

Cite this: *Mater. Adv.*, 2023,
4, 2502

Determination of Pb²⁺ and Cd²⁺ ions in raw milk, honey and groundnut shell using TSAB/MWCNT†

Jayagopi Gayathri,^{id}*^a Sivakumar Sivalingam^{id}*^a and Sanglimuthu Sriman narayanan^b

Herein, we present the fabrication of an *N',N'',N''',N''''*-tetra(salicylidene)-amino benzene-modified multi-walled carbon nanotube electrode (TSAB/MWCNT) for the voltammetric determination of Pb²⁺ and Cd²⁺. FT-IR and H¹-NMR spectroscopic techniques were applied to characterize the TSAB ligand, which was synthesized according to the procedure described in the Experimental section. The surface morphology of the TSAB/MWCNT and MWCNT electrodes was characterized by SEM with EDAX techniques. Electrochemical studies were carried out using EIS and SWASV techniques. The modified TSAB/MWCNT and MWCNT electrodes were compared, and it was observed that the former exhibited significantly enhanced redox peak currents, which implies the increased the conductivity and surface area of the TSAB-coated MWCNT electrode. Under the optimized conditions, square wave anodic stripping voltammetry confirmed the linear range of 1–16 ppb and LOD of 0.13 ppb and 0.4 ppb for Pb²⁺ and Cd²⁺, respectively. The SWASV sensing of Pb²⁺ and Cd²⁺ in spiked raw milk, honey and groundnut shells were analyzed. The results revealed the recovery of 98% to 103% for Pb²⁺ and 96% to 103% for Cd²⁺. Also, the SWASV sample values were consistent with the results from the AAS method.

Received 3rd February 2023,
Accepted 4th May 2023

DOI: 10.1039/d3ma00062a

rsc.li/materials-advances

1. Introduction

Despite the impact of heavy metals such as lead (Pb²⁺), cadmium (Cd²⁺) and zinc (Zn²⁺) in daily life, they are present as environmental pollutants in ground water, industrial discharge, and rice even at a low dose. For instance, the accumulation of Pb²⁺ in the human body can affect neurobehavioral development in children, cause a series of diseases such as muscle paralysis, and contribute to an increase in blood pressure, anemia and kidney injury.¹ Cd²⁺ also affect the lungs and nervous system.² Consequently, the development of convenient, sensitive, selective and reliable methods for the determination of Pb²⁺ and Cd²⁺ is necessary to monitor the environment and protect human health. Currently, several analytical methods are employed to determine Pb²⁺ and Cd²⁺ including atomic adsorption spectroscopy (AAS),^{3–5} inductively coupled plasma-mass spectroscopy (ICP-MS),⁶ and inductively coupled plasma-atomic emission spectroscopy (ICP-AES).⁷ The above-mentioned methods have the capability to detect heavy metal ions even in a low detection range, but their application is limited due to their high cost.⁸ Alternatively, the electro-analytical

technique has been widely applied owing to its low cost and exposure range either in part per billion (ppb) or trillion (ppt) in a few cases.^{9–11} Presently, most researchers focus on detecting heavy metal ions by anodic stripping voltammetry because it is a rapid and insightful electro-analytical method, which is also used to analyze an effective pre-concentration and consequent measurements step.^{12–14} Nevertheless, graphite electrodes usually show limited sensitivity due to their low precise surface area, inadequate electro active centers, and low adsorption capability for ions.¹⁵ Therefore, to improve the adsorption behavior, a modified working electrode is employed to increase the conductivity, stability, sensitivity and specific surface area.¹⁶ For decades, mercury electrodes have been widely used as working electrodes in anodic stripping voltammetry (ASV) owing to their broad cathodic ranges, but they also form amalgams with some heavy metals.¹⁷ Therefore, recently, various modified electrodes have been used to detect harmful metals such as Cd²⁺, Pb²⁺, Hg²⁺, Zn²⁺, and Cu²⁺ present in different sources in the environment by measuring the ASV current based on conductive polymers such as poly(1,5-diaminonaphthalene)/multi-walled carbon nanotubes,¹⁸ BiNP/MWCNT/Nafion/GCE,¹⁹ poly(1,5-diaminonaphthalene)/graphene,²⁰ graphene/polyaniline/polystyrene,²¹ and poly(1,2-diaminoanthraquinone) (PDAAQ).²² Among them, most researchers focus on multi-walled carbon nanotubes (MWCNTs) owing to their good stability and wide range potential.^{23–25} Also, Schiff base materials are highly reactive towards metals, exhibiting outstanding chelating behaviors,

^a Department of Chemistry, Vel Tech Rangarajan Dr.Sagunthala R & D Institute of Science and Technology, Avadi, Chennai, 600062, India.

E-mail: drgayathrij@veltech.edu.in, drsivakumars@veltech.edu.in

^b University of Madras, Guindy Campus, Chennai, 600025, India

† Electronic supplementary information (ESI) available. See DOI: <https://doi.org/10.1039/d3ma00062a>



which are prepared *via* the chemical reaction between aldehyde and amine compounds. In some cases, these materials possess fluorescence properties with biological activity.²⁶ The *in situ* condensation of 3,3'-diaminobenzidine and salicylaldehyde produces the *N,N',N'',N'''*-tetrasalicylideneaminobenzene (TSAB) ligand, which is a symmetrical and chiral molecule.

Herein, we synthesized TSAB and prepared modified TSAB/MWCNT electrodes, which were characterized and their electrochemical behavior for the sensing of Pb²⁺ and Cd²⁺ was investigated. To the best of our knowledge, there are no reports in the literature on the use of TSAB/MWCNT electrodes for the determination of Pb²⁺ and Cd²⁺ to date. The TSAB ligand was characterized *via* Fourier transform infrared (FTIR) and proton nuclear magnetic resonance (¹H-NMR) spectroscopy. The TSAB/MWCNT and MWCNT modified electrodes were characterized by electrochemical impedance spectroscopy (EIS) and cyclic voltammetry (CV). Ultimately, the electrochemical implementation of the sensor was performed by square wave anodic stripping voltammetry (SWASV) using TSAB/MWCNT in acetate buffer solution (pH 5.5) for the analysis of Pb²⁺ and Cd²⁺ in raw milk, honey and groundnut shell samples. The SWASV of the samples results were validated using atomic absorption spectroscopy (AAS).

2. Materials and methods

Herein, the following experimental methods were employed.

2.1. Chemicals and apparatus

MWCNT (98% purity) and graphite rods (diam 3 mm) were purchased from Sigma Aldrich. 2-Amino benzyl alcohol (assay >97.5%), lead(II) acetate trihydrate (pure 99%), cadmium acetate dihydrate (extra pure AR, 99%), acetic acid (min. 99.9%), and sodium acetate anhydrous (extra pure AR 99%) were procured from SRL PVT. Ltd, India. All the listed chemicals were Analar Grade. Doubly distilled water was used for all experiments.

SWASV measurement was carried out using a CHI 660B Analyzer (CH Instruments, USA model). The TSAB/MWCNT modified working electrode was used in the conventional three-electrode system. Scanning electron microscopy (SEM) and energy dispersive X-ray spectroscopy (EDX) were performed using a HITACHI S3400N (JOEL JSM 6360, potential range of 0.3 to 30 kV) instrument. The ¹H-NMR and FT-IR study was carried out using a BRUKER and 650 FT-IR spectrometer (Agilent Technologies), respectively.

2.2. Procedures

2.2.1. Synthesis of *N,N',N'',N'''*-tetra(salicylidene)amino benzene (TSAB). TSAB was prepared using the reported procedure²⁷ with slight modification. Sodium methoxide was added to a stirring suspension of 1,2,4,5-tetraaminobenzene tetrahydrochloride (0.93 g, 3.3 mmol) in 50 mL dry methanol under nitrogen gas until the reaction was complete, resulting in a yellow solution. Subsequently, salicylaldehyde (1.61 g, 13.2 mmol) was added dropwise to the warm solution. The solution was refluxed overnight, the orange solid obtained was

filtered and washed several times with methanol and ether, and air dried.

2.2.2. Modification of TSAB/MWCNT electrode. A graphite surface electrode was used in an earlier report.²⁸ MWCNT was overlaid on the graphite surface electrode. The synthesized TSAB (0.1 mM) was dissolved in an acetonitrile solution. Then, an optimized volume of 10 μL of TSAB was added to the MWCNT electrode surface at 25 °C.

2.2.3. Preparation of supporting medium. Acetate buffer solution (ABS) was prepared by adding a suitable quantity of equimolar (0.1 M) acetic acid and sodium acetate solution in double distilled water. The mandatory pH of the ABS was modifying by adding 0.1 M NaOH and HCl solution.

2.2.4. Standard solutions prepared. 1 mM of Pb²⁺ and Cd²⁺ as stock solutions were prepared individually by dissolving them in pH 5.5 ABS. 10 mM solutions of Pb²⁺ and Cd²⁺ in pH 5.5 ABS were made as the stock solution by sequential dilution. Then 0.05 μL of Pb²⁺ and Cd²⁺ were spiked in a known amount of ABS solution.

2.2.5. Preparation of real groundnut shell, honey and raw milk samples. Raw milk and honey samples were collected from a vegetable market in Avadi (India). Groundnut shells (sample-C) were obtained from a local market in Poonamallee (India). Firstly, the groundnut shells (sample-C) were considerably cleaned with DD water, and subsequently dried at 60 °C in an oven, and then made into powder samples. To each sample, *i.e.*, Sample-A (honey), Sample-B (raw milk) and Sample-C (powdered groundnut shells), 0.1 mM of Pb²⁺ and Cd²⁺ were added using the prepared stock solutions for further experiment. Consequently, these samples were used for the analysis of Pb²⁺ and Cd²⁺ by SWASV and the resultant values were compared with AAS.

2.3. Stripping voltammetric measurement of toxic metal ions

A conventional three-electrode system was used for the voltammetric measurements, with the TSAB/MWCNT-modified electrode, platinum wire and saturated (1.0 M KCl) calomel electrode employed as the working electrode, counter electrode and reference electrode, respectively. The synthesized TSAB was characterized by FT-IR and ¹H-NMR spectroscopy. The conductivity and surface area of the TSAB/MWCNT and MWCNT electrodes were measured by EIS and CV. The Pb²⁺ and Cd²⁺ present in the raw milk, honey and groundnut shell solutions were detected by SWASV using the TSAB/MWCNT-modified electrode and determine at the ppb level.

3. Results and discussion

3.1. Characterization of synthesized TSAB

The FT-IR stretching frequencies of the synthesized TSAB are shown in Fig. S1 (ESI[†]). The absorption band observed at 1620 cm⁻¹ is attributed to the presence of the imine group. A broad range absorption peak was observed between 3385 cm⁻¹ to 3441 cm⁻¹ due to the stretching vibration mode of the -OH group present in water molecules. The stretching frequencies in the range of 1431 cm⁻¹ to 1475 cm⁻¹ and 743 cm⁻¹ to 820 cm⁻¹ are due to presence of an aromatic group. The C-OH stretching



mode observed at 1276 cm^{-1} shifted to a lower frequency as a result of the hydroxyl oxygen atoms in the coordination sites. The $^1\text{H-NMR}$ spectrum of TSAB is presented in Fig. S2 (ESI †). The chemical shift migrated towards the positive side due to the presence of aromatic protons in the range of δ 10.095 to δ 11.801. The OH proton was located at δ 6.553–6.974 ppm. The $^1\text{H-NMR}$ spectra confirmed the formation of TSAB, which were analyzed based on the earlier reports.²⁹

3.2. Investigation of TSAB/MWCNT-modified working electrode

3.2.1. Analysis of SEM with EDAX. The surface morphology of PIGE, MWCNT, TSAB/MWCNT and $(\text{M}^{2+}\text{-TSAB})/\text{MWCNT}$ was identified by SEM and EDAX analysis, as shown in Fig. 1 and Table 1. The SEM image of PIGE shows smooth and regular carbon microspheres (Fig. 1A). EDAX indicated the presence of the C element (Fig. 1E). As shown in Fig. 1B, the SEM image of MWCNT shows a cage-like structure and EDAX indicated that the intensity of C increased, as shown in Fig. 1F. TSAB/MWCNT exhibited a needle crystalline-shaped structure on the PIGE electrode surface, as shown in Fig. 1C. Also, in the EDAX spectrum, intense peaks for C and O were observed, as shown in Fig. 1G. $(\text{M}^{2+}\text{-TSAB})/\text{MWCNT}$ revealed a cloud-shaped

structure due to the adsorption of Pb^{2+} and Cd^{2+} on the electrode surface Fig. 1D and the EDAX in Fig. 1H show the intense peaks of Pb^{2+} , Cd^{2+} , O and C elements.

3.2.2. Electrochemical sensing behavior of PIGE, MWCNT and TSAB/MWCNT. The electrochemical behavior of the modified electrode surface was examined using cyclic voltammetry employing $[\text{Fe}(\text{CN})_6]^{-3/4-}$ as the redox probe. Fig. 2A shows the CV of the bare PIGE, modified MWCNT and TSAB/MWCNT electrodes in the presence of 1 mM $[\text{Fe}(\text{CN})_6]^{-3/4-}$ containing 0.1 mol L^{-1} ABS (pH 5.0) in the potential range of -0.30 V to 90 V at 50 mV s^{-1} . The bare PIGE and MWCNT-modified electrodes showed higher peak to peak potential separation values ($\Delta E_p = 228\text{ mV}$ and 175 mV , respectively) compared with the TSAB/MWCNT-modified electrode ($\Delta E_p = 134\text{ mV}$) (Table 2), where an increase in peak potential separation corresponds to lower electrical conductivity.

Moreover, modifying bare PIGE with MWCNT increased the peak current (I_{pa}) from $66\text{ }\mu\text{A}$ to $100\text{ }\mu\text{A}$. Furthermore, a sharp and very intense peak was observed at $168\text{ }\mu\text{A}$ for TSAB/MWCNT. The observed lower peak potential ΔE_p and enormous improvement in the I_{pa} values (peak current) of TSAB/MWCNT may be due to the combination of TSAB/MWCNT, which increased the surface area and electron transfer between $[\text{Fe}(\text{CN})_6]^{-3/4-}$ and TSAB/MWCNT, indicating that it is electrochemically more active.

In 0.1 M ABS solution, the CV peak current was observed to increase linearly in 1 mM $[\text{Fe}(\text{CN})_6]^{-3/4-}$ for TSAB/MWCNT with the square root of the scan rate ($v^{1/2}$) owing to the fact that the electron migration reaction at this modified (TSAB/MWCNT) electrode is a diffusion-controlled method. The active surface area of each electrode in an irreversible process can be examined based on the Randles–Sevcik equation [31]: $I_p = 2.69 \times 10^5 n^{3/2} A D^{1/2} \nu C$, where n is the number of electrons ($n = 2$), A is the active electrode surface, D is the diffusion coefficient ($7.6 \times 10^{-6}\text{ cm}^2\text{ S}^{-1}$), C is the concentration of $[\text{Fe}(\text{CN})_6]^{-3/4-}$ (1 mM) and ν is the scan rate (V/S). According to the slope ($i_p\text{-}v^{1/2}$), the value of A was measured to be 0.927 cm^2 , 0.338 cm^2 and 0.048 cm^2 for TSAB/MWCNT, MWCNT and PIGE, respectively (Table 2). The obtained results show that the electroactive surface of TSAB/MWCNT increased drastically, which was ~ 19 times higher than that of the untreated PIGE.

3.2.3. Electrochemical impedance spectroscopy of modified electrode. EIS was performed and the changes in the electron transfer resistance (R_{et}) owing on the surface of the fabricated electrode were recorded. The Nyquist plots of the PIGE, MWCNT and TSAB/MWCNT electrodes were measured in 0.1 M ABS solution containing 1 mM of $[\text{Fe}(\text{CN})_6]^{-3/4-}$, where the interfacial electron transfer resistance at the electrode surface is equal to the diameter of the semicircle, which were obtained by EIS.³⁰ The semicircle, which indicates the restrained electron transfer kinetics of the redox probe at the electrode surface,³¹ showed the recorded data obtained in the frequency range of 0.1 Hz to 1.0 MHz using the AC signal amplitude of 5 mV. The impedance figure was fitted using the Randles electrical equivalent circuit, as shown in Fig. 2B, where R_s , R_{et} , C_{dl} and Z_w are the resistance in solution, charge transfer resistance, double-layer capacitance



Fig. 1 Surface morphology. SEM image of PIGE (A), MWCNT (B), TSAB/MWCNT (C), and $[\text{Pb}^{2+}$ and Cd^{2+} -TSAB)/MWCNT (D). EDAX image of PIGE (E), MWCNT (F), TSAB/MWCNT (G) and $[\text{Pb}^{2+}$ and Cd^{2+} -TSAB)/MWCNT (H).



Table 1 SEM and EDX characterization of PIG electrode, MWCNT, TSAB/MWCNT, Cd²⁺ and Pb²⁺-TSAB/MWCNT

Characterization	PIG electrode	MWCNT	TSAB/MWCNT	Cd ²⁺ and Pb ²⁺ -TSAB/MWCNT
SEM	(A) Smooth and regular carbon microspheres	(B) Cage-shape structure	(C) Needle crystalline-shaped structure	(D) Cloud-shaped structure
EDAX	(E) Carbon peak	(F) Increased carbon peak	(G) Carbon and oxygen peaks	(H) Carbon, oxygen, cadmium and lead peaks

and Warburg impedance, respectively. The Cole-Cole plots (Nyquist plot) of PIGE, MWCNT and TSAB/MWCNT were obtained in the presence of 0.1 M ABS containing 1 mM [Fe(CN)₆]^{-3/4-}, as shown in Fig. 2B and Table 2. The Nyquist plot of ferri/ferrocyanide, which undergo redox reactions, shows the significant changes in responses for all three electrodes. A large semicircle diameter ($R_{et} = 4.1$ k Ω) with almost a straight line was observed for PIGE, which indicates the high electron resistance of the bare electrode to the redox, whereas for MWCNT and TSAB/MWCNT R_{et} was calculated to be 2.4 k Ω and 1.07 k Ω , respectively. The result obtained indicates that the bare PIGE, showed huge R_{et} values compared to MWCNT and TSAB/MWCNT. Interestingly, the R_{et} for the modified electrode decreased to about 2.9 k Ω compared to that of the bare PIG, indicating that the TSAB/MWCNT nanocomposite enhanced the electron migration. The reduction in the interfacial electron

Table 2 Electrochemical parameters of PIG electrode, MWCNT, and TSAB/MWCNT using CV and EIS

Characterization	Term	PIG electrode (a)	MWCNT (b)	TSAB/MWCNT (c)
CV	I_p	0.012 cm ²	0.060 cm ²	0.123 cm ²
	I_{pa} (μ A)	66 μ A	100 μ A	168 μ A
	ΔE_p	228 mV	175 mV	138 mV
EIS	R_s	52.36 Ω	38.16 Ω	23.4 Ω
	R_{ct}	7490 Ω	917.2 Ω	194.83 Ω
	θ	0	0.87	0.97

transfer resistance is owing to the good conductivity of the nanocomposite. The results showed that the observed current signal for electro-oxidation increased in the corresponding stripping voltammetric investigations.

3.3. Electrochemical sensing performance for Pb²⁺ and Cd²⁺ ions on MWCNT and TSAB/MWCNT

The electrochemical sensing behavior of Pb²⁺ and Cd²⁺ was detected using stripping voltammetry on the surface of the unmodified (MWCNT without ligand) and modified TSAB/MWCNT electrodes. To detect the sensitivity and selectivity of the unmodified and modified electrodes towards the electrochemical activities of both analytes, the stripping voltammogram of an 8 ppb mixture of Pb²⁺ and Cd²⁺ in pH 5.5 ABS (Fig. 3) was observed. Curve *b* shows the anodic peak current responses using the TSAB/MWCNT electrode in the potential range of -1.2 V to -0.2 V vs. KCl for both metals, which were distinct and very intense peaks, confirming that the TSAB ligand coated



Fig. 2 (A) Cyclic voltammetry and (B) electrochemical impedance spectra on PIGE (curve *a*), MWCNT (curve *b*) and TSAB/MWCNT (curve *c*) in 1 mM [Fe(CN)₆]^{-3/4-} containing 0.1 mol L⁻¹ ABS (pH-5.0) at 50 mV s⁻¹. Inset: Enlarged image of (B) depicting the EIS of TSAB/MWCNT-modified electrode.



Fig. 3 SWASV performed on MWCNT (A) and TSAB/MWCNT (B) with Pb²⁺ and Cd²⁺ (8 ppb) accumulated on electrode surface in 0.1 M acetate buffer solution (pH 5.5).



on MWCNT was responsible for the superior electroanalytical detection performances. In contrast, for both metal analytes, weak peaks were observed at the unmodified MWCNT electrode (curve A). The significant peak for the TSAB ligand, which possesses OH and CH=N groups, is due to its tendency to bind the metal ions present in the electrolyte. Thus, the high-intensity peak observed is due to the presence of -OH and -CH=N- groups in TSAB. The active sites present in the TSAB ligand have sensing performances for the metal (Pb^{2+} and Cd^{2+}) present in the electrolyte, even at a very low concentration

(ppb). The TSAB/MWCNT electrode was further investigated, as illustrated in Scheme 1.

3.4. Utilization of the new sensor

To apply the established sensor, the important parameters in the electrochemical detection of Pb^{2+} and Cd^{2+} , such as supporting medium, pH and preconcentration time were scrutinized for the SWASV analysis of Pb^{2+} and Cd^{2+} (Fig. 4). The TSAB/MWCNT sensor for 8 ppb of Pb^{2+} and Cd^{2+} , as shown in Fig. 4A, exhibited different stripping current peak intensities for various electrolytes including

I. Preconcentration step: (Metal complex)



II. Reduction step: ($\text{M}^{2+} \rightarrow \text{M}^0$)



III. Stripping Step: ($\text{M}^0 \rightarrow \text{M}^{2+}$)



Scheme 1 Mechanism of $[\text{Pb}^{2+}$ and $\text{Cd}^{2+}\text{-TSAB}]/\text{MWCNT}$ electrode.



NH_4NO_3 , KNO_3 , NaNO_3 and acetate buffer at 0.1 M concentration. The most intense peak was observed for acetate buffer (0.78 μA for Pb^{2+} and 0.26 μA Cd^{2+}), while the least intense peak was observed for NH_4NO_3 (0.2 μA for Pb^{2+} and 0.18 μA Cd^{2+}). Therefore, the most suitable electrolyte is acetate buffer among the electrolytes, which was employed in subsequent experiments. The effect of pH using TSAB/MWCNT in 0.1 M acetate medium on the oxidation current for Pb^{2+} and Cd^{2+} was detected by SWASV (Fig. 4B). The acetate buffer pH was constantly maintained in the range of 3.5 to 6.5. The maximum oxidation peak was observed for Pb^{2+} and Cd^{2+} at 5.5 pH, *i.e.*, when the pH increased from 3.5 to 5.5, the oxidation current density was the maximum, increasing the protonation on

the TSAB/MWCNT electrode surface. Further, with an increase in pH, the oxidation current peak density decreased gradually, which may be due to the increase in deprotonation on the TSAB/MWCNT-modified electrode surface. The decrease in the oxidation peak current density observed in the hydrolysis mechanism of Pb^{2+} and Cd^{2+} will resist their accumulation on the TSAB/MWCNT-modified electrode surface at higher pH (5.5–6.5). Therefore, a suitable oxidation current response for detecting the Pb^{2+} and Cd^{2+} ions was achieved in a supporting electrolyte of pH 5.5, which was used in subsequent experiments. The accumulation time was varied from 60 s to 360 s for the detection of Pb^{2+} and Cd^{2+} in 0.1 M acetate buffer at pH 5.5 (Fig. 4C). An increase in the accumulation time

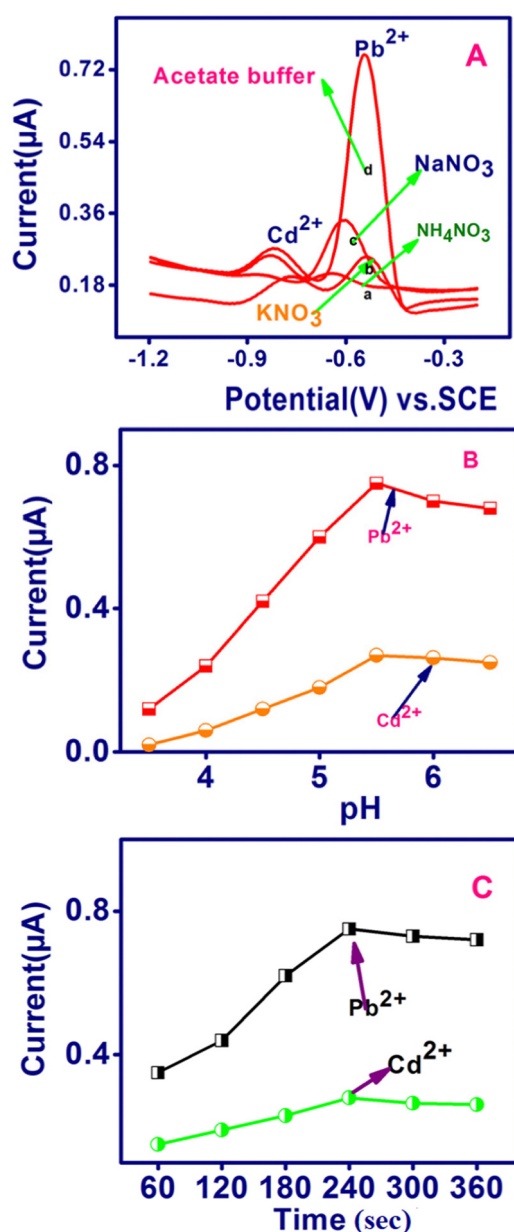


Fig. 4 Effect of (A) supporting electrolyte, (B) pH and (C) preconcentration time on the anodic stripping peak currents of Pb^{2+} and Cd^{2+} (8 ppb each) to evaluate TSAB/MWCNT sensor performance in 0.1 M ABS (pH 5.5).

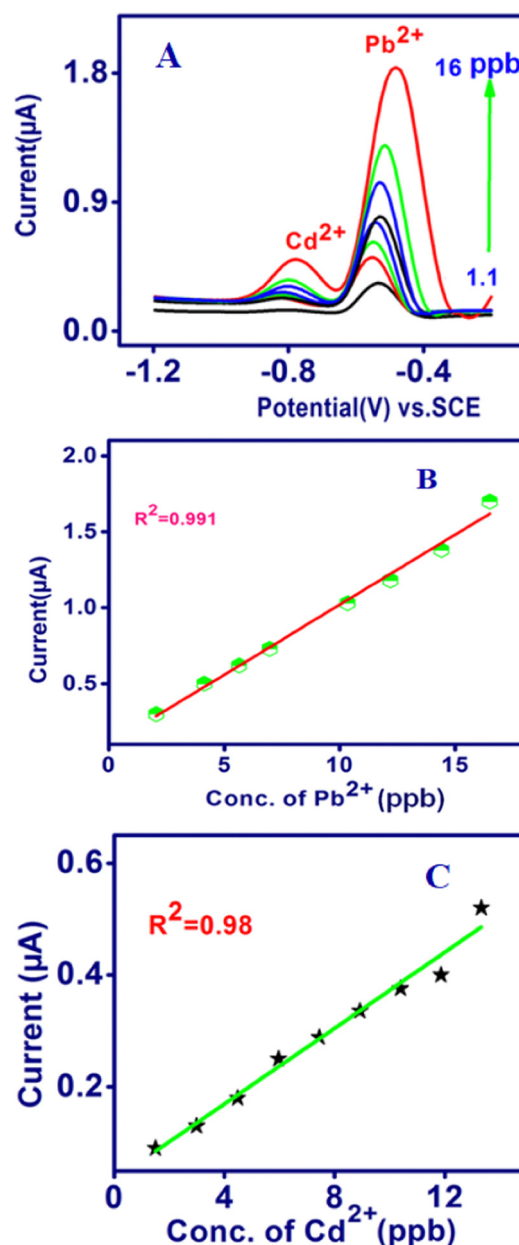
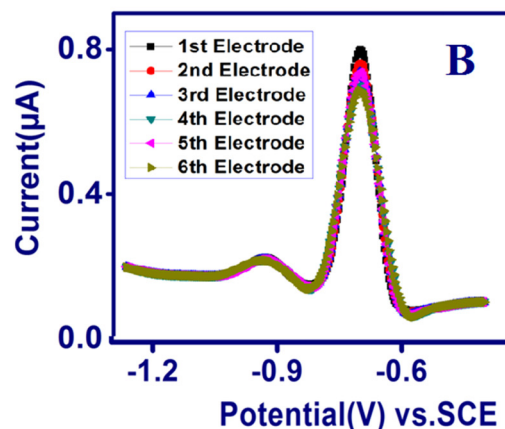
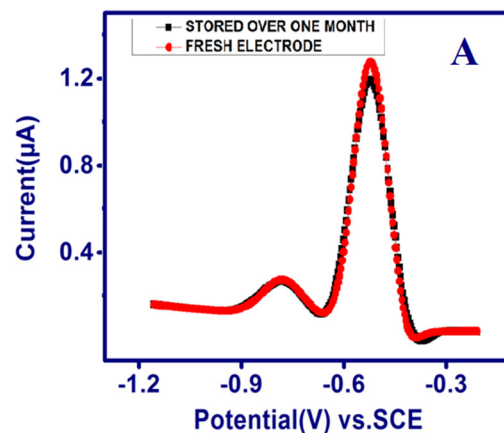


Fig. 5 Simultaneous determination of Pb^{2+} and Cd^{2+} by TSAB/MWCNT using (A) SWASV. Corresponding calibration plots for Pb^{2+} (B) and Cd^{2+} (C) calculated using SWASV.



Table 3 Comparison of TSAB/MWCNT for the detection of Pb²⁺ and Cd²⁺ with already reported works

Modified electrode	Linear range (ppb)		LOD (ppb)		Ref.
	Cd ²⁺	Pb ²⁺	Cd ²⁺	Pb ²⁺	
IONP-COOH/APTES-ITO	10–100	10–100	0.9	0.6	32
Fe ₃ O ₄ /F-MWCNT/GCE	56.2–3372	103.6–6216	5.62	16.58	33
BFO/CPEs	200–20 000	200–20 000	50	30	34
Fe ₃ O ₄ @G ₂ -PAD/MCPE	0.5–80	0.5–80	0.21	0.17	35
SWCNHs	1–60	1–60	0.2	0.4	36
NSRG/GCE	5–3000	5–3000	6.5	1.3	37
nZVI-BPC/GCE	2.0–50	2.0–50	0.19	0.2	38
MWCNT/Bi/PGE	1–80	5–80	0.43	0.27	39
(DTZ)@TiO ₂ /poly(TBA)-(AA)	10–1000	10–1000	10	20	40
polyPCA/GE	40–1000	40–1000	15.4	13.6	41
TSAB/MWCNT	1.1–16	1–16	0.18	0.13	This work

Fig. 6 SWASV curves of TSAB/MWCNT in pH 5.5 ABS containing 12 ppb Pb²⁺ and Cd²⁺ in the presence of 10 ppb Hg²⁺, Ni²⁺, Cu²⁺, Sn²⁺ and Zn²⁺.Fig. 7 Stability and reproducibility of TSAB/MWCNT. (A) Square wave anodic stripping voltammograms of 12 ppb of Pb²⁺ and Cd²⁺ using a freshly prepared electrode (red line) and one stored at room temperature over one month (black line) and (B) SWASV curves continuously recorded with six TSAB/MWCNT prepared under the same conditions in a solution containing 12 ppb of Pb²⁺ and Cd²⁺.Table 4 Investigated results for Pb²⁺ and Cd²⁺ in the presence of different interfering ions, including Hg²⁺, Ni²⁺, Cu²⁺, Sn²⁺ and Zn²⁺, in 0.1 M of ABS solutions spiked with 12 ppb Pb²⁺ and Cd²⁺

Interference ions	Pb ²⁺		Cd ²⁺	
	Peak current (µA)	Relative signal changes (%)	Peak current (µA)	Relative signal changes (%)
No interference ions	1.15	—	0.42	—
Hg ²⁺	1.10	−8.7	0.417	−8.0
Ni ²⁺	1.05	−4.7	0.415	−1.2
Cu ²⁺	1.20	4.3	0.412	−2.0
Sn ²⁺	1.09	−5.3	0.413	−1.7
Zn ²⁺	1.07	−2.8	0.414	−1.5

from 60 s to 240 s increased the oxidation peak current owing to the greater accumulation of Pb²⁺ and Cd²⁺ with a longer accumulation time. Subsequently, with a further increase in the accumulation time to 300 and 360 s, no improvement was observed, which is due to the saturation of the TSAB/MWCNT electrode surface area. Therefore, the maximum accumulation time for the detection of Pb²⁺ and Cd²⁺ using the TSAB/MWCNT-modified electrode is 240 s.

3.5. Analytical behavior of TSAB/MWCNT

In this assessment, the quantitative estimation of the concentrations of the target metal was accomplished using SWASV and analytical curves for the newly modified sensors. The simultaneous detection of Pb²⁺ and Cd²⁺ was performed using SWASV (Fig. 5A), and the calibration data are shown in Fig. 5B and C.



It was observed that under the optimum conditions, well-defined stripping current peaks and potential range were achieved for Pb^{2+} at -0.65 V and Cd^{2+} at -0.8 V. The stripping current peaks were proportional to the Pb^{2+} and Cd^{2+} concentration from 1 to 16 ppb and the linear equations are as follows:

$$\text{Pb}^{2+} \text{ Ip } (\mu\text{A}) = 0.926 \text{ C (ppb)} + 0.9; R^2 = 0.991;$$

$$\text{Cd}^{2+} \text{ Ip } (\mu\text{A}) = 0.415 \text{ C (ppb)} + 0.366; R^2 = 0.991;$$

The sensitivities were $0.926 \mu\text{A}$ per ppb for Pb^{2+} and $0.415 \mu\text{A}$ per ppb for Cd^{2+} , demonstrating that the simultaneous detection of Pb^{2+} and Cd^{2+} on TSAB/MWCNT is feasible. Furthermore, the lowest detection limit ($S/N = 3$) was calculated to be 0.13 ppb and 0.4 ppb for Pb^{2+} and Cd^{2+} , respectively, which is significantly lower than the recommended value for drinking water according to the WHO, *i.e.*, 3 and 10 ppb, respectively.³⁴ Subsequently, the evaluation of TSAB/MWCNT was compared with earlier reports on modified electrodes for the analysis of Pb^{2+} and Cd^{2+} , and data are presented in Table 3. It was found that the fabricated modified electrode exhibited similar or even better analytical performances for the analysis of Pb^{2+} and Cd^{2+} . However, our mercury-free electrodes are simple to prepare and environmentally friendly easily, which can be used for the construction of heavy metal ion electrochemical sensors.

3.6. Interference studies of metal ions

The selectivity of TSAB/MWCNT for the detection of Pb^{2+} and Cd^{2+} was examined by the addition 10 ppb interfering metal such as Hg^{2+} , Ni^{2+} , Cu^{2+} , Sn^{2+} and Zn^{2+} to 0.1 M pH 5.0 acetate buffer containing 12 ppb of Pb^{2+} and Cd^{2+} , as shown in Fig. 6. The stripping peak currents of Pb^{2+} and Cd^{2+} in the presence (I_i) and absence (I_0) of interfering Hg^{2+} , Ni^{2+} , Cu^{2+} , Zn^{2+} and Sn^{2+} and relative signal changes were employed to calculate the RSD using eqn (1), as follows:

$$\text{RSD} = \frac{I_i}{I_0} - 1 \quad (1)$$

The calculated RSD values are shown in Table 4, where the peak currents of Pb^{2+} in the presence of Hg^{2+} , Ni^{2+} , Sn^{2+} and Zn^{2+} interfering metals exhibited a slight decrease in RSD values by 2.8%, 4.6%, 5.3% and 8.7%, respectively. In the case of Cd^{2+} , the Hg^{2+} , Ni^{2+} , Cu^{2+} , Sn^{2+} and Zn^{2+} interfering metals resulted in a decrease of 8.0%, 1.2%, 2.0%, 1.7% and 1.5%, respectively. This can be attributed to the competition of the different metals with the Pb^{2+} and Cd^{2+} analytes and the interfering Hg^{2+} , Ni^{2+} , Cu^{2+} , Sn^{2+} and Zn^{2+} for the active sites present on the TSAB/MWCNT electrode surface.³⁴

3.7. Reproducibility and stability of the TSAB/MWCNT electrode

The reproducibility and stability of the developed TSAB/MWCNT electrode were measured using six modified TSAB/MWCNT electrodes in 0.1 M acetate buffer consisting of 12 ppb of Pb^{2+} and Cd^{2+} (Fig. 7). The six modified electrodes were employed for the analysis of Pb^{2+} and Cd^{2+} using SWASV, as shown in Fig. 7B. The acquired relative standard deviations

(RSDs) of the six electrodes were 2.1% and 3.62% for Pb^{2+} and Cd^{2+} detection, respectively, indicating the excellent reproducibility of the developed electrochemical sensor. The long-term stability of the TSAB/MWCNT electrode was monitored using its anodic current responses for metal ions on a newly prepared TSAB/MWCNT electrode and other electrode was kept at room temperature over a period of one month. As shown in Fig. 7A, the RSDs of the anodic peaks current for the toxic metal ions were determined to be 0.05% for Pb^{2+} and 0.03% for Cd^{2+} .



Fig. 8 SWASV of the TSAB/MWCNT in pH 5.5 ABS containing a 10 ppb mixture of Pb^{2+} and Cd^{2+} at an E_{acc} of -1.2 V and T_{acc} of 240 s in a few environmental samples: honey (A), raw milk (B) and groundnut shell (C).



Table 5 Comparison of the present method and AAS for the analysis of Cd²⁺ and Pb²⁺ (n = 3)

Sample	Metal ions	Square wave anodic stripping voltammetry (SWASV)				Atomic absorption spectroscopy (AAS)	
		Added ppb	Found (ppb)	RSD (%)	Recovery (%)	Found (ppb)	Recovery (%)
Sample-A (honey)	Cd ²⁺	8.0	7.7	1.1	96.0	8.0	100.0
		10	9.8	1.5	98.4	10.0	100.0
	Pb ²⁺	8.0	7.8	1.4	98.0	8.0	100.0
		10	10.9	2.0	99.0	10.0	100.0
Sample-B (raw milk)	Cd ²⁺	8.0	7.6	1.8	96.0	8.0	100.0
		10.0	9.7	1.9	97.0	10.1	101.0
	Pb ²⁺	8.0	8.0	1.5	100.0	8.1	101.0
		10.0	10.1	2.0	101.0	10.2	102.0
Sample-C (ground shell nuts)	Cd ²⁺	8.0	8.2	1.7	102	9.0	113.0
		10.0	10.3	2.3	103	11.4	114.0
	Pb ²⁺	8.0	8.2	2.0	102	9.2	115.0
		10.0	10.3	2.4	103	11.4	118.0

Thus, the SWASV data substantiate the adequate stability and reproducibility of the fabricated MWCNT/TSAB-modified electrode for sensing Pb²⁺ and Cd²⁺.

3.8. Sensing of Pb²⁺ and Cd²⁺ in raw milk, honey and groundnut shell samples

The TSAB/MWCNT sensor was applied for the detection of Pb²⁺ and Cd²⁺ in different real samples, *i.e.*, Sample-A (honey), Sample-B (raw milk) and Sample-C (2 g of groundnut shell), using SWASV. The precision of the quantification of Pb²⁺ and Cd²⁺ was proven by AAS (atomic absorption spectroscopy). The samples were diluted three times in 0.1 M acetate buffer at pH 5.5 for the analysis of 10 ppb Pb²⁺ and Cd²⁺. The results from the analysis of the metal ions in the honey samples (Fig. 8A), raw milk (Fig. 8B), and groundnut shells (Fig. 8C) using SWASV and the recovery values of the diluted samples are exhibited in Table 2. The samples showed a good recovery value of 98% to 103% Pb²⁺ and 96% to 103% Cd²⁺, as illustrated in Table 5. Thus, the constructed TSAB/MWCNT electrode has high precision (recovery) for the detection of Pb²⁺ and Cd²⁺ in environmental samples.

4. Conclusion

TSAB/MWCNT was effectively prepared and utilized as a sensor in ABS for the electrochemical analysis of Pb²⁺ and Cd²⁺. The successful synthesis of TSAB was confirmed by H¹-NMR and FT-IR spectroscopy and the TSAB ligand coated on the MWCNT electrode was characterized by scanning electron microscopy, cyclic voltammetry and electrochemical impedance spectroscopy. Applying the SWASV method, the TSAB/MWCNT electrode was used for sensing Pb²⁺ and Cd²⁺ in raw milk, honey and groundnut shell samples. The suitability of the developed SWASV method using TSAB/MWCNT for the electrochemical detection of Pb²⁺ and Cd²⁺ in raw milk, honey and groundnut shell samples was verified by its recovery value of 98% to 103% of Pb²⁺ and 96% to 103% of Cd²⁺. The samples were validated using atomic absorption spectroscopy and the recovery results were in the range of 100% to 118% of Pb²⁺ and 100% to 113% of Cd²⁺. Furthermore, the TSAB/MWCNT-modified electrode was

established to exhibit a low detection value, good stability and vast linear range. These results demonstrate that TSAB/MWCNT can be used for the detection and has superior potential for the analysis of Pb²⁺ and Cd²⁺ in real samples including raw milk, honey and groundnut shell samples.

Conflicts of interest

There is no conflict of interest.

Acknowledgements

The research work was supported with financial assistance through the DST-PURSE program by the Department of Science and Technology, New Delhi. One of the authors would like to acknowledge the Department of the National center for Nanosciences and Nanotechnology, the University of Madras for providing the facilities to conduct the characterization work.

References

- Z. Fang, J. Růžička and E. Hansen, *Anal. Chim. Acta*, 1984, **164**, 23–39.
- S. Bolisetty, M. Peydayesh and R. Mezzenga, *Chem. Soc. Rev.*, 2019, **48**, 463–487, DOI: [10.1039/C8CS00493E](https://doi.org/10.1039/C8CS00493E).
- K. Yin, Q. Wang, M. Lv and L. Chen, *Chem. Eng. J.*, 2019, **360**, 1553–1563.
- H. Uslu, C. Buyukpinar, T. Unutkan, H. Serbest, N. San and F. Turak, S. Bakirdere, *Microchem. J.*, 2018, **137**, 155–159.
- I. Mounteney, A. K. Burton, A. R. Farrant, M. J. Watts, S. J. Kemp and J. M. Cook, *J. Geochem. Explor.*, 2018, **184**, 1–10.
- J. Wang, E. H. Hansen and M. Miró, *Anal. Chim. Acta*, 2003, **499**, 139–147.
- H. Karami, M. F. Mousavi, Y. Yamini and M. Shamsipur, *Anal. Chim. Acta*, 2004, **509**, 89–94.
- N. G. Beck, R. P. Franks and K. W. Bruland, *Anal. Chim. Acta*, 2002, **455**, 11–22.
- A. Cobelo-García and R. Prego, *Anal. Chim. Acta*, 2004, **524**, 109–114.



- 10 A. Moutcine, C. Laghlimi, Y. Ziat, M. A. Smaini, S. E. E. Qouatli, M. Hammi and A. Chtaini, *Inorg. Chem. Commun.*, 2020, **116**, 1–20.
- 11 A. Moutcine, C. Laghlimi, O. Ifguis, M. A. Smaini, S. E. ElQouatli, M. Hammi and A. Chtaini, *Diamond Relat. Mater.*, 2020, **104**, 1–20.
- 12 M. Baghayeri, H. Alinezhad, M. Fayazi, M. Tarahomi, R. G. Motlagh and B. Maleki, *Electrochim. Acta*, 2019, **312**, 80–88.
- 13 C. B. Liu, X. Y. Chen, B. Y. Zong and S. Mao, *J. Mater. Chem. A*, 2019, **7**, 6616–6630.
- 14 Wei Xionga, Ping Zhanga, Shantang Liua, Yaokang Lv and Da Zhange, *Diamond Relat. Mater.*, 2021, **111**, 108170.
- 15 J. Barton, M. B. G. García, D. H. Santos, P. F. Bolado, A. Ribotti, M. McCaul and D. Diamond, *Microchim. Acta*, 2016, **183**, 503–517.
- 16 R. Y. Jiang, N. T. Liu, S. S. Gao, X. Mamat, Y. H. Su, T. Wagberg, Y. T. Li, X. Hu and G. Z. Hu, *Sensors*, 2018, **18**, 1567–1579.
- 17 L. S. Rocha, J. P. Pinheiro and H. M. Carapuça, *J. Electroanal. Chem.*, 2007, **610**, 37–45.
- 18 H. D. Vu, L.-H. Nguyen, T. D. Nguyen, H. B. Nguyen and D. L. Tran, *Ionics*, 2014, **21**, 571–578.
- 19 S. Palisoc, R. I. M. Vitto and M. Natividad, *Sci. Rep.*, 2018, **9**, 18491.
- 20 T. D. Nguyen, T. T. H. Dang, L. H. Nguyen, D. L. Tran, B. Piro and M. C. Pham, *Electroanalysis*, 2016, **28**, 1907–1913.
- 21 N. Promphet, P. Rattananarat, R. Rangkupan, O. Chailapakul and N. Rodthong, *Sens. Actuators, B*, 2015, **207**, 526–534.
- 22 K. M. Hassan, S. E. Gaber, M. F. Altahan and M. A. Azzem, *Electroanalysis*, 2018, **30**, 1155–1162.
- 23 Y. Tu, Y. Lin, W. Yantasee and Z. Reng, *Electroanalysis*, 2005, **17**, 79–84.
- 24 J. Bankim and A. Srivastava, *Electrochim. Acta*, 2011, **56**, 4188–4196.
- 25 A. M. Golikand, M. Asgari, M. G. Maraghah and E. Lohrasbi, *J. Appl. Electrochem.*, 2009, **39**, 65–70.
- 26 P. Jayaseelan, E. Akila, M. Usha Rani and R. Rajavel, *J. Saudi Chem. Soc.*, 2016, **20**, 625–634.
- 27 S. A. Guzzi and H. S. El Alagi, Pelagia Research Library, *Chem. Sin.*, 2013, **4**, 62–66.
- 28 F. Scholz and B. Lange, *Trends Anal. Chem.*, 1992, **11**, 359–367.
- 29 L. Mishra, K. Bindu and S. Bhattacharya, *Indian J. Chem.*, 2004, **43**, 315–319.
- 30 M. Ibrahim, H. Ibrahim, N. B. Almandil, M. A. Sayed, A.-N. Kawde and Y. Aldaqdouq, *Electroanalysis*, 2020, **32**, 1–11.
- 31 Hossieny Ibrahim and Yassien Temerk, *Talanta*, 2020, **208**, 120362.
- 32 Noorhashimah Mohamad Nor, Sarasijah Arivalakan, Nor Dyana Zakaria, Nithiyaa Nilamani, Zainovia Lockman and Khairunisak Abdul Razak, *ACS Omega*, 2022, **7**, 3823–3833.
- 33 W. Wu, M. Jia, Z. Zhang, X. Chen, Q. Zhang, W. Zhang, P. Li and L. Chen, *Ecotoxicol. Environ. Saf.*, 2019, **175**, 243–250, DOI: [10.1016/j.ecoenv.2019.03.037](https://doi.org/10.1016/j.ecoenv.2019.03.037).
- 34 Y. Beyene, Z. Bitew and F. Fekade, *Mater. Adv.*, 2022, **3**, 5882–5892.
- 35 B. Maleki, M. Baghayeri, M. Ghanei-Motlagh, F. Mohammadi Zonoz, A. Amiri, F. Hajizadeh, A. Hosseinifar and E. Esmailnezhad, *Measurement*, 2019, **140**, 81–88.
- 36 Y. Yao, H. Wu and J. Ping, *Food Chem.*, 2018, **18**, 31521.
- 37 X. Tan, Z. Li, X. Wang, M. Xu, M. Yanga and J. Zhao, *New J. Chem.*, 2022, **46**, 10618–10627.
- 38 M. Amine Djebbi, L. Allagui, M. Slim El Ayachi, S. Boubakri, N. Jaffrezic-Renault, P. Namour and A. Ben Haj Amara, *Zero-Valent, ACS Appl. Nano Mater.*, 2022, **5**, 546–558.
- 39 G. Sirin Ustabasi, I. Yilmaz, M. Ozcan and E. Cetinkaya, *Electroanalysis*, 2022, **34**, 1237–1244.
- 40 R. Sedghi, B. Heidari, H. Javadi and N. Sayyari, *Environ. Nanotechnol., Monit. Manage.*, 2022, **18**, 100670.
- 41 T. Machado Lima, P. Izabela Soares, L. Aguiar do Nascimento, D. Leoni Franco, A. Cesar Pereira and L. Franco Ferreira, *Microchem. J.*, 2021, **168**, 106406.

