



Rapid Plant Uptake of Isothiazolinone Biocides and Formation of Metabolites by Hydroponic Arabidopsis

Journal:	Environmental Science: Processes & Impacts
Manuscript ID	EM-ART-04-2022-000178.R1
Article Type:	Paper



2 3 4 5	1	Rapid Plant Uptake of Isothiazolinone Biocides	
6 7 8	2	and Formation of Metabolites by Hydroponic	
9 10 11 12	3	Arabidopsis	
13	4		
14 15 16	5	Claire P. Muerdter, ^{†,‡} Megan M. Powers, ^{†,‡} Sraboni Chowdhury, ^{†,‡} Alyssa L. Mianecki, ^{†,‡} and	
17 18	6	<i>Gregory H. LeFevre</i> ^{*,†,‡}	
19 20 21 22	7		
23 24 25 26	8		
27 28 20	9	[†] Department of Civil and Environmental Engineering, University of Iowa, 4105 Seamans	
29 30 31	10	Center, Iowa City, Iowa, 52242, United States	
32 33	11	[‡] IIHR—Hydroscience and Engineering, University of Iowa, 100 C. Maxwell Stanley Hydraulics	
34 35 26	12	Laboratory, Iowa City, Iowa, 52242, United States	
30 37 38	13		
39 40	14		
41 42 42	15		
45 44 45	16		
46 47	17		
48 49	18		
50 51 52	19	* Corresponding Author: gregory-lefevre@uiowa.edu; Phone: 319 335 5655, Department of	
52 53 54	20	Civil and Environmental Engineering, 4105 Seamans Center for Engineering, University of	
55 56 57	21	Iowa, Iowa City IA, 52242, United States	
58 59 60		Muerdter et al.	1

22 Abstract

> Isothiazolinones biocides are water-soluble, low molecular weight, nitrogenous compounds widely used to prevent microbial growth in a variety of applications including personal care products and building façade materials. Because isothiazolinones from buildings wash off and enter stormwater, interactions with terrestrial plants may represent an important part of the environmental fate of these compounds (e.g., in green stormwater infrastructure). Using the model plant Arabidopsis thaliana grown hydroponically, we observed rapid (299% within 24 hours), plant-driven removal of four commonly used isothiazolinones: benzisothiazolinone (BIT), chloromethylisothiazolinone, methylisothiazolinone, and octylisothiazolinone. No significant differences in uptake rate occurred between the four compounds; therefore, BIT was used for further detailed investigation. BIT uptake by Arabidopsis was concentration-dependent in a manner that implicates transporter-mediated substrate inhibition. BIT uptake was also minimally impacted by multiple BIT spikes, suggesting constituently active uptake. BIT plant uptake rate was robust, unaffected by multiple inhibitors. We investigated plant metabolism as a relevant removal process. Proposed major metabolites significantly increased in the BIT-exposure treatment compared to control included: endogenous nicotinic acid (confirmed with a Reference Standard) and phenylthioacetohydroximic acid, a possible amino acid-BIT conjugate, and two accurate masses of interest. Two of the compounds (phenylthioacetohydroximic acid and TP 470) were also present in increased amounts in the hydroponic medium after BIT exposure, possibly via plant excretion. Upregulation of endogenous plant compounds is environmentally significant because this demonstrates that BIT impacts plant biology. The rapid plant-driven isothiazolinone removal observed here indicates that plant-isothiazolinone processes could be relevant to the environmental fate of these stormwater compounds.

Muerdter et al.

1 2		
3 4	45	
5 6	46	Environmental Significance Statement
7 8	47	Isothiazolinone biocides are used in outdoor building products, such as paint and building
9 10 11	48	facades, to prevent microbial growth. These chemicals can wash off and enter stormwater, and
11 12 13	49	thus interactions with terrestrial plants may represent an important part of isothiazolinone's
14 15	50	environmental fate. This work uses a hydroponic model plant to demonstrate rapid (<24 h),
16 17	51	concentration-dependent plant uptake of isothiazolinones, including previously untested
18 19 20	52	compounds. This removal is not disrupted by several known plant uptake inhibitors, increasing
20 21 22	53	the likelihood of its environmental relevance. The rapid removal kinetics implicate active uptake,
23 24	54	an understudied mechanism that can inform the environmental behavior of other compounds.
25 26	55	Upregulation of endogenous plant compounds following isothiazolone exposure is
27 28 29	56	environmentally significant because this is an indication that plant biology is impacted.
30 31		
32 33		
34 35		
36 37		
38 39		
40 41		
42		
43 44		
45		
46 47		
48		
49 50		
51		
52		
53 54		
55		
56		
57 58		
59		Muerdter et al.
60		

1. INTRODUCTION

Isothiazolinones are water-soluble, low molecular weight, nitrogenous biocides widely used to prevent microbial growth in a variety of applications including in industrial compounds, personal care products (e.g., shampoo,¹ cosmetics^{2,3}), and building façade materials (e.g., paint⁴). Isothiazolinones from building products wash off of buildings and enter stormwater during precipitation events^{5,6} with concentrations (measured in separate stormwater sewer pipes or an underground stormwater storage pond) of up to 1,600 ng/L benzisothiazolinone (BIT),⁵ 150 ng/L of methylisothiazolinone (MIT),⁷ 41 ng/L chloromethylisothiazolinone (CMI),⁷ and 67 ng/L OIT.⁵ These four compounds (Table S1) are classified by the European Chemicals Agency as 'very toxic to aquatic life with long-lasting effect'; thus, the presence of these biocidal compounds in environmental waters through stormwater runoff could negatively impact wildlife.⁸⁻¹¹ Octylisothiazolinone (OIT) has been reported in three of 17 soil samples collected under home façades in Denmark and was attributed to stormwater.¹² Isothiazolinones can degrade in soil, with relatively short half-lives (e.g., $t_{1/2}$: MIT=0.28, BIT=0.52, OIT=9.3 day).^{12,13} Nevertheless, those half-lives allow sufficient time for plant interaction with these compounds, particularly with repeated dosing through multiple storm events. Due to their high water solubility, isothiazolinones in stormwater can also infiltrate: OIT has been reported in groundwater, attributed to stormwater infiltration.¹⁴ Conventional wastewater activated sludge^{15,16} treatment can decrease isothiazolinone concentrations;^{5,17} however, measurable isothiazolinones have been found in wastewater effluent.⁵ Thus, in locations where recycled wastewater is used for irrigation, isothiazolinones may interact with food crops and potentially be taken up into the plants.¹⁸

Page 5 of 43

1 2

3	
4	
5	
6	
7	
/ 0	
ð	
9	
10	
11	
12	
13	
14	
15	
15	
16	
17	
18	
19	
20	
21	
22	
<u>רר</u> לע	
∠.) \	
24 25	
25	
26	
27	
28	
29	
30	
31	
י כ ככ	
3Z	
33	
34	
35	
36	
37	
38	
39	
10	
-⊤U / 1	
41	
42	
43	
44	
45	
46	
47	
48	
49	
77	
JU 71	
5 I	
52	
53	
54	
55	
56	
57	
58	
50	
72	
oU	

79 The extent and rate of plant uptake of anthropogenic chemicals such as isothiazolinones 80 by vegetation must therefore be understood, both for beneficial applications (*i.e.*, 81 phytoremediation) and to characterize potential exposure risk (*i.e.*, groundwater used as a 82 drinking water source, possible crop uptake during water recycling). Plants are known to take up 83 a variety of anthropogenic chemicals from water^{19,20} via multiple pathways. Some contaminants are taken up by plants passively with water in the transpiration stream.²¹ The transpiration stream 84 85 concentration factor (TSCF, the ratio of the concentration of the chemical in the xylem sap over the concentration in the solution surrounding the plant)²² is a common measurement of 86 87 xenobiotic plant uptake; however, the TSCF does not distinguish between plant uptake routes. Other chemicals, including some organic nitrogen compounds,^{23,24} are transported into the plant 88 89 via transporter proteins that can cause the compound's accumulation rate in the plant to exceed 90 the transpiration rate. For example, uptake of the anticorrosive benzotriazole and tire rubber 91 vulcanizer mercaptobenzothiazole into hydroponic Arabidopsis thaliana (Arabidopsis) exceeds 92 the transpiration rate.^{25,26} Thus, understanding the kinetics of plant uptake of a compound is 93 critical to probe plant uptake mechanisms and predict if plant uptake is likely to occur in a given 94 contact time between the chemical and plant (e.g., during stormwater infiltration through the root 95 zone).

In addition, identifying in-plant transformation products generated following plant uptake
is important to understanding the metabolism of xenobiotic compounds. The classical model of
xenobiotic plant metabolism begins with "Phase I" functionalization of a compound, e.g.,
oxidations such as hydroxylation, hydrolysis, and dealkylation.^{21,27–29} These reactions, mainly
catalyzed by cytochrome P-450 enzymes, create increased-polarity products.^{28,29} The product is
conjugated with plant molecules, such as glutathione, in "Phase II" metabolism,^{29–31} which may

proceed with multiple different pathways for a single xenobiotic²⁷ and generally forms a more
water-soluble product with lowered toxicity.^{30,31} Some xenobiotics, e.g. 2,4,6-tribromophenol
with its hydroxyl group in the parent compound,³² can proceed directly to Phase II metabolism
without Phase I metabolism. Thus, the functional group chemistry of the parent compound is
important to plant metabolism. After Phase II, conjugated metabolites can be more easily
sequestered into vacuoles or bound residues in the plant ("Phase III" metabolism) than the parent
compound, which removes them from harming plant processes.^{29,31}

Nevertheless, recent evidence demonstrates that some xenobiotic compounds do not follow all the phases of classical plant metabolism. For example, recent evidence indicates that conjugated metabolites can subsequently deconjugate^{33,34} and/or be excreted from the plant,²⁵⁻ ^{27,32,35–37} presenting previously unknown exposure routes. Additionally, amino-acid conjugated xenobiotics may closely resemble natural plant compounds and be incorporated into those biosynthesis pathways rather than simply being sequestered. For example, benzotriazole can conjugate with the amino acid alanine to form structural analogues of tryptophan and storage forms of the plant hormone auxin, likely by following the tryptophan and auxin biosynthesis pathways.²⁵ Benzotriazole was not incorporated into proteins,³⁸ but its presence in biosynthetic pathways may still produce important and as-of-yet undocumented plant physiological responses. Thus, xenobiotic plant metabolites may present important consequences for plants / plant consumers, or potential phytoremediation routes.

Despite the potential for plant-isothiazolinone interactions, including both plant uptake and metabolism, knowledge of these interactions is limited. Two papers describe the same set of experiments and are limited to aquatic plants.^{39,40} Both CMI and MIT were rapidly (first measurement at 20 hours) taken up by both a duckweed (*Lemma minor*) and an aquatic fern

Muerdter et al.

Page 7 of 43

1

2	
3	
1	
-	
5	
6	
7	
8	
0	
9	
10	
11	
12	
13	
14	
14	
15	
16	
17	
18	
10	
19	
20	
21	
22	
23	
23	
24	
25	
26	
27	
28	
20	
29	
30	
31	
32	
22	
22	
34	
35	
36	
37	
20	
38	
39	
40	
41	
42	
12	
45	
44	
45	
46	
47	
-T/ 40	
48	
49	
50	
51	
52	
52	
53	
54	
55	
56	
57	
57	
58	
59	
60	

125	(Salvinia braziliensis). ³⁹ The parent CMI was not found in the plant tissue after one day of
126	exposure, suggesting rapid metabolism of the compound. ⁴⁰ Nevertheless, much more information
127	is needed to understand the fate of isothiazolinones interacting with plants. Plant uptake of BIT
128	or OIT to our knowledge has not been studied. Given the presence of these compounds in
129	stormwater, ^{5–7} this represents a critical knowledge gap. Additionally, stormwater-relevant kinetic
130	data are needed. Understanding initial (<20 hours of exposure) plant uptake and metabolism
131	kinetics is critical to determining the environmental fate of these compounds in stormwater
132	situations, when water may rapidly infiltrate into the soil or engineered soil medium. More
133	detailed kinetics will also inform plant uptake mechanisms, including if uptake rate is inducible
134	by repeated biocide exposure, as would occur with repeated rain events. To the best of our
135	knowledge, there are no published data on terrestrial plant uptake and processing of
136	isothiazolinones, which is critical for stormwater flowing across a landscape or in green
137	infrastructure (e.g., bioretention cells ⁴¹).
138	Therefore, the objective of this work was to quantify the plant uptake kinetics of four
139	isothiazolinones [benzisothiazolinone (BIT), chloromethylisothiazolinone (CMI),
140	methylisothiazolinone (MIT), and octylisothiazolinone (OIT)] by Arabidopsis, a model plant, as
141	well as Arabidopsis plant metabolites. Metabolites are defined herein as inclusive of BIT-
142	conjugates and endogenous plant compounds whose presence was significantly increased or
143	decreased following BIT exposure. Because in this work we were primarily interested in BIT

144 plant transformation products, we quantified those metabolites that increased after BIT exposure

145 (i.e., upregulated), such as BIT conjugates with endogenous plant compounds that are typically

146 formed as part of plant xenobiotic metabolism. We hypothesized that all of the tested

147 isothiazolinones would be rapidly taken up by the Arabidopsis plants via a transporter protein

Muerdter et al.

2	
3	
4	
5	
6	
7	
, 0	
8	
9	
10	
11	
12	
12	
13	
14	
15	
16	
17	
10	
10	
19	
20	
21	
22	
23	
2/	
24	
25	
26	
27	
28	
29	
30	
21	
21	
32	
33	
34	
35	
36	
37	
20	
38	
39	
40	
41	
42	
43	
ΔΛ	
44	
45	
46	
47	
48	
49	
50	
51	
51	
52	
53	
54	
55	
56	
57	
57	
50 50	
59	
60	

148 (i.e., not passively transported with the transpiration stream). All four molecules are relatively 149 small (MW=149.6–213.3 Da) and water-soluble (log K_{ow} values of -0.49–2.61), 4,6,42 but feature 150 different functional groups. To determine the detailed kinetics and possible mechanisms of plant 151 uptake of BIT, as a representative isothiazolinone, removal kinetics at a range of starting BIT 152 concentrations were measured. We tested if Arabidopsis BIT uptake was constituently active and 153 probed the route of plant uptake by testing BIT plant uptake in the presence of molecules with 154 similar structures or known plant uptake inhibitors. Additionally, we investigated metabolites in 155 both plant tissue and the hydroponic medium that increased following BIT exposure, and 156 tentatively propose possible structures for these compounds. 157 158 2. METHODS 159 2.1 Chemicals 160 Chemicals used in these experiments include: benzisothiazolinone ("BIT", CAS 2634-33-161 5, Alfa Aesar, 97%), a chloromethylisothiazolinone (CMI) and methylisothiazolinone (MIT) mix 162 (mix of CAS 26172-55-4 and CAS 2682-20-4, Combi-Blocks, Inc., 68%), octylisothiazolinone 163 ("OIT," CAS 26530-20-1, Tokyo Chemical Industry, >98%), L-tryptophan (CAS 73-22-3,

164 Research Products International, >99%), and 1H-benzotriazole (CAS 95-14-7, Fluka Analytical,

165 \geq 98%). Other chemicals are described in the SI. All LC-MS/MS solvents (acetonitrile, water,

166 and formic acid) were Fisher Optima LC/MS grade.

Using previously established protocols for growing Arabidopsis hydroponically,^{25,26}
 liquid plant medium was generated by combining (for 1 L medium): 4.43 g Murashige and

169 Skoog Basal Medium powder (PhytoTech Labs M519), 0.5 g 2-morpholin-4-ylethanesulfonic

Muerdter et al.

acid (MES) free acid monohydrate (Fisher), and deionized water to ~900 mL. Hydroponic
medium pH was corrected to 5.6 with potassium hydroxide, then 5 g sucrose (Research Products
International) and DI water to 1 L was added. Medium pH was rechecked and adjusted to 5.7 as
needed with potassium hydroxide or hydrochloric acid. Before experimental use, the medium
was filter sterilized using a bottle top filter (Corning #431118, 0.22 µm pore size) into an
autoclaved bottle.

176 2.2 Experimental Design

177 2.2.1 Experimental compound selection

Isothiazolinone compounds were selected to represent a variety of chemical features. BIT is the only compound of the four to contain an aromatic ring. The other three tested molecules (CMI, MIT, and OIT) have other functional groups: chloro (CMI), methyl (CMI and MIT), or a carbon chain (OIT). OIT's eight-carbon chain results in a higher log K_{ow} value than the other three molecules (i.e., 2.6 vs log K_{ow} <1 for the other three compounds, Table S1^{6,7,43}); log K_{OW} is known to influence plant uptake rates.^{44,45} All four compounds tested are used as preservatives in building materials and are thus relevant to stormwater.⁷

185 2.2.2 Arabidopsis growth preceding isothiazolinone exposure

Arabidopsis seeds were surface-sterilized using a previously-published bleach
procedure,^{25,26,38,46} with minor modifications (detailed in the SI). Seeds were then grown
aseptically in the sterile hydroponic Murashige and Skoog-based medium above, in washed and
autoclaved Magenta GA-7-3 boxes (Bioworld), also using a previously published procedure with
minor modifications (detailed in the SI).

Muerdter et al.

> The Arabidopsis plants were grown hydroponically in this work to enable the rapid quantification of relevant plant uptake kinetics and route without competition from soil abiotic mechanisms and abundant microbes. Because Arabidopsis is a model plant, the translation of results to other plant species in soil conditions would require further testing. Previous work has demonstrated that such translation is possible and useful: multiple xenobiotic (benzotriazole) metabolites first discovered in hydroponic Arabidopsis were also found in soil-grown strawberries.³⁸ Thus, the Arabidopsis results allowed for the prediction of plant uptake and targeted metabolite discovery in other plant species.

199 2.2.3 Plant isothiazolinone exposure experiments

The exposure experiments were modeled on previous work^{25,26} and described in detail in the SI. Briefly, after a 10–11 day period of growth in unspiked sterile hydroponic medium, the medium was exchanged for sterile, isothiazolinone-spiked medium. A medium-only abiotic control was also created to quantify non-plant related losses (e.g., photolysis, hydrolysis). Each treatment and control consisted of n=3 or n=4. Sampling of the hydroponic medium occurred throughout the duration of the experiment. Except during active sampling, boxes were maintained in a Percival growth chamber alternating between 16 hours light at 23° C and 8 hours dark at 21° C.

208 2.2.4 BIT uptake kinetics under varied concentrations / multiple exposures

To determine if the plant uptake rate of BIT changed with increasing BIT concentration, a plant isothiazolinone exposure experiment was conducted with four starting BIT concentrations run in parallel with an abiotic control. Measured initial concentrations were 8 μ g/L, 112 μ g/L, 678 μ g/L, and 2,127 μ g/L. Medium samples were collected at six timepoints over 48 hours and

Muerdter et al.

Page 11 of 43

1	
2	
3 4	213
5	
6	214
7	
8	215
9	
10	216
11	
12	217
13	
14 15	218
16	210
17	210
18	219
19	220
20	220
21	001
22	221
23	
24	222
25	
26	223
27	
20	224
30	
31	225
32	
33	226
34	-
35	227
36	,
3/	228
38	220
39 40	
41	220
42	
43	
44	230
45	250
46	221
47	231
48	222
49 50	232
50	• • • •
52	233
53	
54	234
55	
56	235
57	
58	

59

60

213 BIT was quantified with LC-MS/MS. The data were fit using nonlinear regression (curve fit, 214 GraphPad Prism 9.0) with zero, first, and second-order equations. Additionally, data from seven 215 other BIT starting concentrations (lowest concentration = $2.4 \mu g/L$), using the same experimental 216 design but with each concentration conducted independently rather than in parallel, were curve 217 fit in the same manner and plotted along with the four rate values from the parallel experiment to 218 make a total of 11 data points. Calculations of observed vs. expected passive (transpiration-219 driven) BIT removal rate from the medium are described in the SI. 220 A repeated exposure kinetics experiment was also conducted with BIT. Medium was 221 spiked to 1,050 µg BIT/L at t=0, and medium samples were taken at 0, 1, 2, 3, 4, 8, and 24 hours. 222 After the 24-hour sample, the plant medium was drained from the boxes (in the laminar flow

hood biosafety cabinet) while retaining the plant biomass (for the plant treatments) by keeping
the lid on the box and tilting until the medium poured out of the box. Then, 25 mL of freshly

spiked 875 µg BIT/L medium (although the aim was to have both spikes be the same

concentration; this is difficult to achieve as evidenced by other similar studies^{25,26}) was added to each box for the second spike. Medium samples were collected using the same timepoints as the first spike, with t=0 as the time of the second spike.

229 2.2.5 Plant metabolomics experiments

To determine plant metabolites increased after BIT exposure, a nominal 200 µg BIT/L
treatment group of Arabidopsis plants was grown in parallel with a "clean" (unspiked with BIT)
positive control group of Arabidopsis plants. Each group had n=9 sample boxes. Additional
details are in the SI, briefly: plant tissue was harvested at 24 h by straining out the liquid medium
as above, inverting the box onto a paper towel, gently patting the plant tissue dry with the paper
towel to remove any residual medium, and freezing the plant tissue at -20 °C until plant tissue

Muerdter et al.

extraction. Freeze drying and plant tissue extraction into a liquid sample followed a previously
published procedure,²⁵ detailed in the SI.

238 2.2.6 Quantification of BIT sorption to plant tissue

Each cleaned and autoclaved magenta box received 25 mL of plant medium spiked to 213 μg BIT/L (the measured concentration targeted to be similar to the nominal 150 μg BIT/L of the inhibitor experiments) and one vial of freeze-dried plant tissue. Given the rapid uptake and phytotransformation of BIT, freeze-dried (lyophilized) Arabidopsis plant tissue (n=4 plant boxes) was used rather than live plants. Freeze-dried plant tissue has been used to quantify sorption of chemicals to plants in other studies.^{46,47} This decouples plant uptake from sorption, while providing a normalized dry-weight basis for sorption and preserving the tissue more than other methods of removing water.⁴⁸ Each vial contained plant tissue grown from 30±2 seeds for 14 days in hydroponic plant growth medium with no biocide present, which had been harvested per the procedure above, placed into a vial, and freeze dried overnight. Hydroponic medium (1.0 mL) was sampled from each box setup described above (213 µg BIT/L spike with freeze-dried plant tissue) at t = 0, 1 h, and 24 h. BIT sample concentration was determined via LC-MS/MS (see below for LC-MS/MS details).

43 252 *2.1*

2.2.7 Competitive inhibition experiments

Two competitive inhibition experiments were performed to test if BIT uptake would be inhibited by chemically similar molecules in the liquid plant medium. Using the procedure for plant isothiazolinone exposure experiments above, benzotriazole and tryptophan were tested separately in mixtures with BIT. Such mixtures are environmentally relevant, as both compounds are known to be taken up by plants,^{25,49} and both tryptophan (an amino acid)-containing

1 2	
2 3 4	258
5 6	259
7 8 0	260
9 10 11	261
12 13 14 15	262
16 17	263
18 19 20	264
21 22	265
23 24	266
25 26 27	267
28 29	268
30 31	269
32 33 34	270
35 36	271
37 38	272
39 40 41	273
41 42 43	274
44 45	275
46 47	276
48 49 50	277
51 52	278
53 54	279
55 56 57	280
58 50	

compounds and benzotriazole (a corrosion inhibitor) are found in stormwater.^{25,50,51} Treatments were: (a) 50 μ g BIT/L, and (b) a nominal mixture of 50 μ g BIT/L and 50 μ g benzotriazole/L. For tryptophan, a nominal mixture of 50 μ g BIT/L and 67.4 μ g L-tryptophan/L (a molar-equivalent concentration to 50 μ g BIT/L) was used for the (b) treatment.

262 2.2.8 Pathway inhibitor experiments

Known inhibitors to plant uptake and xenobiotic metabolic pathways were used to test 3 1 likely BIT plant uptake pathways. Equivalent Arabidopsis-inhibitor experiments to our 5 experimental design were not found in the literature for all of the tested inhibitors. Thus, prior to 5 conducting a full Arabidopsis-inhibitor experiment, we conducted plant health experiments to 7 determine an appropriate inhibitor concentration for our experimental setup to ensure plant 3 health was not significantly impaired (based on visual inspection) by the inhibitor concentration;) full details are in the SI. Then, BIT-inhibitor experiments were run in the same manner as the) plant health experiments, except an abiotic control was added and medium samples were taken non-sacrificially over time. Each inhibitor was tested separately, with a nominal BIT C_0 of 150 μ g BIT/L and 25 mL of plant medium per box. Samples were taken at t = 0 hr, 2 hr, 24 hr, and 2 48 hr, with n=3-4 for each timepoint and no replacement of plant medium. 3

Known inhibitors of several different pathways were used. Diethylpyrocarbonate is an
amino acid,⁵² peptide,⁵³ and sucrose⁵⁴ plant uptake inhibitor. Glycerol is an aquaporin
inhibitor.^{55,56} Quinidine inhibits organic cation transporters (as used in uptake experiments in *Typha latifolia*).⁵⁷ Anthracene-9-carboxylic acid (9-AC) inhibits anion channels⁵⁸ (demonstrated
in algal plasma membranes,^{59,60} and through the inhibition of auxin movement in Arabidopsis⁶¹
and oat coleoptile tissue⁶²). 2,4-dinitrophenol (2,4-DNP) is a cellular metabolism inhibitor,
which is known to inhibit amino acid/peptide uptake of dileucine (for an Arabidopsis membrane

Muerdter et al.

transport protein expressed in yeast),⁶³ glutamine (in excised castor bean cotyledons),⁶⁴ and GlySar (in barley embryos).⁶⁵ 1-aminobenzotriazole (1-ABT) inhibits cytochrome P450,⁶⁶ which is
involved in xenobiotic metabolism in plants.²⁷

285 2.3 Analytical Methods

286 2.3.1 LC-MS/MS methods

All samples except for the metabolomics samples were analyzed via high-performance liquid chromatography (Agilent 1260) coupled to a triple-quadrupole mass spectrometer (LC-MS/MS: Agilent 6460 Triple Ouadrupole MS with Mass-Hunter, version B.07.00) operating in multiple-reaction monitoring (MRM) positive mode and electrospray ionization (ESI) (Table S2). An isotopically labeled (d4) imidacloprid-normalized external calibration curve was used to account for matrix effects during ionization for those samples that were quantified on a mass per L basis. 10 µL of 1.3 mg/L d4 imidacloprid in LCMS-grade acetonitrile were added to each 1 mL of sample or standard. The chromatography column was an Agilent Eclipse Plus C18 (5 µm, 4.6×150 mm) for all but the OIT samples (which used an Agilent XDB-C18 ZORBAX, 3.5 µm, 2.1×50 mm). The mobile phases were Fisher Optima LC/MS Water with 0.1% Optima LC/MS grade formic acid (A) and Fisher Optima LC/MS acetonitrile with 0.1% Optima LC/MS grade formic acid (B). The mobile phase gradients are given in the SI. The injection volume was 10 µL. The MS/MS was set in multiple reaction monitoring mode (MRM, Table S2). Two MRM transitions were used for each compound (Table S2) for quality control. The instrument response was linear throughout the calibration range. Between non-detect values and the lowest non-zero standard with a reliable peak (\leq 500 ng/L, representing \leq 0.02–21% of the

Page 15 of 43

details are provided in the SI.

starting concentrations in the experiments) for the given MS/MS run, the standard curve was

extended and used to estimate the concentration. The signal-to-noise ratio was ≥ 2 for samples

The extracted plant tissues (extraction details in the SI) were analyzed on a Thermo Q-

Exactive Orbitrap High-Resolution Mass Spectrometer using Full MS scans with data-dependent

MS/MS acquisition, in the manner of previous literature.⁶⁷ Both ESI positive and negative modes

were used. Polarity switching was used for the MS scan (i.e., both positive and negative modes

positive and negative modes separately for composited plant tissue extracts (see SI for details).

method (Table S2) were used on the Q-Exactive for both the full scan and data-dependent scans

3.1 (details given in the SI). Because of the explicit experimental design wherein the entire MS

spectra of BIT-exposed plant extracts were compared to unexposed plant extracts, the data can

Metabolomics with Statistics Detect Unknowns with ID using Online Databases and mzLogic"

(Figure S1) was used. The results were filtered to remove background with the "background is

false" filter. Analysis of plant hydroponic medium samples were composite samples from

multiple biological replicates, thus no p-value calculation was possible. In the plant tissue

include both endogenous plant compounds that increased with BIT exposure and conjugates

formed from BIT. The established workflow within Compound Discoverer, "Untargeted

Q-Exactive data of the metabolomics samples were analyzed via Compound Discoverer

were run in the same sample run). Data dependent MS2 (ddMS2) scans were performed in

The chromatography and method parameters from the Agilent Triple Quadrupole MS BIT

of these samples. Specific parameters for the Q-Exactive are given in Tables S3 and S4.

between non-detect and the lowest non-zero standard. Full analytical and quality assurance

2.3.2 Metabolomics investigation via high-resolution mass spectrometry

1	
2 3 1	303
- 5 6	304
7 8	305
9 10	306
11 12	500
13 14	307
15 16	200
17 18	308
19 20	309
21 22	310
23 24	311
25 26 27	312
27 28 29	313
30 31	314
32 33	315
34 35 36	316
30 37 38	317
39 40	318
41 42	319
43 44 45	320
46 47	321
48 49	322
50 51	323
52 53	324
55	
56 57	325
58	

59

60

Muerdter et al.

extracts, compounds were screened for further analysis based on a p-value ≤ 0.05 in the fold-change ratio of the peak area of the treatment (BIT-exposed) plant tissue extracts to the peak area in the positive control (BIT-unexposed) plant tissue extracts. Several hundred compounds met these criteria for both the plant extracts and the medium samples, e.g., 546 compounds increased in BIT-exposed plants (Figure S2), which were then sorted from greatest-to-least peak area ratio for each sample type (plant extract or medium). Results with a peak area ratio of five or greater (i.e., increased five-fold or more in BIT-exposed plants vs. unexposed plants) were selected for further analysis. Features decreased in peak area in BIT-exposed vs. unexposed plants were not examined further in this work. Those compounds were then sorted by retention time, and similar retention time compounds (within 0.04 min) were grouped for further examination as possible in-source fragments of the metabolite. Candidate proposed metabolites were drawn in ChemDraw 20.0 (PerkinElmer) and the exact mass of a proposed formula was compared with the accurate mass of the mass spectrometry results. Fragments were also drawn and compared with the spectra for each proposed compound (SI). 2.3.3 Statistics

GraphPad Prism 9 (GraphPad, La Jolla, CA) was used for all statistics. Matched-pairs t-tests were used to compare two treatments and/or controls, with α =0.05. Departure from the linear null slope at the 95% confidence interval determined if a significant change in compound concentration occurred over time. A repeated measures mixed-effects model in Prism (REML, under one-way ANOVA, α =0.05) with Geisser-Greenhouse correction was used to assess differences in plant uptake rate between the four different biocides, given that some timepoints (0 and 24 h) were the same for all biocides and one timepoint varied (1.5 h vs 2.5 h) for OIT vs the three other biocides.

Muerdter et al.

1
2
3
4
5
6
7
/
8
9
10
11
12
13
14
15
16
17
10
10
19
20
21
22
23
24
25
26
20
2/
28
29
30
31
32
33
34
35
36
37
20
20
39
40
41
42
43
44
45
46
47
48
10
50
50
21
52
53
54
55
56
57
58
50

60

3. RESULTS AND DISCUSSION

350 3.1 Rapid hydroponic removal of isothiazolinone biocides is likely due to transporter-mediated 351 plant uptake

352 The presence of different functional groups and isothiazolinone structure did not 353 significantly (*p*=0.48) alter the Arabidopsis-facilitated removal rate from the hydroponic medium 354 for any of the four tested molecules, given the available data from three timepoints (Figure 1). 355 Although additional data points may reveal subtle differences among uptake rates, all four 356 molecules behaved similarly in our experiments such that \geq 99% of the initial concentration of a 357 given spiked compound was removed from the medium at 24 hours in the presence of Arabidopsis plants (Figure 1). Neither the number of rotatable bonds (previously reported to 358 359 influence plant uptake⁴⁵ and greater in OIT than in the other tested molecules due to the OIT 360 carbon tail), the presence of carbon or methyl groups (i.e., in MIT and CMI,), nor the presence of 361 an aromatic ring (i.e., BIT) impacted isothiazolinone removal extent at 24 hr, and with the given 362 limited data significant uptake rate differences were not observed. Plant-driven process(es) 363 accounted for the majority of isothiazolinone removal, as no significant removal of the four 364 isothiazolinones was found in the abiotic controls (Figure 1, $p \ge 0.05$).



Figure 1: Plant-facilitated depletion kinetics of isothiazolinone biocides from hydroponic medium. OIT (nominal C_0 of 150 µg/L), BIT (nominal C_0 of 100 µg/L), and MIT and CMI (as a mixture, nominal MIT C_0 of 33 µg/L, nominal CMI C_0 of 100 µg/L) were all rapidly taken up by *Arabidopsis thaliana*, at a statistically indistinguishable rate (*p*=0.48 for the plants treatments).

Because the hydroponic plant depletion kinetics of all four molecules were similar, we chose to use BIT as a representative isothiazolinone compound for the remaining more-detailed plant uptake investigation. Removal of BIT from the medium was rapid. Sorption of BIT to plant tissue occurred, representing roughly a quarter of total removal of BIT from the medium (Figure S3); however, the majority of total BIT removal was due to other process(es). It is possible that some BIT is transformed (and thus would appear to be removed) from the medium due to root exudates. Although we did not measure potential abiotic interactions with exudates for BIT directly, previous work in the same experimental setup²⁵ found that Arabidopsis root exudates did not create a significant loss of another organic xenobiotic, benzotriazole. Such rapid depletion of BIT from the hydroponic medium implicates BIT active plant uptake, which likely proceeds through a transporter-mediated process rather than merely passive movement of BIT with the transpiration stream into the plant. Indeed, measured BIT removal from the medium exceeded the expected transpiration rate by 21-fold for a C₀ of 357 μ g/L and

2	
3	
4	
5	
6	
7	
/	
8	
9	
10	
11	
11	
12	
13	
14	
15	
16	
10	
17	
18	
19	
20	
21	
יב רכ	
22	
23	
24	
25	
26	
27	
27	
28	
29	
30	
31	
27	
22	
33	
34	
35	
36	
37	
20	
38	
39	
40	
41	
42	
<u>⊿</u> ⊇	
43	
44	
45	
46	
47	
48	
10	
49	
50	
51	
52	
53	
51	
54	
55	
56	
57	
58	

59

60

384 28-fold for a C₀ of 49 µg BIT/L (calculations shown in the SI). This matches literature precedent 385 for uptake of benzotriazole and mercaptobenzothiazole into Arabidopsis plants, which also 386 substantially exceeded the transpiration rates.^{25,26} Combined with the evidence for BIT 387 phytotransformation based on a likely BIT-conjugate present in plant tissue extracts (described 388 below), it is likely that the primary mechanism for BIT removal from the medium is plant uptake 389 via a transporter protein. Uptake of a variety of other xenobiotics, such as herbicides 2.4-D and 390 paraquat, is known to occur via transporters.^{25,68–70} The ability of transporters to move 391 xenobiotics into plants is attributed to the relatively nonselective nature of the transporters and the similarity of xenobiotics to the intended substrates of the transporters, e.g., amino acids.^{25,68} 392 393

394 3.2 Concentration-dependent BIT uptake by Arabidopsis implicates substrate inhibition

395 For all tested BIT concentrations in the range of 2.4–2,127 µg/L, BIT uptake by 396 Arabidopsis was rapid, resulting in complete or near-complete removal within 24 hours (Figure 397 S4). This rapid removal limited the collection of early data points, but with a total of 252 data 398 points from 11 starting concentrations (Figure S4, summarized in Figure 2), a second-order rate 399 model fit the data best (r² of 0.92 for first-order and 0.95 for second-order, Table S5). A second-400 order rate constant is also consistent with plant uptake of other xenobiotics such as 401 mercaptobenzothiazole into Arabidopsis²⁶ and triclosan into carrot cells.⁷¹ In this work, the BIT 402 removal rate increased between the lowest and second-lowest tested concentration then 403 decreased with increasing initial BIT concentration throughout the remainder of the tested 404 concentration range (initial concentration range: 2.4–2,127 µg/L, Figure 2; representing 252 total 405 samples), a behavior consistent with enzymatic substrate inhibition. Substrate inhibition, wherein 406 the reaction rate increases and then decreases with increasing concentration of substrate instead

Muerdter et al.

of rising to a steady reaction rate following traditional Michaelis-Menten kinetics,⁷² is not uncommon within the normal operating substrate concentration range of enzymes.⁷² Indeed, this inhibition is crucial to the normal functioning of many metabolic processes; for example, substrate inhibition can regulate reaction velocity to a more stable range than would occur without inhibition in order to maintain appropriate concentrations of substrates and/or products.⁷² Other plant uptake studies have reported evidence of substrate inhibition of plant uptake over at least part of the tested range of substrate concentration, e.g., uptake of mercaptobenzothiazole;²⁶ however, substrate inhibition occurred at a higher concentration for mercaptobenzothiazole than for BIT. This result contrasts with the reported plant uptake rate of metformin, which is proposed to enter the plant through transporter proteins but whose plant uptake rate did not vary with initial metformin concentration (possibly due to the concentrations affected by substrate inhibition not being tested).⁵⁷ BIT concentrations tested in this study were slightly higher than typical environmental concentrations: the lowest tested initial concentration in this study was 2.4 μ g/L, and BIT stormwater concentrations have been reported up to 1.6 μ g/L⁵. Lower initial concentrations, e.g. 0.9 µg BIT/L, did not yield full curves from which a reaction rate could be quantified due to the rapid plant uptake/sorption of BIT to below detection limits. Although Figure 2 indicates that the highest BIT plant uptake rates occur at concentrations slightly above the environmentally relevant concentrations, BIT plant uptake was rapid (≥99% removal from the medium within 24 hours) at all tested concentrations (Figure S4) including 0.9 µg BIT/L when the BIT medium concentration reached zero so rapidly that a curve could not be constructed, suggesting the potential for high BIT plant uptake rates at environmentally-relevant concentrations.



repeated BIT spiking experiment (Figure 3) demonstrated that the rate of BIT removal increased

Muerdter et al.

from 0.008 L/(μ g-h) in spike 1 to 0.01 L/(μ g-h) in spike 2. The relatively small difference between the removal rates (25% increase) between the two spikes was statistically significant (p=0.045), but this may in fact be merely due to a slight increase in plant biomass between the first and second spikes (which could not be measured in situ during the experiment) and/or the difference in measured C₀ between the first (1,050 μ g BIT/L) and second (875 μ g BIT/L) spike rather than evidence of induction. Generally, the uptake rate was repeatable and did not increase greatly between spikes, indicating an uptake pathway that is likely constituently active. In contrast, a second spike of mercaptobenzothiazole in similar Arabidopsis systems resulted in a reported rate constant between 500% and 2,100% the spike 1 kinetics at any of the various tested concentration levels²⁶—much greater than what we observed for BIT. The large rate increase reported for mercaptobenzothiazole with repeated spiking may be due to being an inducible process rather than the slight increase between BIT spikes in this study (*i.e.*, 5-21X for mercaptobenzotiazole vs. 1.25X for BIT) that suggests a constituently active process.



Figure 3: Concentration of BIT in the hydroponic medium over time during a repeated spiking 460 experiment. The first spike at t=0 hr was 1,050 μ g BIT/L. The second spike at t=25 hrs was 875

Muerdter et al.

1 2	
3	161
4	401
5	402
6	161
/	404
0 9	405
10	466
11	100
12	
13	467
14	,
15	468
16	100
17	469
19	7 07
20	
21	470
22	170
23	471
24	т/1
25	172
20 27	4/2
27	172
29	4/3
30	171
31	4/4
32	175
33	4/3
34 25	170
36	4/0
37	477
38	4//
39	470
40	4/8
41	470
42	4/9
43 44	100
45	480
46	101
47	481
48	
49	482
50	
51	483
5∠ 5२	
55	484
55	
56	485
57	
58	
59	

461 μ g BIT/L. A second-order reaction rate fit both curves the best (vs. zero- and first-order reaction 462 rates). The reaction rate (k) for the first spike was 0.008 L/(μ g-h) and spike 2 is 0.01 L/(μ g-h). 463 The rates were significantly different (p=0.045) but within 25% of each other. n=4 for all plant 464 treatment data points, and n=1–4 for all no-plant control points. Error bars are ± one standard 465 error, with some error bars obscured by the data point symbol.

3.4 BIT uptake by Arabidopsis is robust, as shown by it being unaffected by several
structurally similar compounds and known inhibitors of plant uptake and metabolism
pathwavs

470 BIT plant uptake proceeds through a pathway that was not significantly affected by 471 multiple tested structurally similar compounds (possible competitive inhibitors) added 472 individually to the plant medium. Neither OIT, benzotriazole, nor tryptophan-all with similar chemical structures to BIT—significantly changed the BIT uptake rate (Figures S5, S6, S7). This 473 474 lack of competitive inhibition suggests either a robust shared uptake pathway⁷⁰ or different uptake pathways⁷⁵ for BIT and the tested molecules. The exact mechanism(s) of uptake, 475 476 however, for the four molecules was not determined. Such mixtures are environmentally relevant because all compounds are known to be taken up by plants,^{25,49} and both benzotriazole (a 477 478 corrosion inhibitor) and tryptophan (an amino acid)-containing compounds are found in 479 stormwater.^{25,50,51} Uptake of a single compound being unaffected by similar organic solutes in 480 plant medium solution has previous precedent, e.g., the influx of amino acids into Zea mays roots was independent of other organic solutes in solution.²⁴ 481 482 Additionally, the tested inhibitors of known plant uptake and metabolism pathways did

Additionally, the tested inhibitors of known plant uptake and metabolism pathways did
not have a significant impact on BIT plant uptake (Figure S8), further demonstrating the
robustness of BIT plant uptake. Following plant health experiments to determine an appropriate
inhibitor concentration, numerous inhibitors (listed here in parentheses) were used at a single

concentration to test pathways known to be important to the plant uptake and metabolism of a variety of compounds. These inhibitors yielded no significant impact on BIT plant uptake (p>0.05): peptide and amino acid uptake (diethyl pyrocarbonate and 2,4-dinitrophenol), sucrose uptake (diethyl pyrocarbonate), aquaporins (glycerol), organic cation transporters (quinidine), anion channels (anthracene-9-carboxylic acid), and cytochrome p450 (1-aminobenzotriazole). Therefore, these pathways do not appear relevant to BIT uptake—at the inhibitor concentrations used in this work. Further testing at higher inhibitor concentrations may reveal that these pathways are involved in BIT uptake; however, these results most clearly suggest that a different pathway or pathways not tested with these inhibitors is/are the mechanism of Arabidopsis BIT uptake. 3.5 Possible plant metabolites observed to increase in response to BIT exposure Following 24 hours of BIT exposure (with α =0.05), there were 546 HRMS features that increased in BIT-exposed plant tissue, and 453 decreased (shown visually in Figure S2). Of the increased compounds, we observed five metabolites (for some metabolites, one metabolite was represented by more than one HRMS feature) upregulated five-fold or greater. Two of the increased metabolites were also present in the plant medium at 24 hours (vs. the unexposed plant medium at t=0), implicating possible excretion of plant metabolites. Two of the proposed upregulated metabolites are endogenous plant compounds rather than transformation products or conjugates of BIT. Nicotinic acid was upregulated by 6-fold in the plant tissue and was initially discovered with a MS2 spectral library match (reported as an mzCloud Best Match score of 86.9 by Compound Discoverer). We subsequently used a commercially-available reference standard and performed a standard addition to confirm the

Muerdter et al.

Page 25 of 43

1

2	
3 4	50
5 6	51
7 8	51
9 10	51
11 12	51
13 14	51
15 16	51
17 18	51
19 20	51
21 22	51
23 24	51
25 26 27	51
27 28 29	52
30 31	52
32 33	52
34 35	52
36 37	52
38 39	52
40 41	52
42 43	52
44 45	52
46 47	52
48 49	52
50 51	53
52 53	52
54 55 56	55
57	
58 59	
60	

)9 compound identify to the highest confidence (Level 1 Confidence⁷⁶). Nicotinic acid is active in 0 numerous normal plant metabolic processes,⁷⁷ including the formation of NADPH.⁷⁸ Thus, the 1 upregulated nicotinic acid in BIT-exposed plants may be due to the increased NADPH needed for glutathione conjugation.^{79–81} cytochrome P450.⁸² or other processes potentially related to 2 3 detoxification metabolism. The other endogenous plant compound significantly upregulated (11-4 fold in plant tissue, 473-fold in hydronic medium) we propose as phenylthioacetohydroximic 5 acid (mass deviation of 3 ppm; 3a confidence⁶⁷). Phenylthioacetohydroximic acid is part of the 6 glucotropaeolin synthesis pathway. Glucotropaeolin is a glucosinolate, which are activated as part of the plant defense system.⁸³ and thus may increase in response to xenobiotics. 7 8 Additionally, the glucotropaeolin pathway requires UDP-glucose, which is known to be used for the glycosylation of xenobiotics.⁸⁴ It is possible that a BIT-glucose conjugate was formed but not 9 20 detected, thus occupying much of the UDP-glucose pool in the plants and causing the 1 accumulation of phenylthioacetohydroximic acid. It is environmentally significant that one or 22 more endogenous plant compounds is proven to be upregulated following exposure to BIT 23 because this demonstrates that xenobiotic compounds found in stormwater can impact 24 fundamental plant biology.

We propose a BIT-amino acid conjugate (BIT-alanine-tyrosine) as a possible structure for another of the compounds upregulated in BIT-exposed plant tissue vs unexposed plant tissue (at a seven-fold increase). The mass deviation of this compound is 8 ppm, with a level 3b confidence⁶⁷ based on MS fragments (no MS2 data) and experimental data. Structures in Table 1 are shown in their unionized state for consistency; however, we would expect (based on the accurate mass measured) that the compound is likely deprotonated in the at ambient conditions (pKa= 4.0)⁸⁵ and tyrosine is known to ionize in either ESI+ or ESI- modes.⁸⁵ Amino acid conjugation is becoming an increasingly well-documented mechanism of xenobiotic metabolism by plants.^{25,68,86} Specifically, conjugation with alanine and tyrosine was also previously reported for Arabidopsis metabolism of di-*n*-butyl phthalate, where the amino acids replaced a hydroxyl group on the parent molecule.⁸⁷ Thus, a similar pathway may occur for BIT. The full metabolic implications of amino acid conjugation of xenobiotics are poorly understood.⁸⁸ Despite the xenobiotic benzotriazole forming tryptophan analogues via amino acid conjugation, benzotriazole was not integrated into plant proteins in Arabidopsis;³⁸ however, other metabolic influences may occur. Lastly, two compounds, TP410 and TP470, both with a plant tissue fold change of 8, could not be structurally resolved at acceptable mass deviations (i.e., 10 ppm or less) and are thus reported here merely as a Level 5 accurate masses of interest. TP470 contains fragments that suggest possible glutathione conjugation (see SI spectra), a well-known plant detoxification process for many xenobiotics.^{30,79,89} Nevertheless, the high mass deviation (199 ppm between TP470 and a BIT glutathione conjugate; see SI) prevents tentative identification. The retention time for TP470 (1.9 min) was consistent and distinct enough to be detected repeatedly in the nine replicates, eluting just after many other compounds early in the sample run (Figure S9). Future work is needed to more conclusively identify these unknown metabolites.

549 Two of the upregulated plant metabolites in the plant tissue, phenylthioacetohydroximic 550 acid and TP470, were also present in the hydroponic medium after 24 hours of BIT exposure 551 (fold changes >1, Table 1), possibly due to plant excretion. There was no significant decrease in 552 the concentration of BIT in the abiotic control medium, demonstrating that minimal if any 553 abiotic BIT transformation occurred. The proposed alanine-tyrosine BIT conjugate was not 554 present as a major metabolite in the medium. This corresponds with previous work in which a

Muerdter et al.

benzotriazole-amino acid conjugate was not excreted into the medium.²⁵ The formation of

compounds,²⁵ but further work would be needed to definitively prove a similar lack of

interaction between plant exudates and BIT.

as well as a lack of any substantial product-to-parent reversion.³³

metabolites through interaction of BIT with exudates in the medium cannot be fully ruled out.

Previous work with benzotriazole did not find that interactions with plant exudates created new

hours, although a small amount was present in BIT-exposed plant tissue (fold change of 1.8 vs

unexposed tissue), and in the medium (1% of the BIT at t=0). This observation indicates BIT

plant uptake occurred followed by near-complete metabolism, consistent with prior literature,⁴⁰

The BIT parent compound was not a major compound in the plant tissue or medium at 24

2	
3	555
4	
5 6	556
7	
8	557
9 10	550
10	558
12	550
13	559
14 15	560
16	
17	561
18	
20	562
21	
22	563
23 24	564
25	504
26	
27	
20 29	
30	
31	
32	
34	
35	
36	
37 38	
39	
40	
41 42	
42 43	
44	
45	
46 47	
48	
49	
50 51	
52	
53	
54	
55 56	
57	
58	
59	

Table 1: Summary of metabolites whose production increased ≥ 5 times more in BIT-exposed Arabidopsis plant tissue vs unexposed plant tissue. BIT itself is also provided for567reference. The BIT structure is shown in blue, and other structures in black. We employed the Schymanski framework of communication of confidence in novel product568discovery.⁷⁶ Accurate and exact metabolite masses, ionization mode, mass spectra, and fragment information are provided in the Supplementary Information (Table S6 and Mass569Spectra section).

Compound Name	Proposed Structure (shown unionized)	Proposed Formula	Confidence Level ⁷⁶	Plant Tissue Fold Change Peak Area in BIT Treatment vs Unexposed Plant Tissue	Medium Fold Change: Peak Area in BIT Treatment at 24 hr vs Unexposed Medium	Retention Time (min)	Ionization mode, Measured m/z [Accurate Mass Deviation (ppm) from proposed ionized formula]
BIT (parent compound, for reference)	NH	C ₇ H ₅ NOS	Level 1: Confirmed with standard	1.8	0.01	14.59	[M+H] ⁺ 151.00928 (3)
Nicotinic Acid	OH CN N	C ₆ H ₅ NO ₂	Level 1: Confirmed with reference standard (additionally, Library Spectrum Match, MS ²)	6	0.4	12.91	[M-H] ⁻ 122.02349 (10)
Phenylthioacetohydroximic acid	N OH SH	C ₈ H9NOS	Level 3a: ⁶⁷ Tentative Candidate (based on MS, fragments, exp. data, MS ²)	11	472	14.77	[M+H] ⁺ 168.04785 (3)

Compound Name	Proposed Structure (shown unionized)	Proposed Formula	Confidence Level ⁷⁶	Plant Tissue Fold Change Peak Area in BIT Treatment vs Unexposed Plant Tissue	Medium Fold Change: Peak Area in BIT Treatment at 24 hr vs Unexposed Medium	Retention Time (min)	Ionization mode, Measured m/z [Accurate Mass Deviation (ppm) from proposed ionized formula]
BIT-Alanine-Tyrosine Conjugate		C ₁₉ H ₂₁ N ₃ O ₅ S	Level 3b: ⁶⁷ Tentative Candidate (based on MS fragments, exp. data)	7	1	13.89	[M+H] ⁺ 403.11679 (8)
TP 470 [Unknown Accurate Mass of Interest Significant Upregulated]	N/A	N/A	Level 5: Accurate Mass of Interest	8	50	1.92	[M-H] ⁻ 470.15134 (N/A)
TP 410 [Unknown Accurate Mass of Interest Significant Upregulated]	N/A	N/A	Level 5: Accurate Mass of Interest	8	0.8	6.73	[M-H] ⁻ 410.86249 (N/A)

Muerdter et al.

4. Conclusions

This work demonstrates that four commonly used isothiazolinones are rapidly removed from plant medium by aseptic Arabidopsis grown hydroponically. This removal is likely primarily due to plant uptake, with no significant difference in removal rate among the compounds. Rapid, likely transporter-mediated removal of isothiazolinones from water is a novel and promising finding for removing isothiazolinones from stormwater even in situations where there may be a fairly short amount of interaction time, e.g., in bioretention cells designed for infiltration of stormwater. In the initial BIT concentration ranges tested in this paper, rapid (≥99% removal from the medium within 24 hours) rates of uptake were found at all concentrations with a pattern suggesting substrate inhibition. Repeated BIT exposures implicated constituently active uptake. These data together suggest the potential for high uptake rates at environmentally relevant concentrations. Further work would be needed to demonstrate the plant uptake kinetics under field conditions. Additionally, isothiazolinone plant uptake was robust even in mixtures with other similar

compounds as potential competitive inhibitors and with the addition of known inhibitors of plant uptake pathways and metabolism. This discovery lends further support to our findings being environmentally relevant in the compound mixtures that occur in stormwater green infrastructure systems. The rapid plant uptake of isothiazolinones coupled with the plant metabolism proposed in this work present the possibility for phytoremediation of BIT and possibly other isothiazolinones. The potential BIT conjugation and upregulation of endogenous plant compounds indicate plant metabolism of BIT and potential to impact plants. Further work is needed to verify the identity of the metabolites identified as accurate masses of interest. Overall, the ability of Arabidopsis to rapidly take up and metabolize isothiazolinone compounds without

Muerdter et al.

2		
3 4	595	visual impacts to plant health indicates the potential for phytoremediation of these compounds
5 6 7	596	from stormwater.
8 9 10	597	
11 12 13	598	Conflicts of Interest
14 15	599	The authors declare no competing financial interest.
16 17 18 19	600	
20 21	601	Acknowledgements
22 23 24	602	C.P.M. was supported by the Iowa Space Grant Consortium under NASA Award No.
25 26 27	603	NNX16AL88H, the University of Iowa Graduate College Post-Comprehensive Fellowship, the
27 28 29	604	University of Iowa Graduate College Summer Fellowship, and the University of Iowa Neil B.
30 31	605	Fisher Environmental Engineering Fellowship.
32 33	606	
34 35 36	607	M.M.P. was supported by the Iowa Biosciences Academy.
37 38	608	
39 40 41	609	This work was supported by the NSF CBET CAREER (1844720), NSF Major Research
41 42 43	610	Instrumentation grant CHE-1919422 for the metabolites work, the University of Iowa Center for
44 45	611	Global & Regional Environmental Research, the University of Iowa Environmental Health
46 47	612	Sciences Research Center (NIH P30 ES005605), and USDA NIFA (2021-67019-33680).
48 49 50	613	The authors thank Reid Simmer for the use of his grow tent, and Lynn Teesch and Vic Parcell of
50 51 52	614	the University of Iowa High Resolution Mass Spectrometry Facility.
53 54	615	
55 56 57 58	616	
59 60		Muerdter et al. 31

2 3	617	Refer	ences	
4 5	017	iterer		
6 7	618	(1)	Hu, K.; Li, HR.; Ou, RJ.; Li, CZ.; Yang, XL. Tissue accumulation and toxicity of	
/ 8 9	619		isothiazolinone in Ctenopharyngodon idellus (grass carp): Association with P-	
) 10 11	620		glycoprotein expression and location within tissues. Environ. Toxicol. Pharmacol. 2014,	
12 13	621		37 (2), 529–535.	
14 15	622	(2)	Elshimy, N ; Thompson, D. The changing faces of contact allergy to methyl	
16 17 18	623		isothiazolinone in children: an interesting case collection. In British Journal Of	
19 20	624		Dermatology; 2016; Vol. 175, p E153.	
21 22	625	(3)	Alvarez-Rivera, G.; Dagnac, T.; Lores, M.; Garcia-Jares, C.; Sanchez-Prado, L.; Lamas, J	•
23 24 25	626		P.; Llompart, M. Determination of isothiazolinone preservatives in cosmetics and	
26 27	627		household products by matrix solid-phase dispersion followed by high-performance liquid	ł
28 29	628		chromatography-tandem mass spectrometry. J. Chromatogr. A 2012, 1270, 41-50.	
30 31 22	629	(4)	Schoknecht, U.; Gruycheva, J.; Mathies, H.; Bergmann, H.; Burkhardt, M. Leaching of	
32 33 34	630		Biocides Used in Façade Coatings under Laboratory Test Conditions. Environ. Sci.	
35 36	631		<i>Technol.</i> 2009 , <i>43</i> (24), 9321–9328.	
37 38 20	632	(5)	Paijens, C.; Bressy, A.; Frère, B.; Moilleron, R. Biocide emissions from building material	s
39 40 41	633		during wet weather: identification of substances, mechanism of release and transfer to the	
42 43	634		aquatic environment. Environ. Sci. Pollut. Res. 2020, 27 (4), 3768-3791.	
44 45	635	(6)	Bollmann, U. E.; Vollertsen, J.; Carmeliet, J.; Bester, K. Dynamics of biocide emissions	
40 47 48	636		from buildings in a suburban stormwater catchment - Concentrations, mass loads and	
49 50	637		emission processes. Water Res. 2014, 56, 66–76.	
51 52	638	(7)	Paijens, C.; Frère, B.; Caupos, E.; Moilleron, R.; Bressy, A. Determination of 18 Biocides	\$
53 54 55 56	639		in Both the Dissolved and Particulate Fractions of Urban and Surface Waters by HPLC-	
58				
59		Muerd	iter et al. 3	2

2 3 4	640		MS/MS. Water, Air, Soil Pollut. 2020, 231 (5), 210.	
5 6	641	(8)	1,2-benzisothiazol-3(2H)-one - Substance Information - ECHA	
7 8	642		https://echa.europa.eu/substance-information/-/substanceinfo/100.018.292 (accessed Ma	ar
9 10 11	643		23, 2021).	
12 13	644	(9)	octhilinone (ISO); 2-octyl-2H-isothiazol-3-one; Substance Information - ECHA	
14 15	645		https://echa.europa.eu/substance-information/-/substanceinfo/100.043.404 (accessed Ap	or
16 17 18	646		2, 2021).	
19 20	647	(10)	2-methylisothiazol-3(2H)-one - Substance Information - ECHA	
21 22	648		https://echa.europa.eu/substance-information/-/substanceinfo/100.018.399 (accessed Ap	r
23 24 25	649		2, 2021).	
26 27	650	(11)	5-Chloro-2-methyl-2H-isothiazol-3-one Substance Information - ECHA	
28 29	651		https://echa.europa.eu/substance-information/-/substanceinfo/100.043.167 (accessed Ap	or
30 31 32	652		2, 2021).	
33 34	653	(12)	Bollmann, U. E.; Fernández-Calviño, D.; Brandt, K. K.; Storgaard, M. S.; Sanderson, H	.;
35 36	654		Bester, K. Biocide Runoff from Building Facades: Degradation Kinetics in Soil. Enviro	n.
37 38 39	655		Sci. Technol. 2017, 51 (7), 3694–3702.	
40 41	656	(13)	Vega-Garcia, P.; Lok, C.; Marhoon, A.; Schwerd, R.; Johann, S.; Helmreich, B. Modell	ing
42 43	657		the environmental fate and behavior of biocides used in façades covered with mortars an	nd
44 45 46	658		plasters and their transformation products. Build. Environ. 2022, 108991.	
40 47 48	659	(14)	Hensen, B.; Lange, J.; Jackisch, N.; Zieger, F.; Olsson, O.; Kümmerer, K. Entry of	
49 50	660		biocides and their transformation products into groundwater via urban stormwater	
51 52	661		infiltration systems. Water Res. 2018, 144, 413-423.	
53 54 55 56 57	662	(15)	Singer, H.; Jaus, S.; Hanke, I.; Lück, A.; Hollender, J.; Alder, A. C. Determination of	
58 59		Muero	dter et al.	33

3 4	663
5 6	664
7 8	665
9 10 11	666
12 13	667
14 15	668
16 17	669
18 19 20	670
21 22	671
23 24 25	672
25 26 27	673
28 29	674
30 31	675
32 33 34	676
35 36	677
37 38	678
39 40 41	679
41 42 43	680
44 45	681
46 47 48	682
49 50	683
51 52	684
53 54 55	685
56 57	
58 59	
60	

biocides and pesticides by on-line solid phase extraction coupled with mass spectrometry
and their behaviour in wastewater and surface water. *Environ. Pollut.* 2010, *158* (10),
3054–3064.

- 666 (16) Chen, Z.-F.; Ying, G.-G.; Lai, H.-J.; Chen, F.; Su, H.-C.; Liu, Y.-S.; Peng, F.-Q.; Zhao, J.-
- 667 L. Determination of biocides in different environmental matrices by use of ultra-high-
- 668 performance liquid chromatography–tandem mass spectrometry. *Anal. Bioanal. Chem.*
- **2012**, *404* (10), 3175–3188.
 - 670 (17) Rafoth, A.; Gabriel, S.; Sacher, F.; Brauch, H.-J. Analysis of isothiazolinones in
- 671 environmental waters by gas chromatography–mass spectrometry. J. Chromatogr. A 2007,
 672 1164 (1), 74–81.
- 673 (18) Wu, X.; Dodgen, L. K.; Conkle, J. L.; Gan, J. Plant uptake of pharmaceutical and personal
 674 care products from recycled water and biosolids: A review. *Science of the Total*
- 675 *Environment*. Elsevier December 1, 2015, pp 655–666.
- 676 (19) Phytoremediation: Transformation and Control of Contaminants; McCutcheon, S. C.,
- 677 Schnoor, J. L., Eds.; Wiley-Interscience: Hoboken, 2003.
- 678 (20) Schnoor, J. L.; Licht, L. A.; McCutcheon, S. C.; Wolfe, N. L.; Carreira, L. H.
- Phytoremediation of Organic and Nutrient Contaminants. *Environ. Sci. Technol.* 1995, 29
 680 (7), 318A-323A.
- 681 (21) Nason, S. L.; Miller, E. L.; Karthikeyan, K. G.; Pedersen, J. A. Effects of Binary Mixtures
 682 and Transpiration on Accumulation of Pharmaceuticals by Spinach. *Environ. Sci. Technol.* 683 2019, 53 (9), 4850–4859.
 - 684 (22) Trapp, S. Modelling uptake into roots and subsequent translocation of neutral and
 685 ionisable organic compounds. *Pest Manag. Sci.* 2000, *56* (9), 767–778.

Muerdter et al.

1 2				
3 4	686	(23)	Jones, D. L.; Darrah, P. R. Influx and efflux of amino acids from Zea mays L. roots and	
5 6 7	687		their implications for N nutrition and the rhizosphere. Plant Soil 1993, 155-156 (1), 87-	
/ 8 0	688		90.	
10 11	689	(24)	Jones, D. L.; Darrah, P. R. Amino-acid influx at the soil-root interface of Zea mays L. and	nd
12 13	690		its implications in the rhizosphere. Plant Soil 1994, 163 (1), 1-12.	
14 15	691	(25)	LeFevre, G. H.; Müller, C. E.; Li, R. J.; Luthy, R. G.; Sattely, E. S. Rapid	
16 17 18	692		phytotransformation of benzotriazole generates synthetic tryptophan and auxin analogs i	n
19 20	693		Arabidopsis. Environ. Sci. Technol. 2015, 49 (18), 10959–10968.	
21 22	694	(26)	LeFevre, G. H.; Portmann, A. C.; Müller, C. E.; Sattely, E. S.; Luthy, R. G. Plant	
23 24 25	695		assimilation kinetics and metabolism of 2-mercaptobenzothiazole tire rubber vulcanizers	3
23 26 27	696		by Arabidopsis. Environ. Sci. Technol. 2016, 50 (13), 6762-6771.	
28 29	697	(27)	Burken, J. G. Uptake and Metabolism of Organic Compounds: Green-Liver Model. In	
30 31 22	698		Phytoremediation: Transformation and Control of Contaminants; McCutcheon, S. C.,	
32 33 34	699		Schnoor, J. L., Eds.; John Wiley & Sons, Inc.: Hoboken, 2003; pp 59-84.	
35 36	700	(28)	Sandermann, H. J. Higher plant metabolism of xenobiotics: the 'green liver' concept.	
37 38	701		<i>Pharmacogenetics</i> 1994 , <i>4</i> (5), 225–241.	
39 40 41	702	(29)	Fu, Q.; Zhang, J.; Borchardt, D.; Schlenk, D.; Gan, J. Direct Conjugation of Emerging	
42 43	703		Contaminants in Arabidopsis : Indication for an Overlooked Risk in Plants? Environ. Sc.	i.
44 45	704		<i>Technol.</i> 2017 , <i>51</i> (11), 6071–6081.	
46 47 49	705	(30)	Coleman, J.; Blake-Kalff, M.; Davies, E. Detoxification of xenobiotics by plants:	
40 49 50	706		chemical modification and vacuolar compartmentation. Trends Plant Sci. 1997, 2 (4),	
51 52	707		144–151.	
53 54	708	(31)	Schröder, P. Exploiting Plant Metabolism for the Phytoremediation of Organic	
55 56 57				
58 59		Muero	lter et al	35

2 3	709		Xenobiotics. In Phytoremediation: Methods and Reviews; Humana Press, 2007; pp 251	l—
4 5 6	710		263.	
7 8	711	(32)	Zhang, Q.; Kong, W.; Wei, L.; Hou, X.; Ma, Q.; Liu, Y.; Luo, Y.; Liao, C.; Liu, J.;	
9 10 11	712		Schnoor, J. L.; et al. Compartmentalization and Excretion of 2,4,6-Tribromophenol	
12 13	713		Sulfation and Glycosylation Conjugates in Rice Plants. 2021.	
14 15	714	(33)	Fu, Q.; Liao, C.; Du, X.; Schlenk, D.; Gan, J. Back Conversion from Product to Parent:	:
16 17 18	715		Methyl Triclosan to Triclosan in Plants. Environ. Sci. Technol. Lett. 2018, 5 (3), 181-1	85.
19 20	716	(34)	Hou, X.; Yu, M.; Liu, A.; Li, Y.; Ruan, T.; Liu, J.; Schnoor, J. L.; Jiang, G.	
21 22	717		Biotransformation of tetrabromobisphenol A dimethyl ether back to tetrabromobisphen	ıol
23 24 25	718		A in whole pumpkin plants. Environ. Pollut. 2018, 241, 331–338.	
26 27	719	(35)	Huynh, K.; Reinhold, D. Metabolism of Sulfamethoxazole by the Model Plant	
28 29	720		Arabidopsis thaliana. Environ. Sci. Technol. 2019, 53 (9), 4901–4911.	
30 31 32	721	(36)	Schröder, P.; Scheer, C. E.; Diekmann, F.; Stampfl, A. How Plants Cope with Foreign	
33 34	722		Compounds. Translocation of xenobiotic glutathione conjugates in roots of barley	
35 36	723		(Hordeum vulgare). Environ. Sci. Pollut. Res. 2007, 14 (2), 114–122.	
37 38 20	724	(37)	Hou, X.; Wei, L.; Tang, Y.; Kong, W.; Liu, J.; Schnoor, J. L.; Jiang, G. Two Typical	
39 40 41	725		Glycosylated Metabolites of Tetrabromobisphenol A Formed in Plants: Excretion and	
42 43	726		Deglycosylation in Plant Root Zones. Environ. Sci. Technol. Lett. 2021.	
44 45	727	(38)	LeFevre, G. H.; Lipsky, A.; Hyland, K. C.; Blaine, A. C.; Higgins, C. P.; Luthy, R. G.	
46 47 48	728		Benzotriazole (BT) and BT plant metabolites in crops irrigated with recycled water.	
49 50	729		Environ. Sci. Water Res. Technol. 2017, 3 (2), 213–223.	
51 52	730	(39)	Krzeminski, S. F.; Brackett, C. K.; Fisher, J. D. Fate of microbicidal 3-isothiazolone	
53 54 55	731		compounds in the environment. Modes and rates of dissipation. J. Agric. Food Chem.	
56 57				
58 59 60		Muero	lter et al.	36

1 2			
3 4	732		1975 , <i>23</i> (6), 1060–1068.
5 6	733	(40)	Krzeminski, S. F.; Brackett, C. K.; Fisher, J. D.; Spinnler, J. F. Fate of microbicidal 3-
7 8 0	734		isothiazolone compounds in the environment. Products of degradation. J. Agric. Food
10 11	735		<i>Chem.</i> 1975 , <i>23</i> (6), 1068–1075.
12 13	736	(41)	Muerdter, C. P.; Wong, C. K.; LeFevre, G. H. Emerging investigator series: The role of
14 15	737		vegetation in bioretention for stormwater treatment in the built environment: Pollutant
16 17 18	738		removal, hydrologic function, and ancillary benefits. Environ. Sci. Water Res. Technol.
19 20	739		2018 , <i>4</i> (5), 592–612.
21 22	740	(42)	Speksnijder, P.; van Ravestijn, J.; de Voogt, P. Trace analysis of isothiazolinones in water
23 24 25	741		samples by large-volume direct injection liquid chromatography tandem mass
26 27	742		spectrometry. J. Chromatogr. A 2010, 1217 (32), 5184–5189.
28 29	743	(43)	Bollmann, U. E.; Tang, C.; Eriksson, E.; Jönsson, K.; Vollertsen, J.; Bester, K. Biocides in
30 31 32	744		urban wastewater treatment plant influent at dry and wet weather: Concentrations, mass
32 33 34	745		flows and possible sources. Water Res. 2014, 60, 64–74.
35 36	746	(44)	Bagheri, M.; Al-jabery, K.; Wunsch, D. C.; Burken, J. G. A deeper look at plant uptake of
37 38	747		environmental contaminants using intelligent approaches. Sci. Total Environ. 2019, 651,
39 40 41	748		561–569.
42 43	749	(45)	Limmer, M. A.; Burken, J. G. Plant Translocation of Organic Compounds: Molecular and
44 45	750		Physicochemical Predictors. Environ. Sci. Technol. Lett. 2014, 1 (2), 156-161.
46 47 48	751	(46)	Müller, C. E.; Lefevre, G. H.; Timofte, A. E.; Hussain, F. A.; Sattely, E. S.; Luthy, R. G.
49 50	752		Competing mechanisms for perfluoroalkyl acid accumulation in plants revealed using an
51 52	753		Arabidopsis model system. Environ. Toxicol. Chem. 2016, 35 (5), 1138–1147.
53 54	754	(47)	Muerdter, C. P.; LeFevre, G. H. Synergistic Lemna Duckweed and Microbial
55 56 57			
58 59		Muer	dter et al

60

2				
3 4	755		Transformation of Imidacloprid and Thiacloprid Neonicotinoids. Environ. Sci. Techno	l.
5 6	756		<i>Lett.</i> 2019 , <i>6</i> (12), 761–767.	
7 8	757	(48)	Das, U. K.; Bordoloi, R.; Ganguly, S. Freeze-drying technique and its wide application	1 in
9 10 11	758		biomedical and pharmaceutical sciences. Res. J. Chem. Environ. Sci 2014, 2 (3), 1-4.	
12 13	759	(49)	Mustafa, A.; Imran, M.; Ashraf, M.; Mahmood, K. Perspectives of Using L-Tryptopha	n
14 15	760		for Improving Productivity of Agricultural Crops: A Review. Pedosphere. Soil Science	e
16 17	761		Society of China February 1, 2018, pp 16–34.	
19 20	762	(50)	D'Acunha, B.; Johnson, M. S. Water quality and greenhouse gas fluxes for stormwater	r
21 22	763		detained in a constructed wetland. J. Environ. Manage. 2019, 231, 1232-1240.	
23 24	764	(51)	Yuan, D.; An, Y.; Wang, J.; Chu, S.; Lim, B.; Chen, B.; Xiong, Y.; Kou, Y.; Li, J.	
25 26 27	765		Dissolved organic matter characteristics of urban stormwater runoff from different	
28 29	766		functional regions during grassy swale treatment. 2019.	
30 31	767	(52)	Li, Z. C.; Bush, D. R. ΔpH -dependent amino acid transport into plasma membrane	
32 33 34	768		vesicles isolated from sugar beet leaves. I. Evidence for carrier-mediated, electrogenic	
35 36	769		flux through multiple transport systems. Plant Physiol. 1990, 94 (1), 268-277.	
37 38	770	(53)	Jamai, A.; Chollet, J. F.; Delrot, S. Proton-peptide co-transport in broad bean leaf tissu	les.
39 40 41	771		<i>Plant Physiol.</i> 1994 , <i>106</i> (3), 1023–1031.	
42 43	772	(54)	Bush, D. R. Proton-Coupled Sucrose Transport in Plasmalemma Vesicles Isolated from	n
44 45	773		Sugar Beet (Beta vulgaris L. cv Great Western) Leaves. Plant Physiol. 1989, 89 (4),	
46 47	774		1318–1323.	
48 49 50	775	(55)	Wen, B.; Li, L.; Liu, Y.; Zhang, H.; Hu, X.; Shan, X. quan; Zhang, S. Mechanistic stud	dies
51 52	776		of perfluorooctane sulfonate, perfluorooctanoic acid uptake by maize (Zea mays L. cv.	
53 54	777		TY2). Plant Soil 2013, 370 (1-2), 345-354.	
55 56 57				
58 59		Muer	dter et al	35

1 2 3				
3 4 5	778	(56)	Zhou, J.; Yang, Z.; Liu, Q.; Liu, Y.; Liu, M.; Wang, T.; Zhu, L. Insights into Uptake,	
5 6 7	779		Translocation, and Transformation Mechanisms of Perfluorophosphinates and	
7 8 9	780		Perfluorophosphonates in Wheat (Triticum aestivum L.). Environ. Sci. Technol. 2019.	
10 11	781	(57)	Cui, H.; Hense, B. A.; Müller, J.; Schröder, P. Short term uptake and transport process for	r
12 13	782		metformin in roots of Phragmites australis and Typha latifolia. Chemosphere 2015, 134,	
14 15	783		307–312.	
10 17 18	784	(58)	Tyerman, S. D. Anion Channels in Plants. Annu. Rev. Plant Physiol. Plant Mol. Biol.	
19 20	785		1992 , <i>43</i> (1), 351–373.	
21 22	786	(59)	Shiina, T.; Tazawa, M. Membrane Biology Ca2+-Activated CI-Channel in Plasmalemma of)f
23 24 25	787		Nitellopsis obtusa; 1987; Vol. 99.	
26 27	788	(60)	Tyerman, S. D.; Findlay, G. P.; Paterson, G. J. Inward membrane current in Chara inflata	:
28 29 30 31	789		II. Effects of pH, Clchannel blockers and NH4+, and significance for the hyperpolarized	ł
	790		state. J. Membr. Biol. 1986, 89 (2), 153-161.	
32 33 34	791	(61)	Thomine, S.; Lelièvre, F.; Boufflet, M.; Guern, J.; Barbier-Brygoo, H. Anion-channel	
35 36	792		blockers interfere with auxin responses in dark-grown arabidopsis hypocotyls. Plant	
37 38	793		<i>Physiol.</i> 1997 , <i>115</i> (2), 533–542.	
39 40 41	794	(62)	Keller, C. P.; Van Volkenburgh, E. The electrical response of Avena coleoptile cortex to	
42 43	795		auxins Evidence in vivo for activation of a Cl- conductance. Planta 1996, 198 (3), 404-	
44 45	796		412.	
46 47 48	797	(63)	Rentsch, D.; Laloi, M.; Rouhara, I.; Schmelzer, E.; Delrot, S.; Frommer, W. B. NTR1	
49 50	798		encodes a high affinity oligopeptide transporter in Arabidopsis. FEBS Lett. 1995, 370 (3),	,
51 52	799		264–268.	
53 54 55 56 57	800	(64)	Robinson, S. P.; Beevers, H. Amino Acid Transport in Germinating Castor Bean	
58 59 60		Muero	dter et al. 3	9

1 2			
3 4 5 6 7 8 9 10 11 12 13	801		Seedlings; 1981; Vol. 68.
	802	(65)	Higgins, C. F.; Payne, J. W. Characterization of active dipeptide transport by germinating
	803		barley embryos: Effects of pH and metabolic inhibitors. Planta 1977, 136 (1), 71-76.
	804	(66)	Chen, Z.; Wang, J.; Chen, H.; Wen, Y.; Liu, W. Enantioselective Phytotoxicity of
	805		Dichlorprop to Arabidopsis thaliana: The Effect of Cytochrome P450 Enzymes and the
14 15 16	806		Role of Fe. Environ. Sci. Technol. 2017, 51 (20), 12007–12015.
17 18	807	(67)	Wiener, E. A.; LeFevre, G. H. White Rot Fungi Produce Novel Tire Wear Compound
19 20	808		Metabolites and Reveal Underappreciated Amino Acid Conjugation Pathways. Environ.
21 22	809		Sci. Technol. Lett. 2022.
23 24 25	810	(68)	Miller, E. L.; Nason, S. L.; Karthikeyan, K. G.; Pedersen, J. A. Root Uptake of
26 27	811		Pharmaceuticals and Personal Care Product Ingredients. Environ. Sci. Technol. 2016, 50
28 29	812		(2), 525–541.
30 31 32	813	(69)	Kubeš, M.; Yang, H.; Richter, G. L.; Cheng, Y.; Młodzińska, E.; Wang, X.; Blakeslee, J.
33 34	814		J.; Carraro, N.; Petrášek, J.; Zažímalová, E.; et al. The Arabidopsis concentration-
35 36	815		dependent influx/efflux transporter ABCB4 regulates cellular auxin levels in the root
37 38	816		epidermis. Plant J. 2012, 69 (4), 640-654.
39 40 41	817	(70)	Hart, J. J.; DiTomaso, J. M.; Linscott, D. L.; Kochian, L. V. Transport interactions
42 43	818		between paraquat and polyamines in roots of intact maize seedlings. <i>Plant Physiol.</i> 1992,
44 45 46 47 48 49 50 51 52 53 54 55 56 57 58	819		<i>99</i> (4), 1400–1405.
	820	(71)	Macherius, A.; Eggen, T.; Lorenz, W.; Moeder, M.; Ondruschka, J.; Reemtsma, T.
	821		Metabolization of the Bacteriostatic Agent Triclosan in Edible Plants and its
	822		Consequences for Plant Uptake Assessment. Environ. Sci. Technol. 2012, 46 (19), 10797-
	823		10804.

1 2			
3 4 5	824	(72)	Reed, M. C.; Lieb, A.; Nijhout, H. F. The biological significance of substrate inhibition: A
5 6	825		mechanism with diverse functions. BioEssays 2010, 32 (5), 422-429.
 7 8 9 10 11 12 13 14 15 16 17 18 19 20 21 22 	826	(73)	Wolfe, N. L.; Hoehamer, C. F. Enzymes Used by Plants and Microorganisms to Detoxify
	827		Organic Compounds. In Phytoremediation; McCutcheon, S. C., Schnoor, J. L., Eds.; John
	828		Wiley & Sons, Inc., 2004; pp 159–187.
	829	(74)	Bryant, C.; DeLuca, M. Purification and characterization of an oxygen-insensitive
	830		NAD(P)H nitroreductase from Enterobacter cloacae. J. Biol. Chem. 1991, 266 (7), 4119-
	831		4125.
	832	(75)	Sopanen, T.; Väisänen, E. Uptake of Glutamine by the Scutellum of Germinating Barley
23 24	833		Grain. Plant Physiol. 1985, 78 (4), 684-689.
25 26 27	834	(76)	Schymanski, E. L.; Jeon, J.; Gulde, R.; Fenner, K.; Ruff, M.; Singer, H. P.; Hollender, J.
28 29	835		Identifying Small Molecules via High Resolution Mass Spectrometry: Communicating
30 31	836		Confidence. Environ. Sci. Technol. 2014, 48 (4), 2097–2098.
32 33 34	837	(77)	Hussein, M. M.; Faham, S. Y.; Alva, A. K. Role of Foliar Application of Nicotinic Acid
34 35 36	838		and Tryptophan on Onion Plants Response to Salinity Stress. J. Agric. Sci. 2014, 6 (8),
37 38	839		p41.
39 40 41	840	(78)	Ashihara, H.; Yin, Y.; Katahira, R.; Watanabe, S.; Mimura, T.; Sasamoto, H. Comparison
41 42 43 44 45 46 47 48 49 50 51 52 53 54	841		of the formation of nicotinic acid conjugates in leaves of different plant species. Plant
	842		Physiol. Biochem. 2012, 60, 190–195.
	843	(79)	Sun, C.; Dudley, S.; McGinnis, M.; Trumble, J.; Gan, J. Acetaminophen detoxification in
	844		cucumber plants via induction of glutathione S-transferases. Sci. Total Environ. 2019, 649,
	845		431–439.
	846	(80)	Gill, S. S.; Anjum, N. A.; Hasanuzzaman, M.; Gill, R.; Trivedi, D. K.; Ahmad, I.; Pereira,
55 56			
57 58 59		Muor	dter et al

1 2				
2 3 4	847		E.; Tuteja, N. Glutathione and glutathione reductase: A boon in disguise for plant abiot	ic
5 6	848		stress defense operations. Plant Physiol. Biochem. 2013, 70, 204-212.	
7 8	849	(81)	Bartha, B.; Huber, C.; Schröder, P. Uptake and metabolism of diclofenac in Typha	
9 10 11	850		latifoliahow plants cope with human pharmaceutical pollution. Plant Sci. 2014, 227, 1	2–
12 13	851		20.	
14 15	852	(82)	Siminszky, B. Plant cytochrome P450-mediated herbicide metabolism. Phytochem. Rev	
16 17 18	853		2006 , <i>5</i> (2), 445–458.	
19 20	854	(83)	Wielanek, M.; Urbanek, H. Enhanced glucotropaeolin production in hairy root cultures	of
21 22	855		Tropaeolum majus L. by combining elicitation and precursor feeding. Plant Cell. Tissue	9
23 24 25	856		Organ Cult. 2006, 86 (2), 177–186.	
25 26 27	857	(84)	Schröder, P.; Collins, C. Conjugating Enzymes Involved in Xenobiotic Metabolism of	
28 29	858		Organic Xenobiotics in Plants. Int. J. Phytoremediation 2002, 4 (4), 247-265.	
30 31 22	859	(85)	Liigand, P.; Kaupmees, K.; Haav, K.; Liigand, J.; Leito, I.; Girod, M.; Antoine, R.; Kru	ve,
32 33 34	860		A. Think Negative: Finding the Best Electrospray Ionization/MS Mode for Your Analy	te.
35 36	861		Anal. Chem. 2017, 89 (11), 5665–5668.	
37 38	862	(86)	Eyer, L.; Vain, T.; Pařízková, B.; Oklestkova, J.; Barbez, E.; Kozubíková, H.; Pospíšil,	Т.;
39 40 41	863		Wierzbicka, R.; Kleine-Vehn, J.; Fránek, M.; et al. 2,4-D and IAA amino acid conjugat	es
42 43	864		show distinct metabolism in Arabidopsis. PLoS One 2016, 11 (7).	
44 45	865	(87)	Cheng, Z.; Sun, H.; Sidhu, H. S.; Sy, N. D.; Wang, X.; Gan, J. Conjugation of Di-n-but	yl
46 47 48	866		Phthalate Metabolites in Arabidopsis thaliana and Potential Deconjugation in Human	
49 50	867		Microsomes. Environ. Sci. Technol. 2021, 55 (4), 2381–2391.	
51 52	868	(88)	Fu, Q.; Dudley, S.; Sun, C.; Schlenk, D.; Gan, J. Stable Isotope Labeling-Assisted	
53 54 55 56	869		Metabolite Probing for Emerging Contaminants in Plants. Anal. Chem. 2018, 90 (18),	
58 59		Muer	dter et al	42

Muerdter et al.

60

1				
2 3 4 5 6 7 8 9 10 11	870		11040–11047.	
	871	(89)	Huber, C.; Bartha, B.; Harpaintner, R.; Schröder, P. Metabolism of acetaminophen	
	872		(paracetamol) in plants-two independent pathways result in the formation of a glutathic	one
	873		and a glucose conjugate. Environ. Sci. Pollut. Res. 2009, 16 (2), 206-213.	
12 13	874			
14 15 16				
17 18				
19 20				
21 22 23				
24 25				
26 27 28				
28 29 30				
31 32				
33 34 35				
36 37				
38 39				
40 41 42				
43 44				
45 46 47				
48 49				
50 51				
52 53 54				
55 56				
57 58				
59		Muer	dter et al.	43