



Cite this: *Environ. Sci.: Atmos.*, 2023, **3**, 328

Development of a microsampler suitable for aerial collection of aerosol particles

Meng-Dawn Cheng, ^a Chih-Hsiang Chien, ^a David E. Graham, ^b and Andrew Harter ^c

A microsampler for collecting aerosol particles on an unmanned aerial system (UAS) was designed and evaluated in the laboratory for dry (no condensed water) and wet (foggy, misty, or rainy) air conditions. This microsampler was produced using a customized impactor design to minimize size, weight, and power requirements uniquely suitable for aerial applications, also maximize the collection efficiency above a specific cut-off size in a short time and concentrate particles for future automated analysis using microfluidics. The computational fluid dynamics (CFD) results suggested a design that provided a higher sampling efficiency with minimal inlet loss. The microsampler was experimentally evaluated using fluorescent polystyrene latex (PSL) particles at four nominal sizes. The experimental results at the four PSL particle sizes agree reasonably well with the collection efficiencies predicted by CFD. We found no effect of humidity on the collection efficiency of the microsampler under conditions at 80% relative humidity. However, when condensed water exists, significant biases on the collection efficiencies were found suggesting that this microsampler should not be used under conditions where there is condensed water in the atmosphere, e.g., foggy, misty, and rainy conditions.

Received 29th July 2022
 Accepted 12th December 2022
 DOI: 10.1039/d2ea00096b
rsc.li/esatmospheres

Environmental significance

An unmanned aerial system (UAS) can seek out and selectively collect aerosol particles *in situ* to significantly increase the signal for source identification, apportionment, and improving understandings of atmospheric science. A microsampler for collecting aerosol particles onboard the UAS would have light weight, small size, low power requirement, and have robust performance under all flight conditions. The microsampler was designed and performed reasonably well, and its collection efficiency agreed with the collection efficiencies predicted by CFD. We found that below 80% relative humidity the collection efficiency was not impacted. However, when condensed water exists, the collection efficiencies could be affected suggesting that this microsampler should not be used under conditions where there is condensed water in the atmosphere, e.g., foggy, misty, or rainy conditions.

1. Introduction

Sampling airborne particles or particulate matter (PM), simply called “aerosols,” from an unmanned aerial system (UAS), quadcopter, balloon, airship, drone, and dropsonde requires a small and high efficiency sampler or collector. There are many commercial aerosol samplers in the market, but none that we know of meets all the requirements for being an aerial aerosol sampler. Small footprint, light weight, and low power consumption are three basic requirements. A number of air quality measurements using UASs have been reported, deploying portable gas analysers and simple particle counters or spectrometers for continuous measurements.¹ However, few aerosol sampling platforms that concentrate and collect

particles for specific analyses have been reported, and they are mostly used in specialized applications.

Aerosol samplers designed for personal exposure studies are typically small. For example, the commercially available sioutas five-stage cascade impactor (SKC, Inc.) was designed to collect respirable particles ranging from 0.25 to 10 µm at a flow rate of 9 litres per minute (LPM).² The impactor is reasonably small with overall dimensions of 8.6 cm × 5.5 cm and a mass of 159 g (<https://www.skinc.com/products/sioutas-five-stage-cascade-impactor>; US Patent 6786105). The impactor (*i.e.*, the main collector body itself) is suitable for applications involving monitoring human respiratory health in which case power requirement is less of an issue. Such a collector would not be desirable for an aerial aerosol sampling application³ that requires a device with overall volumetric dimensions of less than 8 cm³ and weighs less than 75 g, for example, in addition to low power requirement.

Furthermore, the sioutas impactor, like all aerosol collectors, requires a vacuum pump which in turn requires a considerable amount of current through an AC or DC power source (battery).

^aEnvironmental Sciences Division, Oak Ridge National Laboratory, Oak Ridge, TN 37831-6036, USA. E-mail: chengmd@ornl.gov; Tel: +1-865-241-5918

^bBiosciences Division, Oak Ridge National Laboratory, Oak Ridge, TN 37831, USA

^cGeospatial Science and Human Security Division, Oak Ridge National Laboratory, Oak Ridge, TN 37831, USA



A portable, battery-powered pump, in general, would occupy a great portion of the overall system weight and size, because of the power requirement for the impactor operation. Power supply limitations often control the cost and capabilities of UASs;⁴ therefore, a low power requirement is another important consideration in the design of an aerosol sampler for aerial UAS applications.

A few miniature aerosol sizers have been designed to meet the need for extremely small and lightweight systems; thereby, these would be good candidates for aerial particle sampling. For example, a miniature virtual impactor designed to sample aerosol particles smaller than 2.5 μm was incorporated into an air-microfluidic sensor system.⁵ Several chip-based microfluidic sensors have been developed and tested for the toxicity measurements of aerosol particles, particularly when the toxic agents cause oxidative stress. Koehler *et al.*⁶ developed a microfluidic electrochemical sensor for measurement of the aerosol oxidative load, which could have the potential to be further developed for sensing other aerosol components. Brubaker *et al.*⁷ developed and applied a microfluidic device to investigate freezing properties of ice-nucleating particles. Liu *et al.*⁸ also explored the use of microfluidic online assay for aerosol oxidative potentials. Mei and Goldberger developed an 8-channel filter sampler for aerosol collection on a UAS platform for offline, ground-based analysis.⁹ Cazzolara *et al.* designed a 0.6 kg particle collection system composed of an air inlet, impactor, and electric blower to collect pollen and large spores ($>20 \mu\text{m}$) with an airflow of 200 LPM using a multi-copter.¹⁰ Bieber *et al.* used both a cascade impactor (0.69 kg) and an impinger (0.57 kg) particle collector for UAS-based analyses of ice nucleation particles.¹¹ Vacuum pumps sampled air at flow rates from 1 to 9 LPM. Alternatively, Schrod *et al.*¹² built a programmable electrostatic aerosol collector (0.6 kg) to collect ice nucleating particles in a UAS using a pump to sample air at 5 LPM.

Microfluidic sensors require substantially smaller volumes of samples than traditional sensors for analysis and can provide near real-time results. They are ideal for autonomous airborne aerosol research applications. There has been limited development in micro aerosol samplers or collectors. A quarter-sized miniature aerosol cyclone was designed to classify particles larger than 1 μm and 0.3 μm at 300 mL min^{-1} .¹³ Novoselov *et al.*¹⁴ developed a micro-channel collector and reported its performance at higher than 90% collection efficiency on particles greater than 1 μm using their single-loop micro-collector. Damit demonstrated a droplet-based microfluidic sensor for bioaerosol detection, with bioaerosol particles directly deposited into a droplet at the interface on a chip.¹⁵ With the advances in microfluidics detectors in the past couple of decades,¹⁶ future development of microfluidics sensors that can be coupled to a microsampler is also expected to increase.

In this paper, we report a microsampler that was developed specifically for UAS applications. The size and weight are two main constraints of the collector, which is also required to collect aerosol particles of a size greater than 1 μm but less than 10 μm in aerodynamic diameter. The UAS is to be guided and directed into a target area of interest, remotely by wireless

Table 1 Size, weight, and power (SWaP) requirements for the microsampler operated at 1 LPM

Dimension	Size	Weight	Power
Unit	$\text{L} \times \text{W} \times \text{H} (\text{cm}^3)$	Gram	Watts
Sampler	$3.5 \times 3.5 \times 1^a$	15	—
Pump	$4.4 \times 2 \times 1$	48	0.4

^a Note: the protruded inlet adds 1.3 cm to the height.

control. We shall discuss the design and experimental evaluation of the microsampler in this paper without reference to the specific application, except to mention that we limited the weight of the microsampler for our application to no more than 25 g and overall dimensions of 2.5 cm \times 2.5 cm \times 1 cm. Our impactor weighs approximately 10 g in an aluminium body. The sampling flow rate is restricted to 1 LPM using a coreless brush DC diaphragm pump, a 12V DC-powered micropump (Parker Model A.1F17N1.G12VDC). The weight of the overall system (the collector plus pump and electrical wires) is 63 g. These physical dimensions of this microsampler are tabulated in Table 1.

2. Methodology

2.1 Sampler design concept

Fig. 1a displays the schematic diagram of the microsampler. The concept of the design is adopted from right-angle cascade impactors.^{17,18} Conventional impactor samplers such as the micro-orifice uniform deposition impactor (MOUDI) have an inlet and outlet on opposite sides, and the aerosol is forced to make a 180-degree change in direction. However, the right-angle impactor has an unconventional flow pattern, and the flow must make two 90-degree turns to leave the system. For meeting the ultrasmall footprint and weight limitation, our microsampler described in this paper needs only a one-stage impaction as compared to previous right-angle impactors that were slightly bulkier and heavier for our applications. Fig. 1b shows a 3D rendition of the cut-away view of the microsampler above the plane of the collection well and micro exit flow channel.

Fig. 1c displays a fabricated microsampler that has a separated upper body (above the collector plane) and a lower body containing an exit microchannel and a collection well was magnetically attached. Four round N95 permanent magnets sealed both bodies and found no air leak during the evaluation process. It is important for the design of a right-angle impactor that after the flow leaves the impaction surface, the exit reduces the flow to achieve a lower velocity and prevent potential wall loss.¹⁷⁻²¹ The following sections discuss the design of the impactor and the orifice.

2.1.1 General consideration of the impactor design. Our sampler design follows the principle of virtual impaction in collection of aerosol particles. Impaction of particles on a collection substrate is determined using the Stokes number of the particle. In other words, the particle transmission through an impactor is governed by the impactor theory derived previously^{19,20} and formulated as follows:



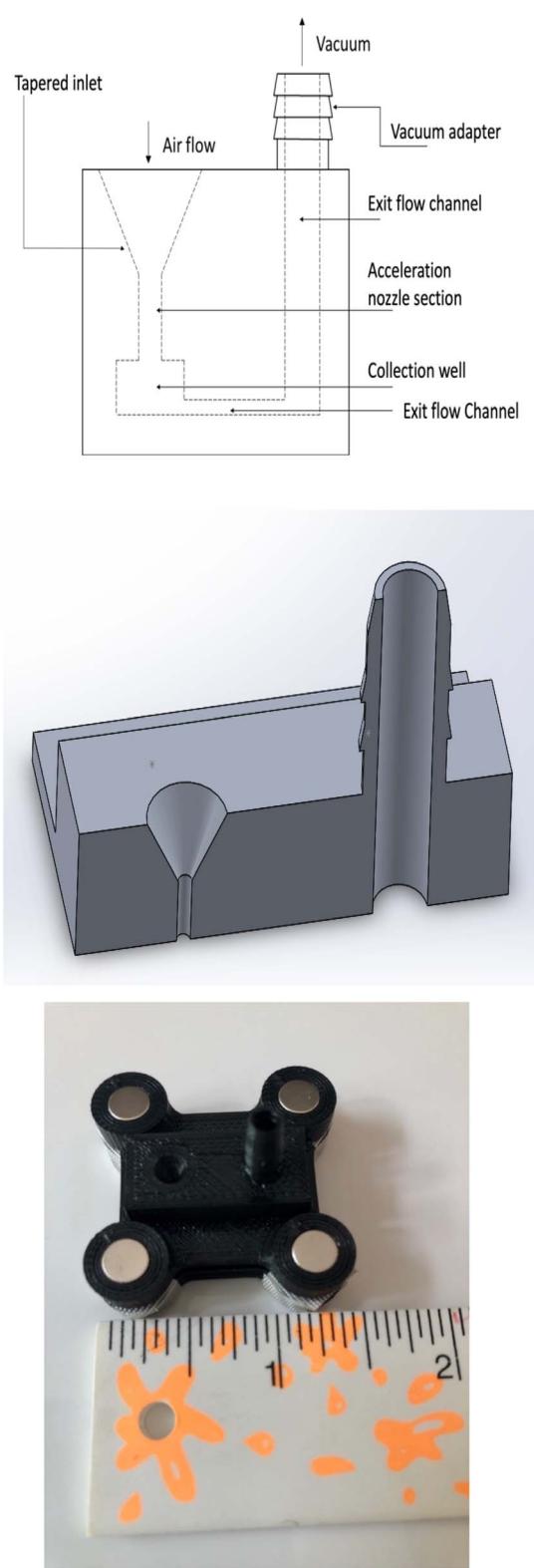


Fig. 1 (a) Schematic diagram of a prototype micro aerosol sampler. (b) 3D rendition of the tapered inlet. (c) Photo showing a 3D-printed microsampl. The ruler is on the cm scale.

$$St = \frac{\rho_p d_p^2 U C_c}{9 \mu D_j}$$

where St is the Stokes number, ρ_p is particle density, d_p is the particle size, U is the free-stream jet velocity relative to the jet radius $\left(\frac{D_j}{2}\right)$ of the impactor, and C_c is the Cunningham correction factor accounting for the particle slippage as their size becomes smaller than the mean free path of the air molecule, *e.g.*, $0.6 \mu\text{m}$ at 20°C .

If the surface of the collection substrate is rigid, the Stokes number for 50% collection efficiency (St_{50}) is 0.24 for a circular jet. The cut size or cut diameter of the impactor for the particle having a 50% collection efficiency (d_{50}) can then be derived as:

$$d_{50} \sqrt{C_c} = \left[\frac{9\pi\mu D_j^3 (St_{50})}{4\rho_p Q} \right]^{0.5}$$

At a design flow rate (Q) of 1.0 LPM, the sampler with a circular jet diameter (D_j) of 0.8 mm will have a jet velocity of 3316 cm s^{-1} , and the jet Reynolds number (Re) is 1759. At 20°C , the air viscosity is 1.81×10^{-4} poise while the air density is $1.2 \times 10^{-3} \text{ g cm}^{-3}$. Therefore, the designed cut size can be calculated to be $0.97 \mu\text{m}$, accordingly.

The impactor geometry, such as the dimensions of the jet-to-plate distance (S), can also influence the cut size and the collection efficiency curve.¹⁹ They also presented a detailed account of how an efficiency curve was affected by the particle size, jet Reynolds number, and or the ratio of the jet-to-plate distance to the nozzle diameter (S/D_j). However, if the S/D_j was designed in the range from 1 to 10 and the Reynolds number is in the 500 to 3000 range, then the penetration curve is less sensitive to the S/D_j . Accordingly, our design value for the parameter S/D_j was chosen to be 1.875 to achieve the sharpest cut-off characteristics possible.

2.1.2 Effects of the sampling inlet. Like all aerosol samplers, the inlet geometry is important for the effective collection of aerosol particles. With the sampler requirement that it be airborne on an aerial moving platform such as a quadcopter, it is also critical that our design minimizes the loss of particles during the collection. Two types of sampling inlets were constructed and numerically evaluated for sampling aerosol particles into the collector. The straight geometry shown in Fig. 2a is convenient for machine fabrication, but such a design is more likely to focus particles closer to the centerline

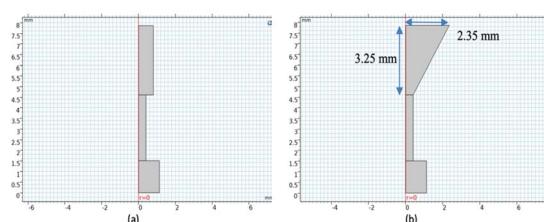


Fig. 2 Impactor geometry with (a) straight orifice and (b) tapered orifice.



than an angled orifice,²² as shown in Fig. 2b, and cause significant particle loss at the entrance of the inlet. We therefore designed a tapered inlet with a 31-degree half angle (*i.e.*, 3.25 mm at cone height and 4.7 mm at cone diameter).

The flow field and particle tracking analyses were conducted using the computational fluid dynamics (CFD) module of the COMSOL Multiphysics V5.5 software. COMSOL solves the stationary general flow field by using a finite element method and fluid dynamics modules. The flow has a Reynolds number estimated to be 1759. Particle tracking analysis was applied to produce the theoretical collection efficiency and visualize the particle flow inside the collector. Using a symmetric, 2-D axisymmetric geometry as shown in Fig. 2, the CFD modelling domain was discretized with a fine tetrahedral mesh. Additional studies using an extra fine-resolution mesh provided by COMSOL were also conducted to investigate the impact of mesh cell resolution on the simulation result.

The results are shown in Fig. 3. The extra-fine resolution composed of 20 064 tetrahedral node elements was generated automatically by COMSOL but it did not yield significant differences in the estimated collection efficiency as shown in Fig. 3. Mesh element quality and size (*i.e.*, resolution) affect the accuracy of the numerical solution. A mesh quality of 1 indicates a perfect regular element and 0 is a degenerated element. The mesh average quality for the simulations was 0.87. The minimum quality acceptable was 0.21. Normal inflow velocity was applied to the inlet, and the design inlet flow rate was 1 LPM. Free boundary conditions were prescribed at the outlet.²³ A no-slip boundary condition was applied to the wall. The flow field was solved in a steady state. To capture the major and minor flow features, the quantitative convergence tolerance was achieved at 10^{-6} .²⁴

After the steady-state flow field was constructed, a time-dependent particle tracking analysis was conducted using the COMSOL particle tracking module. The particle trajectory is calculated by integrating the particle force balance equation and the drag force and gravity force were included in the calculations. Drag force (F_D) is defined as

$$F_D = \frac{\rho_p d_p^2}{18\mu} (u - v)$$

where τ is the particle relaxation time, u is the fluid velocity, and v is the particle velocity. Since the particle Reynolds number is

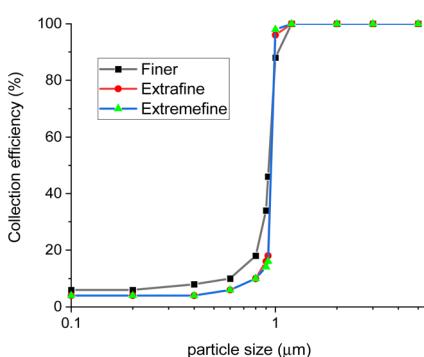


Fig. 3 Collection efficiency curves for a mesh independence study.

low, the Stokes flow was selected for drag coefficient correction. The particle relaxation time is calculated as:

$$\tau = \frac{\rho_p d_p^2}{18\mu}$$

The particle motion can be expressed as

$$\frac{dv}{dt} = \frac{1}{\tau} (u - v)$$

In 1-D flow for example, the trajectory can be expressed as²⁵

$$x(t) = u \left[t - \tau \left(1 - \exp \left(-\frac{t}{\tau} \right) \right) \right]$$

Therefore, the time step size needs to be carefully selected in the scale of particle relaxation time to ensure good simulation accuracy. In this study, the computational time step was set to be equivalent to particle relaxation time.

The 50% collection efficiency cut-off was estimated at 0.97 μm as shown in Fig. 3, consistent with the design value for a micro virtual impactor.

Fig. 4 shows the comparison between the velocity distributions of the two inlets. The flow fields of both inlets appear to be very similar, visually. However, a subtle difference exists and appears to be that a straight orifice led to a higher velocity (as shown in the darker red color) as the gas flow converged into the nozzle section. A higher flow velocity produces a larger inertia for larger particles, *e.g.*, those larger than 50 μm , which could cause blockage to the micro inlet over a short time. Also, larger particle inertia could lead to particle bounce from the substrate as they impact during the collection. The higher converging velocity in the straight inlet could produce a higher loss of particles at the transition from the inlet section to the nozzle section, as we will discuss in the subsequent sections during particle tracing analyses.

2.1.3 Evaluation of collection efficiency as a function of particle size. To analyze the collection efficiency of the micro-sampler, fifty (50) particles of the same size were used in the simulated trajectory tracking. The size was varied in different simulations for the calculation of the collection efficiency for

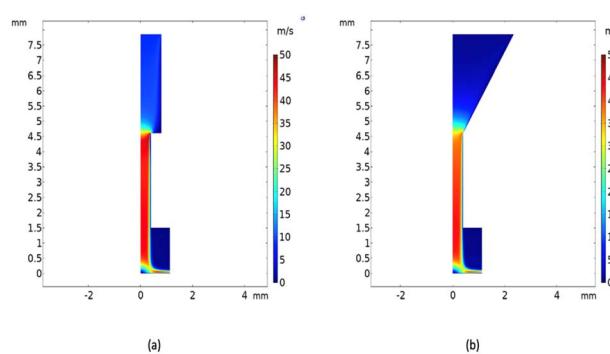


Fig. 4 Gas stream velocity profile for (a) straight orifice and (b) the tapered orifice.



the selected size. These particles were injected uniformly from the top inlet boundary. The particles were then numerically sampled on the wall, substrate (*i.e.*, particle impaction surface) at the bottom of the collection well, and outlet and were counted as n_{wall} , $n_{\text{substrate}}$, and n_{outlet} , respectively. Note that particles were conserved to ensure that all particle trajectories were tracked. The numerical collection efficiency ($\eta_{\text{numerical}}$) is defined in eqn (1) as follows:

$$\eta_{\text{numerical}} = \frac{n_{\text{substrate}}}{n_{\text{wall}} + n_{\text{substrate}} + n_{\text{outlet}}} \quad (1)$$

Fig. 5a shows the collection efficiency ($\eta_{\text{numerical}}$) curves numerically constructed for the two inlet designs. For both straight and tapered inlets, the estimated 50% efficiency cut-off size was between 0.9 and 1 μm , within the tolerance of the designed value of 0.97 μm .

The difference in the geometry of the straight inlet causes insignificant impacts on the collection efficiency compared to that of the tapered inlet as shown in Fig. 5a. Furthermore, the collection efficiency of the straight inlet decreases as the particle size (D_p) increases to be larger than the design cut size (D_{pc}) (see Fig. 5a). The decrease in the collection efficiency for $D_p > D_{pc}$ is dramatic. As shown in Fig. 5a, it decreases from 100% for 1 μm particles to less than 70% for 5 μm particles, which is a 30% loss of collection efficiency. The loss of particles displayed in Fig. 5b is another way to suggest that the micro-sampler should not be used for particles of a size greater than the design cut off.

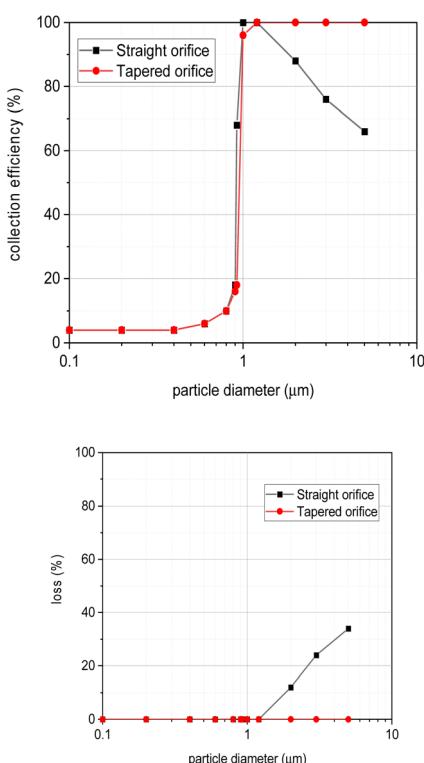


Fig. 5 (a) The collection efficiency curves for the straight and tapered orifice inlets. (b) Calculated particle loss for the two inlet designs.

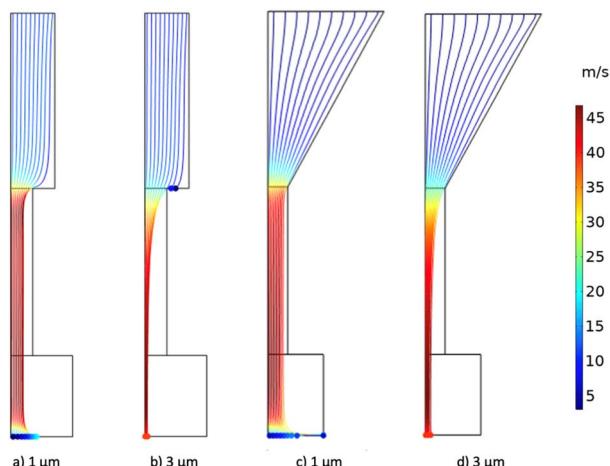


Fig. 6 1 μm and 3 μm particle trajectories for straight and tapered inlets (the color represents the velocity magnitude).

2.1.4 Particle loss analyses. Particle trajectory analysis was used for the analyses of particle loss, and the simulated results are shown in Fig. 6. Two pairs of particle trajectory plots are presented in Fig. 6 for particles of two sizes, 1 and 3 μm . For 1 μm particles, the particle trajectories appear to be closer to the center axis as in the straight inlet (Fig. 6a vs. 6c). The particles sampled by the straight inlet thus had higher momentum than those sampled by the tapered inlet and would lead to uneven collection on the substrate as compared to the tapered inlet. This could contribute to an unwanted collection event.²² Additionally, it is important to note that the straight inlet can cause the loss of particles larger than 1 μm as discussed in the previous section. The loss increases as the particle size increases beyond the designed cut size of 1 μm (Fig. 5b).

The particle trajectories in Fig. 6b illustrate that 3 μm particles were indeed deposited on the wall at the 90-degree angle turn during the transition from the inlet section to the acceleration nozzle section. This led to the particle loss we saw in the early section. The tapered inlet allows the particles to make a smooth transition into the acceleration nozzle section from the inlet section; therefore, all 3 μm particles were able to be collected on the center region of the substrate without random spreading on the substrate (see Fig. 6d).

3. Experimental

The performance of the microsampler was evaluated, experimentally. We will describe the experimental setup in the following sections. Then, we will discuss the collector performance evaluation using polystyrene latex spherical particles.

3.1 Aerosolization of polystyrene latex (PSL) particles

Commercially available fluorescent polystyrene latex (PSL) particles (Fluoresbrite®, Yellow Green Microspheres) from Polysciences, Inc. were used. The nominal diameters of the four monodisperse PSL particles were 0.5 μm , 0.75 μm , 1.0 μm , and 3.0 μm . These geometric sizes were taken as reported by the

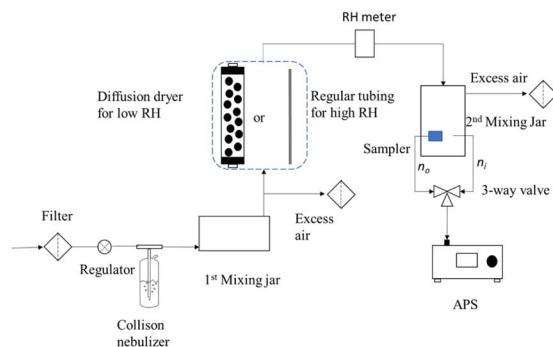


Fig. 7 Experimental setup for collection efficiency evaluation.

vendor, with typical coefficients of variation of 5–10%; sizes were not verified further in our lab. The sizes were benchmarked, however, using a TSI Model 3021 aerodynamic particle sizer (APS) in the aerosol state. Fluorescent particles were used in our study because the particle deposition pattern on the substrate in the sampler can be observed under UV light.

Fig. 7 shows the experimental setup for this study. The commercial fluorescent PSL suspension was diluted with distilled, deionized water ($18\text{ M}\Omega$) to approximately 10^7 particles per mL and atomized in a 3-jet Collison nebulizer. HEPA-filtered building air was supplied to a collison nebulizer to atomize the suspension to generate droplets. The relative humidity (RH) in the building air was generally 4%. 26-psi HEPA-filtered air was supplied to the collison nebulizer to generate droplets with a mass-median-diameter (MMD) of approximately $2.5\text{ }\mu\text{m}$ containing the PSL particles.²⁶ Therefore, the supply pressure was reduced to 20 psi to generate larger than $2.5\text{ }\mu\text{m}$ droplets that could effectively contain $3\text{ }\mu\text{m}$ solid PSL particles. Other than this, 26 psi supply pressure was used throughout for generating PSL particles of 0.5, 0.75, and $1\text{ }\mu\text{m}$.

The nebulized aerosol flow was then directed to a diffusion dryer (TSI Model 3062) to remove water from the PSL particles in the process of forming dry solid PSL test particles. A 1st mixing jar was used to trap excess water before the diffusion dryer, and also excess air was vented through a HEPA filter into a hood. A hygrometer (the RH meter in Fig. 7) was used to monitor the RH condition in the aerosol flow.

The aerosol particles after they pass the diffusion dryer were directed into a 2nd mixing jar that houses the microsampler. The number concentration measured using the APS in the 2nd mixing jar (n_i) was used as the input concentration, while the number concentration measured at the outlet of the microsampler was used as n_o for the collection efficiency calculations [eqn (2)]. Note that an APS pump was used to draw the aerosol particles into the 2nd mixing jar and through the microsampler if a 3-way valve was turned to permit flow through the microsampler.

Again, the diffusion dried PSL particles were directed into a 2nd mixing jar for number concentration and size measurements using the APS (TSI model 3321). As an example, the $3\text{ }\mu\text{m}$ PSL particle population showed an APS-measured size distribution as shown in Fig. 8 as the main peak. There was a smaller

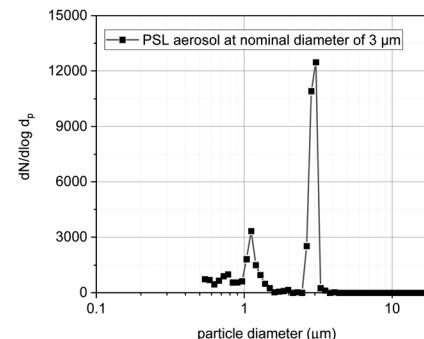


Fig. 8 Size distribution of the $3\text{ }\mu\text{m}$ PSL aerosol.

peak, with a number concentration of about $\frac{1}{4}$ the size of the main peak, located at the proximity of $1\text{ }\mu\text{m}$. We believe the small peaks could be attributed to the impurity of the prepared PSL suspension.

The pressure drop through the sampler was measured with a differential pressure gauge (Dwyer Instruments, Inc.) to be approximately 30.5 cm of water at 1 LPM. The flow rate Q_o of the microsampler was set at 1 LPM. The jar flow rate Q_i was slightly larger than Q_o , typically less than 1%. A correction factor (CF), Q_o/Q_i , was used to correct the calculated collection efficiency. However, the CF value is virtually identical to unity. By applying the correction factor CF, the experimental collection efficiency ($\eta_{\text{experimental}}$) was then computed as follows:

$$\eta_{\text{experimental}} = 1 - \frac{n_o}{n_i * \text{CF}} \quad (2)$$

3.2 Generation of mists

To investigate the performance of the microsampler under a misty condition, a separate experimental setup was established. A misty condition refers to atmospheric fog, drizzle, or rain drops. The collison nebulizer used in the previous experiments generates liquid droplets with sizes much smaller than those found in the natural environment,²⁷ ranging from 1 micron to a few millimetres, and at a number concentration, for example, 10^4 cm^{-3} .

We evaluated the impacts of mists with a different generation technique that allowed the production of larger liquid droplets in the size range that mimic those of natural liquid droplets in the atmosphere. Fig. 9 shows the general setup for producing super-micrometre droplets for the experiments using an ultrasonic generator (Sono-Tek Corp.) with the $0.5\text{ }\mu\text{m}$ fluorescent PSL particles.

The median droplet size (d_d) produced by the ultrasonic generator can be estimated from the following equation:²⁸

$$d_d = 0.34 \left(\frac{8\pi\sigma}{\rho f^2} \right)^{\frac{1}{3}}$$

where σ is the surface tension (N m^{-1}) of the liquid, ρ is the liquid density (kg m^{-3}), and f is the operating frequency of the ultrasonic device in Hz. Assuming that the generator was



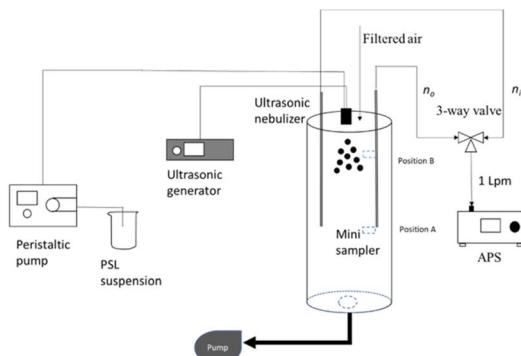


Fig. 9 Experimental setup for the effects of mists on the collection efficiency.

operated at 120 kHz, the median droplet size for water is estimated to be approximately 17 μm , which is in the right size range we were targeting.

Smaller droplets in the micron or submicron size range can rapidly evaporate, when mixed with dried purified air (water content at approximately 200 ppbv available from a ultradry air generator - the Parker Balston Model UDA-300 Compressed Air Dryer) and disappear, while the sizes of large droplets were reduced somewhat as they travel down the chamber. Two positions (labelled as position A and B in Fig. 9) were used to cursorily test the effects of different droplet sizes on the collector performance. Position A is further down in the chamber than position B; thus, evaporation of water droplets will make the liquid droplets smaller at position A than at position B, creating two size conditions for testing the interference of mists on the microsampler. Note that the tests were not designed to quantitatively evaluate the impacts on efficiency, but to learn if the impact exists.

4. Results and conclusions

4.1 Experimental evaluation of collection efficiency using dried PSL aerosol particles

This section reports the experimental collection efficiency curve and compares it to the CFD results. The nominal PSL size reported by the vendor was slightly different from the aerodynamic diameter measured by the APS. Nominal diameters of 0.75, 1.0, and 3.0 μm were measured as 0.81, 1.1, and 2.9 μm in aerodynamic diameter, respectively. The measured nominal diameters were within the 5–10% coefficient of variation for size distribution estimated by the vendor. The lower detection of the APS (TSI Model 3321) limits its ability to accurately measure the nominal 0.5 μm PSL particles. The APS was reported to have consistency issues between summing and correlated modes except for bin 1 (*i.e.*, < 0.523 μm).²⁹ Therefore, the counts in bins 2 and 3 (*i.e.*, 0.542 μm and 0.583) were selected and averaged to represent the particle counts at the size of 0.56 μm .

The experiments were conducted in the chamber under RH conditions of less than 30% at room temperature (*ca.* 20 °C). The predicted collection efficiency curve and collection efficiency calculated using experimental data for the microsampler are

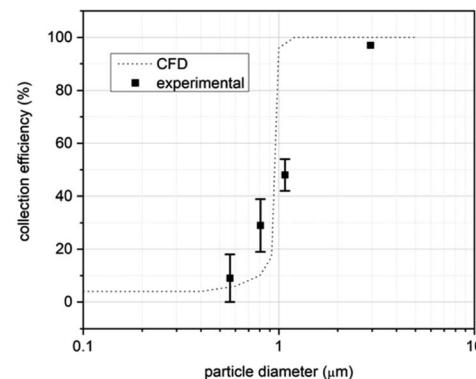


Fig. 10 Particle collection efficiency curve from CFD and experimental data for the microsampler at $\leq 30\%$ RH.

shown in Fig. 10. The testing results for 1.1 μm particles showed a collection efficiency of $48 \pm 6\%$ ($n = 18$), which was close to the designed cut size D_{p50} at 0.96 μm . For particles with an aerodynamic diameter of 2.9 μm , the collection efficiency was $97 \pm 0.3\%$ ($n = 5$). The collection efficiency for 0.56 μm particles was $9 \pm 9\%$ ($n = 4$) because of high penetration of particles smaller than the design cut size of 0.96 μm .

Although Fig. 10 also shows that the experimental data generally agree with the predicted collection efficiency curve, the experimental collection efficiency for the 0.81 μm particles was $29 \pm 10\%$ ($n = 22$) while 10% collection efficiency was predicted. The efficiency for 0.96 μm was predicted to be 50% while the experimental value was at $48 \pm 6\%$ for the APS diameter of 1.1 μm . Knowing the variation of 5–10% in the vendor's manufactured nominal size, we would argue that the experimental efficiency of the impactor agrees well with the design value.

4.2 The effect of high RH on the collection efficiency

Due to the small transport channel passage in this sampler, droplets in misty, foggy, or cloudy air could pose detrimental effects on the sampler's performance. Furthermore, if a significant temperature gradient exists between the ambient air and the interior transport channel in the sampler, water vapor could condense on the channel wall and form droplets, potentially blocking the passageway, and thereby degrading the performance of the sampler. Thus, investigation of the humidity effect on the sampler performance is warranted.

The experimental setup for the investigation of humidity effects was similar to that of the experiment described in Section 3.1. A regular stainless-steel tube (see Fig. 8) was used to replace the diffusion drier and maintain the RH in the aerosol stream at roughly 80%. Fig. 11 shows an example of the particle size distribution for the low ($\leq 30\%$) and high (80%) RH conditions. Under lower humidity conditions ($\leq 30\%$ RH), only one peak was observed for the particle size distribution, and all particles larger than a nominal size of approximately 0.5 μm were virtually unobservable. On the other hand, particles in the size range between 0.5 and 1.0 μm were observed using the APS under the conditions of 80% RH (see Fig. 11) indicating that the



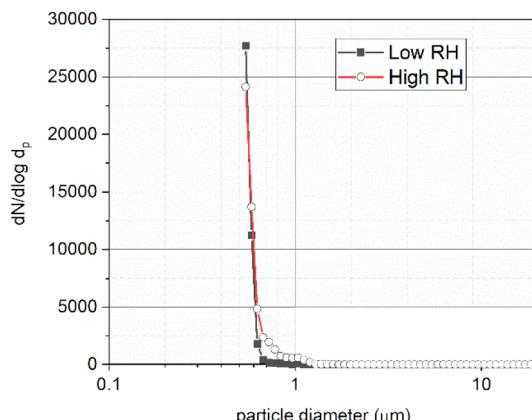


Fig. 11 Size distribution of the PSL aerosol at a nominal diameter of 0.5 μm for the low ($\leq 30\%$) and high (80%) RH conditions.

PSL particles were able to adsorb water vapor onto their surface thereby growing in size. Alternatively, the particles in the size range from 0.5 to 1.0 μm could simply be the water droplets that were unable to evaporate prior to sampling and measurement by the APS since the humidity condition in the aerosol stream was high (80% RH). These results demonstrate the aerosol population that exists under two different RH conditions given the same amount of PSL particles used in the suspension fed to the atomizer.

The collection efficiency of 0.5 μm PSL aerosols was compared under high RH conditions against the same system operated under low RH conditions. The collection efficiency was found to be $9 \pm 9\%$ ($n = 4$) for the low RH conditions, while the collection efficiency was $9 \pm 10\%$ ($n = 6$) for the high RH conditions. Therefore, the collection efficiency of this sampler for the 0.5 μm PSL particles appears to be statistically independent of atmospheric humidity up to 80% of our experimental boundary.

Furthermore, the computational results, as displayed in Fig. 5, show that the highest gas velocity inside the sampler was about 40 m s^{-1} , which is much slower than the speed of sound of 343 m s^{-1} at 20°C . Thus, it is highly unlikely that the effect of aerodynamic cooling exists in this microsampler.³⁰ We do not expect water vapor to condense inside the microsampler in a normal atmospheric environment where RH is less than 80%. The sampler should maintain a collection efficiency close to that predicted for environmental humidity conditions of less than or equal to 80% RH.

4.3 The effects of mists on the collection efficiency

In the previous sections, no condensed water exists in the aerosol stream in our experimental system. We now discuss the impacts of mists (condensed water) on the sampler performance in this section. APS counts for particles with 0.56 μm aerodynamic diameter were used to determine the collection efficiency. Recall that the collection efficiency of this microsampler for the particle size of 0.56 μm was $9 \pm 9\%$ in the absence of mists. The collection efficiency for the particles

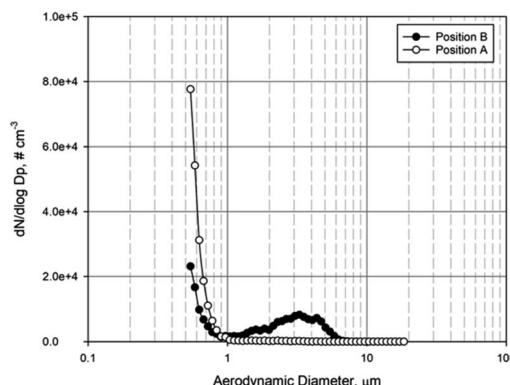


Fig. 12 The size distribution of test aerosol populations in the test chamber at two sampling locations (Note: (A) is the location closer to the chamber exit, while (B) is close to the inlet).

increased to $38 \pm 3\%$ ($n = 6$) at position B (farther away from the exit of the chamber), while the collection efficiency increased to 95% ($n = 2$) at position A (closer to the exit of the chamber). The same 0.56 μm PSL aerosol population has another subpopulation that is in the supermicron size range in a greater quantity at position B than that at position A as shown in Fig. 12. Thus, the results suggest that the sampler performance was impacted and biased by this subpopulation. In other words, the presence of condensed water in the air could significantly affect the sampling performance of this microsampler. We think the wetted wall reduced the size of the inlet and thus increased the travelling speed of the particles through the nozzle thereby artificially increasing the collection efficiency. Use of this microsampler thus should be restricted to conditions where there is no condensed water existing in the atmosphere; simply, no fog, no mist, and no rain in the air.

5. Conclusions

A microsampler for collection of aerosol particles on an aerial platform was designed and evaluated in the laboratory for dry (no condensed water) and wet (foggy, misty, or rainy) air conditions. The CFD results suggested a design that provided a higher sampling efficiency with minimal inlet loss. The microsampler was experimentally evaluated using fluorescent PSL particles at four nominal sizes. The experimental results at the four PSL particle sizes agree reasonably well with the collection efficiencies predicted by CFD. We found no impacts on the collection efficiency of the microsampler under conditions at 80% RH. However, when condensed water exists, biases on the collection efficiencies were found suggesting that this microsampler should not be used under conditions where there is condensed water in the atmosphere, e.g., foggy, misty, and rainy conditions.

Disclaimer

This manuscript has been authored by UT-Battelle, LLC, under contract DE-AC05-00OR22725 with the US Department of



Energy (DOE). The US government retains and the publisher, by accepting the article for publication, acknowledges that the US government retains a nonexclusive, paid-up, irrevocable, worldwide license to publish or reproduce the published form of this manuscript, or allow others to do so, for US government purposes. DOE will provide public access to these results of federally sponsored research in accordance with the DOE Public Access Plan (<https://energy.gov/downloads/doe-public-access-plan>).

Conflicts of interest

There are no conflicts to declare.

Acknowledgements

This research was supported by the Department of Defense/ Defense Threat Reduction Agency of the United States of America. Chris Boring is acknowledged for assistance in the construction of the experimental chamber and Bart Murry is acknowledged for translating the CFD configuration into the CAD drawing and 3D-printing fabrication of the prototype sampler. Oak Ridge National Laboratory is managed by UT-Battelle, LLC for the U.S. Department of Energy under contract DE-AC05-00OR22725.

Notes and references

- 1 T. Francesco Villa, F. Gonzalez, B. Milijevic, Z. D. Ristovski and L. Morawska, An overview of small unmanned aerial vehicles for air quality measurements: present applications and future prospectives, *Sensors*, 2016, **16**, 1072.
- 2 C. Misra, M. Singh, S. Shen, C. Sioutas and P. M. Hall, Development and evaluation of a personal cascade impactor sampler (PCIS), *J. Aerosol Sci.*, 2002, **33**, 1027–1047.
- 3 M.-D. Cheng, Selective collection of airborne particulate matter, *Aerosol Air Qual. Res.*, 2018, **18**(1), 361–365.
- 4 A. Townsend, I. N. Jiya, C. Martinson, D. Bessarabov and R. Gouws, A comprehensive review of energy sources for unmanned aerial vehicles, their shortfalls and opportunities for improvements, *Helijon*, 2020, **6**, e05285.
- 5 I. Paprotny, F. Doering, P. A. Solomon, R. M. White and L. A. Gundel, Microfabricated air-microfluidic sensor for personal monitoring of airborne particulate matter: Design, fabrication, and experimental results, *Sens. Actuators, A*, 2013, **201**, 506–516.
- 6 K. A. Koehler, J. Shapiro, Y. Sameenoi, C. Henry and J. Volckens, Laboratory evaluation of a microfluidic electrochemical sensor for aerosol oxidative load, *Aerosol Sci. Technol.*, 2014, **48**, 489–497.
- 7 T. Brubaker, M. Polen, P. Cheng, V. Ekambaram, J. Somers, S. L. Anna and R. C. Sullivan, Development and characterization of a “store and create” microfluidic device to determine the heterogeneous freezing properties of ice nucleating particles, *Aerosol Sci. Technol.*, 2020, **54**, 79–93.
- 8 F. Liu, N. L. Ng and H. Lu, Emerging applications of microfluidic techniques for *in vitro* toxicity studies of atmospheric particulate matter, *Aerosol Sci. Technol.*, 2021, **55**, 623–639.
- 9 F. Mei, and L. Goldberger, *Time-resolved Aerosol Filter Sampler Instrument Handbook, ARM report DOE/SC-ARM-TR-25*, Pacific Northwest National Laboratory, Richland, WA, 2020.
- 10 C. Cazzolara, M. Ebner, A. Platis, T. Miranda, J. Bange and A. Junginger, A new multicopter-based unmanned aerial system for pollen and spores collection in the atmospheric boundary layer, *Atmos. Meas. Tech.*, 2019, **12**, 1581–1598.
- 11 P. Bieber, T. M. Seifried, J. Burkart, J. Gratzl, A. Kasper-Giebel, D. G. Schmale and H. Grothe, A drone-based bioaerosol sampling system to monitor ice nucleation particles in the lower atmosphere, *Remote Sens.*, 2020, **12**, 552.
- 12 J. Schrod, D. Weber, J. Drücke, C. Keleshis, M. Pikridas, M. Ebert, B. Cvetković, S. Nickovic, E. Marinou, H. Baars, A. Ansmann, M. Vrekoussis, N. Mihalopoulos, J. Sciare, J. Curtius and H. G. Bingemer, Ice nucleating particles over the Eastern Mediterranean measured by unmanned aircraft systems, *Atmos. Chem. Phys.*, 2017, **17**, 4817–4835.
- 13 T.-C. Hsiao, D.-R. Chen and S. Y. Son, Development of mini-cyclones as the size-selective inlet of miniature particle detectors, *J. Aerosol Sci.*, 2009, **40**, 481–491.
- 14 I. V. Novoselov, R. A. Gorder, J. A. Van Amberg and P. C. Ariessohn, Design and performance of a low-cost micro-channel aerosol collector, *Aerosol Sci. Technol.*, 2014, **48**, 822–830.
- 15 B. Damit, Droplet-based microfluidics detector for bioaerosol detection, *Aerosol Sci. Technol.*, 2017, **51**(4), 488–500.
- 16 A. R. Metcalf, S. Narayan and C. S. Dutcher, A review of microfluidic concept and applications for atmospheric aerosol science, *Aerosol Sci. Technol.*, 2018, **52**, 310–329.
- 17 V. A. Marple, B. A. Olson and N. C. Miller, A low-loss cascade impactor with stage collection cups: calibration and pharmaceutical inhaler applications, *Aerosol Sci. Technol.*, 1995, **22**, 124–134.
- 18 V. A. Marple, D. L. Roberts, F. J. Romay, N. C. Miller, K. G. Truman, M. Van Oort and D. Hochrainer, Next generation pharmaceutical impactor (a new impactor for pharmaceutical inhaler testing). Part I: design, *J. Aerosol Med.*, 2003, **16**, 283–299.
- 19 V. A. Marple and B. Y. H. Liu, Characteristics of laminar jet impactors, *Environ. Sci. Technol.*, 1974, **8**, 648–654.
- 20 V. A. Marple and K. Willeke, *Impactor design, Atmos. Environ.*, 1976, **10**, 891–896.
- 21 V. A. Marple, K. L. Rubow and S. M. Behm, A microorifice uniform deposit impactor (MOUDI): description, calibration, and use, *Aerosol Sci. Technol.*, 1991, **14**, 434–446.
- 22 B. Jurcik and H. C. Wang, On the shape of impactor efficiency curves, *J. Aerosol Sci.*, 1995, **26**, 1139–1147.
- 23 S. Hari, Y. A. Hassan and A. R. McFarland, Computational fluid dynamics simulation of a rectangular slit real impactor's performance, *Nucl. Eng. Des.*, 2005, **235**, 1015–1028.



24 J. Tu, G. H. Yeoh, and C. Liu, *Computational Fluid Dynamics: a Practical Approach*, Butterworth-Heinemann, Elsevier, Ltd, Oxford, UK, 3rd edn, 2018.

25 S. K. Friedlander, *Smoke, Dust, and Haze: Fundamentals of Aerosol Dynamics*, Oxford University Press, New York, 2nd edn, 2000.

26 K. R. May, The collision nebulizer: description, performance and application, *J. Aerosol Sci.*, 1973, **4**, 235–243.

27 J. H. Seinfeld, and S. N. Pandis, *Atmospheric Chemistry and Physics: from Air Pollution to Climate Change*, John Wiley & Sons, 2012.

28 S. M. Mahurin and M.-D. Cheng, Generating nanoscale aggregates from colloidal nanoparticles by various aerosol spray techniques, *Nanotox*, 2009, **1**, 130–138.

29 T. M. Peters and D. Leith, Concentration measurement and counting efficiency of the aerodynamic particle sizer 3321, *J. Aerosol Sci.*, 2003, **34**, 627–634.

30 P. Biswas, C. L. Jones and R. C. Flagan, Distortion of size distributions by condensation and evaporation in aerosol instruments, *Aerosol Sci. Technol.*, 1987, **7**, 231–246.

