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Naturally-occurring radioactive materials (NORM) associated with unconventional drilling flowback and produced fluids from the Marcellus Shale have raised environmental health and waste management questions. Radium (Ra) has received considerable attention due to the challenges of removing it from produced fluids. Ra waste management challenges have largely been attributed to ionic strength; however, the apparent solubility of Ra in Marcellus Shale fluids exceeds that observed in other high ionic strength fluids. We herein report the effects of chemical matrices on Ra, as well as other NORM. We find that the concentration of Ba in these fluids is a more important predictor for Ra solubility than ionic strength.

**Title: Partitioning of Naturally-Occurring Radionuclides (NORM) in Marcellus Shale Produced Fluids Influenced by Chemical Matrix**

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**Keywords:** Marcellus Shale, hydraulic fracturing, unconventional drilling, NORM, radium, barium, produced fluids

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**Abstract:** Naturally-occurring radioactive materials (NORM) associated with unconventional drilling produced fluids from the Marcellus Shale have raised environmental concerns. However, few investigations into the fundamental chemistry of NORM in Marcellus Shale produced fluids have been performed. Thus, we performed radiochemical experiments with Marcellus Shale produced fluids to understand the partitioning behavior of major radioelements of environmental health concern (uranium (U), thorium (Th), radium (Ra), lead (Pb), and polonium (Po)). We applied a novel radiotracer, <sup>203</sup>Pb, to understand the behavior of trace-levels of <sup>210</sup>Pb in these fluids. Ultrafiltration experiments indicated U, Th, and Po are particle reactive in Marcellus Shale produced fluids and Ra and Pb are soluble. Sediment partitioning experiments revealed that >99% of Ra does not adsorb to sediments in the presence of Marcellus Shale produced fluids. Further experiments indicated that although Ra adsorption is related to ionic strength, the concentrations of heavier alkaline earth metals (Ba, Sr) are stronger predictors of Ra solubility.

**Main Text:****Introduction:**

The levels of radium (Ra) isotopes and other naturally-occurring radioactive materials (NORM) in flowback and produced fluids from the Marcellus Shale are an emerging environmental concern.<sup>1-9</sup> Recent reports demonstrated that Marcellus Shale produced fluids from unconventional drilling are enriched in Ra isotopes (<sup>228</sup>Ra, <sup>226</sup>Ra, <sup>224</sup>Ra) and devoid of most other long-lived NORM, such as certain U, Th, Pb and Po isotopes.<sup>1, 9-11</sup> Despite research interest in radionuclides contained in Marcellus Shale produced fluids, there have been few laboratory-based studies that characterize the fundamental chemical behavior of NORM in Marcellus Shale produced fluids. Information has been gleaned about the chemical behavior of NORM in these fluids from environmental studies; however, even well designed environmental studies may not accurately characterize the chemical behavior of unconventional drilling wastes due to uncertainty and complexity of environmental sites in a region of the United States that has a long history with the extractive industries. That is, unknown treatment levels (i.e., speciation and composition) of waste source terms, ongoing and legacy pollution from mixed sources (natural gas and coal extraction industries), and misinformation can complicate the interpretation of results from environmental studies of unconventional drilling wastes.<sup>12, 13</sup> Due to the limitations of environmental studies, we decided to explore information about the chemical behavior of NORM in produced fluids through laboratory-based partitioning studies. Although such experiments may lack generalizability, they provide important clues about behavior of radioactive elements without the challenges of accounting for uncontrolled environmental inputs, radioactive decay, and decay product ingrowth.

One of the major chemical parameters of interest for predicting the fate and transport of NORM in unconventional drilling wastes is ionic strength. A wide body of literature exists on the effects of ionic strength on the fate and transport of NORM in the environment. With respect to Ra, adsorption and desorption from sediments has been of great interest to environmental scientist and oceanographers.<sup>14-20</sup> Ra adsorption studies show that Ra generally partitions through an ion exchange mechanism to the aqueous phase as ionic strength increases, yet most empirical data stops at ionic strengths

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comparable to those of seawater. Further, most studies have only explored the effects of ionic strength by considering concentration of Na with little attention to high molarities of alkaline earth metals (Mg, Ca, Sr, Ba). It is unclear whether these models are appropriate for Marcellus Shale produced fluids which may have ionic strengths up to 10 times that of seawater and levels of alkaline earth metals in the g/L range.<sup>8, 9, 21</sup> Despite years of research into NORM, the current understanding of Ra chemistry in Marcellus Shale produced fluids is murky and the understanding of the chemical behavior of U, Th, Pb, and Po in these fluids is largely unexplored.

In this study, we characterize the chemical composition of Marcellus Shale produced fluids and suspended particulates by high resolution transmission electron microscopy (HRTEM) and powder x-ray diffraction (PXRD). We describe the partitioning of U, Th, Ra, Pb, and Po in Marcellus Shale produced fluids using radiotracers and ultrafiltration (0.1  $\mu\text{m}$  filter pore size). We then applied sequential extraction and radiotracer techniques to determine the potential for Ra, Pb, and Po adsorption-desorption on a variety of sediment textures and mineral compositions. Note, we applied an underused radiotracer,  $^{203}\text{Pb}$ , to understand the behavior of trace levels of  $^{210}\text{Pb}$  in these fluids.  $^{203}\text{Pb}$  has been applied as a tracer of Pb in environmental systems;<sup>22</sup> however, lack of access to the isotope has stymied its use. With the increase in medical imaging applications of  $^{203}\text{Pb}$ , in our laboratory in particular,  $^{203}\text{Pb}$  is now more readily attainable. Next, we investigated the effect of ionic strength on Ra using a chemical matrix characteristic of Marcellus Shale fluids. Finally, we explored the role of exchangeable cations (Na, Mg, Ca, Sr, Ba) at a range of concentrations (0.01 M, 0.1 M, and 1 M) on the partitioning of Ra.

**Experimental**

**General.** Sediments were taken from the Dunkard Creek near Wana, West Virginia, USA in May 2014 or from collections at the University of Iowa Quaternary Materials Laboratory. Dunkard Creek sediments were collected from the site of a spill that occurred in 2009.<sup>23</sup> The Dunkard Creek spill was speculated to be related to unconventional drilling and/or coal wastes;<sup>23</sup> however, the etiology of this spill is beyond the scope of this study. Grain size, PXRD, and HRTEM were performed at the University of Iowa

using standard techniques.<sup>24, 25</sup> The Marcellus Shale produced fluids described herein were previously characterized for NORM and matrix composition.<sup>1, 5</sup> Sequential extractions were based on previously published methods.<sup>26</sup> In partitioning and sequential extraction experiments, <sup>226</sup>Ra was determined by <sup>222</sup>Rn emanation,<sup>5</sup> <sup>203</sup>Pb by NaI gamma spectrometry based on standard laboratory methods,<sup>27</sup> and <sup>nat</sup>U, <sup>nat</sup>Th, and <sup>210</sup>Po by preconcentration and alpha spectrometry.<sup>1, 28</sup>

**Grain size.** Sediment grain-size distribution was analyzed using the pipette method with size fraction breaks at <2 $\mu$  (clay), 2-20 $\mu$  (fine silt), 20-50 $\mu$  (coarse silt) and 50 $\mu$ -2mm (sand) at the University of Iowa Quaternary Materials Laboratory.

**PXRD** Sediments, Marcellus Shale produced fluids fines (0.1  $\mu$ m polypropylene filter), and Marcellus Shale produced fluids solid-state evaporites were prepared for power X-ray diffraction (PXRD) by mortar and pestle (under acetone). Diffractograms were collected on a Bruker Advance Diffractometer equipped with CuK $\alpha$  radiation ( $\lambda$  = 1.54056 Å) and a LynxEye detector. Scans were performed from 5-60° 2 $\theta$  with a step size of 0.02° and collection time of 1 sec/step. Major mineral phases present in the solid samples were identified using the Bruker Eva software (**Fig. S1, Fig. S2, Fig. S3, Table S1**).

**High Resolution Transmission Electron Microscopy.** Morphological features of the Fe particulates were determined using high resolution transmission electron microscopy (HRTEM) at the University of Iowa Central Microscopy Research Center. A 1 mg mL<sup>-1</sup> colloidal suspension of the particulates was prepared in ultrapure water and sonicated for 30 min to improve dispersion. Holey carbon copper grids (Ted Pella, Inc 01824, UC-A on holey, 400 mesh Cu) were loaded with 5  $\mu$ L aliquots of the suspension and evaporated to dryness in air. HRTEM investigations were carried out by a JEOL JEM 2100F equipped with a cold Schottky cathode, field emission gun, and operated at a 200 kV accelerating voltage. Gatan DigitalMicrograph software was utilized for image acquisition in congruency with an externally inserted Orius camera cooled to -25 °C.



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157 ***Produced Fluids Filtration U, Th, Ra, Pb, Po.*** <sup>232</sup>U (Eckert and Ziegler 92403), <sup>228</sup>Th (in  
158 <sup>232</sup>U tracer/endogenous), <sup>226</sup>Ra (endogenous), <sup>203</sup>Pb Lantheus Medical Imaging, Billerica,  
159 MA), and <sup>210</sup>Po (from <sup>210</sup>Pb source NPL A13959) were added to Marcellus Shale  
160 produced fluids (20 mL, except <sup>226</sup>Ra 35 mL), stirred (6-18 hours, room temperature, 220  
161 RPM), and ultra-filtered (0.1 µm, filtration Resolve Eichrom). Filters and supernatant  
162 samples requiring analysis by alpha spectrometry were then traced with <sup>nat</sup>U (Eckert and  
163 Ziegler 92564), <sup>230</sup>Th (NIST SRM 4342A), or <sup>209</sup>Po (Eckert and Ziegler 92565) and  
164 prepared using methods previously described.<sup>1</sup> <sup>203</sup>Pb samples were quantitated by NaI  
165 gamma spectrometry. <sup>226</sup>Ra samples were analyzed by RAD7 (DurrIDGE Company, Inc.,  
166 Billerica, MA, USA).

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169 ***Sediment Sequential Extraction of Ra.*** Selected homogenized sediments were placed  
170 into 50 mL conicals (n=3, 3.5 g) (**Fig. S4**). Marcellus Shale produced fluids (35 mL)  
171 were added to each conical before shaking on a rotating table (1 hr, room temperature,  
172 220 RPM). Conicals were then centrifuged (5 min, 3400 x g), and the supernatant was  
173 transferred to clean vials (40 mL, glass VOC). To prevent foaming during counting by  
174 RAD7, anti-foam solution (Sigma, emulsion B, 1 mL 25%) was added to each vial before  
175 topping with H<sub>2</sub>O and hermetically sealing. Ra was then sequentially extracted from  
176 sediment based on methods previously described.<sup>26</sup> Briefly, sediment was treated three  
177 times with water-soluble-phase extractant (0.4 M MgCl<sub>2</sub>, pH 5.5), once with carbonate-  
178 phase extractant (1M NH<sub>4</sub>Ac in 25% v/v acetic acid), once with reducible-phase  
179 extractant (0.01 M NH<sub>2</sub>OH in 25% v/v acetic acid), once with organic-phase extractant  
180 (30% v/v H<sub>2</sub>O<sub>2</sub>, 0.02 M HNO<sub>3</sub>) and finally with acid-digestible phase (4 M HNO<sub>3</sub>).  
181 Extractions were performed at room temperature on a rotating table (1 hr, 220 RPM),  
182 each time centrifuging and transferring supernatant to clean glass vials (40 mL, VOC).  
183 Vials containing acetic acid, H<sub>2</sub>O<sub>2</sub>, or HNO<sub>3</sub> were taken to complete dryness before  
184 topping with H<sub>2</sub>O and hermetically sealing. Samples were held 30+ days to ensure  
185 secular equilibrium between <sup>222</sup>Rn and <sup>226</sup>Ra before counting by RAD7 (DurrIDGE,  
186 Billerica, MA) as previously described.<sup>5</sup> Assessment of adsorption kinetics was



performed on DC6 sediment similarly as described above, yet samples were incubated with Marcellus Shale produced fluids and the extractants for 168 hours *en lieu* of 1 hour.

***Sediment Sequential Extraction of Ba, Pb, Po.*** Sediment from Dunkard Creek was placed into conicals (50 mL, n=3, 2 g). Marcellus Shale produced fluids (20 mL) and radiotracers ( $^{133}\text{Ba}$  Eckert and Ziegler 6133,  $^{203}\text{Pb}$  (Lantheus Medical Imaging),  $^{210}\text{Po}$  (supported by  $^{210}\text{Pb}$ , NPL A13959) traced with  $^{209}\text{Po}$  Eckert and Ziegler 92565)) were added to each conical before shaking on a rotating table (1 hr, room temperature, 220 RPM). Conicals were then centrifuged (5 min, 3400 x g), the supernatant was transferred to clean liquid scintillation (LS) vials (20 mL, glass). LS vials were then counted by a well-type sodium iodide (NaI) detector (Ortec) and gamma-emitting radionuclides were quantitated using standard radioactivity measurement techniques, similarly as described previously.<sup>27</sup> Po was separated and quantitated using standard methods previously described.<sup>1</sup> The radionuclides adsorbed to the sediment were then sequentially extracted based on methods previously described.<sup>26</sup> Extractions were performed at room temperature on a rotating table (1 hr, 220 RPM), each time centrifuging and transferring supernatant to clean glass vials (20 mL, LS).

***Ionic Strength and Ra Partitioning.*** All experiments investigating Ra partitioning were performed on DC6 sediment (n=3, 3.5 g), with the indicated salt solution (35 mL), and  $^{226}\text{Ra}$  standard (37.3 Bq, NIST traceable standard, Eckert & Ziegler 1742-75). Marcellus Shale produced fluids surrogate matrix consisted of reagent grade (or higher) chloride form salts with the following concentrations:  $\text{Na}^+$  (1300 mM),  $\text{K}^+$  (4 mM),  $\text{Mg}^{2+}$  (35 mM),  $\text{Ca}^{2+}$  (320 mM),  $\text{Sr}^{2+}$  (410 mM),  $\text{Ba}^{2+}$  (66 mM),  $\text{Fe}^{3+}$  (0.78 mM),  $\text{Cl}^-$  (4100 mM). Note, surrogate matrix was only used in experiments described in this paragraph. Surrogate matrix was used so that  $^{226}\text{Ra}$  activity (37.3 Bq) could be controlled despite dilution of the matrix solution. Separate select cations ( $\text{Na}^+$ ,  $\text{Mg}^{2+}$ ,  $\text{Ca}^{2+}$ ,  $\text{Sr}^{2+}$ , and  $\text{Ba}^{2+}$ ) were prepared by serial dilution to the following concentrations: 1 M, 0.1 M, and 0.01 M. All samples were shaken on a rotating table (1 hour, room temperature, 220 RPM), centrifuged (3400 x g), and the supernatant was transferred to clean glass vials (40 mL,

217 glass VOC). Anti-foaming solution (1 mL) was added to each vial before topping with  
218 anti-foam (1 mL), hermetically sealing (30+ days), and counting by RAD7.

220 **Results and Discussion:**

221 **NORM in Produced Fluids**

222 Very little is known about the concentrations of most NORM in Marcellus Shale  
223 fluids, and less information is available regarding NORM partitioning between aqueous  
224 and particulate phases. Several studies have developed a general understanding of how  
225 NORM levels in produced fluids change throughout the lifecycle of the unconventional  
226 drilling process.<sup>7-9</sup> For example, Ra isotope concentrations steadily increase concurrently  
227 with the age of the well and concentration of total dissolved solids (TDS) and alkaline  
228 earth metals.<sup>7, 8</sup> In contrast, U levels have been reported to spike briefly at the beginning  
229 of the flowback stage and then decrease over time.<sup>11</sup> Lastly, an analysis of late stage  
230 Marcellus Shale fluids by Nelson *et al.* (2015) reported U, Th, Pb, and Po isotopes were  
231 initially absent, but certain isotopes (<sup>228</sup>Th, <sup>210</sup>Pb, <sup>210</sup>Po) increase by radiochemical decay  
232 product ingrowth processes.<sup>1</sup> Collectively, these reports provide important clues about the  
233 behavior of NORM within Marcellus Shale fluids at the point source, particularly at  
234 increased pressure, higher temperature, and altered redox conditions that occur in the  
235 subsurface environment.

236 It is unclear how U, Th, Ra, Pb, and Po behave when produced fluids enter low-  
237 temperature geochemical environments. Although we anticipate the majority of U and Th  
238 isotopes (with the exception of <sup>228</sup>Th) will remain at depth or be adsorbed to solid waste  
239 products,<sup>1</sup> understanding the solubility of U and Th in Marcellus Shale produced fluids  
240 under surface conditions is important, particularly in cases where there are significant  
241 interactions between the waste forms. To understand the basic partitioning of U, Th, Ra,  
242 Pb, and Po in surface waters, we spiked homogenized Marcellus Shale produced fluids  
243 with appropriate radiochemical tracers. Spiked fluids underwent ultrafiltration (0.1 μm)  
244 and particulate and aqueous phases were analyzed by alpha or gamma spectrometry. U,  
245 Th, and Po associated with the particulates (98%, 97% and >99%, respectively) and Ra  
246 and Pb remained in solution (99% and 97%, respectively) (**Fig. 1A**).

Low particle reactivity of Ra and Pb led us to probe the mineral composition of particulates and dissolved solids in the Marcellus Shale fluids. Powder X-ray diffraction (PXRD) of the solid evaporites from the produced fluids revealed the presence of halite salts containing predominately alkali and alkali earth cations (**Fig. 1B, Fig. S3**). The diffractogram of the Fe particulates include a relatively high background due to fluorescence, but the presence of two broad peaks suggests the presence of an amorphous Fe oxyhydroxide (**Fig. 1C, Table S1**). In addition, weaker reflections correspond to the mineral akaganeite ( $\beta$ -FeO(OH, Cl)), which requires the presence of 1-9 wt% chloride in solution for formation of this phase to be favored over goethite and other Fe<sup>3+</sup> oxyhydroxide species.<sup>29</sup> Akaganeite is most commonly observed as a corrosion product in marine environments,<sup>30</sup> but the high ionic strength of the produced fluids contains concentrations of chloride to support the formation of akaganeite. The high-resolution transmission electron microscopy (HRTEM) image of the Fe particulates reveals the formation of bladed aggregates that are approximately 20-50 nm in width (**Fig. 1 D**). Akaganeite has previously been reported to form elongated aggregates with longitudinal striations and are similar to the morphology of the Fe particulates obtained from the produced fluids.<sup>31, 32</sup> We attribute the low observed adsorption of Pb and Ra to the akaganeite and amorphous iron oxides to matrix effects (competition with exchange cations), similarly as observed for ferrihydrite and goethite.<sup>16</sup>

### Environmental Fate of NORM

The gross levels of radionuclides in unconventional drilling wastes may be inappropriate for assessing environmental impact given the possibility for surface adsorption and sedimentation.<sup>26</sup> To understand the potential fate of radionuclides introduced into the environment, it is important to determine partitioning within soil and colloidal phase reservoirs through sequential extraction experiments.<sup>26, 33</sup> Initially our sequential extraction experiments focused exclusively on Dunkard Creek sediments, as these sediments were collected from a stream that has geology more consistent with sites in the region of the United States overlaying the Marcellus Shale (i.e., sites that are more likely to contact Marcellus Shale produced fluids). We applied sequential extraction techniques to determine <sup>226</sup>Ra concentrations in water-soluble, carbonate, Fe and Mn

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oxides, organic, and strong acid-digestible residue fractions for Dunkard Creek  
sediments.  $^{226}\text{Ra}$  was measured in each sequentially extracted fraction by  $^{222}\text{Rn}$   
emanation, to allow for accurate, trace-level radiochemical experiments (**Fig. S4 of**  
**Supporting Information**). Sequential extractions of  $^{226}\text{Ra}$  (1 hour) indicated >90% of  
 $^{226}\text{Ra}$  remained in the load solution (Marcellus produced fluids) and nearly all the  
residual  $^{226}\text{Ra}$  (>99%) was removed in the water-exchangeable fractions (**Fig. 2A**). The  
high observed solubility of  $^{226}\text{Ra}^{2+}$  led us to explore additional chemical parameters  
important for the adsorption of cations to sediments—kinetics, sediment  
mineralogy/texture, ionic strength, and chemical matrix.

The role of kinetics on the observed solubility was investigated in a time  
dependence study, where we performed the same sequential extraction experiment for  
168 hours (one week) instead of 1 hour. The same trend occurred in the extended reaction  
time, with >90% of  $^{226}\text{Ra}$  removed in the load solution and residual  $^{226}\text{Ra}$  was observed  
in the remaining water-exchangeable fractions such that >99% was recovered. This result  
suggests that kinetics have minimal effects on Ra adsorption-desorption in this system,  
which is in agreement with other studies that find Ra adsorption to sediments occurs  
within minutes.<sup>14, 17</sup>

We further investigated the mineralogy and soil characteristics of the sediments,  
because both characteristics greatly affect the speciation and adsorption of Ra.<sup>26, 34</sup> Soil  
texture and PXRD analysis indicated that four Dunkard Creek sediments were loam to  
sandy-loam, primarily consisting of quartz. Given that loams have lower binding capacity  
for  $^{226}\text{Ra}$  than clay-rich sediments,<sup>15, 16</sup> we included seven different sediments of different  
mineralogy, and diverse texture (97% sand (sediment 7), to 60% clay (sediment 5), to  
86% silt (sediment 3)) (**Fig 3, Fig. S4 of Supporting Information**). Regardless of soil  
texture and mineralogy, >90% of Ra remained in the load solution and nearly all the  
remaining  $^{226}\text{Ra}$  was removed in the water exchangeable fractions. This suggested to us  
that some chemical parameter(s) of the Marcellus Shale produced fluids overwhelm the  
binding capacity of soil minerals.

We suspected the high ionic strength in Marcellus Shale produced fluids could be  
responsible for the low Ra adsorption in this system; therefore, we tested Ra adsorption  
to sediment (DC6) with variable ionic strengths. A surrogate matrix comprised of the

same chemical parameters as Marcellus Shale produced fluids (ionic strength = 4.4 M) was adjusted by serial dilution (1:1, 1:10, 1:100, 1:1000, and 1:10000) and spiked with a known amount of  $^{226}\text{Ra}$  (37 Bq per sample). Sediments incubated with dilute surrogate (0.044 M and less) retained nearly all the  $^{226}\text{Ra}$  (>97%), which agrees with many studies that show  $^{226}\text{Ra}$  is particle reactive at ionic strengths typical of groundwater and freshwater (**Fig. 2B**). However, at an ionic strength of 0.44 M (slightly less than a typical seawater of 0.7 M), 70% of  $^{226}\text{Ra}$  remained in the aqueous phase (**Fig. 2B**). Sediments incubated with undiluted surrogate matrix (4.4 M) retained only 2% of the  $^{226}\text{Ra}$  (**Fig. 2B**). Although Webster *et al.* (1995) have shown that seawater can solubilize approximately 10-25%  $^{226}\text{Ra}$ ,<sup>14</sup> the high solubility of  $^{226}\text{Ra}$  at ionic strengths less than seawater prompted us to probe deeper into the role that chemical composition of the produced fluids plays on the adsorption process.

The chemical matrix of Marcellus Shale produced fluids is dominated by  $\text{Na}^+$ ,  $\text{Mg}^{2+}$ ,  $\text{Ca}^{2+}$ ,  $\text{Sr}^{2+}$ , and  $\text{Ba}^{2+}$  cations and  $\text{Cl}^-$  anions.<sup>8,9</sup> Thus, we tested the effects of individual ions on competition with  $\text{Ra}^{2+}$  for adsorption sites in Dunkard Creek sediments. We incubated sediments with solutions of chloride salts (0.01 M, 0.1 M, and 1 M), resulting in ionic strengths of 0.01 M, 0.1 M, and 1 M for Na solutions and 0.03 M, 0.3 M, and 3 M for alkaline earth solutions. Two clear trends emerged: (1) as cation concentration (and thus ionic strength) increased, the amount of  $^{226}\text{Ra}^{2+}$  in solution increased; and (2) as the ionic radius of the cation approached that of  $^{226}\text{Ra}^{2+}$ , the amount of  $^{226}\text{Ra}^{2+}$  in solution increased (**Fig. 2C**). The effect of concentrations of  $\text{Ba}^{2+}$  is quite dramatic, even at concentrations as low as 10 mM (typical  $\text{Ba}^{2+}$  concentration of late-stage Marcellus Shale produced fluids<sup>8</sup>). Importantly, samples of relatively low ionic strength but considerable levels of  $\text{Ba}^{2+}$ , may not behave as expected based on ionic strength alone. These chemical experiments suggest that in holding ponds containing untreated Marcellus Shale produced fluids,  $\text{Ra}^{2+}$  will not adsorb to solid mineral surfaces according to models developed for seawater matrices. In other words, the chemical composition of Marcellus Shale produced fluids (i.e., high concentrations of divalent alkaline earth metals) enhances the aqueous phase concentrations of  $\text{Ra}^{2+}$ . Consideration of  $\text{Ba}^{2+}$  concentrations, in particular, may be important for understanding the potential for  $\text{Ra}^{2+}$  to adsorb to the sediment-mineral interface in environmental settings (**Fig. 4**).

Furthermore, accurate measurement of  $\text{Ba}^{2+}$  concentration in produced fluids may be more important than ionic strength when assessing treatment goals and potential environmental impact. Future modeling studies will explore the effects of dilution and environmentally relevant anion concentrations (particularly  $\text{SO}_4^{2-}$ ) on bulk transport of  $\text{Ra}^{2+}$  in riparian environments.

Fate and transport of  $^{226}\text{Ra}$  decay products,  $^{210}\text{Pb}$  and  $^{210}\text{Po}$ , from Marcellus Shale produced fluids are important to consider as these isotopes are well-known to bioaccumulate.<sup>35</sup>  $^{210}\text{Pb}$  is challenging to study in environmental systems due to its radiochemical properties. Similarly,  $^{210}\text{Po}$  is often neglected as specific instrumentation and chemical separations (alpha spectrometry and autodeposition) are required for its analysis. Thus instead of measuring  $^{210}\text{Pb}$ , we employed a cyclotron-produced isotope— $^{203}\text{Pb}$ . The short half-life of  $^{203}\text{Pb}$  (51.92 hour) makes this isotope effectively a “massless” tracer and more representative than stable Pb tracers when investigating the fate of naturally-occurring low-mass-abundance (parts-per-quadrillion or less), radioactive Pb isotopes. Sequential extraction experiments were also performed for  $^{203}\text{Pb}$  and  $^{210}\text{Po}$  using a similar process as described above. Results for  $^{203}\text{Pb}$  indicated that  $45 \pm 2\%$  of  $^{203}\text{Pb}$  remained in the produced fluids (**Fig. 2D**). The remainder of the  $^{203}\text{Pb}$  was removed in the carbonate or strong acid phases ( $25 \pm 1\%$  and  $17 \pm 1\%$ , respectively). Isotope-dilution alpha spectrometry indicated that  $^{210}\text{Po}$  (traced with  $^{209}\text{Po}$ ) largely associated with carbonate and strong acid ( $25 \pm 4\%$  and  $52 \pm 2\%$ , respectively) phases (**Fig. 2D**). These results demonstrate that  $^{210}\text{Pb}$  and  $^{210}\text{Po}$  have different environmental fate and transport processes that may lead to radiochemical disequilibrium (non-steady state). More work is needed to understand the potential for these isotopes to migrate through the environment.

**Conclusion**

This study investigated the fundamental partitioning behavior of critical radioelements (U, Th, Ra, Pb, Po) contained in unconventional drilling wastes. In agreement with findings from an earlier study on NORM contained in Marcellus Shale produced fluids, we find that U, Th, and Po are insoluble (particle reactive) and Ra is soluble.<sup>1</sup> It is unclear why the high observed solubility of trace-level Pb (as measured by  $^{203}\text{Pb}$ ) in these fluids differs from earlier observations that  $^{210}\text{Pb}$  was absent in produced



fluids. The studies described herein were performed at room temperature, atmospheric pressure, and normoxic conditions on sediments from West Virginia to maintain relevance to potentially impacted riparian environments. These conditions are inherently different than those within the Marcellus Shale. Future studies should explore the solubility and partitioning behavior of NORM between interstitial fluids and particulate phases, under the pressures, temperatures, and anoxic conditions relevant to unconventional reservoirs.

Without a firm understanding of the basic chemistry of produced fluids, it is difficult to predict the fate and transport of NORM in the environment when these fluids are the source of NORM. We investigated the partitioning behavior of Ra, Pb, and Po from Marcellus Shale produced fluids to the sediment-mineral interface through bench-scale studies. We found that Ra was exceptionally soluble in the Marcellus Shale fluids and >90% was retained in the load solution. The remaining Ra was easily extracted by water-exchangeable washes, such that >99% was extracted in all water-soluble phases. Further experiments indicated that the high solubility of Ra was associated with ionic strength. However, careful consideration into the chemical composition of the Marcellus Shale produced fluids indicated the solubility of Ra was likely attributable to competition with other divalent cations (most profoundly by Ba), rather than ionic strength *per se*. Although wastes may not initially contain  $^{210}\text{Pb}$  and  $^{210}\text{Po}$ , due to radiochemical ingrowth processes, the environmental impacts of  $^{210}\text{Pb}$  and  $^{210}\text{Po}$  must be considered in wastes containing  $^{226}\text{Ra}$ . We found that  $(17 \pm 1\%)$  Pb associated with phases only extractable by strong acid (4 M  $\text{HNO}_3$ ), suggesting that radiochemical disequilibrium between  $^{226}\text{Ra}$  and  $^{210}\text{Pb}$  is likely to occur (even after accounting for losses from  $^{222}\text{Rn}$ ). Even greater portions ( $52 \pm 2\%$ ) of  $^{210}\text{Po}$  associated with strong acid-extractable phases, suggesting radiochemical disequilibrium between  $^{210}\text{Pb}$  and  $^{210}\text{Po}$  should be expected.

The key finding from this study is that Ra is more soluble in Marcellus Shale produced fluids than predicted by conventional adsorption-desorption models. Consistent with adsorption-desorption models generated with seawater matrices, we found that as ionic strength increases, the level of Ra in the aqueous phase increases. However, the level of Ra in the aqueous phases exceeded the amount predicted by seawater models. This suggests that the chemical matrix of produced fluids plays an important role in the



adsorption of Ra to environmental materials. Careful consideration of the levels of divalent cations in Marcellus Shale produced fluids, particularly Ba, is important for assessing the potential impacts of concentrated Marcellus Shale produced fluids releases into the environment.

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**References**

**References**

1. A. W. Nelson, E. S. Eitrheim, A. W. Knight, D. May, M. A. Mehrhoff, R. Shannon, R. Litman, W. C. Burnett, T. Z. Forbes and M. K. Schultz, *Environ. Health Perspect.*, 2015, **123**, 689-696.
2. N. R. Warner, C. A. Christie, R. B. Jackson and A. Vengosh, *Environ. Sci. Technol.*, 2013, **47**, 11849-11857.
3. T. Zhang, K. Gregory, R. W. Hammack and R. D. Vidic, *Environ. Sci. Technol.*, 2014, **48**, 4596-4603.
4. T. Zhang, D. Bain, R. Hammack and R. D. Vidic, *Environ. Sci. Technol.*, 2015, **49**, 2969-2976.
5. A. W. Nelson, D. May, A. W. Knight, E. S. Eitrheim, M. Mehrhoff, R. Shannon, R. Litman and M. K. Schultz, *Environ. Sci. Technol. Lett.*, 2014, **1**, 204-208.
6. V. Brown, *Environ. Health Perspect.*, 2014, **122**, A50-A55.

- 431 7. E. Rowan, M. Engle, C. Kirby and T. Kraemer, *US Geological Survey Scientific*  
432 *Investigations Report 2011*, 2011, **5135**.
- 433 8. L. O. Haluszczak, A. W. Rose and L. R. Kump, *Appl. Geochem.*, 2013, **28**, 55-61.
- 434 9. E. Barbot, N. S. Vidic, K. B. Gregory and R. D. Vidic, *Environ. Sci. Technol.*,  
435 2013, **47**, 2562-2569.
- 436 10. P. F. Ziemkiewicz and Y. Thomas He, *Chemosphere*, 2015, **134**, 224-231.
- 437 11. T. T. Phan, R. C. Capo, B. W. Stewart, J. R. Graney, J. D. Johnson, S. Sharma  
438 and J. Toro, *Appl. Geochem.*, 2015, **60**, 89-103.
- 439 12. E. C. Chapman, R. C. Capo, B. W. Stewart, R. S. Hedin, T. J. Weaver and H. M.  
440 Edenborn, *Appl. Geochem.*, 2013, **31**, 109-118.
- 441 13. W. M. Alley, J. A. Cherry, B. L. Parker and M. C. Ryan, *Trans., Am. Geophys.*  
442 *Union*, 2014, **95**, 264-264.
- 443 14. I. T. Webster, G. J. Hancock and A. S. Murray, *Geochim. Cosmochim. Acta*,  
444 1995, **59**, 2469-2476.
- 445 15. B. A. Powell, T. Miller and D. I. Kaplan, *J. S. C. Acad. Sci.*, 2015, **13**, 4.
- 446 16. M. Sajih, N. D. Bryan, F. R. Livens, D. J. Vaughan, M. Descostes, V.  
447 Phrommavanh, J. Nos and K. Morris, *Geochim. Cosmochim. Acta*, 2014, **146**,  
448 150-163.
- 449 17. M. E. Gonneea, P. J. Morris, H. Dulaiova and M. A. Charette, *Mar. Chem.*, 2008,  
450 **109**, 250-267.
- 451 18. N. C. Sturchio, J. L. Banner, C. M. Binz, L. B. Heraty and M. Musgrove, *Appl.*  
452 *Geochem.*, 2001, **16**, 109-122.
- 453 19. S. Krishnaswami, R. Bhushan and M. Baskaran, *Chem. Geol.: Isot. Geosci. Sect.*,  
454 1991, **87**, 125-136.
- 455 20. S. Tamamura, T. Takada, J. Tomita, S. Nagao, K. Fukushi and M. Yamamoto, *J*  
456 *Radioanal. Nucl. Chem.*, 2014, **299**, 569-575.

457 21. E. C. Chapman, R. C. Capo, B. W. Stewart, C. S. Kirby, R. W. Hammack, K. T.  
458 Schroeder and H. M. Edenborn, *Environ. Sci. Technol.*, 2012, **46**, 3545-3553.

459 22. P. Little and R. D. Wiffen, *Atmos. Environ.*, 1977, **11**, 437-447.

460 23. A. Federman, *Earth Island Journal*, 2012, **Winter 2012**,  
461 [http://www.earthisland.org/journal/index.php/eij/article/what\\_killed\\_dunkard\\_cre](http://www.earthisland.org/journal/index.php/eij/article/what_killed_dunkard_creek/)  
462 [ek/](http://www.earthisland.org/journal/index.php/eij/article/what_killed_dunkard_creek/), (accessed February 2016).

463 24. Soil Survey Staff, in in: Soil Survey Laboratory Methods Manual, U.S.  
464 Department of Agriculture, Natural Resources Conservation Service, National  
465 Soil Survey Center, Lincoln, Nebraska, USA, 1996, Soil Survey Investigations  
466 Report No. 42, Version 3.0 Particle-Size Analysis, Particles <2mm (pipet method)  
467 (3A), Air-Dry Samples (3A1), 35-37.

468 25. D.L. Bish and J.E. Post, *Reviews in Mineralogy*, Mineralogical Society of  
469 America, 20, 1989, *Modern Powder Diffraction*.

470 26. M. Schultz, W. Burnett, K. Inn, J. Thomas and Z. Lin, *J. Res. Nat. Inst. Stand.*  
471 *Technol.*, 1996, **101**, 707-715.

472 27. A. W. Knight, E. S. Eitrheim, A. W. Nelson and M. K. Schultz, *Nukleonika*, 2015,  
473 **60**, 837-845.

474 28. A. W. Knight, E. S. Eitrheim, A. W. Nelson, S. Nelson and M. K. Schultz, *J.*  
475 *Environ. Radioact.*, 2014, **134**, 66-74.

476 29. U. Schwertmann and R. M. Cornell, in *Iron Oxides in the Laboratory*, Wiley-  
477 VCH Verlag GmbH, 2007, 9, 113-119.

478 30. P. Refait and J.-M. Génin, *Corros. Sci.*, 1997, **39**, 539-553.

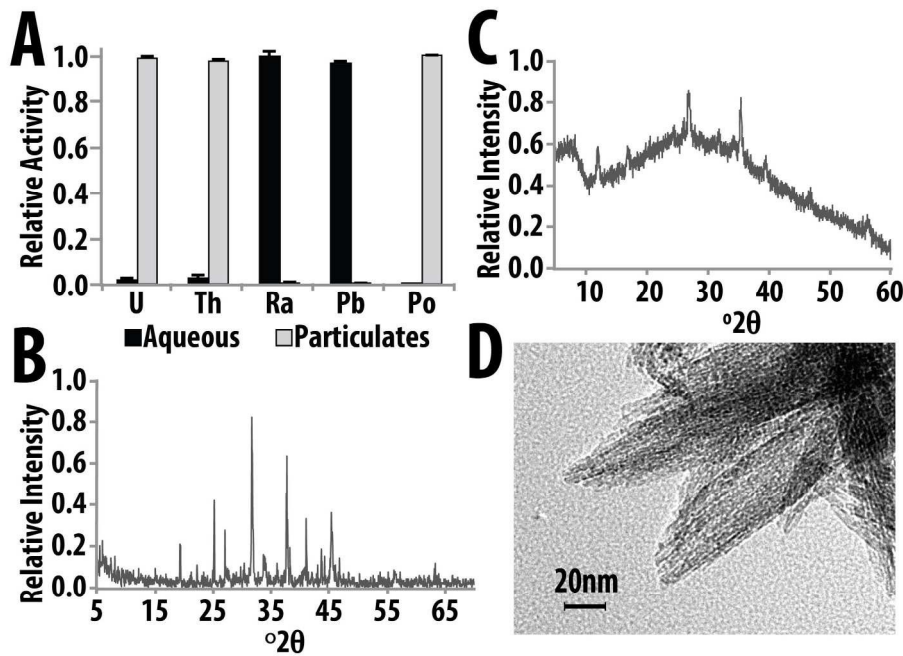
479 31. N. G. Holm, *Origins Life Evol. Biosphere*, 1985, **15**, 131-139.

480 32. M. Mohapatra, L. Mohapatra, S. Anand and B. K. Mishra, *J. Chem. Eng. Data*,  
481 2010, **55**, 1486-1491.

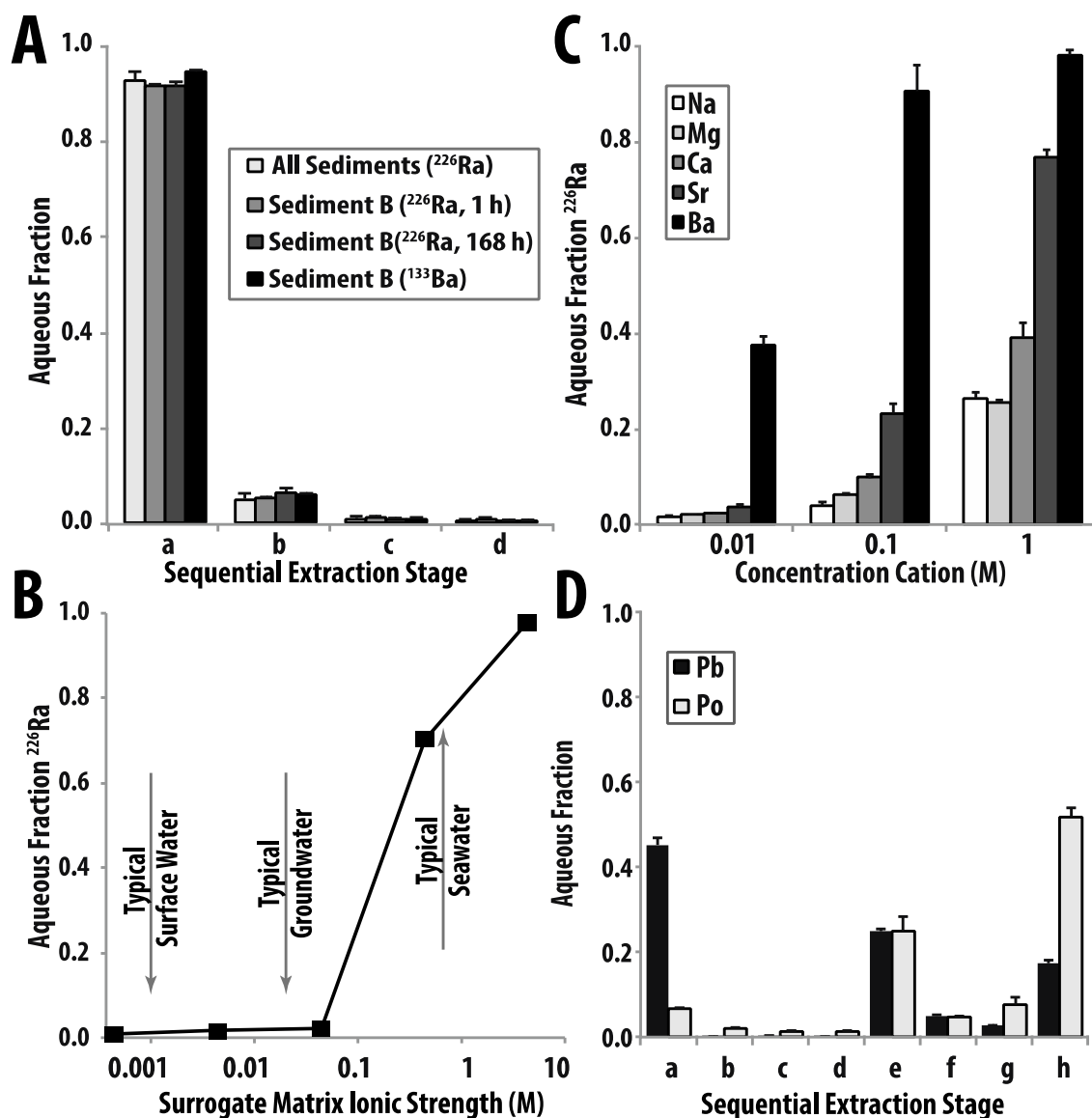
482 33. T. Zhang, R. W. Hammack and R. D. Vidic, *Environ. Sci. Technol.*, 2015, **49**,  
483 9347-9354.

- 1  
2  
3 484 34. D. Langmuir and A. C. Riese, *Geochim. Cosmochim. Acta*, 1985, **49**, 1593-1601.  
4  
5 485 35. R. L. Seiler and J. L. Wiemels, *Environ. Health Perspect.*, 2012, **120**, 1230-1237.  
6  
7  
8 486  
9 487  
10 488  
11 489  
12 490  
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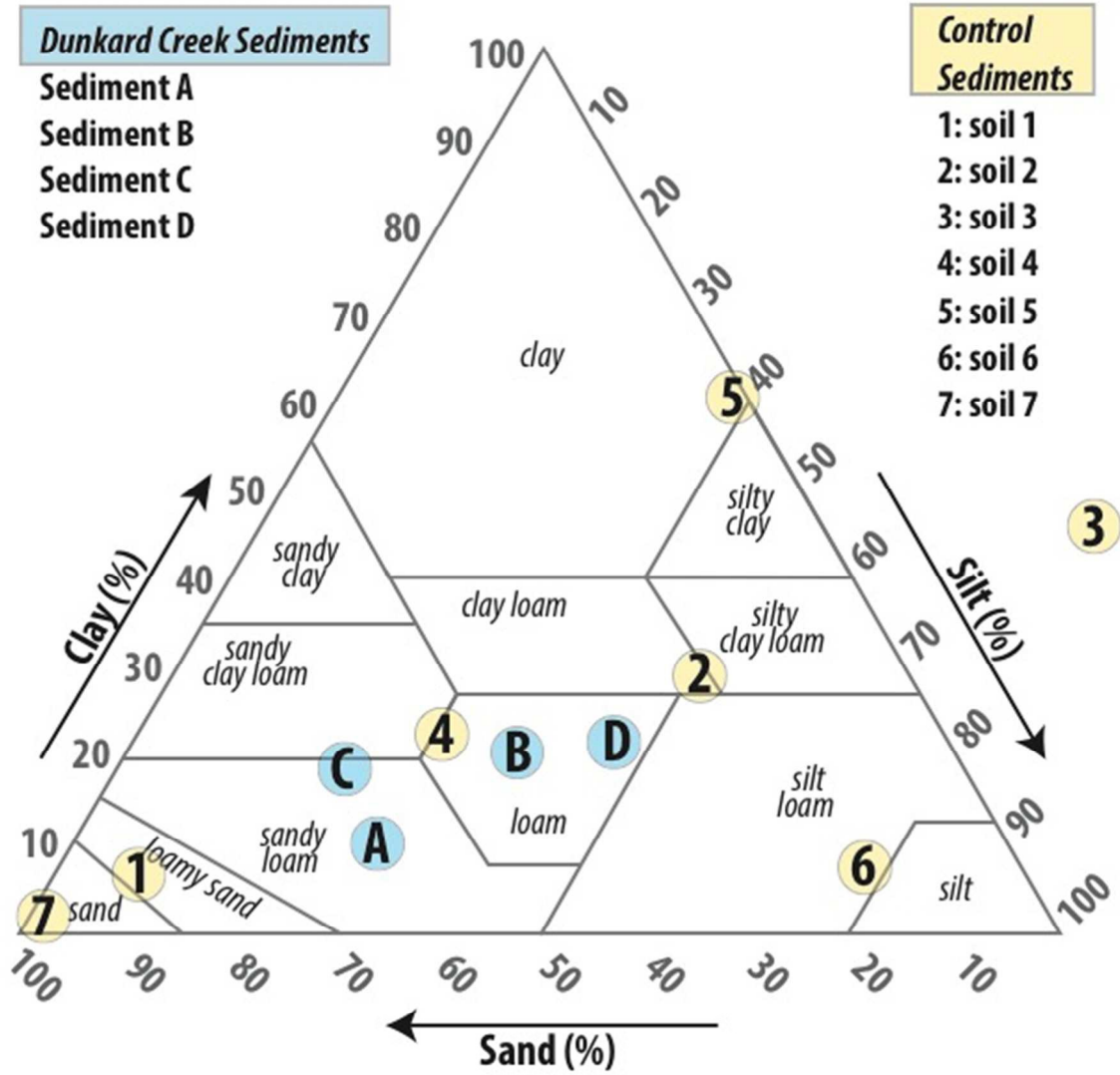
Figures and Figure Legends:



**Fig. 1:** Partitioning within Marcellus Shale produced fluids. (A) Mean recoveries (with standard deviations) of U, Th, Ra, Pb, and Po in aqueous or particulate phases of radioisotope traced, filtered (0.1  $\mu\text{m}$ ) Marcellus Shale produced fluids. (B) pXRD spectrum of evaporated (low heat), filtered (0.1  $\mu\text{m}$ ) produced fluids. (C) pXRD spectrum of filtered (0.1  $\mu\text{m}$ ) produced fluid particulates. (D) TEM image of filtered (0.1  $\mu\text{m}$ ) produced fluid particulates.



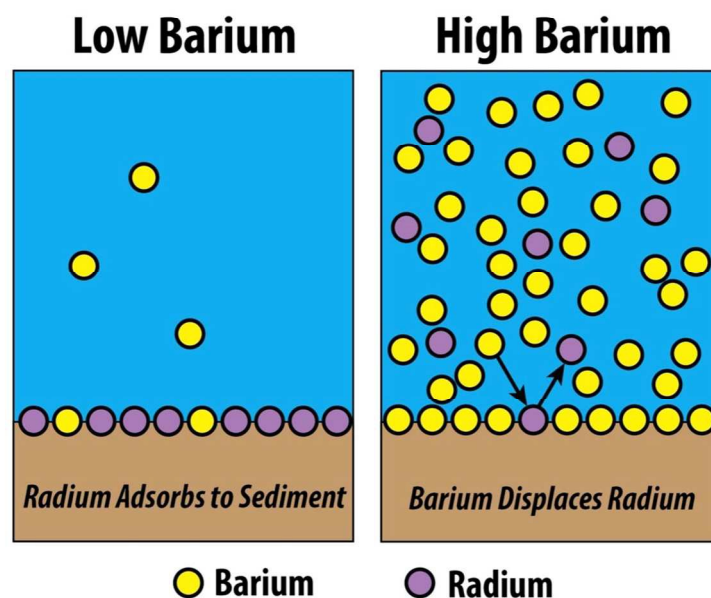
**Fig. 2:** Ra and Ra decay products partitioning between Marcellus Shale produced fluids (or surrogate) and sediments. (A) Sequential extractions of Ra or Ba on select sediments: the first three bars represent mean recoveries with standard deviations of  $^{226}\text{Ra}$  from sequential extractions by  $^{222}\text{Rn}$  emanation; values from left to right indicate Ra removal from 11 different sediments ( $n=3$ ), Dunkard Creek sediment B (1 hour,  $n=3$ ), Dunkard Creek sediment B (168 hour,  $n=4$ ). Fourth bar represents mean recovery with standard deviation of Ba (traced with  $^{133}\text{Ba}$ ) from Dunkard Creek sediment B ( $n=3$ ). (B) Aqueous fraction of surrogate fracturing matrix spiked with  $^{226}\text{Ra}$  (37 Bq) and incubated with Dunkard Creek sediment B at various ionic strengths; values represent means ( $n=3$ ) with standard deviations subsumed within squares. (C) Aqueous fraction of Na, Mg, Ca, Sr, or Ba chloride salts (0.01, 0.1, and 1 M) spiked with  $^{226}\text{Ra}$  (37 Bq) and incubated with Dunkard Creek sediment B; values represent means ( $n=3$ ) with standard deviations subsumed within squares. (D) Mean recovery with standard deviation of  $^{203}\text{Pb}$  ( $n=3$ ) and  $^{210}\text{Po}$  ( $n=3$ , traced with  $^{209}\text{Po}$ ).



**Fig 3.** Grain Size Triangle Plot of Sediments from Quaternary Materials Laboratory at the University of Iowa



## Effect of Barium on Radium Adsorption



**Fig. 4:** Theoretical depiction of Ba and Ra competition for adsorption on sediments in high and low Ba concentrations.