



**Particle size distribution of workplace aerosols in  
manganese alloy smelters applying personal sampling  
strategy**

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3 Toxicological studies have implicated particulate exposure as possibly playing a causal role in  
4 adverse cardiovascular effects. There is an opportunity to study these effects in occupational  
5 settings, but for the better understanding more information on time-specific workplace  
6 exposure to different particulate matter fractions is needed. The present work describes the  
7 results from personal size selective aerosol samplings in two manganese alloy smelters which  
8 will be later used in the epidemiological study which was initiated to examine the prevalence  
9 of early markers of cardiovascular disease with relation to exposure to different particulate  
10 matter fractions. Particle size distribution data is also essential for the better estimation of the  
11 deposited dose in the human respiratory tract as respiratory tract deposition is highly  
12 dependent on the aerosol particle size.  
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## Particle size distribution of workplace aerosols in manganese alloy smelters applying personal sampling strategy

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### Abstract

Air samples were collected by personal sampling with five stage Sioutas cascade impactors and respirable cyclones in parallel among tappers and crane operators in two manganese (Mn) alloy smelters in Norway to investigate PM fractions. The mass concentrations of PM collected by the impactors and the respirable cyclones were critically evaluated by comparing the results of the parallel measurements. The geometric mean (GM) mass concentrations of the respirable fraction and the <10  $\mu\text{m}$  PM fraction were 0.18 and 0.39  $\text{mg m}^{-3}$ , respectively. Particle size distributions were determined using the impactor data in the range from 0 to 10  $\mu\text{m}$  and by stationary measurements by a Scanning Mobility Particle Sizer in the range from 10 to 487 nm. On average 50% of the particulate mass in the Mn alloy smelters was in the range from 2.5 to 10  $\mu\text{m}$ , while the rest was distributed between the lower stages of the impactors. On average 15% of the particulate mass was found in the <0.25  $\mu\text{m}$  PM fraction. The comparisons of the different PM fraction mass concentrations related to different work tasks or different workplaces, showed in many cases statistically significant differences, however, the particle size distribution of PM in the fraction <10  $\mu\text{m}$   $d_{ae}$  was independent of the plant, furnace or work task.

### Keywords

respirable fraction, particulate matter fractions, Sioutas cascade impactor, SMPS, tappers, crane operators

### Introduction

Numerous epidemiological studies have shown associations between exposure to particulate urban air pollution and increased morbidity and mortality from cardiovascular diseases.<sup>1-6</sup> Both short-term and long-term  $\text{PM}_{2.5}$  exposures has been associated with increased incidence of cardiovascular diseases,

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3 but the character of the exposure-response relationship is still uncertain.<sup>7,8</sup> Toxicological studies have  
4 implicated ultrafine particles (UFP, diameter  $\leq 100$  nm) as possibly playing a causal role in the adverse  
5 cardiovascular effects.<sup>9,10</sup> Limited evidence of these relationships has emerged from occupational  
6 settings.<sup>11</sup> Interestingly, Costello *et al.* has shown lately that recent exposure, but not cumulative  
7 exposure to PM<sub>2.5</sub> can be a risk factor for incident ischemic heart disease among aluminum production  
8 workers.<sup>12</sup> More information on time-specific workplace exposure to different PM fractions (including  
9 UFPs) is therefore needed. The manganese (Mn) alloy smelter workers participating in this study are  
10 exposed to PM with Mn content. It is well known that Mn compounds can have neurological effects,<sup>13</sup>  
11 therefore Mn concentration in workplace air is strictly regulated. The current exposure limit values for  
12 inhalable and respirable inorganic Mn compounds in workplace air in Norway are 1.0 and 0.1 mg m<sup>-3</sup>,  
13 respectively.<sup>14</sup>

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20 The health effects resulting from deposition of particulate matter (PM) in the respiratory tract  
21 depend on the dose received, the site of deposition and the body's response to the deposited particles.<sup>15</sup>  
22 One important goal of workplace aerosol measurement should therefore be to ascertain the effective  
23 dose of PM delivered to the lungs. Particle deposition models show that the respiratory tract deposition  
24 is highly dependent on the aerosol particle size.<sup>16,17</sup> For example up to 55% of 20 nm particles with  
25 density of 1 g cm<sup>-3</sup> and spherical shape can deposit in the alveolar region, while the deposition  
26 efficiency of 500 nm particles remains under 10% according to the model given by Bartley and  
27 Vincent.<sup>17</sup>

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32 These models, however, are not applicable for a better dose estimation without the information  
33 on particle size distribution which is usually not available from conventional workplace aerosol  
34 measurements. Cascade impactors are able to provide sharp particle size classifications.<sup>18</sup> Thus, they  
35 can be useful tools when such information is required. Such equipment does have limitations in risk  
36 assessment. One of these is that the information on the collected mass of particles with different sizes  
37 is only one of the most important three metrics which might be used as relevant indicators of worker's  
38 particle exposure, others being particle number and surface area.<sup>19</sup> The evaluation of particle  
39 deposition in the human respiratory tract should also consider the hygroscopic growth of aerosol  
40 particles upon inhalation which has been previously shown by a number of authors.<sup>20-22</sup> Size dependent  
41 chemical characterization may also give valuable information if obtained at the same time.

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47 Cascade impactors have been used for personal and stationary sampling to measure the  
48 particle size distribution in workplace aerosols. In a study personal dust exposure levels and the dust  
49 particle size distribution was measured during various agricultural operations in California by applying  
50 four-stage Sierra Marple 294 cascade impactors.<sup>23</sup> Another study characterized exposures to copper  
51 and zinc oxide using single jet personal cascade impactors with five to six stages in a nonferrous  
52 foundry and compared the results with previous findings obtained using cyclones.<sup>24</sup> Dufresne *et al.*  
53 collected breathing zone samples with an 8-stage Sierra impactor in a magnesium foundry and three  
54 aluminium smelters to assess the distribution of the PM mass and the beryllium content between the  
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3 different size fractions.<sup>25</sup> Particle size distributions of oil mists in fastener manufacturing industry were  
4 determined by using a modified Marple 8-stage cascade impactor.<sup>26</sup> The Sioutas cascade impactor  
5 equipped with quartz-fiber substrates and after-filter was used by Birch *et al.* to collect personal  
6 samples during carbon nanofiber production for the determination of elemental carbon and organic  
7 carbon.<sup>27</sup> The latter model offers a good separation with its after-filter and four stages with cut points  
8 ranging from 0.25 to 10  $\mu\text{m}$  aerodynamic diameter ( $d_{ae}$ ) and it operates at a high flow rate, which can  
9 be maintained by a lightweight personal pump.<sup>28</sup>

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14 There are also applications of stationary cascade impactors in workplace aerosol  
15 characterisation. One example is when Vincent *et al.* used a modified Andersen sampler for the  
16 particle size measurement of aerosols in primary nickel production industry workplaces.<sup>29</sup> Berlinger *et*  
17 *al.* and Chang *et al.* used a stationary 11-stage Berner cascade impactor and a compact cascade  
18 impactor (Harvard CCI) with two stages and a backup filter, respectively for the physicochemical  
19 characterization of welding fumes.<sup>30,31</sup>

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24 Exposure among employees in smelters in Norway has been well assessed both quantitatively  
25 and qualitatively, however, particle size distribution by personal sampling has not yet been  
26 investigated.<sup>32</sup> There has also been little focus on the ultrafine size fraction of PM, although high  
27 exposure levels in the smelting industry could provide a good opportunity to investigate the  
28 association between PM exposure and established cardiovascular disease. The scientific evidence of  
29 such association in occupational exposure situations is still limited.<sup>11</sup>

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34 Because of the importance of the ultrafine particles in the possible cardiovascular effects, not  
35 only personal impactors but personal nanoparticle samplers were considered to be used to assess  
36 smelter workers' exposure to PM. Personal nanoparticle sampler (PENS) was developed in 2011  
37 which consists of a respirable cyclone and a micro-orifice impactor with a 50% cut-off aerodynamic  
38 diameter of 4  $\mu\text{m}$  and 100 nm, respectively.<sup>33</sup> Personal nanoparticle respiratory deposition (NRD)  
39 sampler was also developed in the same year.<sup>34</sup> It collects particles smaller than 300 nm similarly to  
40 their deposition in the human respiratory tract. At the end, however, personal impactors were chosen  
41 to be used at the smelters because these give information on more than only one PM size fraction.

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A study in metal smelters in Norway was initiated to examine the prevalence of early markers  
of cardiovascular disease with relation to exposure to different PM fractions. The present work  
describes the results from personal aerosol samplings in two Mn alloy smelters applying Sioutas  
cascade impactors and respirable cyclones in parallel. The results from the epidemiological study  
where these exposure measurements are to be used will be published later.

## Materials and methods

### *Plants and subjects*

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3 Air samples were collected in two Mn alloy smelters in Norway. Both plants have two smelting  
4 furnaces. Mn alloy smelter 1 is producing silicomanganese (furnace 1) and ferromanganese (furnace 2)  
5 alloys, while Mn alloy smelter 2 is producing ferromanganese only (furnace 3 and 4). Personal air  
6 samples were collected among 38 furnace workers. The main work tasks of the workers were to  
7 control the tapping process (tappers) or to operate the cranes (crane operators). Both tappers and crane  
8 operators were working between 45 and 75 minutes on the tap floor before, during and after tapping of  
9 the molten alloy which was done 3 or 4 times during an 8-hour shift. Tappers usually worked within 3-  
10 5 m from the tapping hole, but they could also sit in a pressurized cabin while following the tapping  
11 process. Crane operators performed their tasks further (5-10 m) from the tapping hole than tappers and  
12 worked mainly after the end of the tapping process. When tappers and crane operators were not  
13 working on the tap floor, they took a rest in the control area where they were not exposed to PM.  
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### 20 21 *Air sampling*

22 Each worker carried three air samplers. A 5-stage Sioutas cascade impactor (SKC, Eighty Four, PA,  
23 USA) ran in parallel with a respirable cyclone (JS Holdings, Stevenage, UK) collecting the respirable  
24 PM fraction (*parallel* respirable fraction) for 6 to 8 hours. Full-shift (approx. 8 hours) air samples were  
25 collected in the breathing zone of workers by a second respirable cyclone (*full-shift* respirable  
26 fraction).  
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30 Taking into account the benefits and disadvantages of the different impactor types, the Sioutas  
31 cascade impactor for personal sampling was chosen. The Sioutas cascade impactor equipped with 25-  
32 mm polytetrafluoroethylene (PTFE) substrates of 0.5  $\mu\text{m}$  pore-size (Pall Corporation, Port  
33 Washington, NY, USA) was operated at 9  $\text{L min}^{-1}$ . At this flow rate particles are separated and  
34 collected on the impactor stages from the top to the bottom in the following aerodynamic particle  
35 diameter ranges (in  $\mu\text{m}$ ): 10–2.5, 2.5–1.0, 1.0–0.5, 0.5–0.25. Particles below the 0.25  $\mu\text{m}$  cut-point of  
36 the last stage are collected on a 2.0  $\mu\text{m}$  pore-size 37-mm PTFE after-filter (SKC, Eighty Four, PA,  
37 USA). More detailed description of the Sioutas cascade impactors have been published earlier.<sup>28,35</sup>  
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43 The respirable cyclones were operated at a flow rate of 2.2  $\text{L min}^{-1}$  that results in a 50% cut-  
44 point diameter of 4.0  $\mu\text{m}$ . The 37-mm plastic cassettes in the cyclones were equipped with 5.0  $\mu\text{m}$   
45 pore-size PVC membrane filters (Millipore Corp., Billerica, MA, USA).  
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47 The 2.2  $\text{L min}^{-1}$  flow required for the respirable samplers was maintained by in-house built  
48 PS103 model personal sampling pumps (National Institute of Occupational Health, Oslo, Norway).  
49 Leland Legacy model high flow personal sampling pumps (SKC, Eighty Four, PA, USA) were applied  
50 to maintain the 9  $\text{L min}^{-1}$  flow for the impactors. These pumps automatically cut-out at a pressure drop  
51 of approximately 500  $\text{mmH}_2\text{O}$ , and in most cases they were unable to operate through the entire 8-  
52 hour shift, stopping after 6-7 hours. Therefore the sampling with the impactors and the cyclones  
53 running in parallel was intentionally stopped and the sampling time was recorded 6-7 hours after the  
54 start of sampling. A second cyclone was used to sample the entire work-shift. The ratio between the  
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PM mass concentrations obtained by the two cyclones was used to adjust data from the impactor to estimate full-shift mass concentrations. In a few cases when the PM mass concentrations were not expected to reach the level which could cause the cut-out of the impactor pump, all the three samplers were let to operate in the whole work shift.

Particle size distributions of the aerosol in the smelters were in addition examined by a scanning mobility particle sizer (SMPS) instrument (model 3034, TSI Inc., Shoreview, USA) in the range from 10 to 487 nm. The SMPS instrument was placed at 3-4 meters from the furnace and run for 6-8 hours during one day in both plants.

### *Analysis*

The collected aerosol particulate masses were determined gravimetrically by a six-place Sartorius Micro model MC5 balance (Sartorius AG, Göttingen, Germany) in a weighing room dedicated to low filter mass measurements, under controlled relative humidity (40±2%) and temperature (20±1 °C) conditions. The balances were calibrated daily. The accuracy and precision of the mass determinations were assessed by weighing certified reference masses (19.989±0.030 and 49.953±0.040mg). The mass detection limits (DLs) calculated as 3 times standard deviation of all field blanks were below 0.01 mg for all kind of substrates and filters used in the study. Both before and after exposure the substrates and filters were stored in the weighing room for at least two days to be acclimatized prior to the weighing procedure.

### *Calculation of the respirable PM mass concentrations from the impactor data*

In order to compare the mass concentrations obtained by the respirable cyclones and the cascade impactors directly, mass concentrations of the respirable PM fraction were calculated from the impactor data. By these calculations sampling efficiencies given by the International Organization for Standard respirable convention were used and the following equations were applied.<sup>36,37</sup>

$$S_{RF_i} = \frac{SE_{LL} + 4 \times SE_{MP} + SE_{UL}}{6} \times S_i \quad (1)$$

where  $S_{RF_i}$  is the mass concentration of the respirable PM collected on the stage  $i$ ; SE is the sampling efficiency according to ISO 7708 at the lower limit (LL), the midpoint size (MP) and the upper limit (UL) of the stage interval;  $S_i$  is the mass concentration of PM collected on the stage  $i$ . The mass concentration of the respirable PM fraction ( $C_{RF}$ ) is calculated by summing up the mass concentrations of the respirable PM collected on all five stages.

$$C_{RF} = \sum_{i=1}^5 S_{RF_i} \quad (2).$$

### Statistics

Logarithmic and arc sine transformation were applied for concentrations and proportions, respectively.<sup>38</sup> Independent sample and paired sample t-tests were performed to compare two independent job groups (e.g., tappers and crane operators) and two dependent variables (e.g., two parallel samples), respectively. Analysis of variance (ANOVA) was used when more than two groups were compared, and the least square difference was calculated in order to assess which groups differed from each other. A two-tailed  $p$ -value  $<0.05$  was considered as the level of statistical significance. IBM® SPSS® statistical programme, version 21.0 (SPSS Inc, Chicago, Illinois, United States) was used for the statistical calculations.

### Results and discussion

Mass concentrations of the different PM size fractions collected by the impactors were compared to mass concentrations of the respirable fraction collected in parallel by respirable cyclones as the first step of the evaluation. Altogether nine PM size fractions were applied in these comparisons, also those ( $<10\ \mu\text{m}$ ,  $<2.5\ \mu\text{m}$ ,  $<1.0\ \mu\text{m}$ ,  $<0.5\ \mu\text{m}$ ) which were calculated from the original five fractions collected on the four impactor stages and the after filter. Pearson's correlation coefficients between the mass concentrations of these fractions and the respirable aerosol fraction collected in parallel are summarised in Table 1. There were statistically significantly high correlations ( $0.61 < r < 0.88$ ,  $p < 0.001$ ,  $N = 37-38$ ) between the mass concentrations of all PM size fractions collected by the impactor and the respirable fraction in the Mn alloy smelters. The respirable fraction calculated from the impactor data was 1.3 times higher than the respirable fraction collected in parallel by the respirable cyclones (Fig. 1). This ratio is within the expectations and suggests that both sampling devices worked according to their specifications.

### Workplace particulate matter mass concentrations

Mass concentrations of the different PM fractions measured in the Mn alloy smelters are summarised in Table 2. The GM mass concentrations of the respirable fraction and the  $<10\ \mu\text{m}$  PM fraction were 0.18 and 0.39  $\text{mg m}^{-3}$  respectively. Johnsen *et al.* found 1.6  $\text{mg m}^{-3}$  GM dust mass concentrations in the Norwegian silicomanganese, ferromanganese and ferrochromium production group, although, they applied 'total dust' sampling cassette which has quite different particle collection characteristics compared to the samplers used in this study.<sup>32</sup> The mass concentrations of the *parallel* and *full-shift* respirable fractions ( $N = 38$ ) correlated very well ( $r = 0.95$ ,  $p < 0.001$ ) in spite of the slightly different sampling times.

It was previously found that the GM concentration of respirable Mn was 28  $\mu\text{g m}^{-3}$  in the same Mn alloy smelters which were investigated in this study.<sup>39</sup> Another study characterising workplace aerosols in the Norwegian Mn alloy production industry by electron microscopy reported that the

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3 submicron size fraction is dominated by MnO and MnSi particles in FeMn and SiMn production,  
4 respectively.<sup>40</sup> The chemical composition of the different PM fractions in the current study will be  
5 analysed later and internationally published in detail elsewhere.  
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#### 8 9 *Particle mass and number size distributions*

10 Mass concentration distributions of the PM in the first and second Mn alloy smelter and the average  
11 mass concentration distribution in both smelters are summarised in Table 3 together with the mass  
12 median diameters (MMAD) and geometric standard deviations (GSD). The MMAD and GSD values  
13 were calculated according to Christopher *et al.*<sup>41</sup> Values estimated under lognormal distribution  
14 assumption are given in the electronic supplement, for sake of completeness. Differential mass  
15 distribution is represented by histogram and “middle of cut-off diameters” method in Fig. 2 as it was  
16 described by Majoral *et al.*<sup>42</sup> On average 50% of the particulate mass in the Mn alloy smelters was in  
17 the range from 2.5 to 10  $\mu\text{m}$ , the rest was distributed between the lower stages of the impactors. On  
18 average 15% of the particulate mass was found in the PM fraction  $<0.25 \mu\text{m}$ .  
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24 The average number and mass size distributions of the PM in the Mn alloy smelters assessed  
25 by the SMPS in the size range from 10 to 487 nm are shown in Fig. 3 and in the electronic  
26 supplement, respectively. In addition, the distribution parameters are given in table S1. The mass  
27 concentrations are not accurate as the density of the particles is not known. The distribution itself,  
28 however, is not affected by the density of the particles as long as it is not a function of particle size,  
29 which is a reasonable assumption for this size range. The particle number and mass distribution curves  
30 look more or less similar close to the different furnaces where the SMPS was running. At the same  
31 time the particle number and mass concentration levels are more different by the different furnaces.  
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37 If the particle size mass distribution in the Mn alloy smelters is compared to the particle  
38 deposition efficiency curves published by Bartley and Vincent, it can be suggested that the deposition  
39 of PM in the workers' lungs may be relatively high compared to the total mass of the fraction  $<10 \mu\text{m}$   
40  $d_{ae}$ , because high percentage of particles can be found in those particle mass fractions which have the  
41 highest deposition efficiencies in the alveolar region.<sup>17</sup>  
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#### 46 *Comparisons of different workplaces and job groups*

47 The measured mass concentrations of the five PM fractions collected by the impactor and the  
48 respirable fraction were compared between the two Mn alloy smelters. The mass concentrations of the  
49 *parallel* respirable fraction ( $GM_{S1} = 0.14 \text{ mg m}^{-3}$ ,  $GM_{S2} = 0.22 \text{ mg m}^{-3}$ ,  $p = 0.024$ ), the 2.5–10  $\mu\text{m}$   
50 ( $GM_{S1} = 0.15 \text{ mg m}^{-3}$ ,  $GM_{S2} = 0.24 \text{ mg m}^{-3}$ ,  $p = 0.035$ ) and  $<10 \mu\text{m}$  ( $GM_{S1} = 0.32 \text{ mg m}^{-3}$ ,  
51  $GM_{S2} = 0.49 \text{ mg m}^{-3}$ ,  $p = 0.050$ ) fractions were statistically significantly higher in Mn alloy smelter 2  
52 compared to Mn alloy smelter 1. Otherwise the differences were not of statistical significance. The  
53 differences in the mass concentrations of the *parallel* respirable fractions might be due to the slight  
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3 differences in the duration of sampling periods in the two smelters. The differences in the mass  
4 concentrations of the 2.5–10  $\mu\text{m}$  and  $<10 \mu\text{m}$  fractions might be explained by the different practises  
5 for sweeping of the settled dust in the two smelters which might cause slight differences in the mass  
6 concentration of the coarse particles.  
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9 Workers at furnace 1 in Mn alloy smelter 1, the only furnace producing silicomanganese, are  
10 exposed to one third of the mass concentrations of 2.5–10  $\mu\text{m}$  ( $\text{GM}_{\text{F1}} = 0.12 \text{ mg m}^{-3}$ ,  
11  $\text{GM}_{\text{F3}} = 0.31 \text{ mg m}^{-3}$ ,  $p = 0.002$ ), 1.0–2.5  $\mu\text{m}$ , 0.25–1.0  $\mu\text{m}$ , 0–0.25  $\mu\text{m}$  ( $\text{GM}_{\text{F1}} = 0.027 \text{ mg m}^{-3}$ ,  
12  $\text{GM}_{\text{F3}} = 0.080 \text{ mg m}^{-3}$ ,  $p = 0.019$ ),  $<10 \mu\text{m}$  and both respirable PM fractions compared to workers at  
13 furnace 3 in Mn alloy smelter 2 producing ferromanganese. The measured mass concentrations of the  
14 *parallel* respirable fraction at furnace 1 was half of the mass concentrations at furnace 4  
15 ( $\text{GM}_{\text{F1}} = 0.11 \text{ mg m}^{-3}$ ,  $\text{GM}_{\text{F4}} = 0.20 \text{ mg m}^{-3}$ ,  $p = 0.034$ ). The differences in workers' exposure to PM at  
16 furnace 1 and 3 might be explained by the different productions; nevertheless furnace 1 was quite  
17 similar to the other two furnaces which are also producing ferromanganese like furnace 3.  
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20 A not completely unexpected finding was that tappers are exposed to statistically significantly  
21 higher mass concentrations of the PM fractions compared to crane operators (Table 5). The 2-3 times  
22 higher mass concentrations of the different PM fractions may be due to differences in duration of work  
23 tasks and distance from the source of PM.  
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26 No statistically significant differences were found between tappers and crane operators by  
27 comparing the proportions of the five PM fractions collected by the impactor. The proportions of the  
28 PM collected by the impactors were comparable among workers exposed at different furnaces or  
29 working in different Mn alloy production plants. These findings suggest that the distribution of the  
30 particles that furnace workers are exposed to is independent of the plant, furnace or work task. On the  
31 other hand one should also take into account the uncertainty of the particle mass distribution  
32 measurements when applying cascade impactors. If the measurement uncertainty is too high, a  
33 potential difference between the distributions cannot be revealed.  
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### 43 Conclusions

44 The GM mass concentrations of the respirable fraction and the  $<10 \mu\text{m}$  PM fraction in the manganese  
45 alloy smelters were 0.18 and 0.39  $\text{mg m}^{-3}$  respectively. On average 50% of the particulate ( $<10 \mu\text{m}$ )  
46 mass was in the range from 2.5 to 10  $\mu\text{m}$ , the rest was distributed between the lower stages of the  
47 impactors. On average 15% of the particulate mass was found in the PM fraction  $<0.25 \mu\text{m}$ . The  
48 comparisons of the different PM fraction mass concentrations related to different work tasks or  
49 different workplaces, showed in many cases statistically significant differences, however, the particle  
50 size distribution of PM in the fraction  $<10 \mu\text{m}$   $d_{\text{ae}}$  was independent of the plant, furnace or work task.  
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8 data.  
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## 12 13 14 References

- 15  
16 1 D. W. Dockery, C. A. Pope, X. P. Xu, J. D. Spengler, J. H. Ware, M. E. Fay, B. G. Ferris and  
17 F. E. Speizer, *N Engl J Med*, 1993, **329**, 1753-1759.
- 18  
19 2 C. A. Pope, D. W. Dockery and J. Schwartz, *Inhal Toxicol*, 1995, **7**, 1-18.
- 20  
21 3 O. Naess, P. Nafstad, G. Aamodt, B. Claussen and P. Rosland, *Am J Epidemiol*, 2007, **165**,  
22 435-443.
- 23  
24 4 H. Chen, M. S. Goldberg and P. J. Villeneuve, *Rev Environ Health*, 2008, **23**, 243-297.
- 25  
26 5 R. D. Brook, S. Rajagopalan, C. A. Pope, J. R. Brook, A. Bhatnagar, A. V. Diez-Roux, F.  
27 Holguin, Y. L. Hong, R. V. Luepker, M. A. Mittleman, A. Peters, D. Siscovick, S. C. Smith,  
28 L. Whitsel and J. D. Kaufman, *Circulation*, 2010, **121**, 2331-2378.
- 29  
30 6 G. Cesaroni, F. Forastiere, M. Stafoggia, Z. J. Andersen, C. Badaloni, R. Beelen, B.  
31 Caracciolo, U. de Faire, R. Erbel, K. T. Eriksen, L. Fratiglioni, C. Galassi, R. Hampel, M.  
32 Heier, F. Hennig, A. Hilding, B. Hoffmann, D. Houthuijs, K. H. Jockel, M. Korek, T. Lanki,  
33 K. Leander, P. K. E. Magnusson, E. Migliore, C. G. Ostenson, K. Overvad, N. L. Pedersen, J.  
34 J. Pekkanen, J. Penell, G. Pershagen, A. Pyko, O. Raaschou-Nielsen, A. Ranzi, F. Ricceri, C.  
35 Sacerdote, V. Salomaa, W. Swart, A. W. Turunen, P. Vineis, G. Weinmayr, K. Wolf, K. de  
36 Hoogh, G. Hoek, B. Brunekreef and A. Peters, *Brit Med J*, 2014, **348**, 16.
- 37  
38 7 R. J. Delfino, C. Sioutas and S. Malik, *Environ Health Perspect*, 2005, **113**, 934-946.
- 39  
40 8 C. A. Pope, III, R. T. Burnett, D. Krewski, M. Jerrett, Y. Shi, E. E. Calle and M. J. Thun,  
41 *Circulation*, 2009, **120**, 941-948.
- 42  
43 9 W. G. Kreyling, M. Semmler and W. Moller, *J Aerosol Med*, 2004, **17**, 140-152.
- 44  
45 10 G. Oberdorster, E. Oberdorster and J. Oberdorster, *Environ Health Perspect*, 2005, **113**, 823-  
46 839.
- 47  
48 11 S. C. Fang, A. Cassidy and D. C. Christiani, *Int J Environ Res Public Health*, 2010, **7**, 1773-  
49 1806.
- 50  
51 12 S. Costello, D. M. Brown, E. M. Noth, L. Cantley, M. D. Slade, B. Tessier-Sherman, S. K.  
52 Hammond, E. A. Eisen and M. R. Cullen, *J Expo Sci Env Epid*, 2014, **24**, 82-88.
- 53  
54 13 D. E. McMillan, *Neurotoxicology*, 1999, **20**, 499-507.
- 55  
56 14 Regulations concerning Action and Limit Values, Norwegian Labour Inspection Authority,  
57 2014, Order No. 704-ENG  
58  
59  
60

- 1  
2  
3 15 J. C. Volkwein, A. D. Maynard, M. Harper, in *Aerosol Measurement: Principles, Techniques, and Applications*, ed. P. A. Kulkarni, P. A. Baron, K. Willeke, John Wiley, Chichester, UK, 4  
5 3<sup>rd</sup> edn., 2011, Workplace aerosol measurements, p. 572.  
6  
7 16 ICRP Human respiratory tract model for radiological protection, *Annals of the International*  
8 *Commission on Radiological Protection (ICRP)*, Pergamon Press, Oxford, UK, 1994  
9  
10 17 D. L. Bartley and J. H. Vincent, *Ann Occup Hyg*, 2011, **55**, 696-709.  
11  
12 18 V. A. Marple, *Aerosol Sci Technol*, 2004, **38**, 247-292.  
13  
14 19 P. T. O'Shaughnessy, *Environ Sci Process Impacts*, 2013, **15**, 49-62.  
15  
16 20 A. Sorooshian, J. Csavina, T. Shinger, S. Dey, F. J. Brechtel, A. E. Saez and E. Betterton,  
17 *Environ Sci Technol*, 2012, **46**, 9473-9480.  
18  
19 21 S. Weinbruch, N. Benker, W. Koch W, M. Ebert, P. A. Drablos, N. P. Skaugset, D. G.  
20 Ellingsen and Y. Thomassen, *J Environ Monit*, 2010, **12**, 448-454.  
21  
22 22 M. Inerle-Hof, S. Weinbruch, M. Ebert and Y. Thomassen, *J Environ Monit*, 2007, **9**, 301-  
23 306.  
24  
25 23 M. J. Nieuwenhuijsen, H. Kruize and M. B. Schenker, *Am Ind Hyg Assoc J*, 1998, **59**, 34-38.  
26  
27 24 H. J. Cohen and B. J. Powers, *Am Ind Hyg Assoc J*, 2000, **61**, 422-430.  
28  
29 25 A. Dufresne, C. Dion, S. Viau, Y. Cloutier and G. Perrault, *J Occup Environ Hyg*, 2009, **6**,  
30 687-697.  
31  
32 26 M. R. Chen, P. J. Tsai, C. C. Chang, T. S. Shih, W. J. Lee and P. C. Liao, *J Hazard Mater*,  
33 2007, **146**, 393-398.  
34  
35 27 M. E. Birch, B. K. Ku, D. E. Evans and T. A. Ruda-Eberenz, *Ann Occup Hyg*, 2011, **55**, 1016-  
36 1036.  
37  
38 28 C. Misra, M. Singh, S. Shen, C. Sioutas and P. A. Hall, *J Aerosol Sci*, 2002, **33**, 1027-1047.  
39  
40 29 J. H. Vincent, G. Ramachandran and S. M. Kerr, *J Environ Monit*, 2001, **3**, 565-574.  
41  
42 30 B. Berlinger, N. Benker, S. Weinbruch, B. L'Vov, M. Ebert, W. Koch, D. G. Ellingsen and Y.  
43 Thomassen, *Anal Bioanal Chem*, 2011, **399**, 1773-1780.  
44  
45 31 C. Chang, P. Demokritou, M. Shafer and D. Christiani, *Environ Sci Process Impacts*, 2013,  
46 **15**, 214-224.  
47  
48 32 H. L. Johnsen, S. M. Hetland, J. S. Benth, J. Kongerud and V. Soyseth, *Ann Occup Hyg*, 2008,  
49 **52**, 623-633.  
50  
51 33 C. J. Tsai, C. N. Liu, S. M. Hung, S. C. Chen, S. N. Uang, Y. S. Cheng and Y. Zhou, *Environ*  
52 *Sci Technol*, 2012, **46**, 4546-4552.  
53  
54 34 L. G. Cena, T. R. Anthony and T. M. Peters, , *Environ Sci Technol*, 2011, **45**, 6483-6490.  
55  
56 35 M. Singh, C. Misra and C. Sioutas, *Atmos Environ*, 2003, **37**, 4781-4793.  
57  
58 36 ISO 7708: Air quality - Particle size fraction definitions for health-related sampling,  
59 International Organization for Standardization, Geneva, Switzerland, 1995  
60

- 1  
2  
3 37 W. C. Hinds, in *Cascade impactor: Sampling and data analysis*, ed. J. P. Lodge and T. L.  
4 Chan, American Industrial Hygiene Association, Akron, Ohio, USA, 1<sup>st</sup> edn., 1986, Data  
5 analysis, p. 45-77.  
6  
7 38 G. W. Snedecor, W. G. Cochran, *Statistical methods*, Iowa State University Press, Ames,  
8 Iowa, USA, 1989, 8<sup>th</sup> edn., p. 289-290.  
9  
10 39 D. G. Ellingsen, S. M. Hetland and Y. Thomassen, *J Environ Monit*, 2003, **5**, 84–90.  
11  
12 40 K. Gjønnnes, A. Skogstad, S. Hetland, D. G. Ellingsen, Y. Thomassen and S. Weinbruch, *Anal*  
13 *Bioanal Chem*, 2011, **399**, 1011–1020.  
14  
15 41 J. D. Christopher, M. Dey, S. Lyapustin, J. P. Mitchell, T. P. Tougas, M. Van Oort, H.  
16 Strickland, and B. Wyka, *Pharm Forum*, 2010, **36**, 812-823.  
17  
18 42 C. Majoral, A. Le Pape, P. Diot and L. Vecellio, *Aerosol Sci Technol*, 2006, **40**, 672-682.  
19  
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21  
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Table 1  
Associations between the mass concentrations of the PM fractions collected in parallel by impactors and respirable cyclones in the Mn alloy smelters

	Parallel respirable fraction		Equation of the line of best fit
	N <sup>a</sup>	r <sup>b</sup>	
< 10 μm	37 <sup>c</sup>	0.88*	y = 0.04+2.02x
< 2.5 μm	37 <sup>c</sup>	0.85*	y = -0.004+1.16x
< 1.0 μm	37 <sup>c</sup>	0.77*	y = -0.010+0.85x
< 0.5 μm	37 <sup>c</sup>	0.73*	y = -0.02+0.72x
< 0.25 μm	37 <sup>c</sup>	0.61*	y = -0.02+0.50x
0.25 – 0.5 μm	38	0.85*	y = 0.003+0.23x
0.5 – 1.0 μm	38	0.71*	y = 0.009+0.13x
1.0 – 2.5 μm	38	0.85*	y = 0.005+0.31x
2.5 – 10 μm	38	0.81*	y = 0.05+0.84x

<sup>a</sup>Number of pairs

<sup>b</sup>Pearson's correlation coefficient

<sup>c</sup> One measurement of the < 0.25 μm fraction was excluded because of its extreme high value (10 times the mass concentration of the respirable fraction)

\* p < 0.001

Table 2  
Mass concentrations (in  $\text{mg m}^{-3}$ ) of the PM fractions measured in the Mn alloy smelters

	N <sup>a</sup>	AM <sup>b</sup>	Min.	Max.	Percentiles		GM <sup>c</sup>	GSD <sup>d</sup>
					10	90		
Parallel respirable fraction <sup>e</sup>	38	0.219	0.046	0.654	0.075	0.469	0.175	1.98
Full-shift respirable fraction <sup>f</sup>	38	0.232	0.042	0.573	0.075	0.467	0.181	2.08
< 10 $\mu\text{m}$ (PM <sub>10</sub> ) <sup>g</sup>	37	0.488	0.116	1.50	0.170	1.10	0.390	1.96
< 0.25 $\mu\text{m}$	37	0.088	0.004	0.609	0.017	0.188	0.051	2.77
0.25 – 0.5 $\mu\text{m}$	38	0.053	0.005	0.162	0.013	0.108	0.039	2.35
0.5 – 1.0 $\mu\text{m}$	38	0.037	0.008	0.115	0.011	0.072	0.028	2.11
1.0 – 2.5 $\mu\text{m}$	38	0.074	0.009	0.239	0.019	0.166	0.055	2.24
2.5 – 10 $\mu\text{m}$	38	0.234	0.018	0.698	0.096	0.497	0.189	2.02

<sup>a</sup> Number of measurements

<sup>b</sup> Arithmetic mean

<sup>c</sup> Geometric mean

<sup>d</sup> Geometric standard deviation

<sup>e</sup> Respirable PM fraction sampled by the respirable cyclone which ran in parallel with impactor

<sup>f</sup> Respirable PM fraction sampled by the respirable cyclone collecting full-shift sample

<sup>g</sup> Sum of all fractions

Table 3  
PM mass fractions (in %) and mass distribution parameters in the Mn alloy smelters calculated from the deposited masses on all stages of the Sioutas cascade impactors

	AM <sup>b</sup>	95% CI <sup>c</sup>	Min.	Max.
Mn alloy smelter 1 (N <sup>a</sup> = 19)				
< 0.25 μm	13.0	11.1 – 14.9	3.6	18.9
0.25 – 0.5 μm	12.8	10.0 – 15.7	5.2	24.1
0.5 – 1.0 μm	8.8	6.7 – 10.9	3.3	19.0
1.0 – 2.5 μm	16.3	12.5 – 20.0	5.3	37.5
2.5 – 10 μm	49.1	43.3 – 55.0	15.7	65.1
MMAD <sup>d</sup>	2.7 μm	2.2 – 3.1 μm	0.9 μm	4.2 μm
GSD <sup>e</sup>	4.5	4.2 – 4.9	3.0	5.4
Mn alloy smelter 2 (N = 18)				
< 0.25 μm	17.3	11.9 – 22.6	2.6	40.8
0.25 – 0.5 μm	10.2	7.3 – 13.0	3.1	22.2
0.5 – 1.0 μm	7.2	5.7 – 8.8	3.4	13.6
1.0 – 2.5 μm	14.9	12.8 – 16.9	8.7	22.5
2.5 – 10 μm	50.5	44.5 – 56.5	33.1	72.1
MMAD	2.6 μm	1.9 – 3.2 μm	0.5 μm	4.8 μm
GSD	4.4	3.9 – 4.9	3.2	5.6
Smelter 1+2 (N = 37)				
< 0.25 μm	15.1	12.3 – 17.8	2.6	40.8
0.25 – 0.5 μm	11.5	9.6 – 13.5	3.1	24.1
0.5 – 1.0 μm	8.0	6.7 – 9.3	3.3	19.0
1.0 – 2.5 μm	15.6	13.5 – 17.7	5.3	37.5
2.5 – 10 μm	49.8	45.8 – 53.8	15.7	72.1
MMAD	2.6 μm	2.2 – 3.0 μm	0.5 μm	4.8 μm
GSD	4.5	4.2 – 4.7	3.0	5.6

<sup>a</sup>Number of measurements

<sup>b</sup>Arithmetic mean

<sup>c</sup>Central 95% confidence interval

<sup>d</sup>Mass median aerodynamic diameter

<sup>e</sup>Geometric standard deviation

Table 4  
Number size distribution parameters in the Mn alloy smelters calculated from the SMPS data

	AM <sup>b</sup>	95% CI <sup>c</sup>	Min.	Max.
Mn smelter 1 furnace 2 (N <sup>a</sup> = 187)				
CMD <sup>d</sup>	27 nm	25 – 28 nm	15 nm	81 nm
GSD <sup>e</sup>	1.94	1.90 – 1.98	1.57	3.31
Mn smelter 2 furnace 1 (N = 101)				
CMD	37 nm	34 – 39 nm	18 nm	104 nm
GSD	2.16	2.12 – 2.20	1.74	2.72
Mn smelter 2 furnace 2 (N = 104)				
CMD	38 nm	34 – 42 nm	18 nm	122 nm
GSD	2.16	2.11 – 2.21	1.67	2.99

<sup>a</sup> Number of size distribution scans

<sup>b</sup> Arithmetic mean

<sup>c</sup> Central 95% confidence interval

<sup>d</sup> Count median mobility diameter

<sup>e</sup> Geometric standard deviation

Table 5  
 Mass concentrations (in  $\text{mg m}^{-3}$ ) of the PM fractions in different job groups in the Mn alloy smelters

	Job group	N <sup>a</sup>	GM <sup>b</sup>	GSD <sup>c</sup>	<i>p</i> -value
<i>Parallel</i> respirable fraction <sup>d</sup>	Tappers	19	0.276	1.66	<0.001
	Crane operators	19	0.110	1.67	
<i>Full-shift</i> respirable fraction <sup>e</sup>	Tappers	19	0.312	1.62	<0.001
	Crane operators	19	0.105	1.64	
< 10 $\mu\text{m}$ (PM <sub>10</sub> ) <sup>f</sup>	Tappers	19	0.601	1.73	<0.001
	Crane operators	18	0.247	1.58	
< 0.25 $\mu\text{m}$	Tappers	19	0.081	2.81	0.003
	Crane operators	18	0.031	2.14	
0.25 – 0.5 $\mu\text{m}$	Tappers	19	0.071	1.68	<0.001
	Crane operators	19	0.021	1.93	
0.5 – 1.0 $\mu\text{m}$	Tappers	19	0.043	1.78	<0.001
	Crane operators	19	0.019	1.92	
1.0 – 2.5 $\mu\text{m}$	Tappers	19	0.090	1.89	<0.001
	Crane operators	19	0.033	1.89	
2.5 – 10 $\mu\text{m}$	Tappers	19	0.277	1.75	<0.001
	Crane operators	19	0.128	1.86	

<sup>a</sup> Number of measurements

<sup>b</sup> Geometric mean

<sup>c</sup> Geometric standard deviation

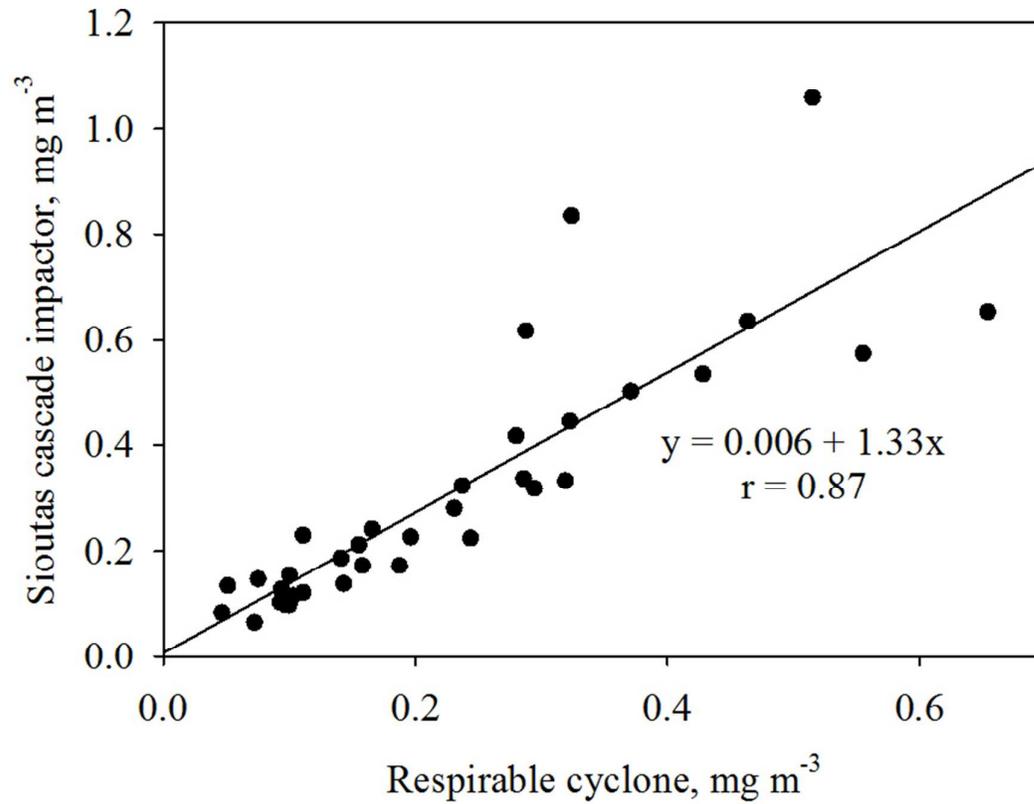
<sup>d</sup> Respirable PM fraction sampled in parallel with impactor

<sup>e</sup> Respirable PM fraction sampled full-shift

<sup>f</sup> Sum of all fractions

## Figures

Fig. 1 The association between the mass concentrations of the respirable fraction calculated from impactor data and collected by respirable cyclones in the Mn alloy smelters



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Fig. 2. Differential mass distribution of particles in the Mn alloy smelters represented by histogram (straight line) and “middle of cut-off diameters” method (dotted line)

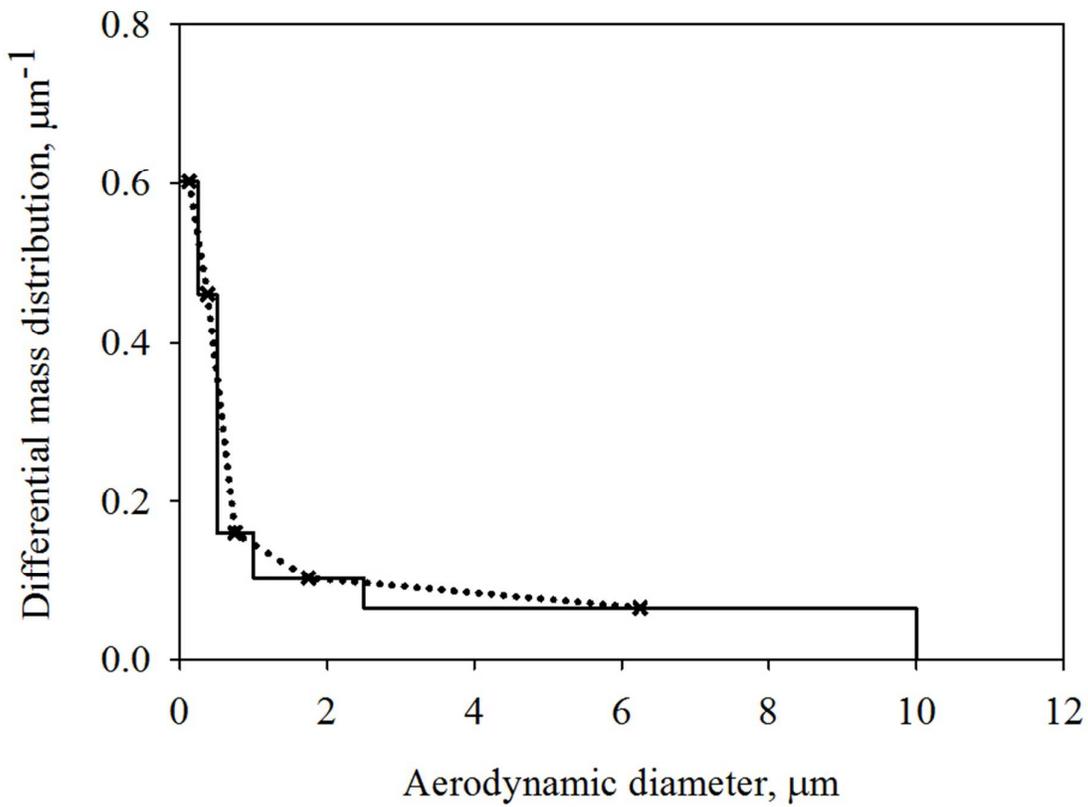


Fig. 3 Particle number concentration distributions measured by SMPS in the Mn alloy smelters

