

Emerging investigator series: Electrochemically-mediated Remediation of GenX Using Redox-Copolymers

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Emerging investigator series: Electrochemically-mediated Remediation of GenX Using Redox-Copolymers

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Per- and polyfluorinated alkyl substances (PFAS) are persistent contaminants that have been continuously detected in groundwater and drinking water around the globe. Hexafluoropropylene oxide dimer acid (tradename GenX) has been used to substitute legacy PFAS, such as PFOA, but its widespread use has caused broad occurrence in water streams at high levels. Here. we evaluate redox-copolymer, poly(4-methacryloyloxy-2,2,6,6-tetramethylpiperidin-1-oxyl-co-4а methacryloyloxy2,2,6,6-tetramethylpiperidine) (PTMA-co-PTMPMA), for the selective electrochemical removal of GenX. The amine functional groups promote the affinity towards anionic PFAS, and the redox-active nitroxide radicals provide electrochemical control for adsorption and desorption. Faster kinetics and higher uptake (>475 mg/g adsorbent) were obtained with the redox-copolymer when applying 0.8 V vs. Ag/AgCl potential compared to open circuit. The copolymer electrosorbents were evaluated over a wide pH range and diverse water matrices, with electrostatic-based mechanisms dependent on the state of protonation of the PFAS. Moreover, we translated the redox-electrodes from a batch to flow-by cell configuration, showing successful adsorption and release of GenX under flow and electrochemical control. Finally, prolonged exposure of GenX at reduction potentials generated smaller PFAS fragments at the redox-electrodes. To fully defluorinate GenX, the copolymer-functionalized electrodes were coupled with a boron-doped diamond (BDD) counter electrode, for integrating separation and defluorination within the same device. The combined system demonstrated close to 100% defluorination efficiency. Thus, we highlight the potential of electroactive redox platforms for the reactive separation of fluorotelomers, and point to future directions for their practical implementation for water treatment.

Water Impact

GenX is a new class of short-chain PFAS that led to increasing environmental concerns due to its ubiquitous occurrence. This study explores the capability of redox-copolymers consisting of TEMPO units to electrochemically capture and remove GenX from water streams. We demonstrate the high capacity of the redox electrodes for GenX adsorption in batch and flow-by cell systems. We also show the enhancement of kinetics under potential and the ability to integrate the electrosorbents with boron-doped diamond (BDD) electrodes during regeneration for GenX defluorination.

1. Introduction

Poly and perfluoroalkyl substances (PFAS) are a large group of anthropogenic organic pollutants that have been extensively used since the 1950s.¹ Their chemical structure is composed of a strong carbon fluoride bond that increases their persistence in the environment, thus promoting the search for effective treatment technologies. Around six thousand PFAS are known, which have been classified by their hydrophilic moieties. The most abundant compounds are perfluoroalkyl acids which possess at least one negatively charged functional group, commonly carboxylic or sulfonic acids. PFAS contains oleophobic and hydrophobic characteristics that are attractive for a range of commercially available products such as firefighting foams, non-stick cookware, food packaging, cosmetics, and many more.²⁻⁵ Consequently, these ubiquitous compounds have been found in soil, landfills, air, and water, and their potential toxicity and bioaccumulation have caused global concern.⁶⁻⁹ Contaminated water is suspected to be the primary route of exposure to humans.¹⁰

Therefore, the U.S. Environmental Protection Agency (EPA) and the European Commission had declared advisory levels for drinking water of 70 ng/L and 0.5 μ g/L, respectively, to reduce the spread of these compounds into consumable water.^{11, 12} PFAS have been detected in aquatic and drinking water at levels that greatly surpass the EPA and European Commission standards.^{8, 13-16} For example, Veneto, Italy, has been one of the places where the abundant concentration of PFAS has been found in drinking water,14 with concentration levels ranging from 319 to as high as 1,475 ng/L.14 In the United States, surface water has been found to contain levels of PFAS up to 2,000 ng/L, and in surface water foams, up to 97,000 ng/L.^{17, 18} Also, in some instances of groundwater, PFAS concentrations have been found at levels as high as 5,200 ng/L.19 Associated with this widespread occurrence, numerous studies have shown their presence in human blood serum and wildlife, which is indicative of the importance of monitoring and implementation of targeted treatment methods.²⁰⁻²²

Most studies of PFAS adsorption and destruction have focused on perfluorooctanoic acid (PFOA) and perfluorooctanesulfonic acid (PFOS). These two compounds were found more broadly and at Paper

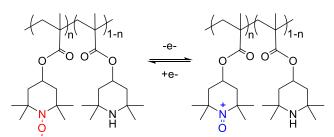


Figure 1: Redox-active PTMA-co-PTMPMA copolymer utilized to increase adsorption uptake via electrochemical input. The nitroxide radical group when oxidized becomes an oxoammonium cation, controlling the electrostatic interactions, while the N-H secondary amine counterpart (PTMPMA) enhances adsorption based on affinity to the amine groups.

much higher concentrations than other PFAS in water,²³ being amongst the original PFAS developed in the early 1940s.¹ However, with increasing stringent regulations for PFOA and PFOS, ¹¹ there has been a rise in the use of short chain PFAS as alternatives. Among these, hexafluoropropylene oxide dimer acid (HFPO-DA, also known as GenX) has seen rapidly increasing use as a direct substitute of PFOA,²⁴ and being utilized in even larger quantities than PFOA, to match the performance of the longer chain precursors.²⁵ While GenX could potentially have lower bioaccumulation properties based on their volatility and C-F distribution, it has increasingly been detected in water, with these short-chain compounds as well as their degradation byproducts persisting in the environment.²⁶

The molecular properties of GenX present several challenges for separation processes and environmental remediation. The shorter aliphatic backbone of GenX, along with the middle ether bond, makes it more hydrophilic when compared with longer chain PFAS. Also, these shorter chain PFAS present higher mobility in the environment, increasing their chances of being found in drinking water and other water streams.^{27, 28} GenX also exhibits high water solubility and a pKa of 2.84, which provides the molecule with a negative charge over a wide pH range.²⁹ Previous studies have targeted GenX for water remediation using conventional treatment techniques such as adsorption, ion exchange, filtration, and reverse

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osmosis.^{30, 31} For adsorption based techniques, short-chain PFAS still suffer from regeneration efficiency limitations,^{30, 32} and limited molecular selectivity. In addition, traditional chemical adsorption methods require the input of additional chemicals and solvents for regeneration, which increases the operating cost of the process as well as chemical footprint.³³

Thus, an electrochemical route for separation and remediation of PFAS can provide an attractive platform that is environmentally friendly, by reducing energy and chemical costs. Recently, electrochemical pathways have been explored for the efficient removal and remediation of emerging contaminants in water by molecularly-tailored leveraging interfaces for selective electrosorption and destruction.³⁴⁻³⁸ In particular, our previous study has leveraged redox-copolymers for the selective binding and integrated defluorination of long-chain PFAS.³⁶ It revealed that three main interactions - hydrophobicity, amine interactions, and electrostatic interactions - synergistically capture/release legacy PFAS such as PFOA and PFOS without the use of chemical regenerant.

However, redox electrosorbents have not yet been evaluated for shorter chain PFAS such as GenX – which is a critical gap for the performance of these redox systems due to the increased occurrence of these new classes of PFAS. Here, we selected a copolymer that combines 4-methacryloyloxy-2,2,6,6-tetramethylpiperidin-1-oxyl (TMA) 4-methacryloyloxy2,2,6,6-tetramethylpiperidine and (TMPMA) repeating units, explicitly PTMA-co-PTMPMA (Figure 1), proven to be effective in adsorbing long-chain PFAS in our previous study.³⁶ Amine functionalities have been regarded as a superior binding site with affinity for negatively charged PFAS,^{39, 40} and the secondary amine functionality (N-H) of the piperidine unit possesses the high pKa value (pKa=11.28),⁴¹ providing protonated and thus positively charged binding sites toward negatively charged GenX in a wide range of pHs. At the same time, the redox couple of nitroxide (N-O●/N=O+) radical/oxoammonium allows for additional enhancement in the electrostatic interactions with the anionic carboxylate group of GenX and serve as a reversible, electrochemically-mediated switch for adsorption and regeneration.

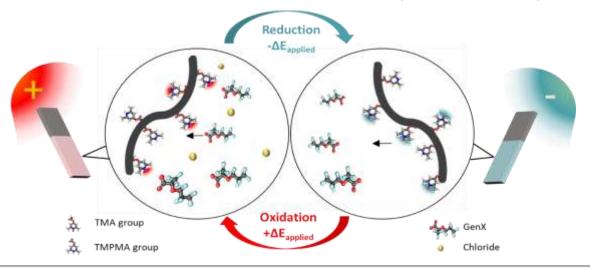


Figure 2: A representative scheme of redox-active electrode of PTMA-co-TMPMA for the electrochemical adsorption and desorption of GenX, in the presence of competing anions. An oxidative potential promotes the selective electrosorption of anionic PFAS, while a reductive potential releases the bound molecules.

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In this work, we investigated redox-electrodes functionalized with PTMA-co-PTMPMA as an electrosorbent for short-chain PFAS, specifically GenX. The redox-electrode system was proven to adsorb GenX selectively when an oxidative potential is applied, and the ability to release bound GenX at a reductive potential (Figure 2). We investigated the balance of electrostatic vs hydrophobic interaction at different pH, and the application of our electrosorbent to a range of environmentally-relevant water matrices. Also, the redox electrodes were translated from a batch analytical scale to a larger flow cell system, to showcase the ability of the redox-electrodes for adsorbing/desorbing GenX through electrochemical modulation for more relevant practical processes. Finally, we examined the reductive degradation of GenX by a combined asymmetric redoxcopolymer//boron-doped diamonds (BDD) configuration, where the BDD counter electrode was shown to be effective for the defluorination of the released GenX.

Materials and Methods

2.1 Materials

Undecafluoro-2-methyl-3-oxahexanoic acid (GenX) (97%) was obtained from Matrix Scientific. All other chemicals were obtained from Sigma Aldrich, VWR, Fisher Scientific, or TCI, and they were used as received. Acetonitrile and ammonium acetate for LC/MS analysis were L.C. grade. Water for LC/MS and ion chromatography (I.C.) analysis was obtained from an ELGA Purelab flex 1 system >18M Ω . The stock solution for each PFAS compound was prepared in DI water at a concentration of 10 mM without supporting electrolytes, and which was diluted to prepare a range of starting solutions for electrochemical tests. These solutions were stored in the fridge (4°C), and the storage time was no longer than six months.

2.2 Analytical methods for PFAS and fluoride detection

A Waters Synapt G2Si with Waters Acquity H-class UPLC LC/MS system was used to detect PFAS concentration in an aqueous solution. All parameters utilized for the LC/MS analysis can be found in the supplementary materials section of this work (Section S1.5). For detecting fluoride content, a Thermo Fisher Dionex 2100 ion chromatography system (I.C.) was utilized. The analytical parameters used have been described elsewhere.³⁶ For every analytical run, a new set of calibration curves were assessed for both LC/MS and (I.C.).

2.3 Electrode preparation and characterization

The polymer-coated electrodes of PTMA-co-PTMPMA/ multi-walled carbon nanotubes (MWCNT) were prepared using a dip-coating technique, which is described in detail in the supplementary information section (Section S1.2). PTMA-co-PTMPMA has been previously synthesized, with a product composed of 51 mol % TMA and 49 mol % TMPMA.³⁶ This copolymer is known for its optimal triple interactions with long-chain PFAS and dispersion with multi-walled carbon nanotubes that serve as an excellent binder to improve the surface area and conductivity. The ink solution ratio was composed of 1:1 PTMA-Co-PTMPMA: MWCNT and was applied onto the stainless-steel mesh electrode via a dip-coating technique and dried with air. The loading mass was maintained at 0.5 mg in an active surface area of 2 cm².

2.4 Electrochemical experiments

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Electrochemical experiments were carried out in a potentiostat (SP-200 Potentiostat, Biologic, and Squidstat Plus, Admiral Instruments). For batch electrosorption and release experiments, a BASi VC-S voltammetry electrochemical cell in a three-electrode configuration was used. Ag/AgCl was used as the reference electrode, and Pt was used as the counter electrode for all adsorption experiments. Batch experiments were performed in 5 mL solutions unless otherwise specified. The typical electrosorption time was 30 min, with a solution containing 20 mM NaCl, unless otherwise indicated. All the adsorption calculations were performed using the total mass of the polymer and multi-walled carbon nanotubes added together. Based on the results obtained in **Figure S12**, MWCNT itself was shown to have some limited adsorption towards GenX.

2.5 Electrochemical Flow cell

For flow cell experimentation, a custom in-house flow-by cell was utilized in a two-electrode configuration, with 16 cm² area (4cm x 4cm) electrodes. The working electrodes were coated with PTMA-co-PTMPMA/MWCNT solution using a drop-casting technique. Either a bare Titanium or a Platinum (Pt) sputter-coated (Pt thickness of 251.0 nm) Titanium plate was used as counter electrode. All the experiments were performed under chronopotentiometry by applying 2 mA for adsorption and -2 mA for desorption with a GenX flow rate of 1 mL/min from a Longer Instruments peristaltic pump.

3. Results and Discussion

3.1 PTMA-co-PTMPMA electrode fabrication and characterization

PTMA-co-PTMPMA was synthesized by oxidation of PTMPMA with MCPBA, as described previously.³⁶ The content of nitroxide radicals was found to be 51 mol %, determined by ultraviolet-visible spectroscopy at 460 nm absorption wavelength.³⁶ The presence of nitroxide and amine functionalities on the electrode was confirmed by X-ray photoelectron spectroscopy (XPS) (Figure S4). The lower binding energy peak (400 eV) corresponds to the N-H functional group, while the N-O• peak can be found at 402 eV. The cyclic voltammogram in Figure 3a indicated that oxidation of the N-O• to N=O+ occurred at 0.83 V vs. Ag/AgCl, while reduction occurred at 0.53 V vs. Ag/AgCl. In addition, contact angle measurements show hydrophobic properties of PTMA-co-PTMPMA, which can be expected to be beneficial for GenX uptake (Figure S5).

3.2 Electrosorption of GenX

Before each electrochemical measurement, we carried out a pretreatment step to ensure all electrodes possessed the same state of charge (fully discharged, with no counter ion bound). For pretreatment, cyclic voltammetry was carried out from 0.0 to 1.2 V vs. Ag/AgCl in a 0.1 M NaClO₄ solution (**Figure 3a**). Then, the electrode was reduced in the same solution by chronoamperometry with a 0.0 V vs. Ag/AgCl for 3 minutes. Without pre-activation, redoxelectrodes showed lower uptake capacities, as observed in **Figure S1**. Based on these results (**Figure S1**), we hypothesized that by fully reducing the electrode, the adsorption sites that promote higher hydrophobic interactions and electrostatic attractions can be activated simultaneously when applying an anodic potential and consequently increase the kinetics for the adsorption reaction. Electrochemical separation of GenX at different potentials was

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carried out in a 0.1 mM + 20 mM NaCl solution. We used 0.1 mM GenX as a way to understand the competitive behaviour of the target PFAS ions towards various electrode chemistries and potentials at higher concentrations. Later, lower concentrations in the ppb range are used for simulated water matrices. The highest uptake capacity for the redox-copolymer functionalized electrodes were observed at 0.8 V vs. Ag/AgCl (Figure 3b). At potentials higher than 0.83 V vs. Ag/AgCl, e.g. above the oxidation peak of PTMA-co-PTMPMA (Figure 3a), the uptake capacity decreased possibly due to competing adsorption of chloride caused by enhanced capacitive effects.

Based on an energy consumption calculation at different potentials (Figure S13a), 0.8 V vs. Ag/AgCl was selected as an optimal operating condition due to lower energetic costs. When comparing adsorption under electric potential at 0.8 V with open circuit adsorption, it was found that applying charge to the system enhanced the adsorption uptake capacity by more than 50 mg/g (Figure 3b). These findings prove the importance of additional electrostatic attractions of the oxoammonium cations in oxidized PTMA-co-PTMPMA towards short-chain GenX PFAS. Equilibrium isotherm study revealed that at different concentrations, our system exhibited an uptake capacity of up to 497 mg/g adsorbent when applying 0.8 V vs. Ag/AgCl for 30 minutes (Figure 3c). The isotherm result was fitted using the Langmuir and Freundlich models. The Langmuir model portraits a better fit based on the correlation coefficient (R²=0.997 for Langmuir, vs. R²=0.934 for Freundlich) (Figure S7 and Table S3).

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Time-dependent measurements were carried out to study adsorption kinetics, at both open circuit (O.C.) and at 0.8 V vs Ag/AgCl, using 2.5 mL of 0.1 mM GenX and 20 mM NaCl (Figure 3d). Interestingly, electrostatic enhancement promoted >95% of GenX removal in just 9 minutes, while for O.C., the reaction reached >95% removal only after 30 minutes. Linear fitting using the pseudo-firstorder and pseudo-second-order kinetics model were used to estimate kinetic parameters. The best-fitting model for 0.8 V vs Ag/AgCl was a pseudo-second-order kinetics model, with a rate constant of 0.00894 g*mg⁻¹ min⁻¹ (Figure S8 and Table S4). In contrast, kinetics results for open circuit experiments showed a better fit with the pseudo-first-order kinetic model, based on the correlation coefficient. The rate constant for the open circuit experiment (based on the pseudo-first-order models) was 0.00268 min⁻¹ (Figure S9 and Table S5). Our results showed comparable kinetics to other state-of-art materials such as ion exchange resins.³²

XPS analysis performed on an adsorbed PTMA-co-PTMPMA/MWCNT electrode after applying a 0.8 V potential under 0.1 mM GenX and 20 mM NaCl indicated a ratio of 9.85 between F⁻ and Cl⁻, which was equal to 0.89 GenX/Cl⁻ ratio, accounting for the number of fluorines in a Gen-X molecule (11). The separation factor was calculated using **Equation 1**, with the results showing a separation factor of 178 for GenX, indicating high selectivity of our redox polymer toward GenX over competing chloride.

Separation Factor_{GenX⁻/Cl⁻} =
$$\frac{\frac{1}{d_{Cl^-}}}{11*C_{GenX^-0}/C_{Cl^-0}}$$
 (1)

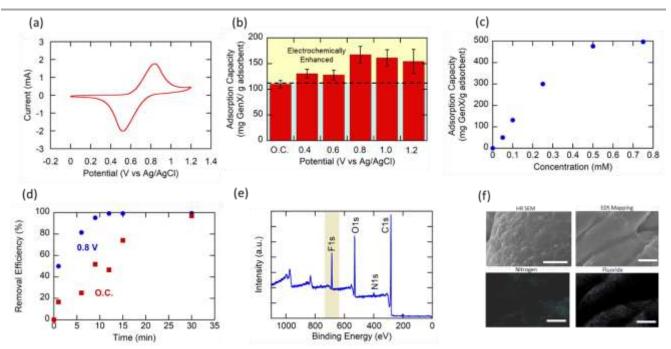
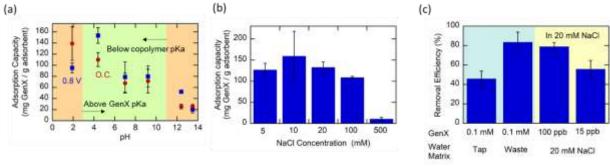
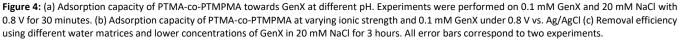


Figure 3: Characterization of the PTMA-co-PTMPMA. (a) Cyclic voltammogram of PTMA-co-PTMPMA in 0.1 M NaClO₄. Scan rate of 10 mV/s in a range from 0.0 to 1.2 V vs. Ag/AgCl. (b) Electrochemical uptake of PTMA-co-PTMPMA at different potentials. The experiments were performed under 0.1 mM GenX and 20 mM NaCl for 30 minutes applying continuous stirring of 300 rpm. Error bars correspond to three consecutive experiments. (c) GenX equilibrium isotherm for PTMA-co-PTMPMA with 0.8 V vs. Ag/AgCl for different GenX concentrations in 20 mM NaCl. (d) Comparison of adsorption kinetics of PTMA-co-PTMPMA at open circuit versus 0.8 V vs. Ag/AgCl. The experiment was performed with 0.1 mM GenX and 20 mM NaCl. (e) XPS survey of an electrode after being adsorbed under 0.1 mM GenX and 20 mM NaCl. Fluoride peak of F1s shows high intensity proving adsorption of the PFAS compound (f) Scanning electron microscopy (SEM) figures of a PTMA-co-PTMPMA electrode after 12 hours of electrosorption at 0.8 V vs. Ag/AgCl. Energy-dispersive X-ray spectroscopy (EDS) mapping highlights fluoride adsorption. The scale bar for the high-resolution image is 5 μm, SEM and EDS mapping on N/F are 20 μm.

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 C_{GenX^-eq} and C_{Cl^-eq} are the initial concentration of their respective species in solution, 0.1 mM for GenX and 20 mM for NaCl, while the ratio of solid phase uptake of q_{F^-} and q_{Cl^-} was estimated from the XPS analysis, 90.78% for F⁻ and 9.22% for Cl⁻. Figure 3e further confirmed the successful electrosorption of GenX based on the intense fluoride peak found after adsorption. Also, Figure 3f depicted a high-resolution scanning electron microscopy (SEM) image of the surface of a new electrode as well as an EDS mapping of PTMA-co-PTMPMA/MWCNT electrode ran for 30 min at 0.8V in 0.1 mM GenX and 20 mM NaCl, where fluoride could be detected after 30 minutes of electrosorption.

3.3 Electroseparation in various water matrices and pH effect

We studied the effect of solution pH on the adsorption of GenX with PTMA-co-PTMPMA. GenX's pKa is 2.84, indicating that it would exist in the anionic form at a higher pH level.²⁹ At pH 1.9, GenX is mostly found in its neutral/protonated form, and hydrophobic interaction is hypothesized to be the dominant interaction for adsorption - note that there was no electrochemically-driven enhancement observed (Figure 4a). At pH>4, the adsorption would be mainly ascribed to electrostatic interactions since now GenX exists in anionic form. Therefore, in the pH range of 4 to 9, applying positive potential exhibited better adsorption compared to open circuit. At the higher pH region >9, competition between GenX and hydroxide counter ions seems to hinder the adsorption of GenX. Also, considering that the pKa of the piperidine group is 11.28,41 the polymer becomes deprotonated at a very high pH, and therefore, hydrophobic interaction becomes dominant binding – we see that the adsorption is decreased when applying 0.8 V at pH 13.5, Figure 4a.

The impact of sodium chloride concentration was studied for 0.1 mM GenX (Figure 4b). Our findings show almost no change in adsorption capacities from 5 to 100 mM NaCl, while the adsorption in highly saline condition⁴² (500 mM NaCl) was poor. The high concentration of competing ions is screening the positive charges on the polymer, thus weakening electrostatic interactions with the GenX species. In addition, different water matrices and low concentrations of GenX were studied to test our system under near environmental conditions (Figure 4c). The water matrices studied were tap water, municipal secondary wastewater effluent (SWE, from the Urbana-Champaign Sanitary District). The water specifications of the secondary wastewater effluent can be found in the supplementary information Section S1.6. The selected concentrations were 100 ppb and 15 ppb of GenX in 20 mM NaCl

based on the high levels of PFAS that had been found in surface and groundwater ^{17, 18, 43}. Tap water and SWE were spiked with 0.1 mM of GenX, and both achieved >40% removal efficiency after 3 hours of electrosorption. Meanwhile, ultra-diluted samples showed >55% removal efficiency under the same experimental conditions.

We also translated the batch scale system into an electrochemical flow-by cell device containing a 16 cm² (4cm x 4 cm) working electrode with 8 mg of PTMA-co-PTMPMA coating, and equal dimension titanium counter electrode for the adsorption and release of GenX. Figure 5 shows the schematics of our flow cell system, and the assembly method is described in the supplementary materials (Section S1.3). Prior to electrochemical GenX adsorption, 0.1 mM GenX solution was cycled through the flow cell apparatus (1 mL/min) at open circuit for 24 hours to allow the system to equilibrate. Electrochemical GenX adsorption was then carried out via application of a constant +2 mA current for 30 minutes, at which 109 mg/g adsorbent was achieved with an energy consumption of 1 kJ/g adsorbent (2 kJ/g copolymer). The linear uptake profile in Figure 5 indicates that complete saturation of the electrode did not occur within 30 minutes and adsorption equilibrium of GenX is likely to be higher given more time. Following adsorption, the flow cell inlet was changed to a pristine solution containing no GenX and allowed to flush for 10 minutes at open circuit. During the solution change-over, very little GenX was observed to release into the pristine solution, and uptake equilibrated to 100 mg/g, indicating little to no concentration equilibrium-based release mechanism.

After adsorption, a constant current of -2 mA was applied, releasing bound GenX into a pristine flowing solution. The system showed a regeneration efficiency of 93% after 30 minutes, with 1.80 kJ/g adsorbent (3.5 kJ/g copolymer) energy consumed. The flow cell results showcase the potential of PTMA-co-PTMPMA electrochemical adsorption platform for continuous adsorption applications in GenX remediation, and highlights the importance of potential in the GenX removal process, where the bound GenX was only released once a reducing current was applied.

3.4 Electrode regeneration and an asymmetric release-destruction system

Once a negative potential is applied, the N=O+ sites were reduced to its original radical (N-O \bullet) state, promoting the release of the GenX

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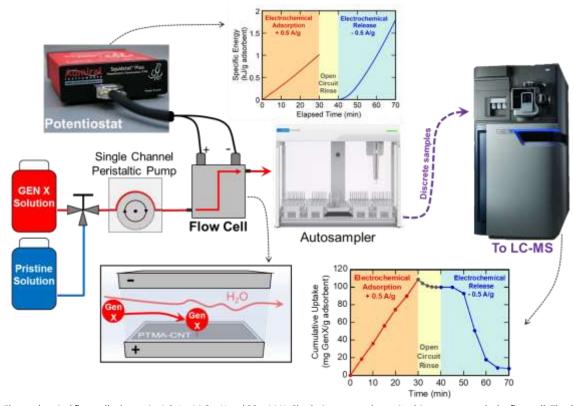


Figure 5: Electrochemical flow cell schematic. A 0.1 mM GenX and 20 mM NaCl solution enters the peristaltic pump towards the flow cell. The GenX solution could be switched to a 20 mM NaCl solution (pristine solution) for desorption. The chronopotentiometric experiments were controlled with a potentiostat and the resulting samples were collected and analysed with LC-MS. The figure also illustrates a flow cell experiment of GenX after pre-treating and pre-equilibrating the PTMA-co-PTMPMA electrode, indicating the cumulative uptake of GenX. The counter electrode used in this experiment was Ti.

molecule by electrostatic repulsion. Different release/regeneration potentials were tested, ranging from -1.5 V to 0.0 V vs. Ag/AgCl for 1 hour (**Figure S11**). The results indicated that decreasing potentials led to higher release of GenX. We selected -1 V vs. Ag/AgCl for our tests as a representative voltage, with a comparison between GenX release at -1 V vs. Ag/AgCl and open circuit shown in **Figure 6a**. Applying a reductive potential was shown to enhance desorption compared to open circuit, with a release of GenX > 80%, after one hour of electrochemical operation (**Figure 6a**).

Next, cyclability tests of our system were performed for 5 cycles of electrosorption and electrochemical release. During electrosorption, a potential of 0.8 V vs Ag/AgCl was applied for 30 min, with a solution of 5 mL of 0.1 mM GenX and 20 mM NaCl for adsorption. For desorption, a -1.0 V vs. Ag/AgCl potential was applied, and the system was reduced for 1 hour into a 20 mM NaCl. This process was repeated for five full cycles, and the results shown in **Figure S16** indicate that the adsorption capacity of our system was maintained. Interestingly, the regeneration efficiency of our system (the amount of GenX recovered during release relative to that adsorbed) decreased considerably, and after four cycles, it was maintained for up to 20% regeneration efficiency (e.g., GenX released).

As such, we hypothesized that partial degradation of GenX could have occurred during reductive regeneration (**Figure S20**). A highresolution mass spectroscopy analysis was performed on a solution that mimics a typical release concentration after an adsorption cycle. This concentration was of 0.024 mM GenX + 20 mM NaCl, and the mass spectra was analysed after applying -1.0 V vs. Ag/AgCl (Figure **S23**) for one hour to evaluate whether there was any electrochemically-mediated degradation (see chromatogram comparison in Figure S22 and HR-MS spectra on Figure S23 and S24). The HR-MS peak analysis seemed to indicate that GenX degradation could potentially occur at the carboxylate group and through breakage of the ether bond. Our results support our hypothesis that GenX could be degrading under reductive potentials and treatment times for regeneration, leading to seemingly low regeneration efficiency for separation, yet the preserved adsorption capacity. A more detailed analysis will be carried out in future studies to develop a full mechanistic model for GenX breakdown at redox-electrodes, and under varying electrochemical conditions.

The consequences from the formation of these new byproducts have not been fully studied, but based on their similar chemical structure as their former PFAS version, these C-F by-products could still be recalcitrant.⁴⁴ As such, full defluorination to fluoride should be ideally achieved as a pathway for complete remediation of these short-chain PFAS. Therefore, we decided to incorporate boron doped diamond (BDD) electrodes into the electrochemical system to increase the defluorination efficiency and further break down into fluoride during the release step. We combined our system with BDD counter electrodes, into an integrated PTMA-co-PTMPMA//BDD configuration for the purpose of reactive separation and defluorination of GenX within a single device. BDD served as an efficient candidate for defluorination since it can generate hydroxyl radicals under high overpotentials that are able to cleave the C-F

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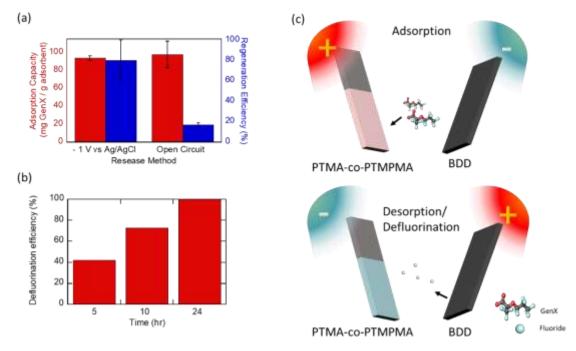


Figure 6: (a) A comparison of the regeneration efficiency of the redox-electrodes at -1 V vs Ag/AgCl vs open circuit. Adsorption was performed for 30 minutes under 0.8 V vs. Ag/AgCl and 0.1 mM GenX and 20 mM NaCl. (b) Defluorination analysis of BDD by time under 10 mA/cm². The defluorination was carried out in a 20 mM NaCl solution, after GenX released from a loaded electrode. (c) Scheme illustrating the synergistic adsorption/defluorination with PTMA-co-PTMPMA as electrosorbent and BDD as a reactive working electrode.

bond of the PFAS,⁴⁵⁻⁴⁷ while displaying significant chemical and electrochemical stability.⁴⁸ Therefore, we evaluated an integrated electrochemical system for first electrosorbing GenX and later simultaneously defluorinating the PFAS during the release step at the counter electrode, as shown in **Figure 6c**. We first adsorbed on PTMA-co-PTMPMA for 30 minutes by applying 0.8 V vs. Ag/AgCl, and then reversed the polarity and desorbed from the PTMA-co-PTMPMA electrode, simultaneously defluorinating with the help of BDD, by applying 10 mA/cm² under continuous stirring (500 rpm). Our results indicate gradual defluorination behaviour with time (**Figure 6b**), and after 24 h operation at 10 mA cm⁻², our system was able to release adsorbed GenX completely, and achieve 100% defluorination efficiency.

Conclusions

demonstrated Here. we the effectiveness of an electrochemically-mediated system for the electrosorption of GenX, using a PTMA-co-PTMPMA redox-copolymer containing nitroxide and amine moieties. The electrochemical removal performance of the redox-copolymers was evaluated across a range of GenX concentrations in different water matrices, pH, and ionic strength, proving the effectiveness of these tailored functional electrodes for the adsorption of GenX. In particular, our work highlights the enhancement of adsorption kinetics under electrochemical conditions, which showed >95% of GenX removal in 9 minutes versus 30 minutes for >95% removal with O.C. At different pH, the adsorption mechanism could be ascribed to varying degrees of hydrophobic affinity or electrostatic attraction, depending on the protonation characteristics of both the electrode and the PFAS. The redoxelectrodes were shown to release GenX and re-adsorb for sequential cycles without significant drops in uptake capacity. We also demonstrated the potential integration of the redoxelectrodes with defluorination systems such as BDD for tandem removal and remediation of GenX for up to 100% defluorination after 24 hours. Finally, the redox-electrodes were translated to a flow cell system, confirming that the electrosorption and release of GenX could be modulated under continuous electrosorption conditions. We envision future studies to improve on the electrochemical reactor engineering, such as optimization of the electrode architecture and flow-cell configuration to maximize mass transfer and enhance the electrosorption kinetics. Further mechanistic studies are also needed for studying the full pathways for short-chain degradation with different redox-electrode materials.

Conflicts of interest

There are no conflicts to declare.

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Notes and references

- 1. A. B. Lindstrom, M. J. Strynar and E. L. Libelo, Polyfluorinated Compounds: Past, Present, and Future, *Environ Sci Technol*, 2011, **45**, 7954-7961.
- Z. Wang, J. M. Boucher, M. Scheringer, I. T. Cousins and K. Hungerbühler, Toward a Comprehensive Global Emission Inventory of C4–C10 Perfluoroalkanesulfonic Acids (PFSAs) and Related Precursors: Focus on the Life Cycle of C8-Based Products and Ongoing Industrial Transition, *Environ Sci Technol*, 2017, **51**, 4482-4493.
- H. D. Whitehead, M. Venier, Y. Wu, E. Eastman, S. Urbanik, M. L. Diamond, A. Shalin, H. Schwartz-Narbonne, T. A. Bruton, A. Blum, Z. Wang, M. Green, M. Tighe, J. T. Wilkinson, S. McGuinness and G. F. Peaslee, Fluorinated Compounds in North American Cosmetics, *Environmental Science & Technology Letters*, 2021, DOI: 10.1021/acs.estlett.1c00240.
- 4. A. G. Paul, K. C. Jones and A. J. Sweetman, A First Global Production, Emission, And Environmental Inventory For Perfluorooctane Sulfonate, *Environ Sci Technol*, 2009, **43**, 386-392.
- A. Timshina, J. J. Aristizabal-Henao, B. F. Da Silva and J. A. Bowden, The last straw: Characterization of per- and polyfluoroalkyl substances in commercially-available plantbased drinking straws, *Chemosphere*, 2021, 277, 130238.
- J. M. Conder, R. A. Hoke, W. d. Wolf, M. H. Russell and R. C. Buck, Are PFCAs Bioaccumulative? A Critical Review and Comparison with Regulatory Criteria and Persistent Lipophilic Compounds, *Environ Sci Technol*, 2008, **42**, 995-1003.
- M. L. Brusseau, R. H. Anderson and B. Guo, PFAS concentrations in soils: Background levels versus contaminated sites, *Science of The Total Environment*, 2020, 740, 140017.
- X. C. Hu, D. Q. Andrews, A. B. Lindstrom, T. A. Bruton, L. A. Schaider, P. Grandjean, R. Lohmann, C. C. Carignan, A. Blum, S. A. Balan, C. P. Higgins and E. M. Sunderland, Detection of Poly- and Perfluoroalkyl Substances (PFASs) in U.S. Drinking Water Linked to Industrial Sites, Military Fire Training Areas, and Wastewater Treatment Plants, *Environmental Science & Technology Letters*, 2016, **3**, 344-350.
- M. E. Morales-McDevitt, J. Becanova, A. Blum, T. A. Bruton, S. Vojta, M. Woodward and R. Lohmann, The Air That We Breathe: Neutral and Volatile PFAS in Indoor Air, *Environmental Science & Technology Letters*, 2021, DOI: 10.1021/acs.estlett.1c00481.
- A. O. De Silva, J. M. Armitage, T. A. Bruton, C. Dassuncao, W. Heiger-Bernays, X. C. Hu, A. Kärrman, B. Kelly, C. Ng, A. Robuck, M. Sun, T. F. Webster and E. M. Sunderland, PFAS

Environmental Science: Water Research & Technology

Exposure Pathways for Humans and Wildlife: A Synthesis of Current Knowledge and Key Gaps in Understanding, *Environmental Toxicology and Chemistry*, 2021, **40**, 631-657.

- 11. Lifetime Health Advisories and Health Effects Support Documents for Perfluorooctanoic Acid and Perfluorooctane Sulfonate, https://www.epa.gov/groundwater-and-drinking-water/supporting-documentsdrinking-water-health-advisories-pfoa-and-pfos).
- 12. Chemicals Strategy for Sustainability Towards a Toxic-Free Environment, https://ec.europa.eu/environment/pdf/chemicals/2020/1

https://ec.europa.eu/environment/pdf/chemicals/2020/1 0/SWD_PFAS.pdf).

- R. H. Anderson, G. C. Long, R. C. Porter and J. K. Anderson, Occurrence of select perfluoroalkyl substances at U.S. Air Force aqueous film-forming foam release sites other than fire-training areas: Field-validation of critical fate and transport properties, *Chemosphere*, 2016, **150**, 678-685.
- G. Pitter, F. Da Re, C. Canova, G. Barbieri, M. Zare Jeddi, F. Daprà, F. Manea, R. Zolin, M. Bettega Anna, G. Stopazzolo, S. Vittorii, L. Zambelli, M. Martuzzi, D. Mantoan and F. Russo, Serum Levels of Perfluoroalkyl Substances (PFAS) in Adolescents and Young Adults Exposed to Contaminated Drinking Water in the Veneto Region, Italy: A Cross-Sectional Study Based on a Health Surveillance Program, *Environmental Health Perspectives*, **128**, 027007.
- 15. G. B. Post, Recent U.S. State and Federal Drinking Water Guidelines for Per- and Polyfluoroalkyl Substances, *Environmental Toxicology and Chemistry*, 2021, **40**, 550-563.
- S. M. Goodrow, B. Ruppel, R. L. Lippincott, G. B. Post and N. A. Procopio, Investigation of levels of perfluoroalkyl substances in surface water, sediment and fish tissue in New Jersey, USA, *Science of The Total Environment*, 2020, 729, 138839.
- X. Bai and Y. Son, Perfluoroalkyl substances (PFAS) in surface water and sediments from two urban watersheds in Nevada, USA, *Science of The Total Environment*, 2021, 751, 141622.
- T. Schwichtenberg, D. Bogdan, C. C. Carignan, P. Reardon, J. Rewerts, T. Wanzek and J. A. Field, PFAS and Dissolved Organic Carbon Enrichment in Surface Water Foams on a Northern U.S. Freshwater Lake, *Environ Sci Technol*, 2020, 54, 14455-14464.
- 19. E. Hepburn, C. Madden, D. Szabo, T. L. Coggan, B. Clarke and M. Currell, Contamination of groundwater with perand polyfluoroalkyl substances (PFAS) from legacy landfills in an urban re-development precinct, *Environmental Pollution*, 2019, **248**, 101-113.
- 20. L. W. Y. Yeung and S. A. Mabury, Are humans exposed to increasing amounts of unidentified organofluorine?, *Environmental Chemistry*, 2016, **13**, 102-110.
- A. Glynn, U. Berger, A. Bignert, S. Ullah, M. Aune, S. Lignell and P. O. Darnerud, Perfluorinated Alkyl Acids in Blood Serum from Primiparous Women in Sweden: Serial Sampling during Pregnancy and Nursing, And Temporal Trends 1996–2010, Environ Sci Technol, 2012, 46, 9071-9079.
- K. A. Pike, P. L. Edmiston, J. J. Morrison and J. A. Faust, Correlation Analysis of Perfluoroalkyl Substances in Regional U.S. Precipitation Events, *Water Research*, 2021, 190, 116685.

37.

Journal Name

- O. Quiñones and S. A. Snyder, Occurrence of Perfluoroalkyl Carboxylates and Sulfonates in Drinking Water Utilities and Related Waters from the United States, *Environ Sci Technol*, 2009, **43**, 9089-9095.
- M. Sun, E. Arevalo, M. Strynar, A. Lindstrom, M. Richardson, B. Kearns, A. Pickett, C. Smith and D. R. U. Knappe, Legacy and Emerging Perfluoroalkyl Substances Are Important Drinking Water Contaminants in the Cape Fear River Watershed of North Carolina, *Environmental Science & Technology Letters*, 2016, **3**, 415-419.
- M. Ateia, A. Maroli, N. Tharayil and T. Karanfil, The overlooked short- and ultrashort-chain poly- and perfluorinated substances: A review, *Chemosphere*, 2019, 220, 866-882.
- 26. S. Brendel, É. Fetter, C. Staude, L. Vierke and A. Biegel-Engler, Short-chain perfluoroalkyl acids: environmental concerns and a regulatory strategy under REACH, *Environmental Sciences Europe*, 2018, **30**, 9.
- 27. K. J, A. Jensen and W. M, Short-chain Polyfluoroalkyl Substances (PFAS). A literature review of information on human health effects and environmental fate and effect aspects of short-chain PFAS, 2015.
- 28. Z. Zhou, Y. Liang, Y. Shi, L. Xu and Y. Cai, Occurrence and Transport of Perfluoroalkyl Acids (PFAAs), Including Short-Chain PFAAs in Tangxun Lake, China, *Environ Sci Technol*, 2013, **47**, 9249-9257.
- S. A. Gannon, W. J. Fasano, M. P. Mawn, D. L. Nabb, R. C. Buck, L. W. Buxton, G. W. Jepson and S. R. Frame, Absorption, distribution, metabolism, excretion, and kinetics of 2,3,3,3-tetrafluoro-2-(heptafluoropropoxy)propanoic acid ammonium salt following a single dose in rat, mouse, and cynomolgus monkey, *Toxicology*, 2016, **340**, 1-9.
- F. Li, J. Duan, S. Tian, H. Ji, Y. Zhu, Z. Wei and D. Zhao, Shortchain per- and polyfluoroalkyl substances in aquatic systems: Occurrence, impacts and treatment, *Chemical Engineering Journal*, 2020, **380**, 122506.
- J. L. Harfmann, K. Tito, R. J. Kieber, G. B. Avery, R. N. Mead, M. S. Shimizu and S. A. Skrabal, Sorption of Hexafluoropropylene Oxide Dimer Acid to Sediments: Biogeochemical Implications and Analytical Considerations, ACS Earth and Space Chemistry, 2021, 5, 580-587.
- W. Wang, A. Maimaiti, H. Shi, R. Wu, R. Wang, Z. Li, D. Qi, G. Yu and S. Deng, Adsorption behavior and mechanism of emerging perfluoro-2-propoxypropanoic acid (GenX) on activated carbons and resins, *Chemical Engineering Journal*, 2019, **364**, 132-138.
- 33. D. Guo, Q. Shi, B. He and X. Yuan, Different solvents for the regeneration of the exhausted activated carbon used in the treatment of coking wastewater, *Journal of Hazardous Materials*, 2011, **186**, 1788-1793.
- K. Kim, S. Cotty, J. Elbert, R. Chen, C. H. Hou and X. Su, Asymmetric Redox-Polymer Interfaces for Electrochemical Reactive Separations: Synergistic Capture and Conversion of Arsenic, *Adv Mater*, 2020, **32**, e1906877.
- 35. Y. Kim, K. Kim, H. H. Eom, X. Su and J. W. Lee, Electrochemically-assisted removal of cadmium ions by redox active Cu-based metal-organic framework, *Chemical Engineering Journal*, 2021, **421**, 129765.
- 36. K. Kim, P. Baldaguez Medina, J. Elbert, E. Kayiwa, R. D. Cusick, Y. Men and X. Su, Molecular Tuning of Redox-

Copolymers for Selective Electrochemical Remediation, *Advanced Functional Materials*, 2020, 2004635.

- L. Bromberg, N. Ozbek, K.-J. Tan, X. Su, L. P. Padhye and T. Alan Hatton, Iron phosphomolybdate complexes in electrocatalytic reduction of aqueous disinfection byproducts, *Chemical Engineering Journal*, 2021, **408**, 127354.
- X. Su, L. Bromberg, K.-J. Tan, T. F. Jamison, L. P. Padhye and T. A. Hatton, Electrochemically Mediated Reduction of Nitrosamines by Hemin-Functionalized Redox Electrodes, *Environmental Science & Technology Letters*, 2017, 4, 161-167.
- M. Ateia, A. Alsbaiee, T. Karanfil and W. Dichtel, Efficient PFAS Removal by Amine-Functionalized Sorbents: Critical Review of the Current Literature, *Environmental Science & Technology Letters*, 2019, 6, 688-695.
- W. Ji, L. Xiao, Y. Ling, C. Ching, M. Matsumoto, R. P. Bisbey, D. E. Helbling and W. R. Dichtel, Removal of GenX and Perfluorinated Alkyl Substances from Water by Amine-Functionalized Covalent Organic Frameworks, *Journal of the American Chemical Society*, 2018, **140**, 12677-12681.
- 41. I. Y. Zhukova, E. N. Papina and I. N. Tyaglivaya, Prospects of Wasteless Technologies of Selective Alcohols Oxidation, *IOP Conference Series: Earth and Environmental Science*, 2020, **459**, 032011.
- 42. U. S. G. Survey, Saline Water and Salinity, https://www.usgs.gov/special-topic/water-scienceschool/science/saline-water-and-salinity?qt-
- science_center_objects=0#qt-science_center_objects).
 K. Hoffman, F. Webster Thomas, M. Bartell Scott, G. Weisskopf Marc, T. Fletcher and M. Vieira Verónica, Private Drinking Water Wells as a Source of Exposure to Perfluorooctanoic Acid (PFOA) in Communities Surrounding a Fluoropolymer Production Facility, *Environmental Health Perspectives*, 2011, **119**, 92-97.
- 44. J. Horst, J. McDonough, I. Ross and E. Houtz, Understanding and Managing the Potential By-Products of PFAS Destruction, *Groundwater Monitoring & Remediation*, 2020, **40**, 17-27.
- M. Pierpaoli, M. Szopińska, B. K. Wilk, M. Sobaszek, A. Łuczkiewicz, R. Bogdanowicz and S. Fudala-Książek, Electrochemical oxidation of PFOA and PFOS in landfill leachates at low and highly boron-doped diamond electrodes, *Journal of Hazardous Materials*, 2021, **403**, 123606.
- 46. Á. Soriano, D. Gorri, L. T. Biegler and A. Urtiaga, An optimization model for the treatment of perfluorocarboxylic acids considering membrane preconcentration and BDD electrooxidation, *Water Research*, 2019, **164**, 114954.
- H. Lin, J. Niu, J. Xu, H. Huang, D. Li, Z. Yue and C. Feng, Highly Efficient and Mild Electrochemical Mineralization of Long-Chain Perfluorocarboxylic Acids (C9–C10) by Ti/SnO2–Sb–Ce, Ti/SnO2–Sb/Ce–PbO2, and Ti/BDD Electrodes, Environ Sci Technol, 2013, 47, 13039-13046.
- 48. M. Panizza, in *Electrochemistry for the Environment*, eds. C. Comninellis and G. Chen, Springer New York, New York, NY, 2010, DOI: 10.1007/978-0-387-68318-8 2, pp. 25-54.

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