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Improving model representation of rapid ozone deposition over soil in the central Tibetan Plateau†

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Ozone soil deposition contributes a major part to the total deposition of ozone on land covered by low vegetation and perturbs the ozone budget on both regional and global scales. Large model-observation divergences in ozone soil deposition require continuous efforts to improve the mechanical understanding and model representation. Observation of ozone deposition over bare soil directly meets the requirement. Here, we performed field observation of ozone deposition over bare soil first available in the Tibetan Plateau (TP) using the aerodynamic gradient method. A top ozone deposition velocity with a daily mean of 0.49 ± 0.11 (1 sd) cm s^{-1} (1 May to 10 July 2019) and an hourly mean maximum across the diel pattern of 0.73 ± 0.67 cm s^{-1} in the afternoon were recorded. Such rapid ozone deposition was mainly attributed to extremely low soil resistance (R_{soil}), which was further regulated by median low soil clay content, dry conditions, and strong solar radiation in the central TP. Parameterization of R_{soil} in the newly developed Stella scheme was demonstrated to be effective according to our verification. An updated scheme was further attained with the inclusion of our observation and better represents the R_{soil} variability than the Stella scheme. More verification is therefore encouraged and hopefully to improve the Stella scheme. Finally, both the Stella scheme and our updated scheme showed great advantages over the oversimplified scheme in current models and should be considered more seriously for the sake of better representation of ozone soil deposition and its variability.

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Environmental significance

Tropospheric ozone is an important gaseous pollutant and a short-lived climate forcer that affects air quality, human health and climate. Ozone soil deposition represents a non-negligible fraction of ozone budget. However, it is highly underestimated by current models and therefore leads to ill-representation of ozone distribution. To explore the soil deposition mechanism and improve the model representation, we performed direct measurements of ozone deposition velocity on bare soil in the central Tibetan Plateau (TP), an ideal experimental field for its pristine nature in terms of weak human perturbation on both land use and ozone photochemistry. For the first time, our measurement recorded in the TP the top ozone deposition velocity on soil, which could be accounted for by measured meteorological and edaphic parameters. An updated parameterization scheme based on Stella *et al.* was summarized and better represented ozone deposition velocity and its environmental variability, relative to the schemes commonly used in chemical transport models. The Stella scheme is therefore worthy of more serious consideration for the sake of better representation of ozone soil deposition and its variability. Our data has also confirmed that ozone soil deposition in the TP is underestimated by current model evaluations.

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1 Introduction

Tropospheric ozone is an important gaseous pollutant and a short-lived climate forcer that plays a key role in atmospheric oxidative capacity, oxidative damage to humans and vegetation, and climate change.^{1–4} Obtaining accurate temporal and spatial variability of ozone is essential for assessing its role. While dense ozone measurement networks have been established in the United States, Europe, and eastern Asia, and great progress has been made in China in recent years, observational constraints on ozone are still absent in other regions, especially in background areas globally.⁵ Chemical transport models (CTMs) are alternative tools to evaluate the temporal–spatial



variability in ozone in those regions where measurement is absent. Simulations of ozone by CTMs are numerical approximations of the budgets of ozone, including chemical reactions that lead to ozone formation and destruction, tropospheric influx from stratosphere–troposphere exchange, deposition loss, and convective and advective transport. However, the simulation representations are limited by our current incomplete understanding of and parameterization methods for these budget items.⁶

Simulation representation of tropospheric ozone, especially ozone at the surface, is sensitive to the representation of ozone dry deposition.^{7–11} The resistance-analogy parameterization framework for ozone deposition developed by Wesely,¹² is widely accepted by different deposition schemes to explain and parameterize ozone deposition on various land cover types.^{13–18} Briefly, the resistance analogy framework conceptualized the dry deposition of ozone or other pollutants as a three-step process including (1) ozone transport downward to a given surface by atmospheric turbulence, (2) mass transfer through the quasi-laminar sublayer around the interface of air and the surface, and (3) uptake on the surface. The ozone deposition velocity, v_d , can be expressed as the reciprocal of the total mass transfer resistance of these three-step processes, namely, aerodynamic resistance R_a , quasi-laminar sublayer resistance R_b , and surface resistance R_c .¹²

Studies comparing different deposition schemes at site-specific sites illustrate that the representation of R_c and its variability cause the largest differences of v_d .^{19,20} Among different deposition schemes, R_a is typically modelled with different empirical formulae of the Monin–Obukhov similarity theory (MOST), and R_b is modelled with a same formula. Wu *et al.*¹⁹ compared $R_a + R_b$ over a forest modelled by four different MOST-based models and found similar results. Combined with the relatively smaller contribution of $R_a + R_b$ than R_c to the total resistance (see Section 3.2 below), more attention is focused on the parameterization of R_c . The parameterization of R_c used in the resistance analogy considers the resistance of ozone uptake by stomatal, cuticular, and ground surfaces, the last of which mainly refers to soil surfaces. Relatively intensive field observations of ozone deposition have been performed over developed canopies.^{21–28} Canopy related stomatal and cuticular uptake are the main ozone deposition paths in these observations. However, studies on crops found that the determining role of stomatal and cuticular resistances highly depends on canopy development and abundance.^{29,30} On surfaces with lower and sparser canopies, such as growing croplands and grasslands, soil resistance (R_{soil}) could contribute up to 55% to total ozone deposition resistance; thus, v_d is more sensitive to R_{soil} for these land cover types.^{31,32}

Ozone soil deposition contributes heavily to the global budget of ozone deposition because a considerable fraction of global land cover is desert, bare soil, and sparse vegetation. However, the modeling of ozone soil deposition remains highly uncertain. Gross model underestimation of ozone deposition on bare soil (equivalent to ozone soil deposition) was consistently found.^{33,34} The large discrepancy of ozone soil deposition and its variability between measurements and simulations

implies the poor representations of ozone soil deposition by current CTMs. The poor representation is mainly attributed to the oversimplified R_{soil} parameterization scheme that uses two-level prescribed values for dry and wet surfaces without more complex response relationships with soil moisture, soil temperature or soil clay content.¹⁹ Modification attempts have been made with the aim of enabling model power to better describe the variability in R_{soil} . Either a positive dependence of R_{soil} on soil moisture or a negative dependence on temperature was reported based on site-specific observations, but the expressions between R_{soil} and moisture or temperature should not be extrapolated to other sites.^{35–37} A few studies introduced a universal corrected function of soil moisture or temperature based on mathematical assumptions; however, field observation validation is rare.^{7,38} Stella *et al.*^{33,34} summarized their six field observations over bare soil and proposed a dual-parameter semiempirical R_{soil} parameterization (known as the “Stella scheme”) involving soil clay content and surface relative humidity (RH_{surf}). In the Stella scheme, RH_{surf} is believed to better represent the varied and complex effect of meteorological parameters, such as surface temperature (T_{surf}), soil moisture, and solar radiation. Soil clay content, as a second influencing factor, reasonably accounts for the spatial heterogeneity of soil uptake reactivity toward ozone and therefore R_{soil} .

Although the Stella scheme seems to have the potential to describe R_{soil} and its variability, limited field observational validation hinders direct extrapolation or CTMs inclusion. The observations of ozone soil deposition were mostly conducted over agricultural fields after harvest to avoid interference of the canopy.³³ However, frequent agricultural activities, such as plowing, fertilization, and irrigation, can cause transformations in surface soil physics and chemical characteristics and potentially soil uptake reactivity toward ozone. Additionally, strong NO_x and/or VOCs emissions from agricultural soil can participate in the fast atmospheric chemical conversion of ozone and therefore significantly interfere with the measurement of ozone soil deposition.^{42,43} Field observation of ozone deposition in environments with negligible NO or highly reactive VOCs sources is ideal in terms of less chemistry interference and human activity perturbation of soil but is sporadically available.³³ In addition, R_{soil} is highly affected by soil clay contents and hydrothermal conditions based on the previous summary, which are spatially heterogeneous in different climate regions. Previous observations were performed mostly in subtropical climate regions. Field observations in more climate regions are therefore highly demanded (Fig. 1). The TP is a typical highland climate region where both soil clay content and moisture are expected to be different from the subtropical climate regions. And TP is characterized as a highly pristine area and covered with sparse and low vegetation, implying weak perturbation of land use and ozone chemistry on ozone soil deposition.

The @Tibet series field campaign supports ozone deposition measurement over the Tibetan Plateau (TP). In this study, we took advantage of the ideal experimental field and performed the first measurement of ozone deposition velocity on bare soil in the TP, acquired from the aerodynamic gradient method. The



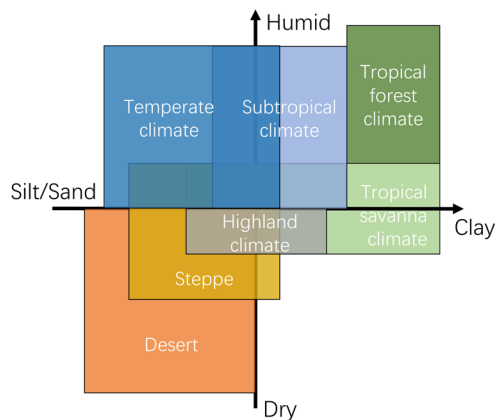


Fig. 1 Simplified concept of climatic regions with different soil clay contents and hydrothermal conditions. Previous measurements of soil resistance (R_{soil}) have been conducted mostly in subtropical climate regions in the figure. Data of climate classification is from Chen *et al.*³⁹ Soil clay content data is from Harmonized World Soil Database v2.0.⁴⁰ Humidity data is referenced to Kummu *et al.*⁴¹

aims of this study are (i) to measure v_d in the TP and to identify the factors controlling v_d and R_{soil} and (ii) to evaluate and improve the Stella scheme by inclusion of our verifications.

2 Methods

2.1 Measurements

The field measurement of ozone deposition was performed from 1 May to 10 July 2019 at Nam Co Comprehensive Observation and Research Station, Chinese Academy of Sciences (CAS) (NMC site, 30°46.30'N, 90°59.31'E, 4730 m a.s.l.), as an essential part of the @Tibet 2019 campaign. Due to power outages, instrument malfunctions and other factors, the valid data cover 50 days. Detailed descriptions of NMC site and @Tibet 2019 campaign can be found in previous studies.^{44–46} NMC site is located in an open flat field ~1 km southeast of Nam Co Lake and more than 15 km to the north and west of the Nyainqêntanglha Mountains. In addition, a sublake of Nam Co Lake is ~280 m away from the site. NMC site is characterized by a continental background of the atmosphere with few surrounding populations, and isolated from Lhasa, the largest city in the TP.

High-altitude and mountainous environment results in a typical arctic–highland climate of this area. The monthly mean air temperature ranges from -7.8 °C to 12.2 °C, and solar radiation reaches 1200 $W\ m^{-2}$ in May.⁴⁷ Annual precipitation of approximately 400 mm is concentrated during the Asian Summer Monsoon (ASM) period from June to September.⁴⁸ In this cold and arid area, the soil is poorly developed because of sporadic vegetation coverage and weak chemical weathering, which is embodied in a loose soil structure and low clay content.⁴⁹

Air temperature and relative humidity (HMP155A, Vaisala, FI) and wind speed (010C, Met One, USA) profiles were measured at 1.8, 3.8, and 5.8 m height, and wind direction (020C, Met One, USA) was measured at 5.8 m. Solar radiation,

including incoming and reflected shortwave and longwave radiation (CNR4, Kipp & Zonen, NL) and $j(NO_2)$ (ultrafast CCD-detector spectrometer, Metcon, GER), were measured. Soil temperature and soil water content (CS655, Campbell Scientific Inc., USA) profiles were also measured at 5, 15, and 25 cm below ground. NO_2 concentration was measured by a sensitive incoherent broadband cavity enhanced absorption spectroscopy NO_2 analyzer.⁵⁰ As the NO concentration was too low to be measured, photostationary state calculation of the NO concentration (NO_{PSS}) was carried out with measured O_3 , NO_2 , and $j(NO_2)$.

Ozone flux was measured using the aerodynamic gradient (AG) method, with characterization of the micrometeorological environment by the eddy covariance (EC). The AG method is based on K -theory, an application of MOST. Similar to Fick's Law, the K -theory assumes that the turbulence flux can be expressed as the product of the turbulence exchange coefficient K and vertical concentration gradient. Therefore, the ozone flux can be calculated as follows:

$$F = -K \frac{d[O_3]}{dz} \quad (1)$$

where F is the ozone flux ($ppbv\ m\ s^{-1}$ or $nmol\ m^{-2}\ s^{-1}$), K is the turbulence exchange coefficient ($m^2\ s^{-1}$), and $\frac{d[O_3]}{dz}$ is the vertical ozone gradient between the two measurement heights ($ppbv\ m^{-1}$ or $nmol\ m^{-2}$). A UV photometric ozone analyzer (Model 49i, Thermo-Environmental Instruments Inc., USA) measured the ozone concentration. Air was sampled at 6.8 m and 1.8 m of the tower through a 27 m Teflon sampling tube with sequential sampling in a 30 min duty cycle. A solenoid valve was used to control the measurement time of 15 min at each height. The measurement gap of 15 min at each height was filled by linear interpolation of adjacent data at the same height. Therefore, continuous ozone gradient was obtained. The sampling tube was covered by black foam for heat preservation and light shading. K is calculated as follows:⁵¹

$$K = \kappa u_* (z_2 - z_1) \left[\ln \frac{z_2 - d}{z_1 - d} - \psi_h \left(\frac{z_2 - d}{L} \right) + \psi_h \left(\frac{z_1 - d}{L} \right) \right] \quad (2)$$

where κ is von Karman's constant (0.4), u_* is the friction velocity ($m\ s^{-1}$), z_1 and z_2 are the lower and greater heights of ozone measurement respectively (m), d is the zero-plane displacement height (m), L is the Obukhov length (m), and $\psi_h(z/L)$ is the integrated stability correction function for heat:

$$\psi_h(z/L) = \begin{cases} 2 \ln \left(\frac{1+y}{2} \right), & z/L < 0 \\ -5z/L, & z/L > 0 \end{cases} \quad (3)$$

where $y = (1 - 16z/L)^{1/2}$. As EC is a reliable method to characterize atmospheric turbulence, u_* and L were provided by the EC method and used in the AG method. The EC method was also used to measure fluxes of momentum, sensible heat (H), latent heat (LE), water vapor, and CO_2 . The EC system includes an integrated CO_2/H_2O open-path gas analyzer and 3D sonic anemometer (IRGASON, Campbell Scientific Inc., USA), electronics for synchronizing gas and wind data (EC100, including



to the premonsoon period, but it still remained at a high level compared with regions of the same latitude. Volumetric soil water content (SWC) zoomed from 0.01 to 0.2 after 1 July. Weakened R_g and enhanced SWC in the ASM period caused a slight decrease in T_a and a significant increase in RH_a . WS showed obvious differences from the premonsoon period, with the peak dropping, but WS mean at night increased.

Diel pattern of key parameters that related to calculation of R_a , R_b , R_{soil} and derivation of v_d according to eqn (6)–(12) and (17) are performed in Fig. 2. Over the entire observation period, the hourly mean T_a ranged from 1.6 ± 4.6 °C at night to 10.6 ± 4.3 °C in the daytime (Fig. 2b). Significant enhancements of T_{surf} compared to T_a occurred with a maximal hourly enhancement of 15.6 °C during daytime. This was caused by strong R_g , of which the hourly mean maximal reached 1252.4 ± 348.5 W



Fig. 2 Diel pattern of meteorological parameters and edaphic parameters associated with ozone deposition velocity derivation and deposition resistance derivation. Diel pattern of (a) solar radiation (R_g), (b) air temperature at 1.8 m (T_a), surface temperature (T_{surf}) and soil temperature at -5 cm (T_{soil}), (c) air relative humidity at 1.8 m (RH_a), surface relative humidity (RH_{surf}) and soil water content (SWC), (d) wind speed at 1.8 m (WS) and wind direction (WD), (e) friction velocity (u_*), (f) atmospheric stability criterion (z/L), (g) turbulence exchange coefficient (K), (h) gradient of ozone between 6.8 m and 1.8 m ($d[O_3]/dZ$), ozone concentration at 6.8 m and ozone concentration at 1.8 m, (i) ozone flux (F), and (j) ozone deposition velocity (v_d). The solid line is the diel mean, and the shading is variation of our observations.



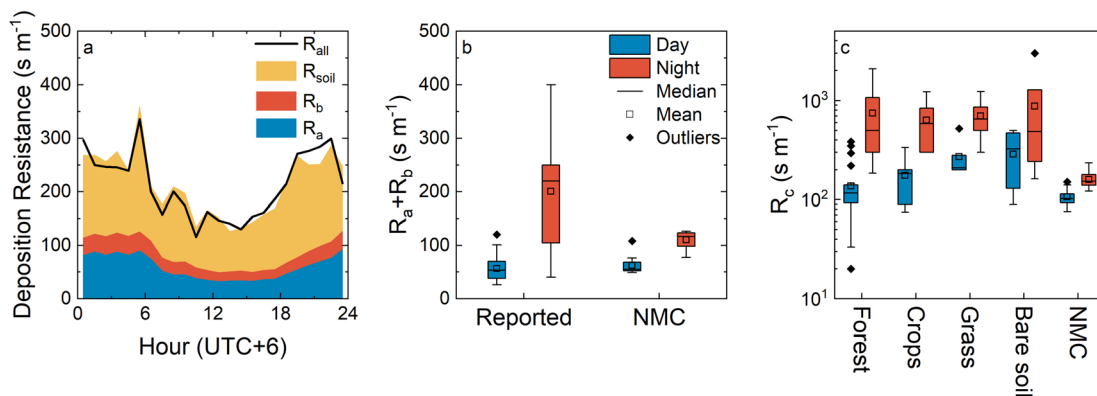


Fig. 3 (a) Median diel pattern of deposition resistance of observation at NMC site. Hourly medians are utilized instead of hourly means due to the skewed distribution of R_{all} and R_{soil} . (b) Comparison of $R_a + R_b$ and reported $R_a + R_b$ during day and night. (c) Comparison of observed R_{soil} and reported R_c over different conditions during day and night. Parts of reported R_c are obtained by $1/v_d$ minus prescribed $R_a + R_b$ (50 s m^{-1} during daytime and 200 s m^{-1} during nighttime). Data from: ^{24,26,58–67}

two mechanisms account for a positive dependence of R_{soil} on moisture as widely observed in field observations and laboratory experiments.²⁷ Our observation at NMC site also showed a significant positive dependence of R_{soil} on RH_{surf} (Fig. 4a). The dry climate at NMC site is conducive to keeping R_{soil} small.

Temperature is a positive-going parameter for ozone uptake, as the reaction of ozone on the surface is endothermic. Thus, R_{soil} decreases with increasing temperature. The mechanical deduction requires an Arrhenius-like response of R_{soil} to temperature, which is verified in field observations but with large fitting uncertainty and variability.^{34–36} The negative relationship between R_{soil} and temperature was also observed at NMC site (Fig. 4b). Notably, the dependence of R_{soil} on temperature could be a misconception due to the highly negative correlation between temperature and moisture. As a result, it is difficult to distinguish the individual effect of temperature from RH_{surf} on R_{soil} . Nevertheless, high T_{surf} is conducive to lower R_{soil} .

As shown in Fig. 5a and b, R_{soil} can be described by an exponential equation of RH_{surf} and an Arrhenius-like function of T_{surf} (eqn (20) and (21)):

$$R_{\text{soil}} = 71.0 \times e^{0.012 \times \text{RH}_{\text{surf}}} \quad (20)$$

$$R_{\text{soil}} = 0.52 \times e^{\frac{12850}{RT_{\text{surf}}}} \quad (21)$$

These two parameterization schemes were tested by simulating v_d with observation-constrained meteorological data by employing eqn (5)–(7) (Fig. 5c). Similar results of v_d are obtained from the two parameterization schemes and both results are closely comparable to the observed v_d in terms of both the magnitude and the temporal variability. Stella *et al.*³⁴ have found that the parameterization scheme with RH_{surf} performed more robustly than the parameterization scheme with T_{surf} . A potential reason for this could be that RH_{surf} itself contains the influence of T_{surf} due to the high correlation between RH_{surf} and T_{surf} , as previously mentioned. Herein, we choose the

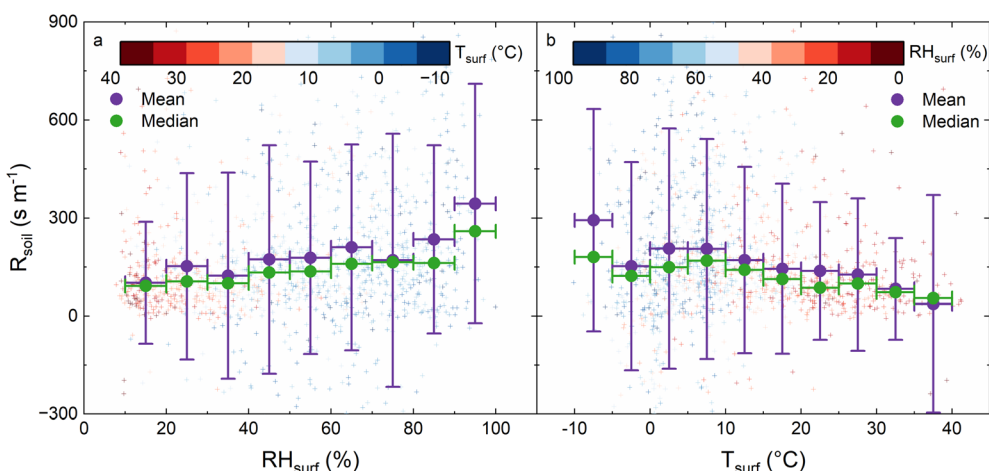


Fig. 4 Dependence of R_{soil} on (a) RH_{surf} and (b) T_{surf} . The block processing ranges of means and medians of RH_{surf} and T_{surf} are 10% and 5 °C, respectively.





Fig. 5 R_{soil} as a function of (a) RH_{surf} and (b) T_{surf} . Dots are block median data with ranges of 10% for RH_{surf} and 5 °C for T_{surf} . (c) Comparison of observed v_d and simulated v_d using R_{soil} as a function of RH_{surf} and T_{surf} with input meteorological data from observations. The shading area for observation represents 1 sd of observed v_d . Time alignments are performed on the data used for comparison.

parameterization scheme with RH_{surf} for further discussion. According to eqn (20), R_{soil} is predicted to be lower than 240 s m^{-1} at NMC site even at a high RH_{surf} range, which explains the rapid ozone deposition at night. In addition, strong solar radiation and dry conditions during the daytime maintain high T_{surf} and low RH_{surf} , which are also conducive to low R_{soil} .

Block medians instead of block means of R_{soil} were chosen to fit the function among R_{soil} and RH_{surf} (or T_{surf}) in this study, since R_{soil} at NMC was positively skewed. The fitting function based on block medians turned out to better reproduce our observations than that based on block means (Fig. 5 and S4†). As a matter of fact, R_{soil} is consistently found to be positively skewed.^{26,65,67,68} Thus, it needs more attention about whether the function among R_{soil} and RH_{surf} (or T_{surf}) are fitted with block medians or block means. Block means of R_{soil} were used in the establishment of the Stella scheme,^{33,34} which should be further examined.

3.4 Evaluation and improvement of the Stella scheme

RH_{surf} explains the variability of R_{soil} at specific sites, while soil clay content determines the site-by-site variability by affecting $R_{\text{soil min}}$ and k . According to the fitted results in eqn (20), $R_{\text{soil min}}$ and k of NMC are 71.0 s m^{-1} and 0.012, respectively. We applied the Stella scheme reversely (eqn (17)–(19)) and derived $R_{\text{soil min}} = 51.0 \text{ s m}^{-1}$ and $k = 0.017$ at NMC. Compared with the observed $R_{\text{soil min}}$ (71.0 s m^{-1}) and k (0.012), the Stella scheme underestimated $R_{\text{soil min}}$ by 28% and overestimated k by 42% (red dots in Fig. 6a and b). By including our observations on $R_{\text{soil min}}$ and k , we updated the Stella scheme (hereafter called the updated Stella scheme, eqn (17), (20) and (21)). To be noted, observations at La Crau were not included fitting in both the Stella scheme and the updated Stella scheme due to the reason mentioned in Stella *et al.*,³³ where the reported soil clay content is not representative.

$$R_{\text{soil min}} = 661 \times (\text{clay content})^{-0.86} \quad (22)$$

$$k = 0.0093 e^{0.0325 \times (\text{clay content})} \quad (23)$$



Fig. 6 Relationships between soil clay content and two coefficients in the Stella scheme, (a) minimum soil resistance ($R_{\text{soil min}}$) and (b) k . The black dots are values collected from previous observations, and the red dots are values from our observation at NMC site. The Stella scheme are shown as black lines, and the updated Stella scheme are shown as red lines. (c) Comparison of observed v_d and simulated v_d using the RA_SS and resistance analogy with the updated Stella scheme (RA_uSS) with input meteorological data from observations. Time alignments are performed on the data used for comparison.



- 52 M. Mauder, M. Cuntz, C. Drüe, A. Graf, C. Rebmann, H. P. Schmid, M. Schmidt and R. Steinbrecher, *Agric. For. Meteorol.*, 2013, **169**, 122–135.
- 53 N. Kljun, P. Calanca, M. W. Rotach and H. P. Schmid, *Bound.-Layer Meteorol.*, 2004, **112**, 503–523.
- 54 S. Walton, M. W. Gallagher and J. H. Duyzer, *Atmos. Environ.*, 1997, **31**, 2915–2931.
- 55 V. Wolff, I. Trebs, T. Foken and F. X. Meixner, *Biogeosciences*, 2010, **7**, 1729–1744.
- 56 J. Rinne, J.-P. Tuovinen, T. Laurila, H. Hakola, M. Aurela and H. Hypén, *Agric. For. Meteorol.*, 2000, **102**, 25–37.
- 57 A. Ito and R. Wagai, *Sci. Data*, 2017, **4**, 170103.
- 58 S. Cieslik, *Environ. Pollut.*, 2009, **157**, 1487–1496.
- 59 R. Wada, S. Yonemura, A. Tani and M. Kajino, *J. Agric. Meteorol.*, 2023, **79**, 38–48.
- 60 D. Fowler, C. Flechard, J. N. Cape, R. L. Storeton-West and M. Coyle, *Water, Air, Soil Pollut.*, 2001, **130**, 63–74.
- 61 G. Gerosa, R. Marzuoli, S. Cieslik and A. Ballarin-Denti, *Atmos. Environ.*, 2004, **38**, 2421–2432.
- 62 T. P. Meyers, P. Finkelstein, J. Clarke, T. G. Ellestad and P. F. Sims, *J. Geophys. Res.: Atmos.*, 1998, **103**, 22645–22661.
- 63 M. Michou, P. Laville, D. Serça, A. Fotiadi and V.-H. Peuch, *Atmos. Res.*, 2005, **74**, 89–116.
- 64 T. N. Mikkelsen, H. Ro-Poulsen, M. F. Hovmand, N. O. Jensen, K. Pilegaard and A. H. Egeløv, *Atmos. Environ.*, 2004, **38**, 2361–2371.
- 65 C. A. Pio, M. S. Feliciano, A. T. Vermeulen and E. C. Sousa, *Atmos. Environ.*, 2000, **34**, 195–205.
- 66 M. Zapletal, P. Cudlín, P. Chroust, O. Urban, R. Pokorný, M. Edwards-Jonášová, R. Czerný, D. Janouš, K. Taufarová, Z. Večeřa, P. Mikuška and E. Paoletti, *Environ. Pollut.*, 2011, **159**, 1024–1034.
- 67 L. Zhang, J. R. Brook and R. Vet, *Atmos. Environ.*, 2002, **36**, 4787–4799.
- 68 L. Zhang, M. D. Moran, P. A. Makar, J. R. Brook and S. Gong, *Atmos. Environ.*, 2002, **36**, 537–560.
- 69 K. M. Willett, R. J. H. Dunn, P. W. Thorne, S. Bell, M. de Podesta, D. E. Parker, P. D. Jones and C. N. Williams Jr, *Clim. Past*, 2014, **10**, 1983–2006.
- 70 A. Smith, N. Lott and R. Vose, *Bull. Am. Meteorol. Soc.*, 2011, **92**, 704–708.

