

Analytical Methods

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**Simultaneous determination of hydroquinone and catechol in
compost bioremediation using a tyrosinase biosensor and artificial
neural networks**

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1
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4 **Abstract**
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6 A biosensor based on tyrosinase immobilized with ordered mesoporous carbon-Au (OMC-Au),
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8 L-Lysine membrane and Au nanoparticles (tyrosinase/OMC-Au/L-Lysine/Au) was combined with
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10 artificial neural networks (ANNs) for the simultaneous determination of catechol (CC) and
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12 hydroquinone (HQ) in compost bioremediation of municipal solid waste. The good performance of
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14 biosensor provided the potential applicability for the simultaneous identification and quantification of
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16 catechol and hydroquinone in real samples, and the combination with ANNs offered a good
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18 chemometric tool for data analysis in respect to the dynamic, nonlinear, and uncertain characteristics of
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20 the complex composting system. Good prediction ability was attained after the ANNs model
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22 optimization, and the direct detection range for catechol and hydroquinone were directly analyzed by
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24 the ANNs model varied between 1.0×10^{-7} and 1.1×10^{-4} M, significantly extended than the linear model
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26 (4.0×10^{-7} to 8.0×10^{-5} M). Finally, the performance of the ANNs model was compared with the linear
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28 regression model. The results demonstrated that the prediction results by the ANNs model were more
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30 precise than those by the linear regression, and the latter was far from accurate at high levels of
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32 catechol and hydroquinone beyond the linear range. All the results showed that the combination of the
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34 biosensor and ANNs was a rapid and sensitive method in the quantitative study of composting system.
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44 **Keywords:** biosensor; DPV signals; catechol; hydroquinone; artificial neural networks; compost
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46 bioremediation.
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20 Introduction

21 Phenolic compounds are widely distributed as environmental pollutants because many of them are
22 resistant to biotic and abiotic degradation. They are mostly derived from various agricultural and
23 industrial activities, including waste discharge from wood preservatives, coking, textiles, plastics, dyes,
24 paper, herbicides industries and the partial degradation of phenoxy contaminants in remediation
25 processes^{1,2}. The toxicity of phenols generated from bioremediation processes, such as composting, can
26 bring on undesirable ecological effects and seriously reduce removal efficiencies³. Catechol (CC) and
27 hydroquinone (HQ) are two isomers of phenolic compounds which are harmful to human health and
28 ecological environment. During the application of composting technology in disposal of municipal
29 solid waste, CC and HQ are generally direct pollutant or by-product of the aromatic pollutant⁴.
30 Therefore, detection and quantification of the toxicity of these phenolic compounds from compost
31 bioremediation of municipal solid waste is a critical issue. Up to now, a great number of analytical
32 methods have been established to determine dihydroxybenzene isomers in compost systems. On the
33 one hand, there are techniques such as high-performance liquid chromatography (HPLC)⁵,
34 spectrophotometry⁶ and gas chromatography⁷, which allow individual identification of phenols, but
35 these procedures usually require specific equipment, laboratory conditions, and are not suitable for
36 on-site analyses. On the other hand, electrochemical methods are applied to detect the hydroquinone,
37 catechol, phenol, resorcinol, cresol. These methods have the advantages of fast response, cheap
38 instrument, low cost, simple operation, timesaving, but the key point lies in improving the sensitivity,
39 selectivity and the potential applicability for the quantification of phenols in real samples. In an attempt
40 to overcome the deficiencies of these analytical methods, the applications of enzyme sensors to specific
41 pollutant detection have been increasingly reported to exhibit superior sensitivity, stability, reusability,

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4 42 selectivity, and portability⁸. Especially the biosensor provided the potential to quantify the pollutant
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6 43 levels in real environmental. The operation efficiency of compost systems will be much improved if
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8 44 enzyme sensors are applied in pollutant detection.
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11 In our previous works, a tyrosinase biosensor was developed for linear calibration and
12
13 46 simultaneous determination of hydroquinone and catechol⁹. The biosensor was evaluated by differential
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15 47 pulse voltammetry (DPV) measurements, which is used to make electrochemical measurements, and
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17 48 the DPV peak currents increased linearly with concentration over the range of 4.0×10^{-7} to 8.0×10^{-5} M,
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19 49 the detection limits of HQ and CC were 5×10^{-8} and 2.5×10^{-8} M (S/N=3), respectively. The sensitivities
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21 50 in the linear calibration regions for low concentration show the following order: 0.4511 A/M (catechol,
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23 51 $n=4$) > 0.338 A/M (hydroquinone, $n=13$). And the electrode showed a rapid and sensitive
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25 52 bioelectrocatalytic response of 65 and 89 s after addition of catechol and hydroquinone, respectively.
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27 53 Using the differential pulse voltammetry (DPV), the wide peak separation and low peak potential
28
29 54 ensured the avoidance of interferences, making this biosensor a potential device for real sample
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31 55 applications. However, the detection procedures are still susceptible to the complex samples containing
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33 56 heterogeneous organic components and certain functional groups, such as phenolic OH and carboxyl,
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35 57 especially in compost system which a variety of organic compounds coexisting, owing to both the
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37 58 redox and sorption of the interfering matrix constituents on the electrode surface⁸. As a result, an
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39 59 unstable baseline or even the overlapped differential pulse voltammetry signal will be obtained with a
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41 60 carbon electrode when it was applied to large quantities of compost samples. Although the data
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43 61 generated by simultaneous determination of phenols compounds from compost bioremediation can be
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45 62 analyzed using the linear regression model, nonlinearities and uncertainties also occur in the process as
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47 63 mentioned above, which restrict the biosensor in practical application. Thus, the quantification
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4 64 capability of the linear model will be limited by the dynamic, nonlinear, and uncertain characteristics of
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6 65 the complex composting system, and will give erroneous results if the linear range is exceeded.
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9 66 Artificial neural network (ANN) are computational models inspired by animal central nervous systems
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11 67 (in particular the brain) that are capable of machine learning and pattern recognition. They are usually
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14 68 presented as systems of interconnected “neurons” that can compute values from inputs by feeding
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16 69 information through the network. It have found extensive utilization in solving many complex
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19 70 real-world problems. ANNs could be deemed as advanced signal processing variants allowing the
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21 71 interpretation, modelling and calibration of complex analytical signals for they can process very
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24 72 nonlinear and complex problems even if the data are imprecise and noisy^{8,10,11}. The combination of the
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26 73 tyrosinase biosensor with ANNs modelling may represent an alternative to classical methods. This
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29 74 approach has already been introduced towards the analysis of phenols. For example, the group of
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31 75 Xavier Cetó and Francisco Céspedes has used this method to manage the sensor signal, and established
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34 76 electronic tongue and Bio-Electronic Tongue (BioET) based on voltammograms correlated to phenol
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36 77 contents in wines¹²⁻¹⁶. In addition, Tang group has used this method to handle the biosensor signal,
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39 78 processing the amperometric signals correlated to enzyme activities or phenol contents in compost
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41 79 system^{8,26}.

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44 80 In this work, the application of ANN technique for evaluation of the DPV signals of
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46 81 multi-component analysis generated by the tyrosinase biosensor for the simultaneous determination of
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49 82 CC and HQ in compost extract samples was explored, which has not been reported. This method
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51 83 combining the advantages of both parts, calibrated the complex overlapping analytical signals and
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54 84 imprecise data from composting samples. The aim of the study was to extend the limited measuring
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56 85 range of the biosensor to a useful and wider working band. This assay provided the potential
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4 86 applicability of the biosensor for the quantification of CC and HQ in compost system, and the
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6 87 development of fast and inexpensive on-line monitoring systems in municipal solid waste compost
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8 88 bioremediation.
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10 11 89 **Experimental**

12 13 14 90 **Apparatus and reagents**

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16 91 Cyclic voltammetric (CV) measurement and differential pulse voltammetry (DPV) measurement
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18 92 were carried out on CHI660B electrochemistry system (Chenhua Instrument, Shanghai, China). Model
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20 93 PHSJ-3F laboratory pH meter (Leici Instrument, Shanghai, China) was used to test pH value. A Sigma
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22 94 4K15 laboratory centrifuge, a vacuum freezing dryer and a mechanical vibrator were used in the assay.
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25 95 The three-electrode system used in this work consisted of a tyrosinase/OMC-Au/L-Lysine/Au/glassy
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27 96 carbon electrode (GCE) as working electrode, a saturated calomel electrode (SCE) as reference
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29 97 electrode and a Pt foil auxiliary electrode. All the work was conducted at room temperature (25 °C)
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31 98 unless otherwise mentioned.
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36 99 Tyrosinase (EC 1.14.18.1, from mushroom as lyophilized powder), catechol and hydroquinone
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38 100 were purchased from Sigma-Aldrich (USA). Tetraethoxysilane (TEOS), L-Lysine, Gold(III) chloride
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40 101 trihydrate ($\text{HAuCl}_4 \cdot 4\text{H}_2\text{O}$, 99.9%) and all other chemicals were of analytical grade and used as received.
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43 102 Phosphate buffer solutions (1/15 M PBS) with different pH 6.98 were prepared by mixing the stock
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45 103 solutions of KH_2PO_4 and $\text{Na}_2\text{HPO}_4 \cdot 12\text{H}_2\text{O}$. All solutions were prepared with double-distilled water.
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48 104 The synthesis of OMCs-Au nanoparticles and the immobilization of tyrosinase on the surface of
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50 105 nanoparticles were achieved according to the procedure introduced by Tang et al⁹.
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53 54 106 **Procedures**

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4 107 The preparation of tyrosinase/OMC-Au/L-Lysine/Au/GCE and the measurements of CC and HQ
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6 108 were carried out as described in our previous work⁹. Briefly, the AuNPs and L-Lysine were
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8 109 immobilized on glassy carbon electrode by electrochemical method. OMC-Au/L-Lysine/Au/GCE was
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10 110 prepared by casting 5.0 μL of the OMC-Au suspension onto the surface of the L-Lysine/Au/GCE,
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12 111 Finally, tyrosinase was immobilized on the electrode surface, as presented in Scheme S1. Au
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14 112 nanoparticles (AuNPs) modified glassy carbon electrode (GCE) due to their high effective surface area,
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16 113 nano-scaled dimension effects, and most importantly, binding affinity with L-Lysine. In addition,
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18 114 L-Lysine provided amino and became the cross-linking agent between AuNPs film and OMC-Au film,
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20 115 and OMC-Au could not only unite with L-Lysine, but also combined with tyrosinase. This makes the
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22 116 enzyme more fixed on the biosensor, accelerates the electron transfer from the enzyme-catalysed redox
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24 117 reaction to electrode surface, and extend its using life as well⁹. Under the optimized condition, 10 mL
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26 118 compost extract samples containing different concentrations of CC and HQ were added into an
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28 119 electrochemical cell, and then the three-electrode system was installed on it. The DPV was recorded
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30 120 from +0.6 to -0.2 V with pulse amplitude of 0.05 V, pulse width of 0.05 s, and pulse period of 0.2 s.
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32 121 The CV was performed between -0.6 and +0.8 V at scan rate of 50 $\text{mV}\cdot\text{s}^{-1}$, sample interval of 0.0001
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34 122 V and quiet time of 2 s.

123 **Preparation of compost extracts**

124 The biosensor simultaneous determination of the CC and HQ concentration was applied in
125 compost bioremediation. The composting process has been introduced previously¹⁷. The components of
126 compost were soil, straw, restaurant leftover, and bran, and the water ratio was 51%. The soil was
127 collected from 100 cm underground on the unfrequented hillside of Yuelu Mountain (Changsha, China),
128 from which large organic scraps were removed. Then aerobic compost was managed 40 days under the

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4 129 condition of 30 °C temperature and 0.033m³·h⁻¹ ventilation. 10 g of compost sample was placed in a
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6 130 flask and 200 mL water was added in. The suspension was agitated on a mechanical vibrator at 200
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9 131 rpm for 2 h. The supernatant was centrifuged at 10000 rpm for 5 min, and then filtered to get the filtrate
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11 132 as the compost extract. All the work was done at room temperature unless otherwise mentioned. The
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13 133 dosage of CC and HQ into each compost extract was controlled using certain volumes of CC and HQ
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16 134 stock solutions.

135 **Data processing**

136 Chemometric processing was done by specific routines in MATLAB 7.0 (Math Works, Natick,
137 MA) written by the authors, using Neural Network Toolboxes to develop the ANN models. SigmaPlot
138 12.0 (Systat Software Inc, California, USA) was used for graphic representations of data and results.

139 The measured data of a total set of 38 samples using the biosensor were divided into three datasets.
140 22 samples for the training set were used to build the proper modeling of the response, 8 samples
141 randomly distributed for the testing set were used to estimate the modeling performances, and another
142 8 extract samples were used to validate the ANN model application. The biosensor DPV responses of
143 compost samples with corresponding CC and HQ concentrations were analyzed using a feed-forward
144 back propagation (BP-ANN). This artificial neural network model for variable selection aims to find an
145 optimal set of inputs that can quickly and successfully classify or predict the desired outputs. It was a
146 feed-forward network combining a back propagation algorithm which was used to train the network
147 according to a learning rule¹⁸. For each sample, a complete DPV was recorded for forming the array
148 and data. In order to reduce the high dimensionality of the recorded signals, to prevent larger numbers
149 from overriding smaller ones, and to prevent premature saturation of hidden nodes, which impedes the
150 learning process, a pre-processing stage was required. There is no one standard procedure for

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4 151 normalizing inputs and outputs¹⁹. But it is recommended that the data be normalized between slightly
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6 152 offset values such as 0.1-0.9 and to avoid saturation of the sigmoid function leading to slow or no
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9 153 learning^{20,21}. For this, the input values of both the training and the test subsets were kept in interval
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11 154 [0.1,0.9] corresponding to the range of the normalized function:

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$$X_i = 0.1 + 0.8 \left(\frac{Z_i - Z_{i \min}}{Z_{i \max} - Z_{i \min}} \right) \quad (1)$$

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16 156 where X_i is the normalized value of input variable, Z_i is the original value, and $Z_{i \max}$, $Z_{i \min}$ are
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18 157 the maximum and minimum original values of primitive data, respectively. After simulation of the
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21 158 networks, the estimated results were reconverted by inverse function of Eq. (1) to be compared with the
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24 159 target values.

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26 160 For complete assessment of model performances, the root mean square error (RMSE) was used,
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29 161 which was calculated between expected and predicted concentration values for each sample (i) and for
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31 162 each of the two analytes (j) considered:

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$$RMSE = \sqrt{\frac{\sum_{ij} (Z_{ij} - \hat{Z}_{ij})^2}{3n - 1}} \quad (2)$$

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37 164 **Results and discussion**

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40 165 **Artificial neural network architecture**

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43 166 In present study, examples of the different curves of current versus time were obtained
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45 167 corresponding to the mixed CC and HQ concentration in spiked compost extract samples. Fig. 1 shows
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48 168 the current response curves for 22 compost extract samples in the training set. The concentrations of
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51 169 CC and HQ in the filtrates both varied from 0.10 to 110 μM . In addition, Fig. 1 presents that a
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54 170 maximum and a minimum signal (any of the 38 currents) of the target concentration were included in
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57 171 the training set, avoiding the need for extrapolation when test the model with the external dataset. It
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59 172 will not give precise results to assign a specific reduction peak potential to each phenolic compound
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4 173 using the statistics of the fitted regression linear model¹², due to some signal overlapping (as shown in
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6 174 Fig. 1). Therefore, BP-ANN method was used to deconvolve the strong overlapping signal and to
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9 175 quantify the concentrations of two phenolic compounds separately, because ANN modelling was
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11 176 considered to be an appropriate chemometric tool for solving overlapping and nonlinear problems,
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14 177 whose structure was designed to imitate the organization of human brain²².

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16 178 “Here Fig. 1”

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19 179 Generally, a BP-ANN comprises three parts: an input layer, an output layer and in between the
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21 180 two layers, there are one or more hidden layers²³. Each layer is formed by a series of interconnected
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24 181 neurons, and the value at each neuron is weighted and transformed by a transfer function²⁴. A
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26 182 simplified scheme of the procedure followed for the measurement and data processing could be seen in
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29 183 Fig. 2. The architecture of the ANN used was defined by these data: the response curves of 22 samples
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31 184 for the training set, the response curves of 8 samples to evaluate model’s response, and another 8
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34 185 extract samples to validate the BP-ANN model application compared with regression liner model. The
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36 186 input layer consisted of a certain number of individual data points of each DPV curve and the output
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39 187 layer consisted of two neurons, namely the two concentrations sought. We used a single intermediate
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41 188 layer, known as the hidden layer, since it was stated that an appropriate level of modelling could be
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44 189 achieved with a single hidden layer in the electrochemical signal resolving process in the relative
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46 190 literature²⁵. So did our experience in previous work also show^{8,26}. Thus, Networks with more than one
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49 191 hidden layer were not considered.

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51 192 “Here Fig. 2”

52 53 54 193 **Network optimization**

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56 194 A study of the BP-ANN architecture was performed in order to optimize the separate

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4 195 quantifications of the two phenols considered. 22 current intensities at specific potentials for the array
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6 196 of DPV were selected as input vector in the BP-ANN, the corresponding concentrations being the
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9 197 targets that the modelling should reach. The learning accomplished (the degree of modelling) was
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11 198 estimated by the root mean square error (RMSE, equation 2). The training process was continued until
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13 199 a preset fitness degree was achieved (RMSE value). Fig. 2 shows the BP-ANN architecture and scheme
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16 200 of this BP-ANN based approach. There are four elements that comprise the ANNs architecture: (a)The
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19 201 number of layers, (b)The number of neurons in each layer, (c)The activation function of each layer,
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21 202 (d)The training algorithm (because this determines the finalvalue of the weights and biases). The
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23 203 number of neurons in each of these two layers is specified by the number of input and output
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26 204 parameters that are used to model each problem so it is readily determined. Therefore, the objective is
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29 205 to find the number of neurons in hidden layer firstly²⁴. Besides, the effects of different transfer function
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31 206 combinations and hidden neuron numbers on the network performance were studied synchronously.
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33 207 Combinations of tan-sigmoidal (Tansig), sat-lineal(Satlin), pure-lineal (Purelin) and log-sigmoidal
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35 208 (Logsig) transfer functions and the hidden neuron numbers (changed from 2 to 16) were tested, as seen
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38 209 on Fig. 3A, with the optimum results of 27 as input neuron number and Levenberg-Marquardt
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41 210 backpropagation (trainlm) as optimization algorithm. Each architecture was retrained five times to get
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43 211 the average RMSEs for the external test set to result in a accurate measure of performance. According
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46 212 to Fig. 3A, the lowest RMSE value was obtained with 10 hidden neurons and Logsig-Purelin as
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49 213 transfer function.

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51 214 Afterwards, the next step was to determine the importance of network inputs and different
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53 215 optimization algorithms. Similarly, the effects of the input neuron and different optimization algorithms
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56 216 on the model performance were evaluated and optimized in parallel. Fig. 3B shows the RMSEs for
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4 217 different input neuron numbers and optimization algorithms with the optimal transfer function
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6 218 combination of Logsig-Purelin and hidden neuron number of 10. According to Fig. 3B, the BP-ANN
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9 219 models with trainbr (Bayesian regularization backpropagation), trainbfg (BFGS quasi-Newton method),
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11 220 traingdm (momentum backpropagation), traingcb (Powell-Beale restarts), traingd (gradient descent
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13 221 backpropagation) and traingdx (backpropagation) as optimization algorithms, respectively, could not
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16 222 meet the performance goal and lowest RMSE. So those algorithms were not taken into account.
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19 223 Trainlm (Levenberg-Marquardt backpropagation) was chosen as the one for the best performance.
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21 224 Once the BP-ANN model was trained, inputs that made relatively small contributions to the variance in
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23 225 our experiment, and it was reasonable that the accuracy of the simulation of the ANN model might
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26 226 increase with more input current values, but the training time was prolonged remarkably with no
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29 227 obvious decrease of RMSE. Therefore, the value number of 9 was selected as the input neuron number
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31 228 with adequate accuracy of simulation.

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34 229 “Here Fig. 3”

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36 230 For all these reasons, the best model was obtained by using a $9 \times 10 \times 2$ network that used a Logsig
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39 231 transfer function in the hidden layer and a Purelin function in the output layer with
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41 232 Levenberg-Marquardt backpropagation (trainlm) as optimization algorithm (shown in Table 1).

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44 233 “Here Table 1”

45 46 234 **Performance of the best ANN**

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49 235 Fig. 4 presented the training performances for the two analytes with the optimal BP-ANN
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51 236 configuration, where the predicted concentrations of the two considered phenols were compared with
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54 237 their expected concentrations. The concentrations of CC and HQ added in compost extract in the
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56 238 experiment both varied between 0.10 and 110 μM . Error bars were plotted by five different retrainings

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4 239 with random reinitialization of weights for the best architecture, giving information about the
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6 240 reproducibility of model. According to Fig. 4, an excellent ability to represent the information on the
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9 241 learning process was obtained with BP-ANN. More valuable was the modelling and prediction
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11 242 capability working with dataset not included in the learning process. Fig. 5 showed the performance of
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14 243 the best ANN on the external testing subset, with data not included in the learning process. Prediction
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16 244 capability of the model could be considered satisfactory due to the very good correlations were
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19 245 obtained in all cases.

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21 246 “Here Fig. 4”

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24 247 “Here Fig. 5”

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26 248 **Comparison of prediction results between regression model and ANN model in composting**
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29 249 **system**

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31 250 In order to compare the performance of the BP-ANN model with the linear regression model in
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34 251 respect to correlation coefficient, adaptability to uncertainty, etc., some compost extract samples were
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36 252 spiked with various amounts of the two phenolic compounds distributed in the range of the
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39 253 experimental design. These were prepared and analyzed employing the BP-ANN model and linear
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41 254 regression model. Both the linear model composed of Eqs. (3) and Eqs. (4) obtained in our previous
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44 255 work⁹ and the BP-ANN model established here were applied into composting system to predict CC and
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46 256 HQ concentrations in eight compost extract samples.

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49 257 $P_{\text{HQ}} = -66.954 - 9.53571g[\text{HQ}]$ (P_{HQ} : μA , $[\text{HQ}]$: M); (R = 0.9565) (3)

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51 258 $P_{\text{CC}} = -88.394 - 13.0811g[\text{CC}]$ (P_{CC} : μA , $[\text{CC}]$: M); (R=0.9771) (4)

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54 259 Practically, there exist a variety of organic compounds in compost extract, such as aromatic,
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56 260 aliphatic, phenolic and quinolic derivatives with varying molecular sizes and properties. It was a

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4 261 complex mixture with diversity, nonlinearity, and uncertain characteristics. In this case, although high
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6 262 specificity and selectivity of biosensor were obtained, when linear model is applied to determine the
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9 263 real samples, the overlapped differential pulse voltammetry signal and the concentration of analyte
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11 264 often exceeds the linear detection range of biosensor, which will affect the accuracy of determination.
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14 265 Therefore, for the sake of obtaining a more applicable and convenient detection method, the
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16 266 combination of biosensors with BP-ANN modelling may turn out to be an alternative tool to classical
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18 267 methods, taking benefit of the advantages of both parts. On one hand, the selectivity, reproducibility
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21 268 and stability of biosensor confirmed the potential applicability for the simultaneous determination of
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24 269 CC and HQ in real environmental samples⁹. On the other hand, the use of ANNs modelling to
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26 270 deconvolve complex signals can enlarge the detection range, and then make the quantification and the
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28
29 271 result analysis more efficient²⁵.

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31 272 In this study, The DPV peak currents of HQ and CC were linear with correlative concentrations
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33 273 over the range from 4.0×10^{-7} to 8.0×10^{-5} M⁹, while the BP-ANN model can directly analyze CC and
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35 274 HQ concentrations varying between 1.0×10^{-7} and 1.1×10^{-4} M. Each of the calibration was done five
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37 275 times with the relative standard deviations (RSD) not more than 5%. Also in this case, the recovery
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39 276 yield for the two phenolic compounds was calculated, which is summarized in Table 2. As can be seen,
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42 277 the recovery yield of CC obtained by linear regression model ranges from 73.9% to 115.2%, while that
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45 278 obtained by BP-ANN model ranges from 96.0% to 115.3%. It is also observed that the recovery yield
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48 279 of HQ calculated by linear regression model ranges from 74.6% to 119.0% , while that calculated by
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51 280 BP-ANN model ranges from 88.15% to 112.0%. As seen on Table S1, the RSD in the linear regression
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54 281 model for CC and in the ANN for CC were 7.73% and 3.7781%, respectively. Although the RSD of
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56 282 linear regression model in the compost extract sample of 4 is lower than the RSD of ANN, the RSD of
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4 283 the rest samples are lower when analyzed by the ANNs model. In addition, the average (RSD) of ANN
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6 284 is lower than the RSD of linear regression model. What's more, the RSD in the linear regression model
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8 285 for CC (21.2004%) is significantly higher than the RSD for the ANN (2.1151%) when sample
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10 286 concentration exceeded the linear range of the biosensor. Correspondingly, as seen on Table S2, the
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12 287 RSD in the linear regression model for HQ and in the ANN for HQ were 10.9592% and 4.8468%,
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14 288 respectively. Obviously, the average (RSD) of ANN is lower than the RSD of linear regression model.
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19 289 The results demonstrated that the prediction results by the ANN model were more precise than the
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21 290 linear regression. The prediction result by linear regression was far from accurate at high levels of CC
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23 291 and HQ beyond the linear range, while the fitting degree of experimental and predicted value using the
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25 292 ANN model were satisfactory (see table 2), thus confirming the BP-ANN model was superior to the
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27 293 linear regression especially for the determination of high levels of CC and HQ in the compost system.
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29 294 Furthermore, the results also showed that the correlation coefficient, adaptability to uncertainty, etc.,
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31 295 obtained after combining the biosensor with BP-ANN were superior to direct linear determination of
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33 296 the CC concentration by the biosensor in the compost system. Obviously, combined with the BP-ANN
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35 297 model, the direct detection range for CC and HQ in the compost system of the biosensor were widened,
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37 298 and the satisfactory results confirmed the potential applicability of the biosensor for quantification of
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39 299 CC and HQ in real compost extract sample determination.
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46 300 "Here Table 2"
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49 301 **Conclusions**

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51 302 In summary, a very good quantification of the two phenolic compounds has been achieved by
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53 303 using the tyrosinase biosensor to get specific signal and BP-ANN as the tool for building the response
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55 304 model. From all the results shown above, it is demonstrated that the combination of tyrosinase
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4 305 biosensor and BP-ANN can give satisfactory quantifications of the CC and HQ concentration
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6 306 simultaneously in composting system with good rapidity and sensitivity. Besides, the direct detection
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8 307 range for CC and HQ of the biosensor was extended to 1.0×10^{-7} - 1.1×10^{-4} M, which was superior to the
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10 308 direct determination by the biosensor with linear data analysis. What's more, this assay provided the
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12 309 potential applicability of the biosensor for the quantification of CC and HQ in composting system
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14 310 though with plenty of interfering substances. In future work, this biosensor combined with artificial
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16 311 neural networks model may be alternatively applied for the quantification of different phenolic
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18 312 mixtures in real contaminated compost samples or other complex environments samples.
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4 **Table captions**
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6 **Table 1.** Optimal results of ANN architecture and training parameters.
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8 **Table 2.** Detailed results obtained for the spiked compost extract samples against added concentrations
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11 of the two phenolic compounds considered. Recovery yield was also expressed for each compost
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14 extract sample.
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Figure captions

Fig. 1 Measured signals were obtained from 22 compost extract samples using in the training set.

Fig. 2 Example of the ANN architecture used to interpret DPV signals. The input vector comprises 9 to 27 individual data points in the DPV curve. The number of hidden neurons ranges from 2 to 16 (for clarity, only 10 are shown here).

Fig. 3 Obtained RMSEs in: (A) prediction for different transfer function combinations and neuron numbers in the hidden layer with input neuron number of 27 and Levenberg-Marquardt backpropagation (trainlm) as optimization algorithm. (B) prediction for different input neuron numbers and optimization algorithms with the optimal transfer function combination of Logsig–Purelin and hidden neuron number of 10.

Fig. 4 Modeling performance achieved for the optimized BP-ANN with 22 samples from the training set. Error bars correspond to 5 different retrainings with random reinitialization of weights for the final architecture. Expected concentrations are plotted against those obtained from the BP-ANN, good correlations were obtained for catechol and hydroquinone.

Fig. 5 Modelling performance of the optimised BP-ANN for the external test set. Expected concentrations are plotted against those obtained by BP-AN. Good correlations were obtained for catechol and hydroquinone.

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Table 1. Optimal results of ANN architecture and training parameters.

Architecture / parameter	Value
Input neuron number	9
Hidden neuron number	10
Output neuron number	2
Transfer function in the hidden layer	Logsig
Transfer function in the output layer	Purelin
Optimization algorithm	Levenberg-Marquardt backpropagation (trainlm)

Table 2. Detailed results obtained for the spiked compost extract samples against added concentrations of the two phenolic compounds considered. Recovery yield was also expressed for each compost extract sample.

Compost extract sample	CC concentration / μM					HQ concentration / μM				
	Added	^L predicted	^B predicted	^L Recovery	^B Recovery	Added	^L predicted	^B predicted	^L Recovery	^B Recovery
1	1.3	1.1 \pm 0.37	1.5 \pm 0.33	84.6%	115.3%	2.5	2.0 \pm 0.46	2.8 \pm 0.36	80.0%	112.0%
2	4.6	5.3 \pm 0.41	5.0 \pm 0.18	115.2%	108.7%	15.5	14.3 \pm 0.39	14.9 \pm 0.28	92.3%	96.1%
3	17.8	17.4 \pm 0.23	17.9 \pm 0.11	97.8%	100.6%	20.5	20.0 \pm 0.44	20.6 \pm 0.37	97.6%	100.5%
4	25.6	27.5 \pm 0.44	28.0 \pm 0.17	107.4%	109.4%	36.3	31.2 \pm 0.40	32.0 \pm 0.16	86.0%	88.15%
5	32.3	30.5 \pm 0.39	31 \pm 0.29	94.4%	96.0%	10.5	12.5 \pm 0.29	8.9 \pm 0.19	119.0%	88.6%
6	39.5	37.7 \pm 0.35	40.3 \pm 0.30	95.4%	102.0%	60.5	57.1 \pm 0.36	58.6 \pm 0.23	94.4%	96.9%
7	59.3	63.5 \pm 0.42	60.2 \pm 0.15	107.1%	101.5%	83.6	65.8 \pm 0.32	85.9 \pm 0.29	78.7%	102.8%
8	95.5	70.6 \pm 0.47	98.4 \pm 0.26	73.9%	103.0%	105.4	78.6 \pm 0.38	109.8 \pm 0.21	74.6%	104.2%

^BBP-ANN model

^Llinear model

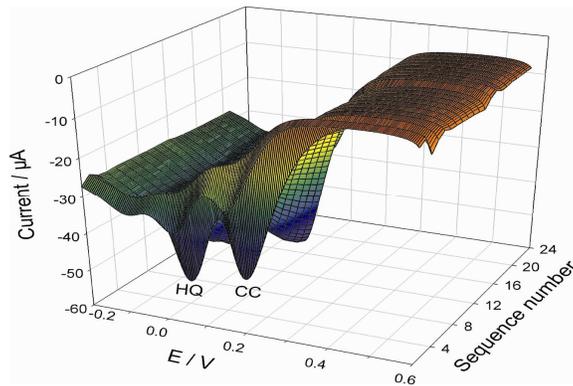


Fig. 1

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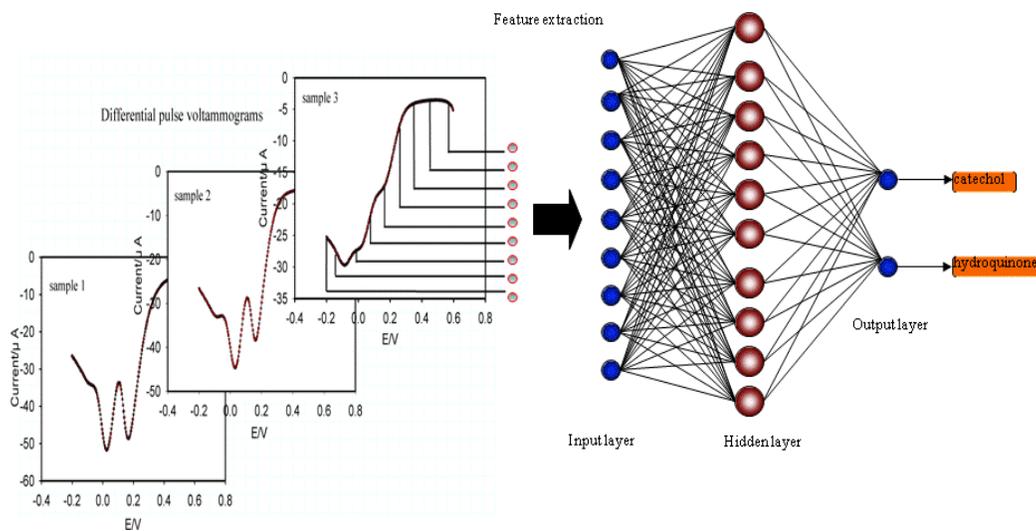


Fig. 2

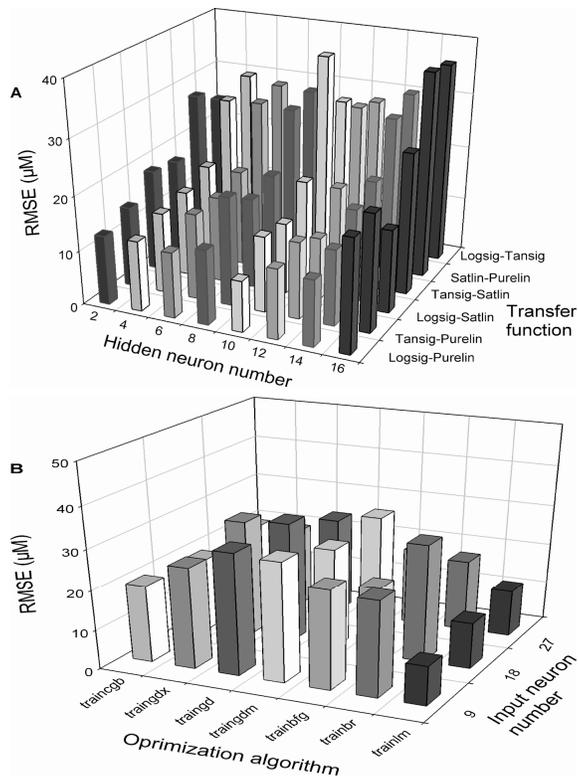


Fig. 3

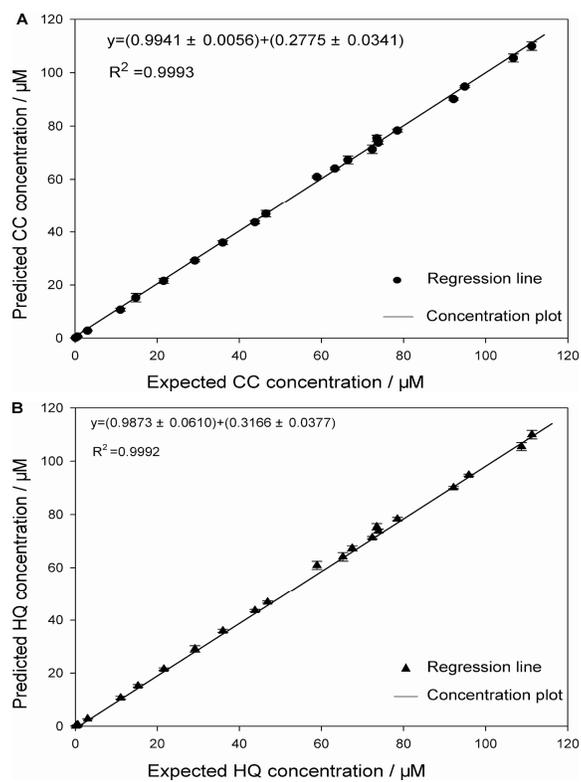


Fig. 4

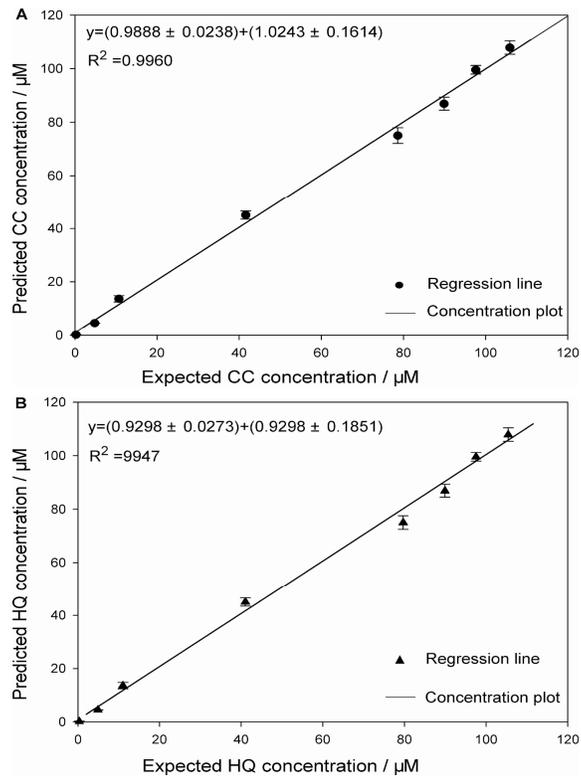
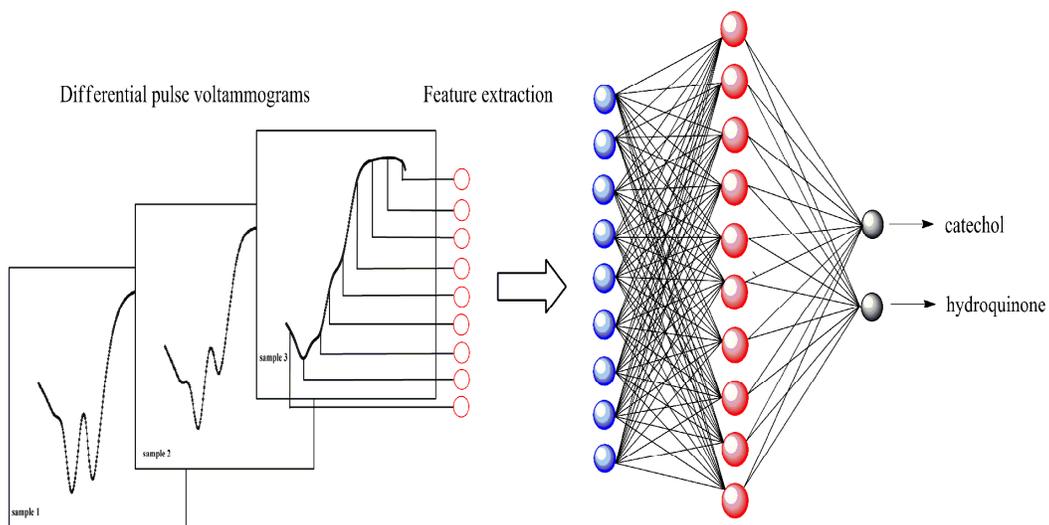


Fig. 5

Contents entry



Selected current intensities of the DPV from biosensor are taken as the input in the ANN. Appropriate weights and biases are applied by the learning algorithm until the targets are satisfied.