



**Selective decolorization of cationic dyes by
peroxymonosulfate: non-radical mechanism and effect of
chloride**

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1 Selective decolorization of cationic dyes by
2 peroxymonosulfate: non-radical mechanism and effect of
3 chloride

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15 **Abstract**

16 How to effectively activate oxidants like hydrogen peroxide (H_2O_2), peroxydisulfate
17 (PDS) and peroxymonosulfate (PMS) is the focus of advanced oxidation processes
18 (AOPs). In this study, we have found that PMS can directly decolorize cationic dyes
19 without activation in a wide pH range (2.0~12.0). In addition, the presence of Cl^-
20 results in the formation of free available chlorine species and thus improves the
21 degradation efficiency as well as broadening the scope of target pollutants. The
22 first-order rate constant increased by 4.3 times when 2.5 mM Cl^- was added as
23 observed by an increase from 0.021 to 0.089 min^{-1} , which further rose to 1.371 min^{-1}
24 when Cl^- was present at a level of 50 mM. Furthermore, it was found that only PMS
25 has such an amazing effect while H_2O_2 and PDS do not. Radical quenching
26 experiments and electron paramagnetic resonance (EPR) studies show that the
27 decolorization of cationic dyes by PMS does not rely on the formation of sulfate
28 radical. A plausible mechanism is that cationic dyes (i.e., RhB^+) first formed a
29 complex with the active component of PMS (HSO_5^-) owing to their electrical
30 interaction. Subsequently, direct electron transfer from cationic dyes to HSO_5^- may
31 occur and probably this is responsible for the bleaching of cationic dyes.

32 **Key Words:** peroxymonosulfate; AOPs; non-radical process; electron transfer;
33 Rhodamine B

34

35 1. Introduction

36 Aqueous effluents containing dyes have increased greatly with their wide applications
37 in many industries^{1,2}. How to best dispose of these dye effluents has been discussed
38 for decades^{1,3}. Normally, the treatment methods used for dye effluents can be divided
39 into three categories: biological, physical and chemical methods³. As physical
40 methods only transfer wastes from one place to another and do not involve any
41 degradation they have received the least attention. Although biological processes that
42 rely on the activity of bacteria are economical and relatively effective, they are
43 usually time consuming and are not applicable for toxic and low biodegradable dyes,
44 and unfortunately, most dyes lie in both of these categories⁴. On the other hand,
45 chemical oxidation is the most effective technique and has received the most
46 attention^{5,6}.

47 Over the past two decades, advanced oxidation processes (AOPs) that depend on the
48 generation of reactive oxygen species (ROS) have been intensively developed for the
49 remediation of wastewater^{2, 7, 8}. Generally, ROS are produced from the reaction
50 between transition metals and oxidants. The three widely applied oxidants are
51 hydrogen peroxide (H₂O₂), peroxydisulfate (PDS) and peroxymonosulfate (PMS).
52 AOPs can be classed as either homogeneous or heterogeneous and, in contrast with
53 the homogenous process which is only effective at acid pH (2.5~4.0), heterogeneous
54 AOPs can be operated in a wide pH range^{9, 10}. In addition, the use of catalysts can
55 avoid the formation of sludge and have much less problems relating to the toxicity of
56 free metal ions. Furthermore, some light sensitive catalysts can be combined with UV,

57 visible, and even, solar light so as to improve the efficiency of pollutant degradation¹¹.
58 Specifically, Wu et al.¹² reported that the decolorization of dyes by phase-pure layered
59 perovskite did not need any extra oxidant or the input of external energy. Similarly, Yu
60 et al.¹³ reported that bismuth silver oxide alone is also effective for the decolorization
61 of Rhodamine B. Indeed, heterogeneous AOPs show many advantages over
62 homogenous AOPs but the fact that the preparation of catalyst is neither easy nor
63 economic should not be ignored. It would be a great advance in AOPs technology if
64 dye effluents could be degraded by oxidants without the use of any transition metal
65 catalysts and we now report our intriguing results in this regard.

66 Our current research, involves evaluating the oxidation ability of the ZnFe₂O₄/PMS
67 coupled system. Unfortunately, it was found that ZnFe₂O₄ cannot activate PMS as
68 there was no obvious change observed when the dosage of ZnFe₂O₄ was altered
69 (Figure S1). Fortunately and interestingly, it was found that PMS can directly lead to
70 the degradation of RhB, without the aid of catalyst or other external energy. Therefore,
71 several important factors that affect the process and the mechanism of this
72 phenomenon were investigated. To the best of our knowledge, this is the first report of
73 the destruction of organic pollutants by PMS alone.

74 **2. Materials and Methods**

75 **2.1 Chemicals and Materials.**

76 Methylene Blue (MB) was purchased from Koch-Light Laboratories Ltd., England.
77 Orange II was bought from Panreac Quimica SA, Barcelona, Spain. Sodium chloride

78 (NaCl) was obtained from Merck, Taiwan. The 5,5-dimethyl-1-pyrrolidine N-oxide
79 (DMPO) was purchased from Aladdin, China. The other main reagents, including
80 Reactive Black 5 (RB5), Rhodamine B (RhB), phenol, methanol, and potassium
81 peroxymonosulfate were bought from Sigma-Aldrich, Taiwan. All reagents used were
82 at least analytical reagent and prepared in distilled water.

83 **2.2 Procedure.**

84 The batch trials were performed in a brown glass bottle and mechanical stirred with
85 the desired concentration of PMS and target pollutants. The volume is 400 mL. When
86 appropriate, the solution pH was adjusted with H₂SO₄ (0.1 M) or NaOH (0.1 M).
87 Samples were removed at predetermined time intervals and analyzed immediately. In
88 some cases, high purity argon was purged during the duration of the reaction in order
89 to create an anaerobic environment. For that, the brown glass bottle was replaced with
90 a novel reactor that has already been reported in our previous work¹⁴.

91 **2.3 Analysis.**

92 The residual concentration of dyes was monitored at the appropriate λ_{\max} using a
93 Hewlett–Packard 8453 diode array spectrophotometer (Agilent). The concentration of
94 phenol was determined by HPLC (Shimadzu 6A) using conditions reported
95 previously¹⁵. The mineralization of pollutants was measured using a TOC analyzer
96 (Sievers 900 Portable).

97 **3. Results and Discussion**

98 **3.1 Effects of major factors.**

99 **3.1.1 Effect of PMS dosage.**

100 As a preliminary step the effect of the dosage of PMS was explored. Figure 1(A)
101 shows the degradation of RhB is apparently PMS dependent. Similar to traditional
102 AOPs, the bleaching of RhB increased with the rise of PMS concentration.
103 Accordingly, the rate constant that was obtained from pseudo first-order kinetic fitting
104 ($R^2 > 0.9$) also increased by 10 times when the PMS concentration increased from 1 to
105 10 mM as shown in Table 1. Nevertheless, there is a difference between our system
106 and that of the usual AOPs. As is well known, there exists an optimum dosage of
107 oxidants in AOPs as they will also react with the generated radicals^{16, 17} but our
108 results suggested at the given PMS dosage from 1.0 mM to 100 mM, the higher the
109 PMS dosage, the better the removal rate.

110 **3.1.2 Effect of RhB concentration.**

111 Figure 1(B) shows that the initial concentration of RhB has insignificant effect on its
112 color removal, and the first-order rate constants are all around $0.022 \pm 0.002 \text{ min}^{-1}$. In
113 this case, the amount of RhB removed would increase almost proportionally with
114 initial dye concentration. Specifically, 9.5, 18.3, 37.8 and 72.5 mg/L of RhB were
115 removed after two hours reaction when initial dye concentrations were 10, 20, 40 and
116 80 mg/L, respectively. This may due to the increased collision chances between PMS
117 and RhB, which leads to the removal of more RhB. Based on the result, it can be
118 concluded that 2.5 mM of PMS is a sufficiently high dosage for 10 - 80 mg/L RhB
119 and therefore over 90% of RhB were bleached in all conditions.

120 **3.1.3 Effect of pH.**

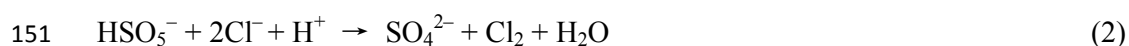
121 It is well known that the solution pH has a great influence on the performance of most

122 AOPs. Generally, while homogeneous AOPs activated by transition metal ions require
123 an acidic pH solution¹⁸, heterogeneous AOPs can operate in a relatively wide pH
124 range but always perform better in slightly acidic or neutral solution¹⁹. To investigate
125 the role of pH in this new kind of decolorization process, the effect of pH at 2.0, 4.5,
126 7.0, 10.0 and 12.0 was explored. Because PMS solution itself is a strong acid solution,
127 the introduction of PMS will immediately lead to the fall of solution pH to around 2.5
128 as indicated in Table S1, when the initial solution pH was 4.5, 7.0 and 10.0.
129 Correspondingly, similar RhB removal efficiencies (96.1% ~ 97.5%) were observed
130 for pH₀ 4.5, 7.0 and 10.0, as shown in Figure 1(C). However, if the solution pH₀ is
131 increased from 10.0 to 12.0, the decolorization percentage declined by 31.4% in 120
132 min reaction and the rate constant dropped from 0.047 to 0.008 min⁻¹. This is
133 probably because that PMS exists mostly in the form of SO₅²⁻ at pH₀ 12.0 and as
134 reported, it is less reactive than HSO₅⁻²⁰. Nevertheless, 66.1% of RhB was
135 decolorized in this case. These results indicate that unlike homogeneous Fenton based
136 AOPs, the bleaching of dyes by PMS can work in a wide pH range.

137 **3.2 Effect of chloride ions.**

138 Much work on the effect of Cl⁻ on the degradation of pollutants in AOPs has been
139 reported as it is observed that both sulfate and hydroxyl radicals can react with Cl⁻ to
140 form chlorine radicals^{21,22}. On the other hand, Cl⁻ is also one of the main components
141 in wastewater so the exploration of the potential effect of Cl⁻ needs investigation. As
142 can be seen from Figure 2, Cl⁻ plays a positive role in our system where the time for
143 total decoloration of 10 mg/L RhB (>96.1%) decreased from more than 120 min to

144 less than 90 min when only 1 mM Cl^- was added. The time can be further shortened
145 by increasing the dose of Cl^- . For example, only 3 min is needed for 100% bleaching
146 of 10 mg/L RhB in the presence of 50 mM Cl^- . Accordingly, with 50 mM Cl^- , the
147 reaction rate constant increased 66-fold to 1.371 min^{-1} , over the rate constant at the
148 absence of Cl^- . It could be surmised that the addition of Cl^- results in the formation of
149 free available chlorine species as shown in Eqs. (1) and (2)^{21, 23}:



152 The observation of the production of small gas bubbles in the reactor supports the
153 formation of gaseous chlorine as suggested by Eq. (2). HOCl and Cl_2 are both widely
154 used as disinfectants and bleaching agents and thus the greatly improved removal of
155 RhB is observed in the presence of Cl^- . Yuan et al.²² reported that in the Co^{2+} /PMS
156 system, the presence of Cl^- leads to more refractory intermediate products as they
157 noticed a new peak in the UV/Vis spectra, but no new peak in the absence of Cl^- .
158 However, in our system, whether Cl^- was present or not, the characteristic peaks of
159 RhB at 260, 355 and 552 nm gradually disappear (see Figure 3) and no new peaks
160 appeared in our system. The continuous decrease of these peaks indicated that the
161 chromophores of RhB as well as the benzene rings are destroyed. The difference in
162 the UV/Vis spectra between the current work and that reported by others may due to
163 the fact that our system does not rely on the generation of radicals.

164 The removal of TOC at different amounts of added Cl^- was also determined as TOC is
165 an important indication of the extent of mineralization of pollutants. It was found that

166 the TOC removal also increased at higher Cl^- concentration. As shown in Figure 4,
167 the TOC removal of 10 mg/L RhB after 120 min is 8.1%, 9.2%, 19.6%, 39.0%, and
168 50.0% at 0, 1, 2.5, 10, and 50 mM Cl^- , respectively. Although PMS alone just leads to
169 the decolorization of RhB, PMS coupled with Cl^- is effective for the mineralization of
170 RhB. As most dye effluents contain Cl^- , our system may provide a simple and
171 effective technology for the treatment of dye effluents.

172 **3.3 Possibility of employing PDS or H_2O_2 .**

173 As discussed above, the results show that PMS alone can effectively degrade dyes.
174 Since PDS and H_2O_2 are also two widely used oxidants, there is a necessity to explore
175 their potential possibility in the non-radical process. Figure 5 illustrates that H_2O_2
176 alone, or in conjunction with NaCl could not degrade dyes. Although PDS did have
177 some effect but was far less effective than PMS and the addition of 5 mM Cl^- did not
178 result in any improvement or inhibition in the removal of RhB. It is concluded that
179 PMS is much more active than the other two oxidants and the structure of the three
180 oxidants can account for the different performances in similar reaction conditions.
181 PDS and H_2O_2 with symmetric structures are more stable than PMS and this is
182 consistent with some reports relating to AOPs. For instance, Guan et al.²⁰ reported that
183 only PMS can be easily activated by CuFe_2O_4 while PDS and H_2O_2 cannot and
184 consequently, RhB was rapidly decolorized in the case of PMS.

185 **3.4 Possibility of removing other pollutants.**

186 In order to investigate the potential application of PMS in the decontamination of
187 other pollutants, phenol and another three dyes: Orange II, MB and RB5 were also

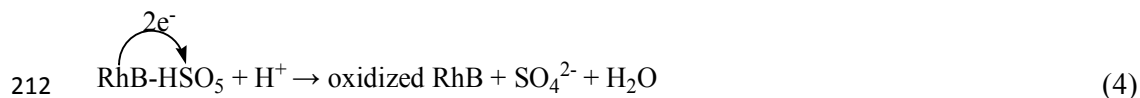
188 treated by PMS based process. Figure 6(A) shows that only MB could be effectively
189 oxidized by PMS alone, while the other pollutants were not removed at all. However,
190 as shown in Figure 6(B), all of these pollutants were readily degraded when 2.5 mM
191 Cl^- were added. These results indicate that PMS alone can selectively oxidize dyes,
192 while its combination with Cl^- can remove different kind of pollutants by the *in-situ*
193 formation of free available chlorine species.

194 **3.5 Mechanism study.**

195 Methanol is a popular radical quencher for hydroxyl and sulfate radicals²⁴, so it was
196 added to explore the reactive species in our system. Figure 7 (A) shows the addition
197 of 2.5 M methanol slightly inhibits the degradation of RhB in PMS/ Cl^- system but has
198 no impact on PMS alone. It was proposed that the decolorization of RhB was not
199 relying on the formation of radicals. To further testify this assumption, electron
200 paramagnetic resonance (EPR) studies using DMPO as a spin-trapping agent were
201 investigated. As can be noticed from Figure 7(B), neither DMPO-OH nor DMPO-SO₄
202 signals were observed in both PMS/RhB and PMS/ Cl^- /RhB systems, indicating that
203 the decolorization of RhB in the two systems is not attributed to radical attack. It
204 should be noted that the weak signals of DMPO adducts are attributed to the
205 background noise.

206 In addition, as identified above, only MB and RhB can be directly decolorized by
207 PMS. It can thus be deduced that non-radical process is only effective for cationic
208 dyes. The active component of PMS is HSO_5^- , and RhB actually exists in the form of
209 cationic (RhB^+) in aqueous solution. When PMS is added to RhB, a strong electrical

210 interaction occurs between the two ions as shown in Eqs. (3) and (4):



213 The two-electron transfer from RhB to PMS occurs, leading to the oxidation of RhB,
214 and a similar process should occur in the degradation of MB. However, as anionic
215 dyes will repel HSO_5^- due to their mutually exclusive electrical interaction, the
216 subsequent electron transfer cannot occur. The proposed mechanism is inconsistent
217 with the recent discovery of Wang et al.²⁵, who reported that PMS could directly
218 oxidize As(III) to As(V). Additionally, as mentioned above, H_2O_2 without activation
219 cannot degrade dyes because H_2O_2 mostly exists in the molecular form in solution.
220 The behavior of H_2O_2 is also a good support for the suggested mechanism. As a result,
221 there is no degradation of phenol (a neutral organic molecule) and anionic dyes. The
222 decreased removal rate at pH 12.0 (Figure 3) also supports the above discussed
223 speculation. At pH 12.0, the attraction of HO^- to RhB^+ is dominant so the reaction as
224 expressed by Eq. (3) will be inhibited and the extent of decolorization of RhB will be
225 decreased. It may be postulated that oxygen is involved in the decolorization of dyes
226 as it was reported that singlet oxygen ($^1\text{O}_2$) will be produced in a similar
227 heterogeneous reaction¹³. However, the decolorization efficiency of RhB did not
228 decrease in the absence of O_2 (Figure S2) when the solution was purged with high
229 purity argon. It was concluded that O_2 did not take part in the bleaching of dyes.

230 4. Conclusions

231 Two systems, namely PMS and PMS/ Cl^- were investigated for degradation of RhB.

232 The results demonstrated that PMS can selectively degrade cationic dyes through a
233 two electron transfer process that does not rely on the generation of radicals. In
234 addition, Cl^- that widely exists in wastewater can accelerate the degradation of dyes
235 by in situ generation of free available chlorine species. Unlike PMS alone, the
236 PMS/ Cl^- system is very effective for all the pollutants studied. It is suggested that
237 either PMS or its combination with Cl^- can serve as an effective and simple
238 technology for the decolorization/degradation of organic contaminants. Additionally,
239 this study suggests that the role of PMS and Cl^- in PMS based AOPs should be
240 reinvestigated and this will be our focus in future work.

241 **Supplementary Material**

242 Supplementary Material associated with this article is available free of charge via the
243 Internet at <http://www.sciencedirect.com>.

244 **Acknowledgments**

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289

Figure captions

290 **Figure 1.** Effects of (A)PMS dosage, (B)RhB concentration and (C) initial pH on the
291 degradation of RhB. Conditions: (A) [RhB] = 10 mg/L, $\text{pH}_0 = 4.5 \pm 0.1$; (B) [PMS] =
292 2.5 mM, $\text{pH}_0 = 4.5 \pm 0.1$; (C) [RhB] = 10 mg/L, [PMS] = 2.5 mM, Time = 120 min.

293 **Figure 2.** Effect of chloride ion on the degradation of RhB by PMS. Conditions: [RhB]
294 = 10 mg/L, [PMS] = 2.5 mM, $\text{pH}_0 = 4.5 \pm 0.1$.

295 **Figure 3.** Changes in the UV-vis spectra of RhB on reaction with PMS in the absence
296 (A) and presence (B) of chloride ion Conditions: [RhB] = 10 mg/L, [PMS] = 2.5 mM,
297 $[\text{Cl}^-] = 2.5 \text{ mM}$, $\text{pH}_0 = 4.5 \pm 0.1$.

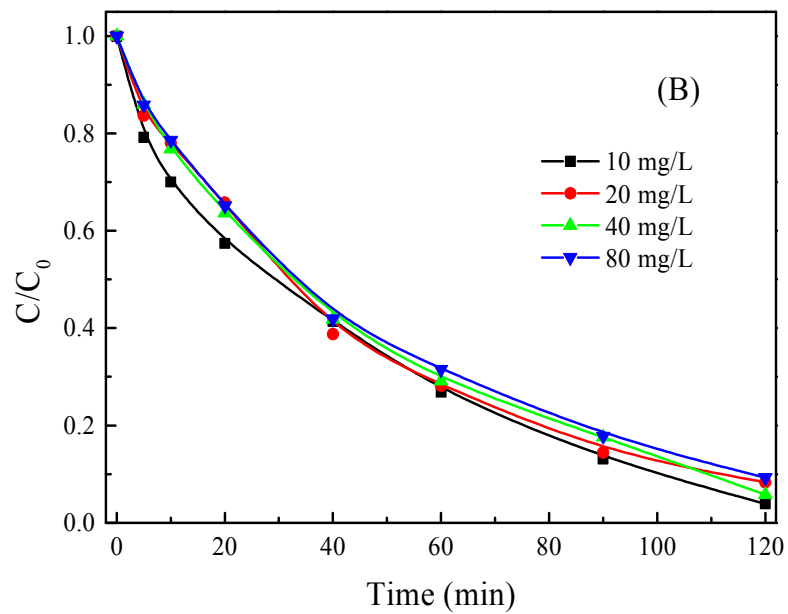
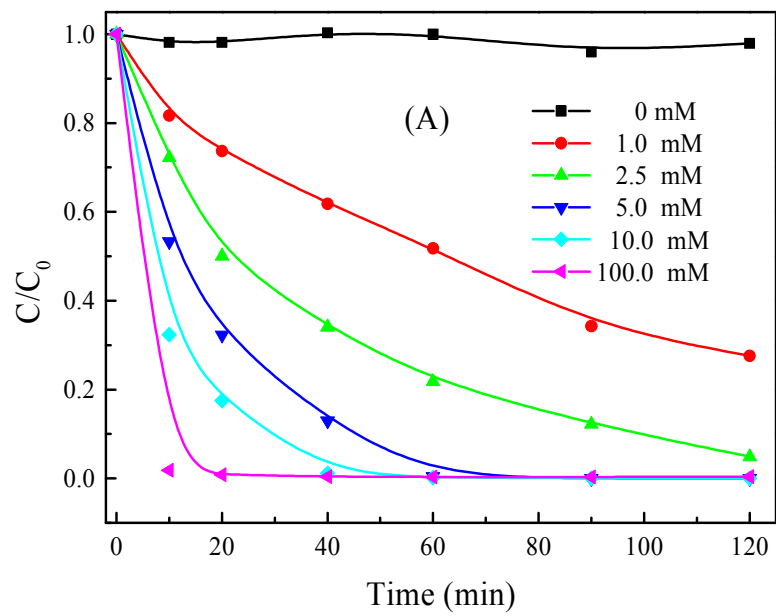
298 **Figure 4.** TOC removal of RhB using different concentrations of chloride ion.
299 Conditions: [RhB] = 10 mg/L, [PMS] = 2.5 mM, $\text{pH}_0 = 4.5 \pm 0.1$. Time = 120 min.

300 **Figure 5.** Degradation of RhB with PDS and H_2O_2 in the presence or absence of
301 chloride ion. Conditions: [RhB] = 10 mg/L, [PMS] = 2.5 mM, $[\text{Cl}^-] = 2.5 \text{ mM}$, $\text{pH}_0 =$
302 4.5 ± 0.1 .

303 **Figure 6.** Degradation of other organic pollutants with PMS (A) or PMS/ Cl^- (B) and
304 UV-vis spectra changes of MB in PMS system (insert figure). Conditions: [MB] =
305 [RB5] = [Orange II] = 10 mg/L, [phenol] = 0.1 mM, [PMS] = 2.5 mM, $[\text{Cl}^-] = 2.5$
306 mM, $\text{pH}_0 = 4.5 \pm 0.1$.

307 **Figure 7.** Effect of methanol on the degradation of RhB (A) and EPR spectra in.
308 PMS/RhB and PMS/RhB/ Cl^- systems (B). Conditions: [RhB] = 10 mg/L, [PMS] = 2.5
309 mM, $[\text{Cl}^-] = 2.5 \text{ mM}$, $\text{pH}_0 = 4.5 \pm 0.1$, [methanol] = 2.5 M, [DMPO] = 10 mM.

310 **Table 1.** Kinetic coefficients under various operating conditions (RhB = 10 mg/L).



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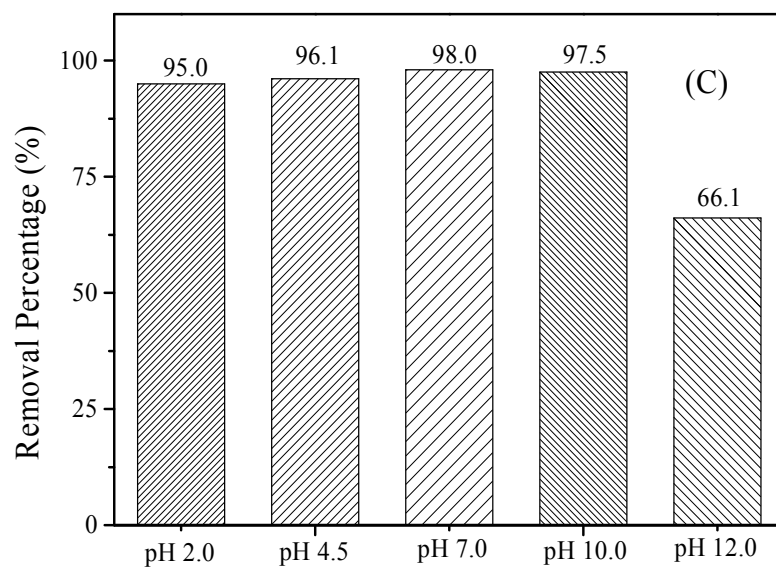


Figure 1. Effects of (A) PMS dosage, (B) RhB concentration and (C) initial pH on the degradation of RhB. Conditions: (A) [RhB] = 10 mg/L, $\text{pH}_0 = 4.5 \pm 0.1$; (B) [PMS] = 2.5 mM, $\text{pH}_0 = 4.5 \pm 0.1$; (C) [RhB] = 10 mg/L, [PMS] = 2.5 mM, Time = 120 min.

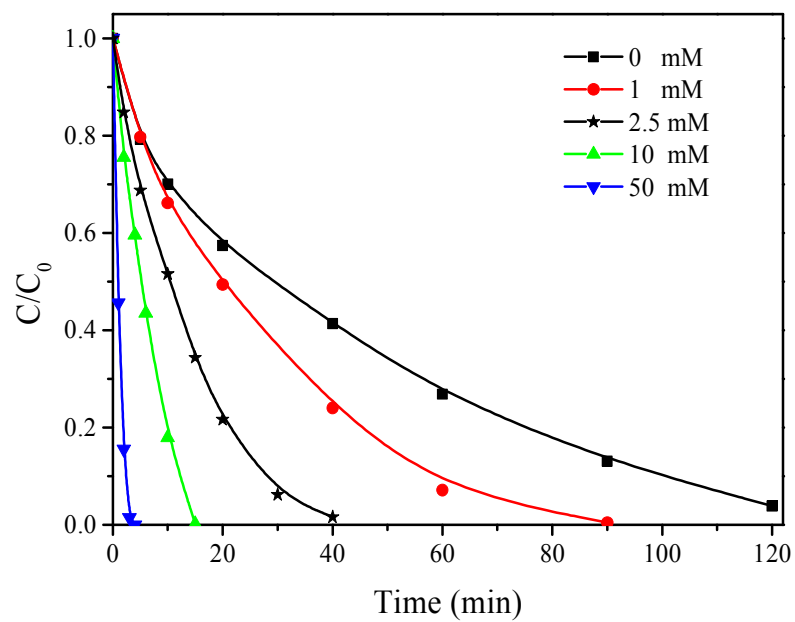


Figure 2. Effects of chloride ion on the degradation of RhB by PMS. Conditions:

[RhB] = 10 mg/L, [PMS] = 2.5 mM, $\text{pH}_0 = 4.5 \pm 0.1$.

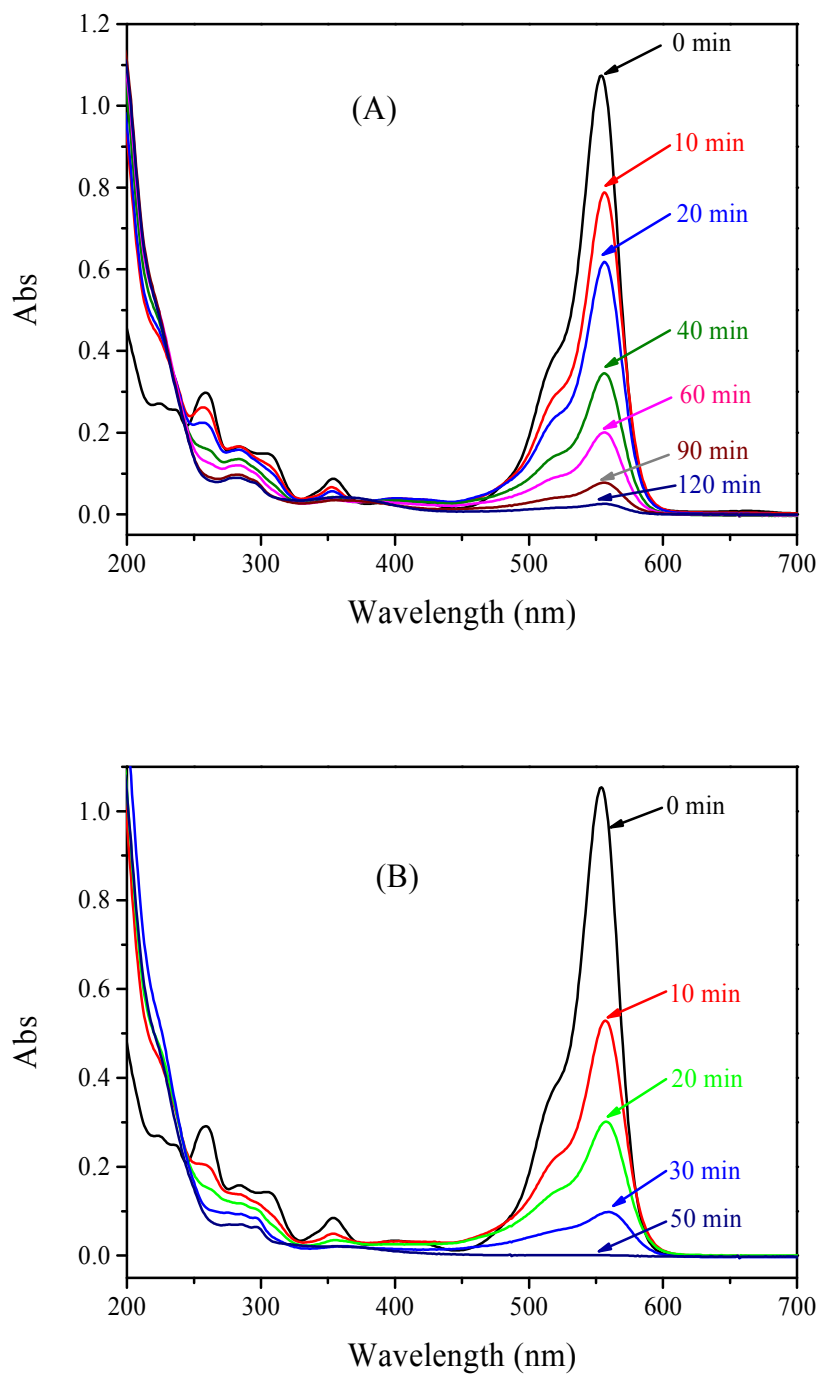


Figure 3. Changes in the UV-vis spectra of RhB on reaction with PMS in the absence (A) and presence (B) of chloride ion Conditions: $[\text{RhB}] = 10 \text{ mg/L}$, $[\text{PMS}] = 2.5 \text{ mM}$, $[\text{Cl}^-] = 2.5 \text{ mM}$, $\text{pH}_0 = 4.5 \pm 0.1$.

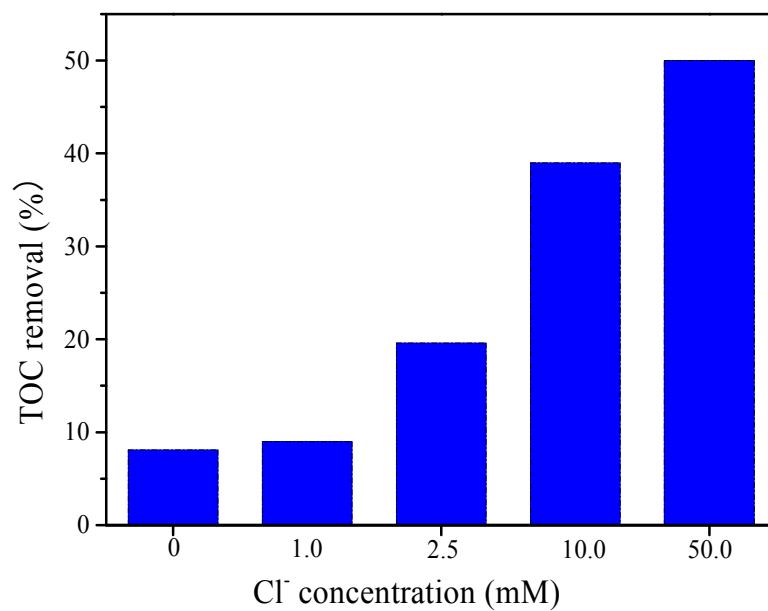


Figure 4. TOC removal of RhB using different concentrations of chloride ion.

Conditions: [RhB] = 10 mg/L, [PMS] = 2.5 mM, $\text{pH}_0 = 4.5 \pm 0.1$. Time = 120 min.

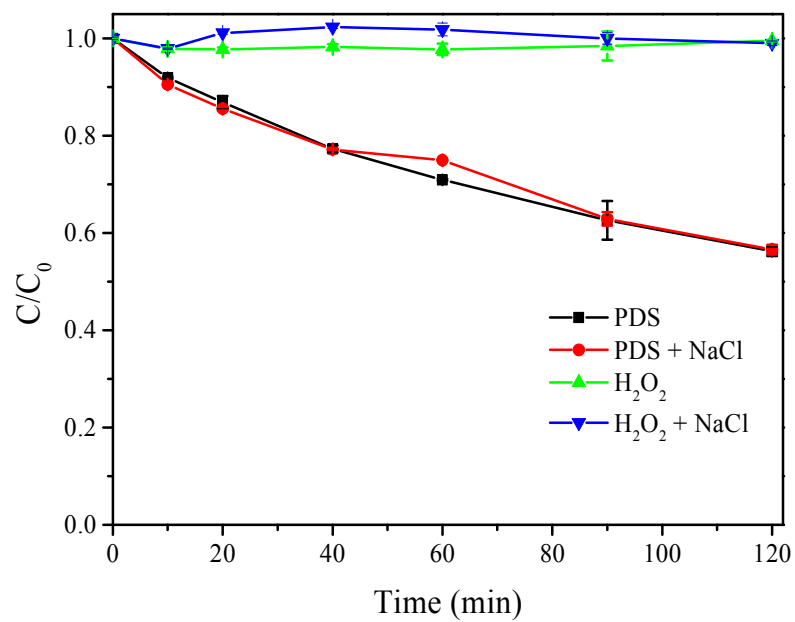


Figure 5. Degradation of RhB with PDS and H₂O₂ in the presence or absence of chloride ion. Conditions: [RhB] = 10 mg/L, [PMS] = 2.5 mM, [Cl⁻] = 2.5 mM, pH₀ = 4.5 ± 0.1.

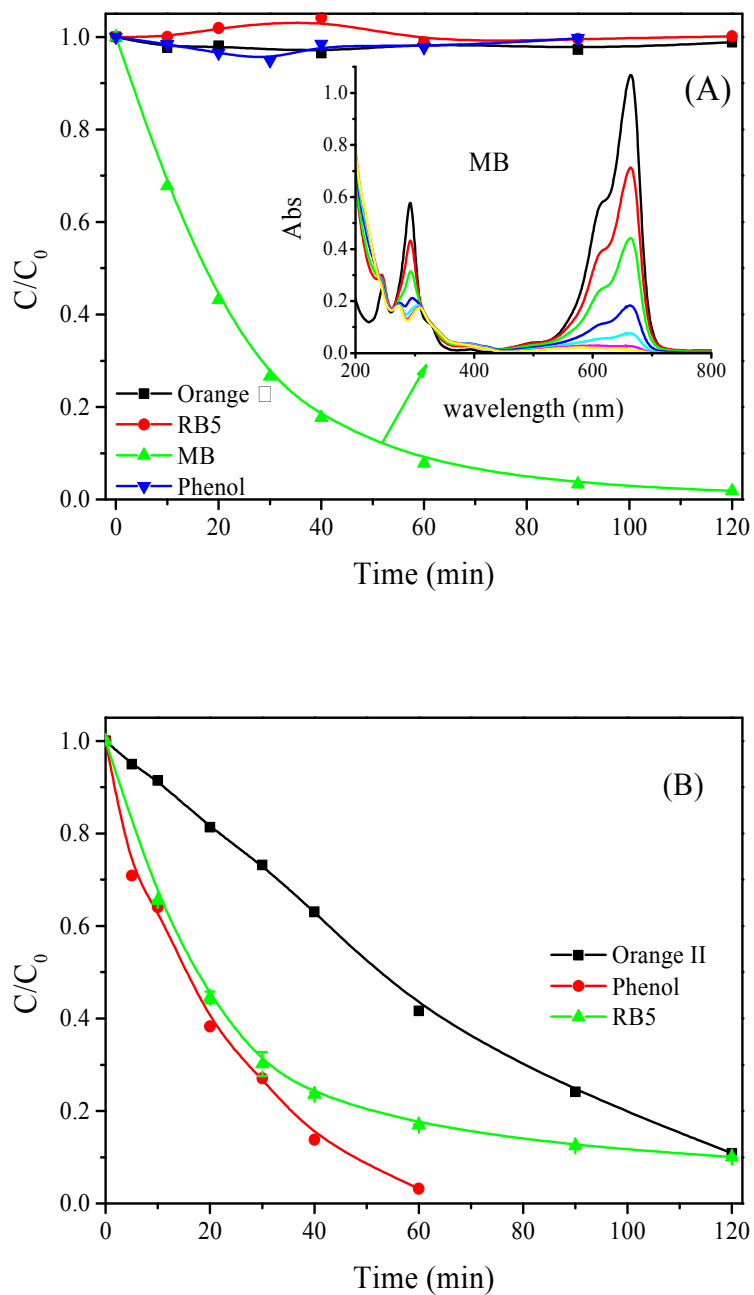
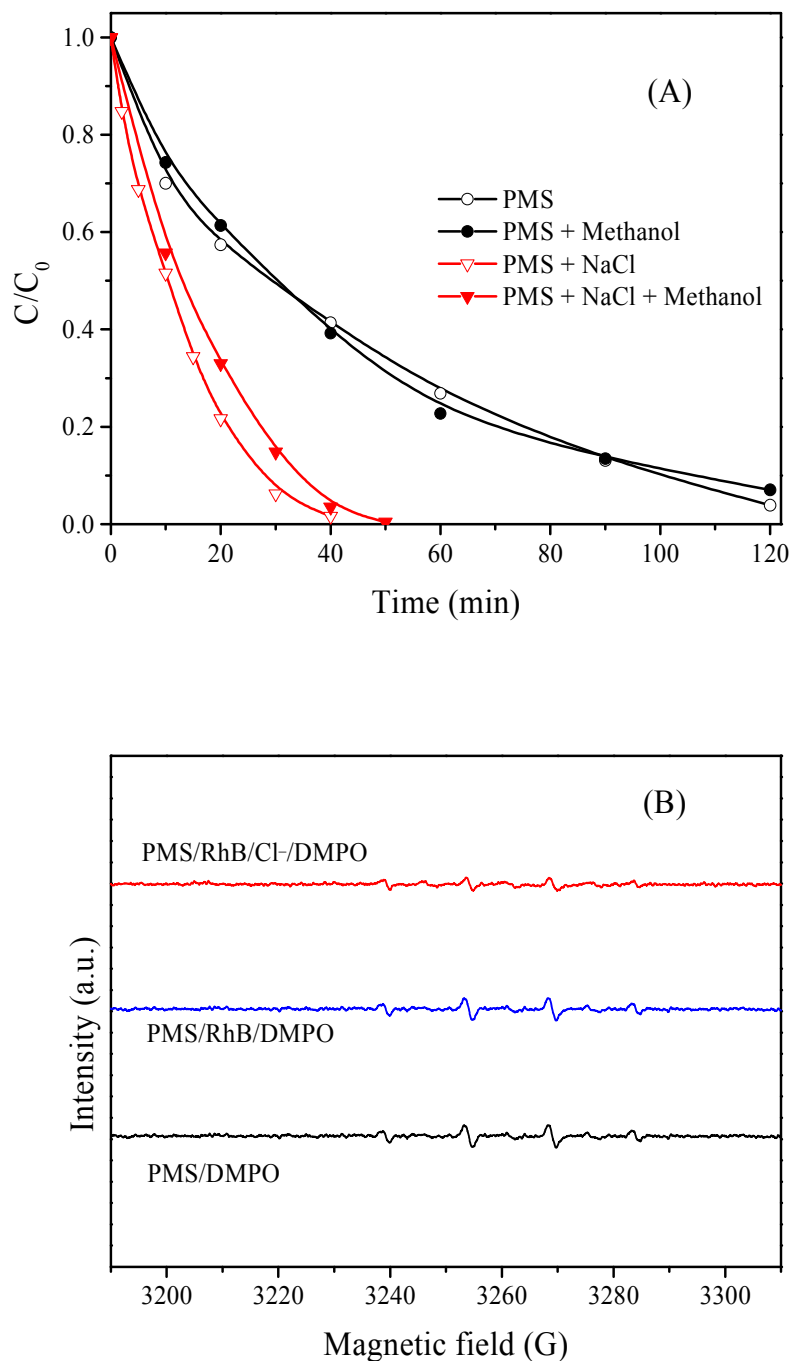


Figure 6. Degradation of other organic pollutants with PMS (A) or PMS/Cl⁻ (B) and UV-vis spectra changes of MB in PMS system (insert figure). Conditions: [MB] = [RB5] = [Orange II] = 10 mg/L, [phenol] = 0.1 mM, [PMS] = 2.5 mM, [Cl⁻] = 2.5 mM, pH₀ = 4.5 ± 0.1.



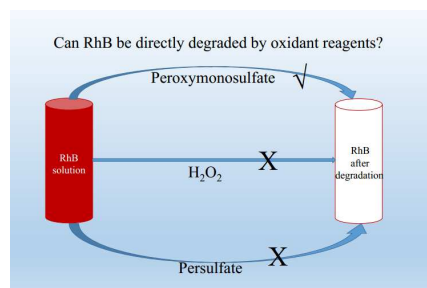
312 **Figure 7.** Effect of methanol on the degradation of RhB (A) and EPR spectra in
313 PMS/RhB and PMS/RhB/Cl⁻ systems (B). Conditions: [RhB] = 10 mg/L, [PMS] = 2.5
314 mM, [Cl⁻] = 2.5 mM, pH₀ = 4.5 ± 0.1, [methanol] = 2.5 M, [DMPO] = 10 mM.

Table 1.

Kinetic coefficients under various operating conditions (RhB = 10 mg/L).

Run No.	[PMS] (mM)	pH ₀	[Cl ⁻] (mM)	k (min ⁻¹)	R ²
1	1.0	4.5	0	0.010	0.963
2	2.5	4.5	0	0.021	0.979
3	5.0	4.5	0	0.050	0.991
4	10.0	4.5	0	0.110	0.975
5	2.5	2.0	0	0.021	0.992
6	2.5	10.0	0	0.047	0.905
7	2.5	12.0	0	0.008	0.982
8	2.5	4.5	1.0	0.042	0.978
9	2.5	4.5	2.5	0.089	0.973
10	2.5	4.5	10.0	0.170	0.973
11	2.5	4.5	50.0	1.371	0.902

Graphical abstract



PMS alone is able to decolorize cationic dyes without the need for any catalysts.