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The facile synthesis, characterization and evaluation of photocatalytic activity of bimetallic FeBiO$_3$ in natural sunlight exposure

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Abstract

In an effort to develop sunlight active photocatalyst for environmental remediation, the phase pure bimetallic oxide, FeBiO$_3$, was synthesized by a facile route. The optical, structural and morphological properties of the synthesized FeBiO$_3$ were compared with both the parent oxides i.e. α-Fe$_2$O$_3$ and Bi$_2$O$_3$. The synthesized powder exhibited strong absorption in the visible region with the appearance of a distinct optical absorption edge at 2.15 eV. The PL, Raman and IR spectroscopic investigations of the synthesized powder, in comparison to that of pure oxides, confirmed the formation of pure phase FeBiO$_3$ and revealed the dominant role of Fe$^{3+}$ ions in controlling the optical properties of the material. The x-ray diffraction analysis (XRD) revealed the hexagonal or rhombohedral geometry with an average crystallite size of 20.2 nm, whereas the x-ray photoelectron spectroscopy (XPS) verified the existence of both Fe and Bi in 3+ oxidation states. The electrochemical evaluation of the synthesized catalysts revealed its excellent stability in the neutral and basic pH in the dark and under illumination however, the catalyst was unstable in the harsh acidic medium (pH = 2). A low resistance to electron transfer and better charge retention ability was witnessed by EIS and chronopotentiometry. The photocatalytic activity of the synthesized catalysts was evaluated in the exposure of complete spectrum and visible region of natural sunlight for the removal of chemically stable substrates such as 2-nitrophenol and 2-chlorophenol. The catalysts showed considerably high activity for the removal of both the substrates in the exposure of natural sunlight, whereas 25% less activity was witnessed in visible region in the same span of time. In addition to degradation, the catalyst also exhibited a substantial activity for mineralization (TOC removal). The catalysts unveiled reproducible activity in the repeated scans. The chemical stability and sustained activity both in the complete spectrum and visible region designate it as a potential candidate for photocatalytic environmental remediation.

Keywords: bimetallic oxide, sunlight, degradation, mineralization, FeBiO$_3$
Introduction

Although photocatalysis is an escalating area in water decontamination however, the associated technical issues such as the non-availability of active narrow bandgap photocatalysts, the use of complicated experimental setup with artificial excitation sources hinders its wide spread commercial applicability. The use of technologically complicated and expensive excitation sources not only restrict this conclusive decontamination process to lab-scale but adds to its cost enormously. It is believed that the cheap and ever renewable sunlight can endorse the wide spread large-scale commercial applicability of photocatalysis by replacing the artificial light sources. The inability of the active photocatalysts like TiO$_2$ (3.2 eV) and ZnO (3.0 eV), to harvest the maximal portion of the solar spectrum has forced the scientific community to investigate new sunlight responsive material with superior activity. In this context, besides the alteration of already existing single metal visible light active oxide semiconducting materials such as Bi$_2$O$_3$, WO$_3$, V$_2$O$_5$, CeO$_2$ etc., for better activity, the exploration of bimetallic oxide semiconductor materials as photocatalysts is an innovative approach. The co-existence of two chemically compatible metals, besides improving the structural properties, can strongly influence the optical and photocatalytic properties of bimetallic semiconductor oxides.$^{1-17}$

Iron bismuth oxide (FeBiO$_3$) with a bandgap of 2.2 eV, chemical stability and optoelectronic properties is a versatile material for aqueous phase photocatalytic applications.$^{18-19}$ The single phase FeBiO$_3$ is ferroelectric and weakly ferromagnetic at room temperature.$^{20-21}$ Although a number of synthetic routes are reported in the literature,$^{22-30}$ however, its phase pure synthesis is still a topic of interest. It has been also reported that the synthesis of FeBiO$_3$ by solid-state reaction at high temperature leads to residual impurities.$^{31}$

The current study is an effort to develop phase pure sunlight responsive active photocatalysts. The bimetallic photocatalyst, FeBiO$_3$, was synthesized by a user friendly approach using surfactant and the metallic ions (1:1 molar ratios) by hydrogel route. The
synthesized catalyst was characterized for optical, structural, electrochemical and morphological properties prior to photocatalytic experiments. The possible variations in the oxidation states of the individual metallic ions during the synthetic process was evaluated by XPS, whereas XRD and EDS were employed to verify the phase purity. Additionally, the phase purity was estimated by comparing the optical and structural properties with that of pure oxides. The stability of the synthesized catalyst was evaluated by cyclic voltammetry in the acidic and basic conditions in the dark and under illumination, whereas the resistance to the charge transfer and charge retention ability was estimated by EIS and chronopotentiometry, respectively. The photocatalytic activity of the synthesized catalyst was evaluated for the removal as well as mineralization of chemically stable and potential pollutants like 2-nitrophenol (2-NP) and 2-chlorophenol (2-CP). A number of chemical analysis tools such as HPLC, TOC, IC and GC-MS were employed to monitor the progress of degradation/estimation of intermediates, progress of mineralization, measurement of released ions and the identification of intermediates. The reusability of the synthesized catalysts was estimated by using the same catalyst for repeated scans.

**Experimental details**

The bimetallic oxide, FeBiO$_3$, was synthesized by hydrolysing the solution of equi-molar ratio Fe$^{3+}$ and Bi$^{3+}$ ions with 0.5 M KOH (Sigma-Aldrich). The ferric nitrate (Fe(NO$_3$)$_3$.9H$_2$O, Sigma-Aldrich) and bismuth nitrate (Bi(NO$_3$)$_3$.6H$_2$O, Sigma-Aldrich) were used as the precursors for Fe$^{3+}$ and Bi$^{3+}$ ions. Prior to hydrolysis, the solution containing metallic ions were stirred with Triton X-100 (used as surfactant). In a typical synthesis the appropriate amount of above mentioned precursors of Fe$^{3+}$ and Bi$^{3+}$ ions containing 0.1 mole of the each ions were dissolved and stirred with 200 ml of 0.1% Triton X-100/deionized water/nitric acid mixture for 2 h. Prior to the addition of 0.5M KOH, the solution was heated to 100$^\circ$C for half an hour. The solution containing the metal ions and surfactant was hydrolysed by the slow drop wise addition of 0.5M KOH solution. The addition continued with stirring (1500 rpm) till pH 9. The hydrogel was heated at 200$^\circ$C until the formation of fine precipitates. The excessive KOH and surfactant was removed by successive addition
of deionized water and decantation. At neutral pH, the precipitates were filtered and washed with 50:50 ethanol-acetone mixture to ensure the complete removal of surfactant. The collected precipitates were dried overnight, crushed and calcined in a muffle furnace at 500°C for 4 h at a heating and cooling rate of 10°C/min. The α-Fe₂O₃, used in the study for the comparison, was synthesized by adopting the procedure detailed above, whereas Bi₂O₃ was pre-synthesized. The detailed synthesis of Bi₂O₃ is mentioned in our previous publications.¹²⁻¹³

For the comparison of optical properties such as solid-state absorption and diffuse reflectance spectra, the measurements were recorded by a Perkin Elmer UV-visible diffuse reflectance spectrophotometer (Lambda 650) in the wide range of 190-900 nm. The direct bandgaps of the materials were evaluated by transforming the reflectance data to F(R), the Kubelka-Munk function. The emission (PL), FTIR and Raman shift spectra of the materials were acquired by a fluorescence spectro-fluorophotometer, RF-5301 PC, Shimadzu, Japan at an excitation wavelength of 360 nm, FTIR spectrometer, Affinity 1, Shimadzu, Japan in 400 to 4000 cm⁻¹ range using KBr pallet and a DXR Raman Microscope, Thermo Scientific, USA, equipped with 532 nm laser as the excitation source at 6mW power, respectively. The XRD patterns of the pre and as-synthesized powders were recorded by Scintag XDS 2000 diffractometer, equipped with a Cu Kα radiation source, in 10 to 80° range. The XPS profile of synthesized FeBiO₃ powder was assimilated in the binding energy range of 0 eV to 1100 eV by X-ray Photoelectron Spectrometer (PHI 5000 VersaProbe II, ULVAC-PHI Inc.). The oxidation states of Fe and Bi were estimated by applying curve fitting and comparing the data with standard values. The morphology of the synthesized powder was inspected by Field Emission Scanning Electron Microscope (FEI, Quanta FEG 450, Quorum Q150R ES, Quorum technologies Ltd. Ashford, Kent, England) at a voltage of 30kV.

The electrochemical behavior and stability of the synthesized FeBiO₃, in the dark and under illumination, was carried out by cyclic voltammetry (CV), electrochemical impedance spectroscopy (EIS) and chronopotentiometry (CP). For CV, EIS and charge-discharge analysis, a VSP multi-channel potentiostat (Bio-logic Science Instrument, USA) equipped with Ec-lab software and a three electrode system namely; glassy carbon (GC), platinum and Ag/AgCl saturated electrodes as working, counter and reference
electrodes, respectively. For the modification of GCE (working electrode), a sonicated dispersion of FeBiO$_3$ in chloroform was coated at the surface. The fitting of EIS Nyquist plots was performed by Zfit (Ec-lab software, Bio-logic Science Instruments, USA). Except for CV analysis which was performed by using 0.1 M (pH 9) phosphate buffer, 0.1 M (pH 7) phosphate buffer and 0.1 M (pH 2) phosphate buffer electrolyte to evaluate the stability under varying pH environment, all the other measurements were executed in 0.1M KCl solution. A 50 watt halogen lamp was used as a light source for the measurements under illumination.

The photocatalytic performance of FeBiO$_3$ was assessed in complete spectrum and visible region of sunlight exposure for the degradation/mineralization of 2-NP and 2-CP by exposing the catalysts/phenolic substrate suspension. Pyrex® glass reactor was utilized as a UV cutoff filter. The dimensions and surface area of the glass reactor were 15.5 (diameter) × 2.5 (height) and 189 cm$^2$, respectively. The optimization of the amount of catalyst was carried out for the removal of 2-NP. In a typical experiment, 150 ml of 2-NP (50 ppm) solution loaded with 10, 30, 70, 100, 200, 300 and 500 mg of the catalyst was exposed to sunlight for a period of 60 min. A linear increase in the degradation of 2-NP was observed till 100 mg of catalyst loading with no significant change afterwards. Based on the study, 100 mg/150 ml suspension was used in the proceeding studies. To establish the adsorption-desorption equilibrium between the catalysts and substrate, prior to exposure in sunlight, the suspension was kept in the dark for 60 min. During the exposure (1000±100 × 10$^2$ lx) the samples were drawn at regular intervals and subjected to HPLC (HPLC, (SPD-20A, Shimadzu Corporation, Japan), TOC (TOC-VCPH total carbon analyzer, Shimadzu Corporation, Japan and IC (Thermo scientific, USA, ion chromatograph, Dionex (ICS-5000 + EG) Eluent Generator) analysis for the monitoring of degradation, mineralization and estimation of released ions during the photocatalytic process.$^{32-36}$

**Results and discussion**

The comparison of the solid state absorption spectra α-Fe$_2$O$_3$, Bi$_2$O$_3$ and FeBiO$_3$ in the broad spectral range of 200 to 900 nm is presented in Fig. 1(a), where compared to pure Bi$_2$O$_3$, a significantly strong absorption of FeBiO$_3$ and α-Fe$_2$O$_3$, in the visible region.
(450-750 nm) is observable. Additionally, an inclination of the absorption pattern of FeBiO$_3$ towards that of $\alpha$-Fe$_2$O$_3$ was also witnessed however, instead of a sharp absorption edge, as observed for $\alpha$-Fe$_2$O$_3$, a consistently increasing absorption trend was perceived. Moreover, FeBiO$_3$ showed even better absorption in the visible region as compared to $\alpha$-Fe$_2$O$_3$. The strong absorption in the visible region by $\alpha$-Fe$_2$O$_3$ is attributed to the inter d-d transitions. In FeBiO$_3$, as $\sim$50% of Bi$^{3+}$ are replaced by Fe$^{3+}$, its properties are strongly influenced by Fe$^{3+}$. It seem quite obvious that the presence of Fe$^{3+}$ entities in the structure and the existence of d-d transitions overshadows the absorption spectra and a comparable rather higher absorption of FeBiO$_3$ was observable. The valence bands of the three powders i.e. $\alpha$-Fe$_2$O$_3$, Bi$_2$O$_3$ and FeBiO$_3$ are constituted by the O$^{2-}$ 2p orbitals, whereas the conduction bands are composed of Fe$^{3+}$ (3d), Bi$^{3+}$ (6p) and Fe$^{3+}$Bi$^{3+}$ (3d + 6p), respectively. These absorption edges represent the band gap excitations in the polymorphs. The graphical evaluation of the bandgap obtained by plotting ($F(R) \times h\nu$)$^{1/2}$ versus $h\nu$ (photon energy) of Bi$_2$O$_3$ is presented in Fig 1(b), whereas that of $\alpha$-Fe$_2$O$_3$ and FeBiO$_3$ are presented in the inset of Fig. 1b. The evaluated bandgaps of 2.1 eV, 2.15eV and 2.7 eV for $\alpha$-Fe$_2$O$_3$, FeBiO$_3$ and Bi$_2$O$_3$, respectively. A mild blue shift compared to the bandgap of $\alpha$-Fe$_2$O$_3$ and a significant red shift compared to that of Bi$_2$O$_3$ demonstrates the leading role of Fe$^{3+}$ states in establishing the band edge of FeBiO$_3$.

Fig. 2 shows the comparison of the normalized PL spectra of $\alpha$-Fe$_2$O$_3$, FeBiO$_3$ and Bi$_2$O$_3$ at an excitation wavelength of 350 nm and slit width of 5nm. It is important to mention here that compared to FeBiO$_3$, the emission intensities of $\alpha$-Fe$_2$O$_3$ and Bi$_2$O$_3$, under identical measurement conditions were significantly higher. The characteristic broad bands were observed that depicted the wide range of transitions in the materials. The maxima’s of the bands for $\alpha$-Fe$_2$O$_3$, FeBiO$_3$ and Bi$_2$O$_3$ at 580 nm, 570.5 nm and 469 nm corresponding to 2.07 eV, 2.21 eV and 2.63 eV, respectively were in accordance with the literature values. These emissions represented the Fe$^{3+}$3d-O$^{2-}$2p, Fe$^{3+}$Bi$^{3+}$ (3d+6p)-O$^{2-}$2p, and Bi$^{3+}$6p- O$^{2-}$2p de-excitations (conduction to valence band) in $\alpha$-Fe$_2$O$_3$, FeBiO$_3$ and Bi$_2$O$_3$, respectively. The appearance of a distinct maxima at a position different than that of the parent oxides ($\alpha$-Fe$_2$O$_3$ and Bi$_2$O$_3$) verified the phase purity of FeBiO$_3$ and additionally, from material and photocatalysis point of view, verified the rearrangement
of the conduction band composed of Fe\(^{3+}\) (3d) and Bi\(^{3+}\) (6p) states to a lower position thus lowering the bandgap of the material.

To ensure the mutual insertion of Fe\(^{3+}\) and Bi\(^{3+}\) in the crystal structure of FeBiO\(_3\) without the formation of other metal oxide offshoots, the FTIR analysis of Bi\(_2\)O\(_3\), FeBiO\(_3\) and α-Fe\(_2\)O\(_3\) was performed and compared in Fig. 3(a). Being rigid in nature, the information regarding the metal-oxygen-metal bond appear in the high energy region of 400-600 cm\(^{-1}\).\(^{40}\) The band at 541.99 cm\(^{-1}\) in the FTIR spectrum of Bi\(_2\)O\(_3\) corresponded to the stretching vibrations of Bi\(^{3+}\)-O-Bi\(^{3+}\) bond, whereas the corresponding vibrations of Fe\(^{3+}\)-O-Bi\(^{3+}\) and Fe\(^{3+}\)-O-Fe\(^{3+}\) bonds in FeBiO\(_3\) and α-Fe\(_2\)O\(_3\) appeared at 553.57 cm\(^{-1}\) and 578.64 cm\(^{-1}\), respectively. The order of the vibrational energy for the corresponding bonds seems quite logical as the insertion of Fe\(^{3+}\) ions in the crystal structure, being lighter in nature, reduces the rigidity of the system compared to that of Bi\(_2\)O\(_3\) thus vibrates at lower energy compared to that of Bi\(^{3+}\)-O-Bi\(^{3+}\) bond. On the contrary, the addition of Bi\(^{3+}\), being heavier than Fe\(^{3+}\) in the structure restricts the magnitude of vibrations therefore shifted to higher energy compared to Fe\(^{3+}\)-O-Fe\(^{3+}\) bond. The shifting in the position of the FTIR therefore validated the mutually shared oxygen positions by Fe\(^{3+}\) and Bi\(^{3+}\) ions in FeBiO\(_3\). The Raman spectroscopy being contemporary to FTIR is also regarded as a powerful tool in estimating the probable structural changes. The comparison of the Raman shift spectra of Bi\(_2\)O\(_3\), FeBiO\(_3\) and α-Fe\(_2\)O\(_3\) is presented in Fig. 3(b), where compared to that of Bi\(_2\)O\(_3\) and α-Fe\(_2\)O\(_3\), a significant shifting in the peak positions for FeBiO\(_3\) is observable. The major peaks representing the A\(_1\) modes for FeBiO\(_3\) appeared at 158.95 cm\(^{-1}\) and 490.64 cm\(^{-1}\), whereas the E modes were identified at 210.05 cm\(^{-1}\), 277.55 cm\(^{-1}\), 391.33 cm\(^{-1}\), 533.07 cm\(^{-1}\) and 588.03 cm\(^{-1}\). These values were in close agreement with that already mentioned in the literature with the discussion on their origin.\(^{41}\) The mismatch between the peaks of FeBiO\(_3\) with that of α-Fe\(_2\)O\(_3\) and Bi\(_2\)O\(_3\) as well as the shifts in the peak positions verified the co-existence of Fe\(^{3+}\) and Bi\(^{3+}\) in the structure.

Fig.4 shows the comparison of the XRD patterns of Bi\(_2\)O\(_3\), FeBiO\(_3\) and α-Fe\(_2\)O\(_3\). The major reflections of α-Fe\(_2\)O\(_3\) appeared at 2θ values of 24.14°, 33.14°, 35.65°, 40.86°, 49.57°, 54.15°, 62.40° and 64.04°, whereas the reflections of appreciable intensity for
Bi$_2$O$_3$ appeared at 25.85°, 27.45°, 28.15°, 32.2°, 33.25°, 35.20°, 37.15°, 42.35°, 46.30°, 52.37° and 61.88°. The XRD pattern of α-Fe$_2$O$_3$ corresponded to rhombohedral geometry with the crystallite size of ~40.2 nm matched with JCPDS # 33-0664. A monoclinic geometry (JCPDS # 41-1449) was assessed for Bi$_2$O$_3$, whereas a crystallite size of ~36.2 nm was calculated by using the parameters of the most intense reflection at 27.45°. The prominent reflections of FeBiO$_3$ were experiential at the 2θ values of 22.45°, 31.95°, 39.55°, 45.55°, 51.71°, 57.10°, 67.25°, 71.17° and 76.09° that corresponded to (101), (110), (021), (202), (113), (300), (220), (303) and (312) indices of rhombohedral structure (JCPDS# 20-0169). An average crystallite size of ~19.2 nm was calculated using the intensity, d-spacing and FWHM values of the intense reflections. The intensity and sharpness of the peaks reflected the well-defined crystal structure of FeBiO$_3$ with high crystallinity.

The XPS wide angle survey scan of as-synthesized FeBiO$_3$ is presented in Fig. 5(a), where the peaks against the corresponding binding energies of the core and splitted levels of Bi, Fe and O are observable. The C1s peak was observed at 285.02 eV. The estimation of the probable variations in the oxidation states of the individual components during the synthesis process, especially Fe and Bi, was performed by the identification of the submerged peaks by deconvolution and curve fitting. The raw and fitted peaks for Bi4f splitted levels are presented in Fig. 5(b). The deconvolution of the Bi4f located the additional low intensity peaks at 157.5 eV (4f$_{7/2}$) and 162.8 eV (4f$_{5/2}$) besides the intense peaks at 159.5 eV (4f$_{7/2}$) and 164.7 eV (4f$_{5/2}$). Such type of behavior is not reported in the literature as the existence of higher oxidation states such as Bi$^{5+}$ leads to the positive increment in the binding energy. Therefore, based on the evaluation of all the possibilities, it was established that the change in the chemical environment resulted in the shifting of the Bi$^{3+}$ peaks to lower energy. The splitted peaks of 4f$_{7/2}$ and 4f$_{5/2}$ peaks due to Bi$^{3+}$-O-Bi$^{3+}$ linkage appeared at 159.5 eV and 164.7 eV, respectively, whereas the Fe$^{3+}$-O-Bi$^{3+}$ shifted the same to lower energy values of 157.5 eV and 162.8 eV, respectively. The fitted Fe2p splitted levels are presented in Fig. 5(c). The peaks of Fe2p$_{1/2}$ and Fe2p$_{3/2}$ levels at 726.23 eV and 712.35 eV, respectively, verified the presence of Fe in 3+ oxidation state. Additionally, the presence of Fe$^{3+}$ characteristic satellite at 718.7 eV and the absence of additional peaks further authenticated the findings. As
presented in Fig. 5(d), the fitting of O1s spectra resolved in three peaks of variable intensities. The peak at ~532.3 eV characterize the chemisorbed surface hydroxyl groups, whereas the main peak at ~530.28 eV represent the skeletal oxygen, whereas the peak due to the singly charged surface O- groups was observed at low energy position of ~528.52 eV.

The comparison of the FESEM images of as-synthesized FeBiO$_3$ at 100,000×, 200,000×, 300,000× and 400,000× are presented in Fig.6 (a-d), respectively. A narrow particle size distribution with the uniformity in the shape was observable that validated the applicability of the used synthetic route and significance of surfactant in controlling the morphology along with other physical factors such as temperature, rate of KOH addition etc. The evaluated crystallite size of ~19.2 nm was also validated by the FESEM analysis. It was also noticed that the larger particles (> 40 nm) were the aggregates of the smaller ones. In the high resolution images (Fig. 6(d)), the particles with spherical shapes are also noticeable. Fig. 7 shows the location of the targeted or focused point for the EDS elemental analysis of the synthesized FeBiO$_3$, whereas the EDS spectra and the elemental composition of the powder is presented in the inset. The analysis revealed the presence of Fe and Bi in 1:1 and also verified the phase purity of the material.

The stability of the synthesized FeBiO$_3$ was examined under various pH conditions i.e. acidic (pH = 2), neutral (pH = 7.0) and basic (pH = 9.0) by cyclic voltammetry (CV) in the dark and under illumination and the comparison is presented in Fig. 8(a) and (b). It is important to mention here that two scans were recorded both in the dark and under illumination however, the results of the 1$^{\text{st}}$ cycle in the dark and under illumination are compared. In 0.1 M pH 2 phosphate buffer electrolyte, two distinct reduction peaks appeared at -0.19 V vs. RHE (-0.51 V vs Ag/AgCl) and -0.57 V vs RHE (-0.89 V vs Ag/AgCl) in the forward cathodic direction both in the dark and under illumination that indicated the dissolution or surface erosion of the powder supported on GCE. The 1$^{\text{st}}$ peak at -0.19 V vs. RHE (-0.51 V vs Ag/AgCl) was attributed to the electro-reduction of surface Bi$^{3+}$-O$^{-}$ type species in the acidic environment leading to the formation of BiO$_2^{-}$ species, whereas the 2$^{\text{nd}}$ peak was assigned to the further reduction of dissolved species BiO$_2^{2-}$ and possible bulk deposition of released Bi. The results similar to our findings are already reported in the literature.$^{43-45}$ In the backward anodic direction,
an oxidation peak due to the oxidation of Fe$^{2+}$ into Fe$^{3+}$ was appeared at +0.29 V vs. RHE (-0.03 V vs Ag/AgCl). The Fe$^{2+}$ ion formed in the forward cathodic scan from the reduction of surface Fe$^{3+}$ however, the broad reduction peak of Bi$^{3+}$-O$^-$ species obscured the reduction peak of Fe$^{3+}$. In the II$^{\text{nd}}$ cycle, probably due to the deposition of released Bi at the electrode surface, the electrode surface was passivated, therefore no reduction peaks were observed in the dark. Additionally, the anodic and cathodic peak currents were substantially decreased. A similar behavior was observed under illumination, however, the reduction peak currents were significantly enhanced as compared to that of dark conditions.

Contrary to acidic medium (pH = 2), in the neutral 0.1 M pH 7 phosphate buffer, the susceptibility of FeBiO$_3$, both in the dark and under illumination, towards reduction was dramatically reduced and a drastic decrease in the cathodic peak intensities elaborated its better stability in the neutral medium. However, both in the dark and under illumination, in the anodic region, two low intensity broad oxidation peaks were noticed at + 0.11 V vs. RHE (-0.5 V vs Ag/AgCl) and +0.45 V vs. RHE (-0.16 V vs Ag/AgCl). As no significant change in the peak intensity was noticed under illumination, the formation of the respective metallic hydroxides under external bias was identified as the source in this regard. In the II$^{\text{nd}}$ cycle, the reduction of dissolved Bi(OH) was appeared at -0.28 V vs. RHE (-0.89 V vs Ag/AgCl) in the dark. The decreased intensity of the oxidation peaks and the absence of the reduction peak under illumination further strengthen the view of the stability of FeBiO$_3$ under illumination.

In basic medium (0.1 M pH 9 phosphate buffer), the background current at positive potential was significantly higher that suppressed the reduction peak current intensities considerably, therefore, the CV’s were recorded in the potential range of +0.73 V vs RHE to -0.27 V vs RHE (0 V vs Ag/AgCl to -1.0 V vs Ag/AgCl). In the forward cathodic scan, three barely detectable reduction peaks were emerged at +0.23 V vs. RHE (-0.5 V vs Ag/AgCl) due to reduction of bismuth oxide, at +0.06 V vs. RHE (-0.67 V vs Ag/AgCl) for the reduction of Fe$^{3+}$ and at -0.21 V vs. RHE (-0.94 V vs Ag/AgCl) for the deposition of released bismuth. No clear oxidation peak appeared in the backward anodic direction. The disappearance of reduction peaks in the 2$^{\text{nd}}$ cycle confirmed the deposition of the released Bi at the electrode surface. Compared to dark, the higher peak currents
intensities under illumination, were attributed to the increased conductivity of the films. The analysis of the CV patterns revealed the adequate stability in neutral and basic pH, whereas FeBiO$_3$ encountered dissolution of both metallic components (Fe and Bi) in harsh conditions of acidic pH.

The comparison of the alternating current impedance measurements of FeBiO$_3$/GCE, in the dark and under illumination, is presented in Fig. 8(c). A measurable decrease in the charge transfer resistance of as-synthesized FeBiO$_3$ was observable both in low and high frequency range as shown in Fig. 8(d) and (e) that depicted the increased conductivity or better charge transfer ability under illumination. In terms of photocatalytic activity, this observation may be correlated to the expected better mobility of charges in the material under illumination.\textsuperscript{17,33} The charge-discharge behavior of FeBiO$_3$/GCE was evaluated at 0.1 A/m$^2$ current density. The enhanced magnitude of the discharging curve, under illumination demonstrated the better charge retaining ability of the material. The charge discharge profiles are shown in Fig. 8(f).

The photocatalytic performance of synthesized FeBiO$_3$ was evaluated for the removal (degradation/mineralization) of 2-NP and 2-CP in complete spectrum and visible region sunlight exposure. The adsorption of the phenolic substrates on as-synthesized FeBiO$_3$ was estimated by retaining the catalyst/substrate suspension in the dark for 1h before the exposure. The FeBiO$_3$ exhibited significantly high adsorption of ~21% and ~17% of 2-NP and 2-CP, respectively. The comparison of the percentage degradation of 2-NP, extracted from the decrease in peak height of 2-NP in HPLC chromatograms, as a function of sunlight (complete spectrum/visible region) is presented in Fig. 9(a). In the initial 30 min of complete spectrum sunlight exposure, ~46% of 2-NP was removed whereas ~15% was removed in visible region of sunlight exposure for the same period. Although ~94% of 2-NP was removed in 120 min, however the complete removal was witnessed in 240 min of complete spectrum sunlight exposure. Probably the scarce access of ROS to the 2-NP substrate molecules due to the decreased concentration is the possible reason for this effect. In 240 min of visible region exposure, ~80% of the substrate was degraded. Although the degradation process was slow in the visible region, however a consistent decrease in the concentration of 2-NP was noticed with the increasing exposure time. The rate constants (k) for the removal of 2-NP under both the conditions
of exposure were estimated by applying the Langmuir-Hinshelwood kinetic model for pseudo first order reactions. The evaluation extracted by plotting $\ln(C_0/C)$ versus the sunlight exposure time is presented in Fig. 9(b). The kinetic model was validated under both the experimental conditions and the rate constants of 0.023 min$^{-1}$ and 0.0059 min$^{-1}$ for the removal in complete spectrum and visible sunlight exposure, respectively. The significant difference in the rate of removal of 2-NP may be attributed to the contribution of the $\leq 5\%$ UV portion in the complete spectrum sunlight. The other plausible is the blocking of the incident visible photons by 2-NP that has extended absorption band in the visible as well as UV region.

The comparison of the time-scale percentage degradation of 2-CP in the complete spectrum and visible region exposure of sunlight is presented in Fig. 9(c). The degradation of 2-CP in the initial 30 min of the exposure in complete spectrum sunlight was comparable to that of 2-NP, whereas marginally higher i.e. $\sim 19\%$ compared to $\sim 15\%$ for 2-NP, in the visible region exposure. With the passage of time, a faster degradation of 2-CP, compared to 2-NP, under the both exposure conditions, was noticed. The probable reason for this effect is the absorption of photons by the 2-NP substrate that is extends from the visible region to UV region, whereas 2-CP has the weak absorption in the UV region below 300 nm. The absorption of photons by the substrate (2-NP) reduces the accessibility of the photons to the catalyst particles. After 240 min of exposure $\sim 100\%$ and $\sim 85\%$ of 2-CP was degraded in complete spectrum and visible region sunlight exposure, respectively. The graphical evaluation of the rate constants ($k$) is presented in Fig. 9(d). The evaluated rate constants for the degradation of 2-CP was 0.025 min$^{-1}$ and 0.0071 min$^{-1}$. Additionally, the Langmuir-Hinshelwood kinetic model was also validated with acceptable correlation.

The comparison of the mineralization (TOC removal) of 2-NP and 2-CP in the complete spectrum and visible region sunlight exposure is presented in Fig. 10 (a). The synthesized FeBiO$_3$ efficiently mineralized both the substrates in the complete spectrum sunlight exposure and reduced the initial TOC contents of 15.6 ppm and 16.8 ppm for 30 ppm of 2-NP and 2-CP to 0.62 ppm and 0.25 ppm, respectively, in 240 min of exposure. The graphical evaluation of rate constants for TOC removal of 2-NP and 2-CP in the exposure of complete spectrum and visible region of sunlight is presented in Fig. 10 (b).
The markedly lower rate of mineralization compared to that of degradation clearly reflected multistep nature of mineralization as compared to degradation process. Additionally, the rapid removal of the substrates also proposes the ease of degradation as compared to mineralization that is accomplished by any major or minor structural change in the substrate changing its identity as a new compound. Therefore, the degradation does not essentially mean the removal rather it is the conversion of substrate into new entity. The time-scale HPLC profiles of all the samples and GCMS analysis of the selected sample confirmed the formation of intermediates during the degradation process. The intermediates are further interacted by the oxidants leading to the formation of smaller fragments and later conversion to CO₂ in successive interactions. The pictorial representation of the above discussion is presented in scheme I.

Scheme I: The plausible mechanism of 2-CP and 2-NP degradation over FeBiO₃ in sunlight exposure.

The fate of the released ions during the degradation/mineralization process was monitored by ion chromatography (IC). Contrary to the expected Cl⁻ and NO₂⁻ ions in the
removal of 2-CP and 2-NP, respectively, ClO\textsuperscript{3-} and NO\textsubscript{3} ions were observed under the subjected instrumental conditions. The presence of these ions was validated by successive analysis of the standards. As estimated by the comparison of the peak heights of the standards and samples, the time-scale profiles of ClO\textsuperscript{3-} and NO\textsubscript{3} ions in complete spectrum and visible region illumination of sunlight are presented in Fig. 11(a). Interestingly, the concentration of these ions was significantly lower than the expected theoretical values. This situation led to the conclusion that the Cl\textsuperscript{-} and NO\textsubscript{2} ions are further interacted by either the photon generated ROS or the reactive sites such as holes (h\textasciitilde) and being transformed to the respective gases, Cl\textsubscript{2} and NO\textsubscript{2}, that escape from the reactor during the course of degradation process. It has been proposed that NO\textsubscript{2} ions undergo photolysis to NO\textsubscript{2}\textbullet in the aqueous system with the generation of hydroxyl radicals. The plausible routs are proposed as below.\textsuperscript{46}

\[
NO_2^- + H^+ + h\nu \rightarrow NO_2^* + HO^* \quad (1)
\]

\[
NO_2^- + HO^* \rightarrow NO_2^* + HO^- \quad (2)
\]

The proposed mechanism indicate that the nitrite ions simultaneously act as producers and consumers of hydroxyl radicals under illumination. The above mechanisms seems equally applicable on the photocatalytic systems as all the species involved are either already present in the system or generated under illumination. Another possibility that we have already discussed in our previous communication,\textsuperscript{5} is the direct donation of electrons by the nitrite ions to the photon generated holes at the surface of photocatalyst.

\[
NO_2^- + h^+ \rightarrow NO_2 \uparrow \quad (3)
\]

The direct evidence for the escape of the NO\textsubscript{2} gas during the degradation process was the formation of brownish-yellow vapors. It is important to mention here that in some of our studies with ZnO, we observed the presence of nitrite ions in substantial
Based on equation 1, it was presumed that the presence of H\(^+\) ions in the solution is vital for the conversion of nitrite ions into NO\(_2\) gas. This hypothesis was verified by measuring the pH during the course of reaction which revealed that the pH of the system persisted in the acidic range i.e. 5.5 to 6.5. These observations validated the proposed mechanism in equation 1. The same analogy is equally probable for the conversion of chloride ions into either Cl\(_2\) or HCl gas in the acidic pH conditions.

The reusability of the synthesized catalyst was evaluated for the removal of 2-NP in the complete spectrum sunlight exposure. The comparison of the three successive scans performed with the same catalyst with fresh 30 ppm 2-NP solution is presented in Fig. 11(b). The catalyst showed acceptable sustained performance and no significant decrease in the activity in the repeated scans.

**Conclusions**

The study revealed the suitability of the adopted procedure for the phase pure and morphology controlled synthesis of FeBiO\(_3\). The presence of both Fe and Bi in the crystal structure of FeBiO\(_3\) imparted better structural, electrochemical and optical properties that classified it as a useful addition for sunlight photocatalytic decontamination applications. The catalyst demonstrated significantly high activity for the removal (degradation as well as mineralization), of 2-NP and 2-CP both in complete spectrum and visible sunlight exposure. The ions released as a consequence of mineralization are further interacted by photon generated reactive oxygen species and released as respective gases.

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References


Figure Captions

Fig. 1  (a) The comparison of the solid-state absorption spectra of FeBiO$_3$, α-Fe$_2$O$_3$ and Bi$_2$O$_3$ in 200-900 nm range, (b) the graphical evaluation of direct band edge position of Bi$_2$O$_3$, and the inset of (b) shows the same for FeBiO$_3$ and α-Fe$_2$O$_3$.

Fig. 2  The comparison of the normalized PL spectra of α-Fe$_2$O$_3$, FeBiO$_3$, and Bi$_2$O$_3$ at an excitation wavelength of 325 nm and slit width of 5nm.

Fig. 3  The comparison of the (a) FTIR and (b) normalized Raman spectra of Bi$_2$O$_3$, FeBiO$_3$, and α-Fe$_2$O$_3$.

Fig. 4  The comparison of the XRD patterns of Bi$_2$O$_3$, FeBiO$_3$ and α-Fe$_2$O$_3$. The dotted lines shows the major reflections of FeBiO$_3$.

Fig. 5  The XPS analysis of FeBiO$_3$ (a) survey scan (b) Bi4f splitted core levels (c) Fe2p splitted core levels and (d) O1s splitted levels.

Fig. 6  The comparison of the FESEM images of FeBiO$_3$ at various resolutions (a) 100,000× (b) 200,000× (c) 300,000× and (d) 400,000×.

Fig. 7  The EDS analysis of the synthesized FeBiO$_3$. The inset shows the EDS spectra and the elemental composition of the components of the sample.

Fig. 8  The electrochemical evaluation of FeBiO$_3$ in the dark and under illumination (a & b) The CV’s at various pH values (c) EIS Nyquist plot of FeBiO$_3$ in dark and under illumination (d & e) exploded view of high and low frequency regions of EIS spectra (f) charge-discharge profile of FeBiO$_3$ at neutral pH.

Fig. 9  The comparison of the percentage degradation and rates of degradation of 2-NP (a, b) and 2-CP (c, d) in the exposure of complete spectrum and visible region of sunlight in the presence of FeBiO$_3$.

Fig. 10  The comparison of (a) TOC removal (ppm) and (b) graphical evaluation of rate constants for TOC removal of 2-NP and 2-CP (c, d) in the exposure of complete spectrum and visible region of sunlight in the presence of FeBiO$_3$.

Fig. 11  (a) The comparison of the major released ions during the degradation of 2-NP and 2-CP in the exposure of complete spectrum and visible region of sunlight in the presence of FeBiO$_3$. (b) Evaluation of the reusability of FeBiO$_3$ for the degradation of 2-NP in complete spectrum sunlight exposure.
Fig. 1  (a) The comparison of the solid-state absorption spectra of FeBiO$_3$, $\alpha$-Fe$_2$O$_3$ and Bi$_2$O$_3$ in 200-900 nm range, (b) the graphical evaluation of direct band edge position of Bi$_2$O$_3$, and the inset of (b) shows the same for FeBiO$_3$ and $\alpha$-Fe$_2$O$_3$. 
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The diagram illustrates the reaction pathways involving oxygen reduction, water oxidation, and the formation of intermediates. The reaction starts with oxygen reduction leading to the formation of $\text{O}_2^-$, which further reacts with intermediates such as 2-CP (2-Chlorophenol) and 2-NP (2-Nitrophenol). Water oxidation produces $\text{H}_2\text{O}^+$, which leads to the formation of $\text{HO}^- + \text{H}^+$. The overall process results in the production of $\text{CO}_2$ and $\text{H}_2\text{O}$.