



**Environmental
Science**
Processes & Impacts

**Photolysis of the Herbicide Dicamba in Aqueous Solutions
and on Corn (Zea Maize) Epicuticular Waxes**

Journal:	<i>Environmental Science: Processes & Impacts</i>
Manuscript ID	EM-ART-02-2021-000058.R1
Article Type:	Paper

SCHOLARONE™
Manuscripts

1
2
3 1 **Photolysis of the Herbicide Dicamba in Aqueous Solutions and on Corn (*Zea***
4
5 2 ***Maize*) Epicuticular Waxes**
6
7 3

8
9 4 Kaitlyn Gruber

10
11 5 Brittany Courteau⁺

12
13 6 Maheemah Bokhoree

14
15 7 Elijah McMahon

16
17 8 Jenna Kotz

18
19 9 Amanda Nienow*

20
21 10
22
23
24
25 11
26
27 12 Gustavus Adolphus College

28
29 13 800 W College Avenue

30
31 14 St. Peter, Minnesota 56082, United States

32
33
34 15
35 16
36
37 17 January 21, 2021

38
39 18 Submitted to Environmental Science: Processes and Impact

40
41 19
42 20
43 21
44 22
45 23
46 24
47 25
48 26 ⁺ Present Address: Department of Chemistry, Iowa State University, 1605 Gilman Hall, Ames,
49 27 IA 50011

50
51 28 *Author to whom correspondence should be addressed: Telephone: (507) 933-7327, fax: (507)
52 29 933-7041, Email: anienow@gustavus.edu.
53 30
54
55
56
57
58
59
60

31 Abstract

32 Dicamba, 3,6-Dichloro-2-methoxybenzoic acid, has been used in agriculture as an herbicide for
33 over fifty years, and has seen an increase in use in the past decade due to the development of
34 glyphosate resistant weeds and soybeans genetically modified to resist dicamba. Despite the
35 previous use of dicamba, many questions remain regarding its environmental fate, especially the
36 new commercial formulations used on genetically modified crops. Here, the photolysis of
37 dicamba, including the commercial formulation Diablo[®], is examined in aqueous solutions of
38 varying water quality and on the surface of corn epicuticular waxes. Dicamba is stable to
39 hydrolysis but degrades under UV light. The photolytic half-life for dicamba photolysis in
40 aqueous solutions at pH 7 irradiated with Rayonet UVB lamps (280-340 nm) was $t_{1/2} = 43.3$ min
41 (0.72 hours), in aqueous solutions at pH 7 in a Q-Sun solar simulator ($\lambda > 300$ nm) was $t_{1/2} =$
42 13.4 hours, and on epicuticular waxes irradiated in the Q-sun solar simulator was $t_{1/2} = 105$
43 hours. Experiments with adjuvants, compounds added into the commercial formulations of
44 dicamba, led to increases in rate constants for both aqueous and wax experiments. In addition to
45 kinetic rate constants, photoproducts were tentatively assigned for the aqueous solution
46 experiments. This work deepens the knowledge of the environmental fate of dicamba including
47 the role surfactants play in chemical reactions and in providing new applications of current
48 methods to examine the photolysis of chemicals sorbed to surfaces.

50 Environmental Significance Statement

51 In 2016, Monsanto announced Roundup Ready 2 soybeans, genetically modified soybeans that
52 can tolerate dicamba and glyphosate. With the increased use of dicamba on soybeans, a crop to
53 which it has not been previously applied, it has become apparent that herbicide drifting causes
54 crop damage in adjacent fields; thus, understanding the environmental fate of dicamba is
55 essential for maintaining ecological health. This project provides details on the environmental
56 fate of dicamba by examining its photochemical reactivity when sorbed to epicuticular waxes, in
57 the presence of adjuvants, and in commercial formulation. The methods used in this work can be
58 extended to examine the photolysis of other pesticides and new pesticide formulations.

59 INTRODUCTION

60 Agrochemicals are widely used around the world. In 2012, 1,182 million pounds of pesticides
61 were used in agriculture in the United States; 678 million pounds of that amount was herbicides
62 dedicated to the eradication of weeds.¹ For over 50 years, dicamba, 3,6-Dichloro-2-
63 methoxybenzoic acid, has been one of the agrochemicals used in the United States, historically
64 mostly on corn or grain crops. With a pKa of 1.87 or 1.94,² dicamba is a weak acid that causes
65 death in most plants by targeting the plants' vascular tissue. Past formulations of dicamba have
66 been listed as a restricted-use herbicide due to high potential to volatilize, leach from soils,
67 persist in groundwater, and to cause widespread contamination of ecosystems.³⁻¹⁰ The use of
68 dicamba decreased upon the rise of glyphosate in the early 2000s, but the recent development of
69 glyphosate resistant weeds means dicamba use is increasing again (as seen by USDA data
70 analyzed on the Pesticide Use Data System of Hygeia Analytics).^{11,12} In 2016, Monsanto
71 announced Roundup Ready 2 soybeans, genetically modified soybeans branded under the name
72 Xtend, that can tolerate both dicamba and glyphosate.¹³ In April 2016, the EPA allowed the use
73 of dicamba in sprays for these soybeans for five years.¹⁴ In July 2016, the Xtend soybeans gained
74 EU import approval.¹⁵ Despite conditional approval and restrictions on use, growing season 2017
75 saw the first use of Xtend soybeans in the US - and by the end of the summer, there were
76 thousands of complaints from farmers with fields adjacent to those using Xtend soybeans. It has
77 since been shown in the scientific literature that dicamba drifting is a probable cause for the crop
78 damage in adjacent fields.¹⁶ After lawsuits by several environmental groups, a federal circuit
79 court revoked the EPA approval to use dicamba in June 2020; a week later, the EPA banned the
80 sale of specific dicamba products but said farmers can spray any dicamba products already in
81 their possession.¹⁷ 2020 already marked the end of federal approval of dicamba products, and
82 manufacturers would already have been submitting new applications for approval for use of the
83 products in 2021.¹⁸ In addition, these rulings only impact the newest formulations of dicamba
84 product. Although the future of dicamba use in agriculture is uncertain, it is imperative to
85 examine the chemistry of the molecule to better understand the environmental fate of the
86 product.

87
88 Biodegradation, photodegradation, and chemical reactions are the three main pathways for
89 eliminating pollutants from the environment. In the literature, dicamba is observed to be stable in
90 aqueous conditions and does not undergo hydrolysis as a degradation pathway.² The

1
2
3 91 biodegradation of dicamba has been examined in many publications.¹⁹⁻²² There are a handful of
4
5 92 papers in the literature exploring the photochemistry of dicamba, but most of the previous work
6
7 93 on the photochemistry of dicamba has been done in aqueous solutions with advanced oxidation
8
9 94 processes (AOP) such as the use of UV/H₂O₂ to degrade the molecule,⁴ on photocatalysis
10
11 95 processes such as using TiO₂ as a catalyst,^{4,23,24} or the photo-Fenton reaction^{2,25} to degrade the
12
13 96 molecule. There has been no work investigating the photochemistry of formulated dicamba on
14
15 97 the surface of crops or examining the role adjuvants, like those added in commercial
16
17 98 formulations, have on the photochemistry of dicamba. Here we examine the photochemistry of
18
19 100 dicamba in aqueous solution (without advanced oxidation processes), in solution with a model
20
21 101 adjuvant, and on the surfaces of epicuticular waxes collected from corn plants. The
22
23 102 photochemistry of the commercial product Diablo[®] was also examined in aqueous solution and
24
25 103 on the epicuticular waxes.

26 104 MATERIALS AND METHODS

27 105 Chemicals and Instrumentation

28
29 106 Analytical standards of dicamba (Sigma-Aldrich, >98.9%) were produced ranging from 0.2 mg/L
30
31 107 to 40 mg/L in Milli-Q water (Milli-Pore). The commercial formulation of dicamba, Diablo[®], was
32
33 108 acquired from Red River Specialties, LLC and used as received. Diablo[®] was diluted from 480
34
35 109 g/L to 15 mg/L in 1 mM pH 7 phosphate buffer. Chromatographic solvents, such as acetonitrile
36
37 110 (ACN, ≥99.9%) and water (>99.9% HPLC-grade), were purchased from Sigma-Aldrich.
38
39 111 Phosphate buffers were prepared using H₃PO₄, NaH₂PO₄·H₂O, or NaHPO₄·3H₂O (Fisher
40
41 112 Scientific) as needed in Milli-Q water and were 1 mM in total phosphate. Natural organic matter
42
43 113 (NOM) was obtained from the International Humic Substance Society (IHSS) collected from the
44
45 114 Suwannee River (1R101N). Actinometry solvents used were methanol (Sigma-Aldrich, ≥99.9%
46
47 115 or EM Science, 99.97%), p-nitroacetophenone (Aldrich, 98%), and pyridine (Sigma-Aldrich,
48
49 116 99.8%). The adjuvant MAKON[®] DA-6 was obtained as a sample from the Stepan Company of
50
51 117 Northfield, IL. MAKON[®] DA-6 is a non-ionic surfactant containing mostly isodecyl alcohol
52
53 118 ethoxylate.

54 120 HPLC analysis for dicamba was performed using 1290 Infinity autosampler and binary pump,
55
56 121 1200 series thermostatted column compartment and 1100 series diode array detector (Agilent)

1
2
3 122 and a reversed-phase 50 mm × 2.1 mm i.d. Eclipse Plus C18, 1.8b micron, S/N: USDAY 40893
4
5 123 column (Agilent). The mobile phase consisted of aqueous %A 1.7 mF phosphate buffer (pH ~3)
6
7 124 and %B ACN in a gradient of 20-50-75-20-20 %B at 0-3.5-4.0-4.01-5 minutes. A flow rate of
8
9 125 0.5 mL/minutes and a run time of 5 minutes were used. Sample injection volume was 20 µL,
10
11 126 column temperature 36°C, and UV-Vis detection set at 220 and 254 nm wavelengths.

12 127 Chromatographic results from HPLC analysis for dicamba are displayed at 220 nm.
13
14 128

15 129 The p-nitroacetophenone/pyridine samples from the quantum yield experiments were analyzed in
16
17 130 the HPLC using the same instrumentation and mobile phase. A reversed-phase 75 x 2.1 mm i.d.
18
19 131 ACE Excel 3 Super C18, S/N: A116046 column (Advanced Chromatography Technologies
20
21 132 LTD) was used with a gradient of 2-80-80-2-2 %B at 0-0.75-1.0-1.01-1.25 minutes. A flow rate
22
23 133 of 0.6 mL/min was used with a stop time of 1.25 minutes, 1 µL sample injection volume, and a
24
25 134 column temperature of 50°C. UV-Vis detection was monitored at 288, 254, and 220 nm
26
27 135 wavelengths. Chromatographic results from the HPLC analysis for p-nitroacetophenone/pyridine
28
29 136 are displayed at 288 nm.

30 137

31 138 Photolysis of Dicamba in Aqueous Solutions

32 139 The photolysis of dicamba was examined with three different irradiation sources: a Q-Sun Xe-1
33
34 140 Solar Simulator with a Daylight-Q filter, a Rayonet RPR-100 Photochemical Reactor (Southern
35
36 141 New England Ultraviolet Company), and outdoors in St. Peter, MN on the Gustavus Adolphus
37
38 142 College campus (44° 20' 0" N, 93° 58' 0" W) in June 2019.

39 143

40
41 144 The photolysis of dicamba in aqueous solution (15 mg/L) with the Q-Sun Xe-1 Solar Simulator
42
43 145 was with a Xe lamp, an irradiance of 0.40 W m² at 340 nm, a nominal cut-on of 295 nm, and at
44
45 146 30 °C. The spectral irradiance of the lamp is shown in Figure S-21 along with the UV-Vis
46
47 147 absorbance spectra of dicamba. Each sample of dicamba was prepared by dissolving the
48
49 148 necessary amount of solid in phosphate buffer (at pH 7), followed by sonication for 20 minutes.
50
51 149 10 mL of the 15 mg/L dicamba solution was added to two glass petri dishes which were then
52
53 150 placed into the Q-Sun solar simulator for irradiation. Simultaneously, two petri dishes containing
54
55 151 PNAP/PYR actinometer were also irradiated (see details on actinometry below). Every two
56
57 152 hours, samples were collected from the four petri dishes: solvent was first added back to the petri
58
59
60

1
2
3 153 dishes until the dishes contained 10 mL again (to account for evaporation), then 100 μ L of
4
5 154 sample was collected into HPLC vials with glass inserts and set aside for HPLC analysis.
6
7 155 Samples were collected every two hours over four 8-hour periods with samples being stored in
8
9 156 the refrigerator at 9 °C between the four periods.

10 157
11
12 158 Multiple photochemical aqueous solution reactions of dicamba (15 mg/L) were conducted using
13
14 159 a 600 mL quartz reaction flask filled with ~500 mL dicamba, held in a Rayonet RPR-100
15
16 160 Photochemical Reactor (Southern New England Ultraviolet Company). Each sample of dicamba
17
18 161 was prepared by dissolving the necessary amount of solid in desired solvent, usually phosphate
19
20 162 buffer (at pH 7), HCl solution (pH 1), or Milli-Q water, followed by sonication for 20 minutes.
21
22 163 In experiments with additions of adjuvant or NOM, the appropriate compound (0.24 mL of a
23
24 164 1.05% (w/w) MAKON® DA-6 stock to make 5 mg/L DA-6 or solid NOM to make solutions of 1
25
26 165 mg/L, 5 mg/L, or 10 mg/L) was added after sonication. For quenching reactions, isopropanol
27
28 166 (final concentration of 1%) or l-histidine (final concentration of 5 mM) was added to the solution
29
30 167 immediately prior to irradiation. In the experiment conducted with H₂O₂, 283 μ L of 30% H₂O₂
31
32 168 was added after sonication of the 15 mg/L dicamba solution. For the experiments using
33
34 169 Minnesota River water, river water was collected using new sample bottles that were rinsed three
35
36 170 times before collection and then river water was filtered using a Millipore glass vacuum filter
37
38 171 with fiber glass filter paper (5 microns). Dicamba was added to the filtered river water to make a
39
40 172 solution of 15 mg/L. For all experiments, solutions were then added to the quartz flask for
41
42 173 experimental irradiation in the Rayonet. Samples were irradiated in a dark room with eight 35 W
43
44 174 low-pressure mercury lamps that emitted light centered at 310 nm. The 310 nm lamps have a
45
46 175 spectral distribution with a full width at half max of 40 nm; the spectral irradiance of the lamps is
47
48 176 shown in Figure S-21. The lamps were uniformly distributed around the vessel, and each sample
49
50 177 was irradiated for at least 60 minutes. One set of experiments was conducted in the manner
51
52 178 described above, but leaving the lights of the Rayonet turned off; this served as the ‘dark control’
53
54 179 to ensure any observed degradation was due to the UV-light. In addition, one set of experiments
55
56 180 was conducted in the Rayonet with UVC lamps (254 nm) to examine the impact of wavelength.

57 181
58
59 182 An outdoor photolysis experiment was conducted in St. Peter, MN on the Gustavus Adolphus
60
183 College campus (44° 20' 0" N, 93° 58' 0" W) on June 12-13, 2019, from 11:10 am – 5:10 pm

1
2
3 184 each day (samples were stored in a refrigerator overnight at 9 °C between the two days). The
4
5 185 average temperature of a control sample (Milli-Q water) during the experimental time period was
6
7 186 31 ± 4 °C. 15 mL samples were placed into 14 quartz test tubes in a 4 × 6 in. black Rubbermaid
8
9 187 notecard box at an upward angle of 45° with the open ends facing east. The irradiated samples
10
11 188 were 15 mg/L dicamba in pH 7 phosphate buffer and 6.5 × 10⁻⁵ M PNAP/0.198 M pyridine in
12
13 189 50:50 Milli-Q water:methanol, the latter as a chemical actinometer. At hour time intervals,
14
15 190 duplicate aliquots of 100 µL were removed and placed into amber HPLC vials. The vials were
16
17 191 immediately stored in the refrigerator (at 9 °C) for later HPLC analysis.
18

19 193 Determination of Dissolved Oxygen Concentration in Aqueous Dicamba Solutions

20
21 194 A modified Winkler Titration method was used to determine the dissolved oxygen (DO)
22
23 195 concentration in oxygen rich and oxygen poor solutions in 1 mM phosphate buffer, pH ~7.²⁶ O₂
24
25 196 (for oxygen rich solutions) or N₂ (for oxygen poor solutions) was introduced into the solutions
26
27 197 via a diffusion stone for 15 minutes before taking 1 mL samples every 10 minutes for 90 minutes
28
29 198 (the length of one photodegradation experiment). Winkler titration solutions were made based on
30
31 199 the literature.^{26,27} Samples were added to a small test tube containing 0.5 mL of hexane. To the
32
33 200 sample layer, 5 µL of manganese sulfate solution and 5 µL of alkali-azide-iodide solution were
34
35 201 added consecutively and allowed to develop for 3 minutes. Concentrated sulfuric acid was added
36
37 202 to the sample layer in 0.1 mL volume. After another 3 minutes, the sample layer was transferred
38
39 203 to a new small test tube with 1-2 drops of starch solution (and 5 drops of Milli-Q Water for the
40
41 204 oxygen sparged solution). The sample was titrated using a 0.000314 M sodium thiosulfate titrant.
42
43 205 Photolysis experiments were conducted as described above with both oxygen sparged and
44
45 206 nitrogen sparged 15 ppm dicamba solutions.
46

47 207 48 208 Photolysis of Dicamba on Corn Wax Surface

49
50 209 Corn wax trials were conducted to simulate conditions of dicamba applied onto the surface of
51
52 210 a corn plant. The following methods were adapted from the work of ter Halle et al.²⁸ The
53
54 211 methods have also been used by others.²⁹⁻³¹ For each set of experiments, approximately 160 corn
55
56 212 plants (Anderson Seed Company, St. Peter, MN) were grown in a greenhouse to the trifoliate
57
58 213 stage, approximately 9-11 days. All corn leaves of each plant were cut off and soaked for two
59
60 214 minutes in 75 mL of dichloromethane. The solution was then filtered with a vacuum filter and

215 approximately 10 mL of solution was deposited onto eight glass petri dishes and allowed to
 216 evaporate. Evaporating the solvent off left a layer of plant wax remaining on the surface of the
 217 petri dishes. Dicamba solutions of 15 mg/L concentration were applied in 10 mL volumes and
 218 left to evaporate the solvent off, leaving a solid layer of dicamba. For the experiments with
 219 adjuvant, solutions of MAKON[®] DA-6 (in concentrations of 1 mg/L, 5 mg/L, 10 mg/L, or 15
 220 mg/L) with 15 mg/L dicamba were prepared and added to the petri dishes in the same way as
 221 above. According to dicamba field application notes, dicamba should be applied to corn at
 222 concentrations less than 0.75 lb/acre,³² applying 10 mL of a 15 mg/L solution to a 5 mm
 223 diameter petri dish gives a 0.68 lb/acre equivalent. Measurement of dicamba recovered from the
 224 surface prior to irradiation shows overall losses (e.g., volatilization, lack of recovery, degradation
 225 in the dark, etc.) are less than 5%. Corn wax plates with applied dicamba samples were exposed
 226 to light (irradiance of 0.40 W m² at 340 nm) in a Q-Sun Xe-1 Solar Simulator with a Daylight-Q
 227 filter (nominal cut-on of 295 nm) at 42 °C. Samples were taken at time intervals ranging from
 228 12-48 hours by reintroducing 10 mL of Milli-Q water before 1 mL aliquots were taken in
 229 duplicate. All dicamba irradiated corn wax samples were analyzed using the same HPLC
 230 parameters as the dicamba aqueous solution samples. The same method was also used for
 231 dicamba with added adjuvant on corn wax surfaces, Diablo[®] on corn wax surfaces, and on
 232 dicamba on glass surfaces.

233

234 Actinometry/Quantum Yield of Dicamba

235 The quantum yield of dicamba under 310 nm lamps in the Rayonet and under the Xe lamps in
 236 the Q-Sun solar simulator and in solutions of varying water quality conditions was determined
 237 using p-nitroacetophenone/pyridine as an actinometer. A 6.5×10^{-5} M PNAP/0.198 M pyridine
 238 solution was made in 50:50 Milli-Q water:methanol and was irradiated using under 310 nm
 239 lamps in the Rayonet RPR-100 photochemical reactor for 90 minutes or in the Q-Sun solar
 240 simulator for 32 hours. In the Rayonet, 1 mL aliquots were sampled in duplicate every 5 minutes
 241 and placed in amber HPLC vials. In the Q-Sun, 100 μ L aliquots were sampled in quadruplicate
 242 every 1-2 hours. The samples were then analyzed in the HPLC using the methods described
 243 above. The quantum yield of dicamba was calculated from Eq 1 (adapted from Leifer eq 6.18):³³

$$244 \quad \phi_D = \frac{k_D}{k_{PNAP}} \left\{ \frac{\sum_{\lambda} \epsilon_{\lambda,PNAP} L_{\lambda}}{\sum_{\lambda} \epsilon_{\lambda,D} L_{\lambda}} \right\} \phi_{PNAP} \quad \text{Eq 1}$$

245

246

247

248

249

250

251

1
2
3 245 where k_D and k_{PNAP} are the first-order reaction rate constants for dicamba and PNAP, $\varepsilon_{\lambda,D}$ and
4
5 246 $\varepsilon_{\lambda,PNAP}$ are the molar absorptivity ($M^{-1} cm^{-1}$) at each wavelength, λ , for dicamba and PNAP, L_λ
6
7 247 is the irradiance at wavelength, λ , ($einstein\ cm^{-2}\ day^{-1}$) and ϕ_{PNAP} is the quantum yield for PNAP
8
9 248 ($\phi_{PNAP} = 0.0169[Pyridine]$). The sums in Eq 1 were summed over wavelengths of 270 – 350 nm
10
11 249 for the Rayonet 310 nm lamps and over wavelengths of 280 – 400 nm for the Q-Sun Xe lamps
12
13 250 (see Figure S-21 for spectra of the lamps). The daily-average solar irradiance spectrum for June
14
15 251 21 at 40° N latitude from Apell and McNeill was used to determine the quantum yield of
16
17 252 dicamba in the outdoor experiments.³⁴
18

253

19 254 Determination of Environmental Half-Lives

20
21 255 The environmental half-life of dicamba was calculated using the GCSOLAR model developed
22
23 256 by Zepp and Cline.³⁵ This program transforms the laboratory data collected (UV-Vis spectra of
24
25 257 dicamba and water samples and quantum yields) into environmentally relevant half-lives. The
26
27 258 following input parameters were used: quantum yield of dicamba in aqueous solution irradiated
28
29 259 in the Q-sun solar simulator, molar absorptivity coefficients of dicamba and MN River water,
30
31 260 longitude of 94° W, clear sky, typical ozone concentration in the atmosphere, default depth
32
33 261 parameters and water refraction index, and sea level. A series of latitudes were used as input:
34
35 262 30°, 40°, 50°, 60° N. The half-lives obtained were integrated over the entire day.
36

263

37 264 Preparation and Analysis of Dicamba Photoproducts

38
39 265 A 40 mg/L solution of dicamba in Milli-Q water was irradiated under 310 nm lamps in the RPR-
40
41 266 100 for 70 minutes, collecting samples every five minutes. The higher concentration and the
42
43 267 RPR-100 were used for photoproduct analysis for analytical measurement reasons. The
44
45 268 photoproduct separation and mass spectrometry was performed for time samples at 0, 35, and 70
46
47 269 minutes using an Agilent 1290 Infinity II LC coupled with an Agilent 6545XT AdvanceBio
48
49 270 LC/Q-TOF. The mass analyzer was calibrated with a standard tuning compound mixture
50
51 271 (Agilent, p/n: G1969-85000). The chromatographic separation was performed with a 50 mm x
52
53 272 2.1 mm i.d. Eclipse Plus C18, 1.8b micron, S/N: USDAY 40893 column (Agilent) with a mobile
54
55 273 phase consisting of %A 1% formic acid in Milli-Q water, %B acetonitrile and a gradient of 2-40-
56
57 274 75-2-2% B from 0.00-3.50-4.00-4.01-5.00 minutes. The flow rate was 0.5 mL/min, 5 μ L sample
58
59 275 injection volume, 36°C column temperature and 50:50 split between the MS and DAD detector
60

1
2
3 276 set at 254 and 220 nm wavelengths. Tandem MS was done in negative ion mode with the
4
5 277 following electrospray ionization parameters: 4000 V capillary voltage, 2000 V nozzle voltage,
6
7 278 325°C gas temperature, 5 L/min drying gas flow, 25 psig nebulizer pressure, 275°C sheath gas
8
9 279 temperature, and 12 L/min sheath gas flow. An 80 V fragmentor voltage for both MS dimensions
10
11 280 was used and had a 45 V skimmer and OCT 1 FR Vpp at 750 V for the MS TOF. For both
12
13 281 dimensions, the m/z range was 50 - 970 m/z with an acquisition rate of 5 spectra/s for dimension
14
15 282 1 and 1.5 spectra/s for dimension 2 (1152 and 3948 transients/spectrum
16
17 283 respectively). MassHunter™ Qualitative Analysis software package, version B.08.00 (Agilent
18
19 284 Technologies) was used to extract mass spectra from the separated photoproducts and to predict
20
21 285 putative photoproduct molecular formulas.

22
23 286
24 287 Gaussian 09³⁶ with WebMO graphical interface, version 17.0.012e, was used to examine the
25
26 288 molecular energies of photoproducts, including possible isomers. All data presented are from
27
28 289 density functional theory calculations for geometry optimization and frequency calculations
29
30 290 using B3LYP functionals with the 6-311+G(2d,p) basis set. The default conditions of 298.15 K
31
32 291 and 1.0 atm were used for thermochemistry calculations.

33 292 Rate Constant Data Analysis

34 293 All reported pseudo-first-order rate constants were obtained from weighted linear least-squares
35
36 294 analysis of the experimental data (regressing $\ln[C]$ versus time, where C equals the molar
37
38 295 dicamba concentration). The weighting factors used for the regression (typically $1/\sigma_Y^2$, where
39
40 296 σ_Y^2 is the variance on each value of $\ln[C]$) were set equal to C^2 due to the transformation
41
42 297 variable.³⁷ The use of this weight reduces the effect of the later time points (which are likely to
43
44 298 have greater variance and error) on the fit of the regression line, yielding more accurate reaction
45
46 299 rate constants. Most aqueous experiments were conducted in at least duplicate, and Table 1 lists
47
48 300 the number of trials for each set of experiments, n. The standard deviations for experiments in
49
50 301 Table 1 with $n \geq 2$ are the standard deviations over the trials, and the standard deviations for
51
52 302 experiments with $n = 1$ is the standard deviation of the regression fit (used as an approximation
53
54 303 as a true standard deviation was unable to be calculated).

55 304

56 305 RESULTS AND DISCUSSION

57
58
59
60

307

308 Aqueous Experiments

309 Kinetics of Dicamba Photolysis

310 The UV-Vis spectra of dicamba as a function of irradiation time is shown in Figure 1. As can be
311 seen in the figure, the absorbance peak near 203 nm, corresponding to dicamba, decreases with
312 increasing irradiation time. There is an isosbestic point near 235 nm formed between the
313 decreasing dicamba peak and the increasing peak with a center near 254 nm.

314

315 Each photolysis reaction of dicamba examined showed an exponential decrease of dicamba,
316 suggesting that the photodegradation follows a first order or pseudo first order reaction
317 mechanism. Figure 2 shows a selection of the photolysis data, illustrating the first order nature of
318 the reaction. The figure shows a photolysis experiment (conditions: 15 mg/L dicamba buffered
319 with 1 mM phosphate buffer at pH 7, irradiated with 310 nm light in the Rayonet) and an
320 experiment conducted in the dark (conditions: 15 mg/L dicamba buffered with 1 mM phosphate
321 buffer at pH 7, with no irradiation). The dark control experiment showed no loss of dicamba over
322 70 minutes and confirmed that dicamba was not lost via evaporation in the aqueous solution
323 photoreaction chamber. The photolysis experiments of dicamba, applying 310 nm irradiation to a
324 15 mg/L solution of dicamba buffered to pH 7, in a quartz photoreaction flask (experiment
325 conducted $n = 6$ times), yielded a photodegradation rate constant of $k = 23.0 \pm 2.9 \text{ day}^{-1}$. Also
326 illustrated in Figure 2 are the kinetic data plots for two additional experiments: irradiation of
327 dicamba in the Rayonet photoreactor with a) an oxygen desaturated solution and b) with addition
328 of 5 mM H_2O_2 . These two reactions will be discussed further below. These four rate constants
329 can be found in Table 1, a summary of all rate constants from this work. The Supporting
330 Information contains all the plots (S1-S19) used to obtain the aqueous rate constants in Table 1.

331

332 Table 1 also provides the rate constants for the experiments completed with aqueous solutions in
333 the Q-Sun solar simulator ($k = 1.3 \pm 0.7 \text{ day}^{-1}$) and outdoors in June 2019 ($k = 0.04 \pm 0.03 \text{ day}^{-1}$).
334 These two experiments also had an exponential decrease of dicamba over irradiation time, but
335 the reaction rates are much slower than those for the reactions conducted in the Rayonet. The
336 slower rate is presumably due to the spectra of light from the two light sources: both the Q-Sun
337 solar simulator and the sun have a cut-off at 295 nm which limits exposure to some of the higher

1
2
3 338 energy wavelengths present in the Rayonet. The quantum yield for dicamba at pH 7 irradiated in
4
5 339 the Q-Sun was determined to be 0.0013, and the quantum yield for dicamba outdoors was
6
7 340 0.0012. These quantum yields, which are calculated relative to photon flux from each light
8
9 341 source, are nearly identical, showing that the Q-Sun solar simulator is a good lab based light
10 342 source to use to examine the environmental fate of dicamba.

11
12 343

13 344 Dicamba photolysis in varied pH and wavelength

14 345 The impact of wavelength on the photodegradation of dicamba was examined in the Rayonet
15 346 reactor by using both UVC (254 nm) and UVB (280 – 340 nm, centered at 310 nm) lamps. The
16
17 347 observed rate constant for the photodegradation of dicamba under 254 nm light was 56.2 ± 2.9
18
19 348 day^{-1} (Figure S2). This rate constant is a factor of 2 larger than the rate constant obtained under
20
21 349 the 310 nm lamps, and two orders of magnitude larger than the reaction in the Q-Sun. The
22
23 350 difference in degradation observed with the different lamps is due to differences in photon flux
24
25 351 and energy of the photons from the two sets of lamps, the differences in absorbance of light by
26
27 352 dicamba at the two wavelengths, and possibly due to different reaction mechanisms. Since UVC
28
29 353 light is not environmentally relevant, no further work was done at this wavelength. Although the
30
31 354 Rayonet UVB light source is not a perfect model for environmental light sources, the Rayonet
32
33 355 UVB light was used for a series of further experiments to more deeply explore the
34
35 356 photochemistry of dicamba as these experiments could be conducted more quickly than
36
37 357 experiments outdoors or in the Q-Sun solar simulator, allowing time for more exploration of the
38
39 358 photochemical reactions. In addition, a quick look at the irradiated samples from the Q-Sun on
40
41 359 the LC-MS showed similar photoproducts to the samples from the samples irradiated with the
42
43 360 Rayonet UVB lamps, suggesting that the mechanisms may be similar.

44
45 361

46 362 The pKa of dicamba has been observed to be about 1.9, and the molecule is primarily
47
48 363 deprotonated at a pH of 7.² Despite the fact that dicamba will be in the carboxylate form in
49
50 364 natural waters, there may be environmental circumstances where the protonated form may be
51
52 365 important. Therefore, experiments of dicamba photolysis in aqueous solutions were conducted
53
54 366 under UVB light in the Rayonet in HCl solutions at pH 1. A rate constant of $20.2 \pm 1.4 \text{ day}^{-1}$ was
55
56 367 observed, suggesting that the reaction is slightly slower in acidic conditions than in the solutions
57
58 368 at pH 7. Although the absorbance of 310 nm light is similar for the solutions buffered at both pH
59
60

1
2
3 369 1 and pH 7 (see Figure S-21), the quantum yield of the protonated form of dicamba ($\phi = 0.048$)
4
5 370 is lower than that of the deprotonated form ($\phi = 0.073$) which leads to the observed slowing of
6
7 371 the reaction under acidic conditions. Thus, the slower reaction at pH 1 is due almost entirely due
8
9 372 to the lower quantum yield for the protonated form. The favoring of photodegradation in the
10
11 373 anionic form has also been observed for other herbicides with a benzoic acid moiety.^{38,39} The
12
13 374 quantum yields obtained here can also be compared to that obtained by Aguer et al.; these
14
15 375 authors obtained a quantum yield of 0.022 for the deprotonated form of dicamba at a wavelength
16
17 376 of 275 nm.² Aguer et al. also observed that the absorbance of acidic solutions of dicamba was
18
19 377 greater than the absorbance of basic/neutral solutions as is shown from this work in Figure S-21.
20
21 378 Although the wavelengths used in the two studies (this one and the one by Aguer et al.) differ
22
23 379 slightly, the quantum yields are found to be on the same order of magnitude. However, both the
24
25 380 quantum yields from the Rayonet experiment and the Aguer et al. paper are an order of
26
27 381 magnitude larger than the quantum yields obtained from the Q-Sun and outdoor experiments due
28
29 382 to the wavelengths of light used in each experiment. For environmental purposes, the quantum
30
31 383 yield from the Q-Sun and/or outdoor experiment should be used.

384 385 Environmental Direct Photolysis Half-Lives of Dicamba

386 Using the quantum yield of dicamba at pH 7 under the Xe lamp of the Q-Sun solar simulator
387 (i.e., $\phi = 0.0013$), the GCSOLAR model was used to determine the environmental direct
388 photolysis half-lives dependent on the season and the degree of latitude in pure water and in
389 Minnesota River water. These direct photolysis half-lives are presented in Table 2. As seen in the
390 table, the half-lives for dicamba range from 3.74 to 97.7 days under these conditions. As
391 expected, half-lives are longer in the winter months and at higher latitudes. Overall, this analysis
392 suggests that dicamba will photodegrade in the environment within days in the typical growing
393 season (summer) and crop locations (latitudes 35° – 45° N) in the United States. These half-lives
394 are on the same order of magnitude as those reported for degradation in aerobic soil in the EPA
395 registration for dicamba.⁴⁰

396 397 Dicamba photolysis in varied oxygen concentration

398 Photodegradation experiments were conducted in oxygen sparged and nitrogen sparged solutions
399 under UVB light in the Rayonet to further examine the role of oxygen in the photodegradation of

1
2
3 400 dicamba. The average dissolved oxygen content in the oxygen sparged solution was 13 ± 4 mg/L
4
5 401 with no significant drop in dissolved oxygen content over the 90 minutes. The average dissolved
6
7 402 oxygen content in the nitrogen sparged solution was 2.8 ± 0.8 mg/L with no significant change in
8
9 403 dissolved oxygen content over the 90 minutes; this corresponds to approximately 1/5th of the
10
11 404 dissolved oxygen in the oxygen sparged solution. It was assumed that the addition of 15 mg/L of
12
13 405 dicamba would not alter the dissolved oxygen concentration in the buffered solutions, and
14
15 406 Winkler titrations were not repeated with dicamba containing aqueous solutions prior to the
16
17 407 photolysis.
18

19 409 The photodegradation rate constant in the nitrogen sparged solution was determined to be $58 \pm$
20
21 410 29 day^{-1} , a factor of three greater than that of the photodegradation rate constant in the oxygen
22
23 411 sparged experiment ($21.6 \pm 4.3 \text{ day}^{-1}$). The kinetic data from the nitrogen sparged solution can be
24
25 412 seen in Figure 2. The inhibiting effect of oxygen has also been observed in the photolysis of 5-
26
27 413 halogenosalicylic acids,⁴¹ compounds with structural similarities to dicamba, and in the
28
29 414 photolysis of chlorothalonil⁴². Using laser flash photolysis and computational chemistry
30
31 415 calculations, the authors of the paper on halogenosalicylic acids were able to show that oxygen
32
33 416 quenched the triplet excited state of the acids.⁴¹ Given the structural similarities of dicamba to
34
35 417 these acids, it is hypothesized that the triplet excited state of dicamba is important in its
36
37 418 photochemistry and that oxygen acts as a quencher of this state. At reduced oxygen levels, the
38
39 419 lifetime of the triplet is longer, leading to faster photoreactions.
40

420 421 Diablo[®] and adjuvants

42 422 The rate constant of Diablo[®] photodegradation in aqueous solution under UVB light in the
43
44 423 Rayonet, $21.6 \pm 1.4 \text{ day}^{-1}$, was statistically consistent with the rates seen with aqueous solution
45
46 424 experiments of dicamba (Table 1). This is not necessarily intuitive as the additional ingredients
47
48 425 in the commercial mix could interact with the dicamba and/or the light to lead to a different
49
50 426 reaction rate and/or mechanism.

51
52 428 As discussed in the introduction, one category of molecules added to commercial herbicides
53
54 429 mixtures are adjuvants, designed to help the solutions spread on surfaces. A series of
55
56 430 experiments were conducted to explore the impact of these adjuvants on the photodegradation of
57
58
59
60

1
2
3 431 dicamba. The concentrations and types of adjuvants used in commercial forms of dicamba are
4 432 proprietary, but there are several surfactants available on the market. Here, a solution of 5 mg/L
5 433 MAKON[®] DA-6, an isodecyl alcohol ethoxylate, and 15 mg/L dicamba was tested to model a
6
7 434 commercial formulation in solution. The rate constant for the model commercial formulation,
8
9 435 $33.1 \pm 7.2 \text{ day}^{-1}$, was significantly higher than the aqueous solution experiments of dicamba.
10
11 436

12
13
14 437 Several papers in the literature have found that nonionic surfactant molecules can photosensitize
15 438 organic molecules such as herbicides.^{43–50} The photosensitization of the dicamba photolysis in
16 439 the solutions with DA-6 could be due to micelle formation changing the environment in solution,
17 440 from energy transfer between the herbicide and surfactant, and/or indirect photolysis from the
18 441 formation of other reactive species. Micelles can form in solutions with surfactant if the
19 442 concentration of surfactant is above the critical micelle concentration (cmc). For Makon DA-6,
20 443 an isodecyl alcohol ethoxylate with 6 ethoxylate units, the cmc is estimated to be above 400
21 444 ppm.⁵¹ For the experiments presented here, with 5 ppm DA-6, no micelles should be formed.
22 445 Experiments verifying this conclusion are described in the Supporting Information. Micelles,
23 446 therefore, should not be the reason for the enhanced photodegradation. No experiments were
24 447 conducted to test the hypothesis that energy transfer may be leading to the photosensitization, but
25 448 this has been observed in the literature on photochemistry and nonionic surfactants.^{45–48}
26
27
28
29
30
31
32
33
34
35

36 450 To investigate whether the adjuvant was contributing to an indirect photolysis pathway,
37 451 quenchers for singlet oxygen (5 mM l-histidine) and hydroxyl radical (1% isopropanol) were
38 452 separately added to solutions of adjuvant/dicamba and irradiated. The rate constants for the
39 453 indirect photolysis experiments of 15 mg/L dicamba with 5 mg/L DA-6 using l-histidine and
40 454 isopropanol as quenchers were $20.2 \pm 2.9 \text{ day}^{-1}$ and $24.5 \pm 5.8 \text{ day}^{-1}$ respectively (Table 1).
41 455 These photodegradation rate constants are consistent with the direct photodegradation rate
42 456 constant of dicamba, suggesting that there may be indirect photolysis occurring in these
43 457 solutions. It remains unclear what reactive species may be forming in these reactions, and how
44 458 they are formed, but it is known that dicamba will react with hydroxyl radicals (the literature
45 459 gives rate constant for the reaction of dicamba with hydroxyl radical as either $k_{\text{OH}} = 1.3 \times 10^9$
46 460 $\text{M}^{-1}\text{s}^{-1}$ or $3.5 \times 10^9 \text{ M}^{-1}\text{s}^{-1}$ at pH 7 and $T = 20 \text{ }^\circ\text{C}$).^{52,53} We also conducted an experiment examining
47 461 the susceptibility of dicamba to hydroxyl radical reactions by irradiating a 15 mg/L dicamba
48
49
50
51
52
53
54
55
56
57
58
59
60

1
2
3 462 solution spiked with 5 mM H₂O₂. When irradiated at 310 nm, H₂O₂ can produce hydroxyl
4
5 463 radicals (quantum yield for this reaction at 308 nm is 1.93±0.39).⁵⁴ As can be seen in Table 1 and
6
7 464 in Figure 2, the photodegradation rate constant for dicamba in the H₂O₂ spiked experiment was
8
9 465 93.6 ± 2.9 day⁻¹, indicating that dicamba photodegradation rates are much faster in the presence
10
11 466 of hydroxyl radicals. Although the H₂O₂ experiment illustrates that dicamba can react with
12
13 467 hydroxyl radicals, it does not provide evidence that hydroxyl radicals are involved in the
14
15 468 experiments with the surfactant DA-6. Overall, it has proven challenging to conclude exactly
16
17 469 why the surfactant experiments show an increase in photodegradation rate.
18

19 471 Dicamba photolysis in varied natural organic matter (NOM) concentration

20 472 Natural organic matter (NOM) is found in all river waters, commonly at concentrations between
21
22 473 1 and 10 mg/L.⁵⁵ The structure and chemical composition of NOM varies due to soil type,
23
24 474 location, biological degradation, and other factors. In these experiments, Suwannee River NOM,
25
26 475 ordered from the International Humic Substances Society, was used as a model NOM. This
27
28 476 NOM largely consists of C (52.47% by mass), O (42.69% by mass), H (4.19% by mass), and N
29
30 477 (1.10% by mass) and should be negatively charged at pH values of 6-9 due to carboxylic acid
31
32 478 functional groups. It has been shown by previous groups that the presence of NOM in water
33
34 479 samples can either increase⁵⁶⁻⁵⁸ or decrease^{38,58-62} photodegradation rates. The increase in
35
36 480 reaction rates is attributed to the ability of NOM to produce reactive oxygen species such as
37
38 481 hydroxyl radicals and/or triplet NOM, and the decrease in rates is typically attributed to
39
40 482 screening of light by the NOM.

41 483
42 484 As can be observed in Table 3, and Figure S-16, an inverse relationship between NOM
43
44 485 concentrations and degradation rates of dicamba was observed in NOM/dicamba experiments.
45
46 486 The addition of 1.2 mg/L NOM to 15 mg/L of dicamba caused no change in reaction ($k = 23.0 \pm$
47
48 487 1.4 day^{-1}); adding 5.2 mg/L of NOM to 15 mg/L dicamba slowed the reaction down and yielded
49
50 488 rate constant of $15.8 \pm 1.4 \text{ day}^{-1}$; and adding 10.0 mg/L of NOM to 15 mg/L dicamba yielded a
51
52 489 rate constant of $13.0 \pm 1.4 \text{ day}^{-1}$. These reaction rate constants suggest that there is no
53
54 490 measurable photosensitization, but it may be likely that light screening is occurring. To test for
55
56 491 the latter, the screening models presented in Leifer's "The Kinetics of Environmental Aquatic
57
58 492 Photochemistry: Theory and Practice" book and in Schwarzenbach, Gschwend, and Imboden's

1
2
3 493 Environmental Organic Chemistry book, were used to correct the measured rate constants for
4
5 494 screening:

6
7 495
$$S(\lambda) = \frac{1 - 10^{-D(\lambda)\alpha(\lambda)z_{mix}}}{2.3z_{mix}D(\lambda)\alpha(\lambda)} \text{ Eq 2}$$

8
9 496 where $S(\lambda)$ is the screening factor, $D(\lambda)$ is the distribution function (equal to 1 in these
10
11 497 experiments), z_{mix} is the vertical distance in a mixed body of water (equal to 4.6 cm in these
12
13 498 experiments), and $\alpha(\lambda)$ is the attenuation coefficient of the medium.^{33,58} $\alpha(\lambda)$ was obtained from
14
15 499 UV-Vis spectra of NOM and dicamba solutions, similar to the work of Espy, et. al.³⁸ Because the
16
17 500 lamps used in the Rayonet photoreactor have a narrow wavelength range, the photodegradation
18
19 501 rate constants obtained in the presence of NOM can be corrected for light screening by dividing
20
21 502 the rate constants by the screening factor:

22 503
$$k_{corr} \approx \frac{k_{obs}^{NOM}}{S(\lambda)} \text{ Eq 3}$$

23
24 504 where k_{obs}^{NOM} is the experimental degradation rate constant for dicamba in the presence of NOM
25
26 505 and k_{corr} is the photodegradation rate constant corrected for light screening. Using the equations
27
28 506 above, the observed rate constants from the NOM experiments were corrected to account for
29
30 507 screening. The corrected rate constants are given in Table 3. As can be seen in the table, the
31
32 508 corrected rate constants match (within experimental error) the photodegradation rate constant of
33
34 509 dicamba in buffered solution. This shows that the observed decrease in photodegradation rate
35
36 510 upon the addition of NOM is accounted for by screening effects.

37 511
38 512 Since the experiments with Suwannee River NOM suggested that screening may be an important
39
40 513 factor in natural water systems, Minnesota River water was collected and used as a matrix for
41
42 514 photodegradation experiments in the lab. The Minnesota River water was collected in St. Peter,
43
44 515 MN in June 2019 and June 2020, and was filtered using a Millipore filtering system prior to
45
46 516 spiking the water with dicamba to a concentration of 15 mg/L. Although the Minnesota River
47
48 517 water was not analyzed, the soil in this area of Minnesota is mostly Lester soil, known to be a
49
50 518 loamy soil with nearly equal amounts of sand and silt, slightly lower clay amounts, and soil
51
52 519 organic matter in the range of 2-5%.⁶³ The rate constant from these experiments, along with the
53
54 520 screening factor, and corrected rate constant are available in Table 3. The rate constants obtained
55
56 521 from these experiments show that photodegradation of dicamba in natural waters is slower than
57
58 522 any of the photodegradation experiments with Suwannee River NOM. Although there is a slight

1
2
3 523 difference between the samples collected from the two different years, the absorbance of light by
4 524 the Minnesota River water is much stronger than in any of the experiments with Suwannee River
5 525 NOM, and the calculated screening factor for the Minnesota River water is 0.580 in 2019 and
6 526 0.884 in 2020. From the 2019 value, one would predict screening is more prevalent in that
7 527 Minnesota River sample as compared to the lab experiments with Suwannee River NOM.
8 528 However, when the rate constant is corrected using the screening factor, the corrected rate
9 529 constant is still slower than the photodegradation rate constant of dicamba in buffered solution.
10 530 The same is observed with the 2020 sample. Hence, screening alone does not account for the
11 531 observed drop in photodegradation rate constant of dicamba in the Minnesota River water.
12
13
14
15
16
17
18
19

20 532
21 533 Adsorption of dicamba to a component in the Minnesota River water is a possibility to explain
22 534 the slower reaction rate in the experiments with the Minnesota River water. Several authors have
23 535 examined the sorption of dicamba to soils.^{64,65} Sakaliene et al., found the adsorption coefficient
24 536 of dicamba to be the lowest for the series of herbicides they examined, with a K_d value ranging
25 537 from 0.03 - 0.08 L/kg depending on the soil type. These low K_d values suggest that adsorption is
26 538 not likely to be the reason for the observed slower reaction rate. This is supported by the fact that
27 539 dicamba and the NOM are both negatively charged at the pH levels in these experiments, and it
28 540 has been shown that negatively charged organic species will adsorb more weakly than the related
29 541 neutral compound to anionic NOM.⁵⁸ In addition, areas of the HPLC peaks from the non-
30 542 irradiated wax samples align with the non-applied solutions, suggesting no adsorption occurs
31 543 upon application. Thus, it remains unclear what is causing the reduced photodegradation rate
32 544 constant of dicamba in the Minnesota River water experiments.
33
34
35
36
37
38
39
40

41 545

42 546 Photolysis on Epicuticular Corn Wax

43 547 Recent studies suggest that the matrix environment in which photolysis studies are conducted
44 548 can affect both photochemical transformation rates and the photoproduct composition.⁶⁶⁻⁷⁰ For
45 549 example, the epicuticular wax of most species absorbs light below 350 nm and also contains
46 550 photosensitizing functional groups such as carboxyl groups.⁷¹⁻⁷³ Thus, in comparison to
47 551 photolysis in water samples or other appropriate controls, photolysis of pesticides sorbed to
48 552 epicuticular wax can be slower (e.g., due to light screening or phase) or faster (e.g., due to
49 553 photosensitization).
50
51
52
53
54
55
56
57
58
59
60

554

555 The experiments with corn waxes are inherently difficult and lengthy due to the time needed to
556 grow the corn, extract and plate the epicuticular wax, and complete the irradiation (so one set of
557 experiments is typically 23-25 days). Dicamba was applied to the surface and irradiated as a
558 solid. Four different sets of corn plants were grown to test the degradation of dicamba on the
559 surface of the waxes. The average rate constant for the photodegradation of dicamba on the
560 surface of corn wax was $0.16 \pm 0.01 \text{ day}^{-1}$. The rate constants observed on the corn wax surfaces
561 a factor of ~ 9 slower than the rate constant from the aqueous solution experiment conducted in
562 the Q-Sun solar simulator, and can be attributed to the solid state of the dicamba during
563 irradiation.

564

565 For comparison, an experiment where dicamba was irradiated after deposition directly on the
566 glass petri dishes was conducted. Other groups interested in the surface photochemistry of
567 herbicides have also used glass or quartz as a comparison.^{2,29,30,66,69} The experiments with
568 dicamba directly deposited on the glass petri dishes yielded no statistically significant reaction.
569 This result can be compared to that of Aguer et al. who found that the photolysis rate constant of
570 dicamba was greater when deposited on clay surfaces compared to the photolysis rate when
571 dicamba was deposited on glass.² In contrast, the herbicides isoproturon and 2,4-D (and related
572 chlorinated phenoxyacetic acids herbicides) and the plant activator acibenzolar S-methyl were
573 found to react faster on glass surfaces than on paraffin wax.^{29,30,66,69} The corn wax used here
574 may have a surface more similar to the clay surfaces used by Aguer compared to paraffin wax
575 surfaces. On the corn waxes, there may be better dispersion across the surface as compared to the
576 glass (perhaps in part because dicamba is anionic in the solid state and glass surfaces are
577 negatively charged).² In addition, Aguer et al. found that dicamba is in a microcrystalline form
578 when in the solid state.² Photolysis in this state is likely slower because less surface is exposed to
579 light and screening can be expected for the central part of the microcrystals. We speculate that
580 the corn wax might allow the dicamba to spread more evenly across the surface.

581

582 Herbicides are generally applied in the field as formulations including the active ingredient and
583 various adjuvants. The most important class of adjuvants is surfactants, used to reduce the
584 surface tension of the spray solution in the field, allowing droplets to spread on the foliage.

1
2
3 585 Today, the most commonly used surfactants used are alkyl ethoxylated surfactants.⁷⁴ A series of
4 586 experiments were conducted to examine the impact adding adjuvant to the epicuticular wax
5 587 surfaces has on the photodegradation of dicamba. A surfactant, MAKON[®] DA-6 (Stepan
6 588 Company) was added to the surface at several concentrations: 1 mg/L, 5 mg/L, 10 mg/L and 15
7 589 mg/L (yielding surface concentrations of 5.1×10^{-4} mg/cm², 2.6×10^{-3} mg/cm², 5.1×10^{-3}
8 590 mg/cm², and 7.7×10^{-3} mg/cm² respectively). Table 1 shows the rate constants from these
9 591 experiments alongside the rate constant from the experiment without adjuvant. Adding adjuvant
10 592 to the experiment increased the photodegradation rate constant; with 1 mg/L DA-6, the rate
11 593 constant increased to 0.26 ± 0.03 day⁻¹ and with 5 mg/L DA-6, the rate constant increased to 0.32
12 594 ± 0.03 day⁻¹. This increase is likely due to the adjuvant breaking apart pesticide aggregates as
13 595 observed by other authors.⁶⁹ However, at higher concentrations of DA-6, the rate constant drops
14 596 again. We believe this is due to experimental limitations; at concentrations of DA-6 above 5
15 597 mg/L, the corn wax layer starts to physically degrade (visibly detaching from the petri dish and
16 598 breaking apart), and some dicamba and DA-6 may be present underneath the wax layer rather
17 599 than at the surface. The corn wax layer also degraded with 15 mg/L DA-6, but the rate constant
18 600 appears higher than the experiment with 10 mg/L DA-6; since the degradation of the wax was
19 601 not controlled, perhaps more dicamba remained on the surface in the 15 mg/L experiment. Also
20 602 shown in Table 1 is an experiment where the commercial product Diablo[®] was applied to the
21 603 surface of corn wax. As with the adjuvant experiments, an increase in photodegradation was
22 604 observed, suggesting that there are other ingredients in the commercial products that lead to
23 605 faster photodegradation on the wax surfaces. Due to the complexity of these experiments, the
24 606 photoproducts and degradation pathways of the reaction on the surface of the epicuticular waxes
25 607 has not been examined.

26 608

27 609 Photoproduct Analysis

28 610 Putative photoproducts were determined using LC-MS-MS analysis of dicamba samples
29 611 irradiated in aqueous solution under UVB light in the Rayonet at times 0, 35, and 70 minutes.
30 612 The putative photoproduct structures, measured mass, calculated mass, error, and formula are
31 613 summarized in Table 4. Proposed structures were determined through fragmentation patterns of
32 614 tandem mass spectra. The chromatogram collected from the LC-MS is shown in Figure S-23 and
33 615 a summary of the computational work related to photoproducts is presented in Tables S-2 and

1
2
3 616 S-3. These photoproduct assignments would be at Level 3: Tentative Candidates on the
4 617 Schymanski scale.⁷⁵

5
6 618
7
8 619 The structure of Photoproduct A was tentatively assigned through the 2D MS fragmentation
9 620 pattern of the peak ion 204.9461 m/z with 162.9534 m/z and 124.9797 m/z peak fragments
10 621 indicating loss of a carboxylic acid group and a chlorine atom respectively (Figure 4).
11
12 622 Photoproduct A has a 0.09 ppm error calculated for the measured mass, calculated mass, and
13 623 chemical formula. The 1D MS chlorine splitting pattern and a calculated 5 double bond
14 624 equivalence confirmed continued presence of the aromatic ring and carboxylic acid as well as
15 625 both chlorine atoms. This putative photoproduct shows a loss of a methoxy group and gain of an
16 626 alcohol group. There is no direct evidence of where the alcohol group adds to the ring, but DFT
17 627 calculations using B3LYP/6-311+G(2d,p) show that the isomer with the alcohol group meta to
18 628 the carboxylic acid is 2.20 kcal/mol lower in energy than the isomer with the alcohol group para
19 629 to the carboxylic acid, and 6.96 kcal/mol lower in energy than the isomer with the alcohol group
20 630 ortho carboxylic acid. This product has also been observed in work examining the
21
22 631 photochemistry of dicamba in the presence of TiO₂.^{4,23}

23
24 632
25
26 633 The structure of Photoproduct B was tentatively assigned using the 1D MS Cl splitting pattern, 5
27 634 double bond equivalence and the evidence of dimerization within the MS after ESI ionization, all
28 635 suggesting that the carboxylic acid remained on the ring. The -0.73 error of the measured mass
29 636 compared to the calculated mass and molecular formula is strong evidence that photoproduct B is
30 637 an addition of an alcohol group in replacement of a chlorine atom, and replacement of the methyl
31 638 ether group with an alcohol group. Time dependent DFT calculations using B3LYP/6-
32 639 311+G(2d,p) showed that the chlorine-carbon bond meta to the carboxylic acid had a larger
33 640 change in bond length upon excitation. DFT calculations were also used to determine the
34 641 molecular energy of three possible isomers of this putative photoproduct (differing by where on
35 642 the ring the two alcohol groups are attached). Two of these isomers are shown in Table 4; isomer
36 643 B1, with the two alcohol groups ortho to the carboxylic acid, had the lowest energy of the three
37 644 isomers examined, and is 1.17 kcal/mol lower in energy than photoproduct B2. This
38 645 photoproduct has also been observed by Fabbri et al. in their work with TiO₂ induce

39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

1
2
3 646 photochemistry of dicamba; these authors also show the product in the form of photoproduct
4
5 647 B1.⁴
6
7 648

8 649 The structure of Photoproduct C was tentatively assigned using the 2D MS fragmentation pattern
9
10 650 of peak ions 234.9564 m/z and 190.9668 m/z. The 5 double bond equivalence suggests that the
11
12 651 carboxylic acid group remains in addition to the ring, and the isomer pattern shows that both
13
14 652 chlorines are still attached. Figure 5a shows the 2D MS fragmentation pattern of the 234.9564
15
16 653 m/z parent peak ion which indicates the presence of a carboxylic acid group from the 175.94 m/z
17
18 654 peak, and two chlorine atoms from the isomer pattern seen in all fragments. The alcohol and
19
20 655 ether group were determined from the 175.94 m/z and 139.97 m/z fragments (Figure 5a and 5b).
21
22 656 The 2D MS fragmentation pattern of the 190.9668 m/z peak ion confirmed the two chlorine
23
24 657 atoms from the isomer pattern visible in all fragments, the loss of an ether group from the 175.94
25
26 658 m/z peak, and the loss of an alcohol group from the 139.97 m/z peak (Figure 5b). DFT
27
28 659 calculations were again used to determine the energetic stability of possible isomers of this
29
30 660 photoproduct. Isomer C1, with the alcohol group para to the carboxylic acid was found to be
31
32 661 0.21 kcal/mol lower in energy than isomer C2.
33

34 662
35 663 The structure of Photoproduct D was tentatively assigned through use of the fragmentation
36
37 664 pattern of the 1D MS with fragment 157.0064 m/z signaling the loss of the carboxylic acid. The
38
39 665 2D MS of the 200.9958 m/z peak ion showed a fragmentation pattern of the loss of the methyl
40
41 666 group from the methyl ether and a subsequent loss of the chlorine from peaks 141.9824 m/z and
42
43 667 107.0078 m/z subsequently. Thus, we can conclude that the photoproduct contains one chlorine
44
45 668 atom, a carboxylic acid group, and a methoxy group. The 5 double bond equivalence and the
46
47 669 evidence of dimerization within the MS also show that the carboxylic acid remains on this
48
49 670 photoproduct. Time dependent DFT calculations using B3LYP/6-311+G(2d,p) were also
50
51 671 completed with photoproduct D, and again the chlorine-carbon bond meta to the carboxylic acid
52
53 672 had a larger change in bond length upon excitation. DFT calculations also show that isomer D2
54
55 673 is 1.25 kcal/mol lower in energy than D1. Both pieces of computational evidence support that the
56
57 674 chlorine meta to the carboxylic acid is the chlorine that has been lost during the formation of this
58
59 675 photoproduct. This product has been observed in work examining the photochemistry of dicamba
60

1
2
3 676 in the presence of TiO_2 but both papers do not assign a location for the alcohol group and
4 677 chlorine atom.^{4,23}

5
6 678
7
8 679 The structure of Photoproduct E was tentatively assigned using 2D MS of the peak ion 232.9865
9 680 m/z and the fragmentation pattern suggesting the loss of the carboxylic acid (173.9715 m/z), the
10 681 loss of a chlorine atom and methyl group (137.9960 m/z), and the loss of an oxygen (125.0238
11 682 m/z). Fabbri et al. also observed this photoproduct in their work with the photodegradation of
12 683 dicamba in the presence of TiO_2 .⁴

13
14 684
15
16 685 The structure of Photoproduct F was tentatively assigned from 1D MS by the lack of a chlorine
17 686 splitting pattern, and the 153.0194 m/z fragment that specified the loss of a carboxylic acid
18 687 group. The presence of the carboxylic acid group was confirmed by the calculated 5 double bond
19 688 equivalence. We draw the putative photoproduct with the two alcohol groups taking the place of
20 689 the missing two chlorine atoms. This is another photoproduct also observed in the work of Fabbri
21 690 et al.⁴ and it has also been observed by Aguer et al.² Both of the photoproducts observed in the
22 691 irradiation of dicamba in aqueous solution in the Aguer paper had a double substitution of
23 692 chlorine for two alcohol groups; perhaps the singly substituted photoproducts were not observed
24 693 due to the experimental conditions/irradiation time.²

25
26 694
27
28 695 Figure 6 shows a proposed transformation pathway for the photodegradation of dicamba in
29 696 aqueous solution. This proposed transformation pathway is similar to that presented by Fabbri et
30 697 al. for the photolytic degradation of dicamba in UV- TiO_2 and UV- H_2O_2 processes,⁴ and is
31 698 supported by our LC-MS-MS data and putative photoproduct assignments. In the transformation
32 699 pathway, the first photoproduct formed is photoproduct C, the addition of an alcohol group to
33 700 dicamba. As with Fabbri,⁴ we posit two different photoproducts forming from photoproduct C –
34 701 photoproduct A, which is a loss of the methoxy group from dicamba, and photoproduct D, loss
35 702 of one chlorine and the addition of an alcohol group at the position meta to the carboxylic acid.
36 703 This isomer of photoproduct D is illustrated as it was found to have the lowest energy in the DFT
37 704 calculations. From photoproduct D, photoproducts B, E, and F form. These photoproducts
38 705 illustrate further loss of chlorine and addition of alcohol groups to the ring. Photoproduct B also
39 706 has a loss of the methoxy group on dicamba. Unlike Fabbri et al., we see no further

1
2
3 707 photoproducts/steps in the proposed transformation pathway. We also saw no evidence for the
4
5 708 oxidation of the methoxy group Fabbri et al. included in their scheme.⁴ We may not be
6
7 709 observing these photoproducts due to the different experimental conditions, analytical
8
9 710 capabilities, and/or the reaction times that were examined.

10 711

11 712 CONCLUSION

13 713 This project provides details on the environmental fate of dicamba, showing that while the
14
15 714 molecule is stable in water and does not undergo hydrolysis, it is susceptible to photolysis when
16
17 715 in aqueous solution, on epicuticular waxes, in the presence of surfactants, and in formulation.

18 716 The quantum yield of dicamba, along with the GCSOLAR program, shows that the
19
20 717 environmental half-life of the compound is on the order of days, especially in the spring and
21
22 718 summer months in which dicamba is applied to fields. The addition of adjuvants causes an
23
24 719 increase in photodegradation of dicamba both in aqueous solution and on epicuticular waxes of
25
26 720 corn. Experiments saturating the solutions with nitrogen gas show that the presence of oxygen
27
28 721 quenches the photodegradation, suggesting that the triplet excited state of dicamba is involved in
29
30 722 the photochemistry. Analysis of photoproducts showed an evolution of aromatic compounds.

31 723

32 724 ACKNOWLEDGMENTS

34 725 This research was funded by the National Science Foundation Grant #1808276 and Gustavus
35
36 726 Adolphus College.

37 727

39 728 We extend our gratitude to Dr. Dwight Stoll, Dr. Gabriel Leme, and Hayley Lhotka for their help
40
41 729 with gathering chromatographic spectra and mass spectra.

42 730

43 731 AUTHOR CONTRIBUTIONS

44
45
46
47 732 **Kaitlyn Gruber:** Data Curation, Formal Analysis, Methodology, Validation, Investigation,
48
49 733 Visualization, Supervision, Writing – Original Draft, Writing – Review & Editing; **Brittany**
50
51 734 **Courteau:** Formal Analysis, Methodology, Validation, Investigation, Visualization, Writing –
52
53 735 Original Draft, Writing – Review & Editing; **Maheemah Bokhoree:** Formal Analysis,
54
55 736 Validation, Investigation, Visualization, Writing – Original Draft, Writing – Review & Editing;
56
57 737 **Elijah McMahon:** Formal Analysis, Investigation, Writing – original draft; **Jenna Kotz:**

58

59

60

1
2
3 738 Formal Analysis, Investigation, Writing – Original Draft; **Amanda Nienow**: Conceptualization,
4
5 739 Data Curation, Funding Acquisition, Formal Analysis, Investigation, Methodology, Project
6
7 740 Administration, Resources, Supervision, Validation, Visualization, Writing – Original Draft,
8
9 741 Writing – Review & Editing

10
11 742 SUPPORTING INFORMATION

12
13
14 743 19 first order kinetic plots showing the data used to obtain rate constants, 1 plot showing UV-Vis
15
16 744 of dicamba in pH 7 phosphate buffer, in solutions containing NOM, and in MN River water, 1
17
18 745 figure showing UV-Vis of dicamba at pH 7 and pH 1 overlaid with the lamp spectra from the
19
20 746 Rayonet and Q-Sun photoreactors, a plot showing the critical micelle concentration of MAKON
21
22 747 DA-6, a LC-MS chromatogram, and two tables with computational data (geometries and
23
24 748 energies).

25 749 FUNDING SOURCE

26
27
28 750 This work has been funded by National Science Foundation Grant #1808276.
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

752 TABLES

753

754 Table 1: Rate Constants for Dicamba Photodegradation and Number of Trials for Each

755 Experiment, n

Conditions	k / day ⁻¹	n
254 nm lamps, pH 7	56.2 ± 2.9	1
No UV-light	No reaction	1
310 nm lamps, pH 7	23.0 ± 2.9	6
310 nm lamps, pH 1	20.2 ± 1.4	2
310 nm lamps, O ₂ sat'd	21.6 ± 4.3	2
310 nm lamps, N ₂ sat'd	58 ± 29	2
310 nm lamps, Diablo®	21.6 ± 1.4	3
310 nm lamps, 5 mg/L DA6	33.1 ± 7.2	2
310 nm lamps, 5 mg/L DA6, IPA	24.5 ± 5.8	2
310 nm lamps, 5 mg/L DA6, histidine	20.2 ± 2.9	1
310 nm lamps, 5 mM H ₂ O ₂	93.6 ± 2.9	1
Outdoors, July 2019, Saint Peter, MN	0.04 ± 0.03	1
Q-Sun Xe lamps, pH 7	1.3 ± 0.7	2
Glass Plate	No reaction	1
Corn Wax	0.16 ± 0.01	1
Corn Wax/Diablo®	0.30 ± 0.03	1
Corn Wax/1mg/L DA-6	0.26 ± 0.03	1
Corn Wax/5 mg/L DA-6	0.32 ± 0.03	1
Corn Wax/10 mg/L DA-6	0.19 ± 0.03	1
Corn Wax/15 mg/L DA-6	0.22 ± 0.01	1

756

757 Table 2: Environmental Direct Photolysis Half-lives of dicamba calculated from GCSOLAR

Season	Latitude (°)	Integrated Half-life (days)	
		DCAM in MN River Water	DCAM in Pure Water
Spring	30	4.23	4.23
	40	4.68	4.68
	50	5.48	5.48
	60	6.83	6.83
Summer	30	3.74	3.74
	40	3.84	3.84
	50	4.08	4.08
	60	4.49	4.49
Fall	30	6.13	6.13
	40	8.21	8.21
	50	12.8	12.8
	60	25.5	25.5
Winter	30	8.63	8.63
	40	14.2	14.2
	50	30.3	30.3
	60	97.7	97.7

758

759

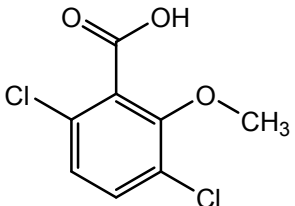
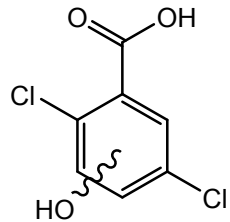
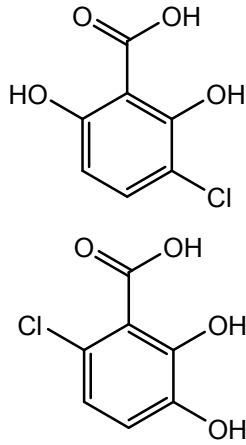
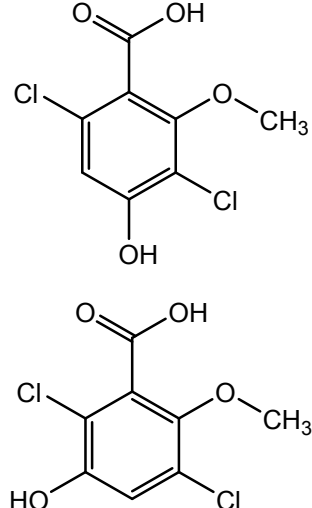
760

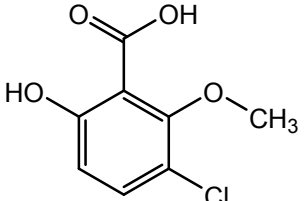
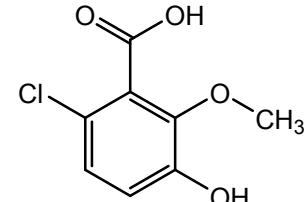
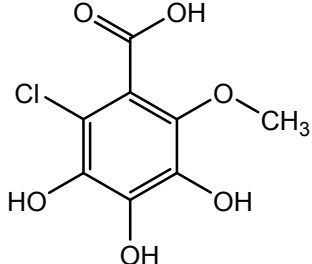
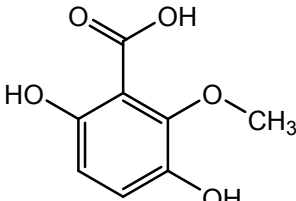
761 Table 3: Experimental (k_{obs}) and Light Screening Corrected (k_{corr}) Rate Constants for Irradiation762 of pH 7 Dicamba. $z_{\text{mix}} = 4.6$ cm and $D(\lambda) = 1$ in calculations of $S(\lambda)$.

Solution	k_{obs} (day ⁻¹)	$S(\lambda)$	k_{corr} (day ⁻¹)
Dicamba	23.0 ± 1.4		
Dicamba + 1.2 mg/L NOM	23.0 ± 1.4	0.958	23.0 ± 1.4
Dicamba + 5.2 mg/L NOM	15.8 ± 1.4	0.756	21.6 ± 1.4
Dicamba + 10.0 mg/L NOM	13.0 ± 1.4	0.619	23.0 ± 1.4
Dicamba in MN River water - 2019	8.9 ± 1.4	0.580	15.8 ± 1.4
Dicamba in MN River water - 2020	14.4 ± 1.4	0.884	15.8 ± 1.4

763

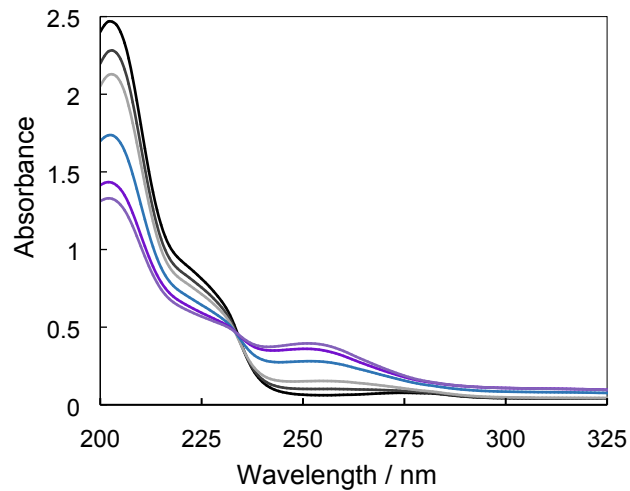
764 Table 4: Data from the LC-MS for Each Putative Photoproduct of the Photolysis of Dicamba (40
 765 mg/L) Irradiated at 310 nm for 0, 35, and 70 min

Photoproduct	Proposed Putative Photoproduct Structure	Measured Mass ESI [M-H] ⁻ m/z	Calculated [M-H] ⁻ m/z	Error (ppm)	Formula	DBE	RT / min
Dicamba		218.9629	218.9616	-2.66	C ₈ H ₆ Cl ₂ O ₃	5	3.227
A		204.9464	204.9459	0.09	C ₇ H ₄ Cl ₂ O ₃	5	2.915
B B1 – top B2 – bottom		186.9806	186.9798	-0.73	C ₇ H ₅ ClO ₄	5	2.737
C C1 – top C2 – bottom		234.9572	234.9565	-0.7	C ₈ H ₆ Cl ₂ O ₄	5	2.109

<p>D D1 – top D2 - bottom</p>	 	200.9964	200.9955	-1.45	$C_8H_7ClO_4$	5	1.919
E		232.9864	232.9853	-2.0	$C_8H_7ClO_6$	5	0.809
F		183.0301	183.0294	-1.05	$C_8H_8O_5$	5	1.795

766

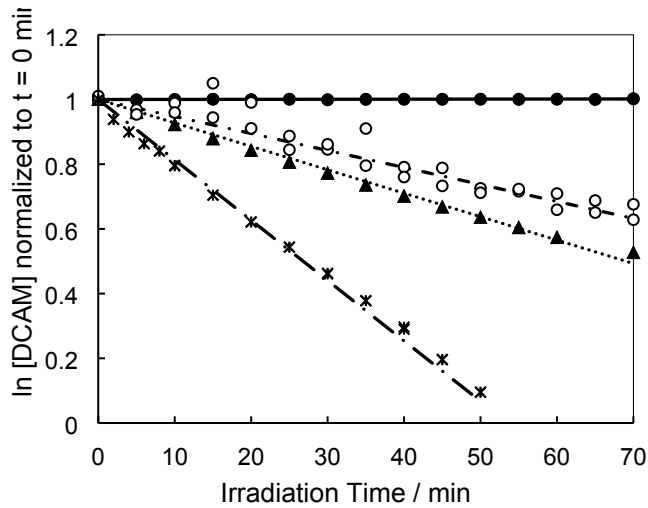
1
2
3 767 FIGURES
4 768
5



6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21 769
22 770

23 771 Figure 1: A selection of UV-Vis spectra of dicamba irradiated at 310 nm. Looking at 200 nm, the
24 772 top black line is non-irradiated dicamba and bottom purple line is dicamba irradiated for 55 min.
25 773 The lines in between were irradiated from 5-50 minutes. The peak at 203 nm corresponds to
26 774 dicamba; absorbance at that wavelength drops over the irradiation time. Another peak at 254 nm
27 775 grows in over the irradiation time. An isosbestic point can be seen near 235 nm.
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

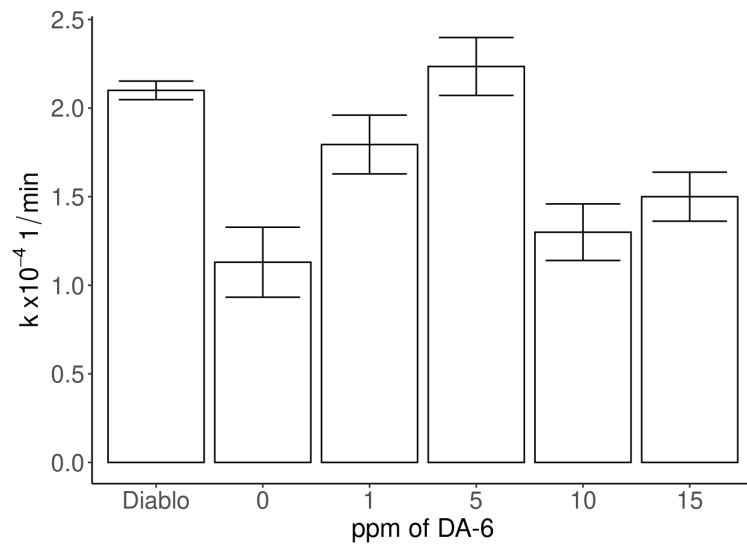
776



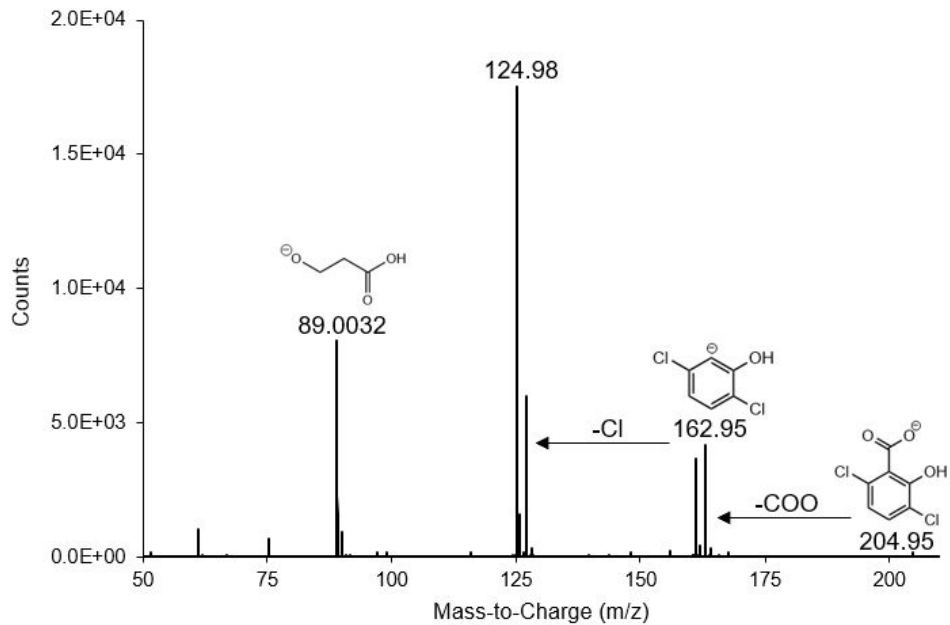
777

778 Figure 2: 15 mg/L at pH 7 dicamba kept the dark (●) and irradiated under 310 nm lamps (▲).
 779 Regression lines yield no reaction for the dark control and $k = 23.0 \pm 2.0 \text{ day}^{-1}$ for the irradiated
 780 sample. Also shown are 15 mg/L at pH 7 dicamba irradiated under 310 nm lamps in oxygen
 781 desaturated solution (○) and with 5 mM H_2O_2 (*). Regression lines give $k = 58 \pm 29 \text{ day}^{-1}$ for the
 782 experiment with oxygen desaturated solution and $93.6 \pm 2.9 \text{ day}^{-1}$ for the 5 mM H_2O_2 experiment.
 783 These rate constants show the decrease in rate due to the removal of oxygen in the oxygen
 784 desaturated solution and the increase in rate due to reaction with hydroxyl radical in the H_2O_2
 785 experiment.

786

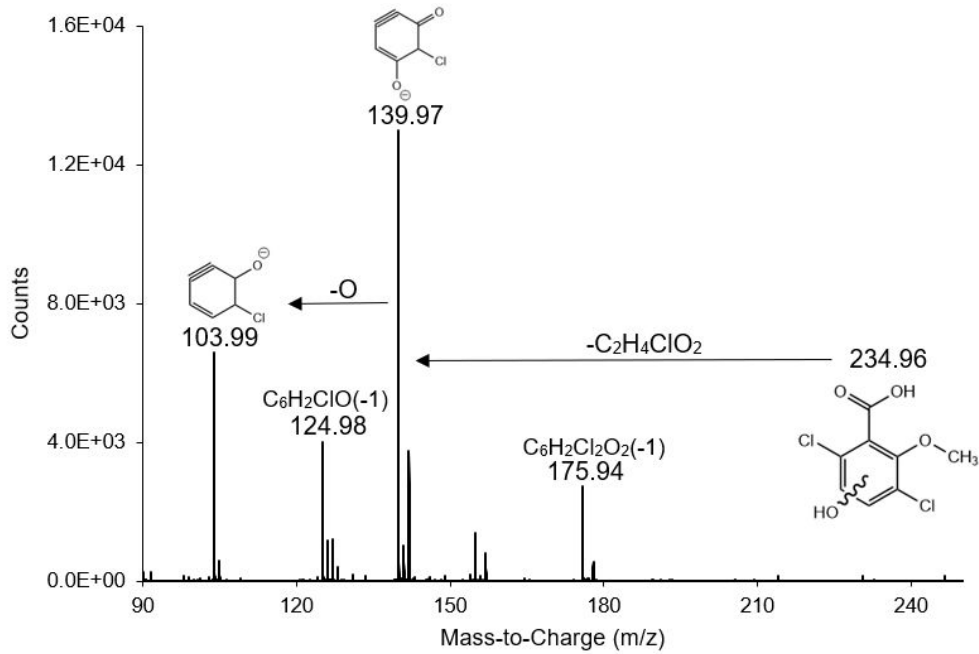


789
790
791 Figure 3: Photodegradation rate constants of dicamba the surfaces of corn wax. Included are:
792 photodegradation rate constant of Diablo[®], a commercial product of dicamba, on the surface of
793 corn wax and photodegradation rate constants of dicamba on the surface of corn wax surfaces
794 with 0 ppm adjuvant DA-6, 1 ppm DA-6, 5 ppm DA-6, 10 ppm DA-6, and 15 ppm DA-6.

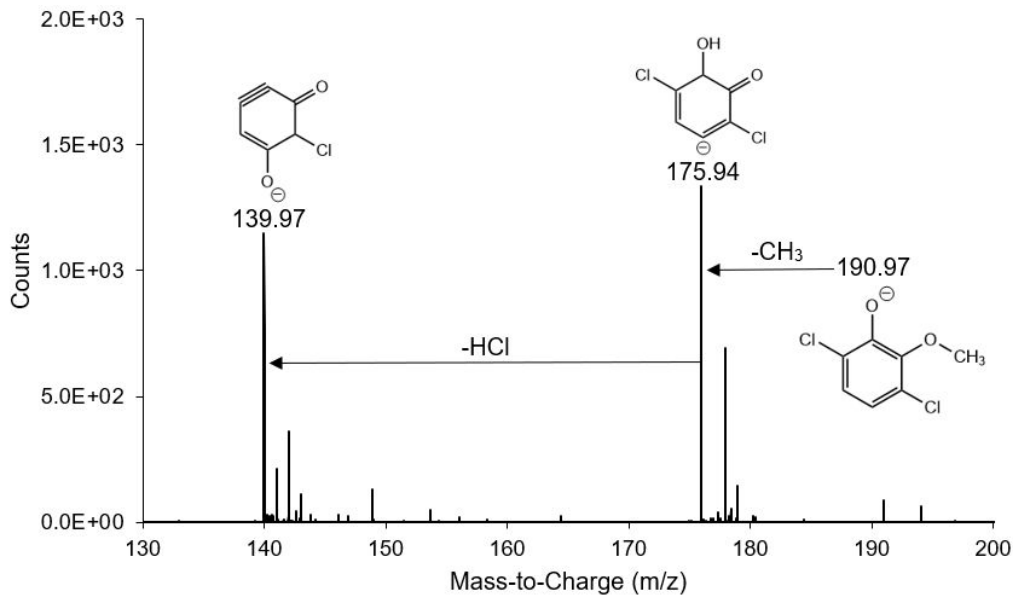


795
796
797
798
799
800
801

Figure 4. 2D MS of photoproduct A, 204.9461 m/z parent ion peak. The fragmentation pattern of the parent peak ion, 204.9461 m/z, indicates the presence the carboxylic acid group due to the 162.95 m/z fragment. The presence of two chlorine atoms was determined from the mass difference and isomer patterns between 124.98 m/z and 162.95 m/z fragments.

802
803804
805

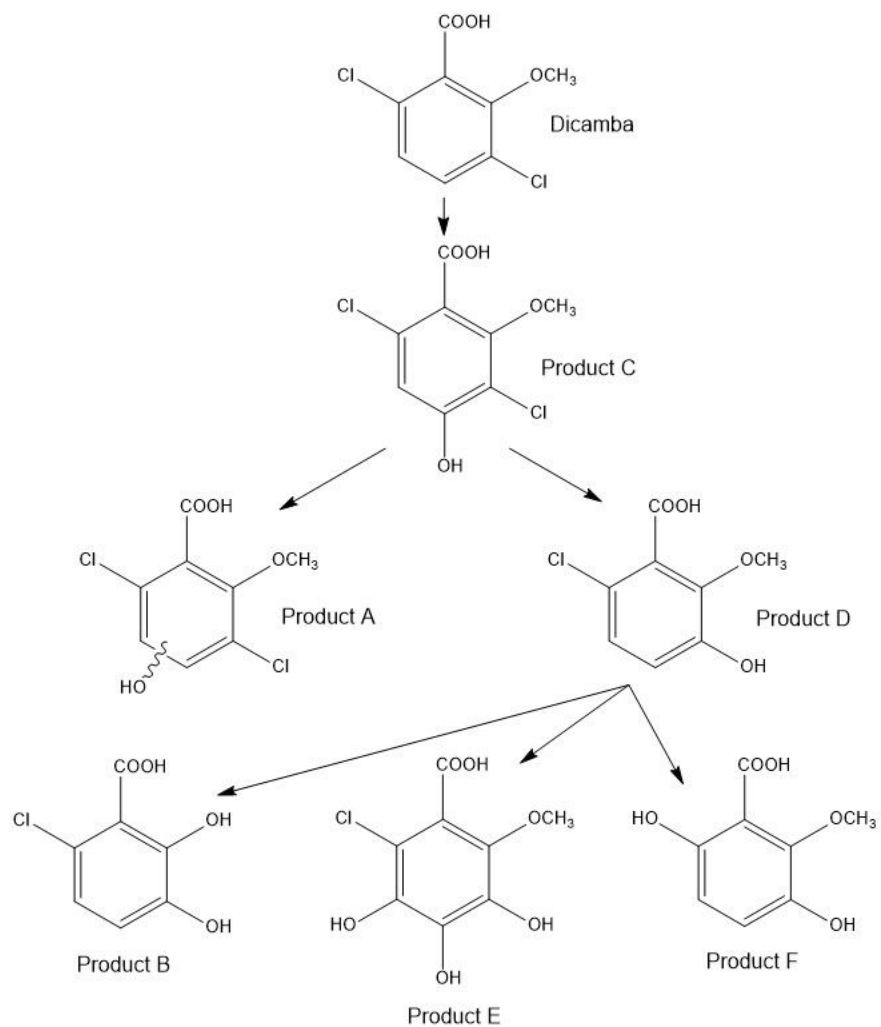
a)



b)

806
807
808
809
810
811
812
813
814
815

Figure 5. 2D MS of photoproduct C, 234.9564 m/z parent peak ion and 190.9668 m/z base peak ion. a) Fragmentation patterns in the 2D MS of the parent peak ion 234.9564 m/z suggest the presence of a carboxylic acid group, two chlorine atoms, and the ether and alcohol groups meta or para from each other. b) The fragmentation pattern in the 2D MS of the base peak ion 190.9668 m/z confirms two chlorine atoms, an ether group from the 175.94 m/z peak, and an alcohol group.



816
817

818 Figure 6. Proposed transformation pathway of photodegradation of dicamba. Proposed
819 transformation pathway was determined from LC-MS-MS data and the similar transformation
820 pathway presented in Fabbri et al.⁴ Isomers with the lowest energy as determined by DFT
821 calculations are shown in the proposed transformation pathway.

1
2
3 822 REFERENCES

- 4 823
- 5 824 1 Atwood, Donald and Paisley-Jones, Claire, *Pesticides Industry Sales and Usage: 2008-2012*
- 6 825 *Market Estimates*, U.S. Environmental Protection Agency.
- 7 826 2 J.-P. Aguer, F. Blachère, P. Boule, S. Garaudee and C. Guillard, Photolysis of dicamba (3,6-
- 8 827 dichloro-2-methoxybenzoic acid) in aqueous solution and dispersed on solid supports,
- 9 828 *International Journal of Photoenergy*, 2000, **2**, 81–86.
- 10 829 3 W. Chu and C. C. Wong, The photocatalytic degradation of dicamba in TiO₂ suspensions with
- 11 830 the help of hydrogen peroxide by different near UV irradiations, *Water Research*, 2004, **38**,
- 12 831 1037–1043.
- 13 832 4 D. Fabbri, A. B. Prevot and E. Pramauro, Analytical monitoring of the photo-induced
- 14 833 degradation of 3,6-dichloro-2-methoxybenzoic acid in homogeneous and heterogeneous
- 15 834 systems, *Research on Chemical Intermediates*, 2007, **33**, 393–405.
- 16 835 5 Behrens, Richard and Lueschen, W. E., Dicamba Volatility, *Weed Science*, 1979, **27**, 486–493.
- 17 836 6 J. F. Egan and D. A. Mortensen, Quantifying vapor drift of dicamba herbicides applied to
- 18 837 soybean, *Environmental Toxicology and Chemistry*, 2012, **31**, 1023–1031.
- 19 838 7 T. C. Mueller, D. R. Wright and K. M. Remund, Effect of Formulation and Application Time of
- 20 839 Day on Detecting Dicamba in the Air under Field Conditions, *Weed Science*, 2013, **61**, 586–
- 21 840 593.
- 22 841 8 S. D. Strachan, M. S. Casini, K. M. Heldreth, J. A. Scocas, S. J. Nissen, B. Bukun, R. B.
- 23 842 Lindenmayer, D. L. Shaner, P. Westra and G. Brunk, Vapor Movement of Synthetic Auxin
- 24 843 Herbicides: Aminocyclopyrachlor, Aminocyclopyrachlor-Methyl Ester, Dicamba, and
- 25 844 Aminopyralid, *Weed Science*, 2010, **58**, 103–108.
- 26 845 9 S. D. Strachan, N. M. Ferry and T. L. Cooper, Vapor Movement of Aminocyclopyrachlor,
- 27 846 Aminopyralid, and Dicamba in the Field, *Weed Technology*, 2013, **27**, 143–155.
- 28 847 10 Baur, J. R. and Bovey, R. W., Ultraviolet and volatility loss of herbicides, *Archives of*
- 29 848 *Environmental Contamination and Toxicology*, 1974, **2**, 275–288.
- 30 849 11 USDA NASS, <https://quickstats.nass.usda.gov/>, (accessed 17 December 2020).
- 31 850 12 Pesticide Use Data System, <https://hygeia-analytics.com/tools/puds/by-crop/>, (accessed 17
- 32 851 December 2020).
- 33 852 13 *Roundup Ready 2 Xtend Soybeans: Currently in Phase IV of Monsanto's R&D Pipeline*,
- 34 853 Monsanto.
- 35 854 14 Erickson, Britt E., *Chemical and Engineering News*, 2016, 94, 15.
- 36 855 15 *Roundup Ready 2 Xtend Soybeans Gain EU Import Approval*, Monsanto, St. Louis, MO, 2016.
- 37 856 16 H. Hatterman-Valenti, G. Endres, B. Jenks, M. Ostlie, T. Reinhardt, A. Robinson, J. Stenger and
- 38 857 R. Zollinger, Defining Glyphosate and Dicamba Drift Injury to Dry Edible Pea, Dry Edible Bean,
- 39 858 and Potato, *HortTechnology*, 2017, **27**, 502–509.
- 40 859 17 Charles, Dan, *Morning Edition, National Public Radio*, 2020.
- 41 860 18 Tong, Scott, *Marketplace, National Public Radio*, 2020.
- 42 861 19 A. M. Fogarty and O. H. Tuovinen, Microbiological degradation of the herbicide dicamba,
- 43 862 *Journal of Industrial Microbiology*, 1995, **14**, 365–370.
- 44 863 20 J. P. Krueger, R. G. Butz and D. J. Cork, Aerobic and anaerobic soil metabolism of dicamba,
- 45 864 *Journal of Agricultural and Food Chemistry*, 1991, **39**, 995–999.
- 46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

- 1
2
3 865 21 E. W. Pavel, A. R. Lopez, D. F. Berry, E. P. Smith, R. B. Reneau and S. Mostaghimi, Anaerobic
4 866 degradation of dicamba and metribuzin in riparian wetland soils, *Water Research*, 1999, **33**,
5 867 87–94.
6
7 868 22 P. W. Milligan and M. M. Häggblom, Biodegradation and Biotransformation of Dicamba
8 869 under Different Reducing Conditions, *Environmental Science & Technology*, 1999, **33**, 1224–
9 870 1229.
10
11 871 23 M. A. Rahman and M. Muneer, Heterogeneous Photocatalytic Degradation of Picloram,
12 872 Dicamba, and Floumeturon in Aqueous Suspensions of Titanium Dioxide, *Journal of*
13 873 *Environmental Science and Health, Part B*, 2005, **40**, 247–267.
14 874 24 A. Bianco-Prevot, D. Fabbri, E. Pramauro, A. Morales-Rubio and M. de la Guardia, Continuous
15 875 monitoring of photocatalytic treatments by flow injection. Degradation of dicamba in
16 876 aqueous TiO₂ dispersions, *Chemosphere*, 2001, **44**, 249–255.
17
18 877 25 E. Brillas, M. Á. Baños and J. A. Garrido, Mineralization of herbicide 3,6-dichloro-2-
19 878 methoxybenzoic acid in aqueous medium by anodic oxidation, electro-Fenton and
20 879 photoelectro-Fenton, *Electrochimica Acta*, 2003, **48**, 1697–1705.
21
22 880 26 A. Shrivastav, G. Sudarsan, P. Bose and V. Tare, Modification of Winkler’s method for
23 881 determination of dissolved oxygen concentration in small sample volumes, *Analytical*
24 882 *Methods*, 2010, **2**, 1618.
25 883 27 In *Standard Methods For the Examination of Water and Wastewater*, American Public Health
26 884 Association, Washington, D.C., 21st Edition., 2005, p. 4:138-4:140.
27
28 885 28 A. Ter Halle, D. Drncova and C. Richard, Phototransformation of the herbicide sulcotrione on
29 886 maize cuticular wax, *Environ Sci Technol Environ Sci Technol*, 2006, **40**, 2989–2995.
30 887 29 P. P. Choudhury, Leaf cuticle-assisted phototransformation of isoproturon, *Acta Physiologiae*
31 888 *Plantarum*, , DOI:10.1007/s11738-017-2471-0.
32 889 30 P. Choudhury, Leaf-cutin assisted phototransformation of 2,4-D ethyl ester, *Indian Journal of*
33 890 *Biochemistry and Biophysics*, 2016, **53**, 227–231.
34
35 891 31 S. C. Anderson, A. Christiansen, A. Peterson, L. Beukelman and A. M. Nienow, Statistical
36 892 analysis of the photodegradation of imazethapyr on the surface of extracted soybean
37 893 (Glycine max) and corn (Zea mays) epicuticular waxes, *Environ. Sci.: Processes Impacts*, 2016,
38 894 **18**, 1305–1315.
39
40 895 32 *Dicamba Herbicide Application Notes*, Drexel Chemical Company.
41 896 33 A. Leifer, *The Kinetics of Environmental Aquatic Photochemistry: Theory and Practice*,
42 897 American Chemical Society, 1988.
43
44 898 34 J. N. Apell and K. McNeill, Updated and validated solar irradiance reference spectra for
45 899 estimating environmental photodegradation rates, *Environmental Science: Processes &*
46 900 *Impacts*, 2019, **21**, 427–437.
47
48 901 35 R. G. Zepp and D. M. Cline, Rates of direct photolysis in aquatic environment, *Environmental*
49 902 *Science & Technology*, 1977, **11**, 359–366.
50 903 36 M. J. Frisch, G. W. Trucks, H. B. Schlegel, G. E. Scuseria, M. A. Robb, J. R. Cheeseman, G.
51 904 Scalmani, V. Barone, B. Mennucci, G. A. Petersson, H. Nakatsuji, M. Caricato, X. Li, H. P.
52 905 Hratchian, A. F. Izmaylov, J. Bloino, G. Zheng, J. L. Sonnenberg, M. Hada, M. Ehara, K. Toyota,
53 906 R. Fukuda, J. Hasegawa, M. Ishida, T. Nakajima, Y. Honda, O. Kitao, H. Nakai, T. Vreven, J. A.
54 907 Montgomery, Jr., J. E. Peralta, F. Ogliaro, M. Bearpark, J. J. Heyd, E. Brothers, K. N. Kudin, V.
55 908 N. Staroverov, R. Kobayashi, J. Normand, K. Raghavachari, A. Rendell, J. C. Burant, S. S.

- 1
2
3 909 Iyengar, J. Tomasi, M. Cossi, N. Rega, J. M. Millam, M. Klene, J. E. Knox, J. B. Cross, V. Bakken,
4 910 C. Adamo, J. Jaramillo, R. Gomperts, R. E. Stratmann, O. Yazyev, A. J. Austin, R. Cammi, C.
5 911 Pomelli, J. W. Ochterski, R. L. Martin, K. Morokuma, V. G. Zakrzewski, G. A. Voth, P. Salvador,
6 912 J. J. Dannenberg, S. Dapprich, A. D. Daniels, Ö. Farkas, J. B. Foresman, J. V. Ortiz, J. Cioslowski,
7 913 and D. J. Fox, *Gaussian 09*, Gaussian, Inc., Wallingford, CT, 2009.
- 8
9 914 37 Green, J.R. and Margerison, D., *Statistical Treatment of Experimental Data*, Elsevier Scientific
10 915 Publishing Company, New York, 1977.
- 11
12 916 38 R. Espy, E. Pelton, A. Opseth, J. Kasprisin and A. M. Nienow, Photodegradation of the
13 917 Herbicide Imazethapyr in Aqueous Solution: Effects of Wavelength, pH, and Natural Organic
14 918 Matter (NOM) and Analysis of Photoproducts, *J Agr Food Chem*, 2011, **59**, 7277–7285.
- 15
16 919 39 H. Barkani, C. Catastini, C. Emmelin, M. Sarakha, A. El Azzouzi and J. M. Chovelon, Study of
17 920 the phototransformation of imazaquin in aqueous solution: a kinetic approach, *J Photoch*
18 921 *Photobio A J Photoch Photobio A*, 2005, **170**, 27–35.
- 19
20 922 40 D. Edwards, *Reregistration Eligibility Decision for Dicamba and Associated Salts*, United
21 923 States Environmental Protection Agency.
- 22
23 924 41 R. Tafer, P. de Sainte-Claire, P. Vicendo, A. Boulkamh and C. Richard, Photochemistry of 5-
24 925 Halogenosalicylic Acids: Evidence of the Triplet Involvement in the Carbene Formation,
25 926 *ChemistrySelect*, 2016, **1**, 4704–4712.
- 26
27 927 42 H. He, D. Xiong, F. Han, Z. Xu, B. Huang and X. Pan, Dissolved oxygen inhibits the promotion
28 928 of chlorothalonil photodegradation mediated by humic acid, *Journal of Photochemistry and*
29 929 *Photobiology A: Chemistry*, 2018, **360**, 289–297.
- 30
31 930 43 F. S. Tanaka, R. G. Wien and E. R. Mansager, Effect of nonionic surfactants on the
32 931 photochemistry of 3-(4-chlorophenyl)-1,1-dimethylurea in aqueous solution, *Journal of*
33 932 *Agricultural and Food Chemistry*, 1979, **27**, 774–779.
- 34
35 933 44 F. S. Tanaka, R. G. Wien and E. R. Mansager, Survey for surfactant effects on the
36 934 photodegradation of herbicides in aqueous media, *Journal of Agricultural and Food*
37 935 *Chemistry*, 1981, **29**, 227–230.
- 38
39 936 45 F. S. Tanaka, R. G. Wien and B. L. Hoffer, Photosensitized degradation of a homogeneous
40 937 nonionic surfactant: hexaethoxylated 2,6,8-trimethyl-4-nonanol, *Journal of Agricultural and*
41 938 *Food Chemistry*, 1986, **34**, 547–551.
- 42
43 939 46 Wei. Chu and C. T. Jafvert, Photodechlorination of Polychlorobenzene Congeners in
44 940 Surfactant Micelle Solutions, *Environmental Science & Technology*, 1994, **28**, 2415–2422.
- 45
46 941 47 F. S. Tanaka, R. G. Wien and R. G. Zaylskie, Photolytic degradation of a homogeneous Triton X
47 942 nonionic surfactant: nonaethoxylated p-(1,1,3,3-tetramethylbutyl)phenol, *Journal of*
48 943 *Agricultural and Food Chemistry*, 1991, **39**, 2046–2052.
- 49
50 944 48 Z. Shi, M. E. Sigman, M. M. Ghosh and R. Dabestani, Photolysis of 2-Chlorophenol Dissolved
51 945 in Surfactant Solutions, *Environmental Science & Technology*, 1997, **31**, 3581–3587.
- 52
53 946 49 K. Huang, G. Lu, Z. Zheng, R. Wang, T. Tang, X. Tao, R. Cai, Z. Dang, P. Wu and H. Yin,
54 947 Photodegradation of 2,4,4'-tribrominated diphenyl ether in various surfactant solutions:
55 948 kinetics, mechanisms and intermediates, *Environmental Science: Processes & Impacts*, 2018,
56 949 **20**, 806–812.
- 57
58 950 50 X. Li, J. Huang, G. Yu and S. Deng, Photodestruction of BDE-99 in micellar solutions of
59 951 nonionic surfactants of Brij 35 and Brij 58, *Chemosphere*, 2010, **78**, 752–759.
- 60

- 1
2
3 952 51P. Mukerjee and K. Mysels, *Critical Micelle Concentrations of Aqueous Surfactant Systems*,
4 953 Office of Standard Reference Data, National Bureau of Standards, 1971.
- 5 954 52K. L. Armbrust, Pesticide hydroxyl radical rate constants: Measurements and estimates of
6 955 their importance in aquatic environments, *Environ Toxicol Chem*, 2000, **19**, 2175–2180.
- 7 956 53X. Jin, S. Peldszus and P. M. Huck, Reaction kinetics of selected micropollutants in ozonation
8 957 and advanced oxidation processes, *Water Research*, 2012, **46**, 6519–6530.
- 9 958 54V. Riffault, T. Gierczak, J. B. Burkholder and A. R. Ravishankara, Quantum yields for OH
10 959 production in the photodissociation of HNO₃ at 248 and 308 nm and H₂O₂ at 308 and 320
11 960 nm, *Physical Chemistry Chemical Physics*, 2006, **8**, 1079.
- 12 961 55J. Michalowski, P. Halaburda and A. Kojlo, Determination of humic acid in natural waters by
13 962 flow injection analysis with chemiluminescence detection, *Analytica Chimica Acta*, 2001,
14 963 143–148.
- 15 964 56J. J. Werner, K. McNeill and W. A. Arnold, Environmental photodegradation of mefenamic
16 965 acid, *Chemosphere*, 2005, **58**, 1339–1346.
- 17 966 57S. Halladja, A. Amine-Khodja, A. ter Halle, A. Boulkamh and C. Richard, Photolysis of
18 967 fluometuron in the presence of natural water constituents, *Chemosphere*, 2007, **69**, 1647–
19 968 1654.
- 20 969 58R. P. Schwarzenbach, P. M. Gschwend and D. M. Imboden, *Environmental Organic Chemistry*,
21 970 John Wiley and Sons, Hoboken, Second Edition., 2003.
- 22 971 59M. Elazzouzi, M. Mekkaoui, S. Zaza, M. El Madani, A. Zrineh and J. M. Chovelon, Abiotic
23 972 degradation of imazethapyr in aqueous solution, *Journal of Environmental Science and*
24 973 *Health, Part B: Pesticides, Food Contaminants, and Agricultural Wastes*, 2002, **B37**, 445–451.
- 25 974 60M. Ramezani, D. P. Oliver, R. S. Kookana, G. Gill and Christopher. Preston, Abiotic
26 975 degradation (photodegradation and hydrolysis) of imidazolinone herbicides., *J. Environ. Sci.*
27 976 *Health, Part B*, 2008, **43**, 105–112.
- 28 977 61A. Latifoglu and M. D. Gurol, The effect of humic acids on nitrobenzene oxidation by
29 978 ozonation and O₃/UV processes, *Water Res*, 2003, **37**, 1879–1889.
- 30 979 62M. Elazzouzi, A. Bensaoud, A. Bouhaouss, S. Guittonneau, A. Dahchour, P. Meallier and A.
31 980 Piccolo, Photodegradation of imazapyr in the presence of humic substances, *Fresenius*
32 981 *Environmental Bulletin*, 1999, **8**, 478–485.
- 33 982 63 *Soil Survey of Nicollet County, Minnesota*, United States Department of Agriculture, Soil
34 983 Conservation Service, 1994.
- 35 984 64O. Sakaliene, S. K. Papiernik, W. C. Koskinen and K. A. Spokas, Sorption and predicted
36 985 mobility of herbicides in Baltic soils, *Journal of Environmental Science and Health, Part B:*
37 986 *Pesticides, Food Contaminants, and Agricultural Wastes*, 2007, **42**, 641–647.
- 38 987 65Johnson, R.M. and Sims, J.T., Influence of surface and subsoil properties on herbicide
39 988 sorption by Atlantic Coastal Plain soils, *Soil Science*, **155**, 339–348.
- 40 989 66M. Sleiman, P. de Sainte Claire and C. Richard, Heterogeneous Photochemistry of
41 990 Agrochemicals at the Leaf Surface: A Case Study of Plant Activator Acibenzolar- S -methyl,
42 991 *Journal of Agricultural and Food Chemistry*, 2017, **65**, 7653–7660.
- 43 992 67A. S. Trivella, S. Monadjemi, D. R. Worrall, I. Kirkpatrick, E. Arzoumanian and Claire. Richard,
44 993 Perinaphthenone phototransformation in a model of leaf epicuticular waxes., *J. Photochem.*
45 994 *Photobiol., B*, 2014, **130**, 93–101.
- 46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

- 1
2
3 995 68S. Monadjemi, A. ter Halle and Claire. Richard, Reactivity of cycloxydim toward singlet oxygen
4 996 in solution and on wax film., *Chemosphere*, 2012, **89**, 269–273.
5
6 997 69L. Su, J. D. Sivey and N. Dai, Emerging investigator series: sunlight photolysis of 2,4-D
7 998 herbicides in systems simulating leaf surfaces, *Environmental Science: Processes & Impacts*,
8 999 2018, **20**, 1123–1135.
9
10 1000 70N. Schippers and W. Schwack, Photochemistry of imidacloprid in model systems, *J. Agric.*
11 1001 *Food Chem.*, 2008, **56**, 8023–8029.
12 1002 71Bianchi, G., Avato, P., and Salamini, F., Surface waxes from grain, leaves, and husks of maize
13 1003 (*Zea mays* L.), *Cereal Chemistry*, 1984, **61**, 45–47.
14 1004 72K. S. Kim, S. H. Park, D. K. Kim and M. A. Jenks, Influence of Water Deficit on Leaf Cuticular
15 1005 Waxes of Soybean (*Glycine max* [L.] Merr.), *International Journal of Plant Sciences*, 2007,
16 1006 **168**, 307–316.
17
18 1007 73P. Cabras, A. Angioni, V. L. Garau, M. Melis, F. M. Pirisi and E. V. Minelli, Effect of epicuticular
19 1008 waxes of fruits on the photodegradation of fenthion, *J Agr Food Chem J Agr Food Chem*,
20 1009 1997, **45**, 3681–3683.
21
22 1010 74D. Lavieille, H. A. Ter, P.-O. Bussiere and C. Richard, Effect of a spreading adjuvant on
23 1011 mesotrione photolysis on wax films, *J Agric Food Chem*, 2009, **57**, 9624–8.
24 1012 75E. L. Schymanski, J. Jeon, R. Gulde, K. Fenner, M. Ruff, H. P. Singer and J. Hollender,
25 1013 Identifying Small Molecules via High Resolution Mass Spectrometry: Communicating
26 1014 Confidence, *Environ. Sci. Technol.*, 2014, **48**, 2097–2098.
27
28 1015
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60