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# Bisphenol A (BPA) toxicity assessment and insights into current remediation strategies

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Bisphenol A (BPA) raises concerns among the scientific community as it is one of the most widely used compounds in industrial processes and a component of polycarbonate plastics and epoxy resins. In this review, we discuss the mechanism of BPA toxicity in food-grade plastics. Owing to its proliferation in the aqueous environment, we delved into the performance of various biological, physical, and chemical techniques for its remediation. Detailed mechanistic insights into these removal processes are provided. The toxic effects of BPA unravel as changes at the cellular level in the brain, which can result in learning difficulties, increased aggressiveness, hyperactivity, endocrine disorders, reduced fertility, and increased risk of dependence on illicit substances. Bacterial decomposition of BPA leads to new intermediates and products with lower toxicity. Processes such as membrane filtration, adsorption, coagulation, ozonation, and photocatalysis have also been shown to be efficient in aqueous-phase degradation. The breakdown mechanism of these processes is also discussed. The review demonstrates that high removal efficiency is usually achieved at the expense of high throughput. For the scalable application of BPA degradation technologies, removal efficiency needs to remain high at high throughput. We propose the need for process intensification using an integrated combination of these processes, which can solve multiple associated performance challenges.

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## 1. Introduction

Society's current consumption model stands out for the growing consumption of industrial compounds in processes that use epoxy resins and polycarbonate plastics, which contain bisphenol A (BPA).<sup>1,2</sup> Among the products consumed around the world are beverage and food packaging, toys, and medical devices.<sup>3,4</sup> These products are consumed in most cultures and social classes encompassing millions of people, resulting in an exponential increase in the release of BPA into the environment, raising concerns about its toxic effects throughout the food chain.<sup>5</sup> Its chemical and physical characteristics are similar to those of hormones and can easily exhibit effects close to estrogen.<sup>6</sup> Owing to this behavior, BPA can generate disturbances in the endocrine system, affecting the regulation and production of hormones.<sup>7</sup>

Other studies also relate its toxic effects to the formation of cancer cells, increased obesity rates, decreased fertility, and propensity for diabetes.<sup>8</sup> These issues have raised interest in eliminating or reducing BPA exposure in various products consumed worldwide. The major concern is mainly linked to its presence in beverage and food packaging, which facilitates leaching into soil and water resources.<sup>9</sup> Transport is favored when products are exposed to higher temperatures or even acidic compounds; in addition to this, the risk of contact is intensified with prolonged and continuous use of packaging.<sup>10</sup> Owing to the damage caused by BPA, the search for new alternatives that aim to replace the product in terms of production has been enhanced and encouraged.<sup>11</sup> The concern also involves the environment, considering its contamination in water and soil, which is intensified by inadequate disposal of various packaging, corroborating its leaching in landfills.<sup>12,13</sup>

Because of its characteristics such as transparency, robustness, and heat resistance, BPA is preferably used in the production of various food-grade plastic containers.<sup>14</sup> The insecurity in using this product is its ability to move into products for human consumption, which leads to constant exposure of living organisms.<sup>15,16</sup> The application of BPA in food-grade plastic packaging has been widely studied and discussed around the world.<sup>17</sup> It is worth noting that although some studies affirm unlikely toxicity at low concentration levels,

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others state that continued exposure over the years causes serious damage, making it highly necessary to substitute new, more environmentally friendly alternatives in place of BPA in food-grade plastics.<sup>18–21</sup>

It was in the 1950s that the commercialization of BPA began in the United States (USA) and then spread throughout Europe.<sup>22</sup> The increase in its production is currently being led by China with a growth rate of approximately 5%, while in Asia it increased by 13% in the early 2000s until 2006.<sup>23</sup> In the USA alone, in the period from 2004 to 2006, around 1 million tons were generated, a number close to that generated in Western Europe.<sup>24</sup> In the USA, BPA is called a toxic chemical with a large volume of use.<sup>25,26</sup> It is worth noting that in poorer countries such as Africa and Latin America, these compounds do not have defined and established guidelines, even though it has been

observed in studies that in countries such as Nigeria (Africa) the same was detected in several sources, highlighting the high risk of exposure in humans.<sup>27</sup> In 2011, the use of 5.5 million tons of BPA was recorded worldwide.<sup>28–31</sup> Inefficient treatments in treatment plants show that domestic sewage, together with manufacturing industries, are the main routes for BPA to enter the environment.<sup>32</sup>

Various technologies are used to remedy its presence in water and soil, and the treatments can be grouped into biological, chemical, and physical techniques. It is worth noting that all technologies have their limitations; therefore, other studies are analyzing the application of hybrid processes as a way to increase the process efficiency and reduce the operating costs. In particular, for BPA reported in the literature, one can observe studies on biological degradation (native microbes and



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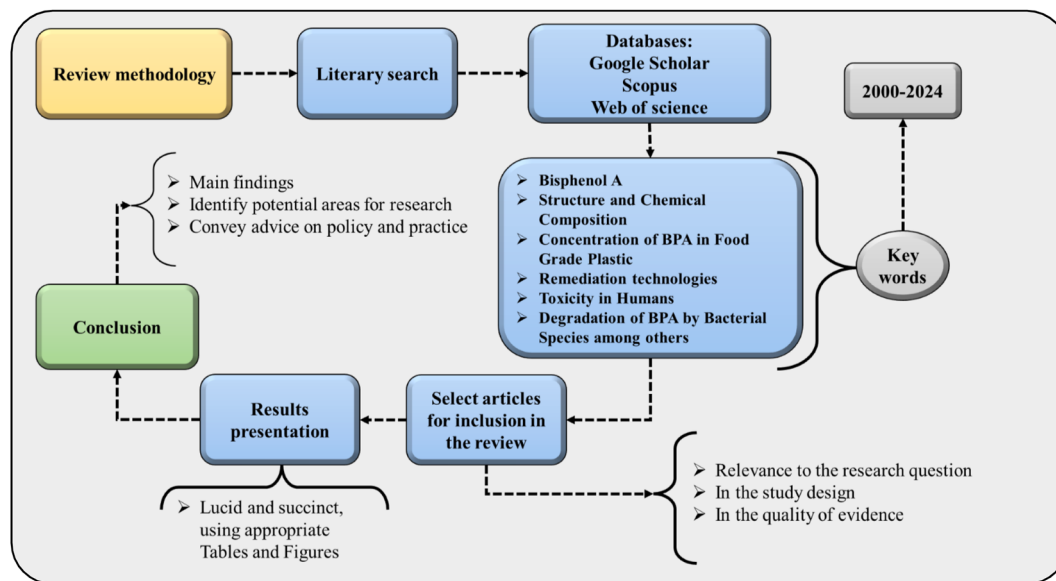


Fig. 1 Steps carried out in the scientific methodology applied in the study.

enzymes),<sup>33,34</sup> biochemical and photo-oxidation,<sup>35–37</sup> membrane separation processes,<sup>38,39</sup> thermal degradation<sup>40–42</sup> and adsorption (graphene, zeolites, and modified carbon).<sup>43–45</sup> These are among the most applicable technologies for decontaminating solutions containing BPA. It is essential to have a critical understanding of the efficiency of these processes across the varied ecospheres of the environment.

This study first analyzes the toxicity of BPA through its detection in environment-grade plastics. As a methodological process, a careful search in the Web of Science, Google Scholar, Scopus, and other databases is performed. A critical and current view is taken to describe the potential for the biodegradation of BPA by different species of bacteria. With this, the potential

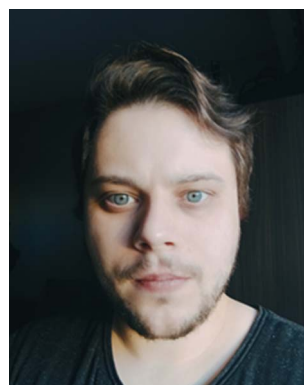
damage along the food chain is described. An in-depth analysis is described to elucidate the contamination and presence of BPA in food-grade plastic packaging. The influences of various storage conditions, such as pH and temperature parameters, are analyzed. To remediate BPA, an understanding of the entry routes (disposal of packaging and wastewater from industrial activities) and the transformation mechanisms is essential. Therefore, this entire process is also analyzed, corroborating the project to improve remediation technologies. Based on the proliferation of BPA in the environment, it was pertinent to review and discuss the current advances in physical, chemical, and biological processes for BPA remediation. To improve our understanding of these processes, an in-depth analysis of the



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mechanisms involved in BPA remediation is discussed for each process. This study aims to update the target audience on the problematic presence of BPA in food-grade plastics, to alert them to the extreme urgency of the need to develop safer and more sustainable alternatives.

## 2. Review methodology

The applied methodology is summarized in Fig. 1. Initially, a literary search was carried out in an extensive database, which includes journals present in the Web of Science, Google Scholar, Scopus, and others. Second, the words used for the search were selected, which include chemical composition and structure, BPA, its concentration in food-grade plastics, the biodegradation of BPA by strains of bacteria, the toxic effects on living organisms, and remediation technologies. The selection of articles that were published over the last 24 years was carried out, seeking to base critical analyses on current studies. For the selected studies, the relevance of the research, the quality of the evidence, and the study design were taken into consideration. The data collected were extracted in a methodical and organized manner to speed up the synthesis and analysis process. All data were carefully examined and collated to generate a comprehensive synopsis regarding the toxicity, concentration, and remediation of food-grade BPA. The results are presented succinctly and lucidly using resources such as figures and tables, explaining the main conclusions drawn. The limitations present in the literature and possible future implications for the area are also provided. Finally, the conclusion provides the main findings of the investigation to describe the main gaps existing in the political, social, and environmental spheres in the study area.

## 3. Food-grade plastics: bisphenol A concentration and sources

Characteristics such as clarity, resistance to impact and heat, and durability confirm that BPA is the most widely used synthetic compound in environmental quality plastic packaging in the world.<sup>46</sup> The major concern is that once present in the environment, BPA is easily leached into groundwater and surface water, especially when exposed to acidic chemical compounds and high temperatures.<sup>47</sup> The manufacturing process is one of the main sources of BPA contamination in food-grade plastics.<sup>48</sup> In polycarbonate, BPA is applied as a monomer, and this process includes the polymerization of the molecules generating a more durable and robust compound.<sup>49</sup> To optimize the durability and resistance, BPA is also applied as an additive in the production of epoxy resins. During these processes, the compound can also be released into the environment. The steps involved in this process are molding and polymerization. The concentration of BPA released from plastics depends on some factors such as acidity, type of plastic, temperature, the manufacturing process used, and duration of exposure time.<sup>50</sup> The storage stage and application of food-grade plastic packaging also influence the sources of BPA.<sup>51</sup> Since food products are packaged in plastics, depending

on storage conditions, BPA can penetrate food, increasing the likelihood of exposure of humans.<sup>52</sup> Another point is that many packages until they reach the final consumer are scratched or damaged, which increases the product contamination process.<sup>53</sup>

The recycling process is also a source of contamination in food-grade plastics, although this also has the positive aspect of reducing waste reduction.<sup>54,55</sup> Residual BPA obtained at the manufacturing stage is generally present in the recycled product and can be transported to other products.<sup>56</sup> The presence of BPA also occurs in other products (toys, medical and dental devices, and water bottles), which increases the sources of food-grade plastics. The intensity of contamination is high when these products are discarded inappropriately.<sup>57</sup> The literature presents several reports that seek to evaluate the concentration of contaminants in food-quality packaging. The results indicate that the quantity varies depending on the pH and temperature, manufacturing process, the substance contained in the container, and the type of plastic. Several studies report its presence in various packaging (water bottles, baby bottles, and food packaging), with the highest concentrations being found in polycarbonate.<sup>58,59</sup> Studies have reported the presence of BPA in breast milk samples (Asia).<sup>60</sup> A study confirmed that the increase in temperature increases the transfer of BPA from the container to the food, which explains why certain samples contain higher concentrations of BPA.<sup>61</sup> Canned products contain a coating that contains epoxy resins. In this investigation, it was observed that the concentration of BPA in foods highly varied, making it possible to detect low and high concentrations.<sup>62,63</sup>

In the literature, it is possible to observe investigations that indicate that low concentrations of BPA do not cause harm to human health; however, many claim that increasing exposure over a long period can cause serious damage, mainly due to its bioaccumulation capacity, which would increase its concentration.<sup>64</sup> Due to this, an increase in research has been observed looking for more ecological and alternative materials to replace BPA in food-grade plastic containers. Some industrial processes are already manufacturing plastic packaging without the presence of BPA; among the alternative materials are stainless steel and glass.<sup>65,66</sup> The great concern is not only associated with human health but also with the impact on the environment; open-air landfills, exposed to high temperatures, are one of the main sources of BPA leaching into the water table, which is intensified in countries with low income.<sup>67</sup> Because of this, it is highly urgent to seek new, safer alternatives to applying BPA to food-grade plastics. The change in population behavior over the last 20 years has also changed the main sources of BPA release, as a result of which scientific advances have made it possible to determine the main transformation pathways, which are described in Table 1.

### 3.1 Exposure to bisphenol A: toxic potential

The commercialization of epoxy resins and the production of polycarbonate plastics are associated with a series of health problems in humans.<sup>81–83</sup> The chemical compound can imitate the estrogen hormone (Fig. 2), and hence, studies report that



Table 1 Main sources of BPA release in the last 20 years and their mechanisms of action in the environment

| BPA release activity into the environment            | Transformation mechanism                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                            | Reference |
|------------------------------------------------------|-------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|-----------|
| Migration of polycarbonates                          | The transport of BPA into liquid foods occurs through polycarbonate <i>via</i> two routes. After the polymer hydrolysis process is catalyzed <i>via</i> hydroxide (manufacturing), residual BPA is released by diffusion into foods in contact with water and simulant. The polycondensation of BPA using phosgene (a highly toxic product) is responsible for the production of polycarbonate, which is of greater economic interest. During the hydrolysis and diffusion of polycarbonate, BPA can migrate, with hydrolysis being much more impactful in terms of release; this is due to pH values and the contact of other compounds (cations). An example was polycarbonate bottles with water stored and with a use time of seven years, when exposed to high temperatures, BPA quickly went into the water. In addition, the speed and concentration of BPA migrated will depend on the chemical composition of the solution | 68        |
| Translocation through microplastics present in water | In the manufacture of resins and polymers, BPA is classified as a common precursor, where microplastic particles manufactured with low-density polyethylene together with polycarbonate, easily release BPA into water media. Additives added to water resources (toxic organic compounds) include microplastics and intensified contamination                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                      | 69        |
| Waste management: inefficient treatments             | In effluent treatment plants, BPA is leached into the hydrolytic of plastic waste. Treatment generally achieves maximum efficiency in removing endocrine disruptive compounds ( $R = 37\text{--}94\%$ ), contaminating water, soil, and air. The release occurs in the decomposition and wear of plastics. Waste paint from electronics also releases BPA due to burning computer circuit boards. Bacteria (soil/sediment) under anaerobic conditions can degrade in four steps forming tribromobisphenol A, dibromobisphenol A, and monobromobisphenol A, followed by BPA as the final product                                                                                                                                                                                                                                                                                                                                     | 70 and 71 |
| Effect of temperature on plastic degradation         | When polymers are subjected to high temperatures, thermal degradation occurs where BPA is synthesized into bis (4-hydroxyphenyl) methane and 4-hydroxybenzoic acid. As a result of this process, the scission of the weak bonds present in the carbonate chain and alcoholysis/hydrolysis are obtained                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                              | 72        |
| Food packaging and canned food                       | Food consumption is the main route of human contamination. Its contact is due to the exposure of animals and the vegetable raw materials used, therefore, bioaccumulation is due to the presence of polymers and food. Studies estimate that consumption <i>via</i> food contaminated with BPA varies from 0.48 to 1.6 g per kg body weight per day, depending on consumption habits, location, and economic and cultural background. The main routes of contamination of water and food are through epoxy resins and polycarbonates                                                                                                                                                                                                                                                                                                                                                                                                | 73 and 74 |
| Electronic waste: disposal and recycling areas       | Developed and underdeveloped countries, especially the latter, are concentrated in large spaces that recycle waste from electronic products, which release large amounts of BPA. This liberation has increased in recent decades due to the constant advancement and dependence of society on electronic products. The recycling of these materials is generally done crudely by dismantling the product and separating the electronic parts (this step uses an incinerator at high temperature), the burning process releases many toxic gases into the atmosphere                                                                                                                                                                                                                                                                                                                                                                 | 75        |
| Dentistry: use of orthodontic adhesives              | Processes related to the dental area use raw materials containing high concentrations of BPA, such as orthodontic adhesives based on Bis-GMA. BPA has an estrogenic behavior that is limited to the double-ring benzoic compound. The breakdown of Bis-GMA releases BPA. Widely used in the restoration of resins, being one of the most dangerous polymers released into the environment. BPA occupies around 70% of the composition of Bis-GMA resins used in enamel restoration and dental obstructions. Its toxicity has the potential for cytotoxicity at the cellular level. Poly-BPA molecules present in the oral route can decompose into BPA, this is accelerated by pH conditions, enzymes and chemical substances present, and temperature                                                                                                                                                                              | 76        |
| Epoxy resins                                         | Prepolymers contain several epoxide groups and are known as epoxy resins. These resins are placed in contact with various curing agents (aliphatic amines, anhydrides, and polyamides), generating cross-linked polymers in industrial processes. Most of the resins are produced with bisphenol-A diglycidyl ether, a very low molecular weight oligomer (the main component of liquid resins)                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                     | 77        |
| Paper coins                                          | High levels of BPA were detected in thermal paper ( $322\text{ g kg}^{-1}$ ). During the handling of received products, contamination can occur dermally, as BPA is sprayed on the leaves (powdery layer). When a receipt is placed next to the coin in the cash register or wallet, or even when a receipt is handled before the money, the contaminant can be transported to the paper currency                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                   | 78 and 79 |
| Use of anti-rust products                            | Products that contain epoxy resin are sources of BPA release, among the most common are epoxy coatings with water and high levels of dust and solids, cans that store drinks and food and are coated with protection, primers used in car bodies automobiles, electrical laminates, composites with fiber-based reinforcement, tools, and castings                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                                  | 80        |



there may be changes in activities related to reproduction, fertility (polycystic ovary syndrome), development, increased incidence of cancer cells, and risk of obesity and diabetes.<sup>84</sup> The effects on the neurological system are intensified in children interfering with learning and behavior.<sup>85</sup> Children exposed to BPA in the prenatal period showed reduced bone density.<sup>86</sup> Exams analyzed children's urine and observed that individuals who had a higher concentration of BPA were more aggressive and hyperactive. Women exposed to BPA in the prenatal period also had a higher rate of mental disorders, such as depression and anxiety. Studies also link the contaminant with disorders in the immune system, decreasing cellular activity and increasing the likelihood of diseases and infections.<sup>87</sup> Concerning the digestive system, it was observed that even at low concentrations, BPA causes dysregulation of the intestinal barrier, reducing the microbiota and composition, which affects the immune system.<sup>88,89</sup> Studies carried out on animals have shown a greater risk of neurodegenerative diseases (Alzheimer's disease, Parkinson's disease, and amyotrophic lateral sclerosis) when they are exposed to BPA, in addition to behavioral disorders and cognitive abnormalities.<sup>90–92</sup> In men, a decrease in semen quality has been observed, directly affecting their fertility.<sup>93</sup> Endocrine-disrupting compounds such as BPA affect endothelial function, resulting in the loss of vascular homeostasis and increasing the chances of cardiovascular diseases.<sup>94</sup>

The genotoxic effect has been recently observed, which is related to the pro-oxidative potential. The increase in oxidative stress is linked to the increase in the production of species that react with oxygen, the accumulation of oxidation compounds (biomacromolecules), and the change in the activities of antioxidant enzymes.<sup>95</sup> This behavior is possible to occur in negative and positive estrogen receptor organisms. Once intracellular oxidation occurs, the chances of DNA damage (double- and single-strand breaks and oxidative damage) increase. Another study describes that BPA, when it also uses non-canonical

estrogen as a receptor, is connected to a membrane and the receptor is coupled to the G protein. Therefore, BPA can modulate transcription factors and ion channels, which generates damage such as mitochondrial stress, damage to the endoplasmic reticulum, and damage to the amyloid polypeptide of human islets.<sup>96</sup> This causes the activation of NFκB in β cells located in the pancreas, generating unregulated glucagon secretion.<sup>96</sup> There are conflicting studies regarding the incidence of asthma in children due to BPA, where three studies show a higher risk of asthma in childhood, and only one reported a low risk.<sup>97</sup>

Damage is also caused to animals, and in the literature, it is possible to observe studies that analyze the damage caused to the reproductive system of boars and sows and relate it to the levels of oxidative stress in the offspring, the quality of the meat, and muscle metabolism.<sup>98</sup> Most studies are related to damage at the level of reproduction (endocrine disruption); however, there is a greater need for analysis regarding the maturation of oocytes and the potential and mechanism of their toxicity in farm animals such as cows and nuts. As in humans, mice that were exposed to BPA showed a reduction in the intestinal microbial composition and richness, where they present distinct metabolic pathways, and thus, the study highlights intestinal toxicity in vertebrates.<sup>99</sup> In aquatic organisms, the effects are similar to those generated in humans, and include the following: decreased life expectancy and alteration of fat metabolism,<sup>100</sup> alteration of fertility,<sup>101</sup> negative regulation of ecdysone genotoxicity,<sup>102</sup> ecdysone upregulation,<sup>103</sup> endocrine-disrupting cytotoxicity,<sup>104</sup> neurodevelopmental toxicity,<sup>105</sup> alteration of retinal morphology due to alteration in thyroid hormone levels,<sup>106</sup> reproductive alteration genotoxicity,<sup>107</sup> modulation of antioxidant enzymes and decreased filtering capacity,<sup>108</sup> and physiological and reproductive changes in fish.<sup>109</sup>

In plants, studies are more limited, however, it is described that BPA generates stress, which produces a large number of

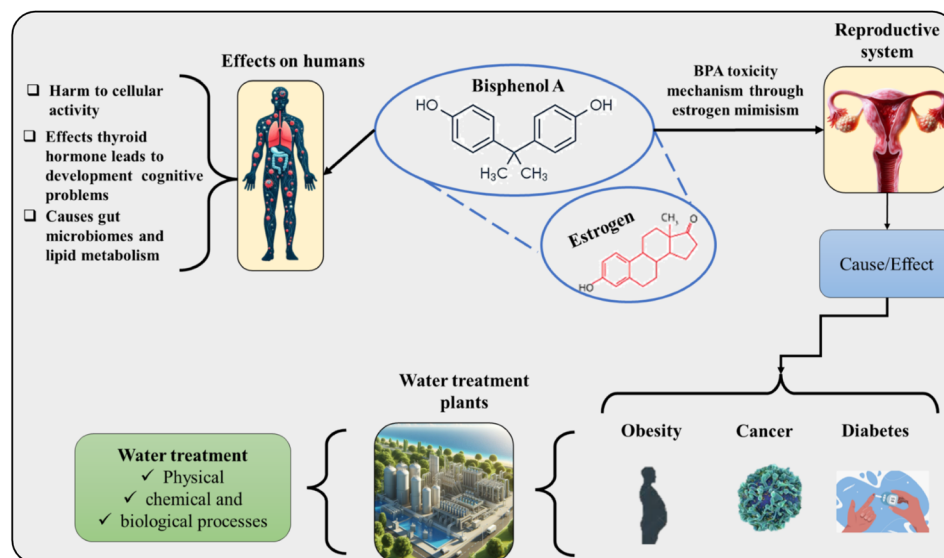


Fig. 2 Mechanism and toxic effects of bisphenol A on human health.



species that react with oxygen.<sup>110</sup> These new species act as signal messengers and provoke various stress responses in plant antioxidant cells.<sup>111</sup> The literature describes that BPA can imitate the cytokine hormone in plants, in addition to the role of ethylene in the oxidative stress process generated by BPA.<sup>112</sup> It is also possible to observe studies that describe different changes at both the biochemical and physiological levels in plant cells.<sup>87</sup> Generally, biochemical reactions correspond to changes in enzymes at temperatures close to the environment. More studies must be carried out to optimize biomonitoring and describe what the risk situations are. Therefore, plant species that are truly sensitive to different concentrations of BPA must be selected, inserted in a natural habitat, and identified biological changes in different phases of development. This will allow us to obtain answers about the relationship between the BPA dose and the associated effects, corroborating the risk assessments. Due to all these damages, there is still considerable discussion regarding the possible effects generated at low concentrations and long periods of exposure, which strengthens the urgency in finding alternative products for BPA.<sup>113,114</sup>

### 3.2 A case study of the fish microbiome and the presence of BPA

Due to the numbers that prove the high human exposure to BPA, the vast majority of studies use animal and human experimental models.<sup>115,116</sup> However, it is observed that water compartments can present high concentrations, and this includes both saltwater and freshwater environments. This fact corroborates the detection of BPA in the tissues of fish species in both cases. An example is in the United Kingdom where, due to the disposal of wastewater, the presence of BPA ranging from 37 to 462 ng L<sup>-1</sup> was detected in river samples.<sup>117</sup> This is not an isolated case, and in the Mediterranean region, the concentrations can reach 30 µg L<sup>-1</sup>, with a minimum of 4.5 µg L<sup>-1</sup>.<sup>118</sup> The same authors analyzed samples on the northeast coast of the Mediterranean, where a minimum concentration of 8.85 and a maximum of 14.76 µg L<sup>-1</sup> (Sea of Marmara and Black Sea) were observed. The authors highlight that the variation is influenced by meteorological variables that correspond to the seasons. Research highlights the influence of the presence of BPA on aquatic biota and its exposure to certain species. The fish species *Dicentrarchus labrax*, *Trachurus*, and *Scomber colias* were analyzed in the northeastern region of the Atlantic Ocean, and both fish tissues showed high levels of BPA proportional to the ingestion of microplastics.<sup>119</sup> As a result, several investigations describe that microplastics are one of the main sources of BPA contamination in aquatic environments.<sup>120</sup>

It is also described that the microbiome of most fish species is potentially sensitive to BPA contamination. Most reports that analyze the impacts of the contaminant on fish are on a laboratory scale using zebrafish as an experimental model. These studies observed that, just as it occurs in mammals, BPA changes the chemical composition and generates behavioral disorders in the living organism.<sup>121</sup> Different genera of zebrafish were exposed for 90 days to a concentration of up to 2 µg L<sup>-1</sup> of

BPA, resulting in a reduction in the pathogenic genus *Hyphomicrobium* and a significant increase in *Lawsonia*.<sup>122</sup> An impoverishment of the microbiota of the intestinal tract was also reported after the presence of BPA; in contrast, in the host model, a response to oxidative injuries and inflammation was observed. Zebra fish of both sexes were analyzed based on their response to the microbiome in the presence of BPA and its other replacement products, 10 days-old larvae showed a decrease in microbial diversity, especially at higher concentrations.<sup>123</sup> BPA derivatives showed diverse effects on the microbiome community. Finally, despite the composition of the microbiota having a behavior opposite to host toxicity, no noticeable effect was observed on the behavior of the larvae.<sup>124</sup> A concentration of 2000 µg L<sup>-1</sup> was added in contact with adult zebrafish for 35 days. In the end, it was reported that the fish showed an increase in muscle triglycerides and gastrointestinal dysbiosis.<sup>125</sup> These studies corroborate the elucidation of how BPA affects the gastrointestinal microbiota of aquatic biota, serving as a support for the comparison of investigations with mammals.<sup>126,127</sup>

## 4. Mechanisms of BPA toxicity in food-grade plastics

The scientific community has continuously studied BPA for two reasons: the first is due to its wide use in the most varied industrial processes, and the second is its ability to contaminate food and drinks, generating toxic effects on animals, humans, and plants. The first point is that BPA is considered an endocrine disruptor due to its ability to imitate estrogen. This hormone is fundamental for the functioning of the body's organs and tissues, in addition to being important for the development in the early stages of growth.<sup>128,129</sup> Once acting like estrogen, BPA alters the functioning of the endocrine system, generating changes in the activity and hormonal production of the human body. As already mentioned, this action generates several problems, among the most common of which are changes in fertility, diabetes, obesity, and cancer.<sup>130</sup> At a cellular level, it alters the cell activities by generating oxidative stress, which can cause damage to membranes, chemical composition, and even ADN.<sup>131</sup> Its toxic potential can occur in different signaling pathways in the human body. One example is the disturbance of enzymatic activities involving lipid metabolism, which causes disorders resulting in problems such as diabetes and obesity.<sup>132–134</sup> One of the specific receptors present in the human body that BPA can interfere with is that of the thyroid hormone, this causes the interruption of signaling, generating health problems, including behavioral and developmental changes.

A study<sup>135</sup> showed that higher levels of thyroid hormone were associated with contact *via* breast milk in newborns, being more evident in female children. Another point is its change in the activities of microorganisms present in the intestinal tract and digestive system, which causes changes and a possible decrease in the immune system, in addition to altering the dynamics of metabolism and other physiological processes that are



involved.<sup>91</sup> Therefore, BPA damage can occur through various mechanisms, and its ability to bind to estrogen receptors is the most discussed in the literature, as a result of which changes in cellular functions and microbial activity of the intestinal system are observed, intensified by direct effects on signaling pathways. This entire process causes considerable damage to health, which corroborates the intense efforts to develop new, more sustainable, and safer alternatives to the BPA present in various consumer products. Therefore, a reduction in exposure routes is achieved, reducing damage throughout the food chain.<sup>136,137</sup>

#### 4.1 Regulatory guidelines and limits for BPA exposure

Due to the serious damage and constant threat of BPA to human health, guidelines have been set and regulatory limits established in some countries. To achieve this, scientific studies were used to protect the society from the potentially harmful effects of the contaminant. The Environmental Protection Agency (EPA), located in the United States, decreed 50  $\mu\text{g}$  per kilogram of body weight per day as a reference dose, and this would be the maximum level of exposure considered non-harmful.<sup>138</sup> The US Federal Drug Administration also decreed guidelines that limit the exposure to contaminants in products for children's use, such as formula packaging and baby bottles. The guidelines established by the European Union were set by the European Food Safety Authority, where a daily amount of 4  $\mu\text{g}$  per kilogram of body weight per day is allowed. This daily dose indicates the limit that can be exposed without causing harm to health.<sup>139</sup> Unlike American legislation, the European Union does not allow the use

of the contaminant in cups, straws, and baby bottles. Canada has defined a maximum concentration of up to 0.005 mg of BPA present in water, which is acceptable; however, Health Canada has warned the population about the presence of BPA in baby bottles and baby formula packaging, warning the preference for smaller packaging sustainable. Countries such as Australia, Japan, and China have created guidelines with tolerance limits for BPA in different environments and packaging. Despite public policies that seek to limit contact with BPA, stricter measures are still needed, especially in less developed countries. Together, debates are needed to confirm whether these limits are truly safe for long-term exposure. Therefore, continued research is recommended to obtain a broader understanding of the real potential health risks together with new, more viable, and safer alternatives to BPA.<sup>140,141</sup>

The presence of BPA in liquids such as beverages generally occurs in products that feature the coating of bottle caps, cans, and other pipes that line the entire water supply.<sup>142</sup> As shown in Table 2, when compared to other solid foods, the presence of BPA in beverages is much less reported and at lower concentrations. Two factors can justify this: first the LODs are high in the most varied analytical methods used, and second, the concentration of coating used in beverage cans is lower than that used in solid foods. BPA has never been observed in liquids packaged in glass; however, it has been detected in plastic packaging and other types of canned goods. BPA has been detected in mineral water packaged with polyethylene terephthalate.<sup>144</sup> The highly low concentrations suggest that the main sources of BPA are bottle closures.<sup>162</sup> The presence of BPA in dairy products can also occur during production due to

Table 2 Presence of BPA in various human consumption products

| Subcategory                                                                 | LOG (ng g <sup>-1</sup> ) | LOQ (ng g <sup>-1</sup> ) | Min-max ( $\mu\text{g kg}^{-1}$ ) | Country and reference      |
|-----------------------------------------------------------------------------|---------------------------|---------------------------|-----------------------------------|----------------------------|
| Energy and soft cola, drinks, cola, juice, beer                             | 0.02                      | 0.02–8.1                  | —                                 | Belgium <sup>143</sup>     |
| Water                                                                       | 0.00075                   | 0.00073–0.102             | —                                 | Italy <sup>144</sup>       |
| Soft drinks                                                                 | 0.18                      | 0.54                      | 0.54–4.98                         | Italy <sup>145</sup>       |
| Soft drinks                                                                 | 0.10                      | 0.40                      | 0.4–10.2                          | Greece <sup>146</sup>      |
| Milk                                                                        | 0.15                      | —                         | 0.99–2.64                         | Spain <sup>147</sup>       |
| Milk                                                                        | 0.30–4.20                 | 1–14.00                   | <1–521                            | Italy <sup>148</sup>       |
| Cheese spreads, milk, brown cheese, and hard cheese                         | —                         | 0.10                      | <0.10                             | Norway <sup>149</sup>      |
| Milk                                                                        | 3                         | 9                         | 3–169                             | Italy <sup>150</sup>       |
| Soft, energy, and beer                                                      | 0.1–9.3                   | —                         | 0.1–3.4                           | Austria <sup>151</sup>     |
| Beans, haricots, red kidney beans, young peas, crushed tomatoes and lentils | 1.1–7.4                   | —                         | 8.5–35                            | Austria <sup>152</sup>     |
| Tomato                                                                      | 15.4–20                   | 51.3–66.9                 | 20.5–115.3                        | Italy <sup>153</sup>       |
| Carrots, peas, bean shoots, artichokes, and mixed vegetables                | 0.016                     | 0.55                      | 8.90–304                          | Spain <sup>154</sup>       |
| Green beans, red pepper, asparagus and lentils mushroom                     | 0.3–1.1                   | 0.9–3.5                   | 7.1–959                           | Spain <sup>155</sup>       |
| Tomato                                                                      | 0.09                      | 0.26                      | 0.31–235.88                       | Italy <sup>156</sup>       |
| Tomato paste and mushrooms                                                  | 0.60                      | 1.9                       | 4.9–66                            | Greece <sup>146</sup>      |
| Pineapple and peaches                                                       | 0.3–1.1                   | 0.9–3.5                   | 6.1–13                            | Spain <sup>155</sup>       |
| Meat                                                                        | —                         | 2–4                       | 6.9–13                            | Sweden <sup>157</sup>      |
| Tripe meat balls                                                            | 0.30–1.10                 | 0.90–3.50                 | 9.2–341                           | Spain <sup>155</sup>       |
| Poultry, meat, offals and game, delicatessen meats                          | 0.02                      | 1.2                       | 0.09–60.19                        | France <sup>158</sup>      |
| Ravioli, corn                                                               | —                         | 0.1                       | 0.94–73.1                         | Belgium <sup>143</sup>     |
| Pasta, bread, burns, flour, and breakfast cereals                           | —                         | —                         | 0.1–0.24                          | Norway <sup>149</sup>      |
| Eggs, oil, products, butter and margarine                                   | 0.21                      | —                         | 0.045–4.51                        | France <sup>159</sup>      |
| Honey                                                                       | 270                       | 800                       | —                                 | Italy <sup>160</sup>       |
| Mustard seeds                                                               | 10                        | 30                        | 10–8350                           | Switzerland <sup>161</sup> |
| Sauces                                                                      | 0.6                       | 1.9                       | —                                 | Greece <sup>146</sup>      |



contact with equipment and other utensils used during food processing. Ruminant feed can also generate BPA contamination, because it contains alkphenols, such as octylphenol and nonylphenol.<sup>163</sup> BPA and its derivatives are present in milk, which are presented in Table 2; in this case, the packaging may play a minor role in contamination. This is because variables such as heat treatments and the various utensils used in production can lead to greater contamination. Fatty dairy products that are secreted by cows can have a high presence of BPA. Few studies report the release of BPA throughout the dairy product production chain, and more studies are needed in this regard. In this same sense, more studies are needed regarding BPA in non-canned fruits or vegetables. In the case of canned products, it is possible to observe a considerable and precise estimate regarding their content, as shown in Table 2.<sup>155,159,164</sup> In general, canned vegetables constitute a highly heterogeneous group of foods (tomatoes, lentils, and mushrooms), together with water, other preservatives, and even oil in some cases. The variation in these products may reflect a high variation in the concentration of BPA. An earlier study<sup>157</sup> observed that people who consume canned vegetables in greater quantities have higher levels of canned vegetables in their urine. Likewise, the presence of BPA in meat occurs due to the migration of the packaging with the solid, and this can be intensified due to the production methods and the animal's diet. Table 2 also describes BPA levels in cereals, which can be detected in plastic packaging and cans; however, more studies must be conducted for these foods.

## 5. Bacterial degradation of bisphenol A

Several studies have been developed, mainly in the last decade, seeking to optimize the BPA decontamination process mainly in

water compartments and the soil. An alternative is the application of colonies of different species of bacteria that have the potential to degrade the chemical compound.<sup>165</sup> Table 3 describes the variety of species and strains analyzed and their degradation potential. The study by Sasaki *et al.* (2005), used the species *Sphingomonas* sp. strain AO1 isolated in a WTP (water treatment plant) located in Asia (Japan). Degradation was efficient since the microorganism used the chemical compound as a carbon source, where intermediates were generated, with a lower toxic potential. The system showed 100% removal for a concentration of 115  $\mu\text{g mL}^{-1}$ . Another study isolated the species *Sphingobium* sp. BiD32 from sediment and soil samples, the consumption of the chemical compound as the only source of energy allowed degradation through a sequence of chemical reactions, generating water and  $\text{CO}_2$  (carbon dioxide) as final products.<sup>184</sup> Similar degradation results were also obtained with other species such as *Rhodococcus* sp. and *Pseudomonas putida*<sup>185</sup> and *Arthrobacter* sp.<sup>186</sup> The efficiency of the process at a cost-benefit level depends on the operating conditions of the environment; therefore, the control and manipulation of variables such as pH, temperature, and quantity of nutrients and water is essential. Despite the positive results and many challenges to be overcome, a limitation is ensuring optimal degradation even in the presence of other chemical compounds. Another aspect is the scale of the process, since under real environmental conditions, it requires a large number of bacteria, coupled with the difficult control of environmental variables, which can limit the proliferation of colonies, affecting degradation.

### 5.1 Degradation mechanism

The process of degradation of the BPA molecule by bacteria involves several routes, which allow the molecule to be broken down into less toxic and simpler compounds. These routes encompass non-enzymatic and enzymatic processes and

Table 3 Biodegradation of BPA in food-grade plastics by bacteria

| Bacterial strains                                      | Concentration             | Bisphenol A removal (%) | Reference |
|--------------------------------------------------------|---------------------------|-------------------------|-----------|
| KU-3 bacterial strains (Chennai coastal region, India) | 250–1000 ppm              | 74                      | 166       |
| K-6 bacterial strains (Chennai coastal region, India)  | 250–1000 ppm              | 78                      | 167       |
| KU-8 bacterial strains (Chennai coastal region, India) | 250–1000 ppm              | 81                      | 167       |
| <i>Aeromonas hydrophila</i>                            | 60–120 $\text{mg L}^{-1}$ | 79                      | 168       |
| <i>Chlorella vulgaris</i>                              | 20 $\text{mg L}^{-1}$     | >50                     | 169       |
| <i>Sphingomonas</i> sp. strain AO1                     | 115 $\mu\text{g mL}^{-1}$ | 100                     | 170       |
| <i>Bacillus</i> sp. GZB                                | 10 $\text{mg L}^{-1}$     | 100                     | 171       |
| <i>Bacillus pumilus</i>                                | 10 ppm                    | 100                     | 172       |
| <i>Pseudomonas putida</i> YC-AE1                       | 1000 $\text{mg L}^{-1}$   | 93                      | 173       |
| <i>Bacillus</i> sp.GZ                                  | 100 $\text{mg L}^{-1}$    | 93                      | 174       |
| <i>Pseudomonas putida</i> KT2440                       | —                         | 46                      | 175       |
| <i>Pseudomonas</i> sp. LBC1                            | 160 $\mu\text{M}$         | 100                     | 176       |
| Bacterial consortium                                   | 50 $\text{mg L}^{-1}$     | 90                      | 177       |
| <i>Sphingobium</i> sp. BiD32                           | 2 $\text{mg L}^{-1}$      | 10                      | 178       |
| <i>Pseudoxanthomonas</i> sp. strain NyZ600             | 50 $\text{mg L}^{-1}$     | 45                      | 179       |
| <i>Sphingomonadaceae</i>                               | 5000 $\mu\text{g L}^{-1}$ | 99                      | 180       |
| <i>Bacillus</i> sp. AM1                                | 25 ppb                    | 84                      | 181       |
| <i>Dracaena sanderiana</i>                             | 100 $\text{mg L}^{-1}$    | 99                      | 182       |
| <i>Sphingomonas</i> sp. strain AO1                     | 115 $\mu\text{g mL}^{-1}$ | 100                     | 170       |
| <i>Chlorella pyrenoidosa</i>                           | 30 $\text{mg L}^{-1}$     | 67                      | 183       |



include a diversity of colony species. The production of BPA hydrolase enzymes is one of the main mechanisms and consists of the ability of these enzymes to break the bonds that unite the BPA molecule.<sup>187</sup> Hydrolases are produced by bacteria that develop when using BPA as a source of energy and carbon. These enzymes were found in several colonies (*Sphingobium japonicum*, *Pseudomonas putida*, and *Rhodococcus* sp.) that degraded the compound efficiently.<sup>188–190</sup> Non-enzymatic processes are also observed, among which biotransformation and adsorption stand out. In the case of adsorption, the interaction of BPA (adsorbate) molecules occurs, which are attracted *via* physical interactions to the surface of the bacteria cells, followed by absorption and degradation. Biotransformation is related to chemical changes that occur in the contaminant through the activity of enzymes produced by bacteria, and the products generated in the end are less complex and less toxic.<sup>191</sup> In this process, metabolic pathways are formed that enable the use of the compound as a source of nutrients. In the literature, the species *Pseudomonas putida* uses BPA as the sole source of energy and carbon; however, the species *Sphingobium japonicum* consumes the compound as part of a broader energy and carbon system.<sup>192,193</sup>

The degradation of BPA and the mechanisms used by bacteria may vary depending on the species used and the environmental conditions (pH, temperature, humidity, *etc.*) in which they were found. In general, the entire mechanism occurs through a combination of non-enzymatic and enzymatic processes, in which BPA bonds are broken.<sup>194,195</sup> Achieving total control over the mechanisms that involve degradation is fundamental in planning new strategies that aim to remedy and mitigate the environmental damage generated by the release of BPA. With the correct species and the mechanisms that the process involves, researchers can achieve maximum efficiency and the best cost-benefit in applying the method.<sup>196</sup>

Owing to the increasing contamination and exposure to compounds containing BPA, there is a growing need and urgency for remediation technologies such as bacterial degradation.<sup>197,198</sup> The colonies of bacteria with the potential to degrade BPA into less toxic compounds have been reported in diverse environments including sediments, treatment plants, and soil samples. Once new intermediate compounds are formed, they can be reused or discarded in safe places.<sup>199</sup> With these practices, it is possible to reduce the volume of waste from plastics and microplastics, corroborating with less exposure, mainly to less toxic chemical molecules. An area of research that encompasses degradation by bacteria and the management of solid waste is the use of bioreactors, and these structures provide the best environmental conditions for rapid colony growth through accelerated metabolism.<sup>200</sup> Bioreactors can isolate cultures where BPA is then added for subsequent degradation.<sup>201</sup> Another natural process is composting. These cases also make use of microorganisms that decompose organic material (garden waste and fruit and vegetable waste). This technique also makes it possible to break down BPA molecules, supporting the management of plastic waste and consequently reducing the environmental impact and mitigating harmful effects on humans, plants, animals, and aquatic organisms.<sup>202</sup>

These processes still need to overcome gaps such as, for example, isolating a certain species or colony capable of degrading several chemical compounds as well as its identification. This is the first step towards achieving ideal conditions for rapid metabolism and the application of a practical and cost-effective method that can be applied to large-volume BPA waste management.<sup>203</sup> To overcome these barriers, degradation with bacteria has great potential in the application of bioreactor development, including the possibility of composting. These are still possible alternatives and examples of sustainable activities in combating contamination by solid waste.

## 6. Overview of technologies for the remediation of BPA-contaminated water

### 6.1 Physical processes

Physical treatment processes such as adsorption, membrane filtration, and electrocoagulation are commonly used to remove BPA contaminants from water and wastewater due to their flexibility, simplicity, and effectiveness. However, methods other than adsorption have drawbacks such as high energy and maintenance costs, lower removal capabilities, and increased risk of membrane fouling.<sup>204</sup> Coagulation or flocculation treatment, a physicochemical process, is employed to remove BPA that attaches to sludge during the primary treatment stage followed by the secondary treatment stage in wastewater treatment.<sup>205</sup> However, the effectiveness of these methods is limited, and its highest recorded removal record rate is 1%. In the secondary treatment stage, BPA is subjected to advanced processes, including adsorption, oxidation, advanced oxidation, and membrane technologies which have high removal efficiency. Among the water treatment processes, coagulation or flocculation is the simplest technology with low installation costs.<sup>206</sup> Over the past decade, adsorption has emerged as the most widely used method for removing BPA. The raw materials for adsorbents, such as graphene, activated carbons, zeolites, and agricultural wastes, are diverse, inexpensive, and easy to maintain. These materials have proven to be promising for large-scale BPA removal applications.<sup>207</sup> Managing adsorbents effectively is essential for the efficient operation of the adsorption process. This includes selection, storage, regeneration, and treatment. Such comprehensive management significantly affects the absorbent's service life, reduces costs, and improves the overall efficiency.<sup>208</sup> Successfully employing adsorption methods for BPA removal requires thorough technical preparation. This includes smooth operation of equipment and diligent monitoring of parameters such as adsorbent saturation, temperature, and flow rate. Continuous assessment ensures efficient BPA removal.<sup>209</sup> Advancements in adsorption techniques promise cost-effective BPA removal. Successful implementation depends on innovation in materials/methods and meticulous management/monitoring throughout the process.

**6.1.1 Membrane processes.** Membrane separation processes (MSPs) are advanced treatments gaining popularity in water and sewage treatment plants for achieving high water



purity levels and facilitating water reuse. These processes encompass two primary mechanisms: sieving and adsorption.<sup>210</sup> Since the 1990s, membrane filtration has gained a lot of interest as a drinking water treatment technology due to its ability to achieve high throughput in a continuous operation mode. There are four main types of the process: reverse osmosis (RO), nanofiltration (NF), ultrafiltration (UF), and microfiltration (MF).<sup>211</sup> NF and RO are usually utilized in removing organic carbon, dissolved solids, and inorganic ions. They can also be used in removing EDCS and micropollutants.<sup>212</sup> They are operated at high pressures when compared to MF and UF membranes.<sup>213</sup> As much as 60 million m<sup>3</sup> per day of water is being treated by RO, while MF and UF have capacities of 20 million m<sup>3</sup> per day, accounting for 60% of total drinking water production.<sup>214</sup> UF is more popular than the other types of membrane separation technologies because it is cheaper, and can handle particles, macromolecules, suspended solids, colloids, bacteria, and viruses.<sup>215</sup> Both NF and RO membranes can effectively remove various types of micropollutants (MPs) not only *via* charge effects and adsorption but also through size exclusion.<sup>216</sup> Various types of membranes employed in treatments of BPA are presented in Table 4.

MF (0.2 µm) and UF (0.05 µm) may retain BPA by the interaction of the chemical with particulates, as the particulate's size is larger than the pore sizes of MF and UF membranes.<sup>222</sup> Up to 92.5% BPA removal efficiency was achieved due to the combined effect of laccase catalysis, membrane rejection, and adsorption.<sup>223</sup> This performance was obtained under optimal conditions of 2 bar pressure, competitive laccase loading (238.8 ± 3.5 µg cm<sup>-2</sup>), and laccase activity (0.6 U cm<sup>-2</sup>) (Li *et al.*, 2020).<sup>223</sup> Wang *et al.*, (2020)<sup>224</sup> proved that in the treatment of Songhua river water spiked with micropollutants, the percentage removal of BPA significantly increased to 88.5%,

surpassing the efficiency of single NF without coagulation, which achieved only 60.7% removal. After the fourth usage cycle, Li *et al.*, (2020)<sup>225</sup> observed a remarkable 99.8% removal efficiency for BPA using combined electrocatalytic oxidation and MF. The DECuO/CM membrane, featuring a permeability of 823.03 L (m<sup>2</sup> h bar)<sup>-1</sup>, was fabricated by depositing CuO onto the carbon membrane. The integration of PAA/PVDF membranes with the catalytic process using nZVI/H<sub>2</sub>O<sub>2</sub> under optimal conditions (50 L m<sup>-2</sup> h<sup>-1</sup>, 10 mM of H<sub>2</sub>O<sub>2</sub>) led to a notable 80% improvement in BPA treatment efficiency.<sup>217</sup>

UF is achieved through pores ranging between 5 and 20 nm in size, with molecular weight cutoffs of 1000 to 100 000 Da.<sup>226</sup> In both MF and UF processes, higher BPA concentrations in the feed solution led to increased adsorption efficiency. The estimated BPA removal rates were 36–72% for MF and 45–67% for UF.<sup>227</sup> To enhance the membrane performance, researchers have incorporated inorganic materials during membrane fabrication alongside MWCNTs and PVC-II. Under optimal conditions of 35 cm<sup>2</sup>, 300 mL, and 100 kPa for 10 minutes, BPA degradation reached 60%. Additionally, the researcher noted that the retention of BPA decreases with the increase in pressure.<sup>218</sup> Tian *et al.*, (2021)<sup>228</sup> developed a piezoelectric membrane using SnS<sub>2</sub> nanosheet-coated carbon nanofibers (CNFs) to generate hydroxyl free radicals for BPA degradation under ultrasound. They found the highest BPA adsorption with the UF membrane, followed by NF and RO membranes. The efficient reduction of BPA degradation was achieved with 97% efficiency within 80 minutes, respectively, through piezocatalysis of 0.5-SnS<sub>2</sub>/CNFs under ultrasonication in the darkness. However, Muhamad *et al.*, (2016b) demonstrated that the UF membrane system effectively removes BPA, with the membrane surface containing SiO<sub>2</sub> nanoparticles and hydroxyl bonding groups of BPA. The variations in feed pH notably influenced

Table 4 Various membrane technologies and their removal efficiencies<sup>a</sup>

| Membrane                                                | Type            | Parameter condition                                                                                                                                                                                                          | Remove efficiency | Reference                                  |
|---------------------------------------------------------|-----------------|------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|-------------------|--------------------------------------------|
| Layer-by-layer (LBL) biocatalytic                       | NF              | 85 cm <sup>2</sup> ; pH 6; operation pressure, 2 bar; temperature, 25 ± 0.5 °C, salt concentration, 1000 ppm; crossflow velocity, 0.2 m s <sup>-1</sup>                                                                      | 92.5%             | (Li <i>et al.</i> , 2020) <sup>223</sup>   |
| Polyamide NF membrane                                   | NF              | 2–3 cm (water level), 25 °C, pH = 7–9.7, 200 rpm for 2 min                                                                                                                                                                   | 88.5%             | (Wang <i>et al.</i> , 2020) <sup>224</sup> |
| Dynamic electrodeposited CuO/carbon membrane (DECuO/CM) | MF              | RhB concentration 300 mg L <sup>-1</sup> , voltage 2.5 V, electrode distant 5.75 cm, flow rate 1.2 mL min <sup>-1</sup> , electrolyte (Na <sub>2</sub> SO <sub>4</sub> ) concentration 5 g L <sup>-1</sup> , and feed pH 5.3 | 98.04%            | (Li <i>et al.</i> , 2020) <sup>225</sup>   |
| PVDF <sub>MW</sub> catalytic-membrane                   | MF              | 12 cm <sup>2</sup> , pH: 7.0, 531.6 eV                                                                                                                                                                                       | 40%               | 217                                        |
| PVC membrane                                            | UF              | 35 cm <sup>2</sup> , 300 mL, 100 kPa for 10 min                                                                                                                                                                              | 60%               | 218                                        |
| PES-SiO <sub>2</sub> hollow fibre membranes             | UF              | 1 bar, pH: 7, feed concentration: 10 and 100 µL <sup>-1</sup>                                                                                                                                                                | 95%               | 211                                        |
| Forward osmosis membrane                                | Forward osmosis | 15 bar, permeate flow of 1.6 mL min <sup>-1</sup> , concentrate flow of 80 mL min <sup>-1</sup> and a RO flux of 7 L m <sup>-2</sup> h <sup>-1</sup>                                                                         | 40%               | 219                                        |
| TW30 membrane                                           | RO              | 15 bar, 10 µg L <sup>-1</sup> , pH: 7                                                                                                                                                                                        | 84%               | 220                                        |
| TW30-4040, polyamide                                    | RO              | NDA                                                                                                                                                                                                                          | 90%               | 221                                        |

<sup>a</sup> NDA: no data available.



BPA elimination, achieving the highest rejection (90%) at pH 7 and the lowest removal (20%) at pH 10. The optimization process of the model identified the following optimum conditions for BPA removal: 1 bar pressure, pH 6.7,  $10 \mu\text{g L}^{-1}$  BPA concentration, and 10 minutes of filtration, resulting in 99.61% removal efficiency. Data from the RSM design confirmed that BPA removal under these optimum conditions was consistent at 99.61%, aligning with experimental values obtained in the UF membrane system.<sup>229</sup>

RO, primarily used for desalination due to its effective particulate rejection capabilities, has been suggested as an additional step in sewage treatment. Multiple research articles have identified RO as the most promising and efficient method for removing organic micropollutants (OMPs), including BPA.<sup>230</sup> In terms of forward osmosis rejection efficiency for BPA, the polyamide thin-film composite membrane (PA-TFC) achieved 25% removal, while the polysulfone (PSf) membrane substrate demonstrated 91% removal from the feed solution.<sup>231</sup> The concentrate produced from municipal wastewater recycling processes, known as reverse osmosis concentrate (ROC), presents notable environmental and health hazards due to its elevated levels of harmful compounds, such as phenolic chemicals like BPA. Under optimal conditions of  $25^\circ\text{C}$ , a  $[\text{Fe(IV)}]/[\text{BPA}]$  ratio of 50, and an initial pH of 8.0 in the ROC, BPA degradation surpassed 90% after a 90 minutes reaction period.<sup>232</sup> However, Moreira *et al.* (2019a) reported BPA removal from water *via* RO membrane separation. The RO membrane showed superior efficiency in removing BPA from samples containing  $10 \mu\text{g L}^{-1}$  at 15 bar pressure, and the removal efficiency rate was 84%.

**6.1.2 Adsorption.** Adsorption is the transfer of adsorbate from the liquid or gas phase to the surface or interface of the adsorbent, representing a surface phenomenon. Adsorption serves as a primary mechanism facilitating the removal of BPA through membrane filtration, along with size exclusion and charge repulsion. This process is anticipated to be influenced by hydrophobic interactions and hydrogen bonding. Notably, adsorption has demonstrated efficacy owing to its affordability, simplicity, environmentally friendly operation, the availability of various adsorbents, and the potential for renewal and reusability.<sup>233</sup> The adsorption rate depends on adsorbent properties, adsorbate nature, and solution conditions. Factors affecting adsorption include temperature, experimental parameters (*e.g.*, time, pH), characteristics of adsorbate and adsorbent, and the presence of other pollutants.<sup>234</sup> Researchers utilize various adsorbents such as zeolites, clay minerals, nanomaterials, graphene, activated carbons, imprinted polymers, biopolymers, and agricultural wastes for removing BPA from water. The effectiveness depends on their adsorption capacity.<sup>235</sup> Adsorbents are grouped into natural ones (*e.g.*, chitosan, clays), activated carbons, graphene-based materials, nanomaterials, composite materials, imprinted polymers, agricultural wastes, hybrid particles, and inorganic-organic-modified bentonite. However, challenges including difficulty in separating very small-sized materials from the treated matrix remain.<sup>236</sup>

The performance of various adsorbents for BPA uptake is summarized in Table 5. The  $\rho$ -phenylenediamine-

functionalized magnetic graphene oxide (PPD-MGO) nanocomposite nanomaterial exhibited good adsorption ability for BPA and good re-usability. Under optimum conditions of  $45^\circ\text{C}$  and pH 7, the maximum adsorption capacity reached  $155.0 \text{ mg g}^{-1}$ . The removal rate was 99.2% after three times of adsorption with the new nanomaterials.<sup>251</sup> Under Langmuir models, self-flocculated powdered activated carbon (PAC-PNIPAM) exhibited the maximum adsorption capacity of BPA by 247.532 and  $116.298 \text{ mg g}^{-1}$  when conducted at  $25^\circ\text{C}$  and  $40^\circ\text{C}$ , respectively.<sup>252</sup> Chen *et al.* (2015) reported that the maximum adsorption capacities of BPA by the Langmuir model were  $182 \text{ mg g}^{-1}$  at pH 5.0 and  $196 \text{ mg g}^{-1}$  at pH 7.0, respectively.<sup>253</sup> The adsorption of BPA onto activated carbons was employed at 298 K and pH 7.0. Carbons W20 and W20N had higher adsorption capacities ( $382.12$  and  $432.34 \text{ mg g}^{-1}$ ) compared to the other tested carbons. Adsorption capacity was influenced by acidic oxygen-containing groups and surface charge density.<sup>254</sup> BPA removal from aqueous solutions *via* graphene adsorption showed a maximum adsorption capacity ( $q_m$ ) for BPA of  $182 \text{ mg g}^{-1}$  at 302.15 K, the highest reported value among carbonaceous adsorbents.<sup>255</sup> The optimal operating parameters for the degradation of BPA were determined:  $0.3 \text{ mL}$  of  $\text{H}_2\text{O}_2$ , pH 3, and 180 min, in the presence of  $\text{Fe}_3\text{O}_4$ @illite. The maximum degradation capacities of  $816 \text{ mg g}^{-1}$ ,  $364 \text{ mg g}^{-1}$ , and  $113 \text{ mg g}^{-1}$  for epoxy BPA concentrations in resin wastewater ( $266 \text{ mg L}^{-1}$ ), synthetic wastewater ( $80 \text{ mg L}^{-1}$ ), and Hefei City Swan Lake ( $25 \text{ mg L}^{-1}$ ), respectively.<sup>256</sup> Moreover, Hernández-Abreu *et al.* (2020)<sup>236</sup> demonstrated that activated carbons (F400 and KLP) had significantly higher maximum BPA adsorption capacities ( $400$  and  $220 \text{ mg g}^{-1}$ , respectively) compared to xerogel ( $78 \text{ mg g}^{-1}$ ). This difference was probably due to the greater microporous volume observed in F400 and KLP-activated carbons. Similarly, at an initial BPA concentration of  $100 \text{ mg L}^{-1}$ , the maximum adsorption capability of  $285 \text{ mg g}^{-1}$  is 9.3 times higher than that of commercially available activated carbon adsorbents.<sup>250</sup> When the BPA concentration exceeded  $342 \text{ mg L}^{-1}$ , the adsorption capacity of multi-walled carbon nanotubes modified with iron oxide and manganese dioxide ( $\text{MWCNTs-Fe}_3\text{O}_4\text{-MnO}_2$ ) reached saturation. The static adsorption capacities were  $76.75 \text{ mg g}^{-1}$  for MWCNTs,  $44.52 \text{ mg g}^{-1}$  for MWCNTs- $\text{Fe}_3\text{O}_4$ , and  $132.89 \text{ mg g}^{-1}$  for MWCNTs- $\text{Fe}_3\text{O}_4\text{-MnO}_2$ .<sup>246</sup> Activated carbon (AC) efficiently removes endocrine-disrupting chemicals from water. It shows high removal rates of BPA even at low initial concentrations ( $1 \text{ mg L}^{-1}$ ; pH: 6.5; adsorbent:  $0.2 \text{ g L}^{-1}$ ) and increased substantial adsorption capacity for BPA, estimated at  $254.9 \pm 16.2 \text{ mg g}^{-1}$ .<sup>238</sup>

Transmission electron microscopy (TEM), Fourier transform infrared (FT-IR) spectroscopy, vibrating sample magnetometry (VSM), X-ray diffraction (XRD), X-ray photoelectron spectroscopy (XPS), thermogravimetric analysis (TGA), energy-dispersive X-ray spectroscopy (EDS), zero-point charge determination ( $\text{pH}_{\text{PZC}}$ ), inductively coupled plasma-optical emission spectroscopy (ICP-OES), and quantum mechanics/molecular mechanics (QM/MM) were performed.

**6.1.3 Coagulation/flocculation treatment.** Electrocoagulation involves passing an electric current through water,

Table 5 BPA treatments using various adsorbents materials

| Adsorbent                                                                                                                                    | Detection of BPA                                                     | Performance                                                                                    | Reference |
|----------------------------------------------------------------------------------------------------------------------------------------------|----------------------------------------------------------------------|------------------------------------------------------------------------------------------------|-----------|
| Calcium alginate/activated carbon (A-AC)                                                                                                     | FESEM, TGA, XRD, FTIR, BET                                           | 368.3 mg g <sup>-1</sup>                                                                       | 237       |
| Activated carbon                                                                                                                             | N <sub>2</sub> adsorption, desorption isotherm, FTIR, zeta potential | 100%                                                                                           | 238       |
| Activated carbon                                                                                                                             | TEM, VSM, EDS, XRD, XPS, FTIR                                        | 100%                                                                                           | 239       |
| Xerogel (RFX), a chemically activated carbon from kraft lignin (KLP), commercial activated carbon (F400)                                     | SEM, FTIR                                                            | F400 = 407 mg g <sup>-1</sup> , KLP = 220 mg g <sup>-1</sup> , xerogel = 78 mg g <sup>-1</sup> | 236       |
| Cellulose acetate (cigarette butt) activated carbon                                                                                          | FTIR, Raman, TGA, XRD, BET, SEM and XPS                              | 364.21 mg g <sup>-1</sup>                                                                      | 240       |
| $\rho$ -Phenylenediamine-functionalized magnetic graphene oxide nanocomposites (PPD-MGO)                                                     | TEM, FT-IR, VSM                                                      | 155.0 mg g <sup>-1</sup> (99.2%)                                                               | 241       |
| Biomass-activated carbon ( <i>Tithonia diversifolia</i> )                                                                                    | SEM and XRD                                                          | 15.69 mg g <sup>-1</sup>                                                                       | 242       |
| nZVI-chitosan                                                                                                                                | FTIR, TEM, XRD, SEM                                                  | 65.16 mg g <sup>-1</sup>                                                                       | 243       |
| Phosphonate <i>Halomonas</i> levan (PhHL)                                                                                                    | FTIR, XPS                                                            | 26.6 mg g <sup>-1</sup>                                                                        | 244       |
| Sulfonic acid functionalized carbonaceous adsorbent (TW-SO <sub>3</sub> H) from tea leaves                                                   | FTIR, SEM, EDS, TGA, EDS, TGA, Raman spectroscopy, zeta potential    | 236.80 mg g <sup>-1</sup>                                                                      | 245       |
| Modification of multi-walled carbon nanotube with iron oxide and manganese dioxide (MWCNTs-Fe <sub>3</sub> O <sub>4</sub> MnO <sub>2</sub> ) | SEM, TEM, EDX, FTIR, VSM                                             | 132.9 mg g <sup>-1</sup>                                                                       | 246       |
| Calcium alginate/activated carbon (A-AC)                                                                                                     | pH <sub>PZC</sub> , FTIR, SEM                                        | 368.3 mg g <sup>-1</sup>                                                                       | 247       |
| Cu-BDC MOFs, Cu-BDC@GrO (graphene oxide)                                                                                                     | XRD, TEM, SEM, EDS, Raman, FTIR, TGA, XPS, Zetasizer and ICP-OES     | 182 mg g <sup>-1</sup>                                                                         | 248       |
| Magnetic vermiculite modified (MV) poly(trimesoyl chloridemelamine) (MP)                                                                     | FTIR, SEM, EDX                                                       | 273.67 mg g <sup>-1</sup>                                                                      | 249       |
| Nitrogen-containing covalent organic framework (PyTTA-Dva-COF)                                                                               | QM/MM, XPS                                                           | 285 mg g <sup>-1</sup>                                                                         | 250       |

Table 6 Studies on BPA removal via coagulation/flocculation

| Method used                              | Operation condition                                                                                                                        | Removal efficiency | Reference |
|------------------------------------------|--------------------------------------------------------------------------------------------------------------------------------------------|--------------------|-----------|
| Peroxy electrocoagulation                | 45 ppm, 0.1 g per L Na <sub>2</sub> SO <sub>4</sub> , 125 g per L H <sub>2</sub> O <sub>2</sub> , pH: 2, 0.11 mA cm <sup>-2</sup> , 45 min | 80.48%             | 258       |
| Photocatalysis                           | 10 mg L <sup>-1</sup> , TiO <sub>2</sub> @MIL-101(Cr), 240 min                                                                             | 99.4%              | 259       |
| Electrocoagulation                       | 3.25 mg L <sup>-1</sup> , pH: 8.5, 12.0 mA cm <sup>-2</sup> , 23 min, 0.308 kW h m <sup>-3</sup>                                           | 98.2%              | 260       |
| Photocatalytic                           | 5.0 mg L <sup>-1</sup> , pH 5.0, 500 W, 0.3 g L <sup>-1</sup> , 180 min                                                                    | 100%               | 261       |
| Electro-fenton                           | 10 ppm, 10, 25 mg L <sup>-1</sup> , pH: 2.9, Na <sub>2</sub> SO <sub>4</sub> , Fe <sub>2</sub> SO <sub>4</sub> , 50 min                    | >90                | 262       |
| Membrane electrochemical microfiltration | 2.0 V, 50 ppm, 0.1 mol L <sup>-1</sup> Na <sub>2</sub> SO <sub>4</sub> , 0.88 min                                                          | 97%                | 263       |
| Microwave-Mn-fenton                      | 100 ppm, 34.0 mg L <sup>-1</sup> H <sub>2</sub> O <sub>2</sub> , 300.0 W, 6.0 min, pH: 4                                                   | 99.7%              | 264       |
| Iron electrode electrocoagulation        | 10 mg L <sup>-1</sup> , 25 °C, 30 V, pH = 7, 1500 mA                                                                                       | 90%                | 265       |

generating coagulant precursors through the electrolytic oxidation of anode materials, usually aluminum or iron. This technology effectively reduces pollutant levels. The performance of various electrocoagulation systems for BPA treatment is summarized in Table 6. Kumar *et al.* (2022)<sup>257</sup> reported that an electrocoagulation (EC) device equipped with aluminum blades was used to eliminate BPA, resulting in the removal of

approximately 42% of the compound. Locust bean gum (LBG) was utilized for coagulation/flocculation as a pre-treatment to reduce BPA levels in leachate prior to advanced treatments. It was reported that at a pH of 7.5, an LBG dosage of 500 mg L<sup>-1</sup> resulted in a reduction efficiency of 76% for BPA.<sup>266</sup> In the modified coagulation sludge process, a BPA concentration of 20 mg L<sup>-1</sup> underwent complete degradation within 120 minutes



Table 7 Previous studies on the chemical process involved in the removal of BPA

| Treatment unit                                                                                       | Performance removal                            | Reference |
|------------------------------------------------------------------------------------------------------|------------------------------------------------|-----------|
| Ozonation                                                                                            | 100%                                           | 272       |
| Ozonation                                                                                            | 17–100%                                        | 273       |
| Ozonation                                                                                            | 40%                                            | 274       |
| Ozonation                                                                                            | 33%                                            | 275       |
| Ozonation                                                                                            | 21.5%                                          | 276       |
| Photocatalytic degradation (Bi <sub>2</sub> WO <sub>6</sub> /g-C <sub>3</sub> N <sub>4</sub> /BPQDs) | 95.6%                                          | 277       |
| Photocatalytic degradation (CPD-TiO <sub>2</sub> )                                                   | 99%                                            | 278       |
| Photocatalytic degradation (Fe-TiO <sub>2</sub> )                                                    | 99.9%                                          | 279       |
| Photocatalytic degradation (Fe-TiO <sub>2</sub> /rGO)                                                | 99.9%                                          | 279       |
| Photocatalytic degradation (MIL-100(Fe)/PDI)                                                         | 100%                                           | 280       |
| Photodegradation                                                                                     | 50–52%                                         | 281       |
| Photodegradation                                                                                     | 58%                                            | 274       |
| UV irradiation                                                                                       | 270 ± 6/10 <sup>-3</sup> min <sup>-1</sup> 55% | 282       |
| UV irradiation                                                                                       | 25–52%                                         | 283       |
| UV irradiation                                                                                       | 75–98%                                         | 281       |

at room temperature, with a pH of 7, a dosage of 3 g L<sup>-1</sup> for PMS, and a dosage of 0.3 g L<sup>-1</sup> for MCS.<sup>267</sup> According to Dolatabadi *et al.* (2021), under optimal conditions—namely, a BPA concentration of 3.25 mg L<sup>-1</sup>, a current density of 12.0 mA cm<sup>-2</sup>, a pH of 8.5, a reaction time of 23 minutes, and an electrical energy consumption of 0.308 kW h m<sup>-3</sup>—the maximum removal efficiency of BPA from the aqueous environment reached 98.2%.<sup>260</sup> The efficacy of electrocoagulation employing iron and aluminum electrodes for eliminating phenol from aqueous surroundings was investigated. The maximum removal efficiencies of phenol were achieved at 60 V, 80 minutes, 5 mg L<sup>-1</sup>, 3000 μS cm<sup>-1</sup>, and pH levels of 5 and 7 with aluminum and iron electrodes, respectively. The removal efficiencies were reported at 94.72% for aluminum electrodes and 98.0% for iron electrodes, respectively.<sup>268,269</sup> Ambauen *et al.* (2020) reported a 99% removal of BPA within the experimental duration of 240 minutes using both tested anode materials (Pt and BDD) and at both tested temperatures (6 and 20 °C), with a current efficiency applied at 86 mA cm<sup>-2</sup>, and BDD where 43 mA cm<sup>-2</sup> was employed. Parastar *et al.* (2018a)<sup>270</sup> reported that under the optimal conditions (10 mg L<sup>-1</sup>, 1500 mA, pH 7, and 0.5 cm), the removal efficiency observed varied between 60% and 90%. This research study concludes that employing the electrocoagulation process with iron electrodes under these optimal conditions can effectively remove BPA from aqueous solutions.

## 6.2 Chemical processes

In wastewater treatment, the primary treatment mostly functions by the sorption of BPAs onto sludge *via* coagulation/flocculation and subsequent sedimentation.<sup>271</sup> However, chemical treatment is also equipped to enhance the BPA removal efficiency. Some of the treatments include ultraviolet irradiation and ozonation,<sup>222</sup> which are summarized in Table 7. Sulfate radical (SO<sub>4</sub><sup>•-</sup>)-based advanced oxidation processes (AOPs) have gained considerable interest in recent years for their capacity to effectively remove recalcitrant organic molecules, including BPA, in water and wastewater due to their

strong oxidation properties. SO<sub>4</sub><sup>•-</sup> has a high standard redox potential (2.5–3.1 V) at a neutral pH and shows greater sensitivity towards certain organic contaminants.<sup>284</sup> SO<sub>4</sub><sup>•-</sup> can be generated by activating persulfate (PS) or peroxymonosulfate (PMS) using many processes including heat, transition metals, alkaline conditions, and ultraviolet (UV) irradiation.<sup>282</sup> The activation by UV irradiation was also considered to be a very successful and promising technique.<sup>285</sup> The degradation rate of BPA by UV irradiation reached 270 ± 6/10<sup>-3</sup> min<sup>-1</sup> with 55% removal from 0.44 mmol L<sup>-1</sup> initial concentration.<sup>282</sup> Furthermore, the ozonation method, which involves a redox mechanism, also has several advantages including high removal efficiency, non-selectivity, and disinfection effects.<sup>222</sup> BPA can be destroyed by a direct reaction with molecular ozone or indirect reaction with hydroxyl radicals, which are generated from the alkaline breakdown of ozone. Hydroxyl radicals are a more powerful oxidant with lesser selectivity, as compared to molecular ozone. Umar *et al.*<sup>272</sup> reported that ozonation achieved up to 100% of BPA removal from water with the initial concentration of a maximum of 0.5 mg L<sup>-1</sup>. Out of all the methods, photocatalytic degradation offers an unlimited potential for BPA degradation, especially when it comes to the choice of the nanocomposite materials.<sup>279</sup> Table 7 shows that the photocatalytic degradation of BPA reached more than 90% efficiencies and can be achieved within a short period.

## 6.3 Integrated processes

Integration treatment was used to enhance the removal of contaminants from various contaminated water or to encounter issues that arise with the single treatment. Integrated or hybrid systems were developed to meet the regulations, such as those, for law enforcement removing substances and producing drinkable water. Another advantage of hybrid systems is their ability to address the limitations of procedures by incorporating alternative methods.<sup>286</sup> Table 8 shows the integration treatment for the removal of BPA. The integration treatment involved a combination of physical, chemical, and biological processes and also the sequential of the single process. From Table 8,

Table 8 Previous studies on the integration process involved in the removal of BPA

| Integration             | Process                                                                                                                                                        | Performance removal                                    | Reference |
|-------------------------|----------------------------------------------------------------------------------------------------------------------------------------------------------------|--------------------------------------------------------|-----------|
| Physical and chemical   | Ceramic membranes incorporating manganese-cobalt oxides (MCOs) to catalyze the ozone-based degradation                                                         | 90.6%                                                  | 287       |
|                         | Membrane filtration coupled with advanced oxidation processes (AOPs) and adsorption. Ceramic membrane filtration coupled with PMS activation and adsorption    | —                                                      | 288       |
|                         | Ultrafiltration-catalysis membrane                                                                                                                             | 95%                                                    | 289       |
|                         | Adsorption and catalytic ozonation: MWCNTs/Fe <sub>3</sub> O <sub>4</sub>                                                                                      | 25 to 75 ppm (98%)                                     | 290       |
|                         | Adsorption and biodegradation                                                                                                                                  | 98.1% at 5 mg per L BPA                                | 291       |
| Physical and biological | Membrane-assisted enzyme treatment                                                                                                                             | 89% removal of BPA                                     | 292       |
|                         | Fenton process and <i>Phanerochaete chrysosporium</i>                                                                                                          | 58.23% at 1 mg per L BPA                               | 293       |
| Chemical and biological | MBR and ozonation                                                                                                                                              | 90% removal of BPA 80 min at 8 mg O <sub>3</sub> per L | 294       |
|                         | An up-flow anaerobic sludge blanket (UASB), submerged aerated biological filters (SABF), and horizontal subsurface flow constructed wetland (HSSF-CW) reactors | 99%                                                    | 295       |

most integration studies showed a great removal of BPA. However, the integration of chemical and biological processes represents only 58.23% of removal. However, this removal percentage surpasses the treatment of 21.59% and 14.48% achieved with *Phanerochaete chrysosporium* alone and the Fenton treatment, respectively.

According to Tarafdar *et al.*, (2022), the effective strategies that can be employed to eliminate BPA are biodegradation and adsorptive processes.<sup>296</sup> However, there is a need to address the issues of increased adsorption ability and rapid adsorption rates for both conventional and non-conventional adsorbents.<sup>297</sup> Abu Hasan *et al.*, (2023) reviewed the proposed biological treatment and adsorption for BPA concentrations in the effluent that exceeds 1 mg L<sup>-1</sup>.<sup>286</sup> For concentrations below 1 mg L<sup>-1</sup>, the utilization of an advanced oxidation process (AOP) and MF may be more appropriate, enabling load reduction and consequently decreasing operational and maintenance expenses. Godiya and Park (2022)<sup>297</sup> stated that the integration of membrane separation with additional oxidation/biodegradation methods has the potential to mitigate contamination and mass transfer constraints. According to Zielińska *et al.* (2019), the integration of chemical and biological gained widespread interest in removing BPA from contaminated water.<sup>298</sup> These methods primarily involve the integration of Advanced Oxidation Processes (AOPs) such as photocatalysis, ozone oxidation, Fenton and photo-Fenton oxidation, and wet air oxidation (WAO) supplementing the conventional biological treatment. Fundamentally, an effective treatment approach for the elimination of particular EDCs should involve the integration and optimization of sophisticated physical, chemical, and

biological treatment techniques.<sup>299</sup> It is essential to conduct an exploration and investigation of every advanced treatment process to identify any unidentified potential limitations.

## 7. Breakdown/removal mechanisms of BPA in water

The remediation process of BPA present in water can be understood from three points of view. The first involves the biodegradation mechanism, the second corresponds to the removal of mass transfer (adsorption) and the third is the degradation of the molecule by advanced oxidative processes.

### 7.1 Adsorption mechanism

The adsorption of BPA and the mechanisms that govern this process depend on some factors including the nature of the pollutant, the chemical and physical properties of the solid, and the adsorption results. The morphology and structure analyzed through characterization techniques (adsorbent) are also necessary to determine the mechanism.<sup>300</sup> The coprecipitation method was used to develop magnetic nanoparticles decorated with graphene oxide nanosheets, and this material was the BPA adsorbent.<sup>301</sup> The main interaction that occurred between the graphene skeleton and the benzene ring of the pollutant was  $\pi$ - $\pi$ . In another study, graphene-coated magnetic iron oxide nanoparticles were used to remove BPA, and the fabricated aerogel (Fe<sub>3</sub>O<sub>4</sub>/GE) was subsequently reduced to ethylenediamine.<sup>302</sup> The interactions were  $\pi$ - $\pi$ , hydrogen bonding, and electrostatic, both present between the adsorbate ring and the conjugated carbon network (Fig. 3). The Fe<sub>3</sub>O<sub>4</sub>/GO adsorbent



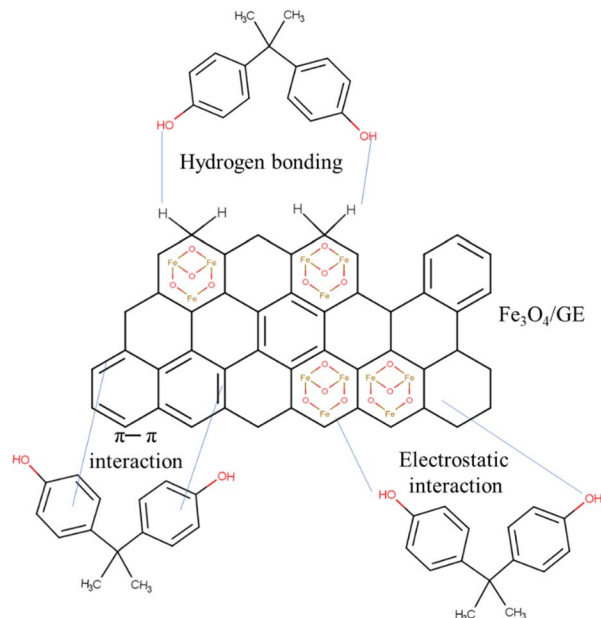


Fig. 3 Representation of the BPA adsorption mechanism onto graphene coated with  $\text{Fe}_3\text{O}_4$  nanoparticles.

was obtained by the hydrothermal method where the BPA ring interacted with the surface *via*  $\pi$ - $\pi$  interactions, and  $\text{sp}^3$  hybridization also corroborated the remediation process.<sup>303</sup> In summary, we can see that the nature of the interactions can be widely varying. However, BPA does not breakdown *per se* but is attached to the adsorbent surface *via* these physico-chemical interactions and thus removed.

## 7.2 Membrane/nano-filtration mechanism

Interactions such as hydrophobicity, hydrogen bonding, and electrostatic interactions are essential in the remediation of BPA in membrane filters or even nanofilters. One study used a synthetic effluent with BPA, which was remediated in a submerged membrane bioreactor.<sup>304</sup> The researchers reported that almost complete removal (99%) was obtained at a low volumetric loading rate (0–21.6 g per  $\text{m}^3$  per day). Biodegradation and adsorption mechanisms were responsible for the efficient result for the membrane use. The adsorption process showed endothermic behavior favorable to high temperatures and hydrophobic character, with van der Waals and electrostatic interactions. These were present between the BPA molecules and the reactor sludge. Electrospinning was used to obtain a 6,6-nanofibrous membrane for BPA remediation.<sup>305</sup> The maximum capacity of  $91 \text{ mg g}^{-1}$  was obtained *via* hydrogen interactions between the adsorbate hydroxyl and the membrane carbonyl and hydrophobic interactions. BPA is nanoscale in size, so size exclusion through the nanofibrous membrane could not support any removal. In another study, a nanofibrous membrane carrying the laccase enzyme was synthesized on polyethersulfone and polyethersulfone membranes to remove BPA.<sup>306</sup> The hydrophilic nature of the laccase membrane enabled 90% removal, being governed by hydrogen bonds of the

amine group present in the membrane with BPA. Hydrophobic interaction between polyethyleneimine/polyethersulfone and the adsorbate and electrostatic interaction with the cationic group of the polyethyleneimine/polyethersulfone and the pollutant molecule. The mechanism of reverse osmosis and nanofiltration membranes was analyzed concerning BPA removal.<sup>307</sup> The authors highlight that membranes made of polyamide present greater rejection of BPA when compared to those made of cellulose acetate. The mechanisms correspond to hydrogen bonding, electrostatic, and hydrophobic interaction (Fig. 3).

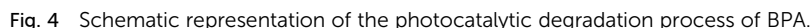
## 7.3 Photocatalytic degradation and its mechanism

The photodegradation of BPA involves mechanisms that correspond to the generation of electron holes present in the conduction band and also the electrons contained in the valence zone after the irradiation step.<sup>308</sup> Studies have concluded that depending on the type of photocatalyst, the oxidation of the valence band and hydroxyls are mainly responsible for the degradation.<sup>309</sup> Therefore, the hydroxyl initiates the degradation by directly attacking the carbon present in the phenyl group, and then 4-isopropylphenyl and hydroquinone are generated as intermediates. In the case of 4-isopropylphenyl, it is again converted to 4-(2-hydroxypropan-2-yl) phenol, and then, after hydroxylation and dehydration, it generates 4-hydroxybenzoic acid and 4-hydroxybenzaldehyde (Fig. 4). Researchers describe that the products can also undergo oxidation producing various organic acids that are also degraded, finally forming products that are not toxic such as water and  $\text{CO}_2$ .<sup>310,311</sup>

## 7.4 Advanced oxidation processes and its mechanism

Due to its properties, ozone is highly oxidizing and is present in the oxidation of several organic compounds, such as BPA. The reactions can be indirect or direct. The oxidation process of BPA and molecular ozone occurs in a basic environment through an indirect reaction, while in an acidic environment, decomposition and oxidation occur through the elimination of radicals in a direct reaction. Molecular ozone also attacks structures rich in electrons (non-protonated amines, activated aromatic compounds, and unsaturated bonds such as  $\text{C}=\text{C}$ ). The presence of phenolic groups increases the chances of a reaction. In this sense, degradation *via* ozonation obtained complete decomposition and 30% mineralization of BPA under acidic conditions.<sup>312,313</sup> After 25 min, the stoichiometric ratio between ozone and BPA was 10.3. The production of ozonoids and the attack of ozone on  $\text{C}=\text{C}$  were the main mechanisms involved. Malonic and oxalic acids accelerate the degradation of ozonides, corroborating with simultaneous mineralization. A nanocomposite ( $\text{AC}/\text{CeO}_2/\text{ZnO}$ ) was synthesized to remove BPA.<sup>314</sup> Under optimal conditions, total carbon removal was observed (time of 60 min, pH of 8, and  $500 \mu\text{g per L}$  catalyst dose) with the use of five degradation products. OH radicals were dominant in the process and increased under basic conditions. The generation of organic acids with the addition of OH reduced the pH. Under pH conditions of 11, BPA removal was 100% using





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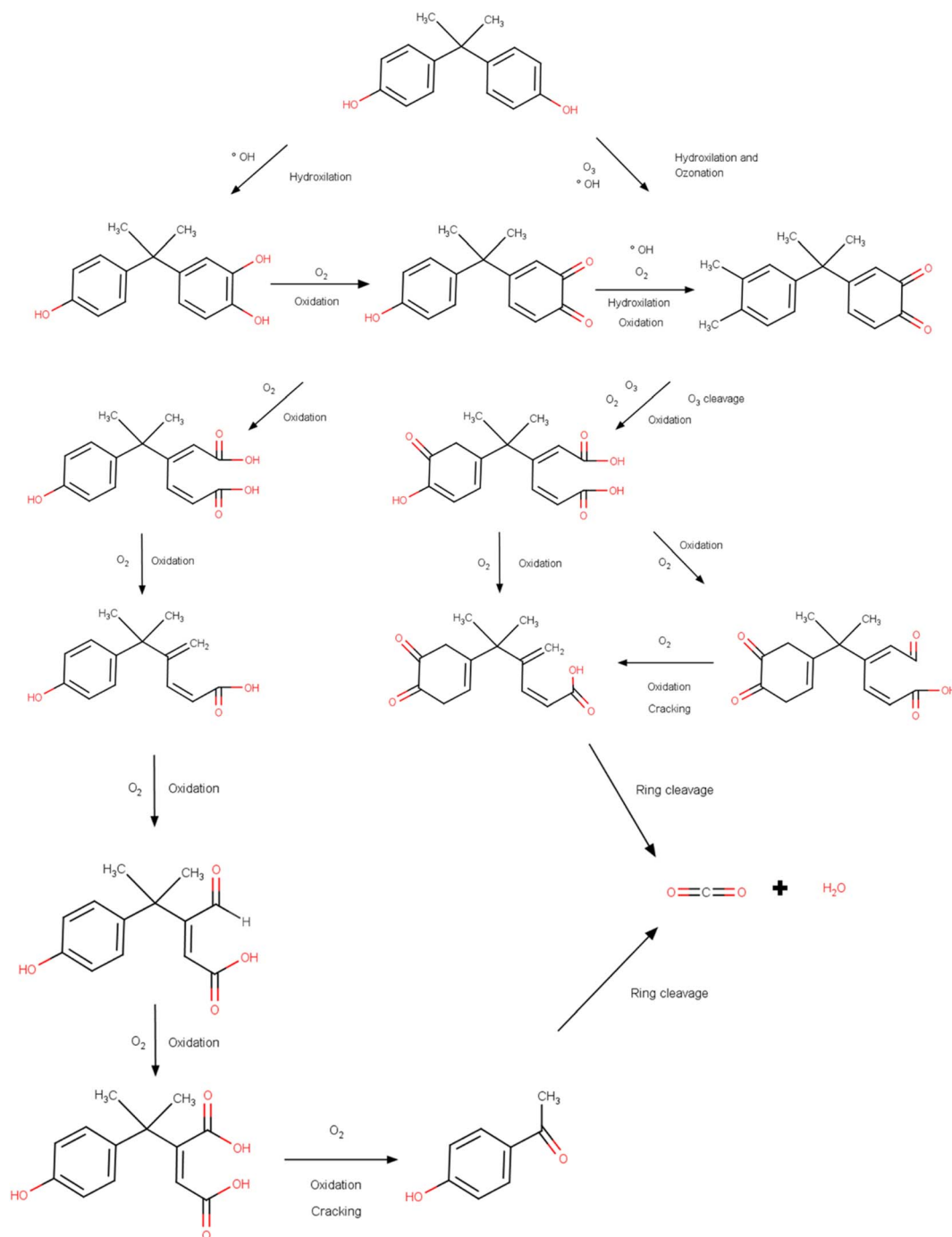


Fig. 5 BPA decomposition mechanism in the presence of  $\text{Fe}_3\text{O}_4\text{-MnO}_2$  and ozone.

## 7.6 Photo-fenton process and its mechanism

The BPA degradation mechanism through the photo-fenton process has been reported by different authors. da Silveira Salla *et al.*<sup>322</sup> have proposed the mechanism of BPA in the presence of  $\text{CuFeS}_2$  and visible light, as given in Fig. 7. First, the BPA is converted into hydroxylated bisphenol, due to the  $\cdot\text{OH}$  radicals. After that the  $\cdot\text{OH}$  radicals tend to interact once again by promoting the BPA bridge cleavage. Leading to the formation of 4-isopropylphenol, 4-hydroxyacetophenone, 2-

methoxyhydroquinone, and 4-(1-hydroxy-1-methylethyl)benzene-1,2-diol. The later intermediates suffer from another hydroxylation, which leads to the ring opening and the formation of acids in the form of oxalic, fumaric, and isobutyric acids. Last, the acids are mineralized forming  $\text{CO}_2$  and  $\text{H}_2\text{O}$ . Other differences generated on the mechanism pathway can be attributed to the catalyst employed. Zhu *et al.* (2018) indicate that the catalyst based on  $\text{Ag}/\text{AgCl}$  nanoparticles onto ferrite can produce electron-hole pairs due to the surface plasmon resonance.<sup>29</sup> These electron pairs are further transferred to the

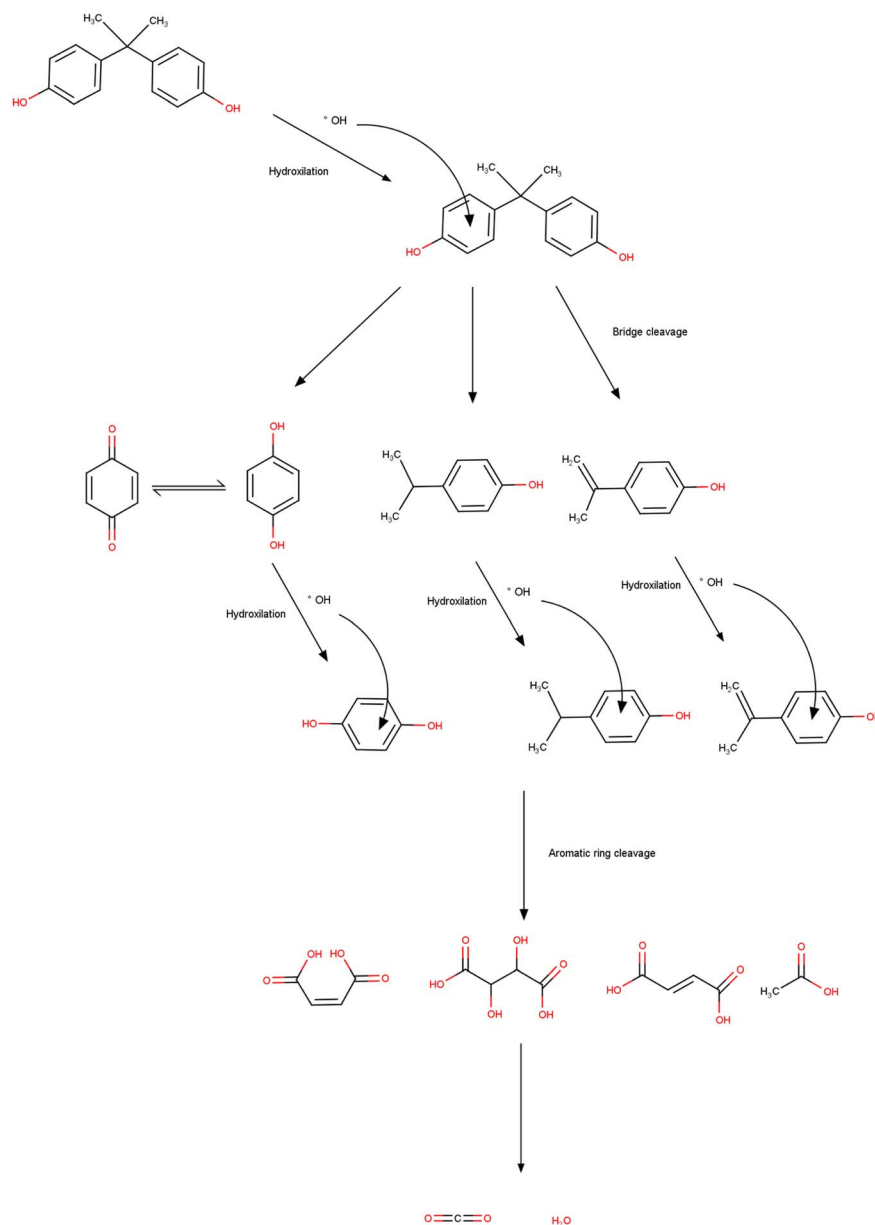


Fig. 6 BPA electrochemical degradation routes.

ferrite, which increases the conversion of  $\text{Fe}^{3+}$  into  $\text{Fe}^{2+}$ . The latter also affects the number of radicals produced due to  $\text{Fe}^{2+}$  reacting with the peroxide ( $\text{H}_2\text{O}_2$ ). The same mechanism was also confirmed by Liu *et al.*<sup>323</sup> when they employed Ag/AgCl/Ferrite-S, indicating that the electron-hole pair generation migrates to the conduction band before acting in the  $\text{Fe}^{3+}$ -to- $\text{Fe}^{2+}$  conversion mechanism.

### 7.7 Biodegradation and its mechanism

Following the other mechanism trends, the degradation of BPA has similar routes regarding the hydroxylation and bridge/aromatic cleavage; however, new routes tend to emerge due to the microorganism present. The removal of BPA in

biodegradation mechanisms involves aerobic and anaerobic conditions, together with the action of microorganisms. The use of *Bacillus* sp. cultivated in anaerobic and aerobic media was applied in the remediation of BPA.<sup>324</sup> In this study, the strain that was extracted obtained the best performance with approximately 51% mineralization using  $5 \text{ mg L}^{-1}$  of pollutant concentration. The aerobic process was the most efficient, generating approximately 5 final products. The entire process was dominated by electron acceptors such as  $\text{Fe}^{3+}$ ,  $\text{SO}_4^{2-}$ , and  $\text{NO}_3^-$  for the anaerobic system, while the acceptors of the aerobic system were  $\text{O}_2$ . The acceptors  $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$  did not influence biodegradation; however,  $\text{Fe}^{3+}$  was the most efficient when added to the anaerobic system. Despite the good performance, the rate was not higher than that obtained in the system



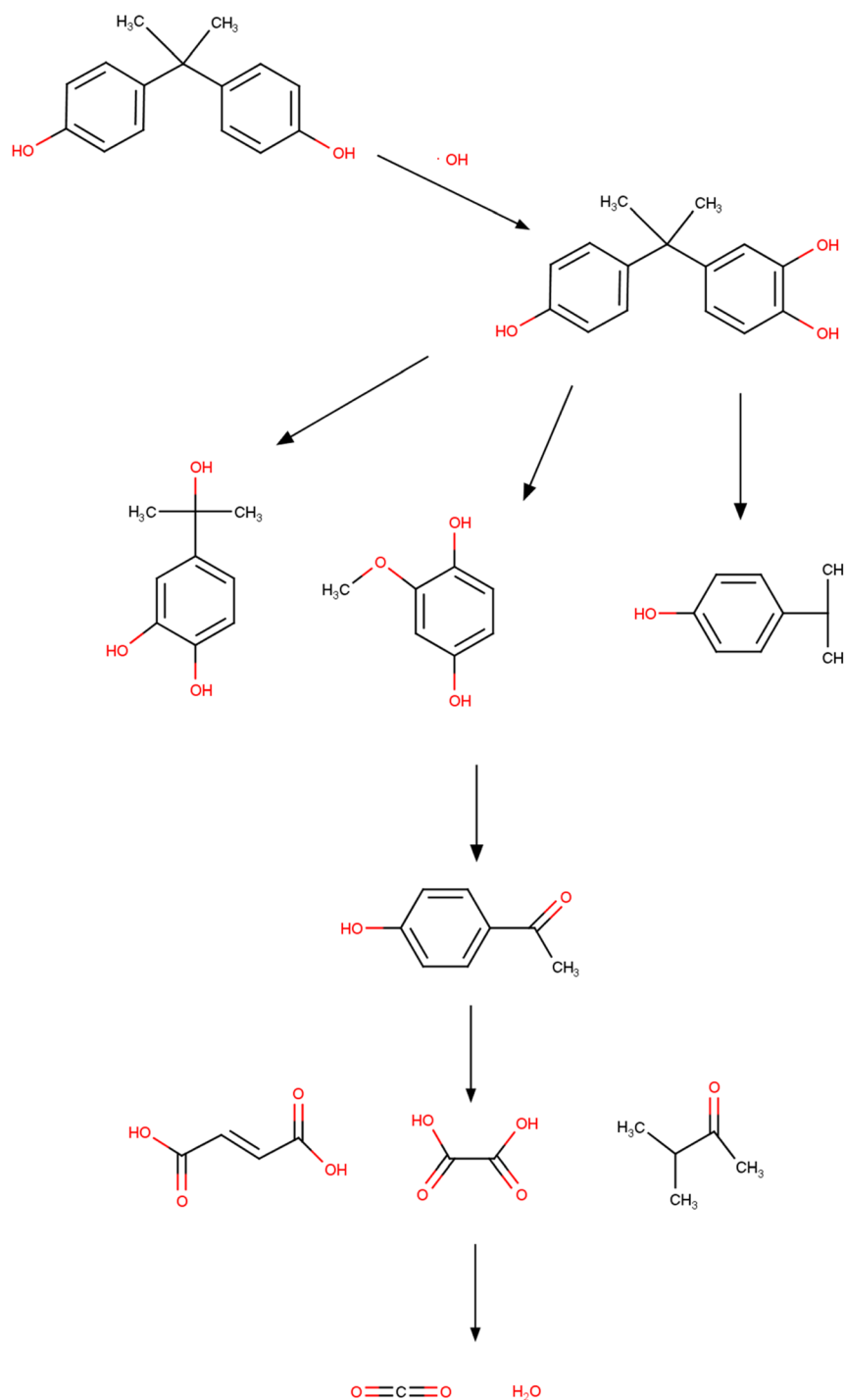


Fig. 7 BPA degradation mechanism for the photo-Fenton reaction.

containing oxygen. In another study, the strain YC-JY1 of *Sphingobium* sp. was obtained to remove BPA; the results show that at a pH of 6.5, degradation was maximum.<sup>325</sup>

Golshan *et al.*<sup>326</sup> investigated the BPA degradation mechanism in the presence of several microorganisms. From all the investigated microorganisms, the author evaluated *Sphingomonas* sp. as given in Fig. 8. At a temperature of 30 °C, the process

involved two mechanisms, the first of which corresponds to the conversion of the pollutant into 1,2-bis(4-hydroxyphenyl)-2-propanol, which is then transformed into 4-hydroxybenzaldehyde and 4'-hydroxyacetophenone, for final use by the bacteria, and the second mechanism corresponds to the conversion of BPA, sequentially generating the accumulation of conversion products (process medium). These products

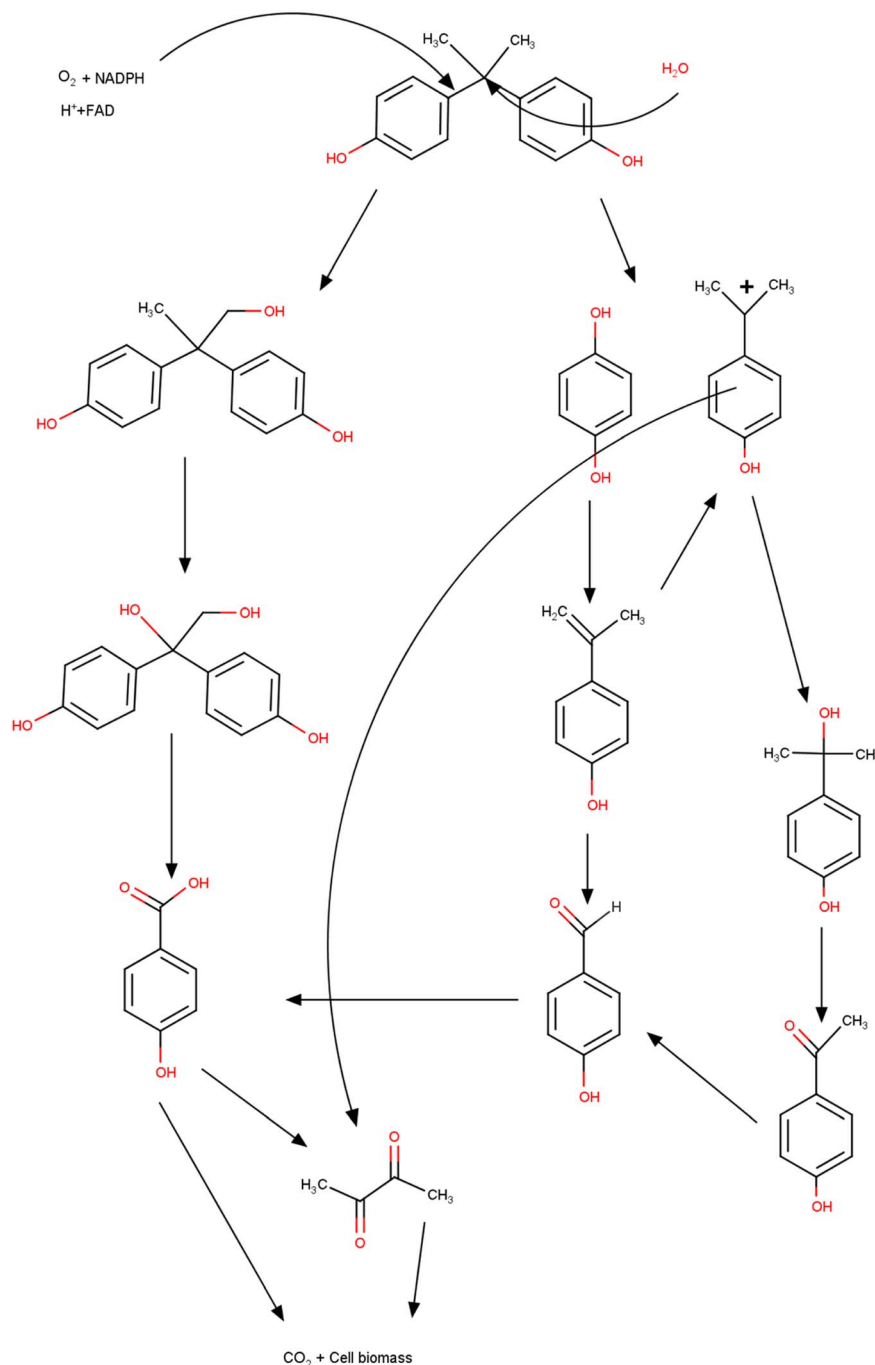


Fig. 8 BPA biodegradation routes with *Sphingomonas* sp. present in the solution.

correspond to the same ones obtained in the degradation and photocatalytic biotransformation of BPA described previously. The addition of salt-tolerant species corresponds to the biological treatment carried out with the concentration of  $50 \text{ mg L}^{-1}$  of pollutant, contained in a sequential batch reactor with saline wastewater.<sup>326</sup>

Similarly, two biodegradation processes that correspond were also found by Wang *et al.*<sup>327</sup> The first is the cleavage of the C=C bond and hydroxylation. In the latter, BPA is first transformed into 2,3-bis(4-hydroxyphenyl)-1,2-propanediol and 2,2-

bis(4-hydroxyphenyl)-1-propanol, which are mineralized in cell biomass and carbon dioxide. The second corresponds to the cleavage step of the C=C bond, where degradation begins in the benzene ring and oxidative cleavage, and aliphatic acids are subsequently generated, which are mineralized in cell biomass and carbon dioxide. In another study, green algae *Desmodesmus* sp. was used to detoxify and biodegrade BPA.<sup>327</sup> The mechanisms encompass three dynamics. The first corresponds to the oxidative cleavage of the pollutant to monophenols, followed by their mineralization. The second corresponds to hydroxylation



generating monohydroxybisphenol-A, followed by hydroxymethylation forming 2-hydroxy-3-hydroxymethylbisphenol-A. The third is glycosylation (a biochemical process that consists of the addition of sugars to proteins, lipids, or other organic compounds, resulting in the formation of glycoconjugates).

## 8. Challenges and future directions

When searching the phrase 'BPA', it is observed that it is a highly studied compound and that it presents a large number of studies, ranging from its presence in the environment, entry routes, and toxic potential in animals, humans, aquatic life and plants, to biological, physical, and chemical processes for their environmental mitigation. All of these efforts and interests are related to their constant presence and release into the environment, due to the means of production and the contemporary culture of consumerism. When analyzing toxic effects, a vast field is observed, where different pathways in the human body can be affected, where it presents a strong correlation with other factors, such as concentration, exposure time, and route of contact. Another aspect is the age at which humans come into contact with BPA. In this sense, it is clear that babies and children in the development phase are the most affected.

Based on this, the implicit mechanisms require better understanding and more investigations, this is one of the challenges that BPA remediation processes face. More reliable technologies for detecting and measuring plastic exposure are also needed. The complexity lies in the different routes of exposure due to its presence in various products, so it can be through dermal contact or oral or olfactory ingestion. Experiencing the availability of highly accurate methods of measuring exposure makes it possible to estimate the level of danger and determine the best interventions to achieve high efficiency in reducing damage. More studies are also needed to analyze and describe the damage along the food chain and prolonged exposure to low concentrations of BPA. Although it is possible to observe studies that show that low concentrations do not present a health risk, other researchers highlight the cumulative potential of BPA over years of exposure. Only by understanding long-term damage and at low concentrations is it possible to establish safe guidelines and regulations by authorities through competent bodies. Among the countless challenges to be overcome, there is also the field that seeks to develop new, more sustainable materials that can replace BPA in the most diverse consumer products. The points observed by the researchers are economic viability and health risk, an example being vegetable polymers. Even if BPA were completely replaced, its presence in the environment would persist for a long time; therefore, more studies are needed to describe its impacts at the aquatic and plant level. These studies must go hand in hand with new BPA mitigation strategies.

Bacterial degradation has been shown to be 45–100% efficient depending on the strain of bacteria and the choice of process parameters. To obtain total optimization of the biodegradation process using bacteria, it is necessary to carry out continuous investigations in different regions of the world, as meteorological variables are different and have a high

influence on colony growth. The seasons must also be taken into account as well as the presence of other chemical molecules in addition to BPA. Regarding other remediation strategies, BPA can be removed from water by physical processes such as membrane filtration, adsorption, and coagulation and by chemical processes such as ozonation and photocatalytic degradation. We see that membrane processes can remove 40–95% BPA in water with a high level of throughput. Other physical processes such as adsorption and coagulation can remove up to 100% BPA in water but throughput is significantly lesser. Chemical processes such as ozonation and photocatalytic degradation have achieved similar performance levels. However, we believe the use of hybrid/integrated processes can simultaneously achieve high removal efficiency at high water throughput due to the intensification of the process. It is encouraged that researchers explore various hybrid technologies for BPA removal from water with the aim of achieving process efficiency and scalability.

## 9. Conclusion

A detailed analysis of the literature was conducted to determine the toxicity levels and concentration of BPA in environment-quality plastic packaging. The degradation process using different species of bacteria made it possible to describe the damage to humans, animals, and plants that BPA can cause through prolonged exposure, and the role that these organisms play. It was observed that food-grade packaging contains considerable concentrations of BPA, a utilization route that had inadvertently led to biomagnification and its ubiquitous nature in the aqueous environment. This has increased the need for remediation techniques to reduce its incidence in the aqueous environment. Researchers should focus their efforts on providing a better understanding of the mechanisms of the degradation of BPA and its toxic effects on the ecosystem. It was observed that physical processes are able to achieve >90% removal efficiency of BPA from water but tend to perform better under a low pollutant load in the influent stream. Chemical processes show wider variation in performance with 20–100% removal efficiency, while integrated processes show 90–100% removal efficiency of BPA from water. It is highly necessary to obtain new routes to replace BPA in various consumer products, without it losing its quality. Therefore, it is highly important to explore new solutions by developing efficient and viable strategies for plastic waste management. In soil, BPA can be effectively remediated by bacterial degradation. Aqueous phase remediation of BPA can be achieved with processes such as membrane filtration, adsorption, coagulation, ozonation, and photocatalytic degradation. It was proposed that hybrid/integrated processes are a strategy that can help achieve high removal efficiency at high wastewater throughput.

## Ethical statement

This article does not contain any studies involving human or animal subjects.



## Data availability

No primary research results, software or code has been included and no new data were generated or analysed as part of this review.

## Conflicts of interest

The authors declare that there are no conflicts of interest.

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Fig. 3–8 were designed using the MarvinSketch software.

## References

- 1 E. B. Jadhav, M. S. Sankhla, R. A. Bhat and D. S. Bhagat, Microplastics from food packaging: An overview of human consumption, health threats, and alternative solutions, *Environ. Nanotechnol., Monit. Manage.*, 2021, **16**, 100608.
- 2 A. Wanag, A. Sienkiewicz, P. Rokicka-Konieczna, E. Kusiak-Nejman and A. W. Morawski, Influence of modification of titanium dioxide by silane coupling agents on the photocatalytic activity and stability, *J. Environ. Chem. Eng.*, 2020, **8**(4), 103917.
- 3 C. Monneret, What is an endocrine disruptor?, *C. R. Biol.*, 2017, **340**(9–10), 403–405.
- 4 E. B. Jadhav, M. S. Sankhla, R. A. Bhat and D. S. Bhagat, Microplastics from food packaging: An overview of human consumption, health threats, and alternative solutions, *Environ. Nanotechnol., Monit. Manage.*, 2021, **16**, 100608.
- 5 E. R. Kabir, M. S. Rahman and I. Rahman, A review on endocrine disruptors and their possible impacts on human health, *Environ. Toxicol. Pharmacol.*, 2015, **40**(1), 241–258.
- 6 C. Sonnenschein and A. M. Soto, An updated review of environmental estrogen and androgen mimics and antagonists, *J. Steroid Biochem. Mol. Biol.*, 1998, **65**(1–6), 143–150.
- 7 H. R. Lee, E. B. Jeung, M. H. Cho, T. H. Kim, P. C. K. Leung and K. C. Choi, Molecular mechanism(s) of endocrine-disrupting chemicals and their potent oestrogenicity in diverse cells and tissues that express oestrogen receptors, *J. Cell. Mol. Med.*, 2013, **17**(1), 1–11.
- 8 J. J. Heindel, B. Blumberg, M. Cave, R. Machtinger, A. Mantovani, M. A. Mendez, *et al.*, Metabolism disrupting chemicals and metabolic disorders, *Reprod. Toxicol.*, 2017, **68**, 3–33.
- 9 *Marine Anthropogenic Litter*, ed. M. Bergmann, L. Gutow, M. Klages, M. Anthropogenic Litter, M. Bergmann, L. Gutow and M. Klages, Springer International Publishing, Cham, 2015, pp. 1–447.
- 10 J. N. Hahladakis, C. A. Velis, R. Weber, E. Iacovidou and P. Purnell, An overview of chemical additives present in plastics: Migration, release, fate and environmental impact during their use, disposal and recycling, *J. Hazard. Mater.*, 2018, **344**, 179–199.
- 11 T. S. Horan, H. Pulcastro, C. Lawson, R. Gerona, S. Martin, M. C. Gieske, *et al.*, Replacement Bisphenols Adversely Affect Mouse Gametogenesis with Consequences for Subsequent Generations, *Curr. Biol.*, 2018, **28**(18), 2948–2954.
- 12 D. Chen, K. Kannan, H. Tan, Z. Zheng, Y. L. Feng, Y. Wu, *et al.*, Bisphenol Analogues Other Than BPA: Environmental Occurrence, Human Exposure, and Toxicity - A Review, *Environ. Sci. Technol.*, 2016, **50**(11), 5438–5453.
- 13 Q. Li, F. Pan, W. Li, D. Li, H. Xu, D. Xia, *et al.*, Enhanced adsorption of bisphenol A from aqueous solution with 2-vinylpyridine functionalized magnetic nanoparticles, *Polymers*, 2018, 1136.
- 14 G. K. Deshwal and N. R. Panjagari, Review on metal packaging: materials, forms, food applications, safety and recyclability, *J. Food Sci. Technol.*, 2020, **57**(7), 2377–2392.
- 15 C. Bach, X. Dauchy, M. C. Chagnon and S. Etienne, Chemical compounds and toxicological assessments of drinking water stored in polyethylene terephthalate (PET) bottles: A source of controversy reviewed, *Water Res.*, 2012, **46**(3), 571–583.
- 16 R. Proshad, T. Kormoker, M. S. Islam, M. A. Haque, M. M. Rahman and M. M. R. Mithu, Toxic effects of plastic on human health and environment : A consequences of health risk assessment in Bangladesh, *Int. J. Health*, 2017, **6**(1), 1–5.
- 17 K. V. S. Rajmohan, C. Ramya, M. Raja Viswanathan and S. Varjani, Plastic pollutants: effective waste management for pollution control and abatement, *Curr. Opin. Environ. Sci. Health.*, 2019, **12**, 72–84.
- 18 C. D. Kassotis, L. N. Vandenberg, B. A. Demeneix, M. Porta, R. Slama and L. Trasande, Endocrine-disrupting chemicals: economic, regulatory, and policy implications, *Lancet Diabetes Endocrinol.*, 2020, **8**(8), 719–730.
- 19 O. O. Ayeleru, S. Dlova, O. J. Akinribide, F. Ntuli, W. K. Kupolati, P. F. Marina, *et al.*, Challenges of plastic waste generation and management in sub-Saharan Africa: A review, *Waste Manage.*, 2020, **110**, 24–42.
- 20 O. Banerjee, S. Singh, I. Saha, S. Pal, M. Banerjee, S. Kundu, *et al.*, Molecular dissection of cellular response of pancreatic islet cells to Bisphenol-A (BPA): A comprehensive review, *Biochem. Pharmacol.*, 2022, **201**, 115068.
- 21 T. S. Horan, H. Pulcastro, C. Lawson, R. Gerona, S. Martin, M. C. Gieske, *et al.*, Replacement Bisphenols Adversely Affect Mouse Gametogenesis with Consequences for Subsequent Generations, *Curr. Biol.*, 2018, **28**(18), 2948–2954.
- 22 M. Y. Lee, I. Ahmed, K. Yu, C. S. Lee, K. K. Kang, M. S. Jang, *et al.*, Aqueous adsorption of bisphenol A over a porphyrinic porous organic polymer, *Chemosphere*, 2021, **265**, 129161.
- 23 T. Geens, L. Goeyens, K. Kannan, H. Neels and A. Covaci, Levels of bisphenol-A in thermal paper receipts from Belgium and estimation of human exposure, *Sci. Total Environ.*, 2012, **435–436**, 30–33.



- 24 B. E. Tokula, A. O. Dada, A. A. Inyinbor, K. S. Obayomi, O. S. Bello and U. Pal, Agro-waste based adsorbents as sustainable materials for effective adsorption of Bisphenol A from the environment: A review, *J. Cleaner Prod.*, 2023, **388**, 135819.
- 25 S. Karuppaiah, R. Annamalai, A. Muthuraj, S. Kesavan, R. Palani, S. Ponnusamy, *et al.*, Efficient photocatalytic degradation of ciprofloxacin and bisphenol A under visible light using Gd<sub>2</sub>WO<sub>6</sub> loaded ZnO/bentonite nanocomposite, *Appl. Surf. Sci.*, 2019, **481**, 1109–1119.
- 26 A. Selvakumar and S. Rangabhashiyam, Biosorption of Rhodamine B onto novel biosorbents from *Kappaphycus alvarezii*, *Gracilaria salicornia* and *Gracilaria edulis*, *Environ. Pollut.*, 2019, **255**, 113291.
- 27 O. Onyekwere, C. J. Okonkwo, A. B. Okoroafor and C. J. Okonkwo, Occurrence and risk assessment of phenolic endocrine disrupting chemicals in shallow groundwater resource from selected Nigerian rural settlements, *Ovidius Univ. Ann. Chem.*, 2019, **30**(2), 101–107.
- 28 J. N. Okereke, O. I. Ogidi and K. O. Obasi, International Journal of Advanced Research in Biological Sciences Environmental and Health Impact of Industrial Wastewater Effluents in Nigeria-A Review, *Int. J. Adv. Res. Biol. Sci.*, 2016, **3**(6), 55–67.
- 29 H. Zhu, Z. Li and J. Yang, A novel composite hydrogel for adsorption and photocatalytic degradation of bisphenol A by visible light irradiation, *Chem. Eng. J.*, 2018, **334**, 1679–1690.
- 30 J. Corrales, L. A. Kristofco, W. Baylor Steele, B. S. Yates, C. S. Breed, E. Spencer Williams, *et al.*, Global assessment of bisphenol a in the environment: Review and analysis of its occurrence and bioaccumulation, *Dose-Response*, 2015, **13**(3), 1–29.
- 31 S. Karuppaiah, R. Annamalai, A. Muthuraj, S. Kesavan, R. Palani, S. Ponnusamy, *et al.*, Efficient photocatalytic degradation of ciprofloxacin and bisphenol A under visible light using Gd<sub>2</sub>WO<sub>6</sub> loaded ZnO/bentonite nanocomposite, *Appl. Surf. Sci.*, 2019, **481**, 1109–1119.
- 32 C. C. Wu, J. N. Shields, C. Akemann, D. N. Meyer, M. Connell, B. B. Baker, *et al.*, The phenotypic and transcriptomic effects of developmental exposure to nanomolar levels of estrone and bisphenol A in zebrafish, *Sci. Total Environ.*, 2021, **757**, 143736.
- 33 N. Shobnam, Y. Sun, M. Mahmood, F. E. Löffler and J. Im, Biologically mediated abiotic degradation (BMAD) of bisphenol A by manganese-oxidizing bacteria, *J. Hazard. Mater.*, 2021, **417**, 1–7.
- 34 M. Sarkhosh, M. Sadani, M. Abtahi, H. Azarpira, H. Alidadi, Z. Atafar, *et al.*, Photo-biological degradation of Bisphenol A, UV/ZnO/Iodide process at the center of biological reactor, *J. Photochem. Photobiol.*, A, 2019, **374**, 115–124.
- 35 R. Frankowski, J. Płatkiewicz, E. Stanisław, T. Grześkowiak and A. Zgoła-Grześkowiak, Biodegradation and photo-Fenton degradation of bisphenol A, bisphenol S and fluconazole in water, *Environ. Pollut.*, 2021, **289**, 117947.
- 36 J. Sharma, I. M. Mishra and V. Kumar, Mechanistic study of photo-oxidation of Bisphenol-A (BPA) with hydrogen peroxide (H<sub>2</sub>O<sub>2</sub>) and sodium persulfate (SPS), *J. Environ. Manage.*, 2016, 12–22.
- 37 G. A. Ashraf, R. T. Rasool, S. Al-Sulaimi, R. U. Rasool, N. Hassan, Z. Ajmal, *et al.*, Construction of type-II scheme SnO@HfC photocatalyst for bisphenol A degradation via peroxymonosulfate activation; DFT and self-cleaning analysis, *Chemosphere*, 2023, **341**, 140095.
- 38 F. Cheng and J. Wang, Removal of bisphenol a from wastewater by adsorption and membrane separation: Performances and mechanisms, *Chem. Eng. J.*, 2024, **484**, 149414.
- 39 A. Shishov, D. Israelyan and A. Bulatov, Automated deep eutectic solvent-based chromatomembrane microextraction: Separation and preconcentration of bisphenols from aqueous samples, *Sep. Purif. Technol.*, 2024, **338**, 126480.
- 40 R. J. Pan, J. Wu, J. Qu, T. Zhang, F. Z. Jiao, M. Zhao, *et al.*, Peak-like three-dimensional CoFe<sub>2</sub>O<sub>4</sub>/carbon nanotube decorated bamboo fabrics for simultaneous solar-thermal evaporation of water and photocatalytic degradation of bisphenol A, *J. Mater. Sci. Technol.*, 2024, **179**, 40–49.
- 41 L. Tian, C. G. Goodyer, J. Zheng and S. Bayen, Thermal degradation of bisphenol A and bisphenol S in water and fish (cod and basa) fillets, *Food Chem.*, 2020, **328**, 126999.
- 42 S. Tiwari, C. L. Gehlot, K. Srivastava and D. Srivastava, Simulation of the thermal degradation and curing kinetics of fly ash reinforced diglycidyl ether bisphenol A composite, *J. Indian Chem. Soc.*, 2021, **98**(6), 100077.
- 43 S. Rani, H. Sabharwal, P. Kumar, A. K. Chauhan, K. S. Khoo and N. Kataria, Comparative behavior of carbon-based materials for the removal of emerging bisphenol A from water: adsorption modelling and mechanism, *Groundw. Sustain. Dev.*, 2024, **25**, 101121.
- 44 T. Ahmad, M. S. Manzar, S. Khan, M. A. Al-Sharafi, J. Georgin, D. S. P. Franco, *et al.*, Enhanced adsorption of bisphenol-A from water through the application of isocyanurate based hyper crosslinked resin, *J. Mol. Liq.*, 2024, **395**, 123861.
- 45 S. A. H. Shah, M. R. Ramachandran, S. A. Mansur, N. M. Saleh and S. Asman, Selective recognition of bisphenol a using molecularly imprinted polymer based on choline chloride-methacrylic acid deep eutectic solvent monomer: Synthesis characterization and adsorption study, *Polymer*, 2023, **283**, 126279.
- 46 S. Chouhan, S. K. Yadav, J. Prakash, Swati and S. P. Singh, Effect of Bisphenol A on human health and its degradation by microorganisms: A review, *Ann. Microbiol.*, 2014, **64**(1), 13–21.
- 47 C. Kubwabo, I. Kosarac, B. Stewart, B. R. Gauthier, K. Lalonde and P. J. Lalonde, Migration of bisphenol A from plastic baby bottles, baby bottle liners and reusable polycarbonate drinking bottles, *Food Addit. Contam.: Part A*, 2009, **26**(6), 928–937.
- 48 C. Campanale, C. Massarelli, I. Savino, V. Locaputo and V. F. Uricchio, A detailed review study on potential effects of microplastics and additives of concern on human health, *Food Addit. Contam.: Part A*, 2020, **17**(4), 1212.



- 49 R. M. Cywar, N. A. Rorrer, C. B. Hoyt, G. T. Beckham and E. Y. X. Chen, Bio-based polymers with performance-advantaged properties, *Nat. Rev. Mater.*, 2022, **7**(2), 83–103.
- 50 *Marine Anthropogenic Litter*, ed. M. Bergmann, L. Gutow, M. Klages, M. Anthropogenic Litter. M. Bergmann, L. Gutow and M. Klages, Springer International Publishing, Cham, 2015, pp. 1–447.
- 51 J. Muncke, Exposure to endocrine disrupting compounds via the food chain: Is packaging a relevant source?, *Sci. Total Environ.*, 2009, **407**(16), 4549–4559.
- 52 V. Mustieles, S. C. D'Cruz, S. Couderq, A. Rodríguez-Carrillo, J. B. Fini, T. Hofer, *et al.*, Bisphenol A and its analogues: A comprehensive review to identify and prioritize effect biomarkers for human biomonitoring, *Environ. Int.*, 2020, **144**, 105811.
- 53 M. A. Coniglio, C. Fioriglio and P. Laganà, Non stances in PET-Bottled Mineral Water, *Springerbriefs in Molecular Science*, 2020, pp. 1–66.
- 54 P. Singh and V. P. Sharma, Integrated Plastic Waste Management: Environmental and Improved Health Approaches, *Procedia Environ. Sci.*, 2016, **35**, 692–700.
- 55 J. N. Hahladakis, C. A. Velis, R. Weber, E. Iacovidou and P. Purnell, An overview of chemical additives present in plastics: Migration, release, fate and environmental impact during their use, disposal and recycling, *J. Hazard. Mater.*, 2018, **344**, 179–199.
- 56 D. K. Schneiderman and M. A. Hillmyer, 50th Anniversary Perspective: There Is a Great Future in Sustainable Polymers, *Macromolecules*, 2017, **50**(10), 3733–3749.
- 57 J. O. Tijani, O. O. Fatoba, O. O. Babajide and L. F. Petrik, Pharmaceuticals, endocrine disruptors, personal care products, nanomaterials and perfluorinated pollutants: a review, *Environ. Chem. Lett.*, 2016, **14**(1), 27–49.
- 58 J. E. Cooper, E. L. Kendig and S. M. Belcher, Assessment of bisphenol A released from reusable plastic, aluminium and stainless steel water bottles, *Chemosphere*, 2011, **85**(6), 943–947.
- 59 C. Kubwabo, I. Kosarac, B. Stewart, B. R. Gauthier, K. Lalonde and P. J. Lalonde, Migration of bisphenol A from plastic baby bottles, baby bottle liners and reusable polycarbonate drinking bottles, *Food Addit. Contam.: Part A*, 2009, **26**(6), 928–937.
- 60 L. M. Iribarne-Durán, F. M. Peinado, C. Freire, I. Castellero-Rosales, F. Artacho-Cordón and N. Olea, Concentrations of bisphenols, parabens, and benzophenones in human breast milk: A systematic review and meta-analysis, *Sci. Total Environ.*, 2022, **806**, 150437.
- 61 H. H. Le, E. M. Carlson, J. P. Chua and S. M. Belcher, Bisphenol A is released from polycarbonate drinking bottles and mimics the neurotoxic actions of estrogen in developing cerebellar neurons, *Toxicol. Lett.*, 2008, **176**(2), 149–156.
- 62 A. Goodson, W. Summerfield and I. Cooper, Survey of bisphenol A and bisphenol F in canned foods, *Food Addit. Contam.*, 2002, **19**(8), 796–802.
- 63 H. C. Erythropel, M. Maric, J. A. Nicell, R. L. Leask and V. Yargeau, Leaching of the plasticizer di(2-ethylhexyl) phthalate (DEHP) from plastic containers and the question of human exposure, *Appl. Microbiol. Biotechnol.*, 2014, **98**(24), 9967–9981.
- 64 J. O. Tijani, O. O. Fatoba, O. O. Babajide and L. F. Petrik, Pharmaceuticals, endocrine disruptors, personal care products, nanomaterials and perfluorinated pollutants: a review, *Environ. Chem. Lett.*, 2016, **14**(1), 27–49.
- 65 N. Bumbudsanpharoke and S. Ko, Nano-Food Packaging: An Overview of Market, Migration Research, and Safety Regulations, *J. Food Sci.*, 2015, **80**(5), R910–R923.
- 66 G. K. Deshwal and N. R. Panjagari, Review on metal packaging: materials, forms, food applications, safety and recyclability, *J. Food Sci. Technol.*, 2020, **57**(7), 2377–2392.
- 67 K. C. Ugoeze, E. O. Amogu, K. E. Oluigbo and N. Nwachukwu, Environmental and public health impacts of plastic wastes due to healthcare and food products packages: A Review, *J. Environ. Sci. Publ. Health*, 2021, **05**(01), 1–31.
- 68 E. J. Hoekstra and C. Simoneau, Release of Bisphenol A from Polycarbonate-A Review, *Crit. Rev. Food Sci. Nutr.*, 2013, **53**(4), 386–402.
- 69 X. Liu, H. Shi, B. Xie, D. D. Dionysiou and Y. Zhao, Microplastics as Both a Sink and a Source of Bisphenol A in the Marine Environment, *Environ. Sci. Technol.*, 2019, **53**(17), 10188–10196.
- 70 K. Y. Foo and B. H. Hameed, An overview of landfill leachate treatment via activated carbon adsorption process, *J. Hazard. Mater.*, 2009, **171**(1–3), 54–60.
- 71 P. Rajasulochana and V. Preethy, Comparison on efficiency of various techniques in treatment of waste and sewage water – A comprehensive review, *Resour.-Effic. Technol.*, 2016, **2**(4), 175–184.
- 72 N. Morin, H. P. H. Arp and S. E. Hale, Bisphenol A in Solid Waste Materials, Leachate Water, and Air Particles from Norwegian Waste-Handling Facilities: Presence and Partitioning Behavior, *Environ. Sci. Technol.*, 2015, **49**(13), 7675–7683.
- 73 S. Almeida, A. Raposo, M. Almeida-González and C. Carrascosa, Bisphenol A: Food Exposure and Impact on Human Health, *Compr. Rev. Food Sci. Food Saf.*, 2018, **17**(6), 1503–1517.
- 74 P. Cao, H.-n. Zhong, K. Qiu, D. Li, G. Wu, H.-x. Sui, *et al.*, Exposure to bisphenol A and its substitutes, bisphenol F and bisphenol S from canned foods and beverages on Chinese market, *Food Control*, 2021, **120**, 107502.
- 75 D. Y. Huang, H. Q. Zhao, C. P. Liu and C. X. Sun, Characteristics, sources, and transport of tetrabromobisphenol A and bisphenol A in soils from a typical e-waste recycling area in South China, *Environ. Sci. Pollut. Res.*, 2014, **21**(9), 5818–5826.
- 76 M. Deviot, I. Lachaise, C. Högg, J. Durner, F. X. Reichl, J. P. Attal, *et al.*, Bisphenol A release from an orthodontic resin composite: A GC/MS and LC/MS study, *Dent. Mater.*, 2018, **34**(2), 341–354.
- 77 D. H. Tsai, F. W. Delrio, A. M. Keene, K. M. Tyner, R. I. MacCusprie, T. J. Cho, *et al.*, Adsorption and conformation of serum albumin protein on gold



- nanoparticles investigated using dimensional measurements and *in situ* spectroscopic methods, *Langmuir*, 2011, **27**(6), 2464–2477.
- 78 C. Liao and K. Kannan, Widespread occurrence of bisphenol A in paper and paper products: Implications for human exposure, *Environmental Science and Technology*, *Environ. Sci. Technol.*, 2011, **45**(21), 9372–9379.
  - 79 T. Geens, L. Goeyens, K. Kannan, H. Neels and A. Covaci, Levels of bisphenol-A in thermal paper receipts from Belgium and estimation of human exposure, *Sci. Total Environ.*, 2012, **435–436**, 30–33.
  - 80 B. E. Tokula, A. O. Dada, A. A. Inyinbor, K. S. Obayomi, O. S. Bello and U. Pal, Agro-waste based adsorbents as sustainable materials for effective adsorption of Bisphenol A from the environment: A review, *J. Cleaner Prod.*, 2023, **388**, 135819.
  - 81 İ. İyigüdoğdu, A. Üstündağ and Y. Duydu, Toxicological evaluation of bisphenol a and its analogues, *Turk. J. Pharm. Sci.*, 2020, **17**(4), 457–462.
  - 82 C. Pivonello, G. Muscogiuri, A. Nardone, F. Garifalos, D. P. Provisiero, N. Verde, *et al.*, Bisphenol A: An emerging threat to female fertility, *Reprod. Biol. Endocrinol.*, 2020, **18**(1), 4–11.
  - 83 C. Sonnenschein and A. M. Soto, An updated review of environmental estrogen and androgen mimics and antagonists, *J. Steroid Biochem. Mol. Biol.*, 1998, **65**(1–6), 143–150.
  - 84 F. S. Vom Saal and L. N. Vandenberg, Update on the Health Effects of Bisphenol A: Overwhelming Evidence of Harm, *Endocrinology*, 2021, **162**(3), 1–25.
  - 85 F. S. Vom Saal and L. N. Vandenberg, Update on the Health Effects of Bisphenol A: Overwhelming Evidence of Harm, *Endocrinology*, 2021, **162**(3), 1–25.
  - 86 J. Liang, L. Pang, C. Yang, J. Long, Q. Liao, P. Tang, *et al.*, Effects of prenatal single and mixed bisphenol exposure on bone mineral density in preschool children: A population-based prospective cohort study, *Ecotoxicol. Environ. Saf.*, 2023, **267**, 115665.
  - 87 Y. Ma, H. Liu, J. Wu, L. Yuan, Y. Wang, X. Du, *et al.*, The adverse health effects of bisphenol A and related toxicity mechanisms, *Environ. Res.*, 2019, 108575.
  - 88 M. Zhu, R. Zeng, D. Wu, Y. Li, T. Chen and A. Wang, Research progress of the effects of bisphenol analogues on the intestine and its underlying mechanisms: A review, *Environ. Res.*, 2024, **243**, 117891.
  - 89 J. Xing, S. Zhang, M. Zhang and J. Hou, A critical review of presence, removal and potential impacts of endocrine disruptors bisphenol A, *Comp. Biochem. Physiol., Part C: Toxicol. Pharmacol.*, 2022, **254**, 109275.
  - 90 Y. Xu, J. Nie, C. Lu, C. Hu, Y. Chen, Y. Ma, *et al.*, Effects and mechanisms of bisphenols exposure on neurodegenerative diseases risk: A systemic review, *Sci. Total Environ.*, 2024, **919**, 170670.
  - 91 A. Tarafdar, R. Sirohi, P. A. Balakumaran, R. Reshmy, A. Madhavan, R. Sindhu, *et al.*, The hazardous threat of Bisphenol A: Toxicity, detection and remediation, *J. Hazard. Mater.*, 2022, 127097.
  - 92 A. Kowalczyk, M. Wrzecińska, E. Czerniawska-Piątkowska, J. P. Araújo and P. Cwynar, Molecular consequences of the exposure to toxic substances for the endocrine system of females, *Biomed. Pharmacother.*, 2022, **155**, 113730.
  - 93 A. Kortenkamp, O. Martin, S. Ermler, A. Baig and M. Scholze, Bisphenol A and declining semen quality: A systematic review to support the derivation of a reference dose for mixture risk assessments, *Int. J. Hyg. Environ. Health*, 2022, **241**, 113942.
  - 94 Y. Liu, Y. Yao, H. Li, F. Qiao, J. Wu, Z. Y. Du, *et al.*, Influence of endogenous and exogenous estrogenic endocrine on intestinal microbiota in zebrafish, *PLoS One*, 2016, **11**(10), 1–12.
  - 95 I. Ďurovcová, S. Kyzek, J. Fabová, J. Makuková, E. Gálová and A. Ševčovičová, Genotoxic potential of bisphenol A: A review, *Environ. Pollut.*, 2022, **306**, 119346.
  - 96 O. Banerjee, S. Singh, I. Saha, S. Pal, M. Banerjee, S. Kundu, *et al.*, Molecular dissection of cellular response of pancreatic islet cells to Bisphenol-A (BPA): A comprehensive review, *Biochem. Pharmacol.*, 2022, **201**, 115068.
  - 97 M. Y. Xie, H. Ni, D. S. Zhao, L. Y. Wen, K. S. Li, H. H. Yang, *et al.*, Exposure to bisphenol A and the development of asthma: A systematic review of cohort studies, *Reprod. Toxicol.*, 2016, **65**, 224–229.
  - 98 I. Bahelka, R. Stupka, J. Čitek and M. Šprysl, The impact of bisphenols on reproductive system and on offspring in pigs – A review 2011–2020, *Chemosphere*, 2021, **263**, 128203.
  - 99 L. Y. Meng, W. F. Tao, J. Li, M. Zhu, D. B. Zhong, J. Zhou, *et al.*, Effects of Bisphenol A and Its Substitute, Bisphenol F, on the Gut Microbiota in Mice, *Biomed. Environ. Sci.*, 2024, **37**(1), 19–30.
  - 100 A. Seyoum and A. Pradhan, Effect of phthalates on development, reproduction, fat metabolism and lifespan in *Daphnia magna*, *Sci. Total Environ.*, 2019, **654**, 969–977.
  - 101 A. Seyoum and A. Pradhan, Effect of phthalates on development, reproduction, fat metabolism and lifespan in *Daphnia magna*, *Sci. Total Environ.*, 2019, **654**, 969–977.
  - 102 R. Planelló, O. Herrero, J. L. Martínez-Guitarte and G. Morcillo, Comparative effects of butyl benzyl phthalate (BBP) and di(2-ethylhexyl) phthalate (DEHP) on the aquatic larvae of *Chironomus riparius* based on gene expression assays related to the endocrine system, the stress response and ribosomes, *Aquat. Toxicol.*, 2011, **105**(1–2), 62–70.
  - 103 R. Planelló, O. Herrero, J. L. Martínez-Guitarte and G. Morcillo, Comparative effects of butyl benzyl phthalate (BBP) and di(2-ethylhexyl) phthalate (DEHP) on the aquatic larvae of *Chironomus riparius* based on gene expression assays related to the endocrine system, the stress response and ribosomes, *Aquat. Toxicol.*, 2011, **105**(1–2), 62–70.
  - 104 B. Yurdakok-Dikmen, Y. Turgut, A. Gunal, R. Uyar, O. Kuzukiran, A. Filazi, *et al.*, In vitro effects of selected endocrine disruptors (DEHP, PCB118, BPA) on narrow-clawed crayfish (*Astacus leptodactylus*) primary cells, *In Vitro Cell. Dev. Biol.: Anim.*, 2020, **56**(9), 783–791.



- 105 I. D. L. Gomes, I. Gazo, D. Nabi, L. Besnardeau, C. Hebras, A. McDougall, *et al.*, Bisphenols disrupt differentiation of the pigmented cells during larval brain formation in the ascidian, *Aquat. Toxicol.*, 2019, **216**, 105314.
- 106 S. N. Volz, R. Poulsen, M. Hansen and H. Holbech, Bisphenol A alters retinal morphology, visually guided behavior, and thyroid hormone levels in zebrafish larvae, *Chemosphere*, 2024, **348**, 140776.
- 107 M. A. Martín-Lara, M. Calero, A. Ronda, I. Iáñez-Rodríguez and C. Escudero, Adsorptive behavior of an activated carbon for bisphenol A removal in single and binary (bisphenol A-heavy metal) solutions, *Water*, 2020, **12**(8), 2150.
- 108 M. Seoane, Á. Cid, C. Herrero and M. Esperanza, Comparative acute toxicity of benzophenone derivatives and bisphenol analogues in the Asian clam *Corbicula fluminea*, *Ecotoxicology*, 2021, **30**(1), 142–153.
- 109 M. Faheem and R. K. Bhandari, Detrimental Effects of Bisphenol Compounds on Physiology and Reproduction in Fish: A Literature Review, *Environ. Toxicol. Pharmacol.*, 2021, **81**, 103497.
- 110 C. Xiao, L. Wang, Q. Zhou and X. Huang, Hazards of bisphenol A (BPA) exposure: A systematic review of plant toxicology studies, *J. Hazard. Mater.*, 2020, **384**, 121488.
- 111 W. Czarnocka and S. Karpiński, Friend or foe? Reactive oxygen species production, scavenging and signaling in plant response to environmental stresses, *Free Radic. Biol. Med.*, 2018, **122**, 4–20.
- 112 İ. İyigündoğdu, A. Üstündağ and Y. Duydu, Toxicological evaluation of bisphenol a and its analogues, *Turk. J. Pharm. Sci.*, 2020, **17**(4), 457–462.
- 113 T. T. Schug and L. S. Birnbaum, Human Health Effects of Bisphenol A, *Molecular and Integrative Toxicology*, 2014, pp. 1–29.
- 114 L. S. Birnbaum, J. R. Bucher, G. W. Collman, D. C. Zeldin, A. F. Johnson, T. T. Schug, *et al.*, Consortium-based science: The NIEHS's multipronged, collaborative approach to assessing the health effects of Bisphenol A, *Environ. Health Perspect.*, 2012, **120**(12), 1640–1644.
- 115 J. W. Brock, Y. Yoshimura, J. R. Barr, V. L. Maggio, S. R. Graiser, H. Nakazawa and L. L. Needham, Measurement of bisphenol A levels in human urine, *J. Exposure Sci. Environ. Epidemiol.*, 2001, **11**(4), 323–328.
- 116 C. Zeng, W. Sweet and Q. Cheng, Ecological Citizenship and Green Burial in China, *J. Agric. Environ. Ethics*, 2016, **29**(6), 985–1001.
- 117 B. Petrie, L. Lopardo, K. Proctor, J. Youdan, R. Barden and B. Kasprzyk-Hordern, Assessment of bisphenol-A in the urban water cycle, *Sci. Total Environ.*, 2019, 900–907.
- 118 K. Ozhan and E. Kocaman, Temporal and Spatial Distributions of Bisphenol A in Marine and Freshwaters in Turkey, *Arch. Environ. Contam. Toxicol.*, 2019, **76**(2), 246–254.
- 119 L. G. A. Barboza, S. C. Cunha, C. Monteiro, J. O. Fernandes and L. Guilhermino, Bisphenol A and its analogs in muscle and liver of fish from the North East Atlantic Ocean in relation to microplastic contamination. Exposure and risk to human consumers, *J. Hazard. Mater.*, 2020, **393**, 122419.
- 120 X. Liu, H. Shi, B. Xie, D. D. Dionysiou and Y. Zhao, Microplastics as Both a Sink and a Source of Bisphenol A in the Marine Environment, *Environ. Sci. Technol.*, 2019, **53**(17), 10188–10196.
- 121 A. B. Javurek, W. G. Spollen, S. A. Johnson, N. J. Bivens, K. H. Bromert, S. A. Givan, *et al.*, Effects of exposure to bisphenol A and ethinyl estradiol on the gut microbiota of parents and their offspring in a rodent model, *Gut Microb.*, 2016, **7**(6), 471–485.
- 122 L. Chen, Y. Guo, C. Hu, P. K. S. Lam, J. C. W. Lam and B. Zhou, Dysbiosis of gut microbiota by chronic coexposure to titanium dioxide nanoparticles and bisphenol A: Implications for host health in zebrafish, *Environ. Pollut.*, 2018, **234**, 307–317.
- 123 T. R. Catron, S. P. Keely, N. E. Brinkman, T. J. Zurlinden, C. E. Wood, J. R. Wright, *et al.*, Host Developmental Toxicity of BPA and BPA Alternatives Is Inversely Related to Microbiota Disruption in Zebrafish, *Toxicol. Sci.*, 2019, **167**(2), 468–483.
- 124 T. R. Catron, S. P. Keely, N. E. Brinkman, T. J. Zurlinden, C. E. Wood, J. R. Wright, *et al.*, Host Developmental Toxicity of BPA and BPA Alternatives Is Inversely Related to Microbiota Disruption in Zebrafish, *Toxicol. Sci.*, 2019, **167**(2), 468–483.
- 125 Y. Liu, Y. Yao, H. Li, F. Qiao, J. Wu, Z. Y. Du, *et al.*, Influence of endogenous and exogenous estrogenic endocrine on intestinal microbiota in zebrafish, *PLoS One*, 2016, **11**(10), 1–12.
- 126 K. P. Lai, Y. T. Chung, R. Li, H. T. Wan and C. K. C. Wong, Bisphenol A alters gut microbiome: Comparative metagenomics analysis, *Environ. Pollut.*, 2016, **218**, 923–930.
- 127 A. B. Javurek, W. G. Spollen, S. A. Johnson, N. J. Bivens, K. H. Bromert, S. A. Givan, *et al.*, Effects of exposure to bisphenol A and ethinyl estradiol on the gut microbiota of parents and their offspring in a rodent model, *Gut Microb.*, 2016, **7**(6), 471–485.
- 128 Y. M. Lee, M. J. Seong, J. W. Lee, Y. K. Lee, T. M. Kim, S. Y. Nam, *et al.*, Estrogen receptor independent neurotoxic mechanism of bisphenol A, an environmental estrogen, *J. Vet. Sci.*, 2007, **8**(1), 27–38.
- 129 T. Wang, J. Han, X. Duan, B. Xiong, X. S. Cui, N. H. Kim, *et al.*, The toxic effects and possible mechanisms of bisphenol A on oocyte maturation of porcine in vitro, *Oncotarget*, 2016, **7**(22), 32554–32565.
- 130 J. Guo, M. H. Zhao, K. T. Shin, Y. J. Niu, Y. D. Ahn, N. H. Kim, *et al.*, The possible molecular mechanisms of bisphenol A action on porcine early embryonic development, *Sci. Rep.*, 2017, **7**(1), 1–9.
- 131 I. Cimmino, F. Fiory, G. Perruolo, C. Miele, F. Beguinot, P. Formisano, *et al.*, Potential mechanisms of bisphenol A (BPA), *Int. J. Mol. Sci.*, 2020, **21**(16), 1–22.
- 132 K. Baralić, K. Živančević, D. Jorgovanović, D. Javorac, J. Radovanović, T. Gojković, *et al.*, Probiotic reduced the impact of phthalates and bisphenol A mixture on type 2



- diabetes mellitus development: Merging bioinformatics with *in vivo* analysis, *Food Chem. Toxicol.*, 2021, **154**, 112325.
- 133 S. Rahmani, N. Pour Khalili, F. Khan, S. Hassani, E. Ghafour-Boroujerdi and M. Abdollahi, Bisphenol A: What lies beneath its induced diabetes and the epigenetic modulation?, *Life Sci.*, 2018, **214**, 136–144.
  - 134 S. Siddique, G. Zhang and C. Kubwabo, Exposure to bisphenol a and risk of developing type 2 diabetes: A mini review, *Emerging Contam.*, 2020, **6**, 274–282.
  - 135 C. Xiong, L. Xu, X. Dong, Z. Cao, Y. Wang, K. Chen, *et al.*, Trimester-specific associations of maternal exposure to bisphenols with neonatal thyroid stimulating hormone levels: A birth cohort study, *Sci. Total Environ.*, 2023, **880**, 163354.
  - 136 R. Meli, A. Monnolo, C. Annunziata, C. Pirozzi and M. C. Ferrante, Oxidative stress and BPA toxicity: An antioxidant approach for male and female reproductive dysfunction, *Antioxidants*, 2020, **9**(5), 405.
  - 137 S. Vigneshwaran, J. Preethi and S. Meenakshi, Interface engineering of ultrathin multi-functional 2D draped chitosan for efficient charge separation on degradation of paraquat – A mechanistic study, *J. Environ. Chem. Eng.*, 2020, **8**(5), 104446.
  - 138 V. Sharma, D. Jain, A. R. Rai, P. Kumari, V. Nagar, A. Kaur, *et al.*, Toxicological assessment and concentration analysis of Bisphenol A in food grade plastics: A systematic review, *Mater. Today: Proc.*, 2023, **95**, 18–25.
  - 139 J. Q. Chen, T. R. Brown and J. Russo, Regulation of energy metabolism pathways by estrogens and estrogenic chemicals and potential implications in obesity associated with increased exposure to endocrine disruptors, *Biochim. Biophys. Acta, Mol. Cell Res.*, 2009, **1793**(7), 1128–1143.
  - 140 N. D. Mandel, B. Gamboa-Loira, M. E. Cebrián, Á. Mérida-Ortega and L. López-Carrillo, Challenges to regulate products containing bisphenol A: Implications for policy, *Salud Publica Mex.*, 2019, **61**(5), 692–697.
  - 141 W. T. Tsai, Human health risk on environmental exposure to bisphenol-A: A review, *J. Environ. Sci. Health C Environ. Carcinog. Ecotoxicol. Rev.*, 2006, **24**(2), 225–255.
  - 142 D. Amiridou and D. Voutsas, Alkylphenols and phthalates in bottled waters, *J. Hazard. Mater.*, 2011, **185**(1), 281–286.
  - 143 T. Geens, T. Z. Apelbaum, L. Goeyens, H. Neels and A. Covaci, Intake of bisphenol A from canned beverages and foods on the Belgian market, *Food Addit. Contam.: Part A*, 2010, **27**(11), 1627–1637.
  - 144 S. Maggioni, P. Balaguer, C. Chiozzotto and E. Benfenati, Screening of endocrine-disrupting phenols, herbicides, steroid estrogens, and estrogenicity in drinking water from the waterworks of 35 Italian cities and from PET-bottled mineral water, *Environ. Sci. Pollut. Res.*, 2013, **20**(3), 1649–1660.
  - 145 E. Fasano, F. Esposito, G. Scognamiglio, F. Di Francesco, P. Montuori, C. R. Amodio, *et al.*, Bisphenol A contamination in soft drinks as a risk for children's health in Italy, *Food Addit. Contam.: Part A*, 2015, **32**(7), 1207–1214.
  - 146 M. N. Tzatzarakis, V. Karzi, E. Vakonaki, M. Goumenou, M. Kavvalakis, P. Stivaktakis, *et al.*, Bisphenol A in soft drinks and canned foods and data evaluation, *Food Addit. Contam., Part B*, 2017, **10**(2), 85–90.
  - 147 N. Casajuana and S. Lacorte, New methodology for the determination of phthalate esters, bisphenol A, bisphenol A diglycidyl ether, and nonylphenol in commercial whole milk samples, *J. Agric. Food Chem.*, 2004, **52**(12), 3702–3707.
  - 148 L. Grumetto, O. Gennari, D. Montesano, R. Ferracane, A. Ritieni, S. Albrizio, *et al.*, Determination of five bisphenols in commercial milk samples by liquid chromatography coupled to fluorescence detection, *J. Food Prot.*, 2013, **76**(9), 1590–1596.
  - 149 A. K. Sakhi, I. T. L. Lillegaard, S. Voorspoels, M. H. Carlsen, E. B. Løken, A. L. Brantsæter, *et al.*, Concentrations of phthalates and bisphenol A in Norwegian foods and beverages and estimated dietary exposure in adults, *Environ. Int.*, 2014, **73**, 259–269.
  - 150 T. Cirillo, G. Latini, M. A. Castaldi, L. Dipaola, E. Fasano, F. Esposito, *et al.*, Exposure to di-2-ethylhexyl phthalate, di-n-butyl phthalate and bisphenol A through infant formulas, *J. Agric. Food Chem.*, 2015, **63**(12), 3303–3310.
  - 151 Z. Brenn-Struckhofova and M. Cichna-Markl, Determination of bisphenol A in wine by sol-gel immunoaffinity chromatography, HPLC and fluorescence detection, *Food Addit. Contam.*, 2006, **23**(11), 1227–1235.
  - 152 J. M. Braun, A. E. Kalkbrenner, A. M. Calafat, J. T. Bernert, X. Ye, M. J. Silva, *et al.*, Variability and predictors of urinary bisphenol a concentrations during pregnancy, *Environ. Health Perspect.*, 2011, **119**(1), 131–137.
  - 153 L. Grumetto, D. Montesano, S. Seccia, S. Albrizio and F. Barbato, Determination of bisphenol A and bisphenol B residues in canned peeled tomatoes by reversed-phase liquid chromatography, *J. Agric. Food Chem.*, 2008, **56**(22), 10633–10637.
  - 154 P. Viñas, N. Campillo, N. Martínez-Castillo and M. Hernández-Córdoba, Comparison of two derivatization-based methods for solid-phase microextraction-gas chromatography-mass spectrometric determination of bisphenol A, bisphenol S and biphenol migrated from food cans, *Anal. Bioanal. Chem.*, 2010, **397**(1), 115–125.
  - 155 A. Alabi, N. Caballero-Casero and S. Rubio, Quick and simple sample treatment for multiresidue analysis of bisphenols, bisphenol diglycidyl ethers and their derivatives in canned food prior to liquid chromatography and fluorescence detection, *J. Chromatogr. A*, 2014, **1336**, 23–33.
  - 156 S. Errico, M. Bianco, L. Mita, M. Migliaccio, S. Rossi, C. Nicolucci, *et al.*, Migration of bisphenol A into canned tomatoes produced in Italy: Dependence on temperature and storage conditions, *Food Chem.*, 2014, **160**, 157–164.
  - 157 I. Gyllenhammar, A. Glynn, P. O. Darnerud, S. Lignell, R. van Delft and M. Aune, 4-Nonylphenol and bisphenol



- A in Swedish food and exposure in Swedish nursing women, *Environ. Int.*, 2012, **43**(1), 21–28.
- 158 S. Gorecki, N. Bemrah, A. C. Roudot, E. Marchioni, B. Le Bizec, F. Faivre, *et al.*, Human health risks related to the consumption of foodstuffs of animal origin contaminated by bisphenol A, *Food Chem. Toxicol.*, 2017, **110**, 333–339.
  - 159 N. Bemrah, J. Jean, G. Rivière, M. Sanaa, S. Leconte, M. Bachelot, *et al.*, Assessment of dietary exposure to bisphenol A in the French population with a special focus on risk characterisation for pregnant French women, *Food Chem. Toxicol.*, 2014, **72**, 90–97.
  - 160 V. Lo Turco, G. Di Bella, A. G. Potorti, A. Tropea, E. K. Casale, M. R. Fede, *et al.*, Determination of plasticisers and BPA in Sicilian and Calabrian nectar honeys by selected ion monitoring GC/MS, *Food Addit. Contam.: Part A*, 2016, **33**(11), 1693–1699.
  - 161 O. Zoller, B. J. Brüsweiler, R. Magnin, H. Reinhard, P. Rhy, H. Rupp, *et al.*, Natural occurrence of bisphenol F in mustard, *Food Addit. Contam.: Part A*, 2015, **33**(1), 137–146.
  - 162 C. Bach, X. Dauchy, M. C. Chagnon and S. Etienne, Chemical compounds and toxicological assessments of drinking water stored in polyethylene terephthalate (PET) bottles: A source of controversy reviewed, *Water Res.*, 2012, **46**(3), 571–583.
  - 163 E. Ferrer, E. Santoni, S. Vittori, G. Font, J. Mañes and G. Sagratini, Simultaneous determination of bisphenol A, octylphenol, and nonylphenol by pressurised liquid extraction and liquid chromatography-tandem mass spectrometry in powdered milk and infant formulas, *Food Chem.*, 2011, **126**(1), 360–367.
  - 164 S. C. Cunha, R. N. Alves, J. O. Fernandes, S. Casal and A. Marques, First approach to assess the bioaccessibility of bisphenol A in canned seafood, *Food Chem.*, 2017, **232**, 501–507.
  - 165 M. Sasaki, J. I. Maki, K. I. Oshiman, Y. Matsumura and T. Tsuchido, Biodegradation of bisphenol A by cells and cell lysate from *Sphingomonas* sp. strain AO1, *Biodegradation*, 2005, **16**(5), 449–459.
  - 166 M. Kamaraj, R. Sivaraj and R. Venkatesh, Biodegradation of Bisphenol A by the tolerant bacterial species isolated from coastal regions of Chennai, Tamil Nadu, India, *Int. Biodeterior. Biodegrad.*, 2014, **93**, 216–222.
  - 167 M. Kamaraj, R. Sivaraj and R. Venkatesh, Biodegradation of Bisphenol A by the tolerant bacterial species isolated from coastal regions of Chennai, Tamil Nadu, India, *Int. Biodeterior. Biodegrad.*, 2014, **93**, 216–222.
  - 168 O. Gulnaz and S. Dincer, Biodegradation of Bisphenol a by *Chlorella vulgaris* and *Aeromonas Hydrophilia*, *J. Appl. Biol. Sci.*, 2009, **3**(2), 79–84.
  - 169 O. Gulnaz and S. Dincer, Biodegradation of Bisphenol a by *Chlorella vulgaris* and *Aeromonas Hydrophilia*, *J. Appl. Biol. Sci.*, 2009, **3**(2), 79–84.
  - 170 M. Sasaki, J. I. Maki, K. I. Oshiman, Y. Matsumura and T. Tsuchido, Biodegradation of bisphenol A by cells and cell lysate from *Sphingomonas* sp. strain AO1, *Biodegradation*, 2005, **16**(5), 449–459.
  - 171 R. Das, G. Li, B. Mai and T. An, Spore cells from BPA degrading bacteria *Bacillus* sp. GZB displaying high laccase activity and stability for BPA degradation, *Sci. Total Environ.*, 2018, **640–641**, 798–806.
  - 172 H. Yamanaka, K. Moriyoshi, T. Ohmoto, T. Ohe and K. Sakai, Degradation of bisphenol a by *Bacillus pumilus* isolated from kimchi, a traditionally fermented food, *Appl. Biochem. Biotechnol.*, 2007, **136**(1), 39–51.
  - 173 Y. Jia, A. Eltoukhy, J. Wang, X. Li, T. S. Hlaing, M. M. Aung, *et al.*, Biodegradation of bisphenol a by *Sphingobium* sp. YC-JY1 and the essential role of cytochrome P450 Monooxygenase, *Int. J. Mol. Sci.*, 2020, **21**(10), 3588.
  - 174 Y. Yang, X. G. Chen, L. Wang, Z. H. Chu, Z. Liu, Y. C. Dong, *et al.*, Sliding wear behavior of in-situ FeAl<sub>2</sub>O<sub>4</sub> matrix nanocomposite coating fabricated by plasma spraying, *Tribol. Int.*, 2015, **81**, 97–104.
  - 175 M. Takeo, J. Akizuki, A. Kawasaki and S. Negoro, Degradation potential of the nonylphenol monooxygenase of *Sphingomonas* sp. NP5 for bisphenols and their structural analogs, *Microorganisms*, 2020, **8**(2), 284.
  - 176 A. A. Telke, D. C. Kalyani, U. U. Jadhav, G. K. Parshetti and S. P. Govindwar, Purification and characterization of an extracellular laccase from a *Pseudomonas* sp. LBC1 and its application for the removal of bisphenol A, *J. Mol. Catal. B Enzym.*, 2009, **61**(3–4), 252–260.
  - 177 E. J. Eio, M. Kawai, K. Tsuchiya, S. Yamamoto and T. Toda, Biodegradation of bisphenol A by bacterial consortia, *Int. Biodeterior. Biodegrad.*, 2014, **96**, 166–173.
  - 178 N. A. Zhou, H. Kjeldal, H. L. Gough and J. L. Nielsen, Identification of Putative Genes Involved in Bisphenol A Degradation Using Differential Protein Abundance Analysis of *Sphingobium* sp. BiD32, *Environ. Sci. Technol.*, 2015, **49**(20), 12231–12241.
  - 179 W. Yue, C. F. Yin, L. Sun, J. Zhang, Y. Xu and N. Y. Zhou, Biodegradation of bisphenol-A polycarbonate plastic by *Pseudoxanthomonas* sp. strain NyZ600, *J. Hazard. Mater.*, 2021, **416**, 125775.
  - 180 S. Oh and D. Choi, Microbial Community Enhances Biodegradation of Bisphenol A Through Selection of *Sphingomonadaceae*, *Microb. Ecol.*, 2019, **77**(3), 631–639.
  - 181 A. López-Moreno, A. Torres-Sánchez, I. Acuña, A. Suárez and M. Aguilera, Representative *Bacillus* sp. AM1 from Gut Microbiota Harbor Versatile Molecular Pathways for Bisphenol A Biodegradation, *Int. J. Mol. Sci.*, 2021, **22**(9), 4952.
  - 182 B. Suyamud, P. Thiravetyan, G. M. Gadd, B. Panyapinyopol and D. Inthorn, Bisphenol A removal from a plastic industry wastewater by *Dracaena sanderiana* endophytic bacteria and *Bacillus cereus* NI, *Int. J. Phytoremediation*, 2020, **22**(2), 167–175.
  - 183 D. He, Y. Zeng and G. Zhou, The influence of microplastics on the toxic effects and biodegradation of bisphenol A in the microalgae *Chlorella pyrenoidosa*, *Aquat. Ecol.*, 2022, **56**(4), 1287–1296.
  - 184 N. A. Zhou, H. Kjeldal, H. L. Gough and J. L. Nielsen, Identification of Putative Genes Involved in Bisphenol A Degradation Using Differential Protein Abundance



- Analysis of *Sphingobium* sp. BiD32, *Environ. Sci. Technol.*, 2015, **49**(20), 12231–12241.
- 185 R. Hu, H. Zhao, X. Xu, Z. Wang, K. Yu, L. Shu, *et al.*, Bacteria-driven phthalic acid ester biodegradation: Current status and emerging opportunities, *Environ. Int.*, 2021, **154**, 106560.
  - 186 L. Ren, Y. Jia, R. Zhang, Z. Lin, Z. Zhen, H. Hu, *et al.*, Insight into metabolic versatility of an aromatic compounds-degrading arthrobactersp. YC-RL1, *Front. Microbiol.*, 2018, **9**, 1–15.
  - 187 A. Ismanto, T. Hadibarata, R. A. Kristanti, L. Maslukah, N. Safinatunnajah and W. Kusumastuti, Endocrine disrupting chemicals (EDCs) in environmental matrices: Occurrence, fate, health impact, physio-chemical and bioremediation technology, *Environ. Pollut.*, 2022, **302**, 119061.
  - 188 W. Dong, Q. Chen, Y. Hou, S. Li, K. Zhuang, F. Huang, *et al.*, Metabolic pathway involved in 2- methyl-6-ethylaniline degradation by *Sphingobium* sp. strain MEA3-1 and cloning of the novel flavin-dependent monooxygenase system meaBA, *Appl. Environ. Microbiol.*, 2015, **81**(24), 8254–8264.
  - 189 C. Filote, M. Roşca, R. M. Hlihor, P. Cozma, I. M. Simion, M. Apostol, *et al.*, Sustainable application of biosorption and bioaccumulation of persistent pollutants in wastewater treatment: Current practice, *Processes*, 2021, **9**(10), 1–38.
  - 190 V. D'Argenio, G. Casaburi, V. Precone and F. Salvatore, Comparative metagenomic analysis of human gut microbiome composition using two different bioinformatic pipelines, *BioMed Res. Int.*, 2014, **2014**, 325340.
  - 191 D. P. Mohapatra, S. K. Brar, R. D. Tyagi and R. Y. Surampalli, Physico-chemical pre-treatment and biotransformation of wastewater and wastewater Sludge - Fate of bisphenol A, *Chemosphere*, 2010, **78**(8), 923–941.
  - 192 W. Dong, Q. Chen, Y. Hou, S. Li, K. Zhuang, F. Huang, *et al.*, Metabolic pathway involved in 2- methyl-6-ethylaniline degradation by *Sphingobium* sp. strain MEA3-1 and cloning of the novel flavin-dependent monooxygenase system meaBA, *Appl. Environ. Microbiol.*, 2015, **81**(24), 8254–8264.
  - 193 V. D'Argenio, G. Casaburi, V. Precone and F. Salvatore, Comparative metagenomic analysis of human gut microbiome composition using two different bioinformatic pipelines, *BioMed Res. Int.*, 2014, **2014**, 325340.
  - 194 L. G. C. Villegas, N. Mashhadi, M. Chen, D. Mukherjee, K. E. Taylor and N. Biswas, A Short Review of Techniques for Phenol Removal from Wastewater, *Curr. Pollut. Rep.*, 2016, **2**(3), 157–167.
  - 195 U. Kües, Fungal enzymes for environmental management, *Curr. Opin. Biotechnol.*, 2015, **33**, 268–278.
  - 196 Y. Jia, A. Eltoukhy, J. Wang, X. Li, T. S. Hlaing, M. M. Aung, *et al.*, Biodegradation of bisphenol a by *sphingobium* sp. YC-JY1 and the essential role of cytochrome P450 Monooxygenase, *Int. J. Mol. Sci.*, 2020, **21**(10), 3588.
  - 197 J. Bhagat, L. Zang, N. Nishimura and Y. Shimada, Zebrafish: An emerging model to study microplastic and nanoplastic toxicity, *Sci. Total Environ.*, 2020, **728**, 138707.
  - 198 S. G. Nkhata, E. Ayua, E. H. Kamau and J. B. Shingiro, Fermentation and germination improve nutritional value of cereals and legumes through activation of endogenous enzymes, *Food Sci. Nutr.*, 2018, **6**(8), 2446–2458.
  - 199 R. V. Moharir and S. Kumar, Challenges associated with plastic waste disposal and allied microbial routes for its effective degradation: A comprehensive review, *J. Cleaner Prod.*, 2019, **208**, 65–76.
  - 200 S. Oh and D. Choi, Microbial Community Enhances Biodegradation of Bisphenol A Through Selection of Sphingomonadaceae, *Microb. Ecol.*, 2019, **77**(3), 631–639.
  - 201 R. Villemur, S. Cristina, J. Ouellette, P. Juteau, F. Lépine, E. Déziel, *et al.*, Biodegradation of endocrine disruptors in solid-liquid two-phase partitioning systems by enrichment cultures, *Appl. Environ. Microbiol.*, 2013, **79**(15), 4701–4711.
  - 202 K. V. S. Rajmohan, C. Ramya, M. Raja Viswanathan and S. Varjani, Plastic pollutants: effective waste management for pollution control and abatement, *Curr. Opin. Environ. Sci. Health.*, 2019, **12**, 72–84.
  - 203 P. A. Opoku, H. Jingyu, L. Yi, D. Ewusi-Mensah and N. Miwornunyuie, Scalability of the multi-anode plug flow microbial fuel cell as a sustainable prospect for large-scale design, *Renew. Energy*, 2023, **207**(1), 693–702.
  - 204 M. Zielińska, I. Wojnowska-Baryła and A. Cydzik-Kwiatkowska, *Bisphenol A Removal from Water and Wastewater*, Springer, 2019.
  - 205 Y. Zhao, C. W. Cho, L. Cui, W. Wei, J. Cai, G. Wu, *et al.*, Adsorptive removal of endocrine-disrupting compounds and a pharmaceutical using activated charcoal from aqueous solution: kinetics, equilibrium, and mechanism studies, *Environ. Sci. Pollut. Res.*, 2019, **26**, 33897–33905.
  - 206 M. S. Muhamad, M. R. Salim, W. J. Lau, Z. Yusop and T. Hadibarata, The removal of bisphenol A in water treatment plant using ultrafiltration membrane system, *Water, Air, Soil Pollut.*, 2016, **227**, 1–12.
  - 207 C. E. Almeida-Naranjo, V. H. Guerrero and C. A. Villamar-Ayala, Emerging contaminants and their removal from aqueous media using conventional/non-conventional adsorbents: a glance at the relationship between materials, processes, and technologies, *Water*, 2023, **15**(8), 1626.
  - 208 B. S. Rath and P. S. Kumar, Application of adsorption process for effective removal of emerging contaminants from water and wastewater, *Environ. Pollut.*, 2021, **280**, 116995.
  - 209 Z. Fang, Y. Hu, J. Cheng and Y. Chen, Continuous removal of trace bisphenol A from water by high efficacy TiO<sub>2</sub> nanotube pillared graphene-based macrostructures in a photocatalytically fluidized bed, *Chem. Eng. J.*, 2019, **372**, 581–589.
  - 210 S. Zinadini, A. A. Zinatizadeh, M. Rahimi, V. Vatanpour and K. Bahrami, Energy recovery and hygienic water production from wastewater using an innovative integrated microbial



- fuel cell-membrane separation process, *Energy*, 2017, **141**, 1350–1362.
- 211 M. S. Muhamad, M. R. Salim, W. J. Lau, Z. Yusop and T. Hadibarata, The removal of bisphenol A in water treatment plant using ultrafiltration membrane system, *Water, Air, Soil Pollut.*, 2016, **227**, 1–12.
  - 212 Y. S. Khoo, P. S. Goh, W. J. Lau, A. F. Ismail, M. S. Abdullah, N. H. M. Ghazali, *et al.*, Removal of emerging organic micropollutants via modified-reverse osmosis/nanofiltration membranes: A review, *Chemosphere*, 2022, **305**, 135151.
  - 213 N. A. Khan, S. Singh, E. A. López-Maldonado, N. Pavithra, P. F. Méndez-Herrera, J. R. López-López, *et al.*, Emerging membrane technology and hybrid treatment systems for the removal of micropollutants from wastewater, *Desalination*, 2023, **565**, 116873.
  - 214 M. S. Muhamad, M. R. Salim, W. J. Lau and Z. Yusop, A review on bisphenol A occurrences, health effects and treatment process via membrane technology for drinking water, *Environ. Sci. Pollut. Res.*, 2016, **23**, 11549–11567.
  - 215 C. Charcosset, Ultrafiltration, microfiltration, nanofiltration and reverse osmosis in integrated membrane processes, *Integr. Membr. Syst. Processes*, 2016, 1–22.
  - 216 Y. S. Khoo, P. S. Goh, W. J. Lau, A. F. Ismail, M. S. Abdullah, N. H. M. Ghazali, *et al.*, Removal of emerging organic micropollutants via modified-reverse osmosis/nanofiltration membranes: A review, *Chemosphere*, 2022, **305**, 135151.
  - 217 L. L. S. Silva, W. Abdelraheem, M. N. Nadagouda, A. R. Maria, D. D. Dionysiou, F. V. Fonseca, *et al.*, Novel microwave-driven synthesis of hydrophilic polyvinylidene fluoride/polyacrylic acid (PVDF/PAA) membranes and decoration with nano zero-valent-iron (nZVI) for water treatment applications, *J. Membr. Sci.*, 2020, 116544.
  - 218 H. Wu, X. Niu, J. Yang, C. Wang and M. Lu, Retentions of bisphenol A and norfloxacin by three different ultrafiltration membranes in regard to drinking water treatment, *Chem. Eng. J.*, 2016, 410–416.
  - 219 R. Valladares Linares, V. Yangali-Quintanilla, Z. Li and G. Amy, Rejection of micropollutants by clean and fouled forward osmosis membrane, *Water Res.*, 2011, 6737–6744.
  - 220 C. G. Moreira, M. H. Moreira, V. M. O. C. Silva, H. G. Santos, D. M. Bila and F. V. Fonseca, Treatment of Bisphenol A (BPA) in water using UV/H<sub>2</sub>O<sub>2</sub> and reverse osmosis (RO) membranes: assessment of estrogenic activity and membrane adsorption, *Water Sci. Technol.*, 2019, **80**(11), 2169–2178.
  - 221 C. G. Moreira, L. C. de Souza, T. C. Castor Neto, G. Gomes, D. M. Bila and F. V. Fonseca, Combined reverse osmosis and UV/H<sub>2</sub>O<sub>2</sub> treatment of aqueous solutions of bisphenol A and 17 $\alpha$ -ethinylestradiol: assessment of estrogenic activity, *Environ. Technol.*, 2023, **44**(20), 3108–3120.
  - 222 Y. Hu, Q. Zhu, X. Yan, C. Liao and G. Jiang, Occurrence, fate and risk assessment of BPA and its substituents in wastewater treatment plant: A review, *Environ. Res.*, 2019, **178**, 108732.
  - 223 X. Li, Y. Xu, K. Goh, T. H. Chong and R. Wang, Layer-by-layer assembly based low pressure biocatalytic nanofiltration membranes for micropollutants removal, *J. Membr. Sci.*, 2020, 118514.
  - 224 P. Wang, F. Wang, H. Jiang, Y. Zhang, M. Zhao, R. Xiong, *et al.*, Strong improvement of nanofiltration performance on micropollutant removal and reduction of membrane fouling by hydrolyzed-aluminum nanoparticles, *Water Res.*, 2020, 115649.
  - 225 C. Li, G. Feng, Z. Pan, C. Song, X. Fan, P. Tao, *et al.*, High-performance electrocatalytic microfiltration CuO/Carbon membrane by facile dynamic electrodeposition for small-sized organic pollutants removal, *J. Membr. Sci.*, 2020, **601**, 117913.
  - 226 P. T. P. Aryanti, A. N. Hakim, S. Widodo, I. N. Widiasta and I. G. Wenten, Prospect and challenges of tight ultrafiltration membrane in drinking water treatment, in *IOP Conference Series: Materials Science and Engineering*, IOP Publishing, 2018, p. 12012.
  - 227 M. Zielińska, A. Cydzik-Kwiatkowska, K. Bułkowska, K. Bernat and I. Wojnowska-Baryła, Treatment of Bisphenol A-Containing Effluents from Aerobic Granular Sludge Reactors with the Use of Microfiltration and Ultrafiltration Ceramic Membranes, *Water, Air, Soil Pollut.*, 2017, **228**(8), 282.
  - 228 W. Tian, J. Qiu, N. Li, D. Chen, Q. Xu, H. Li, *et al.*, Efficient piezocatalytic removal of BPA and Cr (VI) with SnS<sub>2</sub>/CNFs membrane by harvesting vibration energy, *Nano Energy*, 2021, **86**, 106036.
  - 229 M. S. Muhamad, N. Hamidon, M. R. Salim, Z. Yusop, W. J. Lau and T. Hadibarata, Response surface methodology for modeling bisphenol a removal using ultrafiltration membrane system, *Water, Air, Soil Pollut.*, 2018, **229**, 1–11.
  - 230 M. Qasim, M. Badrelzaman, N. N. Darwish, N. A. Darwish and N. Hilal, Reverse osmosis desalination: A state-of-the-art review, *Desalination*, 2019, **459**, 59–104.
  - 231 T. D. Aiyelabegan, S. N. A. Shafie, S. M. Hizam and N. A. H. Nordin, The removal of bisphenol-A from synthetic wastewater using thin-film composite membrane, *Malaysian J. Anal. Sci.*, 2022, **26**(3), 507–519.
  - 232 F. Widhiastuti, L. Fan, J. Paz-Ferreiro and K. Chiang, Oxidative treatment of bisphenol A in municipal wastewater reverse osmosis concentrate using Ferrate (VI), *J. Environ. Chem. Eng.*, 2021, **9**(4), 105462.
  - 233 F. Cheng and J. Wang, Removal of bisphenol a from wastewater by adsorption and membrane separation: Performances and mechanisms, *Chem. Eng. J.*, 2024, 149414.
  - 234 P. Pourhakkak, M. Taghizadeh, A. Taghizadeh, M. Ghaedi, Adsorbent, in *Interface Science and Technology*, Elsevier, 2021, pp. 71–210.
  - 235 A. E. D. Mahmoud, Graphene-based nanomaterials for the removal of organic pollutants: Insights into linear *versus*



- nonlinear mathematical models, *J. Environ. Manage.*, 2020, **270**, 110911.
- 236 A. B. Hernández-Abreu, S. Álvarez-Torrellas, V. I. Águeda, M. Larriba, J. A. Delgado, P. A. Calvo, *et al.*, Enhanced removal of the endocrine disruptor compound Bisphenol A by adsorption onto green-carbon materials. Effect of real effluents on the adsorption process, *J. Environ. Manage.*, 2020, 110604.
  - 237 C. E. Choong, S. Ibrahim and W. J. Basirun, Mesoporous silica from batik sludge impregnated with aluminum hydroxide for the removal of bisphenol A and ibuprofen, *J. Colloid Interface Sci.*, 2019, **541**, 12–17.
  - 238 Y. Zhao, C. W. Cho, L. Cui, W. Wei, J. Cai, G. Wu, *et al.*, Adsorptive removal of endocrine-disrupting compounds and a pharmaceutical using activated charcoal from aqueous solution: kinetics, equilibrium, and mechanism studies, *Environ. Sci. Pollut. Res.*, 2019, **26**, 33897–33905.
  - 239 J. Heo, Y. Yoon, G. Lee, Y. Kim, J. Han and C. M. Park, Enhanced adsorption of bisphenol A and sulfamethoxazole by a novel magnetic CuZnFe<sub>2</sub>O<sub>4</sub>-biochar composite, *Bioresour. Technol.*, 2019, **281**, 179–187.
  - 240 N. S. Alhokbany, M. Naushad, V. Kumar, S. Al hatim, S. M. Alshehri and T. Ahamad, Self-Nitrogen doped carbons aerogel derived from waste cigarette butts (cellulose acetate) for the adsorption of BPA: Kinetics and adsorption mechanisms, *J. King Saud Univ. Sci.*, 2020, 3351–3358.
  - 241 X. Tang, P. Tang, S. Si and L. Liu, Adsorption and removal of bisphenol A from aqueous solution by p-phenylenediamine modified magnetic graphene oxide, *J. Serb. Chem. Soc.*, 2017, **82**(1), 39–50.
  - 242 A. Supong, P. C. Bhomick, M. Baruah, C. Pongener, U. B. Sinha and D. Sinha, Adsorptive removal of Bisphenol A by biomass activated carbon and insights into the adsorption mechanism through density functional theory calculations, *Sustainable Chem. Pharm.*, 2019, 100159.
  - 243 M. H. Dehghani, R. R. Karri, M. Alimohammadi, S. Nazmara, A. Zarei and Z. Saeedi, Insights into endocrine-disrupting Bisphenol-A adsorption from pharmaceutical effluent by chitosan immobilized nanoscale zero-valent iron nanoparticles, *J. Mol. Liq.*, 2020, 113317.
  - 244 G. G. Haciosmanoğlu, T. Doğruel, S. Genç, E. T. Oner and Z. S. Can, Adsorptive removal of bisphenol A from aqueous solutions using phosphonated levan, *J. Hazard. Mater.*, 2019, 43–49.
  - 245 M. A. Ahsan, M. T. Islam, C. Hernandez, H. Kim, Y. Lin, M. L. Curry, *et al.*, Adsorptive Removal of Sulfamethoxazole and Bisphenol A from Contaminated Water using Functionalized Carbonaceous Material Derived from Tea Leaves, *J. Environ. Chem. Eng.*, 2018, 4215–4225.
  - 246 X. Guo, Y. Huang, W. Yu, X. Yu, X. Han and H. Zhai, Multi-walled carbon nanotubes modified with iron oxide and manganese dioxide (MWCNTs-Fe<sub>3</sub>O<sub>4</sub>-MnO<sub>2</sub>) as a novel adsorbent for the determination of BPA, *Microchem. J.*, 2020, 104867.
  - 247 K. Noufel, N. Djebri, N. Boukhalfa, M. Boutahala and A. Dakhouché, Removal of bisphenol A and trichlorophenol from aqueous solutions by adsorption with organically modified bentonite, activated carbon composites: A comparative study in single and binary systems, *Groundw. Sustain. Dev.*, 2020, 100477.
  - 248 M. A. Ahsan, V. Jabbari, M. T. Islam, R. S. Turley, N. Dominguez, H. Kim, *et al.*, Sustainable synthesis and remarkable adsorption capacity of MOF/graphene oxide and MOF/CNT based hybrid nanocomposites for the removal of Bisphenol A from water, *Sci. Total Environ.*, 2019, 306–317.
  - 249 T. A. Saleh, M. Tuzen and A. Sari, Magnetic vermiculite-modified by poly (trimesoyl chloride-melamine) as a sorbent for enhanced removal of bisphenol A, *J. Environ. Chem. Eng.*, 2019, **7**(6), 103436.
  - 250 J. Hao, Q. Zhang, Y. Liu, P. Chen, X. Zheng, X. Zhuang, *et al.*, A novel nitrogen-containing covalent organic framework adsorbent for the efficient removal of bisphenol A from aqueous solution, *J. Taiwan Inst. Chem. Eng.*, 2020, **113**, 204–213.
  - 251 X. Tang, P. Tang, S. Si and L. Liu, Adsorption and removal of bisphenol A from aqueous solution by p-phenylenediamine modified magnetic graphene oxide, *J. Serb. Chem. Soc.*, 2017, **82**(1), 39–50.
  - 252 Z. Gong, S. Li, J. Ma and X. Zhang, Synthesis of recyclable powdered activated carbon with temperature responsive polymer for bisphenol A removal, *Sep. Purif. Technol.*, 2016, **157**, 131–140.
  - 253 X. Chen, T. Fujiwara, S. Fukahori and T. Ishigaki, Factors affecting the adsorptive removal of bisphenol A in landfill leachate by high silica Y-type zeolite, *Environ. Sci. Pollut. Res.*, 2015, **22**, 2788–2799.
  - 254 M. H. Dehghani, A. H. Mahvi, N. Rastkari, R. Saeedi, S. Nazmara and E. Irvani, Adsorption of bisphenol A (BPA) from aqueous solutions by carbon nanotubes: kinetic and equilibrium studies, *Desalination Water Treat.*, 2015, **54**(1), 84–92.
  - 255 J. Xu, L. Wang and Y. Zhu, Decontamination of bisphenol A from aqueous solution by graphene adsorption, *Langmuir*, 2012, **28**(22), 8418–8425.
  - 256 T. Bao, M. M. Damtie, W. Wei, H. N. P. Vo, K. H. Nguyen, A. Hosseinzadeh, *et al.*, Simultaneous adsorption and degradation of bisphenol A on magnetic illite clay composite: Eco-friendly preparation, characterizations, and catalytic mechanism, *J. Cleaner Prod.*, 2021, **287**, 125068.
  - 257 R. Kumar, M. Qureshi, D. K. Vishwakarma, N. Al-Ansari, A. Kuriqi, A. Elbeltagi, *et al.*, A review on emerging water contaminants and the application of sustainable removal technologies, *Case Stud. Chem. Environ. Eng.*, 2022, **6**, 100219.
  - 258 F. Boysan and A. Çavunt, Investigation of bisphenol A removal using peroxy electrocoagulation method, *Desalin. Water Treat.*, 2021, **222**, 426–433.



- 259 Y. Tang, X. Yin, M. Mu, Y. Jiang, X. Li, H. Zhang, *et al.*, Anatase TiO<sub>2</sub>@ MIL-101 (Cr) nanocomposite for photocatalytic degradation of bisphenol A, *Colloids Surf., A*, 2020, **596**, 124745.
- 260 M. Dolatabadi, R. Malekahmadi, A. Ghorbanian and S. Ahmadzadeh, Investigation of electrocoagulation process for efficient removal of bisphenol A from the aqueous environment: Promising treatment strategy, *J. Environ. Health Sustainable Dev.*, 2021, 1275–1283.
- 261 Y. Huang, M. Su, Y. Zhou, D. Chen, Z. Xu, H. Zhang, *et al.*, LiCl-CN nanotubes ceramic films with highly efficient visible light - Driven photocatalytic active for bisphenol A degradation and efficient regeneration process, *Ceram. Int.*, 2020, 26492–26501.
- 262 A. Chmayssem, S. Taha and D. Hauchard, Scaled-up electrochemical reactor with a fixed bed three-dimensional cathode for electro-Fenton process: Application to the treatment of bisphenol A, *Electrochim. Acta*, 2017, **225**, 435–442.
- 263 Z. Pan, F. Yu, L. Li, C. Song, J. Yang, C. Wang, *et al.*, Electrochemical microfiltration treatment of bisphenol A wastewater using coal-based carbon membrane, *Sep. Purif. Technol.*, 2019, **227**, 115695.
- 264 S. Li, G. Zhang, P. Wang, H. Zheng and Y. Zheng, Microwave-enhanced Mn-Fenton process for the removal of BPA in water, *Chem. Eng. J.*, 2016, **294**, 371–379.
- 265 S. Parastar, A. Ebrahimi, H. Mohammadi, N. Mengelizadeh, I. Parseh, N. Rafiei, *et al.*, Removal of BPA from aqueous solutions by electrocoagulation using iron electrodes and optimization, *J. Environ. Health Eng.*, 2018, **5**(3), 264–276.
- 266 A. Aziz, P. Agamuthu and S. H. Fauziah, Removal of bisphenol A and 2, 4-Di-tert-butylphenol from landfill leachate using plant-based coagulant, *Waste Manag. Res.*, 2018, **36**(10), 975–984.
- 267 J. Bai, Y. Li, B. Song and Q. Wang, Activation of peroxymonosulfate by modified coagulation sludge for bisphenol A degradation, *Environ. Sci. Pollut. Res.*, 2022, **29**(52), 78832–78847.
- 268 E. Bazrafshan, H. Biglari and A. H. Mahvi, Phenol removal by electrocoagulation process from aqueous solutions, *Fresenius Environ. Bull.*, 2012, **21**(2), 364–371.
- 269 N. Ambauen, C. Weber, J. Muff, C. Hallé and T. Meyn, Electrochemical removal of Bisphenol A from landfill leachate under Nordic climate conditions, *J. Appl. Electrochem.*, 2020, **50**, 1175–1188.
- 270 S. Parastar, A. Ebrahimi, H. Mohammadi, N. Mengelizadeh, I. Parseh, N. Rafiei, *et al.*, Removal of BPA from aqueous solutions by electrocoagulation using iron electrodes and optimization, *J. Environ. Health Eng.*, 2018, **5**(3), 264–276.
- 271 P. Guerra, M. Kim, S. Teslic, M. Alaei and S. A. A. Smyth, Bisphenol-A removal in various wastewater treatment processes: Operational conditions, mass balance, and optimization, *J. Environ. Manage.*, 2015, **152**, 192–200.
- 272 M. Umar, F. Roddick, L. Fan and H. A. Aziz, Application of ozone for the removal of bisphenol A from water and wastewater – A review, *Chemosphere*, 2013, **90**(8), 2197–2207.
- 273 X. Si, Z. Hu and S. Huang, Combined Process of Ozone Oxidation and Ultrafiltration as an Effective Treatment Technology for the Removal of Endocrine-Disrupting Chemicals, *Appl. Sci.*, 2018, **8**(8), 1240.
- 274 M. Dudziak and E. Burdzik-Niemiec, Comparative studies on elimination of estrogens and xenoestrogens by the oxidation processes, *Ecol. Chem. Eng. A*, 2014, **21**(2), 189–198.
- 275 M. Deborde, S. Rabouan, P. Mazellier, J. P. P. Duguet and B. Legube, Oxidation of bisphenol A by ozone in aqueous solution, *Water Res.*, 2008, **42**(16), 4299–4308.
- 276 T. Garoma, S. A. Matsumoto, Y. Wu and R. Klinger, Removal of Bisphenol A and its Reaction-Intermediates from Aqueous Solution by Ozonation, *Ozone: Sci. Eng.*, 2010, **32**(5), 338–343.
- 277 F. Du, Z. Lai, H. Tang, H. Wang and C. Zhao, Construction of dual Z-scheme Bi<sub>2</sub>WO<sub>6</sub>/g-C<sub>3</sub>N<sub>4</sub>/black phosphorus quantum dots composites for effective bisphenol A degradation, *J. Environ. Sci.*, 2023, **124**, 617–629.
- 278 E. García-Díaz, D. Zhang, Y. Li, R. Verduzco and P. J. J. Alvarez, TiO<sub>2</sub> microspheres with cross-linked cyclodextrin coating exhibit improved stability and sustained photocatalytic degradation of bisphenol A in secondary effluent, *Water Res.*, 2020, **183**, 116095.
- 279 H. Wang, N. Zhang, G. Cheng, H. Guo, Z. Shen, L. Yang, *et al.*, Preparing a photocatalytic Fe doped TiO<sub>2</sub>/rGO for enhanced bisphenol A and its analogues degradation in water sample, *Appl. Surf. Sci.*, 2020, **505**, 144640.
- 280 Z. You, Z. Wang, K. Tian and X. Yao, Activation of persulfate by MIL-100(Fe) coated with perylene diimide heterojunction photocatalyst for efficient removal of Bisphenol A in visible light, *J. Environ. Chem. Eng.*, 2023, **11**(3), 110159.
- 281 Y. Zhang and J. L. Zhou, Occurrence and removal of endocrine disrupting chemicals in wastewater, *Chemosphere*, 2008, **73**(5), 848–853.
- 282 I. Sánchez-Montes, N. Wachter, B. F. Silva and J. M. Aquino, Comparison of UVC-based advanced oxidation processes in the mineralization of bisphenol A: Identification of oxidation by products and toxicity evaluation, *Chem. Eng. J.*, 2020, **386**, 123986.
- 283 B. Huang, X. Li, W. Sun, D. Ren, X. Li, X. Li, *et al.*, Occurrence, removal, and fate of progestogens, androgens, estrogens, and phenols in six sewage treatment plants around Dianchi Lake in China, *Environ. Sci. Pollut. Res.*, 2014, **21**(22), 12898–12908.
- 284 A. Wang, Z. Hua, Z. Wu, C. Chen, S. Hou, B. Huang, *et al.*, Insights into the effects of bromide at fresh water levels on the radical chemistry in the UV/persulfate process, *Water Res.*, 2021, **197**, 117042.
- 285 A. Cai, J. Deng, X. Ling, C. Ye, H. Sun, Y. Deng, *et al.*, Degradation of bisphenol A by UV/persulfate process in the presence of bromide: Role of reactive bromine, *Water Res.*, 2022, **215**, 118288.
- 286 H. Abu Hasan, M. H. Muhamad, S. Budi Kurniawan, J. Buhari and O. Husain Abuzeyad, Managing Bisphenol A



- Contamination: Advances in Removal Technologies and Future Prospects, *Water*, 2023, **15**(20), 3573.
- 287 E. Zhang, P. Zhao, G. Xu, F. Meng, X. Wang, Y. Gao, *et al.*, High efficiency manganese cobalt spinel structure catalytic ozonation ceramic membrane for *in situ* BPA degradation and membrane fouling elimination, *J. Environ. Chem. Eng.*, 2024, **12**(1), 111774.
  - 288 S. Wang, Q. Wu, B. Yan, Y. Guo, W. Xia, J. Li, *et al.*, A novel integrated process of ceramic membrane filtration coupled with peroxymonosulfate activation and adsorption for water treatment, *Sep. Purif. Technol.*, 2022, 120874.
  - 289 J. Xie, Z. Liao, M. Zhang, L. Ni, J. Qi, C. Wang, *et al.*, Sequential ultrafiltration-catalysis membrane for excellent removal of multiple pollutants in water, *Environ. Sci. Technol.*, 2021, 2652–2661.
  - 290 Y. Huang, W. Xu, L. Hu, J. Zeng, C. He, X. Tan, *et al.*, Combined adsorption and catalytic ozonation for removal of endocrine disrupting compounds over MWCNTs/Fe<sub>3</sub>O<sub>4</sub> composites, *Catal. Today*, 2017, **297**, 143–150.
  - 291 Z. Li, Z. Yang, Q. Shi and Y. Sun, Removal of bisphenol A by integrated adsorption and biodegradation using immobilized laccase onto defective PCN-224, *J. Environ. Chem. Eng.*, 2023, **11**(3), 110166.
  - 292 I. Escalona, J. de Grooth, J. Font and K. Nijmeijer, Removal of BPA by enzyme polymerization using NF membranes, *J. Membr. Sci.*, 2014, 192–201.
  - 293 C. Hu, D. Huang, G. Zeng, M. Cheng, X. Gong, R. Wang, *et al.*, The combination of Fenton process and *Phanerochaete chrysosporium* for the removal of bisphenol A in river sediments: Mechanism related to extracellular enzyme, organic acid and iron, *Chem. Eng. J.*, 2018, 432–439.
  - 294 G. Bertanza, R. Pedrazzani, M. Dal Grande, M. Papa, V. Zambarda, C. Montani, *et al.*, Effect of biological and chemical oxidation on the removal of estrogenic compounds (NP and BPA) from wastewater: An integrated assessment procedure, *Water Res.*, 2011, 2473–2484.
  - 295 J. F. d. Oliveira, R. Fia, F. N. Rodrigues, F. R. L. Fia, M. P. d. Matos, L. A. B. Siniscalchi, *et al.*, Quantification, removal and potential ecological risk of emerging contaminants in different organic loads of swine wastewater treated by integrated biological reactors, *Chemosphere*, 2020, **260**, 127516.
  - 296 A. Tarafdar, R. Sirohi, P. A. Balakumaran, R. Reshmy, A. Madhavan, R. Sindhu, *et al.*, The hazardous threat of Bisphenol A: Toxicity, detection and remediation, *J. Hazard. Mater.*, 2022, 127097.
  - 297 C. B. Godiya and B. J. Park, Removal of bisphenol A from wastewater by physical, chemical and biological remediation techniques. A review, *Environ. Chem. Lett.*, 2022, **20**(3), 1801–1837.
  - 298 M. Zielińska, I. Wojnowska-Baryła and A. Cydzik-Kwiatkowska, Integrated Systems for Removal of BPA from Wastewater, in *Bisphenol A Removal from Water and Wastewater*, 2019.
  - 299 M. F. M. A. Zamri, R. Bahru, F. Suja', A. H. Shamsuddin, S. K. Pramanik and I. M. R. Fattah, Treatment strategies for enhancing the removal of endocrine-disrupting chemicals in water and wastewater systems, *J. Water Proc. Eng.*, 2021, 102017.
  - 300 F. M. Mpatani, R. Han, A. A. Aryee, A. N. Kani, Z. Li and L. Qu, Adsorption performance of modified agricultural waste materials for removal of emerging micro-contaminant bisphenol A: a comprehensive review, *Sci. Total Environ.*, 2021, **780**, 146629.
  - 301 Y. Zhang, Y. Cheng, N. Chen, Y. Zhou, B. Li, W. Gu, *et al.*, Recyclable removal of bisphenol A from aqueous solution by reduced graphene oxide–magnetic nanoparticles: adsorption and desorption, *J. Colloid Interface Sci.*, 2014, **421**, 85–92.
  - 302 N. H. Dang, T. H. Tu, V. N. P. Linh, L. T. M. Thy, H. M. Nam, M. T. Phong, *et al.*, Preparation of magnetic iron oxide/graphene aerogel nanocomposites for removal of bisphenol A from water, *Synth. Met.*, 2019, **255**, 116106.
  - 303 K. Ouyang, C. Zhu, Y. Zhao, L. Wang, S. Xie and Q. Wang, Adsorption mechanism of magnetically separable Fe<sub>3</sub>O<sub>4</sub>/graphene oxide hybrids, *Appl. Surf. Sci.*, 2015, **355**, 562–569.
  - 304 B. Seyhi, P. Drogui, G. Buelna, A. Azaïs and M. Heran, Contribution of a submerged membrane bioreactor in the treatment of synthetic effluent contaminated by Bisphenol-A: mechanism of BPA removal and membrane fouling, *Environ. Pollut.*, 2013, **180**, 229–235.
  - 305 M. J. F. Jasni, M. Arulkumar, P. Sathishkumar, A. R. M. Yusoff, N. A. Buang and F. L. Gu, Electrospun nylon 6, 6 membrane as a reusable nano-adsorbent for bisphenol A removal: Adsorption performance and mechanism, *J. Colloid Interface Sci.*, 2017, **508**, 591–602.
  - 306 L. E. Koloti, N. P. Gule, O. A. Arotiba and S. P. Malinga, Laccase-immobilized dendritic nanofibrous membranes as a novel approach towards the removal of bisphenol A, *Environ. Technol.*, 2018, **39**(3), 392–404.
  - 307 S. Yüksel, N. Kabay and M. Yüksel, Removal of bisphenol A (BPA) from water by various nanofiltration (NF) and reverse osmosis (RO) membranes, *J. Hazard. Mater.*, 2013, **263**, 307–310.
  - 308 P. A. K. Reddy, P. V. L. Reddy, E. Kwon, K. H. Kim, T. Akter and S. Kalagara, Recent advances in photocatalytic treatment of pollutants in aqueous media, *Environ. Int.*, 2016, **91**, 94–103.
  - 309 V. Repousi, A. Petala, Z. Frontistis, M. Antonopoulou, I. Konstantinou, D. I. Kondarides, *et al.*, Photocatalytic degradation of bisphenol A over Rh/TiO<sub>2</sub> suspensions in different water matrices, *Catal. Today*, 2017, **284**, 59–66.
  - 310 P. V. L. Reddy, K. H. Kim, B. Kavitha, V. Kumar, N. Raza and S. Kalagara, Photocatalytic degradation of bisphenol A in aqueous media: A review, *J. Environ. Manage.*, 2018, **213**, 189–205.
  - 311 X. Xiao, S. Tu, M. Lu, H. Zhong, C. Zheng, X. Zuo, *et al.*, Discussion on the reaction mechanism of the photocatalytic degradation of organic contaminants from a viewpoint of semiconductor photo-induced electrocatalysis, *Appl. Catal. B*, 2016, **198**, 124–132.



- 312 M. Umar, F. Roddick, L. Fan and H. A. Aziz, Application of ozone for the removal of bisphenol A from water and wastewater—a review, *Chemosphere*, 2013, **90**(8), 2197–2207.
- 313 E. Kusvuran and D. Yildirim, Degradation of bisphenol A by ozonation and determination of degradation intermediates by gas chromatography–mass spectrometry and liquid chromatography–mass spectrometry, *Chem. Eng. J.*, 2013, **220**, 6–14.
- 314 H. Pokkiladathu, S. Farissi, A. Sakkarai and M. Muthuchamy, Degradation of bisphenol A: a contaminant of emerging concern, using catalytic ozonation by activated carbon impregnated nanocomposite-bimetallic catalyst, *Environ. Sci. Pollut. Res.*, 2022, **29**(48), 72417–72430.
- 315 R. Shokoohi, H. Zolghadrnasab, S. Shanehsaz, M. Leili, M. Shaygan and F. Azizi, Efficiency of the catalytic ozonation processes using nanoparticles deposited on pumice in the removal of bisphenol A, *Int. J. Environ. Anal. Chem.*, 2023, **103**(13), 3121–3137.
- 316 H. Zhang, Y. He, L. Lai, G. Yao and B. Lai, Catalytic ozonation of Bisphenol A in aqueous solution by Fe<sub>3</sub>O<sub>4</sub>–MnO<sub>2</sub> magnetic composites: Performance, transformation pathways and mechanism, *Sep. Purif. Technol.*, 2020, **245**, 116449.
- 317 H. Guo, H. Wang, Q. Wu and J. Li, Degradation and mechanism analysis of bisphenol A in aqueous solutions by pulsed discharge plasma combined with activated carbon, *Sep. Purif. Technol.*, 2018, **190**, 288–296.
- 318 M. R. Samarghandi, A. Ansari, A. Dargahi, A. Shabanloo, D. Nematollahi, M. Khazaei, *et al.*, Enhanced electrocatalytic degradation of bisphenol A by graphite/β-PbO<sub>2</sub> anode in a three-dimensional electrochemical reactor, *J. Environ. Chem. Eng.*, 2021, **9**(5), 106072.
- 319 S. Ahmadzadeh and M. Dolatabadi, Modeling and kinetics study of electrochemical peroxidation process for mineralization of bisphenol A; a new paradigm for groundwater treatment, *J. Mol. Liq.*, 2018, **254**, 76–82.
- 320 Y.-h. Cui, L. X. yan and G. Chen, Electrochemical degradation of bisphenol A on different anodes, *Water Res.*, 2009, **43**(7), 1968–1976.
- 321 M. D. Simić, T. P. Brdarić, B. G. S. Rosić, Ĺ. Švorc, D. J. Relić and D. D. Āćimović, Degradation of bisphenol A *via* the electro–Fenton process using nanostructured carbon-metal oxide anodes: Intermediates and reaction mechanisms study, *J. Environ. Chem. Eng.*, 2024, **12**(5), 113369.
- 322 J. da Silveira Salla, K. da Boit Martinello, G. L. Dotto, E. García-Díaz, H. Javed, P. J. J. Alvarez, *et al.*, Synthesis of citrate-modified CuFeS<sub>2</sub> catalyst with significant effect on the photo–Fenton degradation efficiency of bisphenol a under visible light and near-neutral pH, *Colloids Surf., A*, 2020, **595**, 124679.
- 323 Y. Liu, Y. Mao, X. Tang, Y. Xu, C. Li and F. Li, Synthesis of Ag/AgCl/Fe-S plasmonic catalyst for bisphenol A degradation in heterogeneous photo-Fenton system under visible light irradiation, *Chin. J. Catal.*, 2017, **38**(10), 1726–1735.
- 324 G. Li, L. Zu, P. K. Wong, X. Hui, Y. Lu, J. Xiong, *et al.*, Biodegradation and detoxification of bisphenol A with one newly-isolated strain *Bacillus* sp. GZB: kinetics, mechanism and estrogenic transition, *Bioresour. Technol.*, 2012, **114**, 224–230.
- 325 Y. Jia, A. Eltoukhy, J. Wang, X. Li, T. S. Hlaing, M. M. Aung, *et al.*, Biodegradation of bisphenol A by *Sphingobium* sp. YC-JY1 and the essential role of cytochrome P450 monooxygenase, *Int. J. Mol. Sci.*, 2020, **21**(10), 3588.
- 326 M. Golshan, S. Jorfi, N. J. Haghighifard, A. Takdastan, S. Ghafari, S. Rostami, *et al.*, Development of salt-tolerant microbial consortium during the treatment of saline bisphenol A-containing wastewater: Removal mechanisms and microbial characterization, *J. Water Proc. Eng.*, 2019, **32**, 100949.
- 327 R. Wang, P. Diao, Q. Chen, H. Wu, N. Xu and S. Duan, Identification of novel pathways for biodegradation of bisphenol A by the green alga *Desmodesmus* sp. WR1, combined with mechanistic analysis at the transcriptome level, *Chem. Eng. J.*, 2017, **321**, 424–431.

