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1 Ammonium retention by oxidized biochars produced at different
2 pyrolysis temperatures and residence times

3 Bing Wang^{a, b, *}, Johannes Lehmann^{b, c}, Kelly Hanley^b, Rachel Hestrin^b, Akio
4 Enders^b

5 ^a *State Key Laboratory of Environmental Geochemistry, Institute of Geochemistry, Chinese Academy of*
6 *Sciences, Guiyang 550002, China*

7 ^b *Department of Crop and Soil Sciences, Cornell University, 909 Bradfield Hall, Ithaca, NY 14853, USA*

8 ^c *Atkinson Center for a Sustainable Future, Cornell University, Ithaca, NY, USA*

*Corresponding author. Tel.: +86 851 85891611

E-mail addresses: wangbing@vip.gyig.ac.cn (Bing Wang), cl273@cornell.edu (J. Lehmann),
klh54@cornell.edu (K. Hanley), rh482@cornell.edu (R. Hestrin), ae55@cornell.edu (A. Enders).

9 Abstract

10 In order to investigate the effects of pyrolysis conditions and oxidation on the
11 retention potential of ammonium by biochar in aqueous solution, biochars were produced
12 from mixed maple wood at different pyrolysis temperatures (300, 400, 500, 600, 700°C)
13 and residence times (5, 60, 120, 400, 800 min) and adsorption and desorption was
14 determined. Hydrogen peroxide was used to oxidize the biochars to pH values ranging
15 from 7.6 to 2.7, with one set being adjusted to a pH of 7 afterwards. Without oxidation,
16 varying either pyrolysis temperatures or residence times did not have a relevant effect on
17 ammonium adsorption. When oxidized, however, ammonium adsorption was up to 3.6
18 and 1.6 times greater at lower than higher pyrolysis temperatures and shorter than longer
19 residence times, respectively. Neutralizing the oxygen-containing surface functional
20 groups on oxidized biochar to pH 7 further increased ammonium adsorption three to four-
21 fold for biochars originally at a temperature of 500°C and residence time of 5 min, but
22 did not change adsorption of biochars pyrolyzed at 600°C and above and residence time
23 at 400 min and above. Adjusting the pH of unoxidized biochars had no effect on
24 ammonium adsorption. Both pyrolysis temperature and residence time significantly
25 influence the way oxidation changes charge properties with respect to ammonium
26 adsorption by woody biochar.

27 **Keywords:** Biochar; Biomass; Slow pyrolysis; Adsorption; Ammonium nitrogen

28 **1. Introduction**

29 Agricultural non-point source pollution caused by large amounts of nitrogen
30 fertilizers being used with low use efficiency has become a prominent problem which
31 constrains sustainable agricultural development, since a significant portion of fertilizer N
32 is lost from agricultural fields by leaching.¹ Such a loss is not only of economic concern
33 for the farmer, but also bears an environmental consequence of pollution of ground and
34 surface waters including marine ecosystems.² In addition to improved timing and dosing
35 of N applications, also an improved retention of N in the soil will help in enhancing the
36 use efficiency of applied fertilizers.^{3,4}

37 Recently, interest in biochar as a soil amendment has steadily increased and biochar
38 has attracted wide research interest. Biochar has been touted as a soil amendment to
39 improve degraded soils and increase agronomic yield by potentially changing the soil pH,
40 cation exchange capacity (CEC), reducing leaching of nutrients, and supplying nutrients
41 to plants.⁵⁻⁸ Various pyrolysis conditions can result in biochars with different physical
42 and chemical properties, nutritional and agricultural value. Among all these factors,
43 biochar pyrolysis temperature and residence time have been considered to be the greatest
44 overall factors that influence the final characteristics of biochar, such as porous structure,
45 specific surface area and adsorption capacity as well as the retention of nutrients.⁹⁻¹²

46 A lot of research has been carried out to study the effects of pyrolysis condition on
47 the quality and properties of biochar.^{10, 13-17} Although several studies have investigated
48 the adsorption of ammonium, phosphate and nitrate of different biochars,¹⁸⁻²⁰ few
49 research has been conducted with regard to the oxidation effect on ammonium adsorption
50 of biochar having different properties. Some studies found that biochar can be used for

51 recovering excess nitrogen ²⁰ and slowly release essential nutrients to soil in order to
52 improve agricultural properties, ¹⁹ improve N use efficiency, ²¹ and reduce leaching losses
53 of N.²² Key chemical and physical properties of biochar are greatly affected both by
54 choice of feedstock and process conditions (mainly temperature, residence time, heating
55 rate and feedstock preparation). These properties affect the interactions of biochar with
56 the soil as well as its fate in the environment. This underlines the importance of
57 evaluating the effect of pyrolysis conditions on the nitrogen retention potential before
58 land application.

59 Biochar is expected to be highly oxidized after long-term exposure to natural
60 oxidation processes through the formation of oxygen-containing functional groups which
61 can increase the surface acidity and CEC, ²³⁻²⁶ which influences the nutrient retention
62 capacity of biochar. Ammonium retention by biochar may be readily explained by
63 electrostatic adsorption to negatively charged surface functional groups. ^{19, 24} However,
64 there has been little experimental evidence that this oxidation can be replicated by short-
65 term abiotic oxidation in the laboratory under different pyrolysis conditions, which may
66 inform post-production manipulation of biochars. In addition, it is unclear whether the
67 adsorption found with aged biochars in soil is influenced by pyrolysis conditions.
68 Therefore, in order to develop a framework for the selection of biochars, the effects of
69 pyrolysis temperature and residence time on ammonium nitrogen retention by oxidized
70 biochar from aqueous solution need to be better understood.

71 The objectives of this study were to: (1) examine the retention potential of
72 ammonium by biochars that have been oxidized to different extents; (2) establish the
73 optimum pyrolysis temperature and residence time at which oxidized biochars adsorb the

74 most ammonium in aqueous solution.

75

76 **2. Experimental methods and materials**

77 **2.1 Preparation of biochar**

78 Maple wood biochars (20% sugar maple, 80% red maple) which were pyrolyzed at
79 300, 400, 500, 600, 700°C using a modified muffle furnace (Thermo Scientific, Waltham,
80 MA, USA) under argon atmosphere (sweep of 1 L min⁻¹), were ground and sieved to
81 between 149 and 850 µm to obtain a uniform particle size. The residence time in the
82 reactor was 30 min and the heating rate was 2.5°C min⁻¹. Another batch of maple wood
83 biochars was pyrolyzed at 500°C with residence times of 5 min, 30 min, 60 min, 120 min,
84 400 min and 800 min. The physical and chemical properties of biochars are shown in
85 Table 1.²⁷ All feedstocks were dried at 60°C to ca. 10% moisture (w/w) prior to
86 pyrolysis. The glassware and PE centrifuge tubes were acid washed in a hydrochloric
87 acid bath (10%) and rinsed with deionized (DI) water before use. The biochars were
88 oxidized by using H₂O₂ (30% v/v) for two weeks at 30°C, all using a solid-to-liquid ratio
89 of 1:10 (w/v), which we found to result in significant changes in surface charge of
90 biochar produced at 500°C for 30 min (Wang et al., 2015). H₂O₂ was chosen in order to
91 minimize precipitation, complexation or analytical interference which have been
92 observed with other oxidants (e.g., H₂SO₄, HNO₃).²⁸ After oxidation, the H₂O₂ was
93 removed by filtration under suction using a Büchner funnel, fitted with Whatman No.1
94 filter paper, attached to a Büchner flask connected to a Welch Duo-Seal 1400 vacuum
95 pump and the oxidized biochars were rinsed with DI water. The pH values of biochars
96 were determined using a glass electrode (detection limit is 0.01 pH units) with a biochar-

97 to-water ratio of 1:20 (w/v) (Orion 3-Star pH Benchtop; Thermo Electron Corporation,
98 Beverly, MA, USA). After determining the initial pH values of the oxidized biochar
99 samples, the biochars were dried under 60°C for 48 h, and then separated into two
100 identical batches. One batch was utilized as it remained after oxidation and drying; for
101 the other, hydrochloric acid (1.0 M) or sodium hydroxide solution (1.0 M) was used to
102 adjust the pH values to 7.00±0.01, respectively. The pH values were adjusted every 24 h
103 until they reached equilibrium, then suspensions were filtered through Whatman No.1
104 filter paper.

105 The specific surface area (SSA) and pore size distribution of the biochars were
106 evaluated using the ASAP 2020 - Physisorption Analyzer (BET) CO₂ adsorption
107 technique at 273.15 K. Determination of the CEC of biochar is based on the method by
108 Page.²⁹ Elemental H, O, C of biochars were determined on a Temperature Conversion
109 Elemental Analyzer (TC/EA). Proximate analysis was conducted using ASTM D1762-84
110 Chemical Analysis of Wood Charcoal after modification to accommodate biochar
111 reactivity.¹⁶

112

113 2.2 Adsorption experiments

114 Batch adsorption experiments were conducted in PE centrifuge tubes at room
115 temperature. All adsorption experiments were performed in triplicate. 0.5 g of biochar
116 was added into 50-mL centrifuge tubes containing 40 mL of ammonium sulfate solution
117 with 100 mg NH₄-N L⁻¹. The tubes were shaken at 400 rpm in a mechanical shaker for
118 16 h and then centrifuged at 2500rpm for 10 min, and the supernatant was carefully
119 aspirated using a Pasteur pipette. Ammonium N concentrations in the supernatant were

120 determined immediately using a continuous flow analyzer (Bran and Luebbe
121 Autoanalyzer, SPX, Charlotte, NC).

122

123 2.3 Desorption experiments

124 After the adsorption, the solution remaining in the tubes was decanted and 40 mL of
125 ultrapure water ($> 18.2 \text{ M}\Omega\cdot\text{cm}$) which was prepared by a Barnstead E-pure water
126 purifier was added as described for the adsorption experiment described above. This
127 procedure was repeated twice, generating two desorption steps. Ammonium N
128 concentrations in the supernatant were determined within 24 h by using a continuous flow
129 analyzer (Bran and Luebbe Autoanalyzer, SPX, Charlotte, NC).

130

131 2.4 Statistical analysis

132 The results of adsorption and desorption experiments were the average of three
133 replications. The statistical software package SPSS 13.0 (Chicago, IL, USA) was used for
134 descriptive statistics, ANOVA testing (post-hoc LSD analysis at $p < 0.05$) of ammonium
135 adsorption and desorption at different pyrolysis temperatures and residence times. In
136 order clarify the data present in Table 1, we performed regression analyses and added p
137 values into the table.

138

139 3. Results

140 3.1 Biochar properties

141 Increasing the pyrolysis temperature had great effects on proton activity due to
142 decreases in acid functional groups with greater pyrolysis temperature. Fresh biochar pH

143 values ranged from 6.94 at 300°C to 10.20 at 700°C. Biochars produced at higher
144 temperature had high alkalinity, and those at lower temperature had lower alkalinity.
145 After oxidation, the pH dropped to 2.87 and 2.69 for biochars produced at 300°C and
146 400°C, respectively (Fig. 1a). During pH adjustment using NaOH, all of the oxidized
147 biochars produced at these low temperatures dissolved, and no adsorption and desorption
148 data could be obtained. This may be explained by the reaction between sodium hydroxide
149 and acidic functional groups, such as phenolic hydroxyl and carboxyl groups. In contrast
150 to the low-temperature biochars, oxidation had lower effects on pH decrease for biochar
151 produced at higher temperature. After oxidation of biochar produced at 700°C, pH values
152 dropped from 10.2 to only 7.63.

153 In contrast to pyrolysis temperature, increasing the residence time from 5 min to 800
154 min at 500°C had no significant effect on pH values of the unoxidized biochar, which
155 ranged from 8.82 at 5 min to 9.08 at 800 min (Fig. 1f). Interestingly, oxidation
156 significantly reduced the pH of biochars with short residence times, but had little effect
157 on pH of biochar pyrolyzed for 400 min or longer. When oxidized, varying the residence
158 time from 5 min to 80 min at a pyrolysis temperature of 500°C had almost identical
159 effects on pH as varying the pyrolysis temperature from 300°C to 700°C at a residence
160 time of 30 min. This may be explained by the aromatization level of biochar. The high
161 temperature and long residence time makes the carbon highly aromatic.

162 Ash contents significantly increased with higher pyrolysis temperature and residence
163 time. In contrast to ash, volatile contents decreased significantly with higher pyrolysis
164 temperature, when the residence time increased from 60 min to 800 min, the change of
165 volatile contents was not obvious. Fixed carbon contents of biochars significantly

166 increased with increasing pyrolysis temperature, but no significant increase was observed
167 when prolonging pyrolysis. The surface area also increased significantly with increasing
168 pyrolysis temperatures, but there was no alteration when pyrolyzed continuously at
169 500°C. Both pyrolysis temperature and residence time had significant effects on the CEC
170 of biochar. With the increase in pyrolysis temperature, the CEC of biochar decreased
171 from 117.15 cmol kg⁻¹ to 47.11 cmol kg⁻¹. Total C contents increased while total H and O
172 decreased with increasing temperature (Table 1).

173 Biochar yields decreased with increasing pyrolysis temperatures and residence times
174 Increasing the pyrolysis temperature resulted in a significant decrease in biochar yield,
175 and the highest yield was obtained at a temperature of 300°C. For instance, the yield of
176 biochar at 300°C was 62% of the dry feedstock, while at 400°C, the yield decreased to 32%
177 of the dry feedstock. At temperatures of 500, 600 and 700°C, the yield of biochar was
178 reduced to 27%, 26% and 24% of the initial weight, respectively (Fig. 1b). The largest
179 yield loss occurs within the first 120 min of residence time during pyrolysis. Increasing
180 the residence time from 120 min to 800 min decreased the biochar yield only slightly (Fig.
181 1g).

182

183 3.2 Ammonium adsorption

184 In general, fresh biochars had lower ammonium adsorption capacity than oxidized
185 biochars; when oxidized, pH-adjusted biochars had higher adsorption capacity than not
186 pH-adjusted biochars; and oxidized and pH-adjusted biochars pyrolyzed at lower
187 temperatures or for shorter period of residence times had higher adsorption capacity than
188 those pyrolyzed at higher temperatures or for longer period of times. For the fresh

189 biochars, there was no relevant difference among different pyrolysis temperatures. The
190 adsorbed ammonium for the fresh biochars at 300°C and 400°C was 0.87 mg g⁻¹ and 0.58
191 mg g⁻¹, respectively, and for those pyrolyzed at 500°C, 600°C, 700°C 0.52 mg g⁻¹, 0.46
192 mg g⁻¹, and 0.46 mg g⁻¹, respectively (Fig. 1c). When oxidized, biochars pyrolyzed at
193 lower temperatures (300°C, 400°C) had higher adsorption capacity than those at higher
194 pyrolysis temperatures (500°C, 600°C, 700°C). After the pH adjustment, ammonium
195 adsorption significantly increased at lower pyrolysis temperatures ($\leq 500^\circ\text{C}$), but did not
196 change for higher pyrolysis temperatures ($\geq 600^\circ\text{C}$).

197 There was no significant difference in ammonium adsorption for fresh biochars
198 among different residence times. However, neutralizing the oxygen-containing surface
199 functional groups on oxidized biochars to pH 7 increased ammonium adsorption two to
200 three-fold for biochars pyrolyzed from 5 min to 120 min (Fig. 1h).

201

202 3.3 Ammonium desorption

203 In general, ammonium desorption with lower pyrolysis temperature and shorter
204 pyrolysis times was higher than with higher temperature and longer times (Fig. 1d and i).
205 pH adjustment significantly affected the recovery: non-desorbed ammonium remained
206 near 100% if the pH was adjusted, but decreased to 80% if the pH was not adjusted,
207 irrespective of oxidation (Fig. 1e and j). Oxidation did not have a discernable effect on the
208 proportion of desorbed ammonium.

209

210 4. Discussion

211 4.1 Effect of oxidation and pH on ammonium retention

212 Freshly produced biochars typically have very low ability to adsorb ammonium,
213 showing low cation exchange capacity.³⁰ The increase in ammonium adsorption through
214 oxidation by H₂O₂ corroborate the observations from naturally aged charcoals²³ and
215 incubation studies.²⁴ The unchanged ammonium adsorption despite greater oxidation
216 without pH adjustment may be explained by protons and possibly aluminum and reduced
217 iron or other metals that dominate the exchange sites at very low pH values.²⁰ Even
218 carbon-rich wood-based biochars as those investigated here, possess appreciable amounts
219 of ash (1.7%) which contain metals.¹⁶ On the other hand, adjusting the pH to a common
220 pH of 7, significantly increased adsorption with increasing oxidation which corroborate
221 the observations from our previous work.³¹ The reason for this is that after the pH
222 adjustment, the carboxyl and phenolic groups were deprotonated as well as free
223 aluminum and iron may have precipitated as oxides and the negatively charged organic
224 functional groups became the main adsorption sites for ammonium.

225 Abiotic oxidation with peroxide generated oxygen-containing functional groups,
226 which largely improve ammonium adsorption capacity. Another reason for the
227 enhanced ammonium adsorption capacity is the increase of negative charges on the
228 surface of the biochar.³¹ Higher pyrolysis temperatures and longer residence times result
229 in higher degrees of biochar aromaticity. Elemental H/C and O/C ratios can be used to
230 estimate the degree of aromaticity of the carbon structure. Biochars produced at low
231 temperatures had higher H/C ratios, and biochars produced at high temperatures had
232 lower H/C ratios. Previous studies have shown that biochar amendment can increase CEC
233 in soil³² and that when aging and weathering of the biochar occurs the CEC can be
234 further increased.³³ Therefore, the oxygen-containing functional groups and negative

235 charges in the oxidized biochar pyrolyzed under low temperatures and short residence
236 times are higher than those pyrolyzed under high temperatures and long residence times.
237 Previous studies have also pointed out that, initially, biochars produced at high pyrolysis
238 temperatures (800°C) had greater exchangeable cations and CEC than those produced at
239 lower temperatures. However, after 8 weeks of oxidation the CEC of low temperature
240 biochars was significantly higher,³⁴ which is consistent with the results of this study.

241 4.2 Pyrolysis conditions and ammonium nitrogen retention

242 Without oxidation, varying the pyrolysis residence time beyond 5 min had no effects
243 on ammonium adsorption, corollary with a lack of change in pH, elemental ratios, or
244 surface area. The lack of an increase in surface area may be explained by the observation
245 that prolonging heat treatment at 500°C may lead to increased softening of some volatile
246 fractions, forming an intermediate melt which closes and seals off some of the pores,
247 compensating for any creation of pores with longer pyrolysis times.

248 However, the lack of change in ammonium adsorption to unoxidized biochars as a
249 function of pyrolysis temperature is more difficult to explain, as pH, O/C ratios, and
250 surface areas suggested a change in biochar properties. It is possible that the decrease in
251 acid functional groups with increasing pyrolysis temperatures is compensated by an
252 increase in surface area, canceling any temperature effects for the biochars studied here.
253 Increases in surface area with higher pyrolysis temperature are typically observed,¹⁴ as
254 are lower O/C ratios.¹⁶

255 Interestingly, when biochars were oxidized, both residence time and pyrolysis
256 temperature affected ammonium adsorption, and did so to a much greater extent than if
257 they were not oxidized. The much greater adsorption to oxidized biochars produced at

258 lower temperatures and shorter residence times may be the result of greater oxidation,
259 shown by a greater pH drop and increases in the O/C ratios. Greater oxidation with
260 shorter pyrolysis times and lower temperatures upon exposure to H₂O₂ may be explained
261 by the lower degree of fused aromatic C structures typically found under those
262 conditions,^{35,36} also indicated by the higher H/C ratios found in our study. This probably
263 also means that those biochars produced at lower temperatures and shorter residence
264 times will develop cation retention capacity more quickly when added to soil. This may
265 pose a tradeoff with an intent to produce biochars with long turnover times in soil, which
266 are typically produced at higher pyrolysis temperatures.³⁷

267

268 **5. Conclusions**

269 Freshly produced biochars have low ability to adsorb ammonium. Only pyrolysis
270 temperature but not residence times affected oxidation, pH, surface area, and yield of
271 fresh biochar, but none of the two affected the adsorption capacity of fresh biochar.
272 Short-term abiotic oxidation significantly increased ammonium retention of biochars, and
273 under these oxidized conditions, also pyrolysis residence times had significant effects.
274 The greater ammonium adsorption for less pyrolyzed biochars when oxidized may
275 constitute a tradeoff with maximizing biochar persistence where both are desired. Future
276 research should examine whether this also holds for effects of pyrolysis residence times
277 on biochars oxidized over time when they are added to soils.

278

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292

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352

353

Table 1. Physical and chemical properties of biochars.

Process conditions		pH	Proximate analysis (wt.% dry)				Elemental contents					Surface area (m ² g ⁻¹)	CEC (cmol kg ⁻¹)
Pyrolysis temperature (°C)	Residence time (min)		Volatile matter	Ash content	Fixed carbon	C _{tot} (w%)	H (w%)	O (w%)	H/C _{tot} (mol mol ⁻¹)	O/C _{tot} (mol mol ⁻¹)	BET CO ₂		
Original biochar													
300	30	6.94	68.5	0.54	31.0	58.9	5.22	32.9	1.06	0.42	72.4±1.5	117.2±2.3	
400	30	7.60	30.3	1.75	67.9	78.8	3.53	17.0	0.54	0.16	182±1.9	98.6±1.6	
500	30	8.90	21.1	2.16	76.7	91.5	2.81	9.51	0.37	0.08	250±2.0	78.1±1.4	
600	30	9.09	13.4	2.29	84.3	88.9	2.34	5.56	0.32	0.05	298±2.0	72.5±0.9	
700	30	10.2	8.44	2.59	89.0	94.9	1.33	3.92	0.17	0.03	337±2.1	47.1±0.7	
P-value		0.003	0.034	0.029	0.034	0.042	0.005	0.023	0.027	0.049	0.004	0.002	
500	5	8.82	24.3	1.65	74.0	83.3	2.98	11.9	0.43	0.11	222±2.0	122.7±1.8	
500	60	8.84	18.5	1.93	79.6	88.8	2.89	9.18	0.39	0.08	266±2.1	114.1±1.1	
500	120	8.67	17.7	2.04	80.2	86.2	2.79	8.23	0.39	0.07	272±2.1	104.1±0.7	
500	400	8.70	14.5	2.34	83.2	91.6	2.67	7.96	0.35	0.07	294±2.1	93.0±0.6	
500	800	9.08	14.4	2.58	83.0	84.5	2.54	7.06	0.36	0.06	274±1.9	70.7±0.4	
P-value		0.227	0.129	0.016	0.150	1.000	0.013	0.140	0.133	0.169	0.355	0.004	
Oxidized biochar													
300	30	2.87	-	-	-	49.4	5.23	44.3	1.27	0.67	61.5±1.9	143.2±1.9	
400	30	2.69	-	-	-	59.5	3.03	34.6	0.61	0.44	132±1.8	125.8±1.5	
500	30	4.01	-	-	-	73.4	2.80	19.7	0.46	0.20	243±2.0	122.7±1.5	
600	30	7.00	-	-	-	84.0	2.02	8.36	0.29	0.07	287±1.9	117.8±2.1	
700	30	7.63	-	-	-	93.3	1.32	5.33	0.17	0.04	319±2.0	110.9±2.8	
P-value		0.017				0.000	0.015	0.003	0.024	0.008	0.004	0.013	
500	5	3.10	-	-	-	67.3	2.53	25.6	0.45	0.28	174±1.8	131.4±3.1	
500	60	3.94	-	-	-	76.2	2.61	18.6	0.41	0.18	246±1.9	124.6±2.9	
500	120	4.45	-	-	-	77.1	2.50	15.8	0.39	0.15	255±2.0	113.4±1.4	
500	400	7.12	-	-	-	80.5	2.47	12.6	0.37	0.12	267±1.9	96.7±1.1	
500	800	6.99	-	-	-	83.4	1.95	10.1	0.28	0.09	269±1.9	84.9±0.7	
P-value		0.046				0.089	0.021	0.074	0.007	0.104	0.251	0.011	

a C_{tot} is the total carbon of biochar.

b P-value for a linear regression of the effect of pyrolysis temperature and residence time.

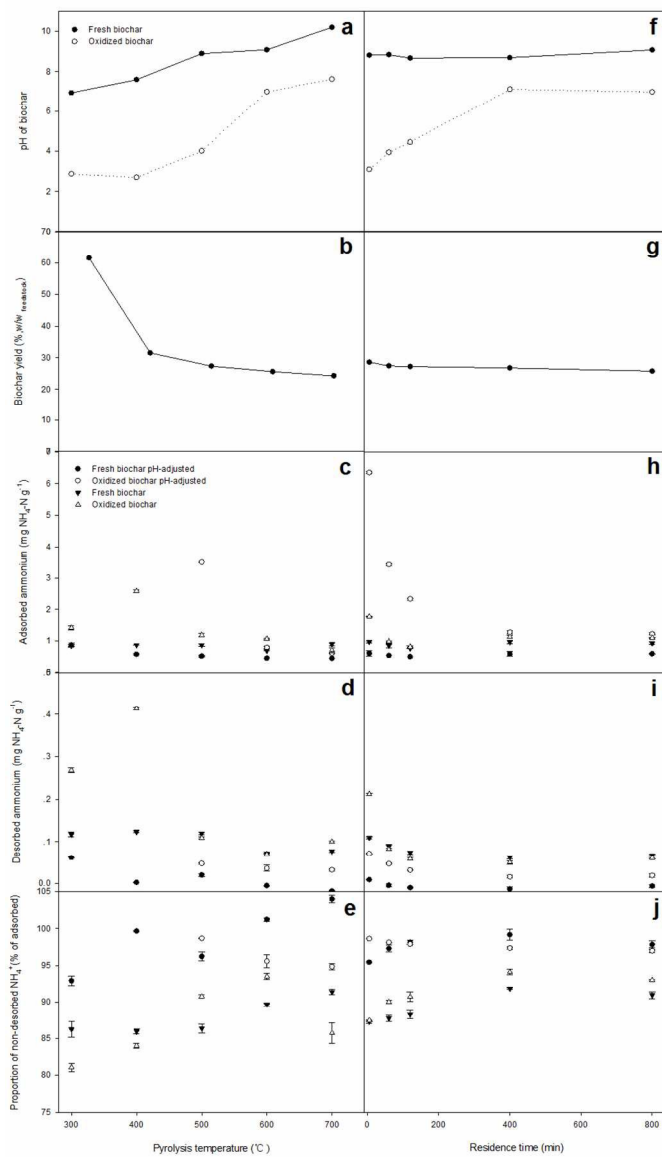


Fig. 1. The effect of pyrolysis temperature (a-e) or residence time in the pyrolyser (g-j) on pH, biochar yield, ammonium adsorption, ammonium desorption and the proportion of non-desorbed ammonium (adsorption and desorption of oxidized and pH-adjusted biochars at low temperatures could not be studied due to their high solubility). Error bars represent standard error of triplicate samples ($n = 3$). Symbols may cover error bars.

171x295mm (150 x 150 DPI)