

This is an *Accepted Manuscript*, which has been through the Royal Society of Chemistry peer review process and has been accepted for publication.

Accepted Manuscripts are published online shortly after acceptance, before technical editing, formatting and proof reading. Using this free service, authors can make their results available to the community, in citable form, before we publish the edited article. This Accepted Manuscript will be replaced by the edited, formatted and paginated article as soon as this is available.

You can find more information about *Accepted Manuscripts* in the **Information for Authors**.

Please note that technical editing may introduce minor changes to the text and/or graphics, which may alter content. The journal's standard <u>Terms & Conditions</u> and the <u>Ethical guidelines</u> still apply. In no event shall the Royal Society of Chemistry be held responsible for any errors or omissions in this *Accepted Manuscript* or any consequences arising from the use of any information it contains.



www.rsc.org/advances

Effect of humic acid on prometryn bioaccumulation and the induction of oxidative stress in zebrafish (*Danio rerio*)

Qian Zhao, Lin Zhu*

MOE Key Laboratory of Pollution Processes and Environmental Criteria / Tianjin Key Laboratory of Environmental Remediation and Pollution Control / College of Environmental Science and Engineering, Nankai University, No. 94 Weijin Road, Nankai District, Tianjin 300071, China

*Corresponding Author: Phone: (86)22-23508807. Fax: (86)22-23508936; E-mail: zhulin@nankai.edu.cn.

RSC Advances Accepted Manuscript

Abstract: Humic acid (HA) is the main component of dissolved organic matter in aquatic ecosystems and affects the bioavailability of contaminants. Prometryn is frequently used to control annual broadleaf and grass weeds and has been widely detected in the aquatic environment. An experiment was conducted to investigate the effect of HA on the chronic toxicity of prometryn in zebrafish. In zebrafish treated with 53.2 μ g·L⁻¹ prometryn (P), prometryn with 5 mg·L⁻¹ HA (P+HA₅), or prometryn with 15 mg L^{-1} HA (P+HA₁₅), catalase (CAT) activity and glutathione (GSH) content in the visceral mass initially increased and then decreased with exposure time. After one day of exposure, malondialdehyde in the visceral mass increased by 58% to 64% in fish treated with prometryn and HA compared to those treated with prometryn alone, however, this difference disappeared during days 10 to 40. Treatment with HA enhanced the bioaccumulation of prometryn by 34% and 40% on day 1 in the P+HA₅ and P+HA₁₅ groups, respectively, possibly due to changes in gill membrane permeability. Nevertheless, the opposite result was observed during days 10 to 40, owing to the presence of more excreta and suspended particulate matter resulting in a reduction in the quantity of free dissolved prometryn in the P+HA₅ and P+HA₁₅ groups. There were significant correlations between CAT activity and prometryn concentration (0.562**, P < 0.01) and between GSH and prometryn concentration $(0.808^{**}, P < 0.01)$ throughout the first 20 days of the experiment, suggesting a role for these antioxidant systems in the detoxification of prometryn in zebrafish.

Keywords: prometryn; humic acid; zebrafish; oxidative stress; bioaccumulation

2

1. Introduction

Prometryn (2,4-bis[isopropylamino]-6-methylthio-s-triazine) is an s-triazine herbicide that is widely used to control many annual gramineous, broad-leaved, and perennial terrestrial weeds¹. In addition, prometryn is also used to remove filamentous algae, aquatic weeds, and other harmful algae in the fish, shrimp, crab, shellfish, and sea cucumber aquaculture industries. Prometryn interferes with photosynthesis by inhibiting electron transport². Prometryn has been prohibited in Europe since 2004³, yet it is still widely used in China⁴, the United States, Canada, New Zealand, and South Africa ⁵. Because of its high chemical stability, prometryn persists in the aquatic environment and has the potential to harm non-target organisms. Ren et al.⁶ detected prometryn in 100% of samples collected from 60 sites in the Bohai Sea, with the maximum concentration of prometryn reaching up to 7.12 ± 0.54 μ g L⁻¹ in a river in Shanghai⁷. Ma et al.⁸ measured the concentration of prometryn at Shanghai waterworks on the Huangpu River for one year and reported that the average concentration of prometryn was the highest in April (1.25 μ g·L⁻¹). Furthermore, the detected concentrations of prometryn ranged from 0.91 to 4.40 $\mu g L^{-1}$ in surface water and exceeded 1 $\mu g L^{-1}$ in ground water in Greece ⁹. Since its stability and persistence, prometryn is also widely residual in the aquatic environment.

Dissolved organic matter (DOM) participates in and is an important regulator of the processes of biogeochemical conversion, which include global nutrient and carbon

RSC Advances Accepted Manuscript

cycling, metal redox reactions, and cation complexation 10 . The concentration of DOM varies greatly in different water bodies and is also dependent upon the season and water depth. Typically, the concentration of DOM ranges from 0 to 15 mg C L^{-1} (denoted as the dissolved organic carbon) in surface water ¹¹. Humic acid (HA) is the main component of DOM in aquatic ecosystems ¹². Previous studies have reported that HA decreases the bioavailability and toxicity of organic contaminants through binding and adsorption ^{13, 14}. However, other studies have indicated that HA results in higher bioavailability and toxicity of organic contaminants by increasing the compound solubility ^{15, 16} or changing fish gill membrane permeability ^{17, 18}. Nevertheless, the mechanism by which HA increases the biological toxicity of organic pollutants remains unclear. Matsuo et al.¹⁵ reported that both P450 1A (CYP 1A) expression and 7-ethoxyresorufin-O-deethylase (EROD) activity in the liver of tambaqui (Colossoma macropomum) were induced 2-fold by crude oil in combination with HA versus crude oil alone. Moreover, previous studies have indicated that DOM enhances Na⁺ flux in Daphnia magna and the gills of fish ^{16, 19} due to increased membrane permeability as a result of DOM adsorption $^{18, 20}$, demonstrating that DOM has direct effects on these organisms ^{21, 22}. Galvez et al. ¹⁷ suggested that the gill epithelium of rainbow trout (Oncorhynchus mykiss) developed a hyperpolarized transepithelial potential (TEP) following exposure to HA (10 mg \cdot L⁻¹). This phenomenon is likely the result of a reduction in the concentration of certain ions, such as Ca^{2+} , which could affect the biological membrane permeability.

Previous studies have mainly focused on the effect of prometryn alone on fish ^{23,}

 24 , however, the effect of HA on prometryn toxicity in fish has not been addressed. Our previous research demonstrated that 15 mg L⁻¹ of HA increased the acute toxicity of prometryn in *Danio rerio* (zebrafish) during a 96 h exposure ²⁵. An HA-accelerated bioaccumulation of prometryn in zebrafish was proposed as an explanation for this increased toxicity. However, this result was inconsistent with other reports ²⁶⁻²⁸, in which HA was observed to reduce the bioavailability of organic contaminants and the associated toxicity.

In this study, a chronic exposure experiment was conducted to investigate the toxicity of prometryn in the presence of HA. It was hypothesized that the chronic toxic effects of prometryn combined with HA would be consistent with the acute effects that were previously observed. Indicators of oxidative stress (malondialdehyde [MDA] content, catalase [CAT] activity, and glutathione [GSH] content) and the bioaccumulation of prometryn were investigated in zebrafish exposed to a combination of prometryn and HA. Additionally, the gill Na⁺/K⁺-ATPase activity and the electrical conductivity of the exposure solution were also evaluated. This work was aimed at providing strong evidence for the enhanced toxicity of an organic contaminant in the presence of DOM in aquatic ecosystem.

2. Materials and methods

2.1. Animals

Zebrafish (body length of 4.6 ± 0.6 cm, weight of 0.49 ± 0.09 g) were bred and maintained in the laboratory in a tank ($95 \times 43 \times 45$ cm) containing dechlorinated tap

water (pH of 7.5 \pm 0.3, temperature of 24 \pm 2°C, dissolved oxygen concentration of 7.83 \pm 0.2 mg·L⁻¹, DOM << 1 mg·L⁻¹) that was continuously aerated. All fish were fed with a commercial fish food (Inch-gold Fish Food Limited Company, Shenzhen, China) at a rate of 1% of body weight per day. The excreta at the bottom of the tank were removed in time. The natural mortality rate was < 1% prior to the beginning of the experiment. Feeding was suspended for 24 h before the exposures. All animal experiments were carried out according to the "Measures for the Administration of Experimental Animals Permit" and the Guidelines of Science and Technology Committee, Tianjin, China.

2.2. Toxicity tests

Prometryn (purity > 97%) was purchased from Zhongshan Import and Export Corporation (Zhejiang, China). The stock solution of prometryn was prepared with dimethyl sulfoxide (DMSO) as the carrier solvent, sonicated for 30 min at 45°C, filtered through a 0.45 μ m membrane, and stored at 4°C. All experimental prometryn solutions were prepared by diluting the stock solution in water. The final concentration of DMSO in the water was always less than 0.05%. The prometryn standard curve was prepared using a prometryn standard substance (Aladdin, USA, purity > 99%, CAS: 7287-19-6). The properties of prometryn were shown in Table 1.

The stock solutions of HA (Sigma, St. Louis, MO, USA;CAS: 1415-93-6) were prepared by dissolving HA into deionized water, stirring for 24 h, standing for 12 h, and filtering through a 0.45 μ m membrane. The concentration of total organic carbon

(TOC, mg C \cdot L⁻¹), which corresponds to the concentration of HA, was measured using an Analytik Jena multi N/C 3100 (Jena, Germany).

Toxicity experiment 1: According to the 96 h LC₅₀ of 5.32 mg L⁻¹ prometryn for zebrafish obtained in a previous acute toxicity experiment, an exposure concentration equal to 1% of the prometryn LC₅₀ for zebrafish (53.2 μ g·L⁻¹) was selected for this study ²⁵. As in surface water HA concentrations are generally in the range of 0 to 15 mg C L^{-1 11}, the HA concentrations selected in this study were 5 and 15 mg L⁻¹ (HA₅ and HA₁₅, respectively). The five treatment groups consisted of the control (CK), 15 mg·L⁻¹ HA alone (HA₁₅), prometryn alone (P), P+HA₅ and P+HA₁₅. Duplicate glass tanks were prepared for each treatment group. Each glass tank (21×26×25 cm) was disinfected using a 2% (*m*/v) K₂MnO₄ solution. Each glass tank was filled with 10 L of exposure solution before 50 zebrafish were placed into the tank. Half of the exposure solution in each tank was renewed every five days during the 40-day experimental period. No dead zebrafish were observed during the exposure.

On days 1, 5, 10, 20, and 40, six zebrafish were randomly sampled from each group. Every three fish were dissected and pooled all organs together (visceral mass) for the analysis of CAT, GSH, and MDA. In addition, on days 1, 5, 10, 20, 30, and 40, six fish were collected, rinsed with deionized water, dried with filter paper, and stored at -80°C for future determination of prometryn bioaccumulation.

The prometryn concentration in the exposure solution was evaluated at 0, 1, 4, 8, 12, 24, 48, and 96 h, and also on days 5, 10, and 15. Furthermore, the zebrafish excreta was collected on days 5, 10, 15, 20, and 25 before renewing the exposure

solution. Following collection, the excreta were freeze-dried and weighed.

Toxicity experiment 2: In order to investigate the effect of HA on zebrafish gill cell membrane permeability, gill Na^+/K^+ -ATPase activity and the electrical conductivity of the exposure solution were determined. Six zebrafish were transferred to 2 L glass beakers containing 1 L of solution which was balanced for 48 h. The treatments and the experimental conditions were consistent with toxicity experiment 1. After one day, gill Na^+/K^+ -ATPase activity and the electrical conductivity of the exposure solution were measured.

2.3. Determination of prometryn concentrations

After freeze-drying for 48 h, the zebrafish were ground to a powder and then 0.2 g of the sample (mixed with 2 g of quartz sand) was extracted using accelerated solvent extraction (ASE) with a Speed Extractor E-916 (BUCHI, Switzerland). Acetonitrile (analytical grade; Kangkede, Tianjin, China) was the extraction solvent. The operating conditions of the ASE were as follows: extraction temperature = 100° C, pressure = 100 bar, heating time = 5 min, static extraction time = 8 min, extraction pool size = 20 mL, purging time = 120 s, and number of extraction cycles = 3. The extract was transferred to pear-shaped bottle and evaporated at 75° C using a rotary evaporator (Heidolph Laborota 4000 efficient; Heidolph, Schwabach, Germany). The residue was dissolved in 2 mL of a mixture of *n*-hexane and ethyl acetate (v/v = 3/2) and filtered through florisil columns (1 g, 6 mL; CNW Technologies, Düsseldorf, Germany). The collected filtrate was dried using a pressure blowing concentrator (Huaruibovuan

MTN-2800W; Huaruiboyuan, Beijing, China). The dried residue was dissolved in 2 mL of methanol and filtered through a 0.45 μ m organic membrane filter. Collected water samples were filtered through a 0.45 μ m organic membrane filter without any pretreatment.

Prometryn concentrations were determined using ultra performance liquid chromatography-tandem mass spectrometry (UPLC-MS) (OA_SPE Waters Xevo TQ-S; Waters Corporation, Milford, MA, USA). The UPLC-MS was performed on an ACQUITY UPLC^R BEH C18 column (1.7 μ m, 2.1×50 mm, Waters, USA) at a temperature of 55°C in a 75% / 25% mixture (ν/ν) of methanol (chromatography grade; Merck, Germany) (A) and 0.1% formic acid aqueous solution (chromatography grade; CNW Technologies; Milli - Q water) (B) at a flow rate of 0.3 mL·min⁻¹, using an injection volume of 10 μ L, a retention time of 0.79 min, and a pressure ripple of 3939-4922 psi. Mass spectrometry conditions were as follows: ion mode of ESI⁺, ionization voltage = 3.0 kV, cone voltage = 25 V, source temperature = 150°C, desolvation temperature = 350°C, cone gas flow = 144 L·h⁻¹, desolvation gas flow = 596 L·h⁻¹, collision gas flow = 0.14 mL·min⁻¹, and prometryn detection pair of 242 > 158, 242 > 200 (quantitative ion pair). The correlation coefficient of the prometryn standard curve obtained was > 0.99.

2.4. Index of toxicology

Six fish from each group were dissected on an ice-cold plate and the visceral mass was excised, washed immediately, and homogenized in a 0.86% physiological saline solution (1:9, w/v). The homogenate was centrifuged at 2,500 g·min⁻¹ for 15 min (Hettich Mikro 200R; Hettich, Tuttlingen, Germany). The supernatant fluid was diluted to appropriate concentrations for further analysis.

MDA content of the supernatant was analyzed using the thiobarbituric acid reactive substances (TBARS) method to evaluate lipid peroxidation. GSH and protein content, and the activities of CAT and Na⁺/K⁺-ATPase, were measured using commercial kits (Jiancheng Bioengineering Institute, Nanjing, China).

2.5. Statistical analysis

The results in the figures and tables were reported as the mean \pm standard error (SE). Statistical analysis of all data was performed using one-way ANOVA with SPSS 20 software. Multiple comparisons were conducted using Duncan's multiple range test and differences were considered significant when P < 0.05. Figures were constructed using Origin 8.5.

3. Results and discussion

3.1. Toxicological parameters

3.1.1. Antioxidant response in zebrafish

On day 1, CAT activity and GSH content in the visceral mass of zebrafish exposed to prometryn alone decreased by 54% and 64% relative to CK. Both CAT activity and GSH content in the visceral mass of zebrafish exposed to HA₁₅ were not significantly different from those in the CK group from days 1 to 20 (P > 0.05). The CAT activity

in the HA₁₅ group increased by 36% compared to the CK group only on day 40. In the P, P+HA₅, and P+HA₁₅ groups, CAT activity and GSH content in the visceral mass of zebrafish increased at first and then decreased during days 1 to 40 (Figure 1 and Figure 2; the data for GSH on day 5 in Figure 2 were lost due to operational errors). On day 1, in the P+HA₁₅ group, CAT activity and GSH content enhanced by 162% and 142%, respectively, compared to those exposed to prometryn alone (P < 0.05, Figure 1 and Figure 2). On the contrary, during days 10 to 40, CAT activity and GSH content in the P+HA₅ and P+HA₁₅ groups were lower than those in the P group (Figure 1 and Figure 2). There was no significant difference in CAT activity or GSH content between the $P+HA_5$ and $P+HA_{15}$ groups during the 40-day experiment, with the exception of day 1 (P < 0.05).

3.1.2. Oxidative stress in zebrafish

The MDA content in the visceral mass of zebrafish increased by 58% and 64% in the P+HA₅ and P+HA₁₅ groups, respectively, relative to the P group on day 1. However, there was no significant difference in MDA content among all groups during days 10 to 40 (P > 0.05, Figure 3). In addition, MDA content in the visceral mass of zebrafish exposed to prometryn alone was not significantly different from that in the CK group throughout the 40 days exposure (P > 0.05, Figure 3).

3.1.3. Na⁺-K⁺-ATPase activity in zebrafish

Biological membrane permeability was estimated by measuring gill Na⁺-K⁺-ATPase activity. The activity of Na⁺-K⁺-ATPase in the gills of zebrafish exposed to HA₅ and HA₁₅ increased by 28% and 19%, respectively, relative to the CK group; the difference between the HA₅ and CK groups was significant (P < 0.05). Gill Na⁺-K⁺-ATPase activities in the P+HA₅ and P+HA₁₅ groups were higher than that in the P group (Figure 4).

3.2. Prometryn accumulation in zebrafish

The recovery of prometryn from zebrafish was between 90% and 115%. Prometryn was not detected in zebrafish from the CK or HA₁₅ treatment groups throughout the 40 days of exposure. On days 1 and 5, prometryn accumulation in zebrafish in the P+HA₅ group increased by 34% and 8%, respectively, while in the P+HA₁₅ group it increased by 40% and 21%, respectively, compared to those in the P group. However, prometryn accumulation in the P+HA₅ and P+HA₁₅ groups decreased by 37% and 64%, respectively, relative to the P group during days 10 to 40 and these differences were significant (P < 0.05, Figure 5).

3.3. Prometryn concentrations in the exposure system

The decrease of prometryn concentration in the exposure solution was less than 10% after five days in all groups. The exposure concentration of prometryn in the P, P+HA₅, and P+HA₁₅ groups gradually decreased over time. Prometryn concentrations in the HA groups were lower than that in the groups without HA, and this difference increased gradually over time (Figure 6). On day 15, prometryn concentrations in the

P+HA₅ and P+HA₁₅ groups decreased by 8% and 12% relative to the P group, respectively (Figure 6). It was noteworthy that an opposite trend was found in the dry weight of fish excreta; the dry weight of fish excreta increasing over time in the HA groups compared to the non-HA groups. On day 25, the dry weight of fish excreta in the P+HA₁₅ group was 1.96 times higher than that in the P group (Figure 7).

3.4. Correlation analysis

Fish metabolic processes result in the production of reactive oxygen species in the presence of stressors such as herbicides, including prometryn, which leads to oxidative stress and damage to the fish (e.g. lipid peroxidation, as indicated by MDA). In response this stress, the antioxidant defense system (CAT and GSH) is activated to clear the reactive oxygen species in order to maintain homeostasis. Accordingly, MDA, CAT, and GSH are typically used as indicators of oxidative stress in order to evaluate the biological toxicity of environmental contaminants. In the P, P+HA₅, and P+HA₁₅ groups, the changes in the trends of CAT activity and GSH content in the visceral mass were coincident with those of prometryn accumulation in zebrafish, suggesting a role for these factors in the process of prometryn detoxification. Pearson correlation analysis indicated that there were significant correlations between prometryn accumulation and CAT activity (0.562^{**} , P < 0.01) and GSH content $(0.808^{**}, P < 0.01)$ during the first 20 days of the experiment (Table 2). Antioxidant parameters are meaningful for the investigation of the presence of, and the response to, certain environmental contaminants that are easily degraded and thus difficult to

1 490 17 01

measure. Therefore, these parameters can be chosen as indicators to evaluate the toxicity of contaminants during short-term exposures.

In the early stages of the exposure (day 1), the content of MDA in the visceral mass of zebrafish in the $P+HA_5$ and $P+HA_{15}$ groups were higher than those in any other group. This suggested that damage as a result of lipid peroxidation was enhanced by HA in fish exposed to prometryn. This result was consistent with that of our previous experiment in which the 96 h acute lethality of prometryn in zebrafish was increased by treatment with 15 mg L^{-1} HA 25 . Meanwhile, prometryn accumulation in fish from the P+HA₅ and P+HA₁₅ groups increased by 34% and 40%, respectively, compared with those in the P group. Correspondingly, GSH content and CAT activity also increased on the first day of exposure, indicating that GSH and CAT play roles in the detoxification of prometryn in the early stages of exposure. Nevertheless, during days 10 to 40, prometryn accumulation, GSH content, and CAT activity in the P+HA₅ and P+HA₁₅ groups decreased relative to the P group. However, MDA levels in the zebrafish did not change significantly during this time. A possible explanation for these changes is homeostatic processes that zebrafish set up the detoxification mechanisms for defense and discharge of prometryn. Eventually, the physiological metabolism processes of the fish would reach a new level of homeostasis through these mechanisms. In addition, the concentration of free dissolved prometryn gradually decreased in the exposure solution, possibly a result of the adsorption of prometryn to the remaining fecal matter and other suspended particulate matters, which may have also eased the level of lipid peroxidation

experienced by the zebrafish.

3.5. Effect of HA on prometryn toxicity in zebrafish

Humic acid is an important type of DOM in aquatic ecosystem and influences the environmental behavior of hydrophobic organic contaminants according to a variety of functional groups such as hydroxyl, carboxyl, phenolic hydroxyl, and enol hydroxyl groups ^{29, 30}. Previous studies have demonstrated that pollutant bioaccumulation and toxicity, in addition to organismal antioxidant capacity, were modified by HA^{26, 31}. On days 1 and 5, prometryn accumulation in zebrafish from the groups treated with HA (P+HA₅ and P+HA₁₅) increased compared to the group exposed to prometryn alone. This may have been due to changes in gill membrane permeability caused by HA^{17,18}. Steinberg et al. ³² suggested that DOM enhanced the bioaccumulation of terbuthylazine (TBA) in fish by interfering with the permeability of the cell membrane. In this study, gill Na⁺-K⁺-ATPase was higher in the P+HA₅ and P+HA₁₅ groups than in the P group, indicating that the permeability of the gill cell membrane was altered by HA. Galvez et al.¹⁷ reported that there was a direct effect on ion transfer and penetration function in the gills of rainbow trout treated with 10 mg C L^{-1} HA due to the enhancement of transepithelial hyperpolarization through the complexation of Ca^{2+} . A reduction in Ca^{2+} concentration, therefore, can lead to changes in the permeability of the gills and affect other ions such as Cl⁻ and Na⁺. Loice et al. ¹⁸ indicated that the permeability of a simulated biological membrane was increased by HA. Other scholars have also suggested that the presence of HA resulted

in changes in sodium metabolism ³³ or promoted the resistance to adverse environmental conditions by regulating Na⁺ flux in fish gills ¹⁷. The electrical conductivity of the exposure solution was measured in the current experiment as there is a close relationship between the permeability of a biological membrane and the ion content of the environment. The results showed that the electrical conductivity of the exposure solution decreased in the presence of HA. Compared with CK, a 7% decrement of electrical conductivity was observed in the solution of P+HA₁₅ (Figure 8). A decreased ion concentration in the exposure solution in the presence of HA likely contributed to changes in the permeability of the biological membranes of the zebrafish.

Xia et al. ³⁰ reported that HA (1 mg·L⁻¹) enhanced the bioaccumulation of perfluoroalkyl substances (PFAS) in *D. magna* by increasing the rate of uptake above the rate of depuration. In addition, the molecular mass of DOM also affects the bioaccumulation of pollutants in organisms. Hudson et al. ³⁴ suggested that high molecular weight (> 1 kDa) of DOM increased the absorption of dissolved cadmium, silver, and mercury in zebra mussels (*Dreissena polymorpha*) compared with low molecular weight of DOM (< 1 kDa).

When the exposure time was prolonged, prometryn bioaccumulation in zebrafish in the three treatment groups (P, P+HA₅ and P+HA₁₅) was changed in the presence of HA. During days 10 to 40 of the current experiment, prometryn bioaccumulation in zebrafish in the group without HA (P) was reduced compared to the groups treated with HA (P+HA₅ and P+HA₁₅). Two potential explanations for this result included an

increased metabolic rate and increased depuration dose of prometryn in zebrafish. Previous studies have confirmed that HA can be used as an indirect or direct energy source by fish and enhance the lifespan and fertility of Daphnia magna under extreme circumstances²². In addition, increased fish activity has been reported following treatment with HA³². A higher quantity of excreta and suspended particulate matter were formed in the exposure solution containing fish treated with HA (P+HA₅ and P+HA₁₅), suggesting a higher metabolic rate among these fish. In the exposure system, the excreta and suspended particulate matter may have interacted with HA and prometryn, resulting in the adsorption of prometryn and thus causing a decrease in the concentration of prometryn in the exposure solution. This reduction of prometryn in the exposed system would eventually reduce the uptake of prometryn by zebrafish. This relationship between prometryn concentrations and quantity of fish excreta was supported by correlation analysis, in which excreta weight was significantly negatively related with the prometryn concentration in the exposure solution (-0.815*, P < 0.05) (Table 2). Furthermore, the gradual adaptation of zebrafish to the exposure environment may also have resulted in the formation of the corresponding detoxification mechanisms. This mechanism may have resulted in an increase in the depuration dose of prometryn in zebrafish that eventually became greater than the uptake dose, especially among fish in the groups treated with HA.

3.6. Biological effects of HA

A wide variety of physiological effects have been reported as a result of DOM

exposure, including changes in biological membrane permeability owing to surface adhesion ^{17, 18}, induction of heat shock protein expression ³⁵, and activation of glutathione-S-transferase (sGST)³⁶ and cytochrome oxidase (CYP 1A)¹⁵. The source, concentration, and molecular weight of DOM also exert an important influence on the response of aquatic organisms. Matsuo et al.¹⁵ found that CYP 1A expression in tambaqui was induced by HA (20 to 80 mg \cdot L⁻¹) and that commercial preparations of HA were more effective than natural organic matter in eliciting this response. It is likely that HA includes certain components that may function as aryl hydrocarbon receptor agonists and result in the induction of CYP 1A expression. Wiegand et al.²⁶ reported that sGST was increased in *Lumbriculus variegatus* following treatment with 25 mg L^{-1} natural DOM, but not with 5 mg L^{-1} natural DOM, while peroxidase activity obviously increased at both concentrations of DOM. In the current study, there was little effect on the antioxidant system in zebrafish exposed to 15 mg L^{-1} HA alone, and the reasons for this need to be studied further.

4. Conclusion

In the early stages of HA and prometryn exposure (day 1), an increased bioaccumulation of prometryn resulted in an increased level of lipid peroxidation in zebrafish. However, during days 10 to 40 of prometryn and HA exposure, HA decreased the uptake of prometryn by zebrafish. During this time HA appeared to alleviate the toxic effects of prometryn exposure on zebrafish. Furthermore, the effects of HA in combination with prometryn were more pronounced at 15 mg \cdot L⁻¹ HA

than at 5 mg·L⁻¹ HA. Based upon the results of this experiment, CAT activity and GSH content are involved in the detoxification of prometryn in zebrafish. Further research is needed to understand the molecular mechanisms of the effects of HA on the toxicity of prometryn in zebrafish.

Acknowledgments

This work was financially supported by Major Science and Technology Program

for Water Pollution Control and Treatment (grant No. 2012ZX07501-003).

References

1. L. Jiang, L. Ma, Y. Sui, S. Han and H. Yang, Journal of Environmental Monitoring, 2011, 13, 1935-1943.

2. J. Zhou, X. Li, Y. Jiang, Y. Wu, J. Chen, F. Hu and H. Li, *Journal of Hazardous Materials*, 2011, 192, 1243-1249.

3. J. Zhou, F. Hu, J. Jiao, M. Liu, H. Li, Journal of Soils and Sediments, 2012, 12, 576-585.

4. J. Zhou, J. Chen, Y. Cheng, D. Li, F. Hu and H. Li, Talanta, 2009, 79, 189-193.

5. S. E. Kegley, B. R. Hill, S. Orme and A. H. Choi. PAN Pesticide Database, Pesticide Action Network, North America (San Francisco, CA, 2010).

6. C. Ren, X. Tian, H. Zhang, Y. Liu, Y. Sun, Y. Xu, X. Gong and M. Wang, *Journal of Chinese Mass Spectrometry Society*, 2013, 34, 353-361. (In Chinese)

7. Z. Li, L. Chen, H. Gao, L. Dong and J. Zhao, *Chinese Journal of Chromatography*, 2006, 24, 267-270. (In Chinese)

8. X. Ma, N. Gao, Q. Li, B. Xu, L. Le and J. Wu, China Watetr & Wastewater, 2006, 22, 1-4. (In Chinese)

9. Z. Vryzas, C. Alexoudis, G. Vassiliou, K. Galanis and M. E. Papadopoulou, *Ecotoxicology and Environmental Safety*, 2011, 74, 174-181.

10. C. E. Steinberg, N. Saul, K. Pietsch, T. Meinelt, S. Rienau and R. Menzel, *Annals of Environmental Science*, 2007, 1, 81-90.

11. E. M. Thurman, Springer Science & Business Media, 2012, Vol. 2.

12. X. Hu, L. Mu, J. Kang, K. Lu, R. Zhou and Q. Zhou, *Environmental Science & Technology*, 2014, 48, 6919-6927.

13. G. Chen, C. Lin, L. Chen and H. Yang, Chemosphere, 2010, 79, 1046-1055.

14. N. H. Song, L. Chen and H. Yang, Geoderma, 2008, 146, 344-352.

15. A. Y. Matsuo, B. R. Woodin, C. M. Reddy, A. L. Val and J. J. Stegeman, *Environmental Science & Technology*, 2006, 40, 2851-2858.

16. C. N. Glover and C. M. Wood, *Physiological and Biochemical Zoology*, 2005, 78, 1005-1016.

17. F. Galvez, A. Donini, R. C. Playle, D. S. Smith, M. J. O'Donnell and C. M. Wood, *Environmental Science & Technology*, 2008, 42, 9385-9390.

18. L. M. Ojwang' and R. L. Cook, Environmental Science & Technology, 2013, 47, 8280-8287.

19. A. Y. Matsuo, R. C. Playle, A. L. Val and C. M. Wood, Aquatic Toxicology, 2004, 70, 63-81.

20. B. Vigneault, A. Percot, M. Lafleur and P. G. Campbell, *Environmental Science & Technology*, 2000, 34, 3907-3913.

21. N. J. Fabian, L. B. Albright, G. Gerlach, H. S. Fisher and G. G. Rosenthal, *Journal of Chemical Ecology*, 2007, 33, 2090-2096.

22. R. Bouchnak and C. E. Steinberg, Limnologica-ecology and Management of Inland Waters, 2010, 40, 86-91.

23. A. Stara, J. Kristan, E. Zuskova and J. Velisek, Pesticide Biochemistry and Physiology, 2013, 105, 18-23.

24. A. Stará, A. Kouba and J. Velíšek, BioMed Research International, 2014, 2014, 1-6.

25. Q. Zhao, C. Wang, X. Yuan and L. Zhu, Journal of Agro-Environment Science, 2015, 34, 653-659. (In Chinese)

26. C. Wiegand, S. Pehkonen, J. Akkanen, O. P. Penttinen and J. V. Kukkonen, Chemosphere, 2007, 66, 558-566.

27. M. Haitzer, S. Höss, W. Traunspurger and C. Steinberg, Aquatic toxicology, 1999, 45, 147-158.

28. P. Qiao and A. Farrell, *Comparative Biochemistry and Physiology Part C: Toxicology & Pharmacology*, 2002, 133, 575-585.

29. D. L. Norwood, R. F. Christman and P. G. Hatcher, Environmental science & technology, 1987, 21, 791-798.

30. X. Xia, Z. Dai, A. H. Rabearisoa, P. Zhao and X. Jiang, Chemosphere, 2015, 119, 978-986.

- 31. J. Akkanen, S. Penttinen, M. Haitzer and J. V. Kukkonen, Chemosphere, 2001, 45, 453-462.
- 32. C. Steinberg, C. Mayr, R. Lorenz, O. Spieser and A. Kettrup, Naturwissenschaften, 1994, 81, 225-227.
- 33. A. Bianchini and C. M. Wood, Environmental Toxicology and Chemistry, 2003, 22, 1361-1367.
- 34. H. A. Roditi, N. S. Fisher and S. A. Sañudo-Wilhelmy, Nature, 2000, 407, 78-80.
- 35. M. A. Timofeyev, C. Wiegand, B. K. Burnison, Z. M. Shatilina, S. Pflugmacher and C. E. Steinberg, Science of
- the Total Environment, 2004, 319, 115-121.
- 36. N. Meems, C. Steinberg and C. Wiegand, Science of the Total Environment, 2004, 319, 123-136.
- 37. A. Kaune, R. Brüggemann and A. Kettrup, Journal of chromatography A, 1998, 805, 119-126.
- 38. K. V. Plakas and A. J. Karabelas, Separation and Purification Technology, 2011, 80, 246-261.

Table lists:

Table 1. Properties of prometryn.

 Table 2. Correlation matrix (Pearson correlation, 2-tailed) between different

 parameters following exposure of zebrafish to prometryn (P) and humic acid (HA) for

 40 days.

Figure captions:

Figure 1. Catalase (CAT) activity (U·mg⁻¹ protein) in the visceral mass of zebrafish exposed to prometryn (P) and humic acid (HA) for 40 days. Data are presented as mean \pm SE (n = 6). Values with different alphabet superscript differ significantly (*P* < 0.05).

Figure 2. Glutathione (GSH) content (mg GSH·g⁻¹ protein) in the visceral mass of zebrafish exposed to prometryn (P) and humic acid (HA) for 40 days. Data are presented as mean \pm SE (n = 6). Values with different alphabet superscript differ significantly (P < 0.05).

Figure 3. Malondialdehyde (MDA) content (nmol·mg⁻¹ protein) in the visceral mass of zebrafish exposed to prometryn (P) and humic acid (HA) for 40 days. Data are presented as mean \pm SE (n = 6). Values with different alphabet superscript differ significantly (P < 0.05).

Figure 4. Gill Na⁺-K⁺-ATPase activity in zebrafish treated with prometryn (P) and humic acid (HA) for 24 h. Data are presented as mean \pm SE (n = 6). Values with

different alphabet superscript differ significantly (P < 0.05).

Figure 5. Prometryn (P) accumulation in zebrafish treated with P and humic acid (HA) for 40 days. Data are presented as mean \pm SE (n = 6). Values with different alphabet superscript differ significantly (*P* < 0.05).

Figure 6. Prometryn concentrations in the exposure systems containing zebrafish treated with prometryn (P) and humic acid (HA) for 15 days.

Figure 7. Dry weight of excreta collected from exposure systems containing zebrafish treated with prometryn (P) and humic acid (HA) for 25 days.

Figure 8. Electrical conductivity in the exposure systems containing zebrafish treated with prometryn (P) and humic acid (HA) for 24 h. Values with different alphabet superscript differ significantly (P < 0.05).

Molecular	Character	Melting	Solubility/mgL ⁻¹	Log <i>K</i> _{ow}	Half-life	Molecular	Vapour
formula		point/°C			period/d	Weight/Da	pressure/mPa
HN/ S/N/N/ H/	White	118-120	33 (25℃)	2.99 ³⁷	390	241.35 ³⁸	0.13 (20°C)
	cry stal						

Table 1. Property of prometryn

40 days.								
	CAT	GSH	Р	MDA	EC	W_E	$\mathbf{P}_{\mathbf{w}}$	Na ⁺ -K ⁺ -ATPase
CAT ^a	1							
GSH^b	0.696**	1						
	0.001							
P^{a}	0.562^{**}	0.808^{**}	1					
	0.004	0.000						
MD A ^a	0.418^*	0.351	0.114	1				
	0.042	0.153	0.596					
EC^{c}	-0.918**	-0.885^{*}	-0.980**	-0.833*	1			
	0.010	0.019	0.001	0.039				
$\mathbf{W}_{\mathrm{E}}^{\mathrm{a}}$	-0.330	-0.598^{*}	-0.430	-0.491*	-0.719	1		
	0.115	0.009	0.036	0.015	0.108			
P_W^d	-0.108	-0.804	-0.491	-0.223	0.962^{**}	-0.815**	1	
	0.739	0.054	0.105	0.485	0.002	0.001		
Na ⁺ -K ⁺ -ATPase ^c	0.585	0.314	0.713	0.666	-0.617	0.981**	-0.775	1
	0.222	0.545	0.112	0.149	0.192	0.001	0.070	

Table 2. Correlation matrix (Pearson correlation, 2-tailed) between different parameters following exposure of zebrafish to prometryn (P) and humic acid (HA) for 40 days.

P, prometryn concentration in zebrafish; EC, electrical conductivity; W_E , dry weight of excreta; P_W , prometryn concentrations in the exposed solution; $Na^+-K^+-ATPase$, $Na^+-K^+-ATPase$ activity in zebrafish gills.

^a, n = 24; ^b, n = 18; ^c, n = 6; ^d, n = 12.

*, correlation is significant at the 0.05 level (2-tailed); **, correlation is significant at the 0.01 level (2-tailed).



Figure 1. Catalase (CAT) activity (U·mg⁻¹ protein) in the visceral mass of zebrafish exposed to prometryn (P) and humic acid (HA) for 40 days. Data are presented as mean \pm SE (n = 6). Values with different alphabet superscript differ significantly (*P* < 0.05).



Figure 2. Glutathione (GSH) content (mg GSH·g⁻¹ protein) in the visceral mass of zebrafish exposed to prometryn (P) and humic acid (HA) for 40 days. Data are presented as mean \pm SE (n = 6). Values with different alphabet superscript differ significantly (P < 0.05).



Figure 3. Malondialdehyde (MDA) content (nmol·mg⁻¹ protein) in the visceral mass of zebrafish exposed to prometryn (P) and humic acid (HA) for 40 days. Data are presented as mean \pm SE (n = 6). Values with different alphabet superscript differ significantly (P < 0.05).



Figure 4. Gill Na⁺-K⁺-ATPase activity in zebrafish treated with prometryn (P) and humic acid (HA) for 24 h. Data are presented as mean \pm SE (n = 6). Values with different alphabet superscript differ significantly (*P* < 0.05).



Figure 5. Prometryn (P) accumulation in zebrafish treated with P and humic acid (HA) for 40 days. Data are presented as mean \pm SE (n = 6). Values with different alphabet superscript differ significantly (*P* < 0.05).



Figure 6. Prometryn concentrations in the exposure systems containing zebrafish treated with prometryn (P) and humic acid (HA) for 15 days.



Figure 7. Dry weight of excreta collected from exposure systems containing zebrafish treated with prometryn (P) and humic acid (HA) for 25 days.



Figure 8. Electrical conductivity in the exposure systems containing zebrafish treated with prometryn (P) and humic acid (HA) for 24 h. Values with different alphabet superscript differ significantly (P < 0.05).



HA made the toxicity of prometryn stronger and then weaker to *Danio rerio* during $1 \sim 40$ days.