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Low-temperature water-assisted crystallization approach to MOF@TiO₂ core–shell nanostructures for efficient dye removal†

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Metal–organic frameworks (MOFs) have shown excellent adsorption and degradation performance towards different kinds of dyes but suffer problems in the desorption process, where adsorbates cannot be completely removed and may cause energy waste or secondary pollution by the repeated washing with eluent, as well as their relatively weak stability in aqueous solutions. Thus, it is very attractive to integrate MOFs with active ingredients to prepare MOF-based composites with dye adsorption and degradation ability. The reported fabrication method of MOF@TiO₂ usually involves high crystallization temperature or contains acid or base additives, which are unfriendly to MOF structures. Here, we report the design of MOF@TiO₂ core–shell nanostructures *via* a surfactant-free amorphous TiO₂ coating followed by water-assisted crystallization at a low temperature (100 °C) below the decomposition temperature of most MOFs. Owing to the mild crystallization conditions, MOF cores are maintained and TiO₂ shells crystallize to an active anatase phase. The obtained MOF@TiO₂ core–shell nanostructures showed excellent dye removal efficiency and stability for anionic dyes, which can remove 98% dye after five adsorption and photodegradation cycles due to the synergistic effect of MOF cores and TiO₂ shells. The design concept and synthetic strategy *via* the water-assisted crystallization method for labile MOF materials encapsulated in functional TiO₂ shells may be used to construct other multifunctional MOF core–shell nanostructures, broadening their applications in diverse fields.

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Introduction

Metal–organic frameworks have shown attractive potential applications in diverse fields, such as adsorption and separation,^{1,2} heterocatalysis,^{3,4} sensors,^{5,6} drug delivery,⁷ and so on.^{8,9} In particular, much attention has been focused on MOFs applied in adsorption, including gas and liquid adsorption for their large surface area and designable pore size.^{10–12} As for liquid adsorption, several MOFs were reported to exhibit excellent performance in the removal of hazardous materials

in wastewater, including methylene blue, mercury, hexavalent chromium Cr(VI) or Cr₂O₇²⁻ and so on.^{13–18} Despite their excellent adsorption performance towards different kinds of dyes, the reuse of MOF adsorbents usually involves the desorption process *via* repeated washing with organic eluent (*e.g.*, methanol, ethanol, acetone and dimethylformamide),^{19–21} or the addition of inorganic salt (*e.g.*, NaNO₃ and NaCl),^{20,22} alkaline or acid,²³ which may cause energy waste and secondary pollution, as well as damage to alkali or acid-sensitive MOFs. Moreover, the removal of organic dyes in sewage *via* physical adsorption and subsequent washing desorption can only transfer hazardous materials from one solvent to another, which is not a green and cost-effective way. Hence, integrating MOF adsorbents with active components that are capable of degrading dye molecules would be very attractive.

TiO₂ is known as one of the most promising heterogeneous photocatalysts for dye photodegradation, which can lead to the complete mineralization of organic carbon into CO₂.^{24–26} While it generally exhibits low surface area, only sensitive to ultraviolet light, rapid electron–hole recombination and relatively poor charge transport property, which result in ultralow quantum efficiencies, hindering its further industrial application.^{27–29} Thus, integrating MOFs and TiO₂ to prepare

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bifunctional composites is expected to combine the advantages of MOFs' high adsorption capacity and TiO_2 's photodegradation activity. Compared to other composite structures, such as TiO_2 in MOF pores and TiO_2 @MOF core-shell structures, MOF@ TiO_2 core-shell nanostructures show advantages. On the one hand, MOF pores will not be blocked by TiO_2 particles, which is possible for the composite structures of TiO_2 in MOF pores.^{30,31} On the other hand, unlike the TiO_2 @MOF core-shell structure, active Ti sites on TiO_2 surfaces will not be covered and light adsorption is not shielded by MOFs for the MOF@ TiO_2 core-shell structure.³² Moreover, the stabilities of MOFs were improved due to the protective TiO_2 shells.³³ It should be noted that the TiO_2 shells need to be porous and permeable to ensure the accessibility of MOF pores.

Generally, to obtain high photocatalytic degradation efficiency, amorphous TiO_2 needs to be crystallized to anatase.³⁴ The conventional method of crystallization is to calcine amorphous TiO_2 at a high temperature (400–600 °C),³⁴ which is fatal to thermally unstable MOF structures. Some researchers prepared MOF@ TiO_2 core-shell composites *via* the post-assembly of prefabricated TiO_2 nanocrystals on the surface of MOF particles.^{35–37} Such a method led to non-uniform TiO_2 shells consisting of aggregated TiO_2 nanocrystals. Moreover, the connection between TiO_2 and the MOF surface was not tight, which made it likely to shed in recycle runs. Several studies prepared MOF@ TiO_2 *via* amorphous TiO_2 coating followed by crystallization, which showed the tight interface of MOF- TiO_2 . The crystallization temperatures in these studies were still above 150 °C, which was unfriendly to the MOF structures and the synthesis conditions were demanding.^{38–41} For example, Xiong *et al.* encapsulated MOFs in TiO_2 shells *via* hydrolysis with the addition of HF followed by crystallization at 180 °C; this method was only applicable for HKUST-1 fabricated from Cu_2O .⁴¹ Given all this, the development of a gentle method to precisely fabricate uniform MOF@ TiO_2 core-shell structures is still a great challenge.

Herein, bifunctional MOF@ TiO_2 core-shell nanostructures were designed and synthesized *via* amorphous TiO_2 coating followed by water-assisted crystallization under mild conditions. For proof of concept, MIL-101 with mesoporous pores was chosen as the core. Specifically, MIL-101 nanocrystals were firstly encapsulated in amorphous TiO_2 shells *via* the classic Stöber method. To crystallize the amorphous TiO_2 shells without the destruction of MOF structures, a water-assisted crystallization method was adopted at low temperature (100 °C) in water rather than calcination at high temperature. As a result, MOF@anatase TiO_2 (MOF@ TiO_2 -A) core-shell nanostructures were successfully obtained and exhibited high adsorption capacity, photodegradation activity, as well as cycle stability.

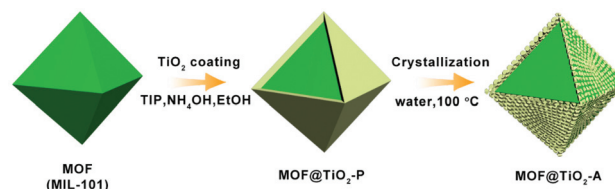
Results and discussion

A core-shell structured MOF@ TiO_2 -A composite was fabricated *via* coating MIL-101 nanocrystals with amorphous TiO_2 shells,

followed by a low-temperature water-assisted crystallization treatment. As shown in Scheme 1, MIL-101(Cr) octahedral nanocrystals were readily synthesized according to a procedure reported in literature.⁴² Amorphous TiO_2 -coated MIL-101 precursors (MOF@ TiO_2 -P) were then obtained by the Stöber method *via* the hydrolysis/condensation of tetraisopropyl titanate (TIP) in the presence of a certain amount of aqueous ammonia. In order to form homogeneous TiO_2 shells, the coating process was carried out in anhydrous ethanol solution with a small amount of ammonia to reduce the hydrolysis rate of TIP. Different from the literature-reported TiO_2 coating processes, in which some surfactants (*e.g.*, hexadecylamine) are usually required to promote the formation of TiO_2 shells,⁴³ here, no surfactant is introduced, and amorphous TiO_2 could be uniformly coated on the surface of MIL-101 crystals due to electrostatic and hydrogen bonding interactions between the positively charged MOF surface and the -OH groups of negatively-charged titanium oligomers.⁴⁴ The subsequent water-assisted crystallization method is in light of the previous literature on achieving the phase transformation of amorphous anatase TiO_2 (*e.g.* nanotubes) to anatase or rutile TiO_2 under hydrothermal conditions.^{34,45–49} Specifically, the precursors of MOF@ TiO_2 -P immersed in water are hydrothermally treated at 100 °C for 24 h. Then, the amorphous TiO_2 shells are expected to be crystallized *in situ* *via* the rearrangement of TiO_6 octahedra to generate a crystalline TiO_2 (anatase)-coated MOF composite with a desirable core-shell heterostructure.

The morphology evolution from MOF to MOF@ TiO_2 -P and finally to MOF@ TiO_2 -A core-shell nanostructures was revealed by scanning electron microscopy (SEM) and transmission electron microscopy (TEM). Fig. 1a–c demonstrate the octahedral configuration of MIL-101 nanoparticles with smooth surfaces. Fig. 1d–f confirmed the successful coating of MIL-101 octahedra with uniform TiO_2 shells (around 10 nm thickness). After the water-assisted crystallization process, smooth amorphous TiO_2 shells were transformed into rough shells composed of ultrafine anatase nanocrystals, as shown in Fig. 1g–i. Notably, the lattice fringes of MIL-101 can be observed clearly in the image of the MOF@ TiO_2 -A core-shell composite (Fig. 1j), illustrating that the MIL-101 crystalline phase is well preserved during the water-assisted crystallization process.

The lattice fringe of 3.1 nm, shown in Fig. 2a, corresponds to the (220) plane of the face-centered-cubic structured



Scheme 1 Schematic representation of the preparation of MOF@ TiO_2 -A core-shell nanostructures. Step 1: coating the MOF nanocrystal with an amorphous TiO_2 shell to obtain MOF@ TiO_2 -P at room temperature. Step 2: water-assisted crystallization of MOF@ TiO_2 -P at 100 °C to obtain the well-crystallized MOF@ TiO_2 -A.

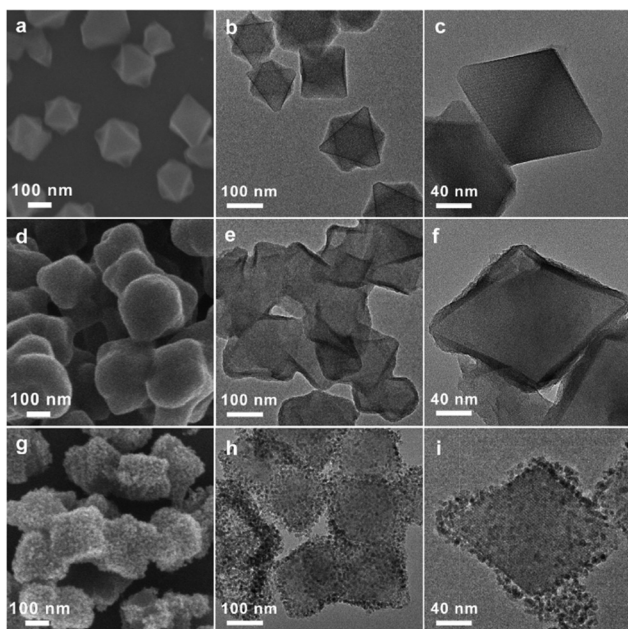


Fig. 1 Morphological changes from MOF to MOF@TiO₂-P and finally to MOF@TiO₂-A core-shell nanostructures. SEM and TEM images of MOF (a–c), MOF@TiO₂-P (d–f), and MOF@TiO₂-A (g–i).

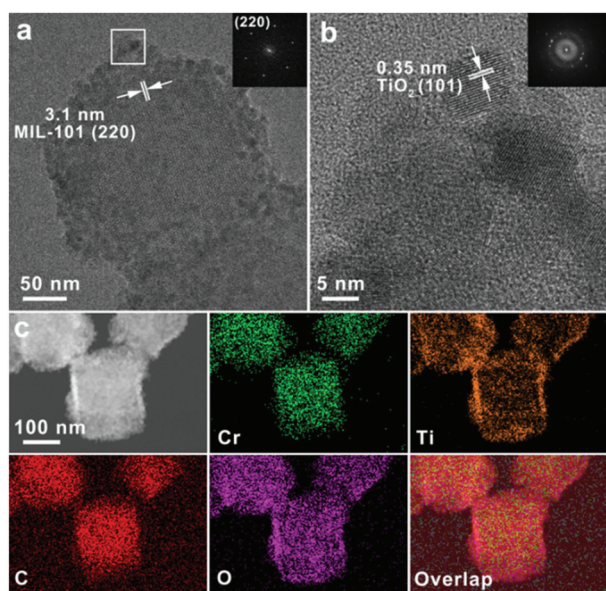


Fig. 2 TEM images of a single MOF@TiO₂-A core-shell nanoparticle (a) and enlarged high-resolution TEM image of the TiO₂ shell corresponding to the rectangular area in a (b). High-magnification HAADF-STEM image of MOF@TiO₂-A (c) and the corresponding EDX elemental mapping results. Insets in a and b correspond to the fast Fourier transform (FFT) images of MOF and TiO₂ crystals.

MIL-101, implying that the crystalline structure of MOF is maintained well after hydrothermal treatment. The fast Fourier transform (FFT) pattern inset further indicates the well-retained crystalline MIL-101 phase. As shown in

Fig. 2b, the enlarged HRTEM image of TiO₂ shell shows the lattice fringe of 0.35 nm, which corresponds to the (101) planes of anatase TiO₂. High-angle annular dark-field scanning TEM (HAADF-STEM) images (Fig. 2c) with energy-dispersive X-ray spectroscopy (EDX) element mappings further confirmed the core-shell nanostructures of MOF@TiO₂-A. It is worth noting that TiO₂ shells and MOF cores are not separate but are connected, implying the formation of an interface structure between the two components. By enlarging the region of the interface structure, as shown in Fig. S1,† an interlaced region rather than a clear boundary was observed, revealing the formation of the MOF-TiO₂ interface phase. Because of the unstable MOF structure under high-energy electrons as well as the large interplanar spacing gap, we could not get an HRTEM image clearly showing the lattice fringes of both MOF and TiO₂ concurrently.

Powder X-ray diffraction (PXRD) patterns of MOF, MOF@TiO₂-P and MOF@TiO₂-A are presented in Fig. 3a. The pattern of MOF@TiO₂-A displays the characteristic diffraction peaks at 25.2°, 38.0°, 47.9° and 54.1° corresponding to (101), (004), (200) and (105) of anatase TiO₂, respectively.⁵⁰ This result confirms the successful crystallization of amorphous TiO₂ shells to anatase *via* the water-assisted crystallization method.⁵⁰ Meanwhile, the retained diffraction peaks of MIL-101 illustrate that the crystalline phase of MOFs is well preserved during TiO₂ coating and water-assisted crystallization process (100 °C, 24 h). It should be mentioned that the reference sample of TiO₂ (named TiO₂-A), prepared *via* the same method as for MOF@TiO₂-A in the absence of MIL-101, presented a characteristic PXRD pattern of the anatase phase (Fig. S2†), further confirming that the phase transformation of amorphous TiO₂ to anatase can indeed occur under such mild treatment conditions. This result is different from the previous literature reports on the transformation of amorphous TiO₂ to

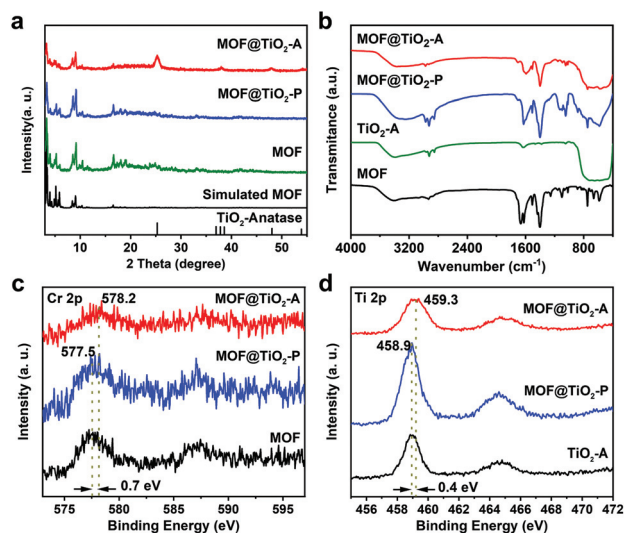


Fig. 3 PXRD patterns (a), FTIR spectra (b), Cr 2p XPS spectra (c) and Ti 2p XPS spectra (d) of the as prepared samples.

the anatase phase in hydrothermal/solvothermal systems, which commonly involve the addition of some organic solvents/templates or mineralizers like HF and relatively high treatment temperatures (*e.g.*, 180 °C).^{38,39,41,51} Hence, the present work demonstrates that using a pure water system (pre-removal of the residual ammonia and ethanol solvent from the coating system) is more beneficial for the subsequent crystallization of the amorphous TiO₂ shells at lower temperatures (100 °C). Fig. 3b shows that the FTIR spectra of MOF@TiO₂-P and MOF@TiO₂-A displayed two strong vibrational bands at 1410 and 1510 cm⁻¹ belonging to the –(O–C–O)– groups within the MIL-101 framework, which are consistent with the characteristic signals of MOFs.^{52,53} The wide band in the region of 450–800 cm⁻¹ could be attributed to TiO₂, confirming the formation of hybrid composites containing MOFs and TiO₂.⁵⁴

The X-ray photoelectron spectroscopy (XPS) survey spectra of MOF@TiO₂-A demonstrated the existence of C, Ti, O and Cr in this hybrid composite (Fig. S3†). As shown in Fig. 3c, the Cr 2p XPS signal of MOF@TiO₂-A had a 0.7 eV shift to a higher binding energy (from 577.5 eV to 578.2 eV) in comparison with that of MOF and MOF@TiO₂-P. Meanwhile, the Ti 2p XPS signal shifted from 458.9 eV to 459.3 eV (Fig. 3d). These results illustrate that a relatively strong interaction could be built between the anatase TiO₂ shells and the MIL-101 cores, possibly related to the formation of an interfacial phase.³⁵ The thermogravimetric (TG) curves of various samples are shown in Fig. S4.† The weight loss below 200 °C is attributed to the desorption of water molecules and –OH in MIL-101 pores.⁵² The steep weight loss of MOF and MOF@TiO₂-A at around 400 °C is due to the structural decomposition of MIL-101. When heated from room temperature to 800 °C in air, the total weight losses of MOF, TiO₂-A and MOF@TiO₂-A were 78.4%, 6.5% and 23%, respectively. Mass percentages of MOF and TiO₂ components in MOF@TiO₂-A were calculated to be 23 wt% and 77 wt% according to the weight loss, consistent with the inductively coupled plasma atomic emission spectra (ICP-AES) results (26 wt% and 74 wt%).

Fig. 4 shows the nitrogen adsorption–desorption isotherms and the pore size distribution curves of MOF@TiO₂-A, TiO₂-A and MOF. The bare MIL-101 presents type I adsorption isotherms, while the noticeable increase at higher pressure ($P/P_0 = 0.9$) is ascribed to the macropores generated from particle

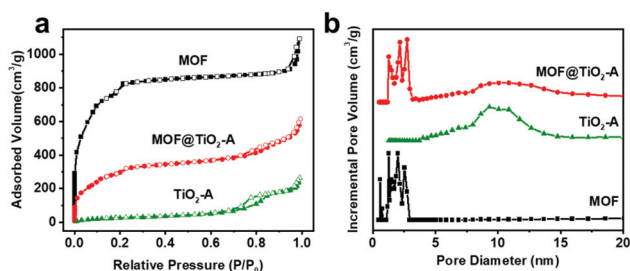


Fig. 4 Nitrogen adsorption–desorption isotherms (a) and pore size distribution curves (b) of MOF, TiO₂-A and MOF@TiO₂-A.

aggregation. N₂ adsorption–desorption isotherms and pore size distribution curves of TiO₂-A and MOF@TiO₂-A indicate the presence of piled pores, which originated from the phase transformation (amorphous TiO₂ to anatase phase) during the process of water-assisted crystallization. This can be further confirmed by the fact that MOF@TiO₂-P precursors show reversible type I adsorption–desorption isotherms (Fig. S5†), which are quite similar to the bare MOF. Combined with the above TEM results, it can be concluded that many piled pores are generated in the TiO₂ shells during the crystallization process. As shown in Table 1, the bare MOF of MIL-101 has a very high BET surface area of 2823 m² g⁻¹ and a large pore volume of 1.69 cm³ g⁻¹. MOF@TiO₂-A exhibited a BET surface area of 1116 m² g⁻¹ and pore volume of 0.95 cm³ g⁻¹, lower as compared to MOF but much higher than TiO₂-A. The relatively high surface area and pore volume of the hybrid composites suggest that the inner pores inherited from MIL-101 are not blocked by external TiO₂ shells, which should be a nice feature for efficient utilization in adsorption and catalysis.

The UV–vis diffuse reflectance spectra (DRS) of MOF, TiO₂-A and MOF@TiO₂-A are shown in Fig. S6a.† TiO₂-A presents strong ultraviolet light absorption with the main peak at 300 nm. MIL-101 presents both ultraviolet and visible light absorption with a narrow peak at 270 nm and two broad peaks at around 440 nm and 590 nm. As expected, MOF@TiO₂-A shows combined adsorption signals, including both the strong ultraviolet absorption and the two broad visible light absorptions. The photoluminescence (PL) spectra of various samples were also investigated for comparing their photogenerated charge carrier separation performances. As shown in Fig. S6b,† both TiO₂-A and the bare MOF exhibited strong emission bands of fluorescence, indicating the serious recombination of the photogenerated electron–hole pairs which can suppress the photocatalytic activity. Interestingly, MOF@TiO₂-A showed much lower intensity of fluorescence emission as compared to the two separated components of MOF and TiO₂-A. The reduced recombination loss might be due to the formation of MOF–TiO₂ heterojunctions at the core–shell interfaces as already revealed in previously reported literature.³⁵

Zeta potential was also measured to study the surface charge of the hybrid composites. As shown in Fig. S7,† MOF@TiO₂-A showed a positive surface charge (+17.5 mV) between that of bare MOF (+35.1 mV) and TiO₂-A (+8.9 mV). The moderate positive surface charge of MOF@TiO₂-A implies the presence of medium electrostatic interactions with anionic adsorbates, which is in favour of the desorption and degra-

Table 1 Comparison of BET surface areas and pore characteristics of MOF, TiO₂ and MOF@TiO₂-A

Sample	S_{BET} (m ² g ⁻¹)	Pore size (nm)	V_{total} (cm ³ g ⁻¹)
MOF	2823	1.1–2.9	1.69
TiO ₂	102	4.0–16.5	0.41
MOF@TiO ₂ -A	1116	1.2–15.5	0.95

dation of adsorbates. This particular surface charge characteristic together with TiO₂ shells may render MOF@TiO₂-A an appropriate material for anionic dye adsorption and degradation.

Despite of their excellent adsorption performances towards different kinds of dyes, previous studies have demonstrated that MOFs commonly suffer from serious problems in completely removing adsorbates, which may also cause energy waste or secondary pollution by repeated washing with eluents.^{19,20,22,23} Here, the integrated MOF@TiO₂-A composite is expected to combine the advantages of the MOFs' high adsorption capacity and TiO₂'s photodegradation activity.

As a typical azo organic dye, Congo red (CR) dye is one of the most common organic contaminants in industrial effluents. As shown in Fig. S8,† the width, depth and thickness of the CR anion are 2.62, 0.74 and 0.43 nm, respectively, which can be adsorbed by MIL-101 owing to the existence of large mesoporous cages with an internal diameter of 3.4 nm and window aperture of 1.4 by 1.6 nm.^{52,55} For checking this point, the removal of CR in wastewater was used as a model reaction to evaluate the adsorption and photodegradation performance of the hybrid composite of MOF@TiO₂-A.

Initial experimental results indicated that the adsorption-desorption equilibrium of CR over MOF@TiO₂-A could be achieved within 50 minutes as shown in Fig. S9.† Hence, the mixture of the hybrid composite and CR were firstly stirred in the dark for 1 h to achieve the adsorption-desorption equilibrium before the lights were turned on. As shown in Fig. 5a, MOF@TiO₂-A had much higher adsorption capability (36%) as compared to P25 (3%) and TiO₂-A (5%), mainly due to its high BET surface area and larger pore volume; it was lower than that of bare MIL-101 (46%) because of its decreased BET surface area as a result of the introduction of TiO₂. After light irradiation, the amount of CR removed over MOF@TiO₂-A reached 98% in 2 h without the addition of any oxides (e.g., H₂O₂), much higher than that of TiO₂-A and P25 (48% and 37%, respectively). Although it shows high adsorption capacity, bare MOF can only remove 86% CR dye in 2 h as a consequence of the poor photodegradation capacity of MIL-101. To investigate the photocatalytic performance of MOF@TiO₂-A, the photodegradation capacities of the obtained samples were calculated by normalization of the remaining CR concen-

tration after the adsorption process in the dark. As shown in Fig. S10a,† MOF@TiO₂-A exhibited the fastest degradation rate compared with P25, TiO₂-A and bare MOF. Fig. S10b† reveals that the photocatalytic degradation follows first-order kinetics, which can be fitted by a linear correlation of $\ln(C_0/C) = kt$, where k is the apparent first-order rate constant.⁵⁶ The calculated rate constant of MOF@TiO₂-A was significantly higher than those of P25, TiO₂-A and bare MOF. These results demonstrate the excellent adsorption capacity and photodegradation efficiency of MOF@TiO₂-A.

MOF@TiO₂-A also showed high photostability, with high dye removal efficiency after 5 recycles (up to 98%) (Fig. 5b). Additional characterization results demonstrated that the core-shell structure of MOF@TiO₂-A remained well after the photodegradation cycles. As shown in Fig. S11,† the PXRD pattern of the used MOF@TiO₂-A is consistent with the fresh sample. In contrast, the framework of bare MOF collapsed after 5 cycles. The N₂ adsorption-desorption measurements revealed only a very slight decrease in the specific surface area (0.5%) and the pore volume (8.4%) of the used MOF@TiO₂-A in comparison with the fresh one (Fig. S12 and Table S1†). For bare MOF, the specific surface area and pore volume decreased to 40.4% and 39.6%, respectively, which may be responsible for the poor photocatalytic stability of pure MIL-101. Based on the above characterization results, it was proposed that the excellent adsorption/photodegradation performance of MOF@TiO₂-A could be mainly attributed to the specific core-shell structure. The relatively high BET surface area and large pore volume, which were inherited from the parent MOF, should contribute to the outstanding adsorption efficiency for CR. The uniform anatase TiO₂ shells with piled pores could provide permeable channels for CR molecule adsorption in the MOF cores, and could also provide effective photoactive sites (e.g., the interface structure between MOF and TiO₂) that can readily decompose the adsorbed CR molecules. In addition, the CR photodegradation mechanism over MOF@TiO₂-A could be hypothesized according to the related literature.^{18,57} Under UV irradiation, photoexcited electron-hole pairs were generated within the conduction band (CB) and valence band (VB) of TiO₂ shells, respectively. The excited electrons oxidize dioxygen to produce superoxide radicals ($\cdot\text{O}_2^-$) in the CB, and positive holes reduce water molecules to produce hydroxyl radicals ($\cdot\text{OH}$) in the VB. The $\cdot\text{O}_2^-$ and $\cdot\text{OH}$ with strong oxidizing ability then oxidize the adjacent CR adsorbed on MOF cores to the mineralized products.

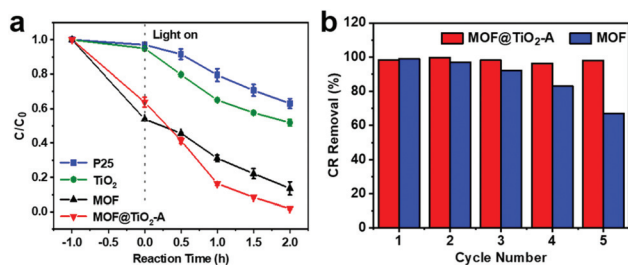


Fig. 5 Kinetic traces for CR adsorption and photodegradation in the presence of P25, TiO₂, MOF and MOF@TiO₂-A (a), and photodegradation stability of recycled MOF and MOF@TiO₂-A (b).

Conclusions

In summary, a multifunctional anatase TiO₂-coated MOF composite was obtained *via* a facile two-step self-assembly approach. The transformation of amorphous TiO₂ into the anatase phase could be achieved under a very mild water-assisted crystallization process, which could lead to the formation of uniform anatase TiO₂ shells with permeable piled channels. The resultant MOF@TiO₂-A core-shell composite

shows excellent photodegradation efficiency and cycle stability for anionic dye CR. The existence of an interface structure between the MOF and anatase TiO₂ may facilitate the separation of photogenerated electron–hole pairs and initiate the degradation of adjacent CR molecules adsorbed by internal MOF cores, thus playing an efficient synergistic role in the adsorption and photodegradation of dye molecules. Furthermore, anatase TiO₂ shells significantly improve the photocatalytic stability of MOF nanocrystals. This simple water-assisted crystallization strategy may provide great opportunities for constructing various photoactive TiO₂-coated MOFs composites with high stabilities under mild conditions, and thus may considerably broaden their application in diverse fields.

Experimental methods

Chemicals and reagents

All the chemicals were used without further purification: Chromium(III) chloride hexahydrate (CrCl₃·6H₂O, Alfa Aesar Reagent Company, 98 wt%), 1,4-benzenedicarboxylate (H₂BDC, TCI, 98 wt%), tetraisopropyl titanate (TIP, Aladdin, 98%), ammonia solution (NH₃·H₂O, Tianjin Yongsheng Fine Chemical Company, 25 wt%), Congo red (CR, Tianjin Guangfu Fine Chemical, 98 wt%), ethanol (EtOH, Beijing Chemical Works, GR). Deionized water was obtained from a Millipore Milli-Q plus system.

Synthesis of MIL-101

In a typical procedure, 266 mg CrCl₃·6H₂O (1.0 mmol) and 166 mg H₂BDC (1.0 mmol) were dispersed in 10 mL deionized water. After vigorous stirring at room temperature for 3 min, the mixed solution was transferred into a Teflon-lined stainless-steel autoclave and heated at 200 °C for 24 h. On completing the reaction and slowly cooling to room temperature, recrystallized H₂BDC was formed in the green solution. The solution was centrifuged at 1000 rpm for 3 min to remove H₂BDC. Subsequently, the supernatant green solution containing MIL-101(Cr) was collected by centrifugation at 5000 rpm for 10 min and washed twice with DMF and ethanol. The obtained products were dried in an oven at 80 °C overnight for further use.

Synthesis of MOF@TiO₂-P

MOF@TiO₂-P core–shell nanoparticles were prepared by the Stöber method. Specifically, 80 mg of as-prepared MIL-101(Cr) was homogeneously dispersed in 10 mL ethanol by ultrasonication, followed by the addition of 0.2 mL of ammonia. The mixture was stirred at room temperature for 5 min to form a uniform dispersion. Then, 0.2 mL of TIP was added to the dispersion dropwise under stirring. After 10 min of reaction, the product MOF@TiO₂-P core–shell nanoparticles were collected by centrifugation and then washed several times with ethanol.

Synthesis of MOF@TiO₂-A

MOF@TiO₂-A was obtained *via* the solvothermal treatment of the as-prepared MOF@aTiO₂ nanoparticles at a relatively low temperature. Specifically, 100 mg MOF@TiO₂-P was homogeneously dispersed in 30 mL deionized water. The dispersion was transferred into a 50 mL Teflon-lined stainless-steel autoclave and heated at 100 °C for 24 h. The obtained product was harvested by centrifuging and washing several times with ethanol, and dried at 80 °C for further use.

Photocatalytic measurements

The photocatalytic performance of the as-prepared samples was investigated by the photodegradation of CR. The CR solution with a concentration of 300 ppm was prepared by dissolving the dye in distilled water. For reaction, 5 mg catalyst was added to 40 mL of CR aqueous solution in a 150 mL beaker with a water jacket to keep the temperature of the beaker at 25 °C. The solution was kept in the dark for 1 h under magnetic stirring to reach the adsorption/desorption equilibrium. Afterwards, the reaction vessel was irradiated under simulated solar light (300 W xenon lamp) and the distance between the reaction solution and the light source was 15 cm. During the reaction, 1 mL of sample was withdrawn every 0.5 h. The remaining concentration of CR was detected by UV-visible spectrophotometer at λ_{\max} .

Characterization

The crystal sizes and morphologies were determined by scanning electron microscopy (SEM) *via* a JSM-7800F (JEOL) electron microscope, and by transmission electron microscopy (TEM) at 200 kV using a Tecnai F20 electron microscope. The crystallinity and phase purity of the samples were characterized by powder X-ray diffraction (XRD) on a Rigaku D-Max 2550 diffractometer using Cu K α radiation ($\lambda = 1.5418 \text{ \AA}$). Chemical composition was determined *via* inductively coupled plasma (ICP) analyses carried out on a PerkinElmer Optima 3300 DV ICP instrument. Nitrogen adsorption/desorption measurements were carried out on a Micromeritics 2020 analyzer at 77.35 K after the samples were degassed at 150 °C under vacuum. The pore size distributions were derived from the adsorption branches using the nonlocal density functional theory (NLDFT) approach. X-ray photoelectron spectroscopy (XPS) studies were conducted using a Thermo ESCALAB 250 spectrometer with monochromatized Al K α excitation. Infrared (IR) spectra of the samples dispersed in KBr pellets were measured on a PerkinElmer Spectrum 430 FT-IR spectrometer. Thermogravimetric (TG) analysis was carried out on a TGA Q500 analyzer in air, with a heating rate of 10 °C min⁻¹. The photoluminescence (PL) spectra of the samples were acquired using a spectrofluorometer (HORIBA Scientific FluoroMax-4). The Zeta potential was measured with a Zetasizer Nano-ZS90 (Malvern Instruments) at 25 °C. A Shimadzu UV-2450 spectrophotometer and a Hitachi U-4100 instrument were used for UV-vis adsorption spectroscopy.

Author contributions

M. J. and X. S. conceived the research project. S. B. and M. L. designed and carried out the experiments. Experiments and data analysis were performed by S. B., M. L., P. S., Z. L., X. S. and M. J. The manuscript was drafted by S. B. and revised by X. S. and M. J.

Conflicts of interest

There are no conflicts to declare.

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