\text{CPDs-II}}$), a $\Phi_{\text{CPDs.TT}}$ value of ~2.0 × 10⁻² 419 420 has been reported (Douki et al., 2000; Patrick, 1977). The ε value for all TT sites within a strand 421 of dsDNA can be calculated as $\varepsilon_{TT} = \varepsilon_{sTT,254} \times (\#TT)$, where ε_{sTT} is the molar absorption coefficient of a single TT site (= $8,400 \text{ M}^{-1} \text{ cm}^{-1}$; Patrick, 1977), and #TT is the number of TT 422 423 sites in a given DNA. Using the two different definitions of Φ_{CPDs} described above, k_{CPDs-I} and 424 $k_{\text{CPDs-II}}$ can be calculated for plasmids as a function of the number of base pairs and TT sites, 425 respectively.

Figure 4 and Figure S6 show a comparison of the calculated k_{CPDs} values (dashed lines) 426 427 with the measured k values for the elimination of transforming activity (symbols) from this 428 study (pUC19) and from the literature (pTZ18R, pBR322, and pWH1266) as a function of the 429 number of (a) base pairs (k_{CPDs-I}) and (b) TT sites ($k_{CPDs-II}$) in the plasmids. In addition, a Φ_{CPDs} , TT-II value of 2.0×10^{-2} was for Figure 4b and Figure S6b. Notably, $k_{\text{CPDs-I}}$ ($k_{\text{CPDs-I}}$ and $k_{\text{CPDs-II}}$) 430 431 were close to k measured when double-mutant E. coli was used as a recipient (differences 432 ranging from factors of 1 to 2.3). This finding indicates that in the double-mutant strain, most of 433 the CPDs formed in the plasmids led to the elimination of transforming activity because the 434 CPDs were not repaired in the corresponding host cell. The k values obtained with single 435 mutants and wild-type E. coli and A. baylyi strains were significantly lower than k_{CPDs} by factors ranging from 5 to 15 (indicated as the ratio of the slopes compared to k_{CPDs} , Figure 4). This 436 437 result strongly suggests a significant role of CPD repair activity in these host cells. One 438 alternative explanation for these trends may be that some DNA damages on the plasmid are 439 located where they do not inhibit the plasmid transformation, which requires further 440 investigation. Overall, good linear trends were found between the measured k values and the 441 number of base pairs and TT sites in the plasmids. These trends can be explained by the fact that 442 more CPDs are formed with increasing gene fragment size, or - more specifically - increasing 443 number of potential pyrimidine dimer sites. The observed linear relation can be further tested in 444 future studies by determining k for plasmids with different gene fragment sizes and 445 compositions, and in host cells with different DNA repair abilities.

446 The rate of pyrimidine dimer formation vs. gene damage measured by qPCR

447 Figure 5 and Figure S7 depict the plots of the measured k values (symbols) for the 448 degradation of target qPCR amplicons as a function of (a) the number of base pairs and (b) the 449 number of TT sites in the genes. In addition, k values were compared with the $k_{\text{CPDs-I}}$ and $k_{\text{CPDs-II}}$ 450 values of the target genes that could be calculated via Eq. 2 and the method described above $(\Phi_{\text{CPDs-I}} \text{ of } 2.4 \times 10^{-3} \text{ for Figure 5a}, 1.0 \times 10^{-3} \text{ for Figure S7a}, \text{ and } \Phi_{\text{CPDs,TT-II}} \text{ of } 2.0 \times 10^{-2} \text{ for}$ 451 452 Figure 5b and Figure S7b). The k data from UV treatment of extracellular and intracellular 453 plasmids as well as the k data from the UV/H_2O_2 treatment of intracellular plasmids (but not 454 extracellular plasmids) from this study and from the literature were included in the plots (Table 455 1). The UV/H_2O_2 data for i-ARGs were included because the damage to intracellular genes 456 during UV/H₂O₂ treatment was exclusively caused by UV, with little contribution from [•]OH. The k values for extracellular plasmids during UV/H_2O_2 treatment were not included because 457

they were much larger compared to those of UV treatment, indicating a significant contribution
of •OH reactions to extracellular gene damage.

460 The following points are noteworthy and follow from the results in Figure 5. First, the 461 intracellular genes (circles) show good linear relations with the numbers of base pairs and TT sites (dashed lines, $r^2 = 0.93$ for both cases), whereas k values for the extracellular genes 462 (triangles) correlated relatively poorly with the numbers of base pairs and TT sites. Second, the 463 number of TT sites ($k = (6.4 \times 10^{-4}) \times (\# \text{ of TT sites}), r^2 = 0.77, n = 25$) yielded a better linear 464 correlation with k as compared to the number of base pairs ($k = (8.9 \times 10^{-5}) \times (\# \text{ of base pairs})$), 465 $r^2 = 0.65$, n = 25) when both extracellular and intracellular genes were considered. This result 466 can again be explained by the fact that UV-induced lesions accumulate most rapidly at TT sites. 467 468 Consistent with this notion, k values reported for tetA (216 and 1200 bp) (Chang et al., 2017) 469 were much lower than the other k values shown in Figure 5a because the *tetA* gene contains 470 fewer TT sites for the same total number of base pairs (Table 1). Third, k values for the gene damage were close to k_{CPDs} . When compared on the basis of total base pairs, the predicted k_{CPDs} . 471 I values exceeded the measured k values by a factor of ~2 when a $\Phi_{\text{CPDs-I}}$ of 2.4 × 10⁻³ (Görner, 472 1994) was used (Figure 5a) but were similar to the measured k values when a $\Phi_{\rm CPDs-I}$ of 1.0 \times 473 10^{-3} (Douki, 2006) was used (Figure S7a). The variation in the CPD quantum yields in literature 474 475 could be attributed to different analytical methods for quantifying the dimeric photoproducts 476 (e.g., acid hydrolysis, chromatographic separation, and quantification by radioactivity or mass 477 spectrometry) or experimental conditions (e.g., DNA concentration, ionic strength). When 478 compared on the basis of total TT sites, the predicted $k_{\text{CPDs-II}}$ values exceeded the measured k 479 values by a factor of 1.5 (Figure 5b). This level of difference in k (i.e., within a factor of 2) can be considered minor, given the assumptions and uncertainties of the parameters in the $k_{\rm CPDs}$ 480 481 calculation. Thus, further attempts to improve the kinetic model for $k_{\text{CPDs-II}}$ prediction by 482 considering the photoproducts of bipyrimidine doublets other than TT were not pursued.

483 Overall, our results indicate that the qPCR method is sensitive enough to detect most major 484 UV-induced DNA damages (e.g., CPDs). This finding is consistent with other studies, which 485 show that the PCR amplification efficiency of genes containing a CPD is drastically reduced (by 486 $\sim 2 \log_{10}$) compared to the intact gene (Sikorsky et al., 2004).

487 Enhanced degradation of extracellular ARGs and its impact on elimination of the 488 transforming activity

489 The kinetic data on the gene damage from this study and from the literature suggest that 490 extracellular genes sometimes undergo more rapid degradation (by a factor of up to 1.7) than 491 the corresponding intracellular genes during UV irradiation (Yoon et al., 2017). This phenomenon was particularly noticeable for some long amplicons (i.e., amp^{R} and kan^{R} : 806–851 492 493 bp long). The higher reactivity of extracellular genes may have been caused by incidental 494 photochemical reactions of metal–DNA complexes. In particular, photolysis of Cu(II)–DNA 495 complexes has been known to produce Cu(I) and reactive oxygen species (ROS) including 'OH, 496 which can initiate oxidative DNA damage (Matzeu and Onori, 1986). Even though the plasmid 497 samples in this study were isolated from E. coli with a purification step (i.e., washing with 498 phosphate buffer), a certain amount of Cu(II) complexed with DNA could have remained. This explanation can be supported by the observation that the degradation rate of e-amp^R amplicons 499 500 during UV irradiation of pUC19 decreased by a factor of 1.1-1.7 (p < 0.05) after the addition of 501 EDTA (0.1 mM) or methanol (10 mM; Figure S8). EDTA can lower photoreactivity of Cu(II)-502 DNA complexes by forming a less photoreactive metal-EDTA complex (Natarajan and 503 Endicott, 1973). Methanol can protect DNA by scavenging 'OH (von Sonntag, 2006). Furthermore, the degradation rate of e-amp^R increased by a factor of ~1.3 (p < 0.05) after the 504 505 addition of CuSO₄ (10 µM; Figure S8). Similar photochemical reactions of Cu(II)-DNA 506 complexes could have proceeded and contributed to the enhanced degradation of the 507 extracellular genes. Nonetheless, this was not the case for the intracellular genes, presumably 508 owing to significant scavenging of UV-induced ROS by intracellular reductants (e.g., sulfur-509 containing proteins). In agreement with these data, the degradation of extracellular ARGs was 510 significantly enhanced by H₂O₂ addition during UV irradiation of plasmids (Figure 2 and Table 511 1), whereas no impact of H_2O_2 addition was observed for intracellular ARGs, owing to nearly 512 complete scavenging of 'OH by intracellular components (Yoon et al., 2017). These 513 observations may in turn provide an explanation for the relatively poor correlations of measured 514 k values with number of base pairs or TT sites for extracellular ARGs in comparison to 515 intracellular ARGs (Figure 5).

516 It is noteworthy that the enhanced degradation of extracellular ARGs beyond the direct UV-517 induced damage does not lead to more rapid elimination of the transforming activity. The rate of 518 elimination of transforming activity was nearly constant across the UV and UV/H₂O₂ treatments 519 of e-ARGs, while the gene degradation rates varied by a factor of up to 1.8 (p < 0.05) for the 520 same treatments (Figure 2 and Table 1). Thus, the damage to extracellular ARGs by UV, [•]OH, 521 or other ROS was detectable by the qPCR method, but in the *E. coli* transformation assay, only 522 the direct UV-induced gene damage could be detected. This finding could be due to efficient 523 repair of the gene damage caused by 'OH (or other ROS) in the *E. coli* transformation system. 524 Single oxidized bases (e.g., 5,6-dihydroxy-5,6-dihydrothymine) are the most frequent type of nucleobase damage in cellular DNA from exposure to 'OH, and most such lesions are 525 526 efficiently removed by the cellular base excision repair system (Cadet and Wagner, 2013). Such 527 DNA repair function does not exist in the qPCR system. In addition to the DNA repair, the 528 difference in fidelity between E. coli DNA polymerase and Taq DNA polymerase could have 529 played a role. E. coli is known to possess specialized lesion bypass DNA polymerases with low 530 fidelity that enable translesion replication of damaged DNA (Maor-Shoshani et al., 2003; Nevin 531 et al., 2017). In contrast, Taq polymerase of the qPCR method is a relatively high-fidelity polymerase with high enough sensitivity to detect even minor single-base lesions (Sikorsky etal., 2004).

534

535 Conclusions

• Under typical UV fluences for disinfection purposes (e.g., 40 mJ/cm²), a ~1 log reduction in
the transforming activity of a plasmid-encoded ARG is expected. To achieve more extensive
elimination of the transforming activity (e.g., >4 log reduction), UV fluence of more than 150
mJ/cm² is required. Addition of H₂O₂ (i.e., the UV/H₂O₂ advanced oxidation process) does not
significantly enhance the efficiency of elimination of the transforming activity.

Efficiency of elimination of the transforming activity for a plasmid-encoded ARG during UV
treatment depends on the rate of formation of CPDs in the plasmid and the repair of such DNA
damage during the transformation process in host cells. Significant capacity for CPD repair is
present in the *E. coli* recipient strain (DH5α) used in this study and is also expected in many
wild-type bacterial cells.

• The rate of formation of CPDs can be calculated by considering the number of TT sites in target gene fragments or whole plasmids and the TT-specific photochemical reaction parameters (i.e., $\Phi_{CPD,TT} = 2.0 \times 10^{-2}$ and $\varepsilon_{TT} = 8,400 \text{ M}^{-1} \text{ cm}^{-1}$ for 254 nm light). Additionally, CPDs can be sensitively detected and quantified by qPCR.

• CPD formation is the major DNA damage mechanism and responsible for the elimination of 551 transforming activity of extra- and intracellular plasmids during UV and UV/H₂O₂ treatment.

• For extracellular plasmids, DNA base oxidation takes place in addition to CPD formation, and these lesions are detectable by the qPCR method. The DNA base oxidation, however, does not

reduce the transforming activity of pUC19 in the *E. coli* recipient strain (DH5 α) utilized here.

555

556 Supporting Information

557 Three texts, two tables, and eight figures are available as supplementary materials and data.

558

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565 **References**

- Alexander, J., Knopp, G., Dötsch, A., Wieland, A., Schwartz, T., 2016. Ozone treatment of
 conditioned wastewater selects antibiotic resistance genes, opportunistic bacteria, and
 induce strong population shifts. Sci. Total. Environ. 559, 103-112.
- Allen et al., 2010, Call of the wild: antibiotic resistance genes in natural environments. Nature
 Rev. 8, 251-259.
- 571 Berendonk, T., Manaia, C.M., Merlin, C., Fatta-Kassinos, D., Cytryn, E., Walsh, F., Bürgmann,
- 572 H., Sørum, H., Norström, M., Pons, M.-N., Kreuzinger, N., Huovinen, P., Stefani, S.,
- 573 Schwartz, T., Kisand, V., Baquero, F., Martinez, J.L., 2015. Tackling antibiotic resistance:
- the environmental framework. Nature reviews Microbiol. 13, 310–317.
- 575 Bioneer, AccuPrep[®] Nano-Plus Plasmid Mini/Midi/Maxi Extraction Kit, User's Guide. Korea,
- 576 2016. Available at <u>http://us.bioneer.com/Protocol/AccuPrep%20Nano-</u>
 577 <u>Plus%20Plasmid%20Extraction%20Kit.pdf</u>.
- 578 Bolton, J.R., Linden, K.G., 2003, Standardization of methods for fluence (UV dose)

- determination in bench-scale UV experiments, J. Environ. Eng., 129, 209–215.
- 580 Cadet, J., Wagner, J.R., 2013, DNA base damage by reactive oxygen species, oxidizing agents,
- and UV radiation. Cold Spring Harb Perspect Biol. 5, a012599.
- 582 Cadet, J., Douki, T., Formation of UV-induced DNA damage contributing to skin cancer
- 583 development, Photochem. Photobiol. Sci., 2018.
- 584 Casali, N., Preston, A. E. coli Plasmid Vectors. Methods and Applications. Humana Press Inc.,
- 585 New York, Springer, 2003.
- 586 Chang et al., Degradation of Extracellular Antibiotic Resistance Genes with UV₂₅₄ Treatment,
- 587 Environ. Sci. Technol., 2017, 51, 6185-6192.
- 588 Chen, H., Zhang, M., 2013. Effects of advanced treatment systems on the removal of antibiotic
- resistance genes in wastewater treatment plants from Hangzhou, China. Environ. Sci.
 Technol. 47, 8157-8163.
- 591 Colomer-Lluch, M., Jofre, J., Muniesa, M., 2011. Antibiotic resistance genes in the 592 bacteriophage DNA fraction of environmental samples, PloS One, 6, e17549.
- 593 Czekalski, N., Berthold, T., Caucci, S., Egli, A., Bürgmann, H., 2012. Increased levels of
 594 multiresistant bacteria and resistance genes after wastewater treatment and their
 595 dissemination into Lake Geneva, Switzerland. Front. Microbiol. 3, 1-18.
- 596 Destiani, R., Templeton, M.R., Kowalski, W., 2017. Relative ultraviolet sensitivity of selected
- antibiotic resistance genes in waterborne bacteria. Environ. Eng. Sci. In press.
 https://doi.org/10.1089/ees.2017.0179.
- 599 Dodd, M.C., 2012. Potential impacts of disinfection processes on elimination and deactivation
- of antibiotic resistance genes during water and wastewater treatment. J. Environ. Monit. 14,
 1754–1771.
- Douki, T., Cadet J., Individual determination of the yield of the main UV-induced dimeric
 pyrimidine photoproducts in DNA suggests a high mutagenicity of CC photolesions,

- 604 Biochemistry 2001, 40, 2495-2501.
- 605 Douki, T. Low ionic strength reduces cytosine photoreactivity in UVC-irradiated isolated DNA,
- 606 Photochem. Photobiol. Sci. 2006, 5, 1045-1051.
- 607 Ferro, G., Guarino, F., Cicatelli, A., Rizzo, L., 2017. β-lactams resistance gene quantification in
- an antibiotic resistant *Escherichia coli* water suspension treated by advanced oxidation with
- 609 UV/H₂O₂. J. Hazard. Mater. 323, 426-433.
- 610 Gerrity, D., et al 2016. Emerging investigators series: prediction of trace organic contaminant
- 611 abatement with UV/H₂O₂: development and validation of semi-empirical models for

612 municipal wastewater effluents, Environ. Sci.: Water Res, 2, 460-473.

- 613 Görner, H., Photochemistry of DNA and related biomolecules: quantum yields and 614 consequences of photoionization. J. Photochem. Photobiol. B. Biology, 1994, 26, 117–139.
- 615 Gurzadyan, G.G., Görner, H., Schulte-Frohlinde, D., Photolesions and biological inactivation of
- plasmid DNA on 254nm irradiation and comparison with 193nm laser irradiation,
 Photochem. Photobiol., 1993, 58, 477-485.
- Hanahan D, 1983, Studies on transformation of *Escherichia coli* with plasmids. J. Mol. Biol.,
 166, 557-580.
- Hanahan, D., Jessee, J., Bloom, F.R., Plasmid transformation of *Escherichia coli* and other
 bacteria. Methods in Enzymology, 1991, 204, 63-113.
- Jacangelo, J.G., Trussell, R.R., 2002. International report: water and wastewater disinfectiontrends, issues, and practices. Water Sci. Technol. Water Supply 2, 147-157.
- Kisker, C., Kuper, J., van Houten, B., Prokaryotic nucleotide excision repair, Cold Spring Harb
 Perspect Biol, 2013, 5, a012591.
- 626 Lee, Y., Gerrity, D., Lee, M., Gamage, S., Pisarenko, A., Trenholm, R.A., Canonica, S., Snyder,
- 627 S.A., von Gunten, U., 2016. Organic contaminant abatement in reclaimed water by
- VV/H_2O_2 and a combined process consisting of O_3/H_2O_2 followed by UV/H_2O_2 : prediction

of abatement efficiency, energy consumption, and byproduct formation. Environ Sci
Technol, 50, 3809–3819.

- 631 Lee, Y., von Gunten, U., 2016. Advances in predicting organic contaminant abatement during
- ozonation of municipal wastewater effluent: reaction kinetics, transformation products, and

changes of biological effects, Environ. Sci.: Water Res, 2, 421-442.

- Luby, E., Ibekwe, A.M., Zilles, J., Pruden, A., 2016. Molecular methods for assessment of
- antibiotic resistance in agricultural ecosystems: prospects and challenges. J. Environ.
 Quality 45, 441-453.
- 637 Lüddeke, F., Heb, S., Gallert, C., Winter, J., Güde, H., Löffler, H., 2015. Removal of total and
- 638 antibiotic resistant bacteria in advanced wastewater treatment by ozonation in combination
- 639 with different filtering techniques. Water Res. 69, 243-251.
- Lusetti, S.L., Cox, M.M., The bacterial RecA protein and the recombinational DNA repair of
 stalled replication forks, Annual Rev. Biochem. 2002, 71, 71-100.
- Maor-Shoshani, A., Ben-Ari, V., Livneh, Z., 2003. Lesion bypass DNA polymerase replicate
 across non-DNA segments. PNAS, 100, 14760-14765.
- 644 Matzeu, M., Onori, G., 1986. Effect of mid-UV radiation on DNA-Cu²⁺ complex: absorption
- and circular dichroism study. Photochem. Photobiol. 44, 59-65.
- 646 Mouret, S., Charveron, M., Favier, A., Cadet, J., & Douki, T., 2008. Differential repair of UVB-
- 647 induced cyclobutane pyrimidine dimers in cultured human skin cells and whole human648 skin. DNA repair, 7(5), 704-712.
- 649 McKinney, C.W., Pruden, A., 2012. Ultraviolet disinfection of antibiotic resistant bacteria and
- their antibiotic resistance genes in water and wastewater. Environ. Sci. Technol, 46,
 13393–13400.
- Miklos, D.B., Hartl, R., Michel, P., Linden, K.G., Drewes, J.E., Hübner, U., 2018. UV/H₂O₂
- 653 process stability and pilot-scale validation for trace organic chemical removal from

- 654 wastewater treatment plant effluents. Water Res, in press.
- 655 Natarajan, P., Endicott, J.F., 1973. Photoredox behavior of transition metal-
- 656 ethylenediaminetetraacetate complexes. A comparison of some group VII metals. J. Phys.
- 657 Chem. 77, 2049-2054.
- 658 Nevin, P., Gabbai, C.C., Marians, K.J., 2017. Replisome-mediated translesion synthesis by a
- 659 cellular replicase. J. Biol. Chem. 292, 13833-13842.
- Palmen et al., Physiological characterization of natural transformation in *Acinetobacter calcoaceticus*, J. General Microbiol, 1993, 139, 295-305.
- 662 Pak, G., Salcedo, D. E., Lee, H., Oh, J., Maeng, S. K., Song, K. G., Hong, S. W., Kim, H.-C.,
- 663 Chandran, K., Kim, S., 2016. Comparison of antibiotic resistance removal efficiencies using
 664 ozone disinfection under different pH and suspended solids and humic substance
 665 concentrations. Environ. Sci. Technol. 50, 7590–7600.
- 666 Patrick, M.H., 1977. Studies on thymine-derived UV photoproducts in DNA-I. Formation and
- biological role of pyrimidine adducts in DNA. Photochem. Photobiol. 25, 357-372.
- 668 Pingound, A., Jeltsch, A., Structure and function of type II restriction endonucleases, Nucleic
- 669 Acids Research, 2001, 29, 3705-3727.
- 670 Pruden, A., 2014. Balancing water sustainability and public health goals in the face of growing
- 671 concerns about antibiotic resistance. Environ. Sci. Technol. 48, 5–14.
- 672 Rizzo, L., Manaia, C., Merlin, C., Schwartz, T., Dagot, C., Poly, M.C., Michael, I., Fatta-
- 673 Kassinos, D., 2013. Urban wastewater treatment plants as hotspots for antibiotic resistant
- bacteria and genes spread into the environment: a review. Sci. Total. Environ. 447, 345e360.
- 675 Shanehbandi D, Saei AA, Zarredar H, Barzegari A., Vibration and glycerol-mediated plasmid
- DNA transformation for Escherichia coli, FEMS Microbiol Lett, 2013, 348, 74–78
- 677 Schneider, C. A., Rasband, W. S., & Eliceiri, K. W., 2012, NIH Image to ImageJ: 25 years of
- 678 image analysis, Nature methods, 9, 671-675.

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- 679 Schulte-Frohlinde, D., Biological consequences of strand breaks in plasmid and viral DNA, Br.
- 680 J. Cancer, 1987, 55, 129-134.
- 681 Sikorsky, J.A., Primerano, D.A., Fenger, T.W., Denvir, J., Effect of DNA damage on PCR 682 amplification efficiency with the relative threshold cycle method, BBRC, 2004, 323, 823-
- 683 830.
- Sinha, R.P., Häder, D.-P., UV-induced DNA damage and repair: a review, Photochem.
 Photobiol. Sci., 2002, 1, 225-236.
- 686 Sousaa, J.M., Macedob, G., Pedrosac, M., Becerra-Castroa, C., Castro-Silvad, S., Pereirac,
- 687 M.F.R., Silva, A.M.T., Nunesa, O.C., Manaia, C.M., 2017. Ozonation and UV254nm
- radiation for the removal of microorganisms and antibiotic resistance genes from urban
- 689 wastewater. J. Hazard. Mater. 323, 434-441.
- Tan, D.T., Shuai, D., 2015. Research highlights: antibiotic resistance genes: from wastewater
 into the environment. Environ. Sci.: Water Res, 1, 264-267.
- 692 Tataurov, A.V., You, Y., Owczarzy, R., 2008. Predicting ultraviolet spectrum of single stranded
- and double stranded deoxyribonucleic acids. Biophys. Chem. 133, 66–70.
- 694 Thomas, C.M., Nielsen, K.M., 2005. Mechanisms of, and barriers to, horizontal gene transfer
- between bacteria, Nature Reviews Microbiology 3, 711-721.
- 696 Vikesland, P.J., et al., 2017. Toward a comprehensive strategy to mitigate dissemination of
 697 environmental sources of antibiotic resistance, Environ. Sci. Technol., 51, 13061-13069.
- Von Sonntag, C., 2006. Free-radical-induced DNA Damage and its Repair. A Chemical
 Perspective. Springer Verlang, Heidelberg, Berlin.
- 700 World Health Organization, 2014. Antimicrobial Resistance: Global Report on Suveillance.701 WHO.
- Yoon et al., 2017. Inactivation efficiency of plasmid-encoded antibiotic resistance genes during
- water treatment with chlorine, UV, and UV/H_2O_2 , Water Res, 123, 783-793.

- 704 Zhang, Y., et al., 2018. Cell-free DNA: A neglected source for antibiotic resistance genes
- spreading from WWTPs. Environ. Sci. Technol. 52, 248-257.

Table 1. Fluence-based rate constants (k_{CPD-I} and k_{CPD-II}) for gene damage quantified by qPCR during treatment of intracellular and extracellular plasmids with UV and UV/H₂O₂.

Gene	#base	#TT	Plasmid	$k, \text{ cm}^2/\text{mJ}^{\text{ c}}$	k _{CPDs}	k _{CPDs}	Ref. ^f	
	pairs ^a	sites ^b	location		(bps),	(TTs),		
			L IX	7 Amo o Amo o má	cm²/mJ "	cm ² /mJ ^c		
			UN	$1.7(+0.00)\times10^{-2}$				
	192	42	Illua Extra	$1.7(\pm 0.09) \times 10$ 1.0(+0.11)×10 ⁻²	3.4×10 ⁻²	3.4×10 ⁻²		
			Intra	$1.9(\pm 0.11) \times 10$ 2.8(±0.06)×10 ⁻²			-	
amp ^R	400	67	Fytra	$2.8(\pm 0.00) \times 10^{-2}$	7.0×10^{-2}	5.5×10^{-2}	This	
(n UC19)			Intra	$3.1(\pm 0.09) \times 10^{-2}$			study	
(poer))	603	96	Fxtra	$5.4(\pm 0.16) \times 10^{-2}$	1.1×10^{-1}	7.9×10^{-2}	study	
			Intra	$7.4(\pm 0.10) \times 10^{-2}$			-	
	851	118	Extra	$1.2(\pm 0.33)\times 10^{-1}$	1.5×10^{-1}	9.7×10 ⁻²		
	100		Enu	1.0(±0.05)×10		2 4 4 2 2		
ori	190	29	Extra	$1.7(\pm 0.09) \times 10^{-2}$	3.3×10^{-2}	2.4×10^{-2}	This	
(pUC19)	390	44	Extra	$2.9(\pm 0.12) \times 10^{-2}$	6.9×10^{-2}	3.6×10^{-2}	study	
ч ,	530	59	Extra	$4.0(\pm 0.24) \times 10^{-2}$	9.3×10 ⁻²	4.8×10^{-2}		
amp^{R}			Intra	6.8×10 ⁻²		9.8×10 ⁻²	Yoon	
(pUC4k)	850	119	Extra	1.1×10 ⁻¹	1.5×10 ⁻¹		et al, 2017	
kan ^R		134	Intra	8.1×10^{-2}	1.4×10 ⁻¹	1.1×10 ⁻¹	Yoon	
(pUC4k)	806		Extra	1.4.10-1			et al,	
				1.4×10 ⁻			2017	
	209	30	Extra	5.5×10 ⁻³	3.7×10 ⁻²	2.5×10 ⁻²	Chang	
$bla_{\text{TEM-1}}$	861	122	Extra	6.8×10^{-2}	1.5×10^{-1}	1.0×10^{-1}	et al,	
	801	125	Extra	0.0×10	1.3×10	1.0×10	2017	
	216	18	Extra	3.8×10 ⁻²	3.8×10 ⁻²	1.5×10^{-2}	Chang et al, 2017	
tetA	1200	80	Extra	5.8×10 ⁻²	2.1×10 ⁻¹	6.6×10 ⁻²		
		1		. O. treatment				
			Intra	$1.6(\pm 0.09) \times 10^{-2}$	_			
	192	42	Fxtra	$1.0(\pm 0.09) \times 10^{-2}$	3.4×10 ⁻²	3.4×10 ⁻²	This study	
<i>amp^R</i> (pUC19)	400	67	Intra	$2.9(\pm 0.14) \times 10^{-2}$	7.0×10 ⁻²	5.5×10 ⁻²		
			Extra	$41(\pm0.12)\times10^{-2}$				
	603	96	Intra	$4.1(\pm 0.10) \times 10^{-2}$	1.1×10 ⁻¹	7.9×10 ⁻²		
			Extra	$6.7(\pm 0.28) \times 10^{-2}$			staaj	
	851	118	Intra	$73(\pm0.30)\times10^{-2}$	1.5×10 ⁻¹			
			Extra	$1.8(\pm 0.06) \times 10^{-1}$		9.7×10 ⁻²		
	100	20	Extra	$2.3(\pm 0.05) \times 10^{-2}$	3.3×10^{-2}	2.4×10^{-2}		
ori	390	44	Extra	$3.9(\pm0.03)\times10^{-2}$	6.9×10^{-2}	3.6×10^{-2}	This study	
(pUC19)		1 77		5.7(±0.14/^10	0.7^10	5.0^10		
	530	59	Extra	$4.7(\pm 0.21) \times 10^{-2}$	9.3×10^{-2}	4.8×10^{-2}	study	

(pUC4k)			Extra	1.7×10 ⁻¹			et al, 2017
kan ^R			Intra	8.0×10 ⁻²			Yoon
(pUC4k)	806	134	Extra	2.5×10 ⁻¹	1.4×10 ⁻¹	1.1×10 ⁻¹	et al, 2017

^aTotal number of base pairs, ^bTotal number of TT sites, ^cFluence-based rate constant measured from this study and literature, ^dFluence-based rate constant for CPD formation calculated using $k_{CPDs} =$ $(2.303 \times \varepsilon_{bp} \times \Phi_{CPD,I}) / U$, $\varepsilon_{bp} = \varepsilon_{sbp} \times (\#base pair) = (15,000 \text{ M}^{-1} \text{ cm}^{-1}) \times (\#base pair)$, and $\Phi_{CPD,I} =$ 2.4×10^{-3} (Gorner, 1994), ^eCalculated using $k_{CPDs} = (2.303 \times \varepsilon_{TT} \times \Phi_{CPD,TT-II}) / U$, $\varepsilon_{TT} = \varepsilon_{sTT,254} \times$ $(\#TT) = (8,400 \text{ M}^{-1} \text{ cm}^{-1}) \times (\#TT)$, and $\Phi_{CPD,TT-II} = 2.0 \times 10^{-2}$ (Douki et al., 2000), ^fSource for the measured *k* values.

pUC19 2686 383 E. coli DH5 α (recA') 6.1×10^{-2} 4.7×10^{-1} 3.1×10^{-1} This study pTZ18R 2861 453 $\begin{bmatrix} E. coli K12 \\ AB2480 (uvrA^{-}) \\ \hline E. coli K12 \\ AB1886 (uvrA^{-}) \\\hline E. coli K12 \\ AB2463 (recA^{-}) \\\hline E. coli K12 \\ AB1157 (wild type) \\\hline E. coli K12 \\ AB2480 (uvrA^{-}, recA^{-}) \\\hline E. coli K12 \\ AB1157 (wild type) \\\hline T.7 \times 10^{-1} \\\hline 4.2 \times 10^{-1} \\\hline 4.2 \times 10^{-1} \\\hline 4.2 \times 10^{-1} \\\hline 4.2 \times 10^{-1} \\\hline 1993 \\\hline $	Plasmid	#base pairs ^a	#TT sites ^b	Host cell for transformation	k, cm ² /mJ ^c	k_{CPDs} (bps), cm ² /mJ ^d	k_{CPDs} (TTs), cm^2/mJ^e	Ref. ^f
$ pTZ18R = 2861 = 453 = 453 = 453 = \frac{E.\ coli\ K12}{AB2480\ (uvrA^{-}, \ recA^{-})} = \frac{5.3 \times 10^{-1}}{AB2480\ (uvrA^{-}, \ recA^{-})} = 5.0 \times 10^{-1} = 3.7 \times 10^{-1} = 4.2 \times 10^{-1} =$	pUC19	2686	383	<i>E. coli</i> DH5α (recA ⁻)	6.1×10 ⁻²	4.7×10 ⁻¹	3.1×10 ⁻¹	This study
pBR322 4361 513 $\begin{array}{c c} E. \ coli \ K12 \\ AB2480 \ (uvrA^{-}, \\ recA^{-}) \\ \hline E. \ coli \ K12 \\ AB1157 \ (wild \\ type) \end{array} \begin{array}{c} 6.7 \times 10^{-1} \\ 7.7 \times 10^{-1} \\ 4.2 \times 10^{-1} \end{array} \begin{array}{c} Gurza \\ an \ et \ a \\ 1993 \end{array}$	pTZ18R	2861	453	<i>E. coli</i> K12 AB2480 (uvrA ⁻ , recA ⁻) <i>E. coli</i> K12 AB1886 (uvrA ⁻) <i>E. coli</i> K12 AB2463 (recA ⁻) <i>E. coli</i> K12 AB1157 (wild type)	5.3×10 ⁻¹ 7.1×10 ⁻² 6.7×10 ⁻² 2.4×10 ⁻²	5.0×10 ⁻¹	3.7×10 ⁻¹	Gurzad an et al 1993
	pBR322	4361	513	<i>E. coli</i> K12 AB2480 (uvrA ⁻ , recA ⁻) <i>E. coli</i> K12 AB1157 (wild type)	6.7×10 ⁻¹ 4.4×10 ⁻²	7.7×10 ⁻¹	4.2×10 ⁻¹	Gurzad an et al 1993

Table 2. Fluence-based rate constants (k_{CPD-I} and k_{CPD-II}) for the elimination of transforming activity of ARGs during UV treatment of extracellular plasmids.

pWH12668890- gAcinetobacter
baylyi 1.1×10^{-1} 1.56- gChang et
al, 2017aTotal number of base pairs, bTotal number of TT sites, cFluence-based rate constant measured from
this study and literature, dFluence-based rate constant for CPD formation calculated using $k_{CPDs} =$
 $(2.303 \times \varepsilon_{bp} \times \Phi_{CPD,I}) / U$, $\varepsilon_{bp} = \varepsilon_{sbp} \times (\#base pair) = (15,000 \text{ M}^{-1} \text{ cm}^{-1}) \times (\#base pair)$, and $\Phi_{CPD,I} =$
 2.4×10^{-3} (Gorner, 1994), eCalculated using $k_{CPDs} = (2.303 \times \varepsilon_{TT} \times \Phi_{CPD,TT-II}) / U$, $\varepsilon_{TT} = \varepsilon_{sTT,254} \times (\#TT) = (8,400 \text{ M}^{-1} \text{ cm}^{-1}) \times (\#TT)$, and $\Phi_{CPD,TT-II} = 2.0 \times 10^{-2}$ (Douki et al., 2000) ffsured k values, gNot available because the full gene sequence is not known.



Figure 1. Concentration of transformants and resulting transforming activity as a function of pUC19 concentration during transformation of amp^R to *E. coli* DH5 α . The error bars represent one-standard deviation of more than three replicate measurements.



Figure 2. Logarithmic relative concentration of the transforming activity (\bigstar) and amp^R qPCR amplicons (192 bps (\checkmark), 400 bps (\bigcirc), 603 bps (\square) and 851 bps (\diamondsuit)) as a function of UV fluence during treatment of (a, b) intracellular and (c, d) extracellular pUC19 with (a, c) UV and (b, d) UV/H₂O₂ ([H₂O₂]₀ = 10 mg/L) at pH 7. The symbols represent the measured data and the error bars represent one standard deviation from triplicate experiments. The lines are linear regressions of the data.



Figure 3. Agarose gel electrophoresis images of extracellular pUC19 plasmids treated with (a) UV and (b) UV/H₂O₂ ($[H_2O_2]_0 = 10 \text{ mg/L}$) as a function of UV fluence (0 – 312 mJ/cm²). The first column shows gel images of standard ladders. The last column shows gel images of the pUC19 plasmid treated by restriction enzyme (*EcoRI*). (c) Logarithmic-scale decreases of the structural integrity (from quantitative analysis of the electrophoresis images) and transforming activity of pUC19 as a function of UV fluence.



Figure 4. UV fluence-based rate constants (*k*) for elimination of transforming activity of plasmids as a function of number of (a) base pairs (k_{CPDs-1}) and (b) TT sites ($k_{CPDs-II}$). The *k* data were obtained from this study (pUC19) and taken from Gurzadyan et al, 1993 (pTZ18r and pBR322), and Chang et al, 2017 (pWH1266) (see Table 2). The long-dash lines indicate the rate constants calculated for CPD formation across the entire plasmid (k_{CPDs}). For (a), a Φ_{CPDs-I} value of 2.4 × 10⁻³ (Görner, 1994) was used to calculate the k_{CPDs-I} for the photons absorbed by entire DNA. For (b), a $\Phi_{CPDs-TT}$ value of 2.0 × 10⁻² (Douki et al., 2000) was used to calculate the $k_{CPDs-II}$ for the photons absorbed by all TT sites within entire DNA. The short-dash and dotted lines are the k_{CPDs} multiplied with slope factors (ranging from 1/15 to 1/5) that were obtained to fit the measured *k* values. See the main text for further explanation.



Figure 5. UV fluence-based rate constants (*k*) for gene damage of extracellular (e-ARGs, triangles) and intracellular (i-ARGs, circles) plasmid-encoded genes as a function of number of (a) base pairs (k_{CPD-I}) and (b) TT sites (k_{CPD-II}). The *k* data were obtained from this study, and taken from Yoon et al., 2017, and Chang et al., 2017 (see Table 1). The long-dash lines indicate the calculated rate constant for the CPD formation (k_{CPDs}). For (a), a Φ_{CPDs-I} value of 2.4 × 10⁻³ (Görner, 1994) was used, and for (b), a $\Phi_{CPDs-TT}$ value of 2.0 × 10⁻² (Douki et al., 2000) was used. The short-dash lines are linear regressions of the *k* values of intracellular genes in which the relative slope of *k* compared to k_{CPDs} are indicated. See Figure 4 and the main text for further explanation.

