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High throughput laser ablation ICP-MS bioimaging of silver distribution in animal organisms and plant tissue after exposure to silver sulfide nanoparticles†

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Laser ablation inductively coupled plasma mass spectrometry (LA-ICP-MS) is a method with high potential to visualize the distribution of elements in different samples, including a variety of organisms. This study aimed to demonstrate the broad application of LA-ICP-MS in nanomaterial-biota fate studies as a high throughput bioimaging tool and to broaden the choice of standard organism models in material-biota interaction research. In this study, operation parameters of LA-ICP-MS were optimised on one organism, woodlice *Porcellio scaber*. Subsequently, the scanning conditions were tested on a range of aquatic (*Girardia tigrina*, *Lumbriculus variegatus*, and *Oncorhynchus mykiss*), terrestrial organisms (*Lumbricus rubellus*, *Porcellio scaber*) and one plant (*Triticum aestivum*) upon exposure to Ag₂S nanoparticles (NPs) and silver nitrate (AgNO₃). Model organisms were exposed in aquatic or terrestrial mesocosm experiments where nominal concentrations of Ag were 10 µg Ag per L of water and 10 mg Ag per kg of soil, respectively. The results showed that both sample preparation and LA-ICP-MS imaging conditions, as optimized on the selected organism (65 µm laser diameter, scan rate 100 µm s⁻¹, measuring duration 35 min), are applicable on different tissues. These LA-ICP-MS imaging conditions enable recognition of the main biological structures and biodistribution of elements of interest. By using fast-screening LA-ICP-MS, we confirmed the presence of Ag₂S NPs on the body surface or in the gut lumen (adsorbed and retained), but not in other internal parts of organisms, which is consistent with our previous toxicokinetic studies. The presence of Ag was also confirmed in some parts of wheat roots. The advantage of this technique is the possibility of sequential use of fast- and slow-scanning steps to optimise the duration of analysis and data processing, whilst also improving cost-effectiveness without compromising the quality of results.

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Introduction

The uptake, tissue distribution, and fate of internalized nanomaterials (NMs) result from complex interactions between the particles and organisms. Unless they dissolve, intact NMs are usually taken up across epithelia by endocytosis-related mechanisms, or in the case of pathology they presumably enter the organism at damaged surfaces.¹ The concern is that the

colloidal behaviour and reactivities of NMs are not the same as those of solutes, and consequently NMs may have different target organs and fates inside the organism compared to traditional chemicals. However, the target organs for NMs are poorly understood, especially across the different anatomies of biological taxa.

The accumulation and fate of metals in organisms exposed to metal-based NMs have been the most studied. Traditional methods of metal analysis are used to analyse a whole tissue, an organ, or sometimes a whole organism to obtain the total metal concentration in the sample. This approach has been useful in developing toxicokinetic models of uptake and excretion of metals from NM exposures.^{2–4} However, the total metal concentration from a tissue digest does not identify if intact particles are present, and even the recent progress on single particle inductively coupled mass spectrometry (sp-ICP-MS) to identify particles in biological samples cannot identify the complex spatial pattern within an organ(s) *in vivo*.

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Different imaging techniques are available to complement these former approaches with spatial information on the location of the metal inside the organism.^{5,6} Among them, the laser ablation inductively coupled plasma mass spectrometry (LA-ICP-MS) technique is becoming a primary analytical tool for imaging the distribution of metals as well as metal-based NPs in cell or tissue samples. Galazzi *et al.*⁷ and Doble *et al.*⁸ reviewed the contribution of ICP-MS-based platforms as tools for evaluating NP behavior in distinct biological samples.

The basic principle of LA-ICP-MS is that the upper layer of the dry sample (whole organism or lyophilized slices of tissue sections) is ablated (*i.e.*, vapourised or sublimed) by energy from a laser and the ions/atoms given off are then detected by inductively coupled plasma mass spectrometry.^{9,10} For a graphical representation of the selected elemental distributions in the sample, the data from LA-ICP-MS measurements are processed and converted into maps. The quality of the maps depends mostly on the laser diameter (LD), scanning speed (SS), and dwell time (DT). In addition, LA-ICP-MS allows scanning either large tissue sections or small spot sizes for a detailed analysis.¹¹ LA-ICP-MS is a very sensitive method with detection limits of imaging in the $\mu\text{g g}^{-1}$ to sub $\mu\text{g g}^{-1}$ range and with a lateral resolution between 5 μm and 200 μm .¹² In addition to providing elemental tissue distribution, LA-ICP-MS can be combined with a single particle (SP) counting technique and can provide indirect visualization of the number and size of NPs at a tissue level together with discrimination between ions and particles.¹³ Namely, in LA-SP-ICP-MS metals in ionic forms behave differently upon ablation and can therefore be separated in post-processing.¹⁴ An important advantage of using LA-ICP-MS is a relatively simple sample preparation. The gold standard in tissue preparation for LA-ICP-MS is a routine histological method of chemical fixation and paraffin embedding (FFPE).¹⁵

LA-ICP-MS has already been applied to different biological samples to study the presence of NPs in cells or their distribution in tissues. For example, Au NPs have been investigated in a mouse leukemic monocyte macrophage cell line, and Ag NPs in alveolar macrophages. SeCd/ZnS quantum dot or Au and Ag NP distribution was studied in fibroblasts.^{6,16,17} At the organ level, Au NPs were inspected in the mouse liver.¹⁸ Doped Fe-oxide particles have been studied in zebrafish larvae and in mouse models showing atherosclerotic plaques.^{19–21} There are also some reports of elemental imaging in plants exposed to NPs.^{8,22} Yamashita *et al.* showed the size and position of Ag and Au nanoparticles in onion cells, and Metarapi *et al.* examined the NP distribution in the roots of Ag⁺ exposed plants.^{14,23} However, much less data exists for invertebrates such as Annelida, Mollusca, and Arthropoda or complex vertebrate samples when NMs are studied. The exceptions are two studies by Böhme *et al.*^{20,24} who provided evidence on the applicability of LA-ICP-MS to study the crustacean *Daphnia magna* and zebrafish *Danio rerio* larvae exposed to engineered nanoparticles. In the study from 2015, they found Al₂O₃, Ag, and Au NMs attached to *D. rerio* larvae chorion and in the gut of *D. magna*.²⁰ In the follow-up study they showed differences in the distribution of Ag, Au, CuO, and ZnO NPs.²⁴

The technique has also been used for studying the elemental distribution profile of organisms²⁵ or imaging different trace elements and isotopes in different biological samples.^{26,27} However, to fully explore the potential of LA-ICP-MS as a reliable tool for biological samples, the selection of optimal conditions for (multi)elemental fast or slow mode mapping is needed.²⁸

In the present study, the application of LA-ICP-MS was tested for a fast screening of the presence and distribution of selected elements, in our case primarily Ag and also some other elements commonly present in biological samples, in different animal and plant tissue samples after exposure to silver sulfide NPs (Ag₂S NPs) in terrestrial or aquatic mesocosm experiments. Ag NPs are among the most commonly used NPs in commercial products, and their production is still on the upswing. Therefore, their deposition in the environment is inevitable, and they are expected to enter the soil as one of the main repositories.²⁹ When Ag NPs enter the soil, the majority are transformed into poorly soluble Ag sulfide NPs (Ag₂S NPs); however they are also available as Ag NPs as well as Ag ions.^{4,30} The distribution of Ag NPs in biota as well as their biokinetics have been studied extensively.^{4,31,32}

We tested the hypothesis that the Ag distribution pattern in Ag₂S NP and AgNO₃ exposed organisms is different. The work here was organized into three steps: (a) method optimization in a single-species experiment with the terrestrial isopod *Porcellio scaber*, then (b) testing the parameters in an environmentally realistic set-up on the same organism, and finally (c) testing selected scanning parameters and sample preparation on a wide range of biological samples [terrestrial species: earthworms (*Lumbricus rubellus*), crustaceans (*Porcellio scaber*), and plants (*Triticum aestivum* L.) and aquatic species: planarians (*Girardia tigrina*), blackworms (*Lumbriculus variegatus*), and the rainbow trout intestine (*Oncorhynchus mykiss*)] to find the best compromise between the scanning speed and image resolution. The data were also used to show the utility of LA-ICP-MS for the fast screening of the metal/NM content in organisms with different anatomy and biological matrices either as a basis for other more advanced techniques or to confirm the outcomes of other approaches (*i.e.*, toxicokinetic studies).

Methods

Nanoparticle characterization

For the single-species tests with woodlice, polyvinylpyrrolidone (PVP) coated sulfidised silver nanoparticle (Ag₂S NP) colloids with a mean size 20 nm were used in the study. Particles were provided by Applied Nanoparticles (Barcelona, Spain). Detailed particle characteristics are provided in the ESI† (Fig. S1). A colloidal suspension of Ag₂S NPs (also provided by Applied Nanoparticles) was used for the terrestrial and aquatic mesocosm experiments. The Ag₂S NPs were also PVP-coated with a reported size by the manufacturer of 20.4 ± 11.9 nm (mean $\pm$ S.D., $n = 613$). Detailed information on the characteristics of these Ag₂S NP colloids is described by Peixoto *et al.*³⁰ Silver nitrate (AgNO₃) was used as a metal salt control (Sigma-Aldrich, 99% purity, CAS 7761–88–8, Germany).



Single-species test and optimization of LA-ICP-MS

Woodlice exposure conditions and sample preparation. For the method optimization process, terrestrial woodlice (*Porcellio scaber*, Isopoda, Crustacea) were exposed to a sub-lethal concentration of Ag in Lufa 2.2 soil as either AgNO₃ (500 mg per kg of soil) or Ag₂S NPs (10 mg per kg of soil). A control group was exposed to Lufa 2.2 soil with no Ag added. Freshly prepared suspensions of Ag₂S NPs and AgNO₃ were mixed with soil using a spatula, and deionized water was added to reach a moisture content of 40% of the water holding capacity. Each experimental unit consisted of a glass jar with a lid containing 20 g of moist soil from the same batch of spiked soil, a piece of dried hazel leaf as a food source (~100 mg), and five randomly selected *P. scaber* specimens (both sexes, weight >25 mg). Mortality was monitored during the 14-day experiment. After 14 days of exposure, isopods were transferred to untreated soil for 24 h to depurate the gut. Afterwards, animals were prepared for histology observation according to the protocol described by Lešer *et al.*^{27,33} Briefly, the whole animals were preserved in Carnoy B fixative (10 ml of acetic acid, 30 ml of chloroform, and 60 ml absolute alcohol), embedded in paraffin, and 8 µm cross sections were stained with hematoxylin and eosin for light microscopy. Histological samples for LA-ICP-MS analysis were unstained.

LA-ICP-MS optimization process. The optimization aimed to find a compromise between mapping (scanning) time and image resolution. LA-ICP-MS imaging that is of low resolution was complemented with histological images in order to better identify anatomy and regions of interest.

The first step of the LA-ICP-MS imaging optimization process was performed on paraffin cross sections of woodlice exposed to AgNO₃, at 500 mg Ag per kg of soil as high concentrations of silver were expected to be found in the digestive tissue. The selected operating parameters were further tested on slices from animals exposed to Ag₂S NPs with a concentration of 10 mg of Ag per kg of soil or an unexposed control soil. The variables tested in the LA-ICP-MS optimization phase are described in Table 1. Four unstained slices of the same animal from each exposure group (control, AgNO₃, and Ag₂S NPs) were used for selecting optimal measuring parameters and to assess the within-animal variation (precision) of the technique.

The instrument for LA-ICP-MS has a laser ablation system (193 nm ArF* excimer; Analyte G2, Teledyne Photon Machines Inc., Bozeman, MT), with a 2-volume ablation cell (HelEx II; He carrier gas flow rate, cell = 0.5 L min⁻¹ and cup = 0.3 L min⁻¹). The LA unit was interfaced with a quadrupole ICP-MS (Agilent 7900, Agilent Technologies, Palo Alto-CA, USA); Ar make-up gas was added before the ICP torch (0.8 L min⁻¹) and MS operation was in time-resolved mode, measuring one point per mass, acquiring the following isotope masses for all analysed samples: ³⁴S, ³⁵Cl, ⁴⁴Ca, ⁶³Cu, ⁶⁶Zn, and ¹⁰⁷Ag. In addition to Ag, these elements were chosen to be measured as they outline the anatomy or are present at specific locations (digestive glands) and thus serve as reference points in the elemental maps. The scan direction in all experiments was horizontal and bidirectional, *i.e.* exchanging the directions from left to right and right

to left, alternately. The lines of odd numbers were recorded in the direction left-to-right, while the lines of even numbers were recorded in the opposite direction, right-to-left, in order to reduce the total analysis time.

LA-ICP-MS analyses of terrestrial and aquatic organisms exposed to Ag₂S NPs and AgNO₃ in a mesocosm set-up

Terrestrial and aquatic mesocosm experiments. After the optimization experiment, the applicability of selected LA-ICP-MS parameters to a range of environmentally relevant organisms was tested. Different terrestrial and aquatic organisms were exposed to Ag₂S NPs or AgNO₃ serving as a metal salt positive control. The detailed terrestrial mesocosm study is described by Peixoto *et al.*³⁰ Briefly, soil was spiked with Ag₂S NPs and AgNO₃ to reach the final nominal concentration of 10 mg Ag per kg of soil. Ten isopods (*P. scaber*), ten mealworms (*Tenebrio molitor*), six earthworms (*L. rubellus*) and four plants (*T. aestivum* L.) were placed in the soil columns with Ag₂S NPs or AgNO₃ spiked soil on the top layer (8.25 cm of Ag-spiked soil on top of 8.25 cm uncontaminated soil) and non-exposed soil for the control group. The bottom layer of all columns was filled with clean soil. Organisms were sampled for different analyses at four timepoints (0, 7, 14, and 28 days) and individual organisms were prepared for LA-ICP-MS when possible. Earthworms and wheat were sampled and prepared for the measurements at all timepoints, and isopods only at days 0 and 7 (as not enough animals survived until day 28). Mealworms were not included in the LA-ICP-MS measurements.

For the aquatic organisms, an indoor modular mesocosm system was used to simulate a stream environment. For a detailed description of the experimental design, see Clark *et al.*³⁴ Briefly, the system was composed of 36 artificial rivers made of glass arranged in triplicate, and four sets of triplicate were assigned to controls, Ag₂S NPs, and AgNO₃. The bottom of each stream was covered by a layer of sediment mixed with ground alder leaves, and streams were filled with an enriched artificial pond water (APW) medium. The following species were introduced in each artificial stream: *G. tigrina* (planarian), *Physa acuta* (snail), *Chironomus riparius* (midge larvae), *L. variegatus* (blackworm), *D. magna* (water flea), and the rainbow trout *O. mykiss* (daphnids and fish were kept separately in submerged plastic chambers). The water was spiked daily to reach a nominal concentration of 10 µg Ag per L of water, and the system operated in recirculation mode. The experiment lasted 14 days, and destructive sampling was performed for all organisms on days 2, 7, and 14 (organisms were also sampled at time 0). Fish were dissected into the intestine, liver, gill, muscle, and remaining carcass.

The selected organisms or tissues from both mesocosm studies were immediately preserved in Carnoy B fixative and histological sections were prepared as described by Lešer *et al.* Sections (8 µm) were stained with eosin and haematoxylin in order to define the organism anatomy/morphology and some were left unstained for the LA-ICP-MS measurements, as in the initial isopod study.³³ For each exposure group (control, NPs, or ionic control) one or two slices per animal were measured.



Table 1 Detailed LA-ICP-MS operating conditions for the analysis of biological samples in the optimization procedure

Abbreviation	Laser ablation	
BD (=BS)	Wavelength	193 nm
	Laser ablation chamber	Fast washout 2-volume HelEx II cell
	Ablation mode	Raster (bidirectional line scanning)
	Beam diameter (=beam size); laser beam spot diameter	65 μm or 35 μm circle
	Line length	2.0–4.0 mm
	Sample dimensions (=sample size)	approx. $2 \times 4 \text{ mm}^2$ or $3.0 \times 4.6 \text{ mm}^2$
	No. of line scans	30–100
	Fluence	1.0 J cm^{-2}
	Repetition rate	10 Hz
	Scanning speed (=scan rate)	10, 100, 200, and $500 \mu\text{m s}^{-1}$
RR	He carrier gas flow rate	0.5 L min^{-1} (cell); 0.3 L min^{-1} (cup)
SS	Ar make-up flow rate	0.8 L min^{-1}
AT	ICP-MS	
	RF power (W)	1500 W
	Isotopes measured	^{34}S , ^{35}Cl , ^{44}Ca , ^{47}Ti , ^{63}Cu , ^{66}Zn , and ^{107}Ag
	Acquisition time/mass	0.05–0.10 s
	Sample period/point (dwell time)	0.5 s
	Point/mass	1
	Measurement mode	Time-resolved TRA
	Ar plasma gas flow rate (L min^{-1})	15 L min^{-1}
	Ar auxiliary gas flow rate (L min^{-1})	0.8 L min^{-1}
DT		

LA-ICP-MS analyses of biological samples. The middle body cross section of invertebrates, parts of the isolated fish intestine, and wheat root parts were analysed. For each organism and each time point, one unstained paraffin section from each exposure group (control, Ag_2S NPs, and AgNO_3) at each time point was analysed with LA-ICP-MS, using optimized operating conditions defined in the process of the optimization, *i.e.*, an investigated sample size of app. $2 \times 4 \text{ mm}^2$, a beam diameter of $65 \mu\text{m}$, and scanning speed of $100 \mu\text{m s}^{-1}$. Up to four animals per exposure group (treatment) were analysed with LA-ICP-MS resulting in a total of 75 analysed samples. At one time point (days of exposure for an individual animal or plant), we analysed only one sample. The remaining non-analysed samples were coded and stored for future analyses if needed.

Results and discussion

In the work presented here, the LA-ICP-MS scanning parameters were optimized on laboratory animals exposed to insoluble Ag_2S NPs, or AgNO_3 , and then tested in realistic aquatic and terrestrial mesocosm exposures on a range of biota with different anatomies and tissue matrices.

Optimization of scanning conditions on AgNO_3 exposed *P. scaber* in a single-species set-up

No mortality was observed in the control or any of the exposed groups in the single species tests with woodlice exposed to AgNO_3 . The optimal fast screening LA-ICP-MS conditions were a compromise between image quality, laser diameter, and scan speed. The “image quality” means that one can identify the

main anatomical structures. The main anatomical structure of the woodlice where ingested and accumulated material is expected, is the digestive system (Fig. S2†). By knowing this, we could optimise the measurement parameters on animals exposed to the high concentration of AgNO_3 (500 mg of Ag per kg of soil). In our study, apart from the element of interest, Ag, elements such as Cu and Ca were also measured to identify animal's internal morphology. The distribution of Ca outlines the animal's body and appendages. The distribution of Cu serves to identify digestive glands with metal-storing cells that accumulate different metals when they are present in bioavailable form in crustaceans' food.^{29,35}

As expected, with a constant laser beam diameter ($65 \mu\text{m}$ circle), the best lateral resolution was achieved using the slowest scan rate ($10 \mu\text{m s}^{-1}$) (Fig. 1A). However, the time of analysis for a typical sample size (approx 10 mm^2) took more than 4.4 h using $10 \mu\text{m s}^{-1}$ scan speed and $65 \mu\text{m}$ beam diameter ($\sim 25 \text{ min mm}^{-2}$). This is not feasible for studies with large numbers of samples where the presence/absence of an element of interest in organs is the first piece of information of interest. Therefore slow LA-ICP-MS scanning should be applied only in cases where more detailed information is needed, as slow LA-ICP-MS scanning remarkably increases the resolution.

When the scan rate was increased by 10 times (*i.e.*, $100 \mu\text{m s}^{-1}$), the time for measuring one sample was reduced significantly to approx. 35 minutes ($\sim 2.5 \text{ min mm}^{-2}$). As seen in Fig. 1B, the resolution is still sufficient, and distinct biological structures of interest could be clearly seen as well as metal-storing cells with accumulated Ag. In Fig. 1C and D, due to a further increase in the scan rate to $200 \mu\text{m s}^{-1}$ and $500 \mu\text{m s}^{-1}$, respectively, the images were not of sufficient quality to define





Fig. 1 Elemental maps of silver (Ag), copper (Cu), and calcium (Ca) in cross sections of woodlice exposed to AgNO_3 with a final concentration of 500 mg Ag per kg of soil. The Cu and Ca maps are used as a reference to identify known biological structures. Different operating conditions were tested: two different beam diameters, 65 μm (A–D) and 35 μm (E) at four different scanning speeds (10, 100, 200, and 500 $\mu\text{m s}^{-1}$) in five paraffine slices of the same animal. The scan direction in all experiments was horizontal and bidirectional, exchanging the directions from left to right and right to left, alternately. In terms of compromise between mapping time and image resolution, the parameters marked with a red square were chosen as optimal. Letter G on the black and white image of the cross section in column A indicates the gut and two parts of digestive glands are circled. The color bar represents low to high element counts per second (CPS) of the measured spot.

the Ag and Cu locations. However, the presence of Ag in the sample could still be confirmed despite the scan time being reasonably short, *i.e.*, 17 min and 7 min, respectively. These latter conditions were therefore deemed satisfactory for getting

information on the presence or absence of elements, but not for identifying the exact tissue location.

When using the fast-screening, $\sim 4 \times 10^3$ data points are produced to analyse a typical sample size (approx. $3 \times 5 \text{ mm}^2$).



This is 5–6 orders of magnitude lower than that in the studies by Yamashita *et al.* and Materapi *et al.*, which used the single particle (LA-SP-ICP-MS) technique for elemental analyses.^{14,23} These authors reported an accumulation of $\sim 2 \times 10^7$ data points for the analysis of a $0.5 \times 0.5 \text{ mm}^2$ sample area, corresponding to $\sim 1.2 \times 10^9$ data points if the same sample size as in our study was analysed.^{14,23} In our work, we have collected substantially fewer experimental data points, which is mainly due to the significantly longer dwell time of the fast-screening method (500 ms), but the resolution quality is still at a sufficient level to study the presence of elements of interest (Ca, Cu for outlining the major biological structures and Ag as an element of interest) in different organisms.

When the laser diameter was smaller ($35 \mu\text{m}$) and a medium scan rate ($200 \mu\text{m s}^{-1}$) was applied (Fig. 1E), the lateral resolution of images on Ag distribution in the sample (size $3.0 \times 4.6 \text{ mm}^2$) was comparable with the resolution seen in Fig. 1B and also the time of the analysis (33 min), corresponding to $\sim 2.4 \text{ min mm}^{-2}$, was comparable. However, the obtained resolution for Cu and Ca was slightly lower in the case of Fig. 1E. Therefore, the settings applied as seen in Fig. 1B ($65 \mu\text{m}$ laser beam diameter and $100 \mu\text{m s}^{-1}$ scan rate) were selected as optimal scanning parameters for a fast screening (35 minutes) and used for analyses of the Ag presence in all other selected organisms exposed in the two mesocosm experiments.

Testing the optimized scanning conditions in single-species tests with *P. scaber* exposed to an environmentally relevant concentration of Ag_2S NPs

After selecting the optimal LA-ICP-MS imaging parameters on *P. scaber* exposed to high concentrations of AgNO_3 (500 mg Ag per kg of soil) these parameters were used to test the sensitivity of the method on *P. scaber* exposed to environmentally relevant concentrations of Ag_2S NPs or AgNO_3 with a final concentration of 10 mg Ag per kg of soil (Fig. 2). From the maps presented in Fig. 2A, it is evident that some Ag was also present in the digestive glands of control animals, but not in the area of the gut. Our previous work has shown that control isopods have some background of Ag.⁴ In the case of exposure to AgNO_3 a strong Ag signal in the area of digestive glands indicates that a much higher amount of silver was present in the digestive system than in the gut indicating accumulation of Ag in digestive glands (Fig. 2B). As expected, Cu is present in the digestive glands (dark red colour). In the case of Ag_2S NP exposure, Ag was also detected in digestive glands and also in the gut confirming the consumption of Ag_2S NPs (Fig. 2C).

Application of the optimized LC-ICP-MS scanning conditions to organisms exposed to Ag_2S NPs and AgNO_3 in aquatic and terrestrial mesocosm experiments

The cross sections of whole earthworms (Fig. 3A), whole isopods (Fig. 3B), and upper parts of wheat roots (Fig. 3C) as well as aquatic organisms – longitudinal sections of blackworm (Fig. 3D), planarian cross section (Fig. 3E), and longitudinal sections of the fish intestine (Fig. 3F) exposed to Ag_2S NPs and AgNO_3 (exposure concentration 10 mg Ag per kg of soil and 10

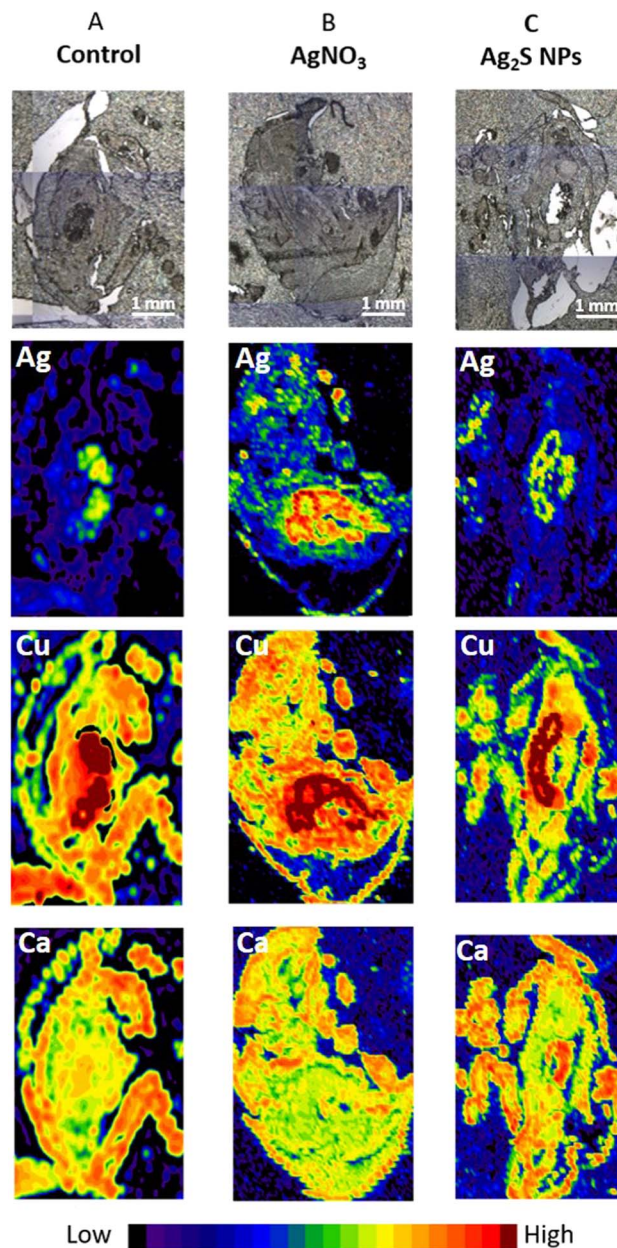


Fig. 2 Silver (Ag), copper (Cu), and calcium (Ca) elemental maps of cross sections of woodlice *Porcellio scaber* exposed to Ag_2S NPs (column C) and AgNO_3 (column B) with a final concentration of 10 mg Ag per kg of soil. One slice of the animal from the control group was also measured (column A). The previously selected optimal LA-ICP-MS measuring conditions were applied (beam size $65 \mu\text{m}$ and scan rate $100 \mu\text{m s}^{-1}$). The Ca and Cu maps are used as a reference to identify known biological structures, Cu especially to identify four digestive gland tubes (the darkest red colour on the Cu maps). The colour bar represents low to high element counts per second (CPS) of measured spots.

μg Ag per L of water) were inspected by LA-ICP-MS using parameters defined in a previous step and described in detail above (Fig. 3E). In Fig. 3, the measurements on the seventh day of exposure for all organisms are presented.

We provide evidence that anatomically very different biological samples could be prepared for LA-ICP-MS analyses





Fig. 3 Silver (Ag), copper (Cu), and calcium (Ca) elemental maps of cross sections of terrestrial organisms *Lumbricus rubellus* (A), *Porcellio scaber* (B), and the *Triticum aestivum* upper root part (C) exposed to 10 mg Ag per kg soil and aquatic organisms *Lumbriculus variegatus* (D), *Girardia tigrina* (E), and the *Oncorhynchus mykiss* intestine (F) exposed to 10 μg Ag per L of water in (F). The measurements on the 7th day of exposure for all organisms are presented. Selected optimal LA-ICP-MS measuring conditions were applied. The Cu and Ca maps are used as a reference to identify known biological structures. The white scale bar line on images of the tissue of all organisms in the first row represents 1000 μm . The color bar represents low to high element counts per second (CPS) of measured spots.

following the protocol for conventional histological sample preparation (*i.e.*, chemical fixation and paraffin embedding). It turned out that by using Carnoy B fixative, major/important biological structures in all samples were sufficiently preserved. A unified sample preparation approach is therefore very welcome when anatomically different organisms are needed to be processed in a relatively short time, such as specimen collections from complex large-scale experiments such as mesocosms or field sampling. In addition, conventional histological sample preparation allows differences in the localization of silver to be identified after exposure to Ag_2S NPs and AgNO_3 or in non-exposed organisms.

All inspected tissues from different organisms were prepared following the same protocol as described in the methods. On the stained sections, the major biological structures could be defined (Fig. S3†), demonstrating that the tissues were sufficiently preserved. In the case of small invertebrates, the whole-body cross section could be analysed, while for larger organisms, distinct organs or plant parts with the expected NP presence were isolated, fixed, and then inspected by LA-ICP-MS (*i.e.*, plant roots and fish intestines). The scanned size of the samples was approx. $2 \times 4 \text{ mm}^2$.

The aim of measuring other elements also in parallel to the element of interest with LA-ICP-MS was to contribute to the identification of major anatomical features in model tissues. The advantage of the method is that the time of measurement is not prolonged when multiple elements present in the sample are analysed (*i.e.* simultaneous elemental analysis). To present the results we have chosen two elements commonly present in all biological tissues, Ca and Cu.

The results show that fast-screening LA-ICP-MS settings (65 μm laser diameter and scan rate $100 \mu\text{m s}^{-1}$), which enable a short measuring time (35 min), are sufficient to obtain enough lateral resolution to locate the main biological structures and biodistribution of elements of interest on the elemental maps.

Elevated concentrations of Ag indicated by a strong signal (red color on images, see Fig. 3, second row) were present only in the specific areas recognized as the gut (confirmed on the histological images – Fig. S3†) of all invertebrate organisms (Fig. 3A, B, D and E), wheat root parts (Fig. 3C) and the lumen of the fish intestine (Fig. 3F). This is likely explained by the adsorption of Ag after Ag_2S NP exposure on the body surface, or as a result of ingested or precipitated materials present in the gut lumen. *P. scaber* did not accumulate Ag and *G. tigrina* showed little Ag accumulation upon Ag_2S NP exposures in the mesocosm tests, which can be the result of eventual excretion of Ag present in the gut lumen, with no Ag internalisation. Adsorption of NMs on the body surface has previously been documented, for example in daphnids.^{30,36} These results are in agreement with the literature reporting that passive translocation of NMs through invertebrate integuments is very unlikely if the barrier is intact.^{2,3,31,37} In the case of plants (*T. aestivum* L.), we confirmed the NP translocation from soil to roots and potentially to other body parts as reported by other authors.^{32,38} In the case of *O. mykiss*, the resolution assured by the fast screening could not be distinguished if Ag is present in the gut lumen or gut epithelium; here also slow LA-ICP-MS scanning of the same region should be provided to answer this question.



It is crucial to emphasize that Ag was measured also in all controls; therefore a background signal was present in all samples (Fig. S4†). When the model organisms were exposed to AgNO₃, elevated concentrations of Ag could also be seen in other body parts, not only the digestive system as in the case of exposure to NPs, as illustrated in Fig. 2 in the case of isopods and Fig. S5† in earthworms (represented as an example).

These findings were also confirmed in the study by Clark *et al.*³⁴ where total metal concentrations in the organs of the fish reflected the lower bioavailability of Ag₂S compared to AgNO₃. Clear differences in Ag biokinetics were also shown in the study by Talaber *et al.*⁴ where Ag bioaccumulation after exposure of two soil invertebrate species to pristine Ag NPs, ionic Ag and Ag₂S NPs (same as those used in our study) depended on the Ag form, soil type and test organism.

In conclusion, the present study confirmed that optimized LA-ICP-MS analyses provide Ag elemental distribution in invertebrate, vertebrate and plant samples exposed to AgNO₃ or Ag₂S NPs. By complementing this information with knowledge on the anatomy and physiology of organisms, the fate of materials/ions in the body or organ can be revealed. This confirms that LA-ICP-MS analyses can provide spatial information about the locations of material assimilation inside the body or on its surface. To our knowledge, this is the first time that the LA-ICP-MS method has been applied to an environmentally relevant mesocosm study and there are no field studies reported for NMs using the LA-ICP-MS method for detection in organisms. However, the present study paves the way for ecological fieldwork, first by demonstrating the method on an organism (isopod) under controlled laboratory conditions and then on a range of organisms in the mesocosm study. Crucially, despite complex environmental matrices such as soil, or sandy sediment in the aquatic mesocosm, these did not interfere with the measurements. Arguably, if organisms are carefully rinsed after collection to remove excess media and quickly preserved in fixative, then this method will have application on field-collected organisms as well. Clearly, fixed samples from mesocosms, or perhaps fieldwork, could be used for LA-ICP-MS to study the fate of metal-based (nano) materials in biota or samples could be stored for an indefinite time for the purpose of future investigations.

Author contributions

Conceptualization: G. M., S. N., D. D., and S. L.; methodology: G. M., S. N., S. L., D. D., M.-B.-F., and R. D. H.; investigation: G. M., S. N., V. K., I. T., S. L., Z. K., P. V.-S., M. B.-F., and R. D. H.; data curation: G. M., S. N., V. K., I. T., S. L., Z. K., P. V.-S., M. B.-F., R. D. H., and D. D.; writing—original draft preparation: G. M., S. N., D. D., and A. J.-K.; writing—review and editing: G. M., A. J.-K., S. N., S. L., Z. K., P. V.-S., R. D. H., and D. D.; funding acquisition: D. D.; EU H2020 project NanoFASE. All authors have read and agreed to the published version of the manuscript.

Conflicts of interest

There are no conflicts to declare.

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References

- 1 R. D. Handy and G. Al-Bairuty, *Ecotoxicology of Nanoparticles in Aquatic Systems*, 2019, pp. 156–168.
- 2 E. J. Petersen, M. Mortimer, R. M. Burgess, R. Handy, S. Hanna, K. T. Ho and P. Holden, *Environ. Sci.: Nano*, 2019, **6**(6), 1619–1656.
- 3 N. W. Van Den Brink, A. J. Kokalj, P. V. Silva, E. Lahive, K. Norrfors, M. Baccaro and C. A. Van Gestel, *Environ. Sci.: Nano*, 2019, **6**(7), 1985–2001.
- 4 I. Talaber, C. A. Van Gestel, A. J. Kokalj, G. Marolt, S. Novak, P. Zidar and D. Drobne, *Environ. Sci.: Nano*, 2020, **7**(9), 2735–2746.
- 5 Z. Qin, J. A. Caruso, B. Lai, A. Matusch and J. S. Becker, *Metallomics*, 2011, **3**(1), 28–37.
- 6 D. Drescher, C. Giesen, H. Traub, U. Panne, J. Kneipp and N. Jakubowski, *Anal. Chem.*, 2012, **84**(22), 9684–9688.
- 7 R. M. Galazzi, K. Chacón-Madrid, D. C. Freitas, L. F. da Costa and M. A. Arruda, *Rapid Commun. Mass Spectrom.*, 2020, **34**, e8726.
- 8 P. A. Doble, R. G. de Vega, D. J. Bishop, D. J. Hare and D. Clases, *Chem. Rev.*, 2021, **121**(19), 11769–11822.
- 9 D. Günther and B. Hattendorf, *Trends Anal. Chem.*, 2005, **24**(3), 255–265.
- 10 D. Clases and R. Gonzalez de Vega, *Anal. Bioanal. Chem.*, 2022, **414**, 7363–7386.
- 11 C. M. Ackerman, P. K. Weber, T. Xiao, B. Thai, T. J. Kuo, E. Zhang and C. J. Chang, *Metallomics*, 2018, **10**(3), 474–485.
- 12 J. S. Becker, M. Zoriy, J. S. Becker, J. Dobrowolska and A. Matusch, *J. Anal. At. Spectrom.*, 2007, **22**(7), 736–744.
- 13 D. Metarapi and J. T. van Elteren, *J. Anal. At. Spectrom.*, 2020, **35**(4), 784–793.
- 14 D. Metarapi, J. T. van Elteren, M. Šala, K. Vogel-Mikuš, I. Arčon, V. S. Šelih and S. B. Hočevár, *Environ. Sci.: Nano*, 2021, **8**(3), 647–656.
- 15 M. Bonta, S. Török, B. Hegedus, B. Döme and A. Limbeck, *Anal. Bioanal. Chem.*, 2017, **409**, 1805–1814.
- 16 M. Wang, L. N. Zheng, B. Wang, H. Q. Chen, Y. L. Zhao, Z. F. Chai, H. J. Reid, B. L. Sharp and W. Y. Feng, *Anal. Chem.*, 2014, **86**, 10252–10256.
- 17 O. Reifschneider, A. Vennemann, G. Buzanich, M. Radtke, U. Reinholz, H. Riesemeier, J. Hogeback, C. Koppen,



- M. Grossgarten, M. Sperling, M. Wiemann and U. Karst, *Chem. Res. Toxicol.*, 2020, **33**, 1250–1255.
- 18 Q. Li, Z. Wang, J. Mo, G. Zhang, Y. Chen and C. Huang, *Sci. Rep.*, 2017, **7**(1), 2965.
- 19 S. Zarco-Fernandez, A. M. Coto-Garcia, R. Munoz-Olivas, J. Sanz-Landaluze, S. Rainieri and C. Camara, *Chemosphere*, 2016, **148**, 328–335.
- 20 S. Böhme, H. J. Stärk, D. Kühnel and T. Reemtsma, *Anal. Bioanal. Chem.*, 2015, **407**, 5477–5485.
- 21 C. Scharlach, L. Müller, S. Wagner, Y. Kobayashi, H. Kratz, M. Ebert and E. Schellenberger, *J. Biomed. Nanotechnol.*, 2016, **12**(5), 1001–1010.
- 22 V. M. Neves, G. M. Heidrich, E. S. Rodrigues, M. S. P. Enders, E. I. Muller, F. T. Nicoloso, H. W. Pereira de Carvalho and V. L. Dressler, *Environ. Sci. Technol.*, 2019, **53**, 10827–10834.
- 23 S. Yamashita, Y. Yoshikuni, H. Obayashi, T. Suzuki, T. Green and T. Hirata, *Anal. Chem.*, 2019, **91**(7), 4544–4551.
- 24 S. Böhme, M. Baccaro, M. Schmidt, A. Potthoff, H. J. Stärk, T. Reemtsma and D. Kühnel, *Environ. Sci.: Nano*, 2017, **4**, 1005.
- 25 S. J. M. Van Malderen, B. Laforce, T. Van Acker, C. Nys, M. De Rijcke, R. de Rycke, M. De Bruyne, M. N. Boone, K. De Schampelaere, O. Borovinskaya, B. De Samber, L. Vincze and F. Vanhaecke, *Anal. Chem.*, 2017, **89**(7), 4161–4168.
- 26 J. S. Becker, *J. Mass Spectrom.*, 2013, **48**(2), 255–268.
- 27 W. X. Wang, *Crit. Rev. Environ. Sci. Technol.*, 2022, **52**(19), 3384–3414.
- 28 J. T. Van Elteren, V. S. Šelih and M. Šala, *J. Anal. At. Spectrom.*, 2019, **34**(9), 1919–1931.
- 29 J. He, D. Wang and D. Zhou, *Sci. Total Environ.*, 2019, **648**, 102–108.
- 30 S. Peixoto, Z. Khodaparast, G. Cornelis, E. Lahive, A. G. Etxabe, M. Baccaro and V. Puentes, *Ecotoxicol. Environ. Saf.*, 2020, **206**, 111405.
- 31 Z. Khodaparast, C. A. van Gestel, A. G. Papadimitis, S. F. Gonçalves, I. Lynch and S. Loureiro, *Sci. Total Environ.*, 2021, **777**, 146071.
- 32 P. S. Tourinho, C. A. van Gestel, A. J. Morgan, P. Kille, C. Svendsen, K. Jurkschat and S. Loureiro, *Ecotoxicology*, 2016, **25**, 267–278.
- 33 V. Lešer, D. Drobne, B. Vilhar, A. Kladnik, N. Žnidaršič and J. Štrus, *Zoology*, 2008, **111**(6), 419–432.
- 34 N. Clark, J. Vassallo, P. V. Silva, A. R. R. Silva, M. Baccaro, N. Medvešček and R. D. Handy, *Sci. Total Environ.*, 2022, **850**, 157912.
- 35 A. Bibic, D. Drobne and J. Strus, *Bull. Environ. Contam. Toxicol.*, 1997, **5**(58), 814–821.
- 36 S. Novak, A. J. Kokalj, M. Hočevár, M. Godec and D. Drobne, *Ecotoxicol. Environ. Saf.*, 2018, **152**, 61–66.
- 37 M. Van der Zande, A. J. Kokalj, D. Spurgeon, S. Loureiro, P. V. Silva, Z. Khodaparast, D. Drobne, N. J. Clark, N. W. Van den Brink, M. Baccaro, C. A. Van Gestel, H. Bouwmeester and R. D. Handy, *Environ. Sci.: Nano*, 2020, **7**(7), 1874–1898.
- 38 E. Kranjc, D. Mazej, M. Regvar, D. Drobne and M. Remškar, *Environ. Sci.: Nano*, 2018, **5**(2), 520–532.

