

This is an *Accepted Manuscript*, which has been through the Royal Society of Chemistry peer review process and has been accepted for publication.

Accepted Manuscripts are published online shortly after acceptance, before technical editing, formatting and proof reading. Using this free service, authors can make their results available to the community, in citable form, before we publish the edited article. This Accepted Manuscript will be replaced by the edited, formatted and paginated article as soon as this is available.

You can find more information about *Accepted Manuscripts* in the **Information for Authors**.

Please note that technical editing may introduce minor changes to the text and/or graphics, which may alter content. The journal's standard <u>Terms & Conditions</u> and the <u>Ethical guidelines</u> still apply. In no event shall the Royal Society of Chemistry be held responsible for any errors or omissions in this *Accepted Manuscript* or any consequences arising from the use of any information it contains.



www.rsc.org/advances

1	
2	
3	
4	Chemical control of struvite scale by a green inhibitor
5	polyaspartic acid
6	
7	Han Li <sup>1</sup> , Sheng-Hui Yu <sup>1</sup> , Qi-Zhi Yao <sup>2</sup> , Gen-Tao Zhou <sup>1*</sup> , Sheng-Quan Fu <sup>3</sup>
8	
9	<sup>1</sup> CAS Key Laboratory of Crust-Mantle Materials and Environments, School of Earth
10	and Space Sciences, University of Science and Technology of China, Hefei 230026, P.
11	R. China.
12	<sup>2</sup> School of Chemistry and Materials Science, University of Science and Technology
13	of China, Hefei 230026, P. R. China.
14	<sup>3</sup> Hefei National Laboratory for Physical Sciences at Microscale, University of
15	Science and Technology of China, Hefei 230026, P. R. China.
16	
17	
18	
19	• Corresponding author: Prof. Dr. Gen-Tao Zhou
20	Email: <u>gtzhou@ustc.edu.cn</u>
21	Tel.: 86 551 63600533
22	Fax: 86 551 63600533
23	

Abstract Many efforts have been made to develop effective chemical inhibitors for struvite scale, which causes a range of operational problems in wastewater treatment industry. Here, the inhibitory capacity of polyaspartic acid (PASP) on the spontaneous precipitation of struvite at pH 9 was investigated. The struvite precipitates were characterized by X-ray diffraction (XRD), field emission scanning electron microscopy (FESEM), and energy dispersive X-ray spectroscopy (EDX). The precipitation experiments dosed with PASP unveiled that PASP is effective in growth inhibition of struvite and its inhibitory capacity is proportional to its concentration, and that PASP also plays a role in the morphological modification of struvite crystals. The effect of several key parameters, including pH, mixing energy, reaction time, and calcium ions on PASP inhibition performance was examined for potentially practical application. The results showed that the inhibitory capacity of PASP is sustainable and efficient. The dissolution experiments dosed with PASP were also performed, and the results showed that PASP can accelerate the dissolution of the preformed struvite, and the capacity increases with its concentration. Therefore, PASP can potentially act as a feasible and environmentally-friendly inhibitor and cleaning agent for struvite scale.

19

1

2

3

4

5

6

7

8

9

10

11

12

13

14

15

16

17

18

20 Keywords: Struvite; Ployaspartic acid (PASP); Scale inhibition; Dissolution;

21 Complexation

22

**RSC Advances Accepted Manuscript** 

# 1 1. Introduction

2

The formation of struvite deposits in wastewater treatment plants (WWTPs) has 3 been widely reported since 1939 when it was first identified in the digested sludge 4 supernatant lines.<sup>1</sup> Struvite, known as magnesium ammonium phosphate hexahydrate 5 (MgNH<sub>4</sub>PO<sub>4</sub> 6H<sub>2</sub>O), crystallizes in the orthorhombic system and adopts a number of 6 natural morphologies including equant, wedge-shaped, short prismatic, and tabular 7 forms.<sup>2</sup> When the concentrations of magnesium, ammonium and phosphate ions 8 9 exceed struvite solubility product, its precipitation occurs. The wastewater usually contains large amounts of phosphorus and nitrogen, and about 80 % of N and 50 % of 10 P originate from urine.<sup>3</sup> The anaerobic digestion further solubilizes organic-P and N to 11  $PO_4^{3-}$  and  $NH_4^+$ , favoring the formation of struvite.<sup>4-5</sup> Struvite crystals tend to form 12 hard scale on process equipment surfaces of WWTPs, such as sludge liquors pipes, 13 pumps, centrifuges and aerators, leading to clogging and breakdowns of these 14 equipment.<sup>1, 6-7</sup> 15

16 In order to eliminate the nuisance, many efforts have been made to deal with the 17 formation of struvite based on four principal approaches. In the first approach, the existing struvite scale is removed by acid washing or chipped away manually with a 18 hammer and a chisel.<sup>8</sup> Obviously, it is a time and manpower consuming maintenance 19 project. The second approach aims at reducing the potential of struvite precipitation 20 21 by lowering supersaturation level. Initially, the digested sludge stream holding high supersaturation is diluted with secondary effluent to reduce the supersaturation, but 22 the mitigation of struvite formation is proved to be limited in practice.<sup>9-10</sup> Recently. 23 chemical dosing of iron (III) salts is employed as a common method to remove the 24 phosphorus in wastewater.<sup>1, 11</sup> However, this process can produce large amounts of 25 sludge and has poor removal ability.<sup>4, 12</sup> Chemical inhibitors, such as 26 ethylenediaminetetraacetic acid (EDTA), sodium polyphosphate, and silicates, have 27 been also used to reduce the magnesium concentration by forming chelates or less 28 soluble substance with magnesium ions in solution.<sup>1,11</sup> Among these inhibitors, EDTA 29

was tested to be quite efficient but usually degrades slowly in the environment.<sup>11, 13-15</sup> 1 This has raised environmental concerns about its role in heavy metal mobilization in 2 groundwater.<sup>13, 16</sup> For the third approach, the inhibition was achieved by selective 3 binding of the inhibitor molecules with some specific crystal faces of struvite, hence 4 decreasing struvite growth rates. For example, Wierzbicki et al.<sup>17</sup> found that 5 phosphocitrate preferentially binding to (101) faces of struvite leads to morphological 6 modification or total growth cessation of struvite when sufficient phosphocitrate is 7 8 used. The fourth strategy is to encourage struvite precipitation by adding MgCl<sub>2</sub> and NaOH into a specific reactor, as the chemical composition and the pH of wastewaters 9 determine the potential of struvite precipitation.<sup>1, 18-19</sup> In this way, the spontaneous 10 precipitation of struvite can be prevented in the WWTPs. 11

Among these approaches, the fourth has been a research focus in recent years 12 since this crystallization technique was also regarded as an effective way to recover 13 phosphorus fertilizer.<sup>20-21</sup> However, the disadvantages for this technique are apparent 14 including : (1) Except for the supernatant of digested sludge containing relatively high 15 16 concentrations of ammonium and phosphate, many wastewater processes cannot fulfill the basic requirement for struvite precipitation.<sup>22</sup> (2) The process has never 17 been proven commercially profitable due to supplementation of magnesium salt and 18 sodium hydroxide.<sup>20, 22</sup> (3) The incorporation of toxic heavy metals, metalloids (e.g., 19 Cr, Zn, and As) and pathogens into struvite crystals can act potentially as a source of 20 pollution when struvite is used as fertilizer.<sup>23-28</sup> (4) In sludge liquors, calcium ion 21 levels can be high relative to magnesium.<sup>29</sup> Calcium ions can interact with phosphate 22 23 to form hydroxylapatite or amorphous calcium phosphates, leading to the inhibition of struvite production and impurity of the recovered product.<sup>20, 30</sup> (5) Undesired struvite 24 fine particles are often generated in the crystallization processes due to the high 25 mixing energy (or turbulence) needed to suspend the growing particles, resulting in a 26 loss of struvite particles for phosphorus recovery.<sup>31</sup> (6) The fertilization efficiency of 27 struvite is not superior to other phosphate based compounds, such as monocalcium 28 phosphate (Ca(H<sub>2</sub>PO<sub>4</sub>)<sub>2</sub>·H<sub>2</sub>O) and dicalcium phosphate (CaHPO<sub>4</sub>·2H<sub>2</sub>O).<sup>32</sup> Hence, 29 Hao et al.<sup>32</sup> pointed out that phosphate recovery should not just focus on struvite. 30

Page 5 of 24

#### **RSC Advances**

1 In view of the harm of struvite scale and the disadvantages of struvite 2 crystallization recovery technique, it is still a challenge to search for effective chemical inhibitors and develop new antiscale techniques. Normally, wastewaters 3 tend to be deficient in magnesium ions.<sup>22</sup> So sequestering magnesium ions will be an 4 ideal way to prevent struvite precipitation. Polyaspartic acid (PASP) is a synthetic 5 polyamino acid and rich in carboxylic functional groups that can combine with metal 6 ions to form metal-PASP species. <sup>16, 33</sup> PASP has been reported to be an effective 7 inhibitor of calcium carbonate, calcium oxalate and calcium phosphate.<sup>34-36</sup> Moreover. 8 PASP is water-soluble, nontoxic, biocompatible, and highly biodegradable, so it is 9 regarded as the most promising green scale inhibitor in industry.<sup>37-38</sup> To the best of our 10 knowledge, however, no research has reported on its scale inhibition to struvite. In the 11 present work, we investigated the inhibitory capacity of PASP on struvite in a 12 dynamic environment. The influence of several physicochemical parameters, 13 14 including pH, mixing energy, reaction time, and the presence of calcium ions on PASP inhibition performance was systematically investigated. The potential of PASP to 15 16 dissolve existing struvite scale was also assessed.

17

## 18 **2. Materials and methods**

19

#### 20 **2.1. Materials**

Magnesium chloride hexahydrate (MgCl<sub>2</sub>·6H<sub>2</sub>O), ammonium dihydrogen phosphate (NH<sub>4</sub>H<sub>2</sub>PO<sub>4</sub>), ammonium chloride (NH<sub>4</sub>Cl), and sodium hydroxide (NaOH) were purchased from Sinopharm Chemical Reagent Co., Ltd, and are of analytical grade. The analytical grade PASP was from Chengdu Ai Keda Chemical Technology Co., Ltd, and the molecular weight is 10000. Deionized water was used in all of the experiments.

27

#### 28 **2.2. Struvite crystallization**

All experiments were conducted at room temperature. A synthetic sludge liquor

was prepared as described by Doyle et al.<sup>39</sup> Concentrations of 4 mM of magnesium 1 and phosphate were used. As ammonium is always in excess relative to magnesium 2 and phosphate in sludge liquor,<sup>22</sup> ammonium chloride was added to produce an excess 3 of ammonium ions. In a typical synthesis procedure, 0.05 g (0.005 mmol) of PASP 4 5 and 0.0407 g (0.2 mmol) of MgCl<sub>2</sub>·6H<sub>2</sub>O were dosed to 40 ml of deionized water in a 50-mL beaker under vigorous stirring to form homogeneous solution A. Then, 0.023 g 6 (0.2 mmol) of NH<sub>4</sub>H<sub>2</sub>PO<sub>4</sub> and 0.0214 g (0.6 mmol) of NH<sub>4</sub>Cl were dissolved in 10 ml 7 8 of deionized water to form solution B. Solution B was introduced into solution A 9 under continuous stirring, and a homogeneous synthetic liquor was obtained, with a molar ratio of 1:1:3 ( $Mg^{2+}:PO_4^{3-}:NH_4^+$ ). Afterwards, the pH of the liquor was adjusted 10 to 9.0 by the addition of 0.5 M NaOH. The beaker was then covered with parafilm to 11 reduce CO<sub>2</sub> interference and NH<sub>3</sub> volatilization, and stirred for 8 h at 360 rpm on a 12 magnetic stirrer. Finally, the product was isolated by centrifugation (1400 g for 3 min), 13 14 washed with absolute alcohol three times, and dried in vacuum at room temperature for 48 h. The same procedures were employed to study the effect of PASP 15 16 concentration, reaction time, mixing energy (stirring speed), pH, and Mg/Ca ratio on 17 struvite formation. All trials were conducted in triplicate in our experiments.

18

#### 19 2.3. Struvite dissolution

20 The struvite used for the dissolution was synthesized via homogeneous precipitation by dissolving 0.508 g of MgCl<sub>2</sub>·6H<sub>2</sub>O and 0.288 g of NH<sub>4</sub>H<sub>2</sub>PO<sub>4</sub> in a 21 50-mL beaker containing 50 ml of deionized water under vigorous stirring. The 22 concentrations of  $Mg^{2+}$ ,  $PO_4^{3-}$ , and  $NH_4^+$  were 50 mM. The pH of this solution was 23 adjusted to 8.0 using 0.25 M NaOH. The beaker was then covered with parafilm and 24 25 kept static. After aging for 4 h, the product was harvested by the similar way 26 described in Section 2.2. The XRD analysis confirmed that the obtained product was pure struvite (data not shown), and the SEM observation revealed that the struvite 27 crystals exhibited coffin-like shape. The dissolution experiments were conducted at 28 room temperature. 0.05 g of PASP was dissolved to 50 ml of deionized water in a 29 50-mL beaker, and then the pH was adjusted to 8.0 by using of 0.25 M NaOH. After 30

the addition of 50 mg of struvite powder, the beaker was covered with parafilm and stirred for 30 min at 360 rpm on a magnetic stirrer. Finally, the remaining precipitate was collected by the way described in Section 2.2. The same procedures were employed in the PASP concentration-dependent dissolution experiments. All trials were also conducted in triplicate.

6

#### 7 2.4. Analytical techniques

8 X-ray diffraction pattern (XRD) was recorded on an X-ray diffractometer 9 equipped with Cu K $\alpha$  irradiation ( $\lambda = 0.154056$  nm, Japan, MapAHF). The 10 morphology and size of the precipitate particles were observed by a field emission 11 scanning electron microscope (FESEM, JEOL JSM-6700F). Energy dispersive X-ray 12 spectroscopy (EDX) analyses of the samples coated with Au were obtained with an 13 EDAX detector installed on the same FESEM.

14

# 15 **3. Results and discussion**

16

# 17 **3.1. Effect of PASP concentration on struvite formation**

In order to understand the effect of PASP on struvite crystallization and growth, a 18 series of experiments with the PASP concentration from 0.0 to 0.8 mM were first 19 conducted. The XRD results of the precipitation products confirmed that all of the 20 precipitates obtained with and without PASP are orthorhombic struvite with space 21 group Pmn2<sub>1</sub> (JCPDS file of No.15-0762), and the representative XRD patterns are 22 shown in Figure 1a and b. The FESEM results of the struvite precipitates are depicted 23 24 in Figure 2. It can be seen from Figure 2a that a large number of rod-shaped crystals were obtained in the absence of PASP, with a length of ca. 30  $\mu$ m and a width of ca. 25 10 µm. When 0.01 mM of PASP was dosed, the product turned out to be 26 arrowhead-shaped with a reduced length of ca. 5 µm (Figure 2b), which is 27 significantly different from the rod-shaped structure. This morphology was held as 28 PASP concentration was further increased to 0.05 and 0.1 mM. However, as the PASP 29

**RSC Advances Accepted Manuscript** 

concentration reached 0.3 mM, the panoramic FESEM image shows that massive 1 2 trapezoidal crystals with a length of ca. 7 µm and less rod-like crystals coexisted (Figure 2c). The further magnified image shows that the "rods" are also trapezoidal 3 (inset in Figure 2c). As PASP concentration raised to 0.5 mM, the products took on 4 5 three shapes, mainly including trapezoid (ca. 7  $\mu$ m) and equilateral triangle (ca. 5  $\mu$ m), and a few rhomboids (ca. 14 µm) (Figure 2d). Further increasing PASP concentration 6 to 0.6 or 0.7 mM did not result in much more change in morphology. In particular, no 7 precipitate could be formed when 0.8 mM of PASP was used. It seems that PASP 8 9 significantly impacts not only on struvite formation but also on its morphogenesis. Figure 1c quantitively depicts the effect of PASP on struvite precipitate. Figure 1c 10 showed that the struvite mass dramatically decreased with the increase in PASP 11 concentration except for 0.01 mM, indicating that the strong inhibition occurred, and 12 the complete inhibition was achieved with the dose of 0.8 mM PASP. Dovle et al.<sup>11</sup> 13 14 tested the effectiveness of EDTA on preventing struvite precipitation. According to their study, as much as 8 mM of EDTA is needed to achieve the complete inhibition of 15 16 struvite precipitation. In this regard, PASP is much more effective than EDTA. Moreover, the size of struvite crystals obtained in the presence of PASP was markedly 17 reduced. This will make struvite easier to be washed away and harder to scale. 18

19 It is well known that PASP is a good chelating agent, and can coordinate with a 20 variety of metal cations to form complex species.<sup>16, 33</sup> Therefore, its coordination with 21  $Mg^{2+}$  can be expressed as follows:

$$Mg^{2'} + nPASP \leftrightarrows Mg(PASP)_n$$
 (1)

23

$$Mg^{2+} + NH_4^+ + H_2PO_4^- + 6H_2O \leftrightarrows MgNH_4PO_4 \cdot 6H_2O \downarrow + 2H^+$$
(2)

where the coordination and precipitation reactions (1) and (2) compete with each other, and the precipitation will predominate with raising pH based on the reaction (2). With the precipitation of struvite, free Mg<sup>2+</sup> ions will decrease and the coordination equilibrium will shift towards disassociation of Mg-PASP complexes, leading to the increase of free PASP. The released PASP molecules can selectively bind to crystal faces of struvite, and thus affect the crystal growth habit (e.g., Figure 2). Therefore, PASP molecules play the dual roles during the scale inhibition: morphological Page 9 of 24

#### **RSC** Advances

modification and precipitation inhibition of struvite. In fact, the similar 1 arrowhead-shaped struvite has been observed in the presence of PASP in our previous 2 biomimetic mineralization experiments,<sup>40</sup> and the selective recognition and binding of 3 PASP onto {010} and {101} faces of struvite crystals were believed to be responsible 4 5 for the formation of the specific morphology. Because these faces have a high density of magnesium cations, and hence provide a positively charged environment for the 6 preferential adsorption and binding of PASP molecules rich in negatively charged 7 carboxyl side chains.<sup>17, 40-41</sup> PASP molecules binding to these faces decrease their 8 growth rate and lead to an enhanced expression of these faces, thus modifying crystal 9 10 morphology. Similarly, trapezoid and equilateral triangle observed in our case may also result from the preferential adsorption and binding of PASP onto some specific 11 faces of struvite crystals. As for the increase of struvite precipitate at 0.01 mM of 12 PASP, it is possible because less PASP molecules in the liquor cannot cause a 13 significant decrease in free  $Mg^{2+}$  by their complexation with  $Mg^{2+}$  ions. In contrast, 14 the low concentration of PASP can induce  $Mg^{2+}$  and  $NH_4^+$  aggregation and 15 16 accumulation around them by their carboxylic groups, and promote struvite nucleation. As a result, the less PASP molecules in the liquor act as a nucleation 17 template, and thus facilitate the precipitation of struvite.<sup>42</sup> The similar effect has also 18 been reported by Elhadj et al.<sup>43</sup> in the system of PASP and calcite. Nevertheless, with 19 the increase of PASP concentration, the coordination between PASP and Mg<sup>2+</sup> 20 predominates over the template function of PASP. Therefore, high concentration of 21 PASP can effectively block struvite precipitation and growth, even completely avoid 22 23 the formation of struvite precipitate.

24

#### **3.2. Effect of reaction conditions on PASP inhibition performance**

The influence of several key parameters, including reaction time, mixing energy, pH, and calcium ions on the scale inhibition performance of PASP was examined. A 0.5 mM of PASP was dosed for each run to guarantee an effective and incomplete inhibition of struvite formation, and other conditions were the same as the PASP concentration-dependent experiments. 1

### 2 **3.2.1. Reaction time**

If PASP was used as an inhibitor to prevent struvite formation in WWTPs, its 3 inhibitory capacity should last enough time to allow wastewater pass through the 4 5 whole pipes before precipitation. Hence, the experiments with stirring time of 8, 16, 24, 48 h were carried out. The precipitates were always formed at different time 6 intervals due to the limited dose of PASP (0.5 mM). The XRD analyses confirmed 7 8 that these precipitates were also struvite (e.g., Figure 1b). The FESEM results showed 9 that the morphology and size of struvite obtained after 16, 24, and 48 h were the same 10 as the 8-hour product (e.g., Figure 2d). It seems that the different stirring time have no effect on the precipitate phase and morphology. Figure 3a shows the plot of 11 precipitate mass versus stirring time. Although the slight increase in precipitate mass 12 was observed from 8 to 16 h, the precipitate mass almost remained constant after 16 h 13 of stirring, indicating that the magnesium-PASP species are stable. Therefore, PASP 14 inhibition performance is effective and sustainable in a longer period of time, 15 16 ensuring wastewater to be treated with less struvite precipitation.

17

#### 18 **3.2.2.** Mixing energy

It is commonly observed that struvite preferentially accumulates in specific 19 locations of a treatment or conveyance system with high mixing energy, such as pipe 20 elbows, pumps, and mixers, rather than uniform deposition.<sup>44</sup> The mechanisms 21 responsible for the preferential accumulation of struvite are associated with the 22 23 mixing energy (or turbulence). In the areas of high mixing energy, CO<sub>2</sub> liberation is enhanced. This can increase pH of the solution, and therefore favor the formation of 24 struvite crystals.<sup>20</sup> More importantly, struvite growth rate was found to be 25 transport-limited.<sup>44</sup> High mixing energy will enhance the mass transfer of solute to the 26 crystals and facilitate struvite crystallization and growth.<sup>44-45</sup> In our experiments, we 27 tested the effect of mixing energy on PASP inhibition performance. The stirring speed 28 ranged from 0 to 960 rpm (i.e., 0, 120, 360, 600, 960 rpm). After 8 h of reaction, no 29 product was obtained in the static experiment (0 rpm) and the solution was always 30

clear. In contrast, the struvite crystals with similar morphology and size were 1 2 harvested under magnetic stirring ranging from 120 to 960 rpm (e.g., Figure 1a and 2d), revealing that the stirring favors the nucleation and growth of struvite. However, 3 no marked changes in precipitate mass were observed with the increase in stirring 4 speed (Figure 3b), further indicating that in the presence of PASP, the high mixing 5 energy cannot also transport enough free  $Mg^{2+}$  to the growth fronts of struvite crystals, 6 and thereby accelerate crystal growth. In other words, the high mixing energy cannot 7 8 weaken the inhibitory capacity of PASP, and PASP can be potentially applied to 9 controlling struvite scaling at different locations, especially those high-mixing 10 environments.

11

#### 12 **3.2.3. Initial pH**

The pH can affect dissolution and supersaturation, as well as morphology, 13 particle size, and purity of struvite.<sup>20, 46</sup> The wastewater may have different pH values 14 depending on the type and source, and the pH cannot be always controlled or 15 maintained throughout struvite crystallization process.<sup>46</sup> Therefore, the effect of initial 16 17 pH (pH<sub>i</sub>) on PASP performance was investigated. Here, the pH<sub>i</sub> ranged from 8 to 11 (i.e., 8, 9, 10, 11). The control experiments without PASP were also conducted. In the 18 absence of PASP, the collected precipitates all were the rod-shaped struvite at  $pH_i$ 19 20 8-10 (Figure 1a and 2a, Figure S1a,b and S2a); At pH<sub>i</sub> 11, the precipitate consists of the rod-shaped struvite and plate-like cattiite [Mg<sub>3</sub>(PO4)<sub>2</sub>·22H<sub>2</sub>O], which were 21 confirmed by the XRD and EDX analyses (Figure S1c and S2b, S3a and b). In the 22 23 presence of PASP, no precipitate was obtained at  $pH_i 8$ , but pure struvite was formed at pH<sub>i</sub>9-11 (Figure 1a, Figure S1d and e), and these struvite crystals gets slenderer 24 with pH<sub>i</sub> (Figure 2d, Figure S2c and d). The variation trend of struvite shape agrees 25 well with the results reported by Ma et al.<sup>46</sup> This is because high precipitation rate and 26 changes in aqueous speciation decrease the concentrations of  $Mg^{2+}(aq)$ ,  $NH_4^{+}(aq)$ , and 27  $PO_4^{3-}$  (aq), resulting in the limiting of crystal growth at higher  $pH_1^{4-}$  Figure 3c shows 28 the effect of  $pH_i$  on the mass of precipitate products in the presence of 0.0 and 0.5 29 mM PASP. It can be seen that the variation trends are similar under the two conditions. 30

**RSC Advances Accepted Manuscript** 

1 That is, the mass increased with  $pH_i$ , and the highest at  $pH_i$  10. This is consistent with the previous observations by Ma and Rouff.<sup>24</sup> The increase in yield with pH<sub>i</sub> can be 2 attributed to an increase in the activity of  $PO_4^{3-}(aq)$  as pH increase promotes the 3 equilibrium shifting from  $H_2PO_4^-$  and  $HPO_4^{2-}$  to  $PO_4^{3-}$  species. However, at pH<sub>1</sub> 11, 4 this effect is neutralized by the reduced activity of  $NH_4^+(aq)$  due to increased 5 formation of NH<sub>3</sub> species, and by the hydrolysis of  $Mg^{2+}(aq)$  to  $MgOH^{+}(aq)$ .<sup>24</sup> It is 6 notable that in the presence of PASP, the precipitation of struvite was significantly 7 8 inhibited at pH<sub>i</sub> from 8 to 11 compared with the controls (Figure 3c). Therefore, PASP 9 can inhibit struvite growth over a large pH<sub>i</sub> range, and additional dose of PASP can 10 achieve the same inhibitory efficiency at higher pH<sub>i</sub>.

11

### 12 **3.2.4.** Mg/Ca ratio

In sludge liquors, calcium levels can be high relative to magnesium.<sup>29</sup> These 13 calcium ions can interact with phosphate or carbonate ions to form additional mineral 14 precipitates, such as hydroxylapatite and calcite.<sup>20</sup> In this context, struvite formation 15 16 can be inhibited if the supply of phosphate is limited. As a chelating agent, PASP can form stable complexes with calcium and magnesium ions in solution.<sup>16, 33</sup> The 17 presence of calcium ions will unavoidably reduce the inhibitory efficiency of PASP on 18 19 struvite. Therefore, the inhibitory experiments at Mg/Ca ratios of 1:0, 2:1, 1:1, and 1:2 20 were performed. Similarly, the precipitates were examined by FESEM and XRD techniques. In the absence of PASP, the precipitates exhibited different mineralogical 21 characteristics over the range of Mg/Ca ratios. Without the addition of Ca, the pure 22 23 rod-shaped struvite was obtained (Figure 1a and 2a). When the Mg/Ca ratio was 2:1, 24 the rod-shaped crystals coated with nanoparticles and irregular aggregates of nanoparticles were formed (Figure S4a). The XRD pattern, despite slight background 25 26 noises, could still well be indexed as struvite (Figure 4a). Therefore, the rod-like 27 crystals can be safely assigned to struvite, whereas the nanoparticles and their aggregates may be an amorphous precipitate. Further increasing Ca concentration 28 (Mg/Ca = 1:1) led to the enhanced output of irregular aggregates with a few rod-like 29 crystals (Figure S4b), and the much more noises can be observed from its XRD 30

pattern (Figure 4b), indicative of the formation of much amorphous matter. When the 1 Mg/Ca ratio reached 1:2, only the huge tabular aggregates were harvested (Figure 2 S4c), and XRD analysis confirmed that the aggregates exhibited an amorphous feature 3 (Figure 4c), indicating that  $Ca^{2+}$  present in the liquor facilitate the formation of 4 amorphous precipitate. To further understand the chemical nature of the amorphous 5 aggregates, EDX analyses were also conducted (Figure S3c-g). Combined with the 6 FESEM images and XRD patterns, the aggregates were identified as amorphous 7 calcium phosphate. Le Corre et al.<sup>29</sup> also obtained amorphous calcium phosphate 8 when they studied the impact of calcium on struvite growth. We also found that the 9 10 addition of Ca inhibited struvite growth but increases the total yield of the precipitates, i.e., the formation of other scales (Figure 3d). However, in the presence of 0.5 mM 11 PASP, only trapezoidal and rod-like struvite was obtained at different Mg/Ca ratios, 12 and no calcium precipitate was harvested (Figure S4d-f, Figure 4d-f). It indicated that 13 14 calcium ions had priority to complex with PASP, resulting in the decrease of PASP inhibition performance on struvite formation. Therefore, a slight increase in struvite 15 16 mass was observed with increasing calcium concentration (Figure 3d). However, the 17 mass of struvite obtained in the presence of PASP was still pretty low at different Mg/Ca ratios compared with the experiments without PASP. That is to say, the effect 18 of calcium on PASP inhibition performance was limited and PASP can still exert 19 20 significant scale inhibition even in the presence of calcium.

21

#### 22 **3.3.** Effect of PASP concentration on struvite dissolution

23 Except for the inhibition of scale formation, the removal of existing struvite scale 24 is another important work in WWTPs. Investigators have tried to find out effective 25 and environmentally safe chelating agents to provide a feasible alternative to acid dissolution, which can cause corrosion in metallic process equipment and piping 26 systems.<sup>47</sup> The previous studies demonstrate that PASP can promote the dissolution of 27 calcium oxalate and calcium phosphate due to its ability to chelate metal ions in 28 solution.<sup>47-49</sup> Therefore, the chelation of PASP with magnesium may also be a driving 29 force for struvite dissolution. Here, the dependence of struvite dissolution on PASP 30

**RSC Advances Accepted Manuscript** 

concentration was examined at pH 8. Figure 5 presents the FESEM images of the 1 2 products before and after dissolution with different concentrations of PASP. The pristine struvite crystals take a coffin-like shape with a length of ca. 30  $\mu$ m and a 3 width of ca. 10  $\mu$ m, and a number of tiny pits and crevices can be seen on lateral sides 4 (Figure 5a). After dispersed in the deionized water for 30 minutes, some tiny pits or 5 crevices were enlarged, but the basic configuration still remained, indicating that the 6 limited dissolution occurred (Figure 5b). Although struvite solubility is low in water, 7 it can be 18 mg  $\cdot$  100 mL<sup>-1</sup> at 25 °C.<sup>20</sup> That is to say, as much as 9 mg of struvite will be 8 dissolved if the dissolution equilibrium is achieved in our case (50 mg struvite + 50 9 10 ml of deionized water). Our dissolution experiment revealed that 8.3 mg (16.6%) of struvite was dissolved in deionized water, approaching to the equilibrium dissolution 11 value (Figure S5). When 0.1 mM of PASP was dosed, the struvite crystals became 12 thinner, containing many grooves and deep carvings (Figure 5c). Noticeably, a couple 13 14 of enlarged corrosion pits could be seen on the surfaces. Some struvite crystals had decreased width and approximately 28.1 mg (56.2%) of struvite was dissolved (Figure 15 16 S5). Therefore, the dose of PASP can effectively enhance struvite dissolution. As the 17 PASP concentration was 0.5 mM, more remarkable dissolution occurred, resulting in the much thinner morphology, i.e., dumbbell-shaped (Figure 5d). Consistently, over 18 44 mg (88%) of struvite was dissolved in this case (Figure S5). Hence, the 19 effectiveness of PASP to promote struvite dissolution was proportional to its 20 concentration. Wu and Grant <sup>16</sup> reported that all the carboxyl groups of PASP will be 21 ionized above pH 6, leading to the strongest complexation with magnesium. The 22 23 complexation reduced the amount of free magnesium ions and therefore broke down 24 the precipitation-dissolution equilibrium, resulting in the enhancement of struvite dissolution (Eq. 2). With increasing PASP concentrations, more complexation will be 25 formed, further promoting the dissolution. Therefore, PASP can be an effective 26 cleaning agent for the existing struvite scale. 27

28

# 29 **4. Conclusions**

1

2 In summary, the precipitation experiments dosed with PASP showed that PASP can not only effectively inhibit the formation and growth of struvite, but also 3 significantly change struvite morphology, leading to the evolution from rod-shape to 4 5 arrowhead-shape, triangle, or trapezoid. Moreover, several key parameters that possibly effect on PASP inhibition performance were also tested, and the results 6 7 revealed that PASP can still exert strong antiscaling on struvite even in a long running 8 time and a large range of mixing energy. Nevertheless, its inhibition potency is pH 9 dependant, and decreases with pH<sub>i</sub>. The presence of calcium ions slightly reduced the inhibition potency of PASP while the dose of PASP prevented the formation of 10 amorphous calcium phosphate. The dissolution experiments dosed with PASP showed 11 that PASP can promote the dissolution of the preformed struvite and the effectiveness 12 increases with its concentration. It appears that PASP has a strong ability to not only 13 14 effectively inhibit the formation and growth of struvite, but also facilitate struvite dissolution. Therefore, PASP can serve as an environmentally friendly scale inhibitor 15 16 and scale cleaning agent.

17

## **18** Acknowledgements

This work was partially supported by the Chinese Ministry of Science and
Technology (No. 2014CB846003), the Natural Science Foundation of China (No.
41372053), and the Specialized Research Fund for the Doctoral Program of Higher
Education (No. 20133402130007).

23

# 24 Electronic supplementary information (ESI)

The characterizations of the samples precipitated at different pH<sub>i</sub> and different Mg/Ca ratios, including XRD patterns, FESEM images, and EDX spectra. Plot of remaining mass in the dissolution experiments.

28

1

# 2 **References**

- 3 1. J. D. Doyle and S. A. Parsons, *Water Res.*, 2002, **36**, 3925-3940.
- 4 2. F. Abbona and R. Boistelle, J. Cryst. Growth, 1979, 46, 339-354.
- 5 3. C. Beal, T. Gardner, W. Ahmed, C. Walton and D. Hamlyn-Harris, in Proceedings
  of AWA Conference, 2007, pp. 1–8.
- 7 4. D. Mamais, P. A. Pitt, Y. W. Cheng, J. Loiacono and D. Jenkins, *Water Environ*.
- 8 *Res.*, 1994, **66**, 912-918.
- 9 5. C. M. Mehta and D. J. Batstone, *Water Res.*, 2013, 47, 2890-2900.
- Mohajit, K. K. Bhattarai, E. P. Taiganides and B. C. Yap, *Biol. Wastes*, 1989, 30, 133-147.
- 7. J. D. Doyle, K. Oldring, J. Churchley and S. A. Parsons, *Water Res.*, 2002, 36, 3971-3978.
- 14 8. S. Williams, *Environ. Technol.*, 1999, **20**, 743-747.
- 15 9. J. Borgerding, J. Water Pollut. Control Fed., 1972, 44, 813-819.
- 16 10. S. A. Parsons and J. D. Doyle, *Water Sci. Technol.*, 2004, 49, 177-182.
- 17 11. J. D. Doyle, K. Oldring, J. Churchley, C. Price and S. A. Parsons, *J. Environ.* 18 *Eng-Asce*, 2003, **129**, 419-426.
- 19 12. K. Fytianos, E. Voudrias and N. Raikos, *Environ. Pollut.*, 1998, **101**, 123-130.
- 20 13. J. L. Means, T. Kucak and D. A. Crerar, *Environ. Pollut. B*, 1980, 1, 45-60.
- 21 14. V. Sykora, P. Pitter, I. Bittnerova and T. Lederer, *Water Res.*, 2001, **35**, 2010-2016.
- 22 15. J. Prywer and M. Olszynski, J. Cryst. Growth, 2013, 375, 108-114.
- 23 16. Y. T. Wu and C. Grant, *Langmuir*, 2002, **18**, 6813-6820.
- 24 17. A. Wierzbicki, J. D. Sallis, E. D. Stevens, M. Smith and C. S. Sikes, *Calcified Tissue Int.*, 1997, 61, 216-222.
- 26 18. I. Stratful, M. D. Scrimshaw and J. N. Lester, *Water Res.*, 2001, 35, 4191-4199.
- 27 19. Y. Ueno and M. Fujii, *Environ. Technol.*, 2001, **22**, 1373-1381.
- 28 20. K. S. Le Corre, E. Valsami-Jones, P. Hobbs and S. A. Parsons, Crit. Rev. Env. Sci.
- 29 *Tec.*, 2009, **39**, 433-477.

1	21. C. C. Wang, X. D. Hao, G. S. Guo and M. C. M. van Loosdrecht, Chem. Eng. J.,
2	2010, <b>159</b> , 280-283.
3	22. L. E. de-Bashan and Y. Bashan, Water Res., 2004, 38, 4222-4246.
4	23. L. Decrey, K. M. Udert, E. Tilley, B. M. Pecson and T. Kohn, Water Res., 2011,
5	<b>45</b> , 4960-4972.
6	24. N. Ma and A. A. Rouff, Environ. Sci. Technol., 2012, 46, 8791-8798.
7	25. A. A. Rouff, Environ. Sci. Technol., 2012, 46, 12493-12501.
8	26. J. R. Lin, N. Chen and Y. M. Pan, Environ. Sci. Technol., 2013, 47, 12728-12735.
9	27. J. R. Lin, N. Chen and Y. M. Pan, Environ. Sci. Technol., 2014, 48, 6938-6946.
10	28. A. A. Rouff and K. M. Juarez, Environ. Sci. Technol., 2014, 48, 6342-6349.
11	29. K. S. Le Corre, E. Valsami-Jones, P. Hobbs and S. A. Parsons, J. Cryst. Growth,
12	2005, <b>283</b> , 514-522.
13	30. X. D. Hao, C. C. Wang, L. Lan and M. C. M. van Loosdrecht, Water Sci. Technol.,
14	2008, <b>58</b> , 1687-1692.
15	31. K. S. Le Corre, E. Valsami-Jones, P. Hobbs, B. Jefferson and S. A. Parsons, Water
16	Res., 2007, 41, 419-425. 己后移
17	32. X. D. Hao, C. C. Wang, M. C. M. van Loosdrecht and Y. S. Hu, Environ. Sci.
18	Technol., 2013, <b>47</b> , 4965-4966.
19	33. S. D. Jiang, Q. Z. Yao, G. T. Zhou and S. Q. Fu, J. Phys. Chem. C, 2012, 116,
20	4484-4492.
21	34. X. X. Sheng, M. D. Ward and J. A. Wesson, J. Am. Chem. Soc., 2003, 125,
22	2854-2855.
23	35. A. Bigi, B. Bracci, S. Panzavolta, M. Iliescu, M. Plouet-Richard, J. Werckmann
24	and D. Cam, Cryst. Growth Des., 2004, 4, 141-146.
25	36. B. Njegic-Dzakula, L. Brecevic, G. Falini and D. Kralj, Cryst. Growth Des., 2009,
26	9, 2425-2434.
27	37. D. Hasson, H. Shemer and A. Sher, Ind. Eng. Chem. Res., 2011, 50, 7601-7607.
28	38. K. L. Dziak and O. Akkus, J. Bone Miner. Metab., 2008, 26, 569-575.
29	39. J. D. Doyle, R. Philp, J. Churchley and S. A. Parsons, Process Saf. Environ.
30	Protect., 2000, 78, 480-488.

- 1 40. H. Li, Q. Z. Yao, Y. Y. Wang, Y. L. Li and G. T. Zhou, Sci. Rep., 2015, 5, No.7718.
- 2 41. Z. Romanowski, P. Kempisty, J. Prywer, S. Krukowski and A. Torzewska, *J. Phys.*3 *Chem. A*, 2010, **114**, 7800-7808.
- 4 42. J. Roque, J. Molera, M. Vendrell-Saz and N. Salvado, *J. Cryst. Growth*, 2004, 262,
  5 543-553.
- 43. S. Elhadj, E. A. Salter, A. Wierzbicki, J. J. De Yoreo, N. Han and P. M. Dove, *Cryst. Growth Des.*, 2006, 6, 197-201.
- 44. K. N. Ohlinger, T. M. Young and E. D. Schroeder, *J. Environ. Eng-Asce*, 1999,
  125, 730-737.
- 10 45. D. Kim, J. Kim, H. D. Ryu and S. I. Lee, *Bioresour. Technol.*, 2009, 100, 74-78.
- 46. N. Ma, A. A. Rouff and B. L. Phillips, ACS Sustain. Chem. Eng., 2014, 2,
  816-822.
- 47. F. Littlejohn, C. S. Grant, Y. L. Wong and A. E. Saez, *Ind. Eng. Chem. Res.*, 2002,
  41, 4576-4584.
- 48. F. Poumier, P. Schaad, Y. Haikel, J. C. Voegel and P. Gramain, *J. Biomed. Mater. Res.*, 1999, 45, 92-99.
- 17 49. S. W. Guo, M. D. Ward and J. A. Wesson, *Langmuir*, 2002, 18, 4284-4291.

18

1	
2	Figure captions
3	Figure 1 Typical XRD patterns of the precipitates from the artificial liquor dosed with
4	0 (a) and 0.5 (b) mM PASP. (c) Effect of PASP concentration on mass of struvite
5	precipitation. All of the precipitation runs were performed at $pH_i$ 9.0 for 8 h.
6	Figure 2 FESEM images of the 8 h struvite crystals from the artificial liquor dosed
7	with 0 (a), 0.01 (b), 0.3 (c), and 0.5 (d) mM of PASP.
8	Figure 3 Mass of the precipitates obtained under different precipitation conditions: (a)
9	different time intervals; (b) different stirring speeds; (c) different $pH_i$ values; (d)
10	different Mg/Ca ratios.
11	Figure 4 XRD patterns of the precipitates at Mg/Ca ratio 2:1 (a), 1:1 (b), 1:2 (c) dosed
12	with 0 mM PASP, and at Mg/Ca ratio 2:1 (d), 1:1 (e), 1:2 (f) dosed with 0.5 mM
13	PASP.
14	Figure 5 FESEM images of struvite crystals before (a) and after (b) 30 min of
15	dissolution dosed with 0 (b), 0.1 (c), and 0.5 (d) mM of PASP.
16	







**RSC Advances Accepted Manuscript** 



2θ (degree)

Figure 4



**RSC Advances Accepted Manuscrip** 



11

Figure 5

24