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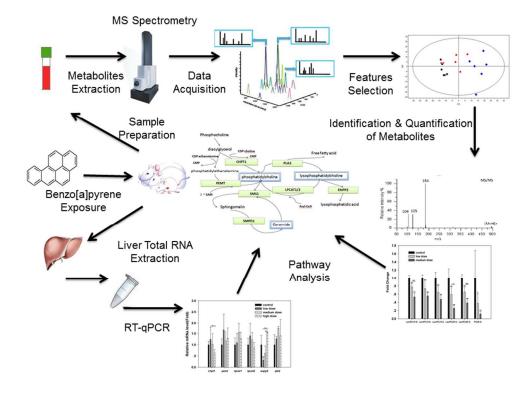
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1	Serum metabolomics analysis reveals oral exposure to
2	benzo(a)pyrene impaired lipid metabolism in rats
3	Xiaoxue Wang, Jie Zhang*, Qingyu Huang, Ambreen Alamdar, Meiping Tian,
4	Liangpo Liu, Heqing Shen*
5	Key Lab of Urban Environment and Health, Institute of Urban Environment,
6	Chinese Academy of Sciences, Xiamen, PR China, 361021
7	*To whom correspondence may be addressed:
8	Prof. Jie Zhang, Institute of Urban Environment, Chinese Academy of Sciences, 1799
9	Jimei Road, Xiamen, 361021, China; Tel/Fax: (86)-592-6190523; E-mail:
10	jzhang@iue.ac.cn
11	Prof. Heqing Shen, Institute of Urban Environment, Chinese Academy of Sciences,
12	1799 Jimei Road, Xiamen, 361021, China; Tel/Fax: (86)-592-6190997; E-mail:
13	hqshen@iue.ac.cn
14	

15 **Running title: Serum** metabolomics analysis of B(a)P-exposed rat

# 16 Short Abstract for Table of Contents

Environmental benzo(a)pyrene (B(a)P) exposure has been associated with a diverse adverse health effects. However, the impact of B(a)P exposure on metabolic network remains obscure. In this metabolomics study, twelve differential metabolic biomarkers were identified in the serum of B(a)P-exposed rats by using ultra-high performance liquid chromatography/mass spectrometry. Our results indicated B(a)P exposure significantly disrupted global lipid metabolism in the rats. These findings may provide useful insights into the mechanisms of B(a)P-mediated toxicity.

24 **ABSTRACT:** Benzo(a)pyrene (B(a)P) is ubiquitously present in the environment. 25 Although its multiple toxicities had been reported, the impact of B(a)P exposure on 26 metabolic network remained obscure. In this study, an ultra-high performance liquid 27 chromatography/mass spectrometry based metabolomics approach was used to 28 investigate the disruption of B(a)P exposure on global serum metabolic profiles in rat. 29 Sprague Dawley rats were orally exposed to 10, 100 and 1000 µg/kg B(a)P for 32 30 days consecutively. Distinct serum metabolomic profiles were found to be associated 31 with the doses. Twelve metabolites were identified as potential biomarkers, which 32 indicated B(a)P exposure disrupted global amino acid metabolism and lipid 33 metabolism, especially phospholipid metabolism and sphingolipid metabolism. Serum 34 lysophosphatidylcholines showed dose-dependent decreases. while serum 35 sphingomyelins presented dose-dependent increases. The expressions of some key genes involved in these pathways were also investigated. Expressions of *enpp2*, sms 36 and *smpd* were significantly altered by high dose of B(a)P exposure. Metabolic 37 38 biomarkers were more sensitive than corresponding gene expression for B(a)P 39 exposure. The findings of this study pointed to novel potential mechanisms in which 40 the identified metabolic pathways involved.

41 KEYWORDS: benzo(a)pyrene; metabolomics; rat; lipid metabolism;
42 lysophosphatidylcholine

# 43 1. INTRODUCTION

Benzo(a)pyrene (B(a)P), a polycyclic aromatic hydrocarbon (PAH), is ubiquitously
present in the environment. It is formed by incomplete combustion of organic
materials in industrial process, automobile emissions. Humans are unavoidably
exposed to B(a)P via the ingestion of contaminated food and water, the inhalation of
particulates in the ambient air and cigarette smoking.<sup>1</sup>

49 B(a)P presents diverse toxicities including hepatotoxicity, neurotoxicity, cytotoxicity, genotoxicity, carcinogenicity and immunotoxicity.<sup>2</sup> A vast number of studies have 50 discussed the mechanisms of B(a)P toxicity. In cells, B(a)P can be metabolized into 51 reactive metabolites via cytochrome P450 (CYP) mediated pathway,<sup>3</sup> and then bind 52 covalently to nucleic acids and proteins, leading to mutation and cell proliferation.<sup>4</sup> A 53 54 number of studies have proved that B(a)P or its metabolites would lead to ROS generation and lipid peroxidation.<sup>5, 6</sup> Recently, increasing evidences have shown that 55 B(a)P might disrupt estrogen receptor activity, testicular steroidogenesis and 56 epididymal function in vivo and in vitro.<sup>7, 8</sup> What's more, endocrine disruption is 57 tightly involved with metabolic disruption through disturbing hormones receptors, 58 intervening peroxisome proliferator-activated receptors (PPAR) and inappropriate 59 activating xenosensors,<sup>9</sup> all of which may be involved in lipid and glucose 60 metabolism. Therefore, we hypothesized that metabolic disruption might be an 61 important mechanism underlying B(a)P toxicity. Moreover, to 62 achieve a comprehensive understanding on B(a)P toxicological effects and mechanisms, a 63 global analysis on the biological responses should be performed at molecular levels 64

65 (gene, protein and metabolite).

66 Omics technologies are capable to acquire the information of global profile, and have 67 been widely used in environmental toxicology. Genomic and proteomic studies have 68 identified several potential B(a)P and other PAHs biomarkers in cells, mice, rats and 69 human.<sup>10-14</sup> However, its influence on global metabolic profiling remained obscure. 70 Metabolic perturbation is often an early event of pollutant induced histopathological 71 change.<sup>15</sup> Metabolomics provides an overview of the metabolic status of a biological 72 system exposed to environmental stress.<sup>16</sup>

A number of metabolomic researches have addressed the adverse effects of B(a)P and other PAHs (e.g. pyrene, phenanthrene, 1,2:5,6-dibenzanthracene ) on metabolic response using several model animals *Manila* clam,<sup>17</sup> fish,<sup>18</sup> earthworm.<sup>19-21</sup> However, these non-mammalian models failed to adequately mimic the human response to B(a)P. Rat is obviously a better model in the toxicity study, but so far there is little discussion about systematic metabolic response to B(a)P in rats.

79 The present study is designed to investigate the adverse effect of B(a)P exposure on global metabolome in rat. An ultra-high performance liquid chromatography/mass 80 81 spectrometry (UHPLC/MS) based metabolomics approach was used to profile and 82 characterize significantly altered metabolites in the sera of the rats orally exposed to B(a)P. Besides, the expressions of key genes involved in altered metabolism pathway 83 84 were examined. This study identified the disrupted metabolic pathways associated with B(a)P exposure, and thus led to a more comprehensive understanding of B(a)P 85 toxicity. 86

# 87 2. METHODS

### 88 2.1 Chemicals

Benzo(a)pyrene (purity>98.5%) and formic acid (HPLC grade) were purchased from
Acros (Morris Plains, NJ, USA). All standards (purity>95%) were purchased from
Sigma-Aldrich (St. Louis, Mo, USA). Methanol (HPLC grade) was obtained from
Fisher Scientific (Fair Lawn, NJ, USA). Distilled water (18.2 MΩ) was obtained from
a Milli-Q system (Beford, MA, USA).

# 94 **2.2 Animal experiments**

A total of 22 Sprague Dawley rats aged six weeks (weight 200  $\pm$  10 g) were 95 96 obtained from Shanghai Laboratory Animal Center, China. Animals were housed separately by sex in stainless-steel cages and acclimatized for one week before 97 98 starting B(a)P exposure. Rats were maintained in an air-conditioned room at the temperature of 26  $\pm 2^{\circ}$ C, a relative humidity of 50  $\pm 5\%$ , and a 12 h light/12 h dark 99 100 cycle. Each animal had ad libitum access to water and pellet diet. All the rats were randomly divided into control and three treatment groups by intragastric B(a)P 101 administration every day. The control group was fed with corn oil. The low, medium 102 and high-dose groups were administered for 32 days with 10, 100 and 1000 µg/kg 103 B(a)P which was dissolved in corn oil, respectively. Although the doses we used were 104 higher than environmental exposure level, they were similar to and even much lower 105 than those used in other peer-reviewed reports on B(a)P.<sup>10, 22, 23</sup> The relatively high 106 107 dose was used to obtain the adverse effect in rats when human model is unavailable, and it helped elicit a detectable response in experimental animals. All animals were 108

109 treated humanely and with regard for alleviation of suffering according to the China Animal Welfare legislation. 110

111 **2.3 Sample collection and preparation** 

After being consecutively treated with B(a)P for 32 days, rats were killed by 112 113 decapitation. Blood was collected from each animal and placed into ice-cold tubes. 114 Serum was obtained by centrifugation (3500  $\times$ g, 10 min at 4 °C) and frozen at -80 °C 115 before further sample preparation and analysis. The livers were removed immediately 116 after sacrifice, rinsed with PBS (room temperature), flash frozen in liquid nitrogen 117 and then stored at -80 °C.

118 For the pretreatment of the serum samples, a volume of 600 µL cold methanol was added to 200 µL serum and was shaken vigorously, and the mixture was stored for 10 119 120 min and subsequently centrifuged at 12,000 ×g for 10 min at 4 °C. The supernatant 121 was filtered through a 0.22 µm syringe filter prior to UHPLC/MS analysis. From each sample, 20 µL sera were mixed and divided into several aliquots as the quality 122 123 controls (QCs). The QCs were periodically injected during sample acquisition batch, and used to evaluate the stability and reproducibility of analytical instrument. 124

125

# 2.4 Metabolic profile Acquisition

126 Serum metabolic profiles were acquired using a UHPLC/orbitrap-MS system (Thermo, USA). A Kinetex C18 column (150 mm×2.1 mm, 2.6 µm) was used for 127 chromatographic separation. The mobile phase consisted of water containing 0.1% 128 129 formic acid (mobile phase A) and methanol containing 0.1% formic acid (mobile phase B). A programmed gradient was used: 5% B increased to 100% B in 16 min and 130

131 held for 4 min, then decreased to 5% in 0.1 min, and finally maintained at 5% B for 3 min. Sample injection volume was 5 µL. Compared with negative ion mode, more 132 133 serum metabolites and higher magnitude of the metabolites were detected under positive ion mode (Fig. S1), hence the mass spectrometer was operated in full-scan 134 135 positive ion mode with a range of 100-1000 m/z. Spray voltage and cone voltage were 136 3.5 KV and 35 V, respectively. Heated capillary temperature and source temperature 137 were 380 °C and 350 °C, respectively. Curtain and auxiliary gas flow were 60 and 35 138 L/h, respectively. Serum samples were run in a randomized fashion to avoid possible 139 uncertainties from artifact-related injection order and gradual changes of instrument 140 sensitivity in whole batch runs. Because the serum samples of high dose treatment 141 were missing during storage, the samples were not included in metabolome analysis. 142 A QC sample and a blank were analyzed once or twice interval of seven samples to identify the sample carryover and check for stability (n=5). QC samples showed a 143 tight cluster in scoring plot (Fig. S2), suggesting the acquired data had high quality 144 145 and were worthy of further multivariate statistical analysis. To carry out MS/MS mode 146 to identify potential biomarkers, argon was used as collision gas, and collision energy 147 was adjusted from 15 to 40 eV for each analyte.

148 **2.5 Metabolome analysis** 

Metabolic raw data were converted into CDF format and then processed by MarkerLynx v4.1 software, which automatically generated a two-dimensional data table of detected variables (m/z, retention time pairs) and their respective intensities (peak areas). Markerlynx parameters were set as follows: mass range 100-1000 m/z, 153 mass tolerance 0.01 Da, retention time window 0.10 min, and noise elimination level 6. The intensity of extracted variables (spectral bins) was normalized to overall 154 intensities for each sample to remove the unwanted systematic bias and correct for the 155 different enrichment factors of serum among individuals.<sup>24</sup> Hence, the intensity of 156 157 peak area of each biomarker normalized by sum was expressed as the relative 158 intensity of this biomarker in serum. Any variables with missing values in more than 80% of the samples were excluded according to the "80% rule".<sup>25</sup> The missing values 159 were substituted with 1/2 minimum values prior to multivariate statistical analysis. 160 161 Finally, the processed data were pareto-scaled and subjected to multivariate statistical analysis using SIMCA-P+ 12.0 software (Umetrics AB, Uppsala, Sweden). Principal 162 component analysis (PCA) was first performed to discover intrinsic treatment-related 163 164 clusters within the datasets. Partial least squares-discriminate analysis (PLS-DA) was 165 further used to improve separation among the groups and screen potential biomarkers. The robustness and validity of the developed PLS-DA model was tested using a 166 200-permutation test (Fig. S3). Variable importance in projection (VIP) represents the 167 extracted variables' ability to discriminate different doses. The variables with VIP 168 values > 1.5 were included in the preset of biomarkers. Kruskal–Wallis test was then 169 used to determine which variables were significantly different between all the three 170 groups. The variables with a p < 0.05 were further subjected to Mann-Whitney test to 171 investigate their changes in either dose group relative to the control. The metabolites 172 173 which presented significant alteration in either dose group were finally selected as potential biomarkers. 174

The identification of potential biomarkers was conducted according to our previous 175 reports.<sup>26, 27</sup> Briefly, metabolites were blasted with m/z values against the human 176 177 metabolome database (HMDB: http://www.hmdb.ca/). Top endogenous metabolites provided by the database searching could be potential biomarkers. MS/MS 178 information was further used to confirm the identification (Table S1). If available, the 179 180 structures of these metabolites were finally confirmed by comparison with commercial standards. Otherwise, the metabolites were tentatively identified by 181 comparing the acquired structure information with metabolite databases and published 182 183 literature.

184 **2.6 Quantitative real-time PCR** 

Quantitative real-time PCR was applied to evaluate the expressions of some key genes 185 186 involved in the lipid metabolism pathway. Total RNA samples were extracted from homogenized liver samples using Total RNA Kit I (OMEGA). Reverse-transcription 187 of cDNA synthesis was performed with 1 µg total RNA using PrimeScript<sup>R</sup>RT reagent 188 Kit (TaKaRa Bio, Otsu, Japan). Real-time PCR was carried out in a 20 µL final 189 volume and performed in duplicate using SYBR Green Master Mix reagents in a 190 Light cycler 480 detection system (Roche Applied Science, Indianapolis, USA) 191 according to the manufacture's protocol. PCR primers were listed in Table 1. The 192 conditions for quantitative PCR were as follows: 95 °C for 10 min followed by 40 193 cycles at 95 °C for 15 s, and 60 °C for 60 s. Gene expression levels were normalized to 194  $\beta$ -actin expression levels. The fold changes (treated/control) of the tested genes were 195 analyzed by the  $2^{-\Delta \Delta Ct}$  method. 196

# 197 **2.7 Statistical analysis**

All analyses were conducted using SPSS Version 18.0 (SPSS Inc., Chicago, USA). 198 199 The relative intensities of the biomarkers in dose groups were further normalized to the respective controls to provide us with the information of fold change. If data were 200 201 not normal distributed, nonparametric tests would be carried out to compare between 202 and within groups. In metabolome analysis, Kruskal-Wallis test was used to 203 determine which variables were significantly different among all the three groups, 204 Mann-Whitney test to investigate their changes in either dose group relative to the 205 control group. Normal distributed data were analyzed using one-way ANOVA; LSD 206 or Tamhane test was used based on homogeneity of variances. Significance was set at 207 *p*≤0.05.

208 **3. RESULTS** 

#### **3.1 QC validation**

The use of biological QC samples provides an important means of monitoring method performance. In our study, QC samples were analyzed to evaluate the sample carryover, the stability and reproducibility of the present method. After data acquisition and preprocess, there were 631 variables remaining in the dataset. CV values of 63.9% variables were<30%, and the CV values of 48.5% variables were<15% (Fig. S4), indicating our method had excellent repeatability and the dataset is worthy for further analysis.

#### 217 **3.2 Metabolome analysis**

218 The PCA scoring plot allowed a primary separation among groups (Fig. 1A). A

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supervised PLS-DA model was further used to discover the difference among groups. The corresponding PLS-DA model with five latent components had a faithful representation of the data (R2X=0.72, R2Y=0.99) and a very good cumulative predictive capacity ( $Q2_{(cum)}=0.76$ ) (Fig. 1B). Furthermore, the medium dose group, low dose group and control group were obviously separated, and the B(a)P group had a sparser cluster than that of control group, perhaps due to different sensitivity to B(a)P of rats in dose groups.

Biomarkers are defined as "putative metabolites responsible for class separation
identified using the loadings and variable importance plots" in metabolomics field.<sup>28</sup>
In the present study, twelve potential biomarkers were tentatively identified (Table 2),
which reflected the impact of B(a)P exposure on metabolic pathways.

230 Most of these biomarkers were involved in lipid metabolism, including five 231 lysophosphatidylcholines (lysoPCs) phosphatidylcholine and one 232 (glycerophospholipid metabolism), palmitic amide (fatty acid metabolism), sphinganine, phytosphingosine and cer(d18:0/14:0) (sphingolipid metabolism). In this 233 study, all the five lysoPCs markedly decreased in a dose-dependent manner in all 234 235 treated groups (Fig. 2). A significant decrease of PC(36:2) was found among all 236 treated groups. We also observed an obvious decrease of cer(d18:0/14:0) in low dose group while an increased concentration of sphinganine and phytosphingosine in 237 238 medium dose group (Fig. 3). Other biomarkers were cytosine and L-valine, which 239 involved in amino acid and nucleotide metabolism, respectively. And both of them were significantly increased in medium dose group. 240

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#### 241 **3.3 Organ coefficients of liver and expression of key hepatic genes**

242 Liver is relevant in toxicology as the primary organ of metabolism and detoxification for B(a)P.<sup>29</sup> No significant alteration of organ coefficients of liver was observed 243 following B(a)P exposure (Figure S5). The alteration in the expression of key hepatic 244 genes was further investigated. Chpt1, pemt, pla2, lpcat1, lpcat2 and enpp2 are 245 246 prerequisites for glycerophospholipid metabolism (Fig. 4). The expression of *chpt1*, pemt, pla2, lpcat1, lpcat2 were not significantly changed upon B(a)P exposure (Fig. 247 S6). The expression of *enpp2* was markedly down-regulated in low dose group and 248 249 up-regulated in high dose groups (Fig. 5), which featured significant low-dose effect 250 and nonmonotonic dose response to B(a)P.

Sphingomelin (SM) is mainly regulated by sphingomyelin synathase and sphingomyelin phosphodiesterase (Fig. 4). Thus, *sms1* and *smpd2* mRNA levels were also measured in livers. Significant hepatic perturbation of SM metabolism was observed in high dose B(a)P exposure. *Sms1* expression was severely suppressed by 1.86-fold in high dose group, while it changed little in either low or medium dose group (Fig. 5). Expression of *smpd2* showed the similar change as *sms1*, it significantly decreased by 1.96-fold in high-dose group.

258 4. DISCUSSION

In this study, we observed B(a)P exposure induced the perturbation of global metabolic system, especially lipid metabolism. Several major and essential structural components of membrane lipid bilayers, including lysoPCs and SMs, were identified as potential biomarkers.

13

PC is a major component of biological membranes. LysoPCs are the intermediate 263 metabolites of PC and they are biologically active lipids regulating a variety of 264 cellular functions. The abnormal levels of lysoPCs may suggest a disturbance of lipid 265 and glucose homeostasis, thus they have been used as a potential diagnostic 266 biomarker in various diseases.<sup>30, 31</sup> In our study, B(a)P exposure caused a 267 268 down-regulated level of PC and lysoPCs, indicating its ability to disrupt glycerophospholipid metabolism in rats. Liver is the major detoxification organ and 269 target organ of B(a)P exposure, and recently increased risk of hepatocellular cancer 270 was associated with B(a)P exposure.<sup>32</sup> Therefore, we further investigated mRNA 271 272 expression of several hepatic genes involved in PC homeostasis. PC can be derived 273 from phosphatidylethanolamine through the action of *pemt* or the CDP-choline pathway catalyzed by *chpt*.<sup>33</sup> However, the hepatic *pemt* and *chpt1* expression seemed 274 275 unchanged by B(a)P exposure. PC is further hydrolyzed by pla2 to generate lysoPCs 276 and free fatty acid via Land's cycle. LysoPCs are converted back to PC catalyzed by *lpcat1*/2, or it is hydrolyzed by *enpp2* to LPA. In our study, *lpcat1*/2 expressions were 277 not significantly changed after B(a)P treatment. However marked non-monotonic 278 dose-response (NMDR) was observed for *enpp2* at transcriptional level. As a member 279 280 of endocrine-disrupting chemicals (EDCs), B(a)P has significant low-dose effect and non-monotonic dose responses.<sup>7, 34</sup> EDCs can have effects at low doses that are not 281 predicted by effects at higher doses and vice versa.<sup>35</sup> NMDR has also been reported 282 for the expression of cytochrome P450 (P450s or CYPs),<sup>36, 37</sup> protein expressions 283 and even hormones level in animals exposed to B(a)P.<sup>38, 39</sup> However, the exact 284

molecular mechanism underlying NMDR of B(a)P exposure remained obscure. 285 Therefore, more efforts are needed to explore the intrinsic causation of low-dose 286 effect of B(a)P. Moreover, *enpp2* is known to stimulate migration of tumor cells, and 287 its overexpression has been associated with a variety of cancers.<sup>40</sup> Previous studies 288 289 revealed PCBs and other toxic chemicals (e.g. 2-mercaptobenzothiazole, benzyl acetate) might up-regulate *enpp2* expression in human PBMC <sup>35</sup> and Panc-1 cells.<sup>41</sup> In 290 this study, our observation that high dose exposure to B(a)P significantly increased 291 enpp2 expression may suggest the association between B(a)P intake and cancer risk, 292 293 but the exact molecular mechanism needs further investigation.

294 Previous evidence has shown that endothelial cells can degrade extracellular lysoPCs to reduce plasma and tissue levels of these pro-inflammatory lipid molecules.<sup>42</sup> 295 296 Interestingly, we found serum lysoPCs were more sensitive than corresponding gene expression for B(a)P exposure. A possible explanation of this discrepancy may be the 297 fact that abnormal levels of serum lysoPCs is not exclusively produced in liver, but 298 could also be an injury signal of other organs.<sup>43</sup> Combing these observations, it's 299 reasonable to assume that the disruption of glycerophospholipid metabolism is critical 300 for the final adverse effects of B(a)P exposure, but further experiments are needed to 301 302 confirm these preliminary findings and investigate in-depth molecular mechanism.

303 Ceramide and sphinganine are known as the major structural components of the 304 plasma membrane. Decreased serum cer(d18:0/14:0) and increased serum 305 sphinganine suggested B(a)P exposure could induce a disruption of membrane 306 distribution. Tekpli *et al.* and Gorria *et al.* also reported an increase in membrane

fluidity in cell models exposed to B(a)P,<sup>44, 45</sup> supporting our results that B(a)P could affect the membrane composition via disturb the sphingolipid metabolism. Moreover, sphingomyelinases hydrolyze sphingomyelin to bioactive lipids (sphinganine and phytosphingosine), and were down-regulated in presence of PAH.<sup>46</sup> This was also in accordance with our observation.

Amino acid and nucleotide metabolism were also disturbed by B(a)P exposure, 312 which were indicated by changed levels of valine and cytosine. L-valine is one of 313 the branched chain essential amino acids (BCAAs) that particularly involved in 314 energy metabolism. Cytosine is the most important base involved in DNA 315 316 methylation. B(a)P has been shown to disrupt DNA methylation patterns in zebrafish embryos and breast cells.<sup>47, 48</sup> The accumulation of cytosine in medium 317 318 dose group may be related to restoration of modified cytosine so that it can replenish cytosine pools after the modified DNA is cut off and resynthesized.<sup>49</sup> 319

320 Our study is not without limitation. In this study, our hepatic total RNA extractant derived from various kinds of cells in both centrilobular and peripherpheral zones. 321 It was possible the alterations in gene expression might be diluted and masked by 322 that from non-responding cells. Therefore, more extrapolation studies are required 323 324 for the assessment of genes expression in various kinds of cells and specific hepatic zones from liver of rats exposed to B(a)P in the future. Besides, the 325 biomarker identified in metabolomics is not exactly the same with the biomarkers 326 used as "prognostic or diagnostic indicators of disease or a sensitive and specific 327 tool for risk assessment".<sup>50</sup> Though biomarkers here provided potential disturbed 328

pathways, they did not specifically respond to B(a)P exposure, therefore they
could not be used as exposure biomarkers in the risk assessment of B(a)P exposure.
Further efforts should aim to discover more specific biomarkers (e.g.
B(a)P-receptor binding related gene or protein) for B(a)P exposure.

### 333 CONCLUSIONS

334 In this study, an UHPLC/MS-based metabolomic approach was conducted to 335 investigate serum metabolic alterations in rats exposed to B(a)P. Obvious metabolic differentiation between the dose and control groups was observed. Twelve metabolites 336 337 were identified as potential biomarkers, including five lysophosphatidylcholines, sphingomyelins, palmitic amide, L-valine and cytosine. The metabolomic results 338 indicated the B(a)P exposure mainly interfered lipid metabolism. Besides, metabolic 339 340 biomarkers were more sensitive than corresponding gene expression for B(a)P exposure. Overall, this work improves our understanding of B(a)P toxicity in 341 mammals, and this preliminary findings demonstrated the great potential of 342 metabolomics in the toxicity research of B(a)P. Since B(a)P is a widely accepted EDC, 343 it's necessary to apply the metabolomics to implement metabolomics profiling of 344 345 hormone-related organs such as breast and prostate. Through combining the 346 metabolomics data and RT-PCR data of these organs, we may obtain a global view of toxic mechanism of B(a)P. 347

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# 352 CONFLICT OF INTEREST

353 The Authors did not report any conflict of interest.

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453

454 Figure 1. Scoring plots of PCA (A) and PLS-DA (B) analysis. ■ control; • low dose
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456

**Figure 2.** B(a)P-induced disruption of glycerophospholipid homeostasis in rats. All the data were expressed as mean  $\pm$  standard deviation (SD). \*p < 0.05, \*\*p < 0.01.

459

**Figure 3.** B(a)P-induced disruption of sphingolipid homeostasis in rats. All the data

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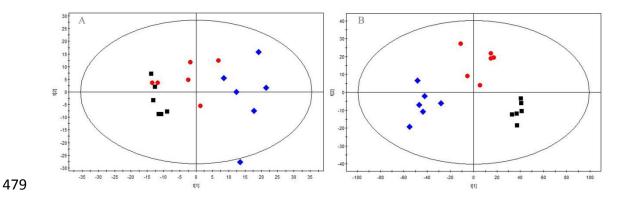
462

463 Figure 4. A schematic representation of the reactions of potential biomarkers involved in glycerophospholipid metabolism and sphingolipid metabolism. PC can be derived 464 phosphatidylethanolamine 465 from catalyzed by phosphatidylethanolamine N-methyltransferase (pemt) or the CDP-choline pathway catalyzed by a final-step 466 enzyme choline phosphotransferase (chpt). Meanwhile, PC is hydrolyzed by 467 phospholipase A 2 (pla2) to generate free fatty acid and LysoPC. The latter can be 468 converted back to PC in the presence of acy-CoA by lysophosphatidylcholine 469 acyltransferase 1/2 (lpcat1/2). Besides that, LysoPC is hydrolyzed by enpp2 to 470 lysophosphatidic acids. Sphingomelin (SM) is mainly regulated by sphingomyelin 471 synathase (sms) and sphingomyelin phosphodiesterase (smpd, also known as 472 sphingomyelinase). Through the action of either acid or neutral sphingomyelinase, 473

- 474 sphingomylein is hydrolyzed to ceramide and phosphocholine. Also, ceramide can be
- 475 synthesized from sphinganine via *de novo* pathway.

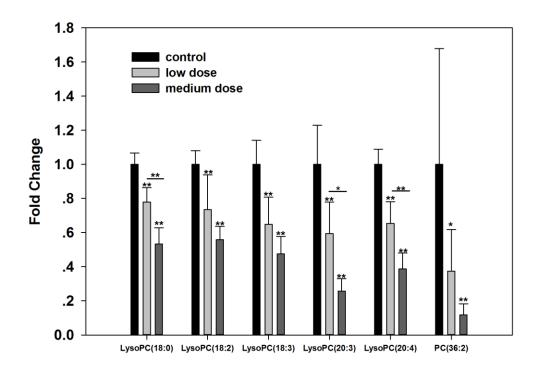
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- 477 Figure 5. Effects of B(a)P exposure on mRNA expression in rat liver. All the data
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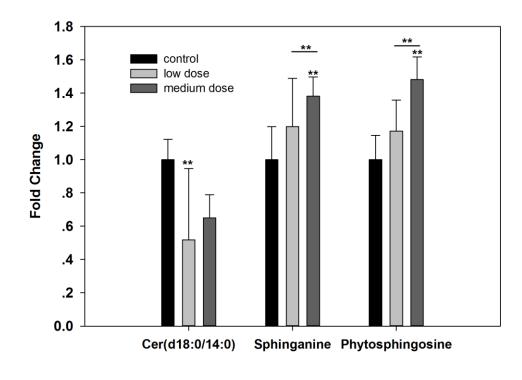


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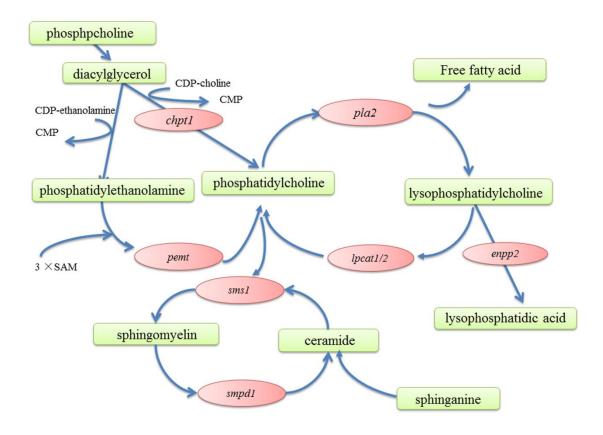
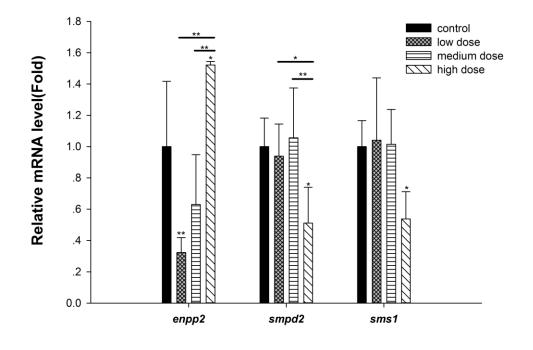


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505

**Table 1.** Primers sequences used for real-time PCR analysis.

Gene	Full name	Forward primer(5'-3')	Reverse primer(5'-3')
chpt1	choline phosphotransferase	TGGCTATCGGTGCTTCAA	TTCCAAATCGCAACACTCCT
pemt	phosphatidylethanolamine N-methyltransferase	CCATTTCCTTCTGGTTCTGG	CTCTGCTCCCACCTTGCTAC
lpcat1	lysophosphatidylcholine acyltransferase 1	TGGCGGTGAGATAGACCTTC	TCAATAGCCTGGAACAAGTCG
lpcat2	lysophosphatidylcholine acyltransferase 2	GGCCCTTTGCTGTAATCTCA	TAGCCACCTTTCCTTTCACG
enpp2	ectonucleotide pyrophosphatase/phosphodiesterase 2	CCTTCAGTCCGAGTTTGACC	GCCGTCCATACAGGAGATGT
pla2	phospholipase A2	GTCACCAACTTGTTCTCAAACCCAT	CAACTCCACCAGAATCTCACT
sms1	sphingomyelin synathase	CAGGAAGCCAAGATGAGGAG	ACAAGATGGTCAGGGCAGTT
smpd2	sphingomyelin phosphodiesterase	GTTCATCCACCACACATCCA	CCACTCTTTCAGTAGGCAGCA

# 507 Table 2. Potential biomarkers of toxicity induced by benzo(a)pyrene.

No.	HMDB ID	Metabolite	Chemical formula	VIP	Multiple comparisons	<i>p</i> value control VS low dose group	control VS medium dose group	Sub-Pathway	KEGG Pathway
1	HMDB10384	LysoPC(18:0)	C26H54NO7P	7.28	0.001	0.004	0.006	Glycerophospholipid metabolism Fatty acid metabolism Sphingolipid metabolism	Lipid metabolism
2	HMDB10386	LysoPC(18:2)	C26H50NO7P	5.28	0.016	0.055	0.006		
3	HMDB10395	LysoPC(20:4)	C28H50NO7P	4.84	0.001	0.004	0.006		
4	HMDB10393	LysoPC(20:3)	C28H52NO7P	2.04	0.001	0.006	0.006		
5	HMDB10387	LysoPC(18:3)	C26H48NO7P	1.79	0.004	0.006	0.006		
6	HMDB08299	PC(36:2)	C44H84NO8P	3.63	0.004	0.037	0.006		
7	HMDB12273	Palmitic amide	C16H33NO	2.16	0.024	0.262	0.011		
8	HMDB00269	Sphinganine	C18H39NO2	3.54	0.01	0.2	0.006		
9	HMDB04610	Phytosphingosine	C18H39NO3	2.98	0.004	0.15	0.006		
10	HMDB11759	Cer(d18:0/14:0)	C32H65NO3	2.58	0.014	0.025	0.006		
11	HMDB00883	Valine	C5H11NO2	3.19	0.002	0.01	0.006	Valine, leucine and isoleucine degradation/ biosynthesis	Amino acid metabolism
12	HMDB00630	Cytosine	C4H5N3O	2.6	0.011	0.423	0.011	Pyrimidine metabolism	Nucleotide metabolism