



**Emerging investigator series: Interdependency of Green
Rust Transformation and the Partitioning and Binding Mode
of Arsenic**

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Environmental Significance Statement

Green rust (GR) is a mixed valent Fe(II,III) hydroxide that can modify contaminant (e.g. arsenic) and nutrient (e.g. phosphate) mobility by adsorption to particle edges. However, GR transforms readily into different Fe (oxyhydr)oxide phases with unknown impacts on chemical partitioning. We tracked the mineral phase, aqueous arsenic concentration, and solid phase arsenic speciation during aging of As(V)- and As(III)-laden GR from an hour to a year. When aged anoxically for a month, GR crystallized, which increased aqueous arsenic to $>60 \mu\text{g/L}$. Further anoxic aging up to a year yielded magnetite and decreased aqueous arsenic to $<5 \mu\text{g/L}$. Oxidic aging rapidly (<1 hour) transformed GR to Fe(III) (oxyhydr)oxides, producing arsenic levels $<1 \mu\text{g/L}$, well below the WHO recommended drinking water limit.

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Emerging investigator series: Interdependency of Green Rust Transformation and the Partitioning and Binding Mode of Arsenic

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Emerging Investigator Series

Abstract

We investigated the impact of aging-induced structural modifications of carbonate green rust (GR), a mixed valent Fe(II,III) (hydr)oxide with a high oxyanion sorption affinity, on the partitioning and binding mode of arsenic (As). Suspensions of carbonate GR were produced in the presence of As(V) or As(III) (i.e. co-precipitated with As(III) or As(V)) and aged in anoxic and oxic conditions for up to a year. We tracked aqueous As over time and characterized the solid phase by X-ray absorption spectroscopy (XAS). In experiments with initial As(V) (4500 $\mu\text{g/L}$, As/Fe=2 mol%), the fresh GR suspension sorbed >99% of the initial As, resulting in approximately 14 ± 8 $\mu\text{g/L}$ residual dissolved As. Anoxic aging of the As(V)-laden GR for a month increased aqueous As to >60 $\mu\text{g/L}$, which was coupled to an increase in GR structural order revealed by Fe K-edge XAS. Further anoxic aging up to a year transformed As(V)-laden GR into magnetite and decreased significantly the aqueous As to <2 $\mu\text{g/L}$. The As binding mode was also modified during GR transformation to magnetite from sorption to GR particle edges to As substitution for tetrahedral Fe in the magnetite structure. These GR structural modifications altered the ratio of As partitioning to the solid ($\mu\text{g As/mg Fe}$) and liquid ($\mu\text{g As/L}$) phase from 2.0 to 0.4 to 14 L/mg for the fresh, month, and year aged suspensions, respectively. Similar trends in GR transformation and As partitioning during anoxic aging were observed for As(III)-laden suspensions, but occurred on more rapid timescales: As(III)-laden GR transformed to magnetite after a day of anoxic aging. In oxic aging experiments, rapid GR oxidation by dissolved oxygen to Fe(III) precipitates required only an hour for both As(V) and As(III) experiments, with lepidocrocite favored in As(V) experiments and hydrous ferric oxide favored in As(III) experiments. Aqueous As during GR oxidation decreased to <10 $\mu\text{g/L}$ for both As(V) and As(III) series. Knowledge of this

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3 interdependence between GR aging products and oxyanion fate improves
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5 biogeochemical models of contaminant and nutrient dynamics during Fe cycling and
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8 can be used to design more effective arsenic remediation strategies that rely on
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10 arsenic sorption to GR.
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1. Introduction.

Green rust (GR) is a mixed valent Fe(II,III) hydroxide that consist of sheets of edge-sharing $\text{Fe}(\text{OH})_6$ octahedra with interlayer water and anions (e.g. Cl^- , SO_4^{2-} , CO_3^{2-}).^{1, 2} GR has been reported in suboxic soils and sediments that are exposed to high Fe(II) fluxes^{3, 4} and it can be associated with microbial activity.^{5, 6} The formation of GR is also frequently observed in engineered systems, particularly as a result of Fe(0) corrosion.⁷⁻⁹ In recent years, GR has gained attention because it is a powerful reductant capable of abiotically transforming many toxic organic^{10, 11} and inorganic contaminants.^{12, 13} In addition, the typically nanocrystalline structure of GR¹⁴ gives rise to a high density of surface sites that effectively bind oxyanions, such as arsenic (As).^{15, 16} The abundance of reactive sorption sites on GR particle edges makes GR an ideal candidate for remediation of water and soils contaminated by As. For example, the controlled production of GR can be used to treat drinking water containing arsenic^{17, 18} and in-situ GR formation can be induced to improve the quality of soils or aquifers with high arsenic content.^{9, 19, 20} However, GR is only stable in a narrow window of redox potential, pH, and solution composition, and in most conditions in natural and engineered environments, GR transforms readily into other Fe (oxyhydr)oxide minerals.^{21, 22} While GR has been shown to be an effective sorbent for As and other oxyanions, less is known about the partitioning of As during GR transformation and how the presence of As impacts the GR transformation products.

Green rust can transform into a variety of products depending on aging time and solution conditions (i.e. the dissolved oxygen (DO) concentration and ionic composition). One of the most common transformation products in anoxic conditions is magnetite,^{5, 23} a highly stable, mixed valent Fe(II,III) oxide that consists of Fe in both tetrahedral and octahedral coordination.²⁴ Transformation under anoxic

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3 conditions is relevant in redoximorphic soils and sediments when redox conditions,
4 that facilitate GR formation remain stable. The anoxic transformation of GR to
5 magnetite has been reported to occur on timescales of hours to weeks in abiotic
6 systems,^{8, 25} with microbial activity and alkaline pH increasing the transformation
7 rates.^{5, 26} However, strongly sorbing oxyanions, such as phosphate (P) and silicate
8 (Si), have been shown to slow down the kinetics of magnetite formation from GR to
9 over several months.^{23, 25, 27} Oxidic transformation of GR to form different Fe(III)
10 (oxyhydr)oxides occurs quite rapidly compared to anoxic transformation,¹ which is
11 attributed to the high instability of Fe(II) in the presence of DO. Transformation under
12 oxidic conditions is relevant to natural systems exposed to temporal redox variations,
13 such as paddy soils. Complete GR oxidation by DO to goethite and lepidocrocite has
14 been reported in less than an hour, with increased DO concentrations favoring
15 lepidocrocite over goethite.²⁶ Similar to anoxic GR transformation, strongly sorbing
16 oxyanions can modify the kinetics and pathways of GR oxidation. Phosphate and Si
17 decreased the rate of GR oxidation by DO, stabilizing the GR structure for several
18 hours in air.²⁷ Furthermore, P and Si modified the end products of GR oxidation by
19 DO, with mixtures of poorly-crystalline hydrous ferric oxide and fully oxidized, ferric
20 GR favored instead of lepidocrocite and goethite.^{23, 27}

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These diverse transformations can alter key properties of the solid phase that govern the sorption reactivity, including specific surface area, coordination of surface sites, surface charge, and the average Fe oxidation state. Furthermore, if GR transforms into magnetite, aging would produce tetrahedral Fe sites in the mineral structure, which can be occupied by tetrahedral oxyanions, leading to strong, multinuclear sorption complexes.²⁸⁻³⁰ Despite the potential for changes in sorption reactivity induced by GR transformation, systematic investigations of the fate of

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3 oxyanions during GR aging and, in turn, the impact of oxyanions on GR
4 transformation pathways, are absent from the literature. Understanding how As
5 partitions to the solid or liquid phase and how the As sorption configuration changes
6 during GR aging is critical since GR transformation could result in aqueous As levels
7 below or above target drinking water limits (i.e. the World Health Organization 10
8 $\mu\text{g/L}$ provisional limit) if GR is implemented in an Fe-based As remediation strategy.
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17 In this work, we generated carbonate green rust (GR- CO_3) in the presence of
18 As(V) or As(III) and aged the As-laden suspensions in anoxic and oxic conditions
19 from an hour to a year. Arsenic was selected as the oxyanion probe compound for
20 several reasons. First, since As is redox active, changes in As oxidation state can
21 provide information on possible redox reactions involving Fe that occur during GR
22 transformation. Second, As yields high quality X-ray absorption spectra to facilitate
23 identification of sorption configuration. Third, the sorption reactivity of As(V) and
24 As(III) are comparable to other environmentally relevant oxyanions (i.e. P and Si),
25 which aids in extending the results of this work to systems containing other
26 oxyanions.³¹⁻³³ We integrated aqueous As measurements during GR aging
27 experiments with molecular-scale characterization of the Fe mineral phase and As
28 bonding environment by Fe and As K-edge X-ray absorption spectroscopy. Our
29 results advance the understanding of the fate of sorbed oxyanions during the
30 transformation of GR. This information is critical to refine models of Fe and oxyanion
31 cycling in natural systems exposed to spatial and temporal redox variation (i.e. paddy
32 soils and marine sediments) and can be used to improve As remediation strategies that
33 rely on GR as a reactive sorbent.
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2. Methods

2.1. Synthesis of green rust suspensions

Suspensions of GR were generated by Fe(0) electrolysis, also called Fe(0) electrocoagulation (EC), following previously reported synthesis procedures.³⁴ Briefly, GR suspensions were produced in 200 mL electrolyte solutions open to the atmosphere by generating 3 mM total Fe(II) by Fe(0) electrolysis. The initial solutions all contained 2 mM NaHCO₃ and 10 mM NaCl, which were added from 0.3 M NaHCO₃ and 0.5 M NaCl stock solutions, and the pH was set to 8 before electrolysis. The solutions also contained either 4500 µg/L As(V) (60 µM, As/Fe = 2 mol%) or 1125 µg/L As(III) (15 µM, As/Fe = 0.5 mol%). The initial DO of the electrolyte solution was set to 3.0 mg/L by bubbling N_{2(g)} and was measured with a Hach LDO probe. The production of GR was carried out by applying an electric current of 200 mA to the Fe(0) EC cell, which corresponds to an Fe(II) production rate of 300 µM Fe(II)/min (10 min total electrolysis time). The consumption of DO by electrochemically generated Fe(II) resulted in a final DO <0.1 mg/L, which is required for mixed valent Fe(II,III) (hydr)oxide formation.³⁴ During the production of GR by Fe(0) electrolysis, the pH was adjusted manually using pipettes with small additions of 0.01 or 0.1 M HCl or NaOH to maintain the initial value of 8.0 +/- 0.3. Immediately after electrolysis, samples were collected (t = 0 in the aging series) to measure the aqueous As concentration and to characterize the initial solid phase. X-ray diffraction measurements (Figure ESI 1) confirmed that the initial As-laden suspensions consisted of GR-CO₃ based on the positions of the basal Bragg peaks. No evidence of chloride GR (GR-Cl) nor arsenic GR (GR-As) was observed, consistent with previous work.³⁴

2.2 Aging experiments

Several 200 mL batches of As(V) or As(III)-laden GR were prepared for the aging experiments because sampling for aqueous As measurements and solid phase characterization during aging (described below) required approximately 25-30 mL per time point. For anoxic aging experiments, GR suspensions were produced and transferred immediately to an anoxic chamber (95% N₂, 5% H₂). The 200 mL GR suspensions (3 mM total Fe) were divided into separate smaller 50 mL containers without any additional electrolyte solutions and were sealed and aged within the anoxic chamber. For oxic aging experiments, freshly prepared GR suspensions were transferred into 50 mL containers and the DO was set to ≈ 9.0 mg/L (air saturated conditions, verified using Luminescent Dissolved Oxygen sensor). After sealing each 50 mL container, no additional methods were used to keep the suspensions sterile. The separate 50 mL containers of GR suspensions were well mixed with a rotary shaker at 150 ± 50 rpm. Samples for aqueous As measurements were collected at various time points from an hour to a year by withdrawing aliquots of the suspensions in the anoxic chamber (anoxic aging) or open to the atmosphere (oxic aging). Samples were passed through 0.22 or 0.45 μm filters and the filtered solution was reserved for analysis of aqueous As. Details on the aqueous arsenic measurements are provided in the electronic supplementary information (ESI). The dissolved Fe in the fresh GR suspensions was also measured. The solution pH was measured during anoxic and oxic aging and was allowed to drift from the initial value of 8.0 through the experiments. Each aging experiment was performed at least twice, with early time points (< 1 month) replicated in independent experiments 3 or 4 times.

2.3. X-ray absorption spectroscopy

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2.3.1 Data collection. Fe K-edge X-ray absorption spectroscopy (XAS) data were collected on filtered samples (preparation described in ESI) at the DUBBLE beam line (BM-26a) of the European Synchrotron Radiation Facility (ESRF, Grenoble, FR). The design of the DUBBLE beam line is described in Borsboom et al. (1998)³⁵ and Nikitenko et al. (2008).³⁶ Fe K-edge XAS data were recorded at room temperature in transmission mode out to k of 13 \AA^{-1} using ion chambers for measurements of I_0 and I_t . The vertical dimension of the X-ray beam during data collection was 1 mm and the horizontal dimension was 2 mm. Harmonic rejection mirrors were used to prevent second-order harmonics. Beam calibration was performed by setting the maximum of the first derivative of an Fe(0) foil to 7112 eV. The X-ray absorption near edge structure (XANES) region was measured with 0.35 eV steps, whereas step sizes of 0.05 \AA^{-1} were used for the extended X-ray absorption fine structure (EXAFS) region. Two to 4 scans were collected for each sample, depending on data quality.

As K-edge XAS data were collected at beam line 4-1 of the Stanford Synchrotron Radiation Lightsource (SSRL, Menlo Park, USA). Spectra were recorded at liquid nitrogen temperature ($\approx 80 \text{ °K}$) in fluorescence mode out to k of 13.5 or 14 \AA^{-1} using a Lytle detector or solid-state PIPS detector. The vertical dimension of the X-ray beam during data collection was 1 mm and the horizontal dimension was 4 to 8 mm. To prevent second-order harmonics, the X-ray beam was detuned 50%. Beam calibration was performed by setting the maximum of the first derivative of an Au(0) foil to 11919 eV. The XANES was measured with 0.35 eV steps and step sizes of 0.05 \AA^{-1} were used for the EXAFS region. Four to 10 scans were collected for each sample, depending on data quality.

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3 During data collection for both Fe and As K-edge spectra, changes in line
4 shape and peak position indicative of beam-induced redox reactions were examined
5 and no artifacts were observed. Spectra were aligned, averaged, and background-
6 subtracted using SixPack software³⁷ following standard methods described
7 previously.³⁸ The EXAFS spectra were extracted and given k^3 -weighting and were
8 Fourier-transformed over the k -range 3 to 12 or 13 \AA^{-1} using a Kaiser-Bessel window
9 with dk of 3 \AA^{-1} .

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19 2.3.2 *Fe K-edge EXAFS analysis.* The k^3 -weighted Fe K-edge EXAFS spectra
20 were analyzed by principal component analysis (PCA) and iterative transformation
21 factor analysis (ITFA) in the range of 2 – 12 \AA^{-1} using the ITFA software package.³⁹⁻
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41 The minimum of the indicator (IND) function during the PCA was used to
determine the number of principal (independent) components that reproduced the
main variance in the aging data set. The relative concentrations of the principal
components derived by PCA were varied to reconstruct each of the 14 experimental
Fe K-edge EXAFS spectra. The ITFA program requires that the relative concentration
be constrained to 1.0 for at least one of the components.⁴¹ Therefore, we constrained
the relative concentrations of the longest aged sample in the As(III) anoxic series and
the longest aged sample in the As(V) oxic series. These samples showed the highest
relative loading of the respective component in the Varimax rotation of the ITFA
procedure. Components found by ITFA to have a relative concentration <0.1 are not
reported. The sum of the relative concentrations of components derived by ITFA were
normalized to 1.0. The independent components that made up the data set were
compared with the EXAFS spectra of Fe-bearing references that were collected as
part of our previous work.³⁴ These Fe-bearing reference minerals were analyzed by

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3 shell-by-shell fits (Table ESI 1), with details on the fitting procedure reported in the
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5 ESI.
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8 *2.3.3 As K-edge XANES analysis.* The fraction of sorbed As(III) and As(V) in
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10 each sample was quantified by linear combination fits (LCFs) of the As K-edge
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12 XANES spectra using the SixPack software.³⁷ The XANES LCFs were performed
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14 over the range of 11860 to 11880 eV, with only non-negative percentages allowed,
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16 but without constraining the sum of the components to 1. Since the selection of
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18 particular references used in the LCFs can impact the fit-derived fractions,⁴² we
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20 performed three sets of XANES LCFs for each sample using three sets of references:
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22 As(III) and As(V) adsorbed to i) green rust, ii) 2-line ferrihydrite and iii) magnetite.
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24 The fraction of As(III) and As(V) in each experimental sample is reported as the
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26 average and standard deviation of these three sets of LCFs. The spectra of the As(III)
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28 and As(V) adsorption references used in the LCFs were collected at the same beam
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30 lines in identical conditions as the current data set. In addition to adsorption
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32 references, we report the spectrum of As(V) co-precipitated with magnetite (AsV Co-
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34 PPT), which was synthesized by mixing Fe(II) and Fe(III) salts in the presence of
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36 As(V) following previous procedures.⁴³
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42 *2.3.4 As K-edge EXAFS analysis.* Shell-by-shell fits were used to derive As
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44 sorption configurations for select aging series samples and reference materials. Since
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48 coordination geometry by shell fits, our analysis focused primarily on samples
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50 containing a single As oxidation state. Theoretical curve fits were performed from 1
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52 to 3.5 Å in R+ΔR-space, based on algorithms derived from IFEFFIT.⁴⁴ Parameters
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54 varied in the fits typically included the interatomic distance (R), the coordination
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56 number (CN), the mean squared atomic displacement parameter (σ^2), and the change
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3 in threshold energy (ΔE_0) for one sample. Phase and amplitude functions for single
4 and multiple scattering paths were calculated using FEFF6⁴⁵ and included As-O, As-
5 O-O and As-Fe paths derived from the structure of scorodite.⁴⁶ In preliminary fits, the
6 CN and σ^2 were found to be highly correlated, which produced high fit-derived
7 standard errors in these fitting parameters. Therefore, following previous work,^{38, 47}
8 we constrained σ^2 in the second shell fits (As-Fe path) to reduce the high correlations.
9 We constrained σ^2 to 0.009 for samples that contained a major fraction of green rust
10 or magnetite in the Fe K-edge EXAFS data, whereas σ^2 was set to 0.010 for samples
11 that contained mainly Fe(III) precipitates. For all fits, the passive electron reduction
12 parameter, S_0^2 , was set to 1.0, consistent with previous work.^{47, 48} The goodness-of-fit
13 was assessed based on the R-factor, which is reported along with fit-derived variables.
14 The R-factor is defined as the mean square difference between the fit and the data on
15 a point-by-point basis: $R = \sum_i (\text{data}_i - \text{fit}_i)^2 / \sum_i (\text{data}_i)^2$. An R-factor <0.05 is
16 considered to reflect a reasonable fit.⁴⁹
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3. Results

3.1. Changes in aqueous arsenic during aging.

Figure 1 shows the changes in dissolved As concentration of As(V)- and As(III)-laden GR suspensions aged in anoxic and oxic conditions. For the As(V) series, the residual aqueous As of the fresh (i.e. $t = 0$) sample was $14 \pm 8 \mu\text{g/L}$, which was >99% removal of the initial $4500 \mu\text{g/L}$ As(V) ($60 \mu\text{M}$). When the As(V)-laden GR was aged for a day anoxically, the aqueous As concentration increased to $>60 \mu\text{g/L}$ and it remained at this level for up to a month of anoxic aging. However, after a year of anoxic aging, the aqueous As concentration decreased substantially to $<2 \mu\text{g/L}$, the lowest As level in this aging series. These changes in aqueous As during GR aging correspond to ratios of As partitioning to the solid ($\mu\text{g As/mg Fe}$) and liquid ($\mu\text{g As/L}$) phase of 2.0, 0.4, and 14 L/mg for the fresh, month, and year aged samples, respectively. When As(V)-laden GR was aged in oxic conditions, the behavior of dissolved As differed from anoxic aging conditions. Rather than increasing on day and month timescales, dissolved As decreased rapidly from $14 \pm 8 \mu\text{g/L}$ to $<1 \mu\text{g/L}$ after an hour of aging (i.e. partitioning ratio increased from 2.0 to $>25 \text{ L/mg}$). The solution pH also decreased rapidly to <6 during oxic aging (Figure ESI 2) due to H^+ production from Fe(II) oxidation, which contrasts the stable pH (7 to 8) in the anoxic aging series. The As concentration for the oxic aging series remained $<1 \mu\text{g/L}$ throughout the duration of the experiment.

For the As(III) series, the residual As concentration for the fresh (i.e. $t = 0$) GR sample was $370 \pm 16 \mu\text{g/L}$, which corresponded to 67% removal of the initial $1125 \mu\text{g/L}$ As(III) ($15 \mu\text{M}$). When As(III)-laden GR was aged anoxically for a day, the aqueous As concentration decreased to $<150 \mu\text{g/L}$, which contrasts the behavior of aqueous As in the As(V) anoxic aging series. The As concentration decreased further

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3 when As(III)-laden GR was aged for a year, leading to a final As level below the
4 WHO 10 $\mu\text{g/L}$ limit. When the As(III) samples were aged oxically, aqueous As
5 decreased to $<10 \mu\text{g/L}$ after an hour. Although the pH of the anoxic As(III) series did
6 not change significantly, the pH of the oxic aging series decreased to <6 after an hour
7 (Figure ESI 2). Similar to the analogous As(V) samples, continued aging in oxic
8 conditions further decreased aqueous As, with the lowest As concentration ($< 1 \mu\text{g/L}$)
9 measured for the sample aged the longest (≈ 5 months).
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21 3.2. Solid phase Fe speciation.

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24 *3.2.1 Iterative transformation factor analysis.* Differences in the line-shape
25 and phase of the major oscillations of the Fe K-edge EXAFS spectra (Figure 2)
26 indicate systematic changes in solid phase Fe speciation as a function of aging
27 conditions (i.e. time, DO concentration, presence of As(V) or As(III)). Principal
28 component analysis (PCA) of the 14 Fe K-edge EXAFS spectra in the data set,
29 including the initial samples, revealed a minimum in the indicator (IND) function
30 with four independent components (Figure ESI 3). Therefore, the 14 aging sample
31 spectra were reconstructed using linear combinations of four independent components.
32 Iterative target transformation (ITT) was used to extract the spectra of the four
33 endmembers, which are compared to the EXAFS spectra of Fe-bearing reference
34 minerals in Figure 3.
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49 Component 1 was a clear match to the reference spectrum of GR based on
50 several characteristic features (Figure 3A), including the first asymmetric oscillation
51 from 3.5 to 5.5 \AA^{-1} and the subtle beat features near 7.8 \AA^{-1} and 8.5 \AA^{-1} . The Fourier
52 transforms of these spectra showed small differences in the amplitude of the second
53 shell peak arising from Fe-Fe pairs from edge-sharing FeO_6 octahedra, which
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3 suggests that Component 1 is a GR phase that is slightly less crystalline in the
4 direction of the sheets than the reference GR. Component 2 strongly resembled the
5 Fe K-edge EXAFS spectrum of magnetite, particularly the characteristic shape of the
6 doublet in the first oscillation from 3.5 to 6.0 Å⁻¹. In addition, Component 2 matched
7 the subtle features of the magnetite reference spectrum at $k > 8.0$ Å⁻¹, including the
8 shoulder at 8.2 Å⁻¹ and the peaks centered near 9.5, 10.7, and 11.3 Å⁻¹. However,
9 several oscillations in Component 2 had lower amplitudes relative to the magnetite
10 reference spectrum. These differences are manifest in the Fourier transform as a lower
11 amplitude of the shoulder at larger R in the second shell peak, which is indicative of
12 Fe-Fe pairs linked through corner sharing polyhedra. Therefore, these results suggest
13 that Component 2 is slightly less crystalline than the magnetite reference.
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28 Component 3 reproduced all major and minor features of the EXAFS
29 spectrum of lepidocrocite, such as the asymmetric first oscillation from 3.5 to 5.5 Å⁻¹,
30 the small beat near 7.8 Å⁻¹, and the intense oscillation at $k > 8.5$ Å⁻¹. The major peaks
31 of the reference lepidocrocite spectrum have slightly larger amplitude than those of
32 Component 3. This difference in amplitude corresponded to a lower amplitude of the
33 first and second shell peaks of the Fourier transform, which is consistent with slightly
34 higher crystallinity in the reference lepidocrocite spectrum relative to Component 3.
35 Component 4 displayed several features characteristic of poorly-ordered Fe(III)
36 precipitates, including a relatively symmetric first oscillation from 3.2 to 5.2 Å⁻¹, a
37 small beat feature near 5.5 Å⁻¹, and broad oscillations with low amplitude at $k > 8.5$ Å⁻¹.
38 Although there is good agreement between Component 4 and the EXAFS spectrum
39 of 2-line ferrihydrite (2LFh), some mismatches are present in the oscillations at 8.2 Å⁻¹
40 and at $k > 10$ Å⁻¹. In these regions, Component 4 is a better match to the Fe K-edge
41 EXAFS spectrum of silicate-rich hydrous ferric oxide (Si-HFO), which was prepared
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3 by Fe(III) hydrolysis in the presence of strongly-sorbing Si. As shown in the Fourier
4 transform, the shape and amplitude of the second shell peak of Component 4 lie
5 intermediate between that of 2LFh and Si-HFO. Since the structures of 2LFh and Si-
6 HFO differ mainly in the presence of more corner-sharing Fe-Fe bonds for 2LFh
7 (indicated by the arrow in Figure 3H), Component 4 represents a disordered Fe(III)
8 precipitate with fewer corner-sharing Fe-Fe bonds than 2LFh, but higher Fe-Fe
9 polymerization than Si-HFO.
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19 Considering the similarity of the ITFA-derived independent components to the
20 EXAFS spectra of Fe-bearing reference minerals, we herein refer to Component 1 as
21 GR, Component 2 as magnetite, Component 3 as lepidocrocite, and Component 4 as
22 hydrous ferric oxide (HFO).
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28 *3.2.2 Relative concentrations of Fe phases in the aging samples.* The relative
29 concentrations of GR, magnetite, lepidocrocite and HFO derived by ITFA for all
30 aging samples are shown Figure 4 (fit reconstructions are given in Figure 2). The
31 initial As(V)-laden GR suspension was found by ITFA to consist of approximately
32 equal fractions of GR (0.49) and HFO (0.51). When the As(V)-laden suspension was
33 aged anoxically for a day, the relative concentration of GR increased to 1.0 at the
34 expense of HFO. No change in solid phase Fe speciation was detected when these
35 samples were aged from a day to a month (GR = 1.0). However, anoxic aging for a
36 year produced solids that consisted of a major fraction of magnetite (0.54), with the
37 remainder being HFO. When aged in oxic conditions, the As(V)-laden GR suspension
38 transformed rapidly to a mixture of lepidocrocite (0.61) and HFO (0.39) in an hour.
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The relative concentration of lepidocrocite derived by ITFA increased systematically
from 0.61 to 1.0 when As(V)-laden GR was aged for a day (0.80), a week (0.85), and
a month (1.0) in oxic conditions.

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3 For the As(III) series, the initial sample was found by ITFA to consist of
4 mixtures of GR (0.61) and HFO (0.39), which is similar to the analogous As(V)
5 sample. When As(III)-laden GR was aged anoxically for a day, ITFA indicated a
6 transformation of the solids to magnetite (0.45) and HFO (0.55). When aged
7 anoxically for a year, the relative concentration of magnetite detected by ITFA
8 increased to 1.0, with no other Fe phases detected. Similar to the As(V) series, the
9 As(III)-laden GR samples aged in oxic conditions transformed rapidly into Fe(III)
10 precipitates, but the transformation products differed. In the As(III) series, the major
11 phase making up the solids for all aging samples was HFO instead of lepidocrocite.
12 After an hour, a day, and a month of aging in oxic conditions, the relative
13 concentration of HFO in the As(III) series was 1.0, 0.87, and 0.82, respectively.
14 Although still a minor fraction of solids in the As(III) series, lepidocrocite was
15 detected at relative concentrations that increased systematically (0, 0.13, 0.18) with
16 increased oxic aging (an hour, a day, a month), which was also observed in the
17 analogous As(V) samples.
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40 3.3. Arsenic oxidation state.

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42 For all samples in the As(V) anoxic aging series, the position of the XANES
43 absorption maxima (11875 eV) matched that of the As(V) references (Figure 5,
44 Figure ESI 4). Shoulders in the XANES indicative of As(V) reduction to As(III) were
45 not observed for the initial sample, nor for the samples aged in anoxic conditions for a
46 week or a month. However, the As(V)-laden GR sample aged anoxically for a year
47 displayed a clear shoulder in the absorption edge at 11871 eV, which is consistent
48 with the presence of a significant fraction of As(III). In addition, this sample
49 displayed unique post-edge oscillations that matched the As(V) co-precipitated with
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3 magnetite (AsV Co-PPT Mag) reference. The XANES LCFs for this aging series
4 confirmed the predominance of As(V) in all samples (>97% As(V), Table 1) except
5 for the sample aged for a year, which contained 44±10% As(III). We also note that for
6 this sample, which contained only As(V) in the initial electrolyte solution, the residual
7 aqueous As is expected to be primarily As(III) since the sorption affinity of As(V) is
8 orders of higher than As(III)³³ and aqueous As(V) could displace sorbed As(III) if it
9 were present in solution. The XANES spectra for all As(V)-laden GR samples aged in
10 oxic conditions had a maximum near 11875 eV, consistent with As(V). The presence
11 of As(III) was not supported by the XANES spectra of any As(V) sample aged
12 oxically, with the LCFs yielding >98% As(V) for each sample.

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26 In contrast to the As(V)-laden GR samples, the As(III) aging series in both
27 anoxic and oxic systems consisted of samples having major fractions of As(III) and
28 As(V), suggesting significant As(III) oxidation. For the initial sample in the As(III)
29 series, which was produced by forming GR in the presence of As(III), the XANES
30 spectrum displayed a double peak that was found by LCFs to contain 25±7% As(V)
31 (Table 1). This result is consistent with As(III) oxidation during GR formation by the
32 reactive Fenton-type oxidants produced during Fe(II) oxidation by DO (i.e. Fe(IV)).⁵⁰
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51 The fractions of As(III) and As(V) determined by XANES LCFs varied slightly
when As(III)-GR was aged in anoxic conditions (Table 1), but the fraction of As(V)
in the sample aged a year (23±5% As(V)) was consistent with the initial sample
(25±7% As(V)). Furthermore, both As(III) anoxic aging samples displayed the unique
post-edge oscillations present in the As(V) Co-PPT magnetite reference and the
As(V) sample aged for a year in anoxic conditions. This result suggests a change in
the As coordination environment for both As(III) and As(V) samples when magnetite
was detected in the Fe K-edge EXAFS spectra. The XANES spectra of the As(III)-

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3 laden GR samples aged under oxic conditions showed systematic changes consistent
4 with As(III) oxidation. The amplitude of the XANES peak centered at 11871 eV,
5 which is indicative of As(III), decreased continuously in the samples aged for an hour
6 and a day in oxic conditions. By a week of oxic aging, the As(III) peak in the XANES
7 spectrum was absent in favor of the As(V) peak at 11875 eV, which was similar to the
8 sample aged a month in oxic conditions. The XANES LCFs confirmed the increase in
9 As(V) for samples aged oxically for an hour ($31\pm 5\%$), a day ($40\pm 8\%$), a week
10 ($93\pm 15\%$) and a month ($97\pm 12\%$).
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24 3.4. Arsenic coordination environment.

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26 *3.4.1 As K-edge EXAFS spectra.* Figure 6 presents the As K-edge EXAFS
27 spectra of the As(V)- and As(III)-laden GR samples aged in anoxic and oxic
28 conditions. Consistent with the XANES spectra, the major changes in the EXAFS
29 spectra of the anoxic As(V) series occurred from a month to a year. Only subtle
30 differences from the initial sample were present in the EXAFS spectra of samples
31 aged for up to a month. Characteristic features in the EXAFS spectra appeared in the
32 As(V) sample aged anoxically for a year (highlighted by * symbols in Figure 6),
33 including more pronounced shoulders in the first and third oscillations and a beat near
34 5.8 \AA^{-1} . These features were also prominent in the EXAFS spectrum of the As(V) Co-
35 PPT magnetite reference, which suggests similar coordination environments. When
36 aged oxically, the As(V)-laden GR samples displayed nearly identical EXAFS spectra.
37 This result is consistent with the similar As K-edge XANES and Fe K-edge EXAFS
38 spectra of this aging series.
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56 Several oscillations in the EXAFS spectra of As(III)-laden GR samples aged
57 anoxically differed relative to the initial sample. For example, as aging increased to a
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3 year, the first oscillation flattened and the second oscillation grew more round with a
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5 shoulder near 6.2 \AA^{-1} . Since the XANES spectra did not show a significant change in
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7 the fractions of As(III) and As(V) between the initial sample and the anoxic aging
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9 samples, the differences in the EXAFS spectra are likely due to a difference in second
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11 shell (i.e. As-Fe) bonding environment. This result is consistent with the presence of
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13 unique post-edge oscillations in the XANES spectra for the aged samples that were
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15 absent from the initial sample. When aged in an oxic environment, the EXAFS
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17 spectra of samples aged an hour and a day were similar, consistent with the similar
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19 fractions of As(III) and As(V) in these samples (Table 1). The EXAFS spectra of
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21 samples aged oxically for a week and a month showed a narrowing of each oscillation,
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23 especially the last two peaks at 9.8 \AA^{-1} and 11.8 \AA^{-1} . The samples aged oxically for a
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25 week and a month in both As(III) and As(V) series all exhibited similar EXAFS
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27 spectra, which is consistent with the similar predominance of As(V) in these samples,
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29 suggesting comparable coordination environments.
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35 *3.4.2 Shell-by-shell fits.* The shell-by-shell fitting output (solid black lines) is
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37 overlain to the experimental data (dotted red lines) of select samples and references in
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39 Figure 7. Table 2 summarizes the results of the shell-by-shell fits. The first shell
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41 fitting results for the As(V) samples aged anoxically up to a month were similar. For
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43 example, the As-O coordination number (CN_{As-O}) and interatomic distance (R_{As-O}) for
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45 each sample was 4.3 to 4.4 and 1.69 \AA , which is consistent with As(V) in tetrahedral
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47 coordination. The fits of the second shell of these samples using an As-Fe path
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49 yielded a similar R_{As-Fe} of 3.39 to 3.41 \AA , regardless of aging time. However, the fit-
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51 derived CN_{As-Fe} for these samples decreased from 2.1 in the initial sample, which is
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53 consistent with As(V) binding to GR particle edges in the binuclear corner-sharing
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55 (2C) geometry,^{52, 53} to 1.4 to 1.5 in the samples aged a week and a month. It is
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3 important to note that the CN_{As-Fe} values for these samples are still within the high fit-
4 derived standard error on CN_{As-Fe} , but a higher number of As-Fe atomic pairs in the
5 initial sample is consistent with its slightly larger second shell amplitude (Figure 7)
6 relative to the samples aged a week and a month.
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12 Consistent with the unique XANES and EXAFS features of As(V)-laden GR
13 aged anoxically for a year, the shell-fits for this sample differed relative to the other
14 samples in this aging series. The CN_{As-O} returned by the fit (4.0 ± 0.6) was lower than
15 the other aging samples, the disorder term (σ^2_{As-O}) was higher ($0.005 \pm 0.001 \text{ \AA}^2$) and
16 the R_{As-O} was also higher ($1.71 \pm 0.01 \text{ \AA}$), which is in agreement with the presence of a
17 significant fraction of three-fold coordinated As(III) (Tables 1 and 2). Fits of the
18 second shell of this sample required two As-Fe paths positioned at $R_{As-Fe1} = 3.43 \pm 0.03$
19 \AA and $R_{As-Fe2} = 3.60 \pm 0.07 \text{ \AA}$. These R_{As-Fe} values are within the fit-derived standard
20 error of those obtained in the fit of the As(V) Co-PPT magnetite reference ($R_{As-Fe1} =$
21 $3.46 \pm 0.02 \text{ \AA}$ and $R_{As-Fe2} = 3.64 \pm 0.03 \text{ \AA}$) and are significantly longer than the fit-
22 derived R_{As-Fe} for As(V) adsorbed to magnetite ($3.34 \pm 0.02 \text{ \AA}$, Table 2). However, the
23 sample aged anoxically for a year had lower CNs on these As-Fe paths than the As(V)
24 Co-PPT magnetite reference, which could be a consequence of the high fraction of
25 As(III) in the aging sample.
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45 Despite the differences in the As K-edge XAS data of the fresh As(V)- and
46 As(III)-laden GR suspensions, these samples aged oxically produced identical shell-
47 by-shell fits for almost every sample. The fits of the As(V)-laden GR samples aged in
48 oxidic conditions for an hour, a day, a week and a month all yielded $R_{As-O} = 1.69 \text{ \AA}$ and
49 $CN_{As-O} = 4.3$ to 4.7 , which indicates As(V) in tetrahedral coordination. The second
50 shell fits of these samples were indistinguishable within standard errors ($R_{As-Fe} = 3.34$
51 \AA , $CN_{As-Fe} = 2.4$ to 2.5 , Table 2), and are consistent with a similar ^{29}Si As(V) surface
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3 complex. For the As(III)-laden GR samples aged oxically for a week and a month,
4 which were determined by XANES LCFs to consist of dominantly As(V), the shell-
5 by-shell fits yielded similar structural parameters as the As(V) oxidic aging samples.
6 For the As(III) samples, the fits yielded $CN_{As-O} = 4.3$ to 4.5 and $R_{As-O} = 1.69$ Å for the
7 first shell and $CN_{As-Fe} = 2.3$ to 2.4 and $R_{As-Fe} = 3.32$ to 3.33 Å for the second shell.
8 These fitting parameters are consistent with As(V) binding in the 2C geometry to
9 Fe(III) precipitate surfaces, which is similar to As(V) oxidic aging samples.^{52, 53}
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4. Discussion

4.1. Structural modifications of GR during aging.

The characterization data indicated that GR transformation during aging depended on aging time and solution conditions (i.e. DO concentration and initial As oxidation state). Modifications of the initial GR suspensions included: i) increased structural ordering, ii) transformation of GR to magnetite and iii) GR oxidation to form Fe(III) (oxyhydr)oxides. These three structural modifications during aging are discussed in the following subsections.

4.1.1 Increased GR ordering during anoxic aging. For the As(V) series, the Fe K-edge EXAFS analysis suggested a contribution of both GR and HFO in the initial sample (Figure 4), which was also identified in the initial sample of the As(III) series. However, when the As(V)-laden GR sample was aged for a week and a month, the HFO fraction determined by ITFA was absent and only GR was detected. One explanation for the presence of HFO in the initial As(V)- and As(III)-laden GR samples identified by ITFA is the formation of a mixture of two Fe phases: poorly-ordered Fe(III) precipitates and GR. Another interpretation is that the ITFA routine produced a fraction of HFO to account for an increase in GR crystallinity upon aging, likely caused by Ostwald ripening (i.e. the growth of larger particles at the expense of smaller ones). We argue for the latter interpretation for two reasons. First, shell-by-shell fits of the Fourier-transformed As K-edge EXAFS spectra of the initial As(V)-laden GR sample and the samples aged up to a month indicated nearly identical $R_{\text{As-Fe}}$ values (3.39 to 3.41 Å), which is the $R_{\text{As-Fe}}$ expected for ^{235}U As(V) adsorption to GR particle edges.¹⁶ If present at significant fractions in the initial sample, HFO would bind As(V) effectively because of its higher specific surface area than GR,⁵⁴ which would decrease the $R_{\text{As-Fe}}$ distance ($R_{\text{As-Fe}} = 3.34$ Å for the oxidized GR samples,

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3 Table 2). Second, shell-by-shell fits of the Fe K-edge EXAFS spectra (Table ESI 1,
4 Figure ESI 5) of the initial and aged As(V)-laden GR samples yielded similar values
5 of $R_{\text{Fe-Fe}}$ (3.17 to 3.18 Å) and $\sigma^2_{\text{Fe-Fe}}$ (0.007 to 0.008 Å²), which is inconsistent with a
6 mixture of GR ($R_{\text{Fe-Fe}} = 3.18$ Å)³⁴ and Fe(III) precipitates (edge-sharing $R_{\text{Fe-Fe}} = 3.05$
7 to 3.06 Å in 2LFh⁵⁵ and lepidocrocite⁵⁶). An increase in edge-sharing Fe-Fe atomic
8 pairs in the aged sample ($\text{CN}_{\text{Fe-Fe}} = 6.4 \pm 0.8$, Table ESI 1) relative to the initial sample
9 ($\text{CN}_{\text{Fe-Fe}} = 3.8 \pm 0.6$), rather than a decrease in $\sigma^2_{\text{Fe-Fe}}$, suggests GR aging increases
10 structural order along GR sheets. While increased GR ordering was only observed in
11 the As(V) series, we expect that this process would be detected in the As(III) series if
12 sampling was performed at shorter timescales (i.e. before GR transformed to
13 magnetite in a day). The crystallization of GR during aging is also consistent with the
14 presence of dissolved Fe in the fresh (i.e. $t = 0$) GR suspension (0.36 ± 0.06 mM
15 dissolved Fe; $\approx 10\%$ of the total Fe), which is well known to catalyze Fe phase
16 crystallization.⁵⁴

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4.1.2 GR transformation into magnetite. In both As(V) and As(III) aging series, the final product of anoxic aging was magnetite, but the timescales of magnetite formation depended on the oxidation state and concentration of As. The Fe K-edge EXAFS data indicated GR transformation to magnetite required more than a month for the As(V)-laden GR suspension, whereas As(III)-laden GR transformed to magnetite in a day. The different rates of anoxic GR transformation in the As(V) and As(III) experiments can be explained by the orders of magnitude higher sorption affinity of As(V) than As(III) for Fe precipitate surfaces³³ and the higher As/Fe solids ratio in the As(V) series. Many studies report the transformation of metastable GR to magnetite,^{8, 23, 57, 58} with this reaction likely proceeding through a dissolution and re-precipitation pathway rather than solid state transformation.²¹ Strongly-sorbing

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3 oxyanions, such as As(V), bind to GR particle edges and stabilize the GR structure,
4 which decreases the rate of dissolution (and re-precipitation).^{23, 27} We propose that the
5 higher affinity of As(V) than As(III) for Fe (hydr)oxide surfaces, combined with the
6 higher As/Fe solids ratio in the As(V) experiments, stabilized GR. Weakly bound
7 As(III) is more easily mobilized from GR surfaces, which facilitates GR dissolution
8 and re-precipitation as magnetite.
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17 It is noteworthy that when magnetite was first observed in both As(V) (a year)
18 and As(III) (1 day) aging series, a fraction of HFO was also detected by ITFA.
19 Consistent with our interpretation of the ITFA-derived fraction of HFO in the initial
20 GR samples, the detection of magnetite and HFO by ITFA most likely reflects
21 differences in the crystallinity between magnetite particles, rather than the formation
22 of a mixture of two Fe phases. Finally, the Fe K-edge EXAFS analysis indicated that
23 the magnetite formed by GR transformation was lower in crystallinity than the
24 reference magnetite synthesized by mixing Fe(II) and Fe(III) salts, which produces
25 nanoparticulate magnetite.⁵⁴ The EXAFS spectrum of magnetite formed by GR
26 transformation also matches well the EXAFS spectrum of a magnetite produced by
27 Fe(0) electrolysis in the absence of dissolved carbonate, which had a crystallite size
28 determined by transmission electron microscopy to be <10 nm.³⁴ Therefore, we
29 propose the magnetite formed by GR transformation in our study is nanoparticulate
30 with a primarily crystallite size < 10 nm.
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4.1.3 Formation of Fe(III) precipitates from GR oxidation. The most rapid GR
transformations occurred in the oxic aging series, with complete GR oxidation in less
than an hour for both As(V) and As(III) experiments. Although our oxic experiments
were performed with the fresh GR suspensions, we expect the aged GR suspensions
that were more crystalline than the fresh GR would also oxidize rapidly (i.e.

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3 timescales of minutes to hours). While the instability of GR in the presence of DO in
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5 our experiments agrees well with previous work,²⁷ the reaction products of the As(V)
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7 and As(III) oxic aging experiments are different than the goethite end product often
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9 reported for GR oxidation by DO.⁵⁹ The different products of GR oxidation in our
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11 experiments (lepidocrocite and HFO) compared to previous studies (goethite) can be
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13 attributed to a number of factors, including the relatively dilute total Fe concentration
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15 in our experiments (3 mM), the high concentration of DO (air saturated conditions),
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17 and possibly the decrease in pH from Fe(II) oxidation by DO (Figure ESI 2).
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22 Although GR was completely oxidized by DO at similar timescales in the
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24 As(V) and As(III) experiments, the reaction products in the As(V) and As(III) series
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26 were different. For the As(V) series, GR oxidation yielded primarily lepidocrocite,
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28 whereas the dominant end product of As(III)-laden GR oxidation was HFO. One
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30 explanation for the different end products in the As(V) and As(III) oxic aging
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32 experiments relates to the different sorption affinity of As(V) and As(III) for Fe(III)
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34 precipitates.³³ The oxidation of Fe(II) by DO in the presence of strongly sorbing
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36 oxyanions, such as As(V) and P, has been shown to lead to the sequential formation
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38 of poorly-ordered solids that take up oxyanions at high As/Fe or P/Fe molar ratios,
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40 followed by moderately crystalline lepidocrocite once these oxyanions are removed
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42 from solution.⁶⁰⁻⁶² By contrast, more weakly sorbing ions, such as As(III) and Si, are
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44 removed at lower oxyanion/Fe ratios, leaving the majority of Fe(II) to oxidize and
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46 polymerize in the presence of surface-poisoning oxyanions (i.e. As(III)) that prevent
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48 lepidocrocite formation.^{61, 63} This interpretation is consistent with higher aqueous As
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50 concentration of the initial As(III)-laden GR sample (370±16 µg/L) relative to the
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52 initial As(V)-laden GR sample (14±8 µg/L), indicating Fe(III) polymerized in the
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54 presence of more surface-poisoning As in the As(III) experiments.
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4.2. Behavior of arsenic during green rust transformation.

Modifications of the GR structure during aging depended on the initial As oxidation state and concentration, but in turn, these structural modification also resulted in changes in the partitioning and local bonding environment of As. These results suggest an interdependent relationship between the GR transformation pathway and the fate of As. When As(V)-laden GR was aged anoxically up to a month, the Fe K-edge EXAFS data suggested an increase in GR crystallinity, which corresponded with a significant increase in aqueous As (Figure 1). These data are consistent with a decrease in reactive surface area per mass of GR during anoxic aging due to Ostwald ripening. The structural data suggest increased ordering along GR sheets during aging, which would lower the density of adsorption sites since As(V) binds primarily to GR particle edges.^{15, 16} With respect to the As bonding environment, fits of the second shell of the Fourier-transformed As K-edge EXAFS spectra of As(V)-laden GR aged up to a month returned identical R_{As-Fe} values, but a slight decrease in the CN_{As-Fe} (2.1 ± 0.8 initially to 1.4 ± 0.7 after a month). These results could indicate that As(V) released from the particle edge during GR crystallization is associated with the solid phase by occupying interlayer sites. Intercalation of As(V) would not alter the average R_{As-Fe} because no Fe atoms would occur in the first few coordination shells of As, but the average CN_{As-Fe} would decrease, consistent with our data. However, we note that this interpretation is speculative because of the high fit-derived standard error on CN_{As-Fe} , which is a weakness of EXAFS spectroscopy.^{49, 64}

The anoxic transformation of GR to magnetite was coupled to a dramatic decrease in the aqueous As concentration for both the As(V) and As(III) aging series.

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3 Based on systematic changes in the As K-edge XAS data, the formation of magnetite
4 also coincided with a modification in the As coordination environment. For example,
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6 distinct post-edge oscillations appeared in the As K-edge XANES spectra and
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8 characteristic features in the As K-edge EXAFS data were present in all samples that
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10 contained magnetite. These unique XANES and EXAFS features arise from a change
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12 in the As bonding environment from adsorption in the ²C geometry at GR particle
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14 edges to occupation of the tetrahedral iron site in magnetite. Previous studies have
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16 shown that As(V) and As(III) can substitute for tetrahedral iron in magnetite, leading
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18 to strong multinuclear sorption complexes that more effectively remove As from
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20 solution than surface complexes.^{30, 43, 65} Therefore, the transformation of GR to
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22 magnetite enhances As uptake because magnetite has a higher density of sorption
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24 sites capable of binding As in presumably stronger uptake modes (i.e. multinuclear
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26 complexes).

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33 In both As(V) and As(III) oxic aging series, the transformation of GR to
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35 Fe(III) precipitates decreased substantially the aqueous As concentration. Even
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37 though the Fe K-edge EXAFS analysis indicated an increased fraction of
38
39 lepidocrocite with time for both aging series, which implies a decrease in the specific
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41 surface area of the solids, the As concentration still decreased with time. These results
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43 most likely reflect changes in the As oxidation state and solution pH during aging,
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45 and the larger specific surface area of the solid aging products compared to GR. For
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47 the As(V) oxic aging series, the As K-edge XAS data showed no evidence for
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49 changes in oxidation state with time. However, the shell-by-shell fits indicated a
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51 decrease in $R_{\text{As-Fe}}$ from $3.40 \pm 0.03 \text{ \AA}$ in the initial GR sample to $3.34 \pm 0.03 \text{ \AA}$, which is
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53 consistent with the difference in the average $R_{\text{Fe-O}}$ between GR (2.08 \AA) and
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55 lepidocrocite (2.00 \AA). This result suggests As(V) binds to the initial GR and its oxic
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3 aging products in a similar ^{235}C geometry.^{16, 52} In contrast to the As(V) series, the
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5 As(III) samples showed changes in As oxidation during oxic aging, with complete
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7 As(III) oxidation in a week. The oxidation of As(III) in this aging series is attributed
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9 to a combination of Fenton-type oxidants produced by Fe(II) reactions with DO at
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11 early timescales (minutes to hours) and surface catalyzed As(III) oxidation by DO at
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13 longer timescales (hours to days).⁶⁶ The orders of magnitude higher sorption affinity
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15 of As(V) than As(III) for Fe(III) precipitates³³ is consistent with the decreased
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17 aqueous As in the As(III) aging series. Furthermore, the rapid drop in solution pH
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19 measured for both As(V) and As(III) oxic aging series, caused by H^+ production
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21 during Fe(II) oxidation and Fe(III) polymerization,⁶⁷ is also expected to improve As
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23 uptake especially for samples consisting of dominantly As(V). Sorption of negatively
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25 charged As(V) ($\text{pK}_{\text{a}1} = 2.3$, $\text{pK}_{\text{a}2} = 6.8$)⁶⁷ to Fe (oxyhydr)oxides is most favorable at
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27 $\text{pH} < 4$,^{31, 68, 69} where the positive surface charge of these minerals ($\text{PZC} = 6$ to 8)⁷⁰
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29 enhances electrostatic attraction. Furthermore, lower pH would favor increased
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31 adsorption if As(V) and As(III) bind to the surface through a ligand exchange
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33 mechanism that generates OH^- . Although the pH of the As(V) series was slightly
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35 higher than the As(III) series (Figure ESI 2), which could be due to increased
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37 buffering by $\text{H}_2\text{AsO}_4^-/\text{HAsO}_4^{2-}$ or OH^- production during adsorption, the pH of both
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39 oxic aging series decreased rapidly to < 6 . In particular, the pH of the As(III) oxic
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41 aging series was near the optimum pH for As(V) sorption after several weeks of aging,
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43 which explains the $< 1 \mu\text{g/L}$ residual As concentration in the final As(III) oxic aging
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45 sample consisting dominantly As(V) (Table 1).
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56 4.3. Relevance to natural and engineered settings. 57 58 59 60

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3 Our results revealed the interdependent relationship between the GR
4 transformation pathway and the fate of oxyanions, which must be considered in
5 biogeochemical models of contaminant (e.g. As) and nutrient (e.g. P and Si) mobility
6 during Fe cycling. Depending on aging time, redox conditions, and the oxidation state
7 and concentration of As, GR transformed into a variety of end products, including
8 poorly-ordered Fe(III) precipitates, lepidocrocite and magnetite. These structural
9 modifications were tightly linked to aqueous As concentrations, with anoxic GR
10 transformation to magnetite and oxic GR transformation to Fe(III) (oxyhydr)oxides
11 both sequestering As more effectively than the initial GR phase. Since As(V) and
12 As(III) are chemical analogues of P and Si with respect to charge and sorption
13 reactivity with Fe (oxyhydr)oxides,³³ our results suggest that GR transformation in
14 anoxic conditions to magnetite would reduce the mobility and bioavailability of As, Si,
15 and P, with the potential for enhanced oxyanion immobilization by occupying
16 tetrahedral Fe sites.^{30, 43} Furthermore, we expect that the presence of Si and P would
17 lead to similar GR oxidation products (i.e. HFO), but these ions would decrease the
18 As/Fe ratio of the solid phase because they compete effectively for sorption sites on
19 Fe(III) precipitate surfaces. We also provide compelling evidence that GR
20 transformation to magnetite can abiotically reduce As(V) to As(III). This constraint
21 on the reduction capacity of GR is particularly important because GR minerals are
22 thought to play an important role in the cycling of elements in Precambrian oceans.^{71,}
23 ⁷² When GR was exposed to oxic conditions, our results suggest that rapid
24 transformation to Fe(III) precipitates would also decrease oxyanion mobility by
25 producing solids with higher sorption site density and by lowering solution pH.
26 However, we note that the decrease in pH upon GR oxidation in our experiments,
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3 which likely enhanced As uptake, might not occur in natural systems with higher
4 buffering capacity.
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8 With respect to As remediation, Fe-based engineered solutions can rely on
9 mixed valent Fe(II,III) (oxyhydr)oxides to decrease aqueous As.^{8, 9, 18, 20} In this study,
10 we found that structural modifications of the initial GR resulted in differences
11 between As concentrations spanning 10 µg/L, the WHO provisional limit of As in
12 drinking water. For example, in the As(V) series, the initial GR phase removed As to
13 near the WHO limit, but after anoxic aging for a day and a month, aqueous As
14 increased to concentrations ~5 times higher than the WHO limit. Further anoxic aging
15 for a year resulted in a decrease in aqueous As to levels ~5 times lower than the WHO
16 limit. This temporal variation in aqueous As during GR formation must be taken into
17 account when designing GR-based remediation schemes, and especially the high
18 potential for As release during GR crystallization before magnetite formation in
19 anoxic systems. Another design consideration is that the release of As during GR
20 crystallization might be desired in some situations. For example, it could be
21 advantageous to employ a sorbent with properties that favor low cost and simple As
22 sorption reversibility. Using a sorbent that releases As easily would decrease the
23 amount of solid waste to be managed and would facilitate separating As from the
24 solid phase, making waste reuse or reprocessing safer and simpler. The release of As
25 from magnetite formed in the anoxic GR aging experiments is expected to be more
26 challenging than GR, making reuse of the As-rich magnetite solids more difficult.
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51 While we found that the initial GR phase removed high fractions of As from
52 solution, the residual As concentration decreased when GR transformed to magnetite
53 and Fe(III) precipitates. This conclusion appears to be inconsistent with previous
54 studies designating GR as one of the most effective Fe-based sorbents for As. This
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3 discrepancy arises from comparing sorbent reactivity based on sorption capacity
4 rather than the residual dissolved As concentration. For GR, the As sorption capacity
5 has been reported to be larger than other Fe (oxyhydr)oxides, presumably because GR
6 favors polymeric As surface species at high aqueous As concentrations.¹⁵ However,
7 comparisons based on minimized residual As for a given sorbent amount (or for a
8 given treatment cost) can be more meaningful than sorption capacity in the context of
9 drinking water treatment since the sorption capacity is often obtained in laboratory
10 experiments with unrealistically high As concentrations (i.e. several mg/L). Therefore,
11 if the type of Fe phase can be tuned, we propose that magnetite or Fe(III)
12 (oxyhydr)oxides should generally be targeted to optimize As treatment, with a notable
13 exception. Remediation schemes based on pre-synthesized Fe (oxyhydr)oxides coated
14 on sands or used in a filter columns could exploit the metastability of GR to improve
15 As removal. Coating sand grains with stable Fe mineral phases would limit sorbent
16 reactivity to the particle exterior. Since GR transforms to magnetite and Fe(III)
17 precipitates likely by dissolution and reprecipitation,²¹ a high fraction of the Fe atoms
18 making up the solid could be involved in binding As after GR transformation, and in
19 the case of GR transformation to magnetite, the As uptake mode could shift from
20 adsorption to GR particle edges to substitution for tetrahedral Fe and (partial)
21 incorporation.^{30, 43}
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5. Conclusions

In this study, we found an interdependent relationship between the pathways of GR transformation and the fate of As. The end products of GR transformation depended on the oxidation state and concentration of As (in addition to aging time and DO concentration), and the partitioning and uptake mode of As depended on the GR transformation products. In anoxic conditions, GR transformed to magnetite over timescales of a day (As(III) series) to a year (As(V) series), which decreased the aqueous As concentration to $<10 \mu\text{g/L}$ and resulted in a transition from As adsorbed to GR particle edges to As taken up in multinuclear complexes by substituting for tetrahedral Fe. In oxic conditions, GR transformed to primarily lepidocrocite (As(V) series) or HFO (As(III) series) in an hour, which decreased the aqueous As concentration to $<10 \mu\text{g/L}$, but did not modify the As uptake geometry (similar ^{2}C As(V) surface complexes before and after oxic aging). These results advance our understanding of the impact of Fe cycling on oxyanion mobility in the environment and can be used to improve predictions of As removal efficiency in Fe-based As remediation strategies.

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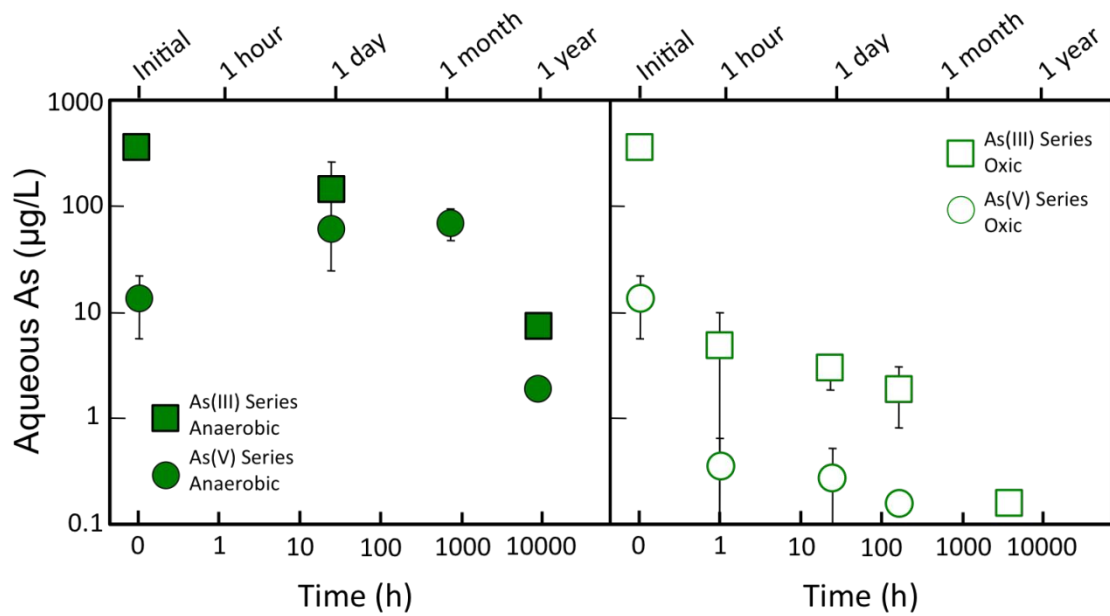


Figure 1: Aqueous As concentration during aging of green rust suspensions in anoxic (solid symbols) and oxic (open symbols) environments. The circles represent experiments with initial As(V) and the squares represent experiments with initial As(III). Data points and error bars represent the average and standard deviation of replicate aqueous As measurements, with some error bars smaller than the symbol.

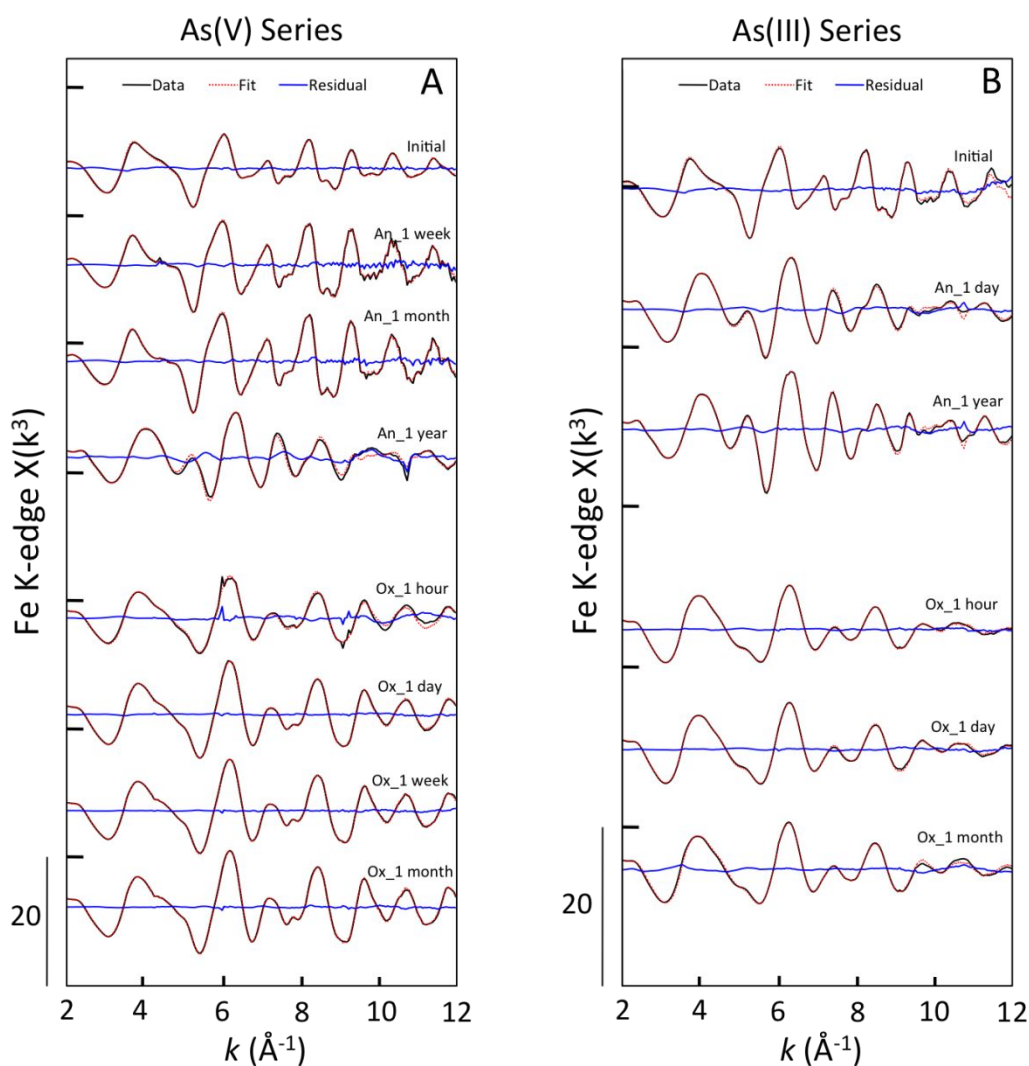


Figure 2: Fe K-edge EXAFS spectra of green rust aging samples in As(V) experiments (left panel) and As(III) experiments (right panel). The ITFA fitting output is shown in dotted red lines and the data is given in solid black lines. The fitting residual is given in solid blue lines. The anoxic aging series (e.g. An_1 week) are given above the oxic aging samples (e.g. Ox_1 hour).

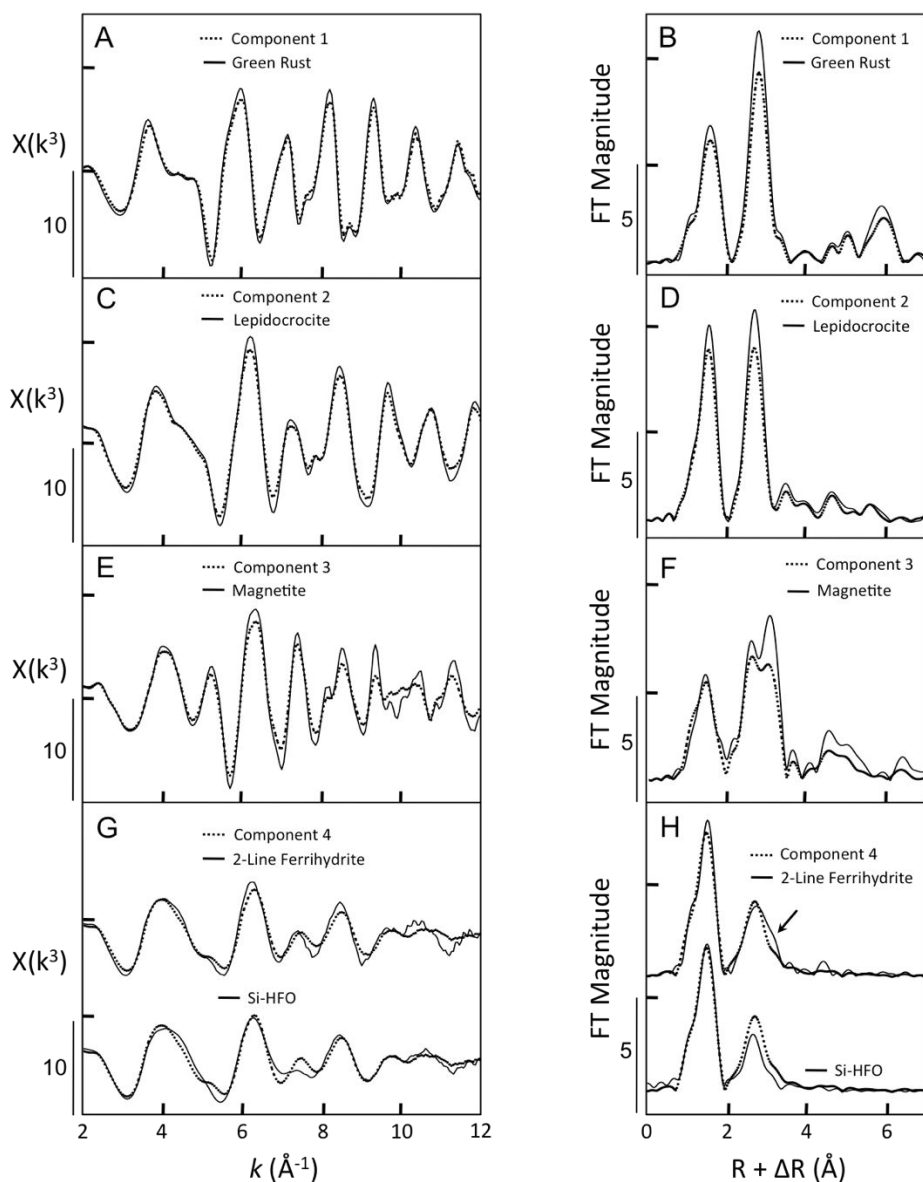


Figure 3: Comparison of ITFA-derived components and Fe K-edge EXAFS spectra of Fe-bearing reference materials. The EXAFS spectra are given in the left panel and the corresponding Fourier transforms are shown in the right panel. The dotted black lines represent the components and the solid black lines represent measured reference spectra. The arrow in H highlights features in the Fourier-transformed EXAFS spectrum of 2-line ferrihydrite that arise from corner-sharing Fe-Fe bonds.

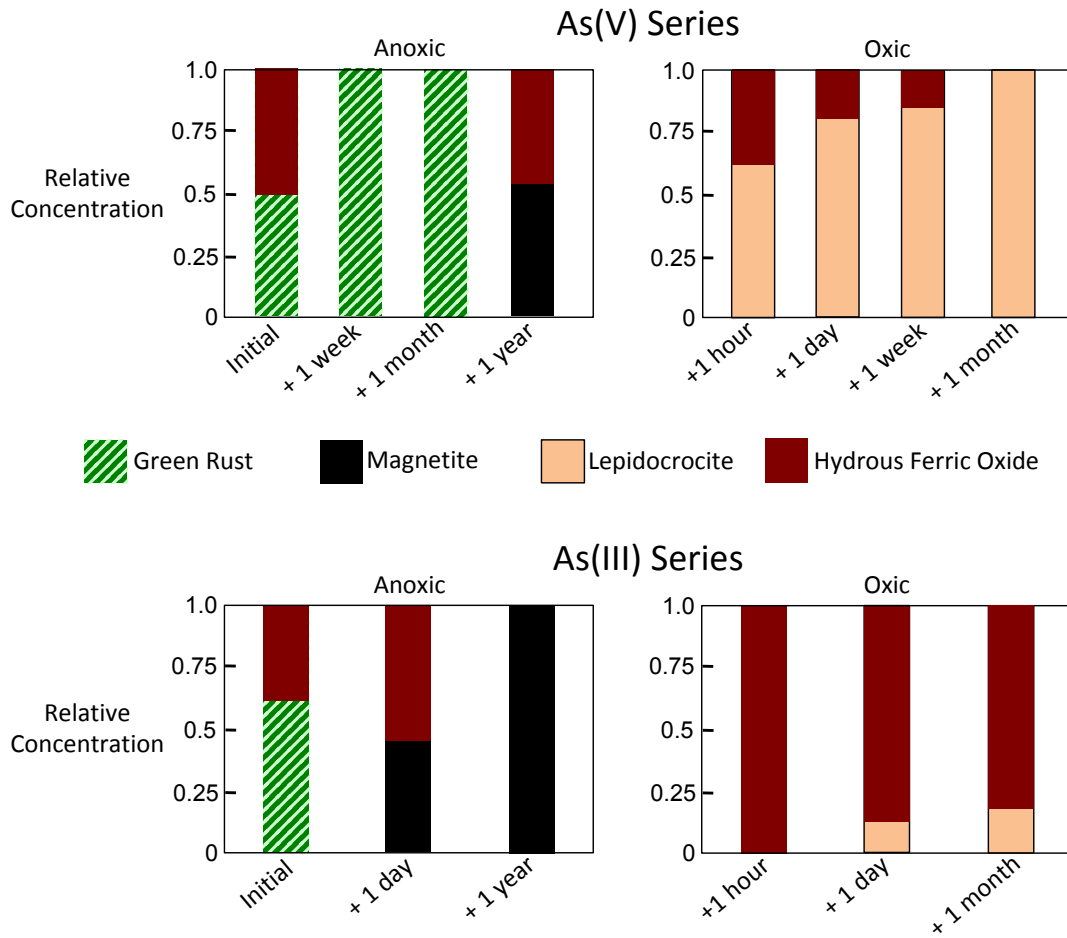


Figure 4: Relative concentrations of Component 1 (green rust), Component 2 (magnetite), Component 3 (lepidocrocite) and Component 4 (hydrous ferric oxide) derived by ITFA.

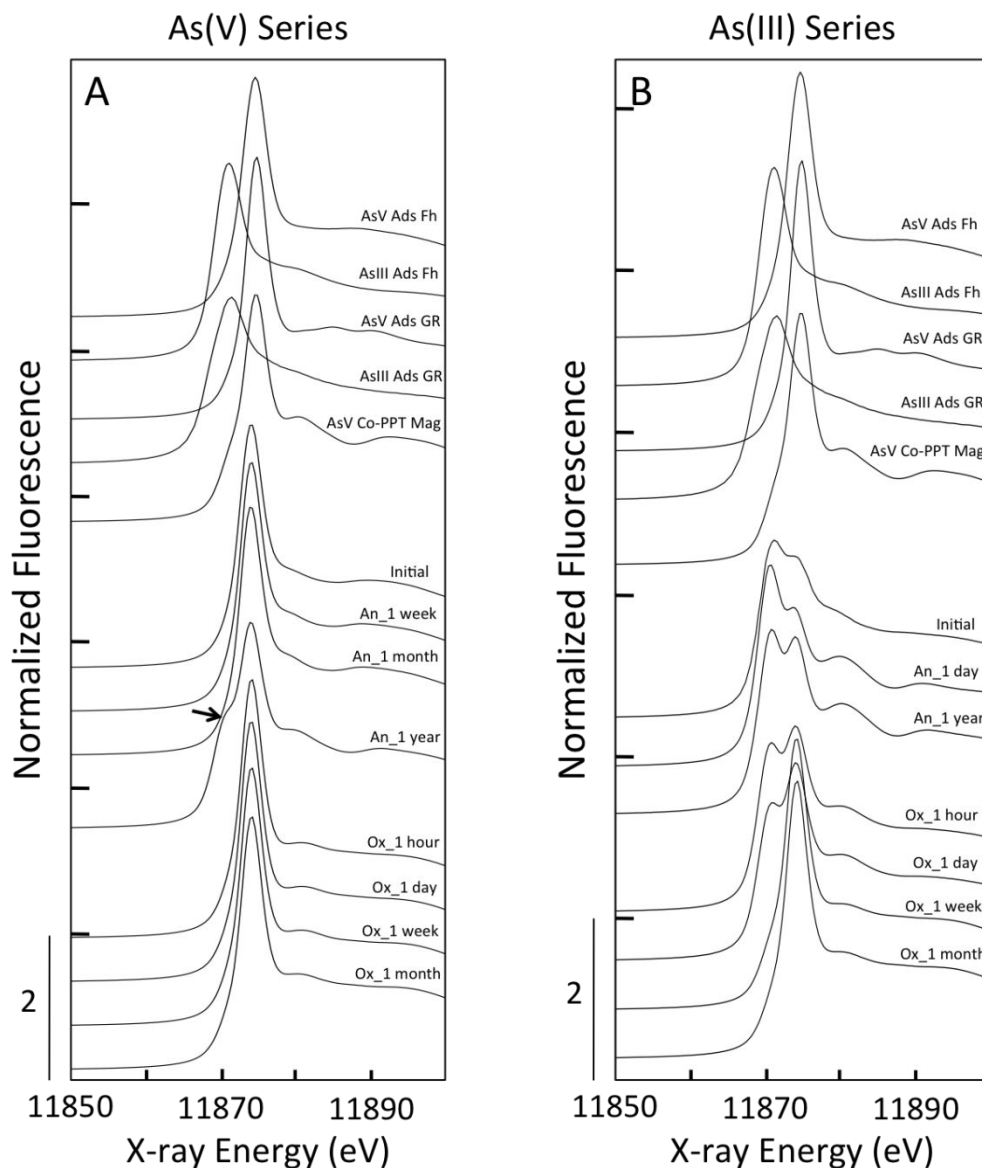


Figure 5: As K-edge XANES spectra for green rust aging samples in As(V) experiments (A) and As(III) experiments (B). The experimental data are given below As(III) and As(V) bearing reference materials. The anoxic aging series (e.g. An_1 week) are given above the oxic aging samples (e.g. Ox_1 hour). The arrow in A highlights a shoulder in the anoxic aging sample resulting from the presence of As(III).

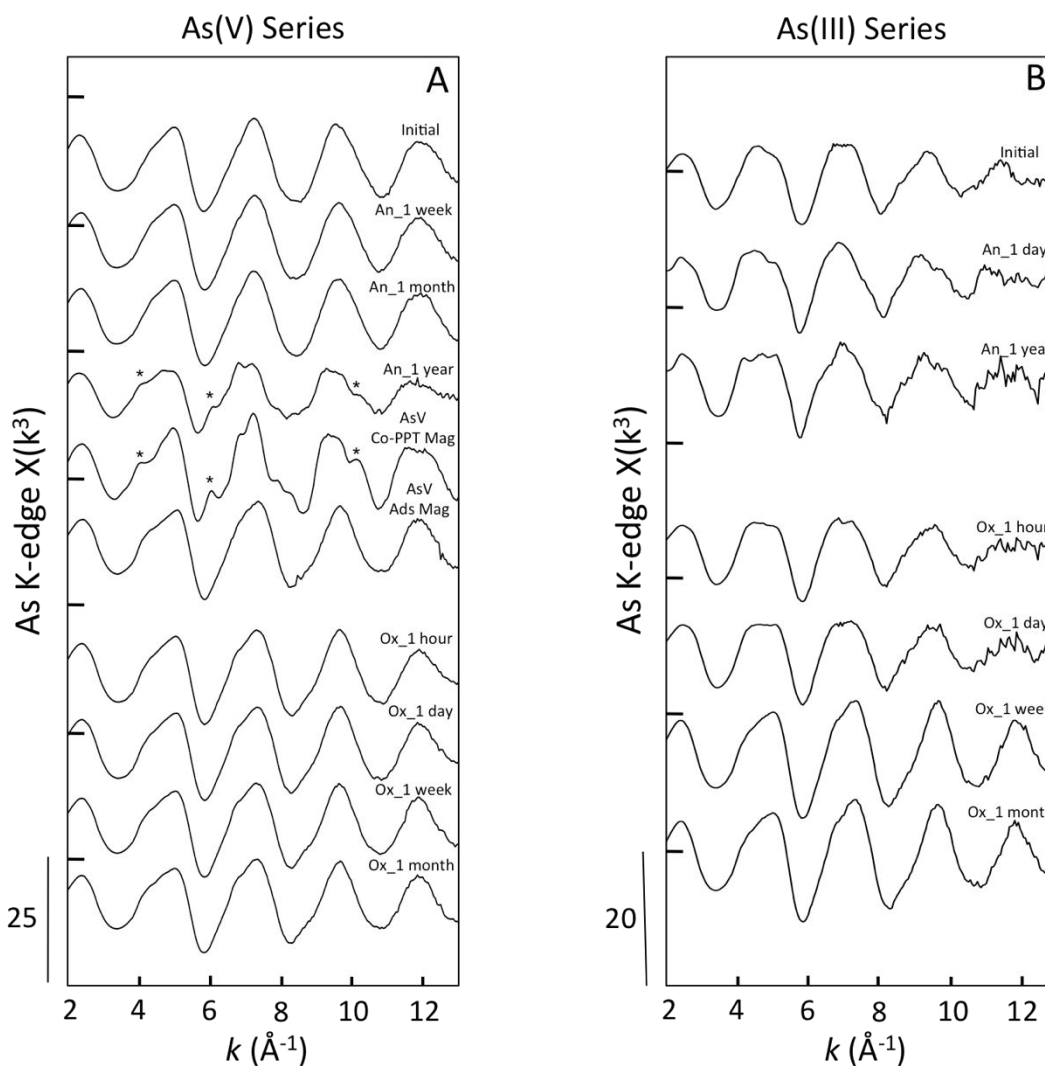


Figure 6: As K-edge EXAFS spectra of aging samples in As(V) experiments (A) and As(III) experiments (B). Reference spectra of As(V) co-precipitated with magnetite (AsV Co-PPT Mag) and adsorbed to magnetite (AsV Ads Mag) are given for comparison. The * symbols indicate EXAFS features consistent with As occupying the tetrahedral Fe site in magnetite.

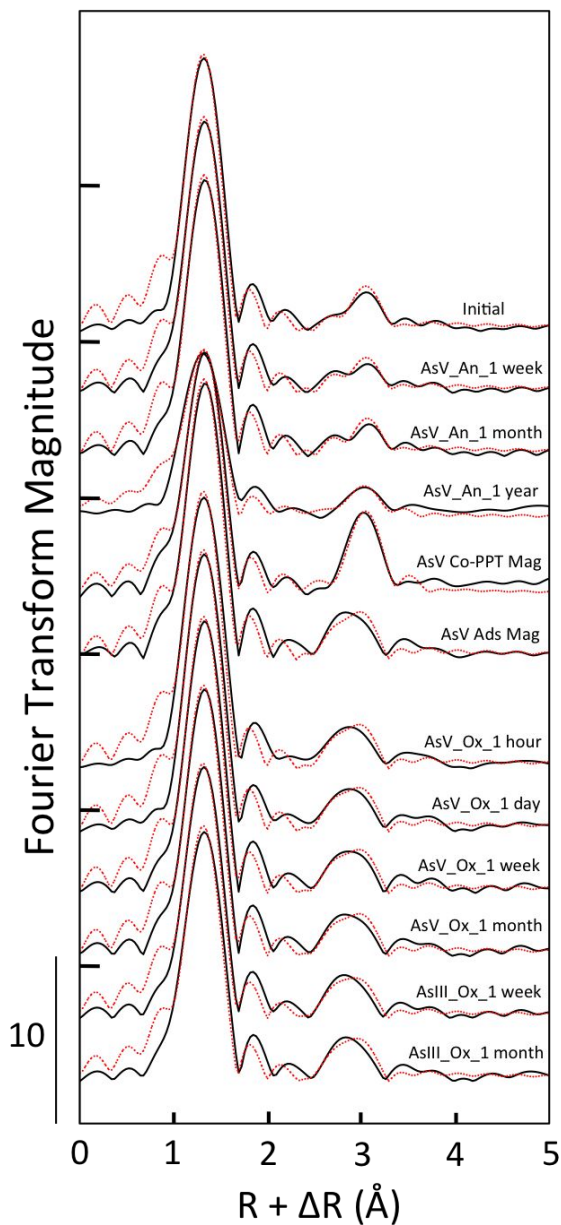


Figure 7: Fourier-transformed As K-edge EXAFS spectra of select aging samples in As(V) and As(III) experiments. The shell-by-shell fitting output (dotted red lines) is overlain to the data (solid black lines). In addition to aging samples, the fits of As(V) co-precipitated with magnetite (AsV Co-PPT Mag) and adsorbed to magnetite (AsV Ads Mag) are given.

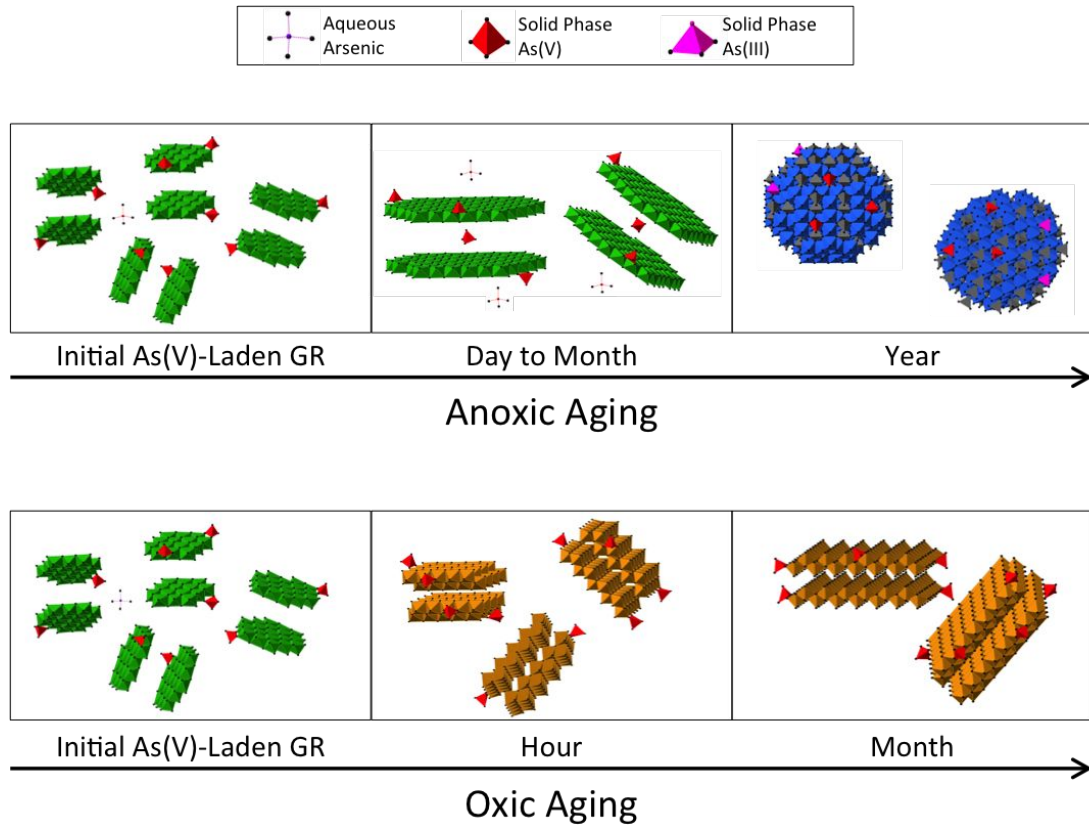


Figure 8: Models of the transformation of As(V)-laden green rust (GR) over time in anoxic (top panels) and oxic conditions (bottom panels). The initial GR transforms into magnetite (blue spherical crystals) in a year in the anoxic system, whereas lepidocrocite (brown layered crystals) forms in an hour in the oxic system. Experiments with As(III)-laden GR followed similar trends.

Table 1: Summary of As K-edge XANES Fitting Results

Sample	Aging Conditions	% As ³⁺	% As ⁵⁺
As(V) Series	Initial	7 ± 8	97 ± 10
	Anoxic +1 week	6 ± 7	99 ± 9
	Anoxic +1 month	7 ± 7	98 ± 9
	Anoxic +1 year	44 ± 10	62 ± 11
	Oxic + 1 hour	5 ± 6	101 ± 10
	Oxic + 1 day	6 ± 7	100 ± 10
	Oxic + 1 month	8 ± 8	98 ± 12
As(III) Series	Initial	78 ± 7	25 ± 7
	Anoxic +1 day	94 ± 6	9 ± 4
	Anoxic +1 year	82 ± 6	23 ± 5
	Oxic + 1 hour	73 ± 7	31 ± 6
	Oxic + 1 day	65 ± 7	40 ± 8
	Oxic +1 week	13 ± 12	93 ± 15
	Oxic + 1 month	9 ± 9	97 ± 12

Table 2: Summary of As Shell-by-shell Fits for Select Aging Samples and References

Sample	Atomic Pairs	CN	R (Å)	σ^2 (Å ²)	ΔE_0 (eV)	R-Factor
As(V)_Initial	As-O	4.4 (0.5)	1.69 (0.01)	0.003(0.001)	3.8 (2.0)	0.019
	As-O-O	12	1.82(R_{As-O}) = 3.08	σ^2 (As-O)		
	As-Fe1	2.1 (0.8)	3.40 (0.03)	0.009		
As(V)_Anoxic 1 Week	As-O	4.4 (0.5)	1.69 (0.01)	0.003 (0.001)	5.4 (1.8)	0.016
	As-O-O	12	1.82(R_{As-O}) = 3.08	σ^2 (As-O)		
	As-Fe1	1.4 (0.7)	3.39 (0.04)	0.009		
As(V)_Anoxic 1 Month	As-O	4.3 (0.4)	1.69 (0.01)	0.002 (0.001)	5.6 (1.8)	0.016
	As-O-O	12	1.82(R_{As-O}) = 3.08	σ^2 (As-O)		
	As-Fe1	1.5 (0.7)	3.41 (0.03)	0.009		
As(V)_Anoxic 1 Year	As-O	4.0 (0.6)	1.71 (0.01)	0.005 (0.001)	1.9 (2.4)	0.032
	As-O-O	12	1.82(R_{As-O}) = 3.11	σ^2 (As-O)		
	As-Fe1	2.7 (1.2)	3.43 (0.03)	0.009		
	As-Fe2	1.3 (1.4)	3.60 (0.07)	σ^2 (As-Fe1)		
As(V) Co-PPT Magnetite	As-O	4.2 (0.5)	1.70 (0.01)	0.003 (0.001)	4.5 (1.9)	0.015
	As-O-O	12	1.82(R_{As-O}) = 3.09	σ^2 (As-O)		
	As-Fe1	7.7 (1.5)	3.46 (0.02)	0.009		
	As-Fe2	3.9 (1.9)	3.64 (0.03)	σ^2 (As-Fe1)		
As(V) Ads Magnetite	As-O	4.2 (0.5)	1.69 (0.01)	0.002 (0.001)	6.0 (2.0)	0.019
	As-O-O	12	1.82(R_{As-O}) = 3.08	σ^2 (As-O)		
	As-Fe1	2.2 (0.8)	3.34 (0.02)	0.009		
As(V)_Oxic 1 Hour	As-O	4.7 (0.6)	1.69 (0.01)	0.003 (0.001)	4.4 (2.3)	0.024
	As-O-O	12	1.82(R_{As-O}) = 3.08	σ^2 (As-O)		
	As-Fe1	2.4 (1.0)	3.34 (0.03)	0.010		
As(V)_Oxic 1 Day	As-O	4.5 (0.5)	1.69 (0.01)	0.003 (0.001)	4.5 (2.1)	0.021
	As-O-O	12	1.82(R_{As-O}) = 3.08	σ^2 (As-O)		
	As-Fe1	2.4 (0.9)	3.34 (0.03)	0.010		
As(V)_Oxic 1 Week	As-O	4.3 (0.05)	1.69 (0.01)	0.002 (0.002)	4.3 (1.9)	0.018
	As-O-O	12	1.82(R_{As-O}) = 3.08	σ^2 (As-O)		
	As-Fe1	2.5 (0.8)	3.34 (0.03)	0.010		
As(V)_Oxic 1 Month	As-O	4.3 (0.5)	1.69 (0.01)	0.002 (0.001)	4.0 (1.9)	0.017
	As-O-O	12	1.82(R_{As-O}) = 3.08	σ^2 (As-O)		
	As-Fe1	2.4 (0.8)	3.34 (0.03)	0.010		
As(III)_Oxic 1 Week	As-O	4.3 (0.5)	1.69 (0.01)	0.003 (0.001)	2.9 (2.0)	0.018
	As-O-O	12	1.82(R_{As-O}) = 3.08	σ^2 (As-O)		
	As-Fe1	2.3 (0.8)	3.32 (0.03)	0.010		
As(III)_Oxic 1 Month	As-O	4.5 (0.5)	1.69 (0.01)	0.003 (0.001)	3.7 (2.0)	0.019
	As-O-O	12	1.82(R_{As-O}) = 3.08	σ^2 (As-O)		
	As-Fe1	2.4 (0.8)	3.33 (0.03)	0.010		

CN represents the coordination number, R the interatomic distance, σ^2 the mean squared atomic displacement and ΔE_0 the change in threshold energy. Fitting parameters allowed to float are given with fit-determined standard errors in parenthesis, while constrained parameters appear without a parenthesis.

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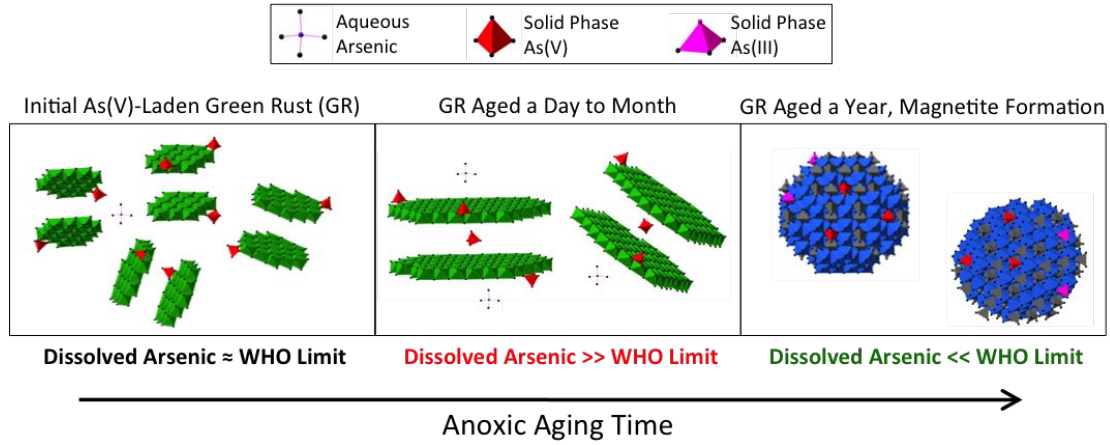
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Table of Contents Entry



Structural transformations of green rust and the partitioning of arsenic were tracked during aging for up to a year