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Dynamics of residential indoor gas- and particle-phase water-soluble organic carbon: measurements during the CASA experiment†

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Previous time-integrated (2 h to 4 h) measurements show that total gas-phase water-soluble organic carbon (WSOC_g) is 10 to 20 times higher inside homes compared to outside. However, concentration dynamics of WSOC_g and total particle phase WSOC (WSOC_p)—are not well understood. During the Chemical Assessment of Surfaces and Air (CASA) experiment, we measured concentration dynamics of WSOC_g and WSOC_p inside a residential test facility in the house background and during scripted activities. A total organic carbon (TOC) analyzer pulled alternately from a particle-into-liquid sampler (PILS) or a mist chamber (MC). WSOC_g concentrations ($215 \pm 29 \mu\text{g-C m}^{-3}$) were generally 36× higher than WSOC_p ($6 \pm 3 \mu\text{g-C m}^{-3}$) and 20× higher than outdoor levels. A building-specific emission factor (E_f) of 31 mg-C h^{-1} maintained the relatively high house WSOC_g background, which was dominated by ethanol ($46 \mu\text{g-C m}^{-3}$ to $82 \mu\text{g-C m}^{-3}$). When we opened the windows, WSOC_g decayed slower (2.8 h^{-1}) than the air change rate (21.2 h^{-1}) and E_f increased (243 mg-C h^{-1}). The response (increased E_f) suggests WSOC_g concentrations are regulated by large near surface reservoirs rather than diffusion through surface materials. Cooking and ozone addition had a small impact on WSOC, whereas surface cleaning, volatile organic compound (VOC) additions, or wood smoke injections had significant impacts on WSOC concentrations. WSOC_g concentration decay rates from these activities (0.4 h^{-1} to 4.0 h^{-1}) were greater than the normal operating 0.24 h^{-1} air change rate, which is consistent with an important role for surface removal.

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Environmental significance

People spend a considerable amount of time indoors, and thus the indoor environment is a potentially important exposure location. As part of the Chemical Assessment of Surfaces and Air (CASA) experiment, influences on the concentration dynamics of gas and particle water-soluble organic carbon (WSOC) in indoor air were investigated in a residential test facility. A system for semi-continuous, real-time monitoring of gas and particle WSOC was used to provide quantitative insights into the concentration dynamics of organic matter in indoor air and the role of surfaces as a source, sink and reservoir.

Introduction

People spend the majority of their time indoors,¹ where they are exposed to elevated air concentrations of many organic

compounds. In fact, total gas-phase water soluble organic carbon (WSOC_g) concentrations are 10 to 20 times higher inside homes compared to outside, as demonstrated by a study of 13 eastern U.S. homes with 2 h to 4 h integrated measurements.²

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Drawing from studies of individual water-soluble compounds, we expect indoor (total) WSOC_g concentrations would be elevated due to limited outdoor ventilation³ and that WSOC_g concentration dynamics would be influenced by a plethora of indoor sources such as material off-gassing,^{4–8} oxidation chemistry,^{9,10} human occupancy,^{8,11} cooking,¹² cleaning,¹³ fire-place usage,¹⁴ and other human activities.^{6,8} We also expect WSOC_g to partition to indoor particles, surface films, and other reservoirs, prolonging and altering indoor exposures.^{15–19} Additionally, we expect the surface partitioning and chemistry of WSOC_g (*i.e.*, oxygenated VOCs or OVOCS) to depend on relative humidity, surface properties, and hygroscopicity of surface soiling.²⁰ For this reason, it is not clear whether emission rates of WSOC_g from indoor surfaces will be unaffected (as they are for VOCs),^{21–23} or will be enhanced (as they are for semi-volatile organic compounds)^{24,25} when air concentrations are reduced, *i.e.* due to increased outdoor ventilation. Reactive uptake of water soluble organics by atmospheric waters (*i.e.*, in aerosols, fogs and clouds) is an important source of secondary organic aerosol (SOA) outdoors,^{26–28} and reactive uptake may also play a role in the chemistry of damp indoor surfaces. However, only a few of these inferences have been tested in indoor environments, so the sources, sinks, reservoirs, and dynamics of WSOC_g and particle-phase WSOC (WSOC_p) in indoor air are not well understood. Furthermore, an understanding the composition of WSOC is important to developing a comprehensive understanding of the chemistry, and yet there is little data to assess the species mass closure of WSOC in indoor air.^{2,29}

Insights into the composition and dynamics of WSOC_g have been provided by recent time-resolved measurements of *individual* compounds in real indoor environments. Proton transfer reaction-mass spectrometry (PTR-MS)^{8,30–38} and high-resolution time-of-flight chemical ionization mass spectrometry (HR-ToF-CIMS)^{6,29,31–33,39} have been used to identify formulas consistent with a variety of water soluble organic compounds indoors. For example, Tang *et al.* performed PTR-MS measurements in a university classroom and reported masses consistent with water-soluble aldehydes (*e.g.*, acetaldehyde and 4-oxo-pentanal), ketones (*e.g.*, acetone and hydroxyacetone), alcohols (*e.g.*, methanol and ethanol), organic acids (*e.g.*, formic acid and acetic acid) and precursors to WSOC_g (*i.e.*, terpenes).⁸ Furthermore, Tang *et al.* described significant source rates from human occupancy (57%), supply air (35%) and material surfaces (8%).⁸ Nonoccupant sources such as material off-gassing were important contributors to the top 20 most abundant compounds that Tang *et al.* measured which were mainly oxygenated gases such as organic acids (*e.g.*, acetic acid), ketones (*i.e.*, acetone), alcohols (*e.g.*, methanol), aldehydes (*e.g.*, acetaldehyde) and C8 aromatics;⁸ surface emissions likely included emissions from surface film reservoirs and not just the underlying materials.^{16,19,40} Liu *et al.* used HR-ToF-CIMS and reported formulas consistent with nearly 100 organic acids including monoacids, diacids, hydroxy acids, carbonyl acids and aromatic acids, with an average measured carboxylic acid mixing ratio of 6.8 ppbv in a university classroom.⁶ Note: we define mixing ratio as the ratio of the number density of

a gaseous compound to the dry air number density (approximately 2.43×10^{19} molecules per cm³ at standard ambient temperature and pressure) expressed here as parts-per-billion by volume (ppbv). Duncan *et al.* deployed a HR-ToF-CIMS in a single home in North Carolina (NC) and reported formulas consistent with 23 water soluble organic compounds: organic acids, glycols, parabens and methyl esters.²⁹ Duncan *et al.* quantified a small number of organic acids and determined the most abundant species were acetic ($30 \mu\text{g m}^{-3}$ to $125 \mu\text{g m}^{-3}$), formic ($15 \mu\text{g m}^{-3}$ to $53 \mu\text{g m}^{-3}$) and lactic acids ($2.5 \mu\text{g m}^{-3}$ to $13.5 \mu\text{g m}^{-3}$).²⁹ Using integrated mist chamber (MC) sampling and total organic carbon (TOC) analysis in 13 homes in the eastern U.S., Duncan *et al.* determined that those same acids comprised 30% to 54% of WSOC_g ($87 \mu\text{g-C m}^{-3}$ to $215 \mu\text{g-C m}^{-3}$) in those homes.² To our knowledge, that was the first time that total airborne WSOC species mass closure was attempted in homes. However, WSOC species mass closure has not been reported for real-time indoor sampling and a more comprehensive, real-time mass balance is warranted.

Real-time gas-surface dynamics of some common water soluble organic gases has been investigated in field studies in real indoor environments where they were shown to decay at rates faster than the concurrent air change rate (*i.e.*, defined herein as air change rate with outdoor air; ACR; h^{−1}) due to substantial surface removal.^{29,33} For example, surfaces in an art museum (wood, glass, wallboard, concrete, paint, *etc.*) were found to be significant sinks for emitted acetic and formic acids. Those surfaces became net sources when outdoor ventilation increased.³³ CIMS measurements captured rapid decays and rebounds of small water-soluble organic acids and alcohols measured at the University of Texas at Austin test house (UTest House) at the House Observations of Microbial and Environmental Chemistry (HOMEChem) experiment; this was attributed to substantial gas-surface partitioning dynamics.¹⁹ Duncan *et al.*, observed substantial loss rates coefficients (1.6 h^{-1} to 2.2 h^{-1}) for compounds with oxygen to carbon ratios greater than 0.75 using HR-ToF-CIMS real-time monitoring even though the home outdoor ventilation rate was only 0.4 h^{-1} (ACR).²⁹ Decay rates were elevated for 13 compounds (C1 to C7 organic acids and possible glycols and parabens) with air conditioning (AC) usage likely due to enhanced surface removal in cold, damp AC ducts and loss to AC condensate.²⁹ Since then, decay and rebounding of water soluble organic compounds with AC cycling has been observed in other indoor environments^{19,41} and in modeled results.⁴² Additionally, partitioning of water soluble organic gases to wet aerosols and subsequent reaction is important to the chemistry of outdoor SOA formation and similarly, reactive uptake in surface-associated water could be important indoors, where surface soiling is common and surface to volume (S/V) ratios are large, especially in damp homes.⁴³ Thus, time-resolved WSOC_g measurements may prove valuable to the assessment of partitioning and chemistry in indoor environments.

WSOC_p concentrations, behavior and composition in residential air are even less well understood than WSOC_g.⁴⁴ WSOC_p can be of outdoor origin (*i.e.*, through infiltration and outdoor ventilation)^{45–47} or indoor origin (*i.e.*, primary sources and

secondary formation through oxidation chemistry).^{17,48–55} Outdoors, WSOC_p can account for the majority of organic carbon (OC) in aerosols in non-pristine locations ranging from 40% to over 90% by mass on a carbon basis.⁴⁴ The composition of organic carbon in indoor aerosols is poorly characterized,⁵⁶ but some fraction is predicted to be water-soluble.^{44,57,58} Duarte and Duarte tabulated a list of the few studies on WSOC_p in different indoor scenarios including measurements of $0.3 \pm 0.02 \mu\text{g-C m}^{-3}$ in a church and $0.7 \mu\text{g-C m}^{-3}$ to $1.4 \mu\text{g-C m}^{-3}$ in retirement homes.⁴⁴

For these reasons, we investigated the dynamics of airborne total water-soluble organic carbon (WSOC_g and WSOC_p) at the unoccupied and unfurnished National Institute of Standards and Technology (NIST) Net-Zero Energy Residential Test Facility (NZERTF) during unperturbed periods and a variety of scripted activities as part of the Chemical Assessment of Surfaces and Air (CASA) experiment. This study provided a unique opportunity to measure total airborne WSOC concentrations and dynamics and assess the role of building materials and their associated reservoirs in a well-characterized building that lacked the typical complexity of occupied residences (*i.e.*, high human occupancy, regular activities, and furnishings). These measurements provide useful insights into WSOC concentration mass closure, distribution (gas-, particle-, and surface-phases), perturbation by activities and modulation by indoor surfaces.

Materials & methods

As part of the CASA experiment at the NZERTF, an on-line total carbon analyzer alternated every 6 min between a mist chamber (MC) and particle-into-liquid sampler (PILS) for semi-continuous, real-time monitoring (4 s resolution) of (total) WSOC_g and WSOC_p, respectively. Measurements were conducted from March 7th, 2022 to April 11th, 2022, during business days (Monday to Friday) and hours (8 am to 6 pm) inside the unfurnished, unoccupied NZERTF located on the NIST Campus in Gaithersburg, MD. Measurements were also obtained immediately outside the NZERTF on April 11th.

Net-zero energy residential test facility (NZERTF)

The NZERTF is 3-bedroom, 3-bathroom home with a basement (380 m³), main house (1st and 2nd floors; 624 m³)⁵⁹ and attic (99 m³), and S/V ratio of 2.7 (m⁻¹) associated with the superficial surface areas (1681 m²) of painted walls, painted ceilings, and wooden floors in the main house. The building was tightly constructed with low-VOC emitting building materials, particularly those with no formaldehyde additives (NAF) and with low acetic acid emissions.^{59,60} The NZERTF was constructed of NAF paints/coatings, adhesives/sealants, insulation binders, and plywood, and with materials tested using VOC emissions testing standards set by the California Department of Public Health and/or the California Air Resources Board (CARB).⁶¹ Furthermore, gypsum board was mechanically fastened without use of adhesives and only adhesives/sealants with limited quantities of acetic acid were used when needed.⁶¹ The building

ventilation system was operated at an outdoor ACR of 0.24 h^{-1} under normal conditions to meet American Society of Heating, Refrigerating, and Air-Conditioning Engineers (ASHRAE) Standard 62.2. A variable heat recovery ventilator (HRV) exchanged indoor air with outdoor air. A heating and air conditioning (HAC) system constantly circulated air among rooms of the main house and basement with an internal circulation rate of about 1.3 h^{-1} . Filters rated at a minimum efficiency reporting values (MERV) of 7 and 8 were used in the HRV and HAC systems, respectively. Air circulated between the attic and the house *via* passive vents. The house temperature and RH were generally 24 °C and 30% under normal unperturbed conditions.

Experimental design of CASA

The CASA experiment was designed to assess how air concentrations and surfaces responded to perturbations in conditions. There were 8 types of perturbation experiments conducted: cooking, cleaning, or injections, *i.e.*, pesticide products, a mixture of volatile organic compounds (VOCs) called “CASA cocktail,” wood smoke, ammonia (NH₃), carbon dioxide (CO₂), or ozone (O₃). Detailed descriptions of perturbations performed inside the NZERTF are summarized in ESI Section S1.† Specific days were dedicated to each experiment type where activities or injections were made repeatedly 2 to 4 times under normal outdoor ventilation (windows and doors closed). Some experiments were repeated under both low and high RH conditions on different days. Wood smoke injections occurred late in the campaign to avoid changing the house background before other experiments. Prior to, and in between, perturbation experiments, the house was operated in a standard mode to allow the building conditions to stabilize and allow researchers to measure the house background, which reflects emissions from the building itself.

Air sampling for WSOC_g and WSOC_p

Indoor air was sampled for WSOC inside the dining room on the 1st floor of the NZERTF (Section S2, Fig. S1†) using a MC, PILS, and TOC analyzer (Fig. 1) located on the enclosed porch connected to the test house (Fig. S1 and S2†). Sample lines were extended into the dining room of the house through drilled holes in the blocked door shared between the porch and dining room in the locations noted on Fig. S1.† The sampling inlets (Fig. S3†) were positioned 2 m above the floor and extended 2 m into the dining room. A 7.6 m long 0.9525 cm inner diameter (ID; 3/8 inch ID) polytetrafluoroethylene (PTFE) tube was used to collect 25 L min^{-1} of indoor air into the MC. A 12 m long 0.9525 cm outer diameter (OD; 3/8 inch OD) refrigeration grade copper tube was used to collect 13 L min^{-1} of indoor air into the PILS from the same location. Indoor air was drawn continuously by the MC and PILS throughout each experiment. Occasionally equivalent sample lines were extended outdoors to enable collection of residential outdoor WSOC concentrations.

Mist chamber (MC) – WSOC_g

WSOC_g was continuously scrubbed into a 1.2 mL min^{-1} flow of water by pulling 25 L min^{-1} of filtered indoor air through

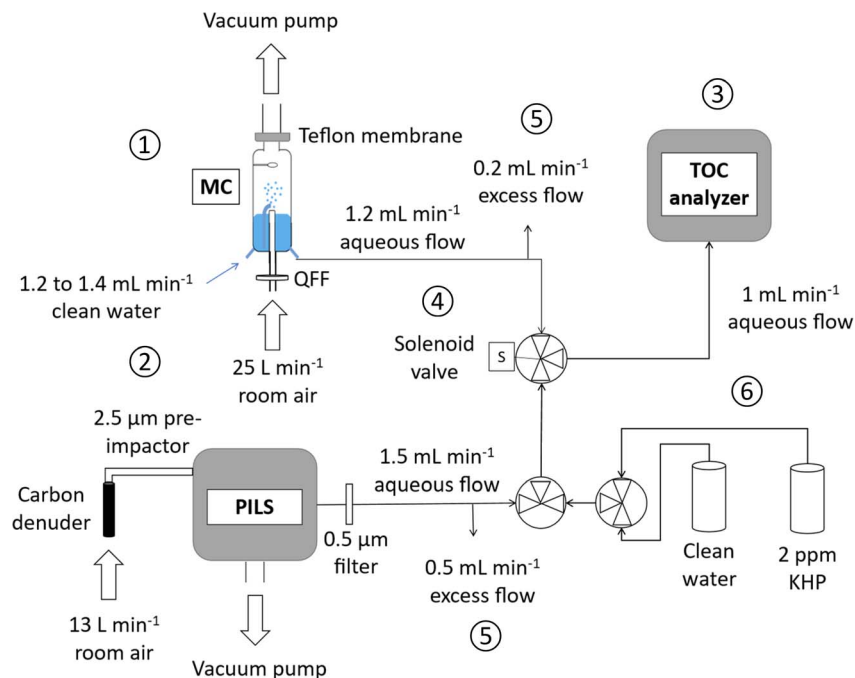


Fig. 1 MC/PILS-TOC system for sampling of total airborne WSOC. The MC (1) and PILS (2) continuously collect WSOC_g and WSOC_p, respectively, from room air into aqueous flows. These were analyzed in an on-line TOC analyzer (3) alternatively via a 6 min timed solenoid 3-way valve switch (4). Sample flow was sent to excess (5) when either instrument was not on-line with the TOC analyzer. Clean water or 2 ppm_m potassium hydrogen phthalate (KHP) standard was directed to the TOC via manual 3-way valves (6) for water flushes and calibrant checks as needed (QFF: quartz fiber filter).

a 190 mL modified mist chamber, also known as Cofer scrubber,^{57,58} equipped with a 10 mL water reservoir (Fig. 1). Water exiting the MC was analyzed semi-continuously for TOC, alternating at 6 min intervals with TOC analysis of the PILS sample flow. For some experiments, only the MC was sampled, and near-real-time results are reported here. MCs use a continuous, fine aqueous mist providing large air–water surface area to efficiently scrub water soluble gases from the sampled air into a refluxing water reservoir that can be analyzed.^{62,63} Field deployed MCs have been mostly operated as integrated samplers both outdoors^{26,64} and indoors,^{2,29} but at least one study used a system of two MCs in parallel for real-time sampling of WSOC outdoors.⁶⁵ For the present study, a single MC was modified and configured to operate as a real-time sampler to capture WSOC_g dynamics when coupled to an on-line TOC analyzer. To convert the MC into a real-time sampler, a new exit nozzle was added opposite from the MC inlet nozzle to accommodate a continuous 1.2 mL min^{-1} outlet flow (peristaltic pump-controlled) that was supplied to the TOC analyzer. The nozzle nearest the inner capillary tube where water is drawn into the mist was used as the inlet nozzle (Fig. S4†). The MC was initially loaded with 10 mL of deionized water daily and the meniscus level of bulk water inside the mist chamber was maintained by controlling the water flow rate at the MC inlet nozzle using a peristaltic pump set between 1.2 mL min^{-1} to 1.4 mL min^{-1} to replace water removed in exit flow and loss to evaporation.

The MC sampled particle-free indoor air by placing a quartz fiber filter (QFF; 47 mm) upstream; this filter provided a low-

pressure drop that does not interfere with creating a mist as described by Duncan *et al.* 2019.²⁹ Particle-free air entering the glass mist chamber encountered a refluxing water mist from the 10 mL reservoir of bulk water. A hydrophobic Teflon membrane (47 mm) at the air flow exit prevented loss of water droplets; which refluxed back into the water reservoir.⁶³ The MC was operated with water reservoir and mist droplet residence times of 8.3 min and 0.43 s, respectively; the droplet residence time refers to the time a sprayed droplet in the mist is suspended and in contact with sampled air.

Several quality control procedures were implemented (Section S3†). Briefly, dynamic blanks were collected by running the MC at the same water flow rate while setting the air flow rate to zero (*i.e.*, without generating mist). The WSOC_g limit of detection (LOD), estimated as three times the standard deviation (σ) of the dynamic blank, was $0.5 \mu\text{g-C m}^{-3}$. The collection efficiency of WSOC_g in MC samplers is largely controlled by Henry's law partitioning constants (K_H).^{2,63,66} The mean ($n = 10$) collection efficiency of the MC operated in continuous sampling mode was 57% for the WSOC_g mixture found in indoor air inside an empty university seminar room/teaching kitchen at UNC, determined by operating two real-time MCs in series. It was 63%, 75%, 74%, and 41% for glyoxal (K_H : $4\,000\,000 \text{ M atm}^{-1}$), formic acid (K_H : 8000 M atm^{-1}), acetic acid (K_H : 4000 M atm^{-1}), and acetone (K_H : 31 M atm^{-1}), respectively. The 57% collection efficiency, determined in the university teaching kitchen, was used as an effective collection efficiency to correct measurements from the NZERTF under the assumption that the relative composition of WSOC_g was similar in both locations.

Experimental set-up (Fig. S5A†), sample data (Fig. S5B†), collection efficiency calculation (eqn (S1)†) and tabulated collection efficiency results (Table S1†) are shown in Section S3.† Based on the TOC accuracy ($\pm 15\%$), air flow uncertainty ($\pm 5\%$) and relative standard deviation of dynamic blank concentrations ($\pm 1\%$), the uncertainty for WSOC_g measurements was estimated, by propagation of error, to be $\pm 16\%$.

Particle-into-liquid sampler (PILS) – WSOC_p

WSOC_p was continuously measured using a PILS and TOC analyzer at 6 min intervals, alternating with TOC analysis of MC sample flow. The PILS uses supersaturated water vapor to condense water on aerosol particles, increasing particle diameters and impaction collection efficiencies. The aerosol is impacted onto a liquid water flow that can be analyzed.^{67,68} PILS instruments have previously been coupled to on-line TOC analyzers to measure WSOC_p in outdoor environments, and the method is well-detailed in previous literature.^{69–71} The PILS sampled indoor air through a parallel-plate carbon strip denuder (Sunset Laboratory Inc.) to remove organic gases (breakthrough: 11%; Section S3†) and a 2.5 μm aerodynamic diameter cut point round jet impactor (Brechtel Manufacturing Inc.) that removed larger particles, so that the instrument collected WSOC_p in $\text{PM}_{2.5}$ into a 1.5 mL min^{-1} aqueous flow. The aqueous sample flow was filtered of insoluble particles using a 0.5 μm in-line liquid PEEK (polyetheretherketone) filter before introduction to the TOC analyzer.

Measures to describe the quality control of WSOC_p are summarized here and provided in more detail in Section S3.† For the dynamic blank, a filter was placed at the PILS sample air inlet to remove particles from the sampled air. The LOD was 2.0 $\mu\text{g-C m}^{-3}$ calculated as 3σ of the dynamic blank. These LOD values are higher than reported for outdoor WSOC_p measurements,⁶⁹ likely because more variable (and higher) indoor WSOC_g concentrations resulted in more variable (and higher) dynamic blank values. WSOC_p measurements were corrected for WSOC_g breakthrough. The PILS collection efficiency is reported to be 97% in the particle size range of 0.03 μm to 10 μm .⁶⁷ The on-line PILS-TOC method was previously shown to agree within 10% of traditional integrated filter methods;^{72,73} and have an estimated precision of 10%.⁶⁹

Total organic carbon analysis

The TOC was calibrated one week prior to the start of the study using commercial potassium hydrogen phthalate (KHP) calibration standards (range: 0.25 ppm_m to 5 ppm_m), a sucrose verification standard (1 mg-C L^{-1}) and verified calibration blanks (low TOC water); calibration checks were performed throughout the campaign. Note: we define ppm_m as the grams of compound per million grams of solution. The TOC was operated to perform real-time analysis with $\pm 15\%$ accuracy, 2% precision, 4 s analysis time, 3.25 min response and a dynamic analytical range of 0.20 ppb_m to 10 ppm_m . The TOC analyzer sampled the liquid flows at a flow rate of 1.1 mL min^{-1} through an approximately 1 m long, 0.0254 cm ID, 0.1588 cm OD (0.01 inch ID, 1/16 inch OD) PTFE tubing; excess flow from the MC or

PILS was sent to waste or collected in vials and stored frozen. Due to the TOC instrument response time (3.25 min) and the residence time in tubing (about 45 s), our system had an approximately 4 min system response time to changes in WSOC in indoor air.

Daily operation

Each sampling day began with either a 20 min calibrant check and/or 20 min clean water flush of the TOC analyzer before collection of dynamic blanks and sampling of indoor air. A 20 min calibrant check was performed 1 to 2 times a week using a 2.0 ppm_m KHP solution followed by a 20 min de-ionized water flush. The measured calibrant concentration was $1.915 \text{ ppm}_m \pm 0.076 \text{ ppm}_m$ over the duration of the study. MC and PILS dynamic blanks ($n = 3$) were then measured alternately in 6 min intervals for a total of 36 min for blank-correction of WSOC in indoor air. Next, WSOC_g and WSOC_p in indoor air were continuously collected, switching every 6 min between the MC and PILS, respectively, for the duration of the day. The last 2 min of data before each valve switch was used. Measurements of indoor air typically began at least 36 min prior to experimental perturbations in the house to obtain average ($n = 3$) unperturbed WSOC_g and WSOC_p concentrations (house background). On days with early perturbation start times, dynamic blank collection was skipped (2 to 3 times a week) to ensure enough time to collect indoor background measurements before perturbations began. A single day was selected for extended collection of the KHP standard, clean water, and dynamic blanks (Fig. S6†). In addition, airborne WSOC was measured indoors and then outdoors for 2 h each on a single day (April 11th) toward the end of the campaign (Section S4, Fig. S7†).

Calculations

The building-specific WSOC_g emission rate (E_f) without the introduction of purposeful sources was calculated using eqn (1):

$$E_f = \text{ACR} \times V \times (C_{\text{in}} - C_{\text{out}}) \quad (1)$$

where V is the volume of the main house, C_{in} is the mean WSOC_g house background concentration, and C_{out} is the outdoor WSOC_g concentration measured on April 11 (mean: $5 \pm 1 \text{ } \mu\text{g-C m}^{-3}$; see Fig. S7†).

Additional measurements

The following additional measurements were performed in the NZERTF and are described in detail in Section S5.† Gas-phase mixing ratios of C2 to C4 water-soluble aldehydes, ketones, and alcohols were measured using a Vocus proton-transfer-reaction (PTR-MS) mass spectrometer.⁷⁴ A fast-scanning gas chromatograph⁷⁵ was periodically interfaced to the inlet of the PTR-MS which provided chromatographic separation for each sampled ion and enabled correction for interferences. The gas-phase mixing ratios of C1 to C5 water-soluble carboxylic acids (*i.e.*, formic, acetic, propanoic, butanoic, and pentanoic acids) were measured using a Time-of-Flight Chemical Ionization

Mass Spectrometer (ToF-CIMS) equipped with iodide (I^-) reagent ions.^{76,77} Mixing ratio uncertainties for formic, acetic, propanoic, and pentanoic acids, ethanol and acetaldehyde are estimated to be within 20%; they are likely factor of two for acetone, butanone, and hydroxyacetone. Formaldehyde and formic acid mixing ratios were measured on the 1st floor and outside the NZERTF using a quantum cascade tunable infrared direct absorption spectrometer (QC-TILDAS; LOD of 0.1 ppb_v). O₃ mixing ratios were measured using an ozone analyzer (LOD: 0.5 ppb_v) on the 1st floor and outside the NZERTF.

Results

WSOC in house background

The vast majority of WSOC in the house background was in the gas phase (Section S6, Table S2†). On average, the house background WSOC_g concentration was 36 times higher than WSOC_p. In other words, only 3% of total airborne WSOC was in the particle phase (F_p). The mean WSOC_g and WSOC_p in the house background were $215 \mu\text{g-C m}^{-3} \pm 29 \mu\text{g-C m}^{-3}$ and $6 \mu\text{g-C m}^{-3} \pm 3 \mu\text{g-C m}^{-3}$, respectively. Using an average organic molecular weight to carbon weight for gaseous water soluble organic matter (WSOM_g) of 2.1, estimated as described below, and an organic molecular weight to carbon mass for particle-phase water-soluble organic matter (WSOM_p) of 2, previously reported for outdoor aerosol,^{44,78} the house background concentrations of WSOM_g and WSOM_p were $452 \mu\text{g m}^{-3}$ and $12 \mu\text{g m}^{-3}$, respectively. The 2.1 value is the average molecular weight-to-carbon mass ratio for the compounds identified in this study and for elemental formulas of species detected *via* positive mode high-resolution quadrupole time-of-flight mass spectrometry with an electrospray ion source (ESI-QTOF-MS) in indoor mist chamber sampling by Duncan *et al.*² (Table 2 in Duncan *et al.*).²

WSOC_g in the house background was substantial (Table S2†), suggesting the importance of contributions from house reservoirs such as building materials to indoor air WSOC concentrations. The indoor-outdoor ratio (I/O) of WSOC_g measured on April 11th was approximately 19 (outdoor WSOC_g concentration: $5 \mu\text{g-C m}^{-3} \pm 1 \mu\text{g-C m}^{-3}$; Fig. S7†); indoor WSOC_p was consistently equal to or greater than the outdoor concentration (WSOC_p I/O ≥ 1 using the LOD as an upper-bound estimate for outdoor WSOC_p concentrations below LOD). The reported I/O ratio could be underestimated because it was measured on a day when the house was more open than usual (*i.e.*, frequently opened or propped open front door) resulting in lower indoor air concentrations of airborne WSOC; indoor WSOC_p concentrations were near or below LOD (Fig. S7†). The fact that a dramatic reduction in the WSOC_g concentration was also observed with window opening, discussed later, also supports the conclusion that there is a substantial indoor source of WSOC_g.

The building-specific emission rate of WSOC_g was roughly 31 mg-C h^{-1} . Normalized to the area associated with painted walls, painted ceilings, and wooden floors in the main house, the surface area-specific E_f was approximately $18 \mu\text{g-C h}^{-1} \text{ m}^{-2}$. Average whole-house emission factors for various VOCs which

included water-soluble organic acids, alcohols, aldehydes, ketones and glycols were previously measured over a 15 month sampling period 1 year after the NZERTF was constructed (2013).⁵⁹ When summed, the average emission factors from January to July 2014 equated to an average WSOC_g emission rate of 148 mg-C h^{-1} (using only measured compounds with $K_H > 10 \text{ M atm}^{-1}$);⁵⁹ note this study did not directly measure total WSOC and so the sum of measured water-soluble species provides a lower bound for total WSOC_g emissions. Our E_f suggests at least a 79% reduction in (total) WSOC_g emissions in the 8 years that followed; more substantial reduction likely occurred in the first few years after construction.⁷⁹

The sum of measured water-soluble species accounted for 51% to 72% of the WSOC_g house background mass (Fig. 2; Table S3†). The remaining approximately 28% to 49% is expected to include organic compounds with K_H values from 10^1 M atm^{-1} to greater than 10^3 M atm^{-1} , which we verified we collected with efficiencies of 41% to greater than 75%. The most abundant species from each compound class were ethanol ($46 \mu\text{g-C m}^{-3}$ to $82 \mu\text{g-C m}^{-3}$), acetic acid ($12 \mu\text{g-C m}^{-3}$ to $24 \mu\text{g-C m}^{-3}$), acetone ($11 \mu\text{g-C m}^{-3}$ to $16 \mu\text{g-C m}^{-3}$), and acetaldehyde ($7 \mu\text{g-C m}^{-3}$ to $9 \mu\text{g-C m}^{-3}$), which have been measured in a variety of indoor air studies and have known building- and human-derived sources of acetic acid,^{2,6,29,32,33,80,81} acetone,^{2,30,32,33,80–84} acetaldehyde.^{23,30,33,81–87} Notably, they could be emitted from building materials such as wood-based products,^{86–89} paints,^{33,90,91} and finishes (*e.g.*, adhesives).^{59,92,93} Oxidation chemistry is also a potential source.^{34,94–96}

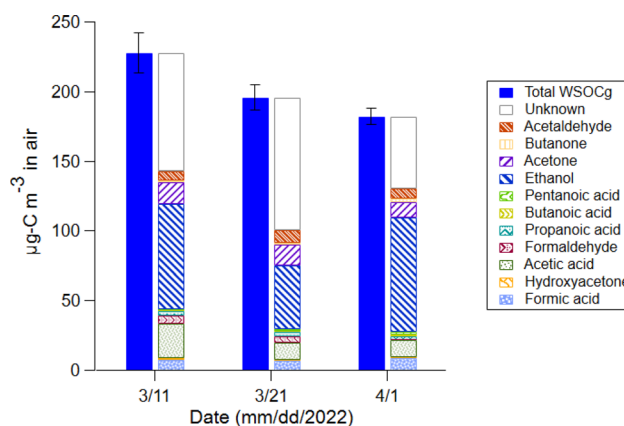


Fig. 2 WSOC_g (blue bar) measured with the MC with TOC analyzer and reconstructed WSOC_g (multicolored stack) which was taken to be the sum of carbon in measured individual water-soluble organic gases ($K_H > 13 \text{ M atm}^{-1}$) and the remaining mass of carbon (difference between WSOC_g and sum of species; white) from unperturbed periods (*i.e.*, the house background) on March 11th, March 21st, and April 1st (see Tables S2 and S3†). Compounds are stacked from most (bottom) to least (top) water soluble. The concentration of total WSOC_g was corrected for the effective WSOC_g collection efficiency (57%). Error bars represent $\pm 1\sigma$. "House background periods" refer to time periods immediately before conditions were perturbed by scripted activities (typically 1 h to 2 h before the perturbation), 12 h to 16 h after the last perturbation, and sometimes shortly (1 h to 2 h) after perturbations that did not induce an observable change in WSOC_g.

WSOC during perturbations

Fig. 3, 4 and Table 1 show the effect of several perturbations on $WSOC_g$ and $WSOC_p$ in the house (NZERTF). Some perturbations increased (e.g., cooking with vinegar, wood smoke additions,

CASA cocktail addition, ozone injection, and closing windows) or decreased (e.g., surface cleaning and window opening) $WSOC$, but others did not result in a measurable change in airborne $WSOC$ (e.g., pesticide product additions, inorganic

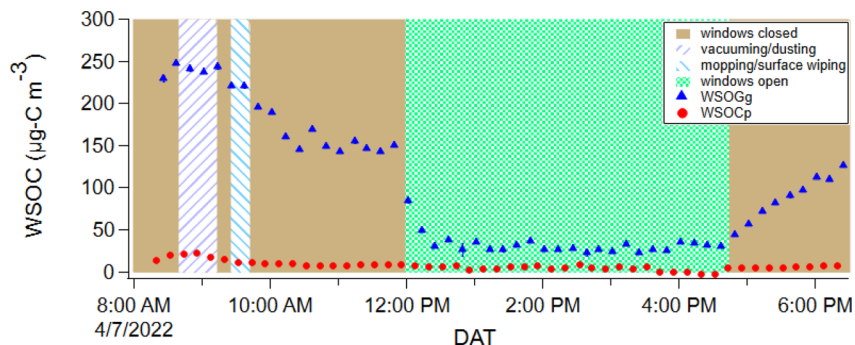


Fig. 3 Indoor $WSOC_g$ (blue triangles) and $WSOC_p$ concentrations (red dots) during surface cleaning of the wood smoke contaminated house (cleaning times indicated by purple and blue-bands), window opening (green band) and subsequent window closing (brown band). Error bars represent measurement uncertainties ($\pm 1 \sigma$) and are mostly smaller than the data point markers (DAT = date and time).

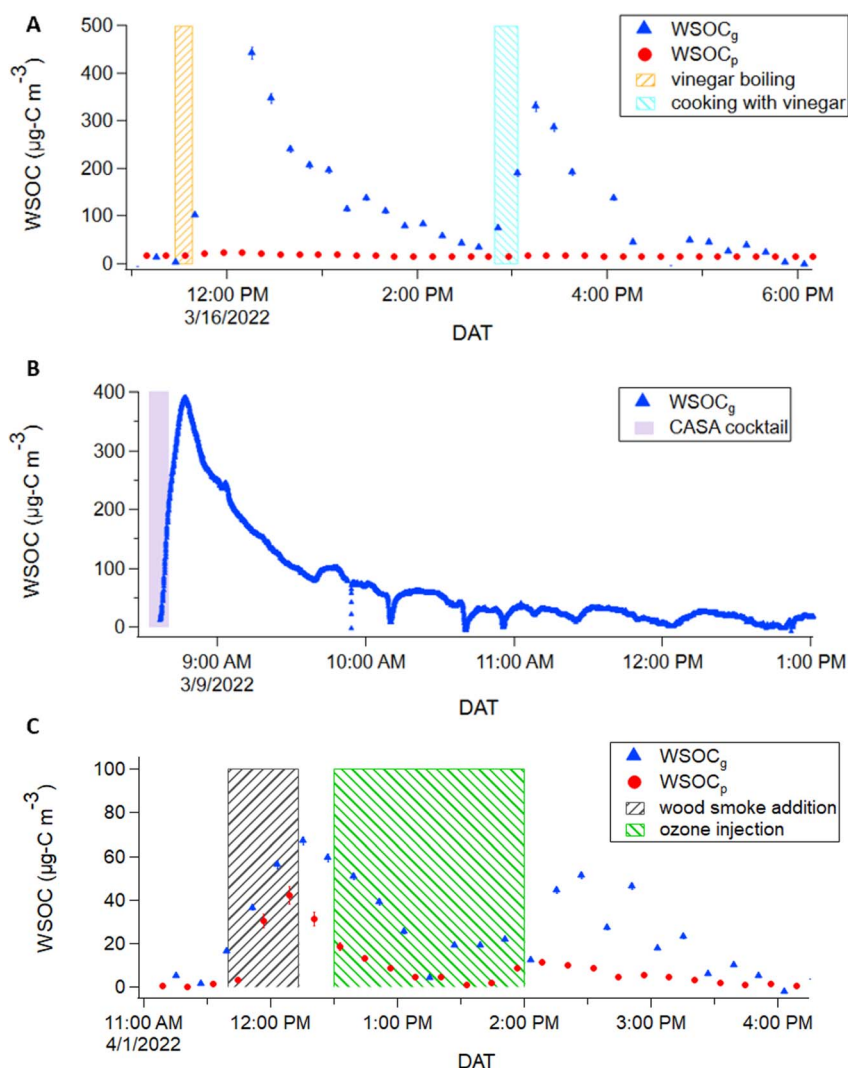


Fig. 4 Indoor $WSOC_g$ (blue triangles) and $WSOC_p$ concentrations (red dots) in indoor air during selected perturbations.

Table 1 Change in WSOC_g and WSOC_p peak heights (Δ WSOC) during perturbations compared to average unperturbed house background concentration^a

Perturbation	Duration (min)	Amount added	Δ WSOC based on peak ($\mu\text{g-C m}^{-3}$)		House background ($\mu\text{g-C m}^{-3}$)	
			WSOC _g	WSOC _p	WSOC _g	WSOC _p
Surface cleaning	45	—	(–) 75	9	215 $\pm$ 29	6 $\pm$ 3
Air cleaning	60	—	No change	No change		
Window opening	315	—	(–) 119	(–) 4		
Window closing	—	—	98	3		
Cooking without vinegar	40	Varied by meal ^b	No change	No change		
Cooking with vinegar	10–15	30 mL	329–444	No change		
CASA cocktail	8	0.5 mL of each chemical	100–390	20		
Wood smoke addition	2	0.5 g woodchips burned	25–220 ^c	16–100		
Pesticide addition	0.5	425–502 g sprayed	No change	No change		
Ozone injection	90	70 ppb peak levels	25–80	7–12		
Acid/base addition	30	500 g min ^{–1} CO ₂ 40 mg min ^{–1} NH ₃	No change	No change		

^a The “(–)” denotes a decrease in WSOC, positive values denote an observed maximum peak above the house background, and “no change” denotes perturbations that did not induce a measurable change in WSOC in our system. ^b See ESI Section S1. ^c The 220 $\mu\text{g-C m}^{-3}$ peak was from a smoke injection on the morning of March 29th that was much higher than any other observed peak for this perturbation type despite similar mass of woodchips burned. The next highest peak concentration was 93 $\mu\text{g-C m}^{-3}$.

acid/base additions, portable air cleaner operation and selected cooking activities that did not include vinegar).

Surface cleaning inside the wood smoke contaminated NZERTF on April 7th caused a reduction in WSOC_g concentrations (Fig. 3). First, a small peak of WSOC_p (8 $\mu\text{g-C m}^{-3}$) was observed as researchers were in the house preparing to clean, presumably because of resuspension of particles from researchers' movement.⁹⁷ Then, following vacuuming, dusting, mopping and surface wiping, WSOC_g concentrations decayed at a rate of 0.42 h^{–1}, decreasing by 39% after surfaces were cleaned, which may be due to removal of WSOC reservoirs on wood smoke-contaminated surfaces. Li *et al.* observed concentration reductions of individual water-soluble organic gases, specifically formic acid and formaldehyde, during the same time period inside the NZERTF and previously drew the same conclusion.⁷⁷ Air cleaners used on a separate day after smoke injections (March 24) did not result in an observable change in WSOC (see Table 1, Section S7 and Fig. S8†).

The “Window Opening and Closing” perturbation (Fig. 3) provided evidence consistent with emissions and air concentrations being influenced by surface reservoirs. WSOC_g concentrations were diluted upon window opening and reduced by 81% from 147 $\mu\text{g-C m}^{-3}$ to 28 $\mu\text{g-C m}^{-3}$ before rebounding after windows were closed, increasing from 28 $\mu\text{g-C m}^{-3}$ to 125 $\mu\text{g-C m}^{-3}$. The dramatic decrease with window opening and subsequent rebound provided additional evidence that WSOC in building materials/reservoirs are responsible for the high house background. WSOC_g decayed at a rate of 2.8 h^{–1} post window opening due to dilution with outdoor air and an increased air change rate (outdoor ACR = 21.2 h^{–1}; see Section S8, Fig. S9 and S10†). The decrease in air concentrations also likely increased emissions from surfaces reservoirs, causing WSOC_g concentrations to decay slower than they would if the emission rates were constant. The estimated E_f of WSOC_g

during the steady-state period of window opening was 301 mg-C h^{–1}, which is much greater than the steady-state E_f associated with the typical house background (31 mg-C h^{–1}) and before windows were opened (21 mg-C h^{–1}). Approximately 90 minutes after closing windows, the WSOC_g concentration had nearly returned to pre-perturbation levels. The characteristic rate for WSOC_g rebound from window closing was 0.8 h^{–1}, which is faster than the ventilation rate (outdoor ACR). Gas exchange with surface reservoirs is likely responsible for this rapid rise in the indoor air concentration of WSOC_g.^{16,19,40} Effects on WSOC_p concentrations were too small to measure but concentrations appear to slightly decrease upon window opening and increase after window closing. It is important to note that the measurement of such a high outdoor ventilation rate in the fully open house (*i.e.*, during the window opening perturbation; Section S1†) is expected to be somewhat uncertain due to increased variations in interzonal mixing, but even an error of 50% would still produce a substantially increased E_f during the period of window opening.

The change in WSOC from cooking activities depended on the food items cooked. Cooking bacon, peppers, and tater tots (pan fry or air fry) did not result in a measurable response in either gas or particle phase WSOC, likely due to low water-solubility of cooking emissions (*e.g.*, cooking oils and long-chain fatty acids). However, cooking with vinegar did produce a large WSOC_g response due to acetic acid emissions. The first peak (approximately 450 $\mu\text{g-C m}^{-3}$) in Fig. 4a shows a large WSOC_g peak from boiling 2 tablespoons of balsamic vinegar to produce a balsamic glaze. The second peak (approximately 350 $\mu\text{g-C m}^{-3}$) shows the signal from cooking a similar glaze with 2 tablespoons of balsamic vinegar, but with vegetables included in the dish. The smaller peak is likely due to the other food ingredients absorbing some of the acetic acid leading to less vinegar evaporating from the glaze. Although an increase in

indoor $\text{PM}_{2.5}$ was observed during the second cooking event (with vegetables; up to $10 \mu\text{g m}^{-3}$ to $15 \mu\text{g m}^{-3}$), WSOC_p did not increase.

The addition of the CASA cocktail (Fig. 4b) and wood smoke (Fig. 4c) significantly increased indoor WSOC concentrations over background levels, but pesticide addition did not result in a measurable change in WSOC (Table 1). The pesticide product contained water insoluble or sparingly soluble chemicals that were not expected to be captured in our system (see ESI Section S1†). Injection of the CASA cocktail, which contained many compounds that are relatively water soluble, resulted in a $100 \mu\text{g-C m}^{-3}$ to $390 \mu\text{g-C m}^{-3}$ increase in WSOC_g (Table 1; Fig. 4b). Wood smoke injections increased WSOC_g and WSOC_p by $25 \mu\text{g-C m}^{-3}$ to $220 \mu\text{g-C m}^{-3}$ and $16 \mu\text{g-C m}^{-3}$ to $100 \mu\text{g-C m}^{-3}$, respectively (Table 1; Fig. 4c). WSOC_p concentrations were highest for wood smoke, with gas to particle concentration ratios of approximately 1 to 2. The WSOC system was run for only 1 of the 2 high RH (approximately 74% RH) smoke injections, precluding investigation of an RH effect on airborne WSOC from wood smoke injections. Li *et al.* reported the bulk organic chemical composition of the CASA campaign wood smoke aerosols was 95% organic, 3% nitrate, 1% sulfate, and 1% ammonium, and that gaseous wood smoke emissions included formic acid, acetic acid, and formaldehyde.⁷⁷ Based on Li *et al.* organic aerosol data⁷⁷ and an OM:OC conversion factor of 2.0 derived from fireplace wood combustion,^{78,98} about 32% to 67% of wood smoke organic aerosol mass (OM) is water soluble.

The addition of inorganic species (*i.e.*, ozone, NH_3 , or CO_2 ; Section S9†) had moderate to no impact on WSOC levels measured by our system. Although the MC chamber is well-suited for collecting oxidation products,⁶³ ozone additions did not consistently produce a measurable increase in WSOC_p or WSOC_g . A change in WSOC in our system was observed for 3 out of the 9 ozone additions that took place during times of WSOC sampling, perhaps due to the high loading of WSOC_g in the house air to begin with (house background). In some cases spikes in WSOC were observed when ozone was added shortly after wood smoke injections (Table 1, Fig. 4c); the mixing ratio of O_3 in the main house was typically about 10 ppb_v when not intentionally perturbed. Additions of basic NH_3 and acidic CO_2 did not induce a measurable change in WSOC, but measurements of gas-phase total inorganic carbon (TIC) in the system became more sensitive to indoor CO_2 after NH_3 injection, presumably due to pH elevation of MC water as discussed in Section S9 (see Fig. S11†).

WSOC_g peaks decayed at rates ranging from 0.4 h^{-1} to 4.0 h^{-1} across the perturbations (Table 2). These rates are 1.7 to 16.7 times faster than the ventilation rate, suggesting the presence of surface reservoirs that amplified the loss rates of WSOC_g in air. When surface reservoirs and gas-phase concentrations approach steady-state, outdoor ventilation and exfiltration become the dominant removal mechanisms. Decay rates were influenced by internal mixing (1.3 h^{-1}), surface removal rates, and ventilation (0.25 h^{-1}), slowing with time. For example, decay of the CASA cocktail injection was rapid to begin with (*initial* decay rate: $2.6 \text{ h}^{-1} \pm 0.02 \text{ h}^{-1}$), perhaps associated

Table 2 Decay rate for each perturbation that resulted in a WSOC_g peak^a

Perturbation	Peak decay rates (h^{-1})	House parameters (h^{-1})	
	WSOC_g	Internal mixing	Outdoor ACR
Wood smoke	4.0–0.4	1.3	0.24
Window opening	2.8		
CASA cocktail	2.6–1.0		
Cooking with vinegar	2.0–1.0		
Ozone	1.6		
Window closing ^b	0.80–0.20		
Surface cleaning	0.42		

^a The house background WSOC signal was subtracted to isolate the contribution from perturbations. (A) Cooking with vinegar, (B) CASA cocktail (VOC mixture) injection; only WSOC_g was measured to focus on gas-phase and (C) wood smoke followed by ozone addition. Error bars represent measurement uncertainties ($\pm 1 \sigma$) and are mostly smaller than the data point markers. ^b Rebound rate during WSOC_g house background recovery.

with some mixing with the rest of the house, but also uptake by surfaces (see Fig. S12†). The initial peak decay rate held for about 15 minutes before the decay rate reduced (*later* decay rate: $1.7 \pm 0.02 \text{ h}^{-1}$), indicating that surface reservoirs were beginning to shift away from being a net sink, but had not yet become a net source (see Fig. S12†). Long-term decay (*i.e.*, longer than 2 h to 3 h) of WSOC when concentrations were just elevated over

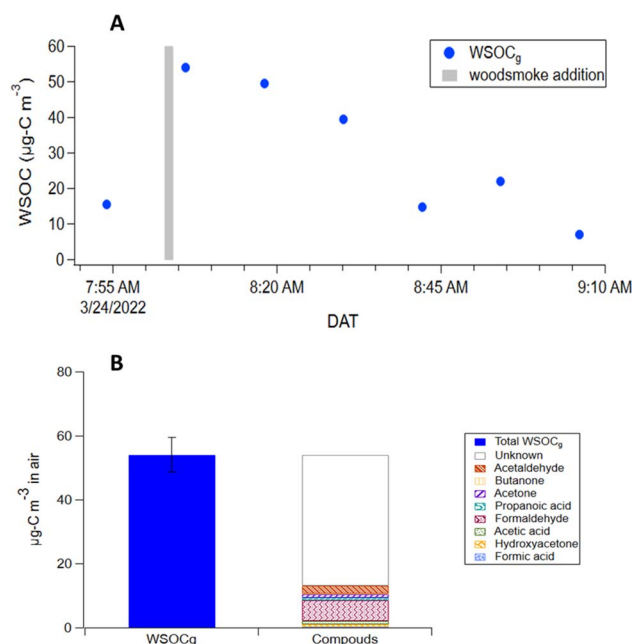


Fig. 5 (A) WSOC_g concentrations in indoor air during wood smoke injection. (B) Mass closure of peak WSOC_g concentration observed during wood smoke injection. Specifically, panel (B) includes WSOC_g (blue bar) measured with the MC with TOC analyzer and reconstructed WSOC_g (multicolored stack) which was taken to be the sum of carbon in measured individual water-soluble organic gases ($K_H > 13 \text{ M atm}^{-1}$) and the remaining mass of carbon (difference between WSOC_g and sum of species; white).

normal conditions and associated variability were not captured in our system. After internal mixing is complete and there is neither net surface uptake nor removal, decay rates reduce to the ventilation rate, consistent with tracer gas (SF_6) decay.

Wood smoke injections introduced water soluble organic gases in indoor air that were not detected in the house background. Based on the mass closure on peak WSOC_g concentrations ($54.1 \mu\text{g-C m}^{-3}$) during wood smoke injections, 75% (by mass) of WSOC_g in the wood smoke was uncharacterized and 25% was composed of gases also detected in the house background (Fig. 5). The most abundant of the characterized fractions was formaldehyde ($6.5 \mu\text{g-C m}^{-3}$) and acetaldehyde ($2.6 \mu\text{g-C m}^{-3}$), which comprised 17% of peak WSOC_g during the perturbation.

Discussion

The concentrations and dynamics of WSOC_g and WSOC_p were semi-continuously monitored in real-time. Airborne WSOC in the test house background mostly resided in the gas phase and not particle phase (see Table S2†), which was predicted by Weschler and Nazaroff to be true for individual compounds that partition strongly to surfaces.¹⁶ The small fraction in the particle phase ($F_p = 0.03$) also suggests that indoor particles in the test house would have a small impact on WSOC_g dynamics,¹⁶ likely because airborne particles are removed *via* outdoor ventilation and contribute negligibly to total surface area and volume compared to that of the building structure.^{16,99} WSOC_p ranged from $3 \mu\text{g-C m}^{-3}$ to $10 \mu\text{g-C m}^{-3}$ in the house background. This is comparable in magnitude to the average $7 \mu\text{g-C m}^{-3}$ of organic carbon (OC) in $\text{PM}_{2.5}$ that Polidori *et al.* measured in 173 occupied homes across CA, TX, and NJ as part of the Relationship of Indoor, Outdoor and Personal Air (RIOPA) study;⁵⁵ although the Polidori *et al.* measurements would have included water insoluble organic compounds such as oily fatty acids from cooking emissions that would not have been captured in our system.

The house background of WSOC_g ($215 \mu\text{g-C m}^{-3}$ on average) compares to WSOC_g concentrations reported for some occupied homes.² Duncan *et al.* performed extensive measurements of WSOC_g concentrations in 13 homes in the eastern U.S., reporting a range of $87 \mu\text{g-C m}^{-3}$ to $215 \mu\text{g-C m}^{-3}$ and I/O ratios of 10 to 20.² The Duncan *et al.* study homes were regularly occupied, cooked in, cleaned, more loosely constructed and likely built of more conventional materials than the NZERTF.² The CASA WSOC_g concentrations (this study) were on the upper end of the Duncan *et al.* range, which may be due, at least in part, to the NZERTF being more air tight (0.24 h^{-1}) than homes investigated by Duncan *et al.* (0.3 to 0.8 h^{-1}). Although buildings constructed of low-VOC emitting materials tend to maintain lower VOC concentrations than typical buildings,^{85,100–102} outdoor air change rates are known to have the strongest influence on indoor VOC levels.^{85,103,104} The CASA WSOC_g house background was also similar to the sum of measured reactive organic carbon species during the HOMEChem study when that home was unoccupied (median indoor = $223 \mu\text{g-C m}^{-3}$; median outdoor = $54 \mu\text{g-C m}^{-3}$).¹⁰⁵ However, it must be noted that the

HOMEChem measurements included some compounds with low water solubility and did not include all water soluble species.

The high WSOC_g concentration from house background emissions is also consistent with observations of individual water soluble organic gases that contributed to “high baselines” of VOCs in other houses.^{81,89} For example, Arata *et al.* and Liu *et al.* determined individual water soluble organic gases (*i.e.*, acetic acid, formic acid and methanol) dominated house background emissions inside the unoccupied UTest House⁸¹ and an occupied California home.⁸⁹ Methanol was not directly measured during our study because of the insensitivity of iodide-HR-ToF-CIMS to mono-alcohols⁷⁶ and the poor transmission of low mass-to-charge ions through the ion quadrupole guide of the Vocus PTR-MS.⁷⁴

Building-associated emissions (*e.g.*, structural materials, finishing materials, and paint) and/or surface reservoirs, maintained a high house background concentration of WSOC_g which was dominated by ethanol (23% to 45% of WSOC_g). Ethanol mixing ratios in indoor air are typically associated with human-activities (*e.g.*, spikes in ethanol during cooking, cleaning, and consuming alcohol) and not continuous material emissions,^{81,89} so the consistently high background of ethanol inside the unoccupied NZERTF is unusual. In a previous indoor air quality study on the NZERTF, ethanol was identified as the dominant VOC and had an average emission factor of 64 mg-C h^{-1} that remained constant over a 15 month sampling period (June 2013 through July 2014). During the previous study, Poppendieck *et al.* observed large spikes in ethanol mixing ratios during maintenance of the ground source heat pump (GSHP).¹⁰² The potential for contaminated building materials to be a long-term source of ethanol warrants further investigation.

Typically, small organic acids have been the most abundant WSOC_g in measured indoor air.^{2,29,94} For example, Duncan *et al.* determined formic, acetic and lactic acids made up 30% to 54% of WSOC_g across studied (occupied) homes and the only alcohols identified were butanol and 2-butoxyethanol.² In the current study, measured organic acids accounted for 12% to 16% of WSOC_g . The $7 \mu\text{g-C m}^{-3}$ to $9 \mu\text{g-C m}^{-3}$ of formic acid and $12 \mu\text{g-C m}^{-3}$ to $24 \mu\text{g-C m}^{-3}$ of acetic acid reported here for the NZERTF fall into reported ranges for the acids in indoor air,^{2,5,6,29,94,106} but the acetic acid concentrations were about 3 to 4 times smaller than the average concentrations reported by Duncan *et al.* ($34 \mu\text{g-C m}^{-3}$ to $87 \mu\text{g-C m}^{-3}$).² This, in part, may be related to materials selection in NZERTF construction. Both formaldehyde and acetic acid were reported to have lower concentrations in the NZERTF than other new/existing homes.^{59,107} In addition, occupancy, regular human activities, and consumer product usage likely played a role in higher acetic acid concentrations reported in the Duncan *et al.* study.²

A total of 28% to 49% of WSOC_g in the NZERTF house background and 75% of WSOC_g during contrasting chemical wood smoke infiltration was unidentified at the molecular level. The unidentified house background mass may include: lactic acid, propylene glycol, ethyl acetate, butanol, 2-butoxyethanol, hexanone, 4-oxopentanal, various amines, amides, and organophosphates, which have been identified (based on elemental

formula) in previous mist-chamber studies in residences,²⁹ phenol and benzaldehyde,⁵⁹ which were measured previously in the NZERTF,⁵⁹ and propanol and methanol, quantified in other indoor studies.^{30,80} The unidentified mass during wood smoke infiltration may include glycolic acid, phenols, furanones, furfurals, oxobutanoic acid, and hydroxybenzaldehyde.^{14,108–111}

Water is expected to be an important component of shallow and deep surface reservoirs and has been demonstrated to facilitate surface uptake of water-soluble organic gases.²⁰ It is worth noting that the aqueous concentration of WSOC_g scrubbed into the MC bulk water under unperturbed conditions was 4.4 mg-C L⁻¹, on average (correcting for collection efficiency). The WSOC (aq) concentrations in MC water could provide an order-of-magnitude estimate of WSOC concentrations in NZERTF aqueous surface films from the partitioning of WSOC_g in the test house, particularly in deep reservoirs that are hard to access and measure; however, aqueous concentrations of ionizable species (*i.e.*, acids/bases) may depend on pH of surface-associated water.

Implications

Decay rates that were larger than the ventilation rate, followed by concentration rebound when windows and doors were closed after periods of elevated outdoor ventilation, demonstrate the substantial role that surface reservoirs play in WSOC_g concentration dynamics (*i.e.*, the building surfaces themselves, irrespective of house activities). Previous studies point to differences in the behavior of VOCs compared to semi-volatile organic compounds (SVOCs) in response to window opening and closing; comparison with these past studies aids our understanding of the behavior of WSOC_g. Although adequate ventilation reduces airborne concentrations, it does not reduce or generally influence building material emission rates of nonpolar VOCs because slow bulk diffusion within deep reservoirs in the material limits the emission rate.^{21–23} SVOC emission rates are typically enhanced when the concentration gradient between the material and air increases (*i.e.*, by increasing ventilation), suggesting emission is not limited by diffusion in the material (*i.e.*, SVOCs are readily available from near surface reservoirs).^{24,25} This is because SVOCs are more uniformly distributed in their host materials and evaporate more slowly at the surface than VOCs.^{24,25} We observed that WSOC_g emissions are enhanced at elevated ventilation, similar to the emission behavior of SVOCs and individual polar VOCs (*e.g.*, formaldehyde).^{22,85,104} This behavior suggests that WSOC (or oxygenated VOCs) are readily available from near surface reservoirs and emissions are not limited by diffusion in the materials. The indoor WSOC_g concentration reduced by 81% from 147 µg-C m⁻³ to 28 µg-C m⁻³ during the period of elevated outdoor ventilation associated with open windows which was 5.6 times higher than outdoors (*I/O* = 5.6). If emission rates were instead constant and independent of the ventilation rate, the *I/O* ratio would have approached 1.0 and the WSOC_g concentration would have been reduced by 95% to 7 µg-C m⁻³.

Limitations

Although this study was performed inside a test house, data was discussed in the context of related indoor measurements from literature that were often performed in real, occupied homes. The homes across various indoor air studies are not identical buildings and will have considerable variations in conditions (*e.g.*, ventilation rate, dimensions, contents, materials, outdoor air composition, and human activities). Major differences between the NZERTF and typical homes include a lack of substantial occupancy, furniture, exposure history (*e.g.*, years of cooking emissions), the use of low-VOC emitting building materials in construction, and building envelope air leakage rates at the lower end of the range for typical homes. Therefore, comparisons to the measurements taken inside the NZERTF test house are used to provide insights into how the building (*i.e.*, walls, flooring, and ceilings) in the absence of typical features of occupied residences may contribute to and control airborne WSOC, and how perturbations induce spikes of WSOC above the house background. The high house background prevented measurements of relatively small changes in WSOC_g. Also, it should be noted that a QFF was used to remove particles upstream of the MC sampler. Organic gases can adsorb to QFFs, reducing measured WSOC_g concentrations until gases come into equilibrium between the QFF and the gas phase. Thus, reported WSOC_g concentrations should be considered lower bound values.

Disclaimer

Any equipment, instruments, software, or materials mentioned in this paper are solely for the purpose of accurately describing the experimental procedure. Mentioning these items does not constitute an endorsement or recommendation by NIST, nor does it imply that they are the best available options. The policy of NIST is to use the International System of Units in all publications. In the document, however, some units are presented in the system prevalent in the relevant discipline.

Data availability

All data are available on request.

Author contributions

MW: writing – original draft, review and editing; conceptualization, methodology, investigation, formal analysis, data curation, visualization. GCM: writing – review and editing, conceptualization, methodology, investigation, supervision. KB: writing – review and editing, methodology, supervision. JL: writing – review and editing, investigation, formal analysis, data curation. JCD: writing – review and editing, investigation, formal analysis. HNH: writing – review and editing, investigation, formal analysis, data curation. KM: writing – review and editing, investigation, formal analysis. LM: writing – review and editing, investigation, formal analysis. MEV: writing – review and editing, conceptualization, methodology, supervision,

funding acquisition, project administration. DKF: writing – review and editing, conceptualization, methodology, supervision, funding acquisition, project administration. JA: writing – review and editing, conceptualization, methodology, supervision, funding acquisition. DP: writing – review and editing, conceptualization, methodology, investigation, resources, project administration. BJT: writing – review and editing, conceptualization, methodology, validation, supervision, funding acquisition.

Conflicts of interest

There are no conflicts of interest to declare.

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References

- 1 N. E. Klepeis, W. C. Nelson, W. R. Ott, J. P. Robinson, A. M. Tsang, P. Switzer, J. V. Behar, S. C. Hern and W. H. Engelmann, The National Human Activity Pattern Survey (NHAPS): a resource for assessing exposure to environmental pollutants, *J. Exposure Anal. Environ. Epidemiol.*, 2001, **11**, 231–252.
- 2 S. M. Duncan, K. Sexton, L. Collins and B. J. Turpin, Residential water-soluble organic gases: chemical characterization of a substantial contributor to indoor exposures, *Environ. Sci.: Processes Impacts*, 2019, **21**, 1364–1373.
- 3 N. Yamamoto, D. G. Shendell, A. M. Winer and J. Zhang, Residential air exchange rates in three major US metropolitan areas: results from the Relationship Among Indoor, Outdoor, and Personal Air Study 1999–2001, *Indoor Air*, 2010, **20**, 85–90.
- 4 Z. Liu, T. Rios-Carvajal, M. Ceccato and T. Hassenkam, Nanoscale chemical mapping of oxygen functional groups on graphene oxide using atomic force microscopy-coupled infrared spectroscopy, *J. Colloid Interface Sci.*, 2019, **556**, 458–465.
- 5 J. Zhang, P. J. Lioy and Q. He, Characteristics of aldehydes: concentrations, sources, and exposures for indoor and outdoor residential microenvironments, *Environ. Sci. Technol.*, 1994, **28**, 146–152.
- 6 S. Liu, S. L. Thompson, H. Stark, P. J. Ziemann and J. L. Jimenez, Gas-Phase Carboxylic Acids in a University Classroom: Abundance, Variability, and Sources, *Environ. Sci. Technol.*, 2017, **51**, 5454–5463.
- 7 R. Reiss, P. B. Ryan, P. Koutrakis and S. J. Tibbetts, Ozone reactive chemistry on interior latex paint, *Environ. Sci. Technol.*, 1995, **29**, 1906–1912.
- 8 X. Tang, P. K. Misztal, W. W. Nazaroff and A. H. Goldstein, Volatile Organic Compound Emissions from Humans Indoors, *Environ. Sci. Technol.*, 2016, **50**, 12686–12694.
- 9 W. W. Nazaroff and C. J. Weschler, Indoor ozone: Concentrations and influencing factors, *Indoor Air*, 2022, **32**, e12942.
- 10 M. Kruza, A. C. Lewis, G. C. Morrison and N. Carslaw, Impact of surface ozone interactions on indoor air chemistry: A modeling study, *Indoor Air*, 2017, **27**, 1001–1011.
- 11 C. J. Weschler, Roles of the human occupant in indoor chemistry, *Indoor Air*, 2016, **26**, 6–24.
- 12 H. R. Katragadda, A. Fullana, S. Sidhu and Á. A. Carbonell-Barrachina, Emissions of volatile aldehydes from heated cooking oils, *Food Chem.*, 2010, **120**, 59–65.
- 13 W. W. Nazaroff and C. J. Weschler, Cleaning products and air fresheners: exposure to primary and secondary air pollutants, *Atmos. Environ.*, 2004, **38**, 2841–2865.
- 14 J. J. Schauer, M. J. Kleeman, G. R. Cass and B. R. Simoneit, Measurement of emissions from air pollution sources. 3. C1–C29 organic compounds from fireplace combustion of wood, *Environ. Sci. Technol.*, 2001, **35**, 1716–1728.
- 15 S. M. Duncan, K. G. Sexton and B. J. Turpin, Oxygenated VOCs, aqueous chemistry, and potential impacts on residential indoor air composition, *Indoor Air*, 2018, **28**, 198–212.
- 16 C. J. Weschler and W. W. Nazaroff, Semivolatile organic compounds in indoor environments, *Atmos. Environ.*, 2008, **42**, 9018–9040.
- 17 N. Carslaw, T. Mota, M. E. Jenkin, M. H. Barley and G. McFiggans, A significant role for nitrate and peroxide groups on indoor secondary organic aerosol, *Environ. Sci. Technol.*, 2012, **46**, 9290–9298.
- 18 C. Liu, G. C. Morrison and Y. Zhang, Role of aerosols in enhancing SVOC flux between air and indoor surfaces and its influence on exposure, *Atmos. Environ.*, 2012, **55**, 347–356.
- 19 C. Wang, D. B. Collins, C. Arata, A. H. Goldstein, J. M. Mattila, D. K. Farmer, L. Ampollini, P. F. DeCarlo, A. Novoselac, M. E. Vance, W. W. Nazaroff and J. P. D. Abbatt, Surface reservoirs dominate dynamic gas-surface partitioning of many indoor air constituents, *Sci. Adv.*, 2020, **6**, eaay8973.
- 20 M. Webb, L. Cui, G. Morrison, K. Baumann, J. D. Surratt, Z. Zhang, J. Atkin and B. J. Turpin, The fate of organic peroxides indoors: quantifying humidity-dependent uptake on naturally soiled indoor window glass, *Environ. Sci.: Processes Impacts*, 2023, **25**, 1031–1048.

- 21 S. K. Brown, Volatile organic pollutants in new and established buildings in Melbourne, Australia, *Indoor Air*, 2002, **12**, 55–63.
- 22 F. Caron, R. Guichard, L. Robert, M. Verrielle and F. Thevenet, Behaviour of individual VOCs in indoor environments: How ventilation affects emission from materials, *Atmos. Environ.*, 2020, **243**, 117713.
- 23 S. B. Holøs, A. Yang, M. Lind, K. Thunshelle, P. Schild and M. Mysen, VOC emission rates in newly built and renovated buildings, and the influence of ventilation – a review and meta-analysis, *Int. J. Vent.*, 2018, **18**, 1–14.
- 24 J. Pei, Y. Yin, J. Cao, Y. Sun, J. Liu and Y. Zhang, Time dependence of characteristic parameter for semi-volatile organic compounds (SVOCs) emitted from indoor materials, *Build. Environ.*, 2017, **125**, 339–347.
- 25 Z. Liu, W. Ye and J. C. Little, Predicting emissions of volatile and semivolatile organic compounds from building materials: A review, *Build. Environ.*, 2013, **64**, 7–25.
- 26 C. J. Hennigan, M. H. Bergin, A. G. Russell, A. Nenes and R. J. Weber, Gas/particle partitioning of water-soluble organic aerosol in Atlanta, *Atmos. Chem. Phys.*, 2009, **9**, 3613–3628.
- 27 J. D. Blando and B. J. Turpin, Secondary organic aerosol formation in cloud and fog droplets: a literature evaluation of plausibility, *Atmos. Environ.*, 2000, **34**, 1623–1632.
- 28 Y.-H. Lin, Z. Zhang, K. S. Docherty, H. Zhang, S. H. Budisulistiorini, C. L. Rubitschun, S. L. Shaw, E. M. Knipping, E. S. Edgerton, T. E. Kleindienst, A. Gold and J. D. Surratt, Isoprene epoxydiols as precursors to secondary organic aerosol formation: acid-catalyzed reactive uptake studies with authentic compounds, *Environ. Sci. Technol.*, 2012, **46**, 250–258.
- 29 S. M. Duncan, S. Tomaz, G. Morrison, M. Webb, J. Atkin, J. D. Surratt and B. J. Turpin, Dynamics of Residential Water-Soluble Organic Gases: Insights into Sources and Sinks, *Environ. Sci. Technol.*, 2019, **53**, 1812–1821.
- 30 S. Liu, R. Li, R. J. Wild, C. Warneke, J. A. de Gouw, S. S. Brown, S. L. Miller, J. C. Luongo, J. L. Jimenez and P. J. Ziemann, Contribution of human-related sources to indoor volatile organic compounds in a university classroom, *Indoor Air*, 2016, **26**, 925–938.
- 31 C. Wang, J. M. Mattila, D. K. Farmer, C. Arata, A. H. Goldstein and J. P. D. Abbatt, Behavior of Isocyanic Acid and Other Nitrogen-Containing Volatile Organic Compounds in The Indoor Environment, *Environ. Sci. Technol.*, 2022, **56**, 7598–7607.
- 32 D. J. Price, D. A. Day, D. Pagonis, H. Stark, L. B. Algrim, A. V. Handschy, S. Liu, J. E. Krechmer, S. L. Miller, J. F. Hunter, J. A. de Gouw, P. J. Ziemann and J. L. Jimenez, Budgets of organic carbon composition and oxidation in indoor air, *Environ. Sci. Technol.*, 2019, **53**, 13053–13063.
- 33 D. Pagonis, D. J. Price, L. B. Algrim, D. A. Day, A. V. Handschy, H. Stark, S. L. Miller, J. de Gouw, J. L. Jimenez and P. J. Ziemann, Time-Resolved Measurements of Indoor Chemical Emissions, Deposition, and Reactions in a University Art Museum, *Environ. Sci. Technol.*, 2019, **53**, 4794–4802.
- 34 X. Lyu, Y. Huo, J. Yang, D. Yao, K. Li, H. Lu, Y. Zeren and H. Guo, Real-time molecular characterization of air pollutants in a Hong Kong residence: Implication of indoor source emissions and heterogeneous chemistry, *Indoor Air*, 2021, **31**, 1340–1352.
- 35 J. Qiu, D. Xie, Y. Li, Y. Qu, Y. Liu, T. Zhu and Y. Liu, Dibasic esters observed as potential emerging indoor air pollutants in new apartments in Beijing, China, *Environ. Sci. Technol. Lett.*, 2021, **8**, 445–450.
- 36 S. J. Solomon, G. W. Schade, J. Kuttippurath, A. Ladstätter-Weissenmayer and J. P. Burrows, VOC concentrations in an indoor workplace environment of a university building, *Indoor Built Environ.*, 2008, **17**, 260–268.
- 37 Y. Qu, D. Xie and Y. Liu, Emissions of Volatile Organic Compounds from Human Occupants in a Student Office: Dependence on Ozone Concentration, *ACS Environ. Au*, 2024, **4**, 3–11.
- 38 R. Sheu, C. Stönnner, J. C. Ditto, T. Klüpfel, J. Williams and D. R. Gentner, Human transport of thirdhand tobacco smoke: A prominent source of hazardous air pollutants into indoor nonsmoking environments, *Sci. Adv.*, 2020, **6**, eaay4109.
- 39 D. J. Price, D. A. Day, D. Pagonis, H. Stark, A. V. Handschy, L. B. Algrim, S. L. Miller, J. A. de Gouw, P. J. Ziemann and J. L. Jimenez, Sources of Gas-Phase Species in an Art Museum from Comprehensive Real-Time Measurements, *ACS Earth Space Chem.*, 2021, **5**, 2252–2267.
- 40 R. Meinighaus, L. Gunnarsen and H. N. Knudsen, Diffusion and sorption of volatile organic compounds in building materials—impact on indoor air quality, *Environ. Sci. Technol.*, 2000, **34**, 3101–3108.
- 41 D. K. Farmer, M. E. Vance, J. P. D. Abbatt, A. Abeleira, M. R. Alves, C. Arata, E. Boedicker, S. Bourne, F. Cardoso-Saldaña, R. Corsi, P. F. DeCarlo, A. H. Goldstein, V. H. Grassian, L. Hildebrandt Ruiz, J. L. Jimenez, T. F. Kahan, E. F. Katz, J. M. Mattila, W. W. Nazaroff, A. Novoselac, R. E. O'Brien, V. W. Or, S. Patel, S. Sankhyan, P. S. Stevens, Y. Tian, M. Wade, C. Wang, S. Zhou and Y. Zhou, Overview of HOMEChem: house observations of microbial and environmental chemistry, *Environ. Sci.: Processes Impacts*, 2019, **21**, 1280–1300.
- 42 H. Schwartz-Narbonne, J. P. D. Abbatt, P. F. DeCarlo, D. K. Farmer, J. M. Mattila, C. Wang, D. J. Donaldson and J. A. Siegel, Modeling the Removal of Water-Soluble Trace Gases from Indoor Air via Air Conditioner Condensate, *Environ. Sci. Technol.*, 2021, **55**, 10987–10993.
- 43 A. Manuja, J. Ritchie, K. Buch, Y. Wu, C. M. A. Eichler, J. C. Little and L. C. Marr, Total surface area in indoor environments, *Environ. Sci.: Processes Impacts*, 2019, **21**, 1384–1392.
- 44 R. M. B. O. Duarte and A. C. Duarte, On the Water-Soluble Organic Matter in Inhalable Air Particles: Why Should Outdoor Experience Motivate Indoor Studies?, *Appl. Sci.*, 2021, **11**, 9917.

- 45 M. M. Lunden, T. W. Kirchstetter, T. L. Thatcher, S. V. Hering and N. J. Brown, Factors affecting the indoor concentrations of carbonaceous aerosols of outdoor origin, *Atmos. Environ.*, 2008, **42**, 5660–5671.
- 46 E. Diapouli, A. Chaloulakou and P. Koutrakis, Estimating the concentration of indoor particles of outdoor origin: A review, *J. Air Waste Manage. Assoc.*, 2013, **63**, 1113–1129.
- 47 A. M. Johnson, M. S. Waring and P. F. DeCarlo, Real-time transformation of outdoor aerosol components upon transport indoors measured with aerosol mass spectrometry, *Indoor Air*, 2017, **27**, 230–240.
- 48 C. J. Weschler and H. C. Shields, Indoor ozone/terpene reactions as a source of indoor particles, *Atmos. Environ.*, 1999, **33**, 2301–2312.
- 49 M. S. Waring and J. A. Siegel, Indoor secondary organic aerosol formation initiated from reactions between ozone and surface-sorbed D-limonene, *Environ. Sci. Technol.*, 2013, **47**, 6341–6348.
- 50 W. L. Brown, D. A. Day, H. Stark, D. Pagonis, J. E. Krechmer, X. Liu, D. J. Price, E. F. Katz, P. F. DeCarlo, C. G. Masoud, D. S. Wang, L. Hildebrandt Ruiz, C. Arata, D. M. Lunderberg, A. H. Goldstein, D. K. Farmer, M. E. Vance and J. L. Jimenez, Real-time organic aerosol chemical speciation in the indoor environment using extractive electrospray ionization mass spectrometry, *Indoor Air*, 2021, **31**, 141–155.
- 51 G. C. Morrison, A. Eftekhari, P. S. J. Lakey, M. Shiraiwa, B. E. Cummings, M. S. Waring and B. Williams, Partitioning of reactive oxygen species from indoor surfaces to indoor aerosols, *Environ. Sci.: Processes Impacts*, 2022, 2310–2323.
- 52 M. Kruza, G. McFiggans, M. S. Waring, J. R. Wells and N. Carslaw, Indoor secondary organic aerosols: Towards an improved representation of their formation and composition in models, *Atmos. Environ.*, 2020, **240**, 117784.
- 53 P. F. DeCarlo, A. M. Avery and M. S. Waring, Thirdhand smoke uptake to aerosol particles in the indoor environment, *Sci. Adv.*, 2018, **4**, eaap8368.
- 54 J. M. Mattila, P. S. J. Lakey, M. Shiraiwa, C. Wang, J. P. D. Abbatt, C. Arata, A. H. Goldstein, L. Ampollini, E. F. Katz, P. F. DeCarlo, S. Zhou, T. F. Kahan, F. J. Cardoso-Saldaña, L. H. Ruiz, A. Abeleira, E. K. Boedicker, M. E. Vance and D. K. Farmer, Multiphase Chemistry Controls Inorganic Chlorinated and Nitrogenated Compounds in Indoor Air during Bleach Cleaning, *Environ. Sci. Technol.*, 2020, **54**, 1730–1739.
- 55 A. Polidori, B. Turpin, Q. Y. Meng, J. H. Lee, C. Weisel, M. Morandi, S. Colome, T. Stock, A. Winer, J. Zhang, J. Kwon, S. Alimokhtari, D. Shendell, J. Jones, C. Farrar and S. Maberti, Fine organic particulate matter dominates indoor-generated PM_{2.5} in RIOPA homes, *J. Exposure Sci. Environ. Epidemiol.*, 2006, **16**, 321–331.
- 56 *Handbook of Indoor Air Quality*, ed. Y. Zhang, P. K. Hopke and C. Mandin, Springer Singapore, Singapore, 2021.
- 57 N. Daher, A. Ruprecht, G. Invernizzi, C. De Marco, J. Miller-Schulze, J. B. Heo, M. M. Shafer, J. J. Schauer and C. Sioutas, Chemical characterization and source apportionment of fine and coarse particulate matter inside the refectory of Santa Maria Delle Grazie Church, home of Leonardo Da Vinci's "Last Supper", *Environ. Sci. Technol.*, 2011, **45**, 10344–10353.
- 58 S. Hasheminassab, N. Daher, M. M. Shafer, J. J. Schauer, R. J. Delfino and C. Sioutas, Chemical characterization and source apportionment of indoor and outdoor fine particulate matter (PM_{2.5}) in retirement communities of the Los Angeles Basin, *Sci. Total Environ.*, 2014, **490**, 528–537.
- 59 D. G. Poppendieck, L. C. Ng, A. K. Persily and A. T. Hodgson, Long term air quality monitoring in a net-zero energy residence designed with low emitting interior products, *Build. Environ.*, 2015, **94**, 33–42.
- 60 W. M. Healy, C. Gates, A. H. Fannery and B. Pettit, *Design Challenges of the NIST Net Zero Energy Residential Test Facility*, National Institute of Standards and Technology, Gaithersburg, MD, 2015.
- 61 Berkeley Analytical Associates LLC, Bernheim + Dean Inc and White + GreenSpec, *High Performance Indoor Air Quality Specification for Net Zero Energy Homes*, National Institute of Standards and Technology, Gaithersburg, MD, 2015.
- 62 W. R. Cofer, V. G. Collins and R. W. Talbot, Improved aqueous scrubber for collection of soluble atmospheric trace gases, *Environ. Sci. Technol.*, 1985, **19**, 557–560.
- 63 R. S. Spaulding, R. W. Talbot and M. J. Charles, Optimization of a mist chamber (Cofer scrubber) for sampling water-soluble organics in air, *Environ. Sci. Technol.*, 2002, **36**, 1798–1808.
- 64 N. Sareen, A. G. Carlton, J. D. Surratt, A. Gold, B. Lee, F. D. Lopez-Hilfiker, C. Mohr, J. A. Thornton, Z. Zhang, Y. B. Lim and B. J. Turpin, Identifying precursors and aqueous organic aerosol formation pathways during the SOAS campaign, *Atmos. Chem. Phys.*, 2016, **16**, 14409–14420.
- 65 C. H. Anderson, J. E. Dibb, R. J. Griffin, G. S. W. Hagler and M. H. Bergin, Atmospheric water-soluble organic carbon measurements at Summit, Greenland, *Atmos. Environ.*, 2008, **42**, 5612–5621.
- 66 C. J. Hennigan, M. M. H. El-Sayed and A. Hodzic, Detailed characterization of a mist chamber for the collection of water-soluble organic gases, *Atmos. Environ.*, 2018, **188**, 12–17.
- 67 D. A. Orsini, Y. Ma, A. Sullivan, B. Sierau, K. Baumann and R. J. Weber, Refinements to the particle-into-liquid sampler (PILS) for ground and airborne measurements of water soluble aerosol composition, *Atmos. Environ.*, 2003, **37**, 1243–1259.
- 68 R. J. Weber, D. Orsini, Y. Daun, Y. N. Lee, P. J. Klotz and F. Brechtel, A Particle-into-Liquid Collector for Rapid Measurement of Aerosol Bulk Chemical Composition, *Aerosol Sci. Technol.*, 2001, **35**, 718–727.
- 69 A. P. Sullivan, R. J. Weber, A. L. Clements, J. R. Turner, M. S. Bae and J. J. Schauer, A method for on-line measurement of water-soluble organic carbon in ambient

- aerosol particles: Results from an urban site, *Geophys. Res. Lett.*, 2004, **31**(13), DOI: [10.1029/2004GL019681](https://doi.org/10.1029/2004GL019681).
- 70 A. P. Sullivan, N. Hodas, B. J. Turpin, K. Skog, F. N. Keutsch, S. Gilardoni, M. Paglione, M. Rinaldi, S. Decesari, M. C. Facchini, L. Poulain, H. Herrmann, A. Wiedensohler, E. Nemitz, M. M. Twigg and J. L. Collett Jr, Evidence for ambient dark aqueous SOA formation in the Po Valley, Italy, *Atmos. Chem. Phys.*, 2016, **16**, 8095–8108.
 - 71 Y. Jin, C. Yan, A. P. Sullivan, Y. Liu, X. Wang, H. Dong, S. Chen, L. Zeng, J. L. Collett Jr and M. Zheng, Significant Contribution of Primary Sources to Water-Soluble Organic Carbon During Spring in Beijing, China, *Atmosphere*, 2020, **11**, 395.
 - 72 Y. Miyazaki, Y. Kondo, N. Takegawa, Y. Komazaki, M. Fukuda, K. Kawamura, M. Mochida, K. Okuzawa and R. J. Weber, Time-resolved measurements of water-soluble organic carbon in Tokyo, *J. Geophys. Res.*, 2006, **111**(D23), DOI: [10.1029/2006JD007125](https://doi.org/10.1029/2006JD007125).
 - 73 A. P. Sullivan and R. J. Weber, Chemical characterization of the ambient organic aerosol soluble in water: 2. Isolation of acid, neutral, and basic fractions by modified size-exclusion chromatography, *J. Geophys. Res.*, 2006, **111**(D5), DOI: [10.1029/2005JD006486](https://doi.org/10.1029/2005JD006486).
 - 74 J. Krechmer, F. Lopez-Hilfiker, A. Koss, M. Hutterli, C. Stoermer, B. Deming, J. Kimmel, C. Warneke, R. Holzinger, J. Jayne, D. Worsnop, K. Fuhrer, M. Gonin and J. de Gouw, Evaluation of a New Reagent-Ion Source and Focusing Ion-Molecule Reactor for Use in Proton-Transfer-Reaction Mass Spectrometry, *Anal. Chem.*, 2018, **90**, 12011–12018.
 - 75 M. S. Claflin, D. Pagonis, Z. Finewax, A. V. Handschy, D. A. Day, W. L. Brown, J. T. Jayne, D. R. Worsnop, J. L. Jimenez, P. J. Ziemann, J. de Gouw and B. M. Lerner, An *in situ* gas chromatograph with automatic detector switching between PTR- and EI-TOF-MS: isomer-resolved measurements of indoor air, *Atmos. Meas. Tech.*, 2021, **14**, 133–152.
 - 76 B. H. Lee, F. D. Lopez-Hilfiker, C. Mohr, T. Kurtén, D. R. Worsnop and J. A. Thornton, An iodide-adduct high-resolution time-of-flight chemical-ionization mass spectrometer: application to atmospheric inorganic and organic compounds, *Environ. Sci. Technol.*, 2014, **48**, 6309–6317.
 - 77 J. Li, M. F. Link, S. Pandit, M. H. Webb, K. J. Mayer, L. A. Garofalo, K. L. Rediger, D. G. Poppendieck, S. M. Zimmerman, M. E. Vance, V. H. Grassian, G. C. Morrison, B. J. Turpin and D. K. Farmer, The persistence of smoke VOCs indoors: Partitioning, surface cleaning, and air cleaning in a smoke-contaminated house, *Sci. Adv.*, 2023, **9**, eadh8263.
 - 78 B. J. Turpin and H.-J. Lim, Species contributions to PM_{2.5} mass concentrations: revisiting common assumptions for estimating organic mass, *Aerosol Sci. Technol.*, 2001, **35**, 602–610.
 - 79 J. S. Park and K. Ikeda, Variations of formaldehyde and VOC levels during 3 years in new and older homes, *Indoor Air*, 2006, **16**, 129–135.
 - 80 E. Gallego, X. Roca, J. F. Perales and X. Guardino, Determining indoor air quality and identifying the origin of odour episodes in indoor environments, *J. Environ. Sci.*, 2009, **21**, 333–339.
 - 81 C. Arata, P. K. Misztal, Y. Tian, D. M. Lunderberg, K. Kristensen, A. Novoselac, M. E. Vance, D. K. Farmer, W. W. Nazaroff and A. H. Goldstein, Volatile organic compound emissions during HOMEChem, *Indoor Air*, 2021, **31**, 2099–2117.
 - 82 S. H. Shin and W. K. Jo, Volatile organic compound concentrations, emission rates, and source apportionment in newly-built apartments at pre-occupancy stage, *Chemosphere*, 2012, **89**, 569–578.
 - 83 W. Liu, J. Zhang, L. Zhang, B. Turpin, C. Weisel, M. Morandi, T. Stock, S. Colome and L. Korn, Estimating contributions of indoor and outdoor sources to indoor carbonyl concentrations in three urban areas of the United States, *Atmos. Environ.*, 2006, **40**, 2202–2214.
 - 84 X. M. Wu, M. G. Apte, R. Maddalena and D. H. Bennett, Volatile organic compounds in small- and medium-sized commercial buildings in California, *Environ. Sci. Technol.*, 2011, **45**, 9075–9083.
 - 85 E. L. Hult, H. Willem, P. N. Price, T. Hotchi, M. L. Russell and B. C. Singer, Formaldehyde and acetaldehyde exposure mitigation in US residences: in-home measurements of ventilation control and source control, *Indoor Air*, 2015, **25**, 523–535.
 - 86 T. Salthammer, Acetaldehyde in the indoor environment, *Environ. Sci.: Atmos.*, 2023, **3**, 474–493.
 - 87 E. Pibiri, A. Omelan, E. Uhde and T. Salthammer, Effect of surface covering on the release of formaldehyde, acetaldehyde, formic acid and acetic acid from particleboard, *Build. Environ.*, 2020, **178**, 106947.
 - 88 A. Schieweck, Very volatile organic compounds (VVOC) as emissions from wooden materials and in indoor air of new prefabricated wooden houses, *Build. Environ.*, 2021, **190**, 107537.
 - 89 Y. Liu, P. K. Misztal, J. Xiong, Y. Tian, C. Arata, R. J. Weber, W. W. Nazaroff and A. H. Goldstein, Characterizing sources and emissions of volatile organic compounds in a northern California residence using space- and time-resolved measurements, *Indoor Air*, 2019, **29**, 630–644.
 - 90 C. E. Stockwell, M. M. Coggon, G. I. Gkatzelis, J. Ortega, B. C. McDonald, J. Peischl, K. Aikin, J. B. Gilman, M. Trainer and C. Warneke, Volatile organic compound emissions from solvent- and water-borne coatings – compositional differences and tracer compound identifications, *Atmos. Chem. Phys.*, 2021, **21**, 6005–6022.
 - 91 J. C. Chang, R. Fortmann, N. Roache and H. C. Lao, Evaluation of low-VOC latex paints, *Indoor Air*, 1999, **9**, 253–258.
 - 92 C. H. Halios, C. Landeg-Cox, S. D. Lowther, A. Middleton, T. Marczylo and S. Dimitroulopoulou, Chemicals in European residences - Part I: A review of emissions, concentrations and health effects of volatile organic compounds (VOCs), *Sci. Total Environ.*, 2022, **839**, 156201.

- 93 O. Wilke, O. Jann and D. Brödner, VOC- and SVOC-emissions from adhesives, floor coverings and complete floor structures, *Indoor Air*, 2004, **14**(Suppl 8), 98–107.
- 94 R. Reiss, P. B. Ryan, S. J. Tibbetts and P. Koutrakis, Measurement of organic acids, aldehydes, and ketones in residential environments and their relation to ozone, *J. Air Waste Manage. Assoc.*, 1995, **45**, 811–822.
- 95 C. J. Weschler, Ozone in indoor environments: concentration and chemistry, *Indoor Air*, 2000, **10**, 269–288.
- 96 T. Salthammer, Analytical chemistry of carbonyl compounds in indoor air, *Analyst*, 2023, **148**, 3432–3451.
- 97 A. R. Ferro, R. J. Kopperud and L. M. Hildemann, Source strengths for indoor human activities that resuspend particulate matter, *Environ. Sci. Technol.*, 2004, **38**, 1759–1764.
- 98 J. J. Schauer, Source contributions to atmospheric organic compound concentrations: emissions measurements and model predictions, PhD Dissertation, California Institute of Technology, 1998, DOI: [10.7907/3fph-hy50](https://doi.org/10.7907/3fph-hy50).
- 99 W. W. Nazaroff and C. J. Weschler, Indoor acids and bases, *Indoor Air*, 2020, **30**, 559–644.
- 100 M. Tuomainen, A.-L. Pasanen, A. Tuomainen, J. Liesivuori and P. Juvonen, Usefulness of the Finnish classification of indoor climate, construction and finishing materials: comparison of indoor climate between two new blocks of flats in Finland, *Atmos. Environ.*, 2001, **35**, 305–313.
- 101 H. Jarnstrom, K. Saarela, P. Kalliokoski and A. Pasanen, Reference values for indoor air pollutant concentrations in new, residential buildings in Finland, *Atmos. Environ.*, 2006, **40**, 7178–7191.
- 102 D. G. Poppendieck, L. C. Ng, A. K. Persily and A. T. Hodgson, Long term air quality monitoring in a net-zero energy residence designed with low emitting interior products, *Build. Environ.*, 2015, **94**, 33–42.
- 103 T. Salthammer, F. Fuhrmann, S. Kaufhold, B. Meyer and A. Schwarz, Effects of climatic parameters on formaldehyde concentrations in indoor air, *Indoor Air*, 1995, **5**, 120–128.
- 104 H. Willem, E. L. Hult, T. Hotchi, M. L. Russell, R. L. Maddalena and B. C. Singer, *Ventilation Control of Volatile Organic Compounds in New U.S. Homes: Results of a Controlled Field Study in Nine Residential Units*, Lawrence Berkeley National Laboratory (LBNL), Berkeley, CA, United States, 2013.
- 105 J. M. Mattila, C. Arata, A. Abeleira, Y. Zhou, C. Wang, E. F. Katz, A. H. Goldstein, J. P. D. Abbatt, P. F. DeCarlo, M. E. Vance and D. K. Farmer, Contrasting chemical complexity and the reactive organic carbon budget of indoor and outdoor air, *Environ. Sci. Technol.*, 2022, **56**, 109–118.
- 106 H. Destailhats, M. M. Lunden, B. C. Singer, B. K. Coleman, A. T. Hodgson, C. J. Weschler and W. W. Nazaroff, Indoor secondary pollutants from household product emissions in the presence of ozone: A bench-scale chamber study, *Environ. Sci. Technol.*, 2006, **40**, 4421–4428.
- 107 A. K. Persily, L. C. Ng, D. G. Poppendieck and S. J. Emmerich, Ventilation, IAQ and Filtration in a Net Zero Energy House, *Air Media Magazine*, 2018.
- 108 S. Tomaz, T. Cui, Y. Chen, K. G. Sexton, J. M. Roberts, C. Warneke, R. J. Yokelson, J. D. Surratt and B. J. Turpin, Photochemical cloud processing of primary wildfire emissions as a potential source of secondary organic aerosol, *Environ. Sci. Technol.*, 2018, **52**, 11027–11037.
- 109 A. R. Koss, K. Sekimoto, J. B. Gilman, V. Selimovic, M. M. Coggon, K. J. Zarzana, B. Yuan, B. M. Lerner, S. S. Brown, J. L. Jimenez, J. Krechmer, J. M. Roberts, C. Warneke, R. J. Yokelson and J. de Gouw, Non-methane organic gas emissions from biomass burning: identification, quantification, and emission factors from PTR-ToF during the FIREX 2016 laboratory experiment, *Atmos. Chem. Phys.*, 2018, **18**, 3299–3319.
- 110 L. E. Hatch, W. Luo, J. F. Pankow, R. J. Yokelson, C. E. Stockwell and K. C. Barsanti, Identification and quantification of gaseous organic compounds emitted from biomass burning using two-dimensional gas chromatography–time-of-flight mass spectrometry, *Atmos. Chem. Phys.*, 2015, **15**, 1865–1899.
- 111 S. Kundu, K. Kawamura, T. W. Andreae, A. Hoffer and M. O. Andreae, Molecular distributions of dicarboxylic acids, ketocarboxylic acids and α -dicarbonyls in biomass burning aerosols: implications for photochemical production and degradation in smoke layers, *Atmos. Chem. Phys.*, 2010, **10**, 2209–2225.