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1 **The effects of *P. aeruginosa* ATCC 9027 and NTA on phytoextraction of Cd by**
2 **ramie (*Boehmeria nivea* (L.) Gaud)**

3 Jieli Xie^{a,b}, Yunguo Liu^{a,b,*}, Guangming Zeng^{a,b}, Huan Liu^{a,b}, Bohong Zheng^c, Hui Tang^{a,b}, Weihua
4 Xu^{a,b}, Zhichao Sun^{a,b}, Xiaofei Tan^{a,b}, Jian Nie^{a,b}, Zhengjiang Jiang^{a,b}, Chao Gan^{a,b}, Shufan Wang^{a,b}

5

6 ^a College of Environmental Science and Engineering, Hunan University, Changsha 410082, P.R. China

7 ^b Key Laboratory of Environmental Biology and Pollution Control (Hunan University), Ministry of
8 Education, Changsha 410082, P.R. China

9 ^c School of Architecture and Art Central South University, Central South University, Changsha 410082,
10 PR China

11 *Corresponding author: Yunguo Liu; Tel.: + 86 731 88649208; Fax: + 86 731 88822829;

12 E-mail address: xjlhnu@163.com

13

14 Abstract

15 In pot experiments, the effects of *Pseudomonas aeruginosa* ATCC 9027 and
16 nitrilotriacetic acid (NTA) on Cd phytoextraction from contaminated soil by
17 *Boehmeria nivea* (L.) Gaud (ramie) was investigated. Ramie was grown in a sandy
18 soil in the presence of 30 mg kg⁻¹ Cd and 50 mg kg⁻¹ Cd, respectively. Experimental
19 pots were amended with *P. aeruginosa* ATCC 9027 or NTA at different levels (5, 10
20 and 20 mmol kg⁻¹) weekly. The results showed that the inoculation of *P. aeruginosa*
21 ATCC 9027 alleviated the Cd-induced damages, resulting in promotion of ramie
22 growth, improvement of antioxidative enzymes activities and increase of total
23 Cd-uptake by ramie. By contrasting 30 and 50 mg kg⁻¹ Cd treatments, the inoculation
24 of *P. aeruginosa* ATCC 9027 increased accumulation in root ranging from 54% to 96%
25 and 13% to 104% in 30 and 50 mg kg⁻¹ Cd soils, respectively. The average
26 accumulation of Cd with *P. aeruginosa* ATCC 9027 was about 1.95-fold (30 mg kg⁻¹
27 Cd) and 1.54-fold (50 mg kg⁻¹ Cd) compared to the corresponding NTA treatments.
28 When added with NTA, the accumulation of Cd in shoot of ramie was higher than the
29 controls, but the inhibition of plant growth and related enzyme activities were
30 observed. The experimental results demonstrated that *P. aeruginosa* ATCC 9027 can
31 greatly enhance phytoremediation efficiency. Besides, the results also indicated that *P.*
32 *aeruginosa* ATCC 9027 was suitable than NTA to improve the efficiency of ramie
33 under cadmium stress in practical applications.

34 **Keywords:** Phytoextraction; *Pseudomonas aeruginosa* ATCC 9027; Nitrilotriacetic
35 acid; Ramie; Cadmium

36 1. Introduction

37 Cadmium (Cd) is a major anthropogenic pollutant derived from agricultural and
38 industrial activities, including wastewater irrigations, mining and smelting of
39 metalliferous ores.¹ Due to its non-degradability, chemical mobility and high toxicity

40 to biota, Cd can transfer through food chains and then cause various diseases to plants,
41 animals and even human beings.² Given that Cd contamination has posed an
42 unprecedented threat to a wide range of ecosystem and human health, more and more
43 attention has been globally focused on the mechanisms of Cd contamination and
44 remediation technologies.^{3,4}

45 Phytoremediation, a technology of applying vegetations to remediate
46 contaminated soils, is generally considered as a low-cost, eco-friendly approach
47 which has gained considerable interests worldwide.^{5, 6} Although, large amounts of
48 plant species could hyperaccumulate heavy metals in their tissues, there still exist
49 limitations of phytoremediation in practice such as a lower effectiveness than
50 mechanical methods, phytotoxicity, low biomass production and limited contaminant
51 absorption.⁷ Given this, the success of phytoremediation of heavy metals depends not
52 only upon the potential of the plants' tolerance to high concentrations of heavy metals,
53 but also upon a large plant biomass.⁸ In fact, the accumulation effect and tolerance of
54 the plant still need to be strengthened in the actual repair applications, and adding
55 exogenous substances gradually became the focus of the phytoremediation in recent
56 years. Several chemical amendments, including ethylene diamine tetracetic acid
57 (EDTA), citric acid (CA) have been used to promote either phytostabilization or
58 phytoextraction process.⁹ As is known, EDTA is proved to be the most effective
59 chelating agent, which is widely applied to remediate heavy metal contaminated
60 soil.¹⁰⁻¹² However, due to the low biodegradability and high solubility, EDTA leads to
61 high environmental risk of heavy metal leaching to groundwater.¹³ In order to
62 construct a clean and environmental friendly remediation in practical applications,
63 biodegradable chelants and metal-tolerant plant-microbe have been the objective of
64 particular attention. Therefore, selection of suitable chelants for the solubilization of
65 heavy metals must be the first issue to be considered to increase extraction efficiency.

66 Recently, the focus of researches on chelant-enhanced phytoextraction has been
67 shifted to some biodegradable chelating agent such as nitrilotriacetic acid (NTA),

68 which has been used as detergents in the last 50 years. NTA can improve the uptake
69 of metals by plants and limit leaching of metal into deeper soil.¹⁴ Several studies have
70 been performed using NTA as a ligand to improve the efficiency of metal
71 phytoextraction. As reported early, NTA performed effectively in desorbing Cu, Pb
72 and Zn from soils, increasing Cu, Pb and Zn uptakes in shoots of *Festuca*
73 *arundinacea*, and improving Cd accumulation and translocation in *Siegesbeckia*
74 *orientalis* L.¹⁵⁻¹⁷ Nevertheless, little information is available about the addition of
75 NTA to ramie under cadmium stress. In addition, in the remediation of contaminated
76 soil, another promising alternative to amendments could be the utilization of
77 microbe-mediated processes, because the microbial metabolites in the rhizosphere can
78 facilitate plant metal uptake by altering the bioavailability and mobility.⁵
79 growth-promoting bacteria can be exploited to facilitate phytoremediation.^{18, 19}
80 Besides, Plant-associated bacteria can accelerate metal uptake and plant growth due
81 to its feasibility of microorganisms for bioaccumulating metals from contaminated
82 soils or its influences on metal mobilization/immobilization.²⁰ In addition, compared
83 with some chemical amendments living around the plant surface, the microbial
84 metabolites are more biodegradable, and less toxic and the microbes may be possible
85 to produce plant growth substances such as siderophores,
86 1-aminocyclopropane-1-carboxylic acid (ACC) deaminase, and these substances
87 improve the growth of the plant in metal contaminated soils.^{21, 22} Combination of
88 microbes and plants has been applied to the cleanup of contaminated soils.^{5, 22} Despite
89 a large number of literatures concerning the application of bacteria or endophytes in
90 various plants, little information is available on the response of *P. aeruginosa* ATCC
91 9027 of ramie under Cd stress. As a consequence, researches about the effect and
92 mechanism of the application of *P. aeruginosa* ATCC 9027 and NTA is urgently
93 needed.

94 *Boehmeria nivea* (L.) Gaud (ramie) was applied as the study plant which is a
95 Cd-tolerant species with large biomass and fast growth rate.²³ Although there are

96 some previous researches concerning the response of ramie to Cd toxicity in
97 hydroponic condition, little information is available on the Cd accumulation and
98 tolerance mechanism of ramie in the presence of microbe and NTA.²⁴ The main
99 objective of the research was (i) to investigate the potential ability of ramie in
100 enhancing phytoremediation of Cd by application of *P. aeruginosa* ATCC 9027 and
101 NTA; (ii) to explain the influence of *P. aeruginosa* ATCC 9027 and NTA on
102 phytoremediation by analyzing physiological parameters and relevant enzymatic
103 antioxidants of ramie; (iii) to compare *P. aeruginosa* ATCC 9027 and NTA,
104 choosing a better way on practical application in phytoremediation of cadmium
105 polluted soils in the future.

106 **2. Materials and methods**

107 **2.1. Experimental design and treatments**

108 The experimental soil was collected from the superficial layer (depth: 0–25 cm)
109 originating from Taozi Lake, which located at Hunan University (Changsha, China).
110 Soil samples were air-dried, ground, and sieved to < 2 mm prior to use. The
111 physicochemical properties of experimental soil are shown in Table 1. Then the soil
112 was uniformly spiked with 30 and 50 mg Cd kg⁻¹ (from solutions of Cd(NO₃)₂·4H₂O)
113 respectively. After incubated for 1 month, each 2 kg of the treated soil was filled into
114 3L plastic pots. Ramie seeds were obtained from Chinese Academy of Agricultural
115 Sciences, Hunan, China. After the ramie seedlings acclimated for 1 week, plants were
116 inoculated with the gram-negative strain *P. aeruginosa* ATCC 9027 (S1, S2, S3). S1
117 represented the addition of strain one time, while S2 represented twice and S3
118 represented three times, each added for one week apart. Simultaneously, 1 week
119 before harvested, the pots were correspondingly treated with different concentrations
120 of NTA (0, 5, 10, and 20 mmol kg⁻¹ soil in a 200 mL solution) to the surface of the
121 soil. The plants were performed in triplicates and conducted in a completely
122 randomized design following fifteen treatments in 30 mg kg⁻¹ Cd (Cd30) and 50 mg

123 kg⁻¹ Cd (Cd50) soil. The fifteen treatments were detailed presented in Table 2. All of
124 the experiments were carried out in naturally illuminated greenhouse with 16h light
125 period at a minimal light intensity of 300 $\mu\text{mol m}^{-2}\text{s}^{-1}$, temperature of 25 ± 3 °C and
126 60–70% relative humidity. The ramie seedlings were cultivated in Cd contaminated
127 soil for 3 weeks. After harvested, the plants were separated into roots, stems and
128 leaves, and frozen immediately in -80 °C for further analysis.

129 **2.2. Microbial culture**

130 The gram-negative strain *P. aeruginosa* ATCC 9027 which used as foreign
131 substances in the work was procured from the American Type Culture Collection. It
132 was maintained on 3–4 °C peptone agar slants and activated at 30 °C before used. *P.*
133 *aeruginosa* ATCC 9027 inoculums from peptone agar slant were transferred to 50 mL
134 mineral salt medium (MSM) with 0.5 g L⁻¹ yeast extract in a 250 mL Erlenmeyer
135 flask and performed at 37 °C on a gyratory shaker at 200 rpm for 24 h.²⁵ Then 5 mL
136 of the enriched cell suspension was further transferred to 100 mL MSM in 500 mL
137 Erlenmeyer flask containing 20g L⁻¹ glucose as the sole carbon source.²⁵ This
138 inoculated culture medium was grown at 37 °C for 72 h under shaking conditions
139 (200 rpm) which was composed of 5.0 g L⁻¹ NH₄Cl, 0.5 g L⁻¹ MgSO₄·7H₂O, 2.5 g
140 L⁻¹ K₂HPO₄, 5.0 g L⁻¹ Na₂HPO₄, with a pH of 6.8.²⁶ The bacterial cells were
141 collected by centrifugation (7000 rpm) at 4 °C for 15 min, washed twice with
142 physiological water and obtained an inoculum with approximate absorbance value
143 (OD₆₀₀) of 0.6 (approximate 10⁸ CFU⁻¹ mL).²⁷ 5 mL of this strain was used for
144 inoculation of each pot.

145 **2.3. Metal analysis**

146 Upon harvesting, the samples were washed with deionized water and the roots
147 were then rinsed with 5 mM CaCl₂ for approximately 5 min to remove the metals
148 absorbed.¹³ The plants were separated into roots, stems, and leaves. Then the samples

149 were oven-dried, milled, and digested with a mixture of HNO₃/ HClO₄ (3:1, v/v) by
150 graphite digestion instrument (SISP, DS-360, China). The Cd concentration of each
151 solution was determined using flame atomic absorption spectroscopy (FAAS)
152 (AAAnalyst 700, Perkin Elmer, USA). The translocation factor (TF) is defined as the
153 total metal content in plant from root to shoot.²⁸

154 **2.4. Determination of chlorophyll and malondialdehyde (MDA) content**

155 The chlorophyll content of ramie leaf was determined using the acetone method.
156 ²⁹ Frozen leaf tissues were homogenized in 80% ice-cold acetone in dark and then
157 centrifuged at 2000 rpm for 10 min. Then, chlorophyll content was determined
158 spectro-photometrically on the supernatant at wavelength of 646 nm and 663 nm.²⁹

159 The MDA content of leaves was determined using the thiobarbituric acid (TBA)
160 method.³⁰ Frozen leaf tissues (0.5 g) were homogenized with 10 mL 10% (w/v)
161 trichloroacetic acid (TCA). The homogenate was centrifuged at 10,000 rpm for 10
162 min. Then 2 mL of the aliquot of the supernatant and 2 mL of 10% TCA containing
163 0.5% (w/v) TBA were added. The mixtures were incubated at 95 °C for 30 min and
164 then cooled quickly in an ice-bath. The samples were centrifuged at 10,000 rpm for 15
165 min and the absorbance of the supernatant was measured at 532 nm and corrected for
166 nonspecific absorbance at 600 nm. The concentration of MDA was calculated using
167 155 mM⁻¹ cm⁻¹ as extinction coefficient.

168 **2.5. Soil enzyme activities**

169 Urease activity was determined according to the method suggested by Tabatabai
170 and Bremner (1972), using 5.00g of soil (d.w.).³¹ Triplicate samples of air-dried soil
171 were measured to mix with 1ml toluene for 15 min. Afterwards they were added to
172 10ml of 10% urea and 20ml citrate buffer (pH 6.7), and then set in the incubator at
173 38 °C for 24h. After the incubation, the mixtures were immediately filtered. The
174 filtrate was measured to determine urease activity. The absorbance was analyzed in

175 the supernatant at 578 nm and expressed as NH_4^+-N ($\text{mg kg}^{-1} \text{ h}^{-1}$).³²

176 **2.6. Enzymatic analyses**

177 The activity of antioxidant enzyme superoxide dismutase (SOD) and catalase
178 (CAT) was determined with an assay kit purchased from Nanjing Jian Cheng
179 Bioengineering Institute, Nanjing, China.

180 Fresh leaves (0.2 g) were homogenized in 4 mL ice-cold 50 mM phosphate
181 buffer (pH 7.0-7.4). After the centrifugation at 3500-4000 rpm at 4 °C for 10 min, the
182 supernatant of homogenate was measured to determine SOD assays. For CAT assays,
183 fresh leaves (0.2 g) were homogenized in 1.8 mL ice-cold normal saline (NS). After
184 the centrifugation at 2500 rpm at 4 °C for 10 min, the supernatant was taken for
185 detection. Total soluble protein content was determined by following the method of
186 Bradford (1976), using bovine serum as standard.³³

187 **2.7. Statistical analysis**

188 The data were performed by using standard statistical software (SPSS 12.0), and
189 values were presented as the mean values \pm SD of three replications. Evaluation of
190 significant differences among different treatments was analyzed using one-way
191 ANOVA followed by Duncan's multiple-range test, with $p < 0.05$ indicating statistical
192 significance.

193 **3. Results and discussion**

194 **3.1. Plant growth and biomass**

195 After grown for 3 weeks, the average height of plants was 51 cm, and total dry
196 weight was 4.4 g per pot, and the average root weight accounted for 38.3% of the total
197 biomass. As Fig. 1 exhibited, the growth state of ramie showed differences among the
198 different treatments. The addition of NTA to the soil inhibited the plant biomass when

199 plants were exposed to Cd30 and Cd50, but no visible toxic symptoms appeared
200 except the treatment of 50N20. In the 50N20 group, whitish-brown chlorosis and
201 necrosis appeared. While the addition of strain was found to significantly enhance
202 plant growth. Not only was there a positive biomass production, but also there was a
203 higher and verdant growth. Moreover, the biomass of ramie treating with strain even
204 exceeded the biomass production under unpolluted soil treatment.

205 As seen from Fig. 2, significant differences in the biomass of shoot and root
206 were observed among the 15 treatments ($p < 0.05$). The biomass of ramie was
207 decreased after the addition of different levels of NTA in Cd30 and Cd50 treatments,
208 but there was an exception in the group of 30N5, which had an increased biomass of
209 10% compared to the control plants. However, when strain was added to the soil, the
210 total dry biomass was increased ranging from 15.8% to 33.1% and 14.3% to 30.7%
211 under the level of Cd30 and Cd50 respectively. These results demonstrated that high
212 concentration of NTA inhibited ramie biomass, while the application of *P. aeruginosa*
213 ATCC 9027 could enhance the total dry biomass in presence of Cd contamination.

214 The increased biomass of ramie with the application of *P. aeruginosa* ATCC 9027
215 may be beneficial for the removal of cadmium, because the more biomass means it
216 can pick up more contaminants. The possible mechanisms involved in ramie growth
217 promotion by *P. aeruginosa* ATCC 9027 could be explained in two different ways:
218 Firstly, the indirect promotion of ramie growth occurred when *P. aeruginosa* ATCC
219 9027 prevented or decreased some of the deleterious effects of phytopathogenic
220 organism.⁵ Besides, *P. aeruginosa* ATCC 9027 can also directly promote plant
221 growth by providing with a compound that is synthesized by the bacterium or by
222 further facilitating the uptake of nutrients (especially small molecules such as sugars,
223 amino acids, organic acids) from the plant.^{5,34} The inhibition biomass under 10 and 20
224 mmol kg^{-1} NTA was that metal phytotoxicity did occur due to the desorption and
225 dissolution effects by NTA.³⁵ Analogously, negative effects of NTA on plant growth
226 were reported in many studies.^{36,37}

227

228 **3.2. Effect of Cd on physiological characteristics**

229 The previous studies suggested that the inhibition of malondialdehyde (MDA) or
230 the degradation of chlorophyll was responsible for the growth restraint induced by
231 Cd.³⁸ In addition, Cd uptake by plants has been reported to induce extensive lipid
232 peroxidation, which reflected the degree of cell membrane damage caused by oxygen
233 free radicals.³⁹

234 Not surprisingly, it can be clearly seen in Fig. 3, the MDA content was 34.3
235 nmol g⁻¹ FW in unpolluted soil treatment, but reached up to 63.7 nmol g⁻¹ FW and
236 74.1 nmol g⁻¹ FW in Cd30 and Cd50 treatments, respectively. The NTA treatments
237 exhibited a linear enhancement of MDA content which was in accordance with the
238 increase in concentration of NTA. There was a slight decline in MDA content at low
239 NTA concentrations (5 mmol kg⁻¹) but higher MDA content was detected in ramie
240 when treated with 10 and 20 mmol kg⁻¹ NTA compared to the controls. The increase
241 of MDA content is probably due to its poisonous derivatives and the deleterious effect
242 of H₂O₂.⁴⁰ In addition, it can be seen that MDA contents in the leaves of ramie with
243 different levels of strain were lower than the controls, although differences were not
244 statistically significant. Similarly, *Serratia nematodiphila* LRE07, a endophytic
245 bacteria, significantly attenuated the content of MDA in *Solanum nigrum* L.⁴¹ The
246 lower level of MDA in leaves with the application of strain revealed that bacterial
247 inoculation can alleviate the damage on the cell membrane caused by Cd stress.

248 The chlorophyll content in plants was determined to elucidate the toxic effect of
249 Cd or exogenous chelants on photosynthesis system in ramie (Fig. 3b). Chlorophyll
250 content in leaves of ramie showed no significant alteration (p>0.05) when added with
251 NTA under Cd30, but decreased when ramie was exposed to Cd50. In contrast, both
252 in Cd30 and Cd50 treatments, there were a slight increase in chlorophyll content with
253 the addition of *P. asaeruginosa* ATCC 9027. These results meant that ramie suffered

254 strong stress with NTA while alleviated with *P. asaeruginosa* ATCC 9027. These are
255 in consistent with some previous studies, which have also reported that bacterial strain
256 could positively influence the chlorophyll contents of host plant under abiotic
257 stresses.^{34,42}

258 3.3. Cd accumulation and distribution in ramie

259 Fig. 4 presented the effects of 15 different treatments on accumulation of
260 cadmium in different tissues of ramie. The *P. aeruginosa* ATCC 9027 or NTA
261 improved the accumulation of Cd in shoots and roots of ramie to the different degrees.
262 Generally speaking, in different tissues of ramie, Cd accumulations in roots were
263 considerable higher than in shoots, which was in agreement with previous reports on
264 ramie.⁴³ The performance of NTA did not display obvious effect in Cd uptake.
265 Although increasing concentrations of NTA under Cd30 and Cd50 led to an increase
266 of the total Cd accumulations in plants, there was an inhibition of Cd accumulation in
267 the roots of ramie when compared with Cd30 and Cd50 treatments (except for the
268 treatment of 30N20). But the accumulation of Cd in shoot with NTA treatments was
269 significantly improved compared with Cd50 treatments. The highest Cd accumulation
270 in shoot with NTA was 274.5 and 405.1 mg kg⁻¹ DW in 30N20 and 50N20 treatments,
271 respectively. From Fig. 4, the infected strain plants in the presence of Cd had higher
272 Cd concentration in tissues compared with non-inoculation controls, especially in the
273 roots. And the higher concentration of Cd was observed in plants under Cd50 than
274 that under Cd30 with the addition of strain. The highest Cd accumulation in ramie
275 with *P. aeruginosa* ACCT 9027 was 748.3 and 1112.6 mg kg⁻¹ DW in 30S3 and 50S2
276 treatments, respectively. Besides, the inoculation of *P. aeruginosa* ATCC 9027
277 increased accumulation in root ranging from 54% to 96% and 13% to 104% in 30 and
278 50 mg kg⁻¹ Cd soils, respectively. The average uptake of Cd in root with *P.*
279 *aeruginosa* ACCT 9027 was increased by approximately 1.54- to 1.96-fold and 1.13-
280 and 2.04-fold under Cd30 and Cd50 treatments, respectively. Furthermore, the

281 average accumulation of Cd with *P. aeruginosa* ATCC 9027 was about 1.95-fold
282 (Cd30) and 1.54-fold (Cd50) compared to the corresponding NTA treatments.

283 The TF of the heavy metal Cd and the applied chelating agents NTA and *P.*
284 *aeruginosa* ATCC 9027 are depicted in Table 2. Compared to the Cd30 and Cd50
285 treatments, the addition of NTA tended to significantly increase Cd concentration in
286 stems and leaves, indicating that NTA enhanced Cd translocation from roots to shoots.
287 More interestingly, the TF of 50N (50 mg kg⁻¹ Cd and NTA) was higher than 30N (30
288 mg kg⁻¹ Cd and NTA). NTA increased TF compared to the controls ranked
289 17.4–36.2%, 74.5–81.8% at Cd30 and Cd50, respectively. The increase of TF might
290 be attributed to the fact that NTA facilitated Cd movement from roots to shoots. This
291 is the greatest advantage of NTA compared to other chelating agent for the
292 remediation of contaminated soils. Because ramie obviously absorbed Cd in root, so
293 the application of *P. aeruginosa* ATCC 9027 to soil caused no obvious difference of
294 TF.

295 The present investigation confirmed that strain is a better effective chelator than
296 NTA in accumulating Cd as well as increasing its availability for plant uptake. The
297 ability of NTA to desorb metals from the soil was lower in comparison to strain due to
298 the low affinity constants of its complexes with Cd.¹⁵ This is consistent with previous
299 research which have also reported that bacteria inoculation could enhance plant to
300 absorb heavy metals.^{44, 45} Overall the microbial activities in the root soils enhance the
301 efficiency of phytoremediation mechanisms under Cd stress soil by two
302 complementary ways: (i) Plant associated microbes reduce the mobility or availability
303 of pollutants in the rhizosphere; (ii) The microbes confer plant metal tolerance and/or
304 increase the plant biomass production in order to remove the pollutants.^{5, 21} This can
305 be interpreted as that the treatments with *P. aeruginosa* ATCC 9027 in ramie can
306 produce iron chelators called siderophores in response to low iron levels in plants.⁵
307 Plant growth-promoting bacteria may synthesize siderophores which can sequester
308 and solubilize iron from the soils and provide it to plant cells.^{21, 34} However, further

309 investigations on how the plant-associated metabolites producing microbes influence
310 the heavy metal mobilization and its uptake by plants in contaminated soils are needed.
311 These processes were therefore reasoned that the strain had an exceptional capacity to
312 accumulate Cd in the developed root system in plants. In addition, NTA acted as a
313 chelating agent which was useful to facilitate Cd movement. And the results are in
314 agreement with some previous studies which had also reported that the addition of
315 NTA could promote the mobilization of heavy metals.^{46,47} The increased TF by NTA
316 was probably due to the following reasons. Firstly, plants accumulate free metals in
317 their roots in the time period before chelant application. Secondly, with the
318 application of a chelating agent, metals are complexed within the roots and
319 translocated as metal chelates.⁴⁸

320 **3.4. Soil enzyme activities**

321 Many previous studies correlated with the toxicity of heavy metals on enzyme
322 activities especially urease activity in soil are available in literatures.⁴⁹⁻⁵¹ Additionally,
323 enzyme activities have been suggested as sensitive indicators of soil quality, which
324 indicated that urease activity can be considered as the biochemical index which
325 reflects the degree of soil Cd pollution.⁵²

326 Fig. 5 illustrated the urease activity changes in the soil during the plant growth.
327 When NTA was added to the soil, the urease activity significantly decreased with
328 increasing culture time (Fig. 5a and Fig. 5b). It suggested that the urease activity
329 tended to be decreased with increasing concentration of NTA when exposed to Cd30,
330 and the lowest urease activity of ramie appeared in 30N20 treatment after 3 weeks.
331 When treated with Cd50, the low dose of NTA (5 mmol kg^{-1}) enhanced urease
332 activity, while the high dose of NTA (20 mmol kg^{-1}) decreased urease activity. The
333 results suggested that urease activity was lower at high concentration of NTA than in
334 other treatments. Besides, the higher the concentration of NTA is, the more obvious
335 effect can be seen. Figs. 5c and 5d showed that the urease activity was maximized

336 after 3 weeks (except for 30S2) when exposed to Cd30 and Cd50. Moreover, there
337 was a net increase in urease activity from 1 week to 3 weeks with the increase of
338 strain concentration. It means that strain-infection positively influenced the urease
339 activity of soil, Furthermore, there was a negatively correlation between the Cd
340 content and the urease activity which was also confirmed by Stpniewska et al.⁵³ The
341 increase of urease activity in soil with the application of strain could be explained that
342 bacteria can produce a variety of low molecular weight organic acids such as chelate
343 compounds or complexes, and consequent release of active urease molecules.⁵⁴

344 **3.5. Effect of Cd on activated oxygen metabolism**

345 Antioxidant enzymes (SOD, CAT) removing the cells from active oxygen
346 species were determined as pivotal enzymes (Fig. 6). SOD was considered as first
347 defense barrier against reactive oxygen species (ROS) as it acted on superoxide
348 radicals.⁵⁵ It is essential to control the levels of ROS for their cellular damage
349 activities. Fig. 6a showed that SOD activity in leaves of ramie which was treated with
350 NTA was reduced, and the maximum decrease rate in the treatment of 30N20 and
351 50N10 were up to 14.6% and 18.6%, respectively. When strain was inoculated, SOD
352 activity was increased with increasing concentration of strain, and the maximum
353 increase rate was up to 26.3% in the 30S3 treatment and 37.1% in the 50S3 treatment
354 compared to Cd treatments, respectively. Fig. 7 represents CAT activity in leaves of
355 ramie which treated with NTA and strain. It can be seen that strain induced obvious
356 decrease in CAT activity. In the present study, when ramie seedlings were submitted
357 to different levels of NTA (5, 10, 20 mmol kg⁻¹), obvious decrease in the activities of
358 CAT was observed with the addition of NTA by 5 and 10 mmol kg⁻¹ while an
359 increase was observed at NTA concentration of 20 mmol kg⁻¹. Fig. 7b showed that
360 CAT activity was lower under strain treatment than that under single Cd treatment.
361 The increase of CAT activity was probably due to the fact that heavy metals
362 stimulated the synthesis of enzyme.

363 The results indicated that NTA and antioxidative enzymes activities were
364 negatively correlated, while ATCC 9027 could promote ramie against Cd
365 phytotoxicity via improving antioxidative enzymes activities. Similar effects of
366 bacteria on plant antioxidative system have also been reported for Cd in drunken
367 horse grass and Zn in ryegrass.⁵⁶ This could be explained that SOD dismutates two
368 superoxide radicals to oxygen and H₂O₂ and thus maintained superoxide radicals in
369 steady state level.⁵⁶ The reason for decrease of SOD activity with NTA might be the
370 inactivation of enzyme by H₂O₂.⁴⁰ The increase of SOD activity with strain was
371 attributed to the synthesis of enzyme protein.⁵⁷ Moreover, some traits, such as
372 production of siderophore and antioxidative enzymes, of bacteria may be the possible
373 reason of enhancing the activities of antioxidative enzymes in plants.⁵⁸ CAT exists in
374 mitochondria and peroxisomes where it decomposes H₂O₂ to water and oxygen. The
375 increasing in CAT activity was probably due to the fact that heavy metals stimulated
376 the synthesis of enzyme, the decline of CAT activity might be attributed to
377 inactivation of enzyme by ROS.⁵⁹ These results indicated that the inoculation with
378 beneficial microbes assisted plants to alleviate heavy metal stress through enhancing
379 the activities of antioxidant enzymes.

380 4. Conclusions

381 The results demonstrated that the addition of NTA effectively increased the Cd
382 translation from root to shoot, whereas showed no obvious effect in Cd uptake, even
383 plant growth and related enzyme activities were inhibited. However, the inoculation
384 of *P. aeruginosa* ATCC 9027 alleviated these Cd-induced damages, resulting in
385 promotion of ramie growth, improvement of antioxidative enzymes activities and
386 increase of total Cd-uptake by ramie. Additionally, the average accumulation of Cd by
387 ramie with *P. aeruginosa* ATCC 9027 treatment was much higher than that of NTA
388 treatments. The improvement in antioxidative enzymes activities and urease activity
389 in soil after the inoculation of *P. aeruginosa* ATCC 9027 are probably the main

390 mechanisms involved in Cd phytotoxicity reduction. Besides, the results showed that
391 the inoculation with beneficial microbes assisted plants to alleviate heavy metal stress
392 through enhancing the activities of antioxidant enzymes. All these results indicated
393 that *P. aeruginosa* ATCC 9027 may be an effective remedy for Cd contaminated soils
394 and a promising candidate for practical application on phytoremediation of Cd
395 contaminated soils. However, additional studies regarding the interaction of
396 plant-bacterial-metals in polluted soils are needed in the further investigations.

397

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Figure captions:

Fig. 1 Photographs of ramie in the presence of Cd pollution after being transferred into pot for 3 weeks (a: 30mg kg⁻¹ Cd and NTA, b: 50mg kg⁻¹ Cd and NTA, c: 30mg kg⁻¹ Cd and strain, d: 50mg kg⁻¹ Cd and strain).

Fig. 2 The differences in leaf, stem and root biomass of 15 treatments after 3 weeks of growth in Cd contaminated soil (30 and 50 mg kg⁻¹ Cd). All the values are mean of triplicates ± SD.

Fig. 3 Changes of malondialdehyde (MDA) in leaves of ramie added with NTA and *P.aeruginosa* ATCC 9027 (a) exposed to 30, 50 mg kg⁻¹ Cd. Changes of chlorophylla in leaves of ramie added with NTA and *P. aeruginosa* ATCC 9027 (b) exposed to 30, 50 mg kg⁻¹ Cd. All the values are mean of triplicates ± SD.

Fig. 4 Changes in Cd amounts in the leaf, stem and root of ramie under different treatments. All the values are mean of triplicates ± SD.

Fig. 5 Change of urease activity in different treatments. (a) represents soil with NTA exposed to 30 mg kg⁻¹ Cd; (b) represents soil with NTA exposed to 50 mg kg⁻¹ Cd; (c) represents soil with *P. aeruginosa* ATCC 9027 exposed to 30 mg kg⁻¹ Cd; (d) represents soil with *P. aeruginosa* ATCC 9027 exposed to 50 mg kg⁻¹ Cd. All the values are mean of triplicates ± SD.

Fig. 6 Changes of superoxide dismutase activity (SOD) in leaves of ramie added with NTA (a), and infected by *P. aeruginosa* ATCC 9027 (b) exposed to 30, 50 mg kg⁻¹ Cd. All the values are mean of triplicates ± SD.

Fig. 7 Changes of catalase activity (CAT) in leaves of ramie added with NTA (a), and infected by *P. aeruginosa* ATCC 9027 (b) exposed to 30, 50 mg kg⁻¹ Cd. All the values are mean of triplicates \pm SD.





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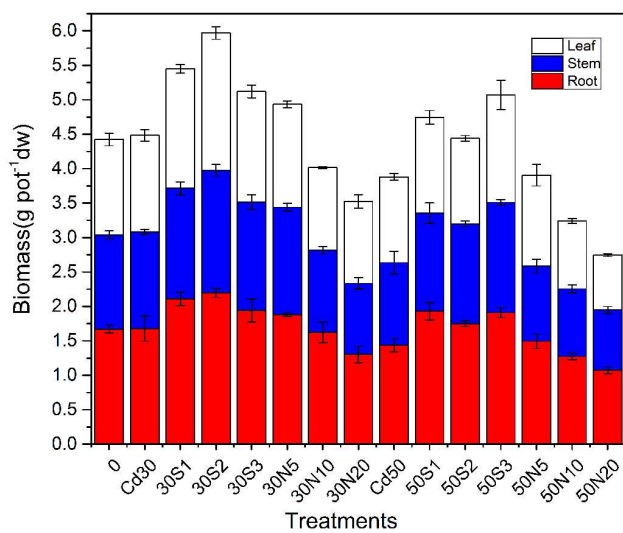


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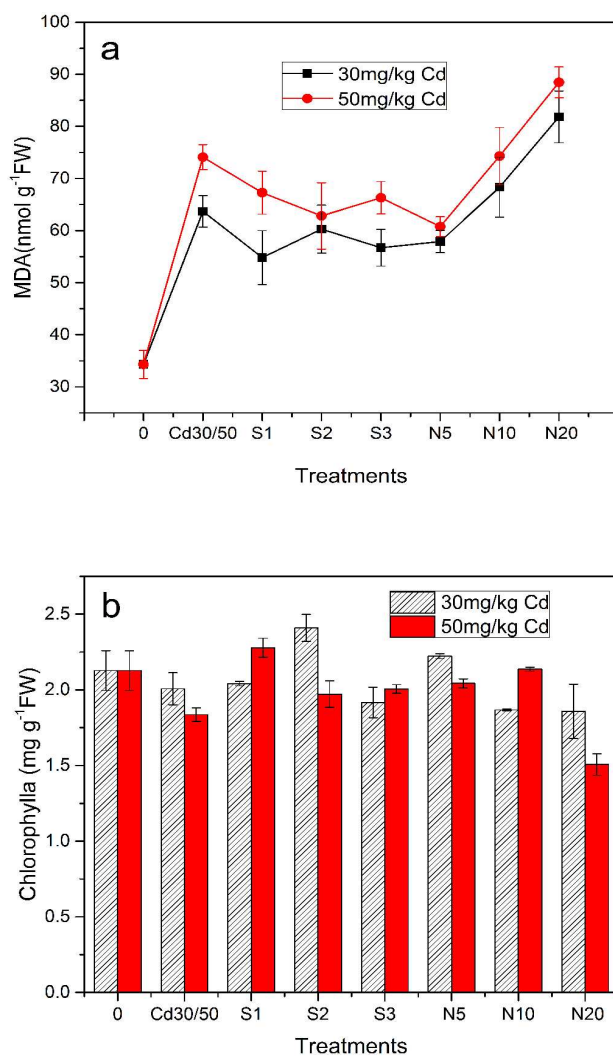


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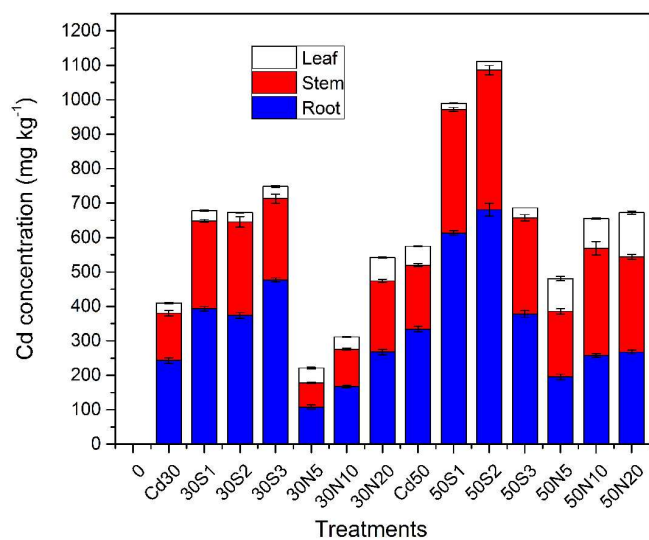
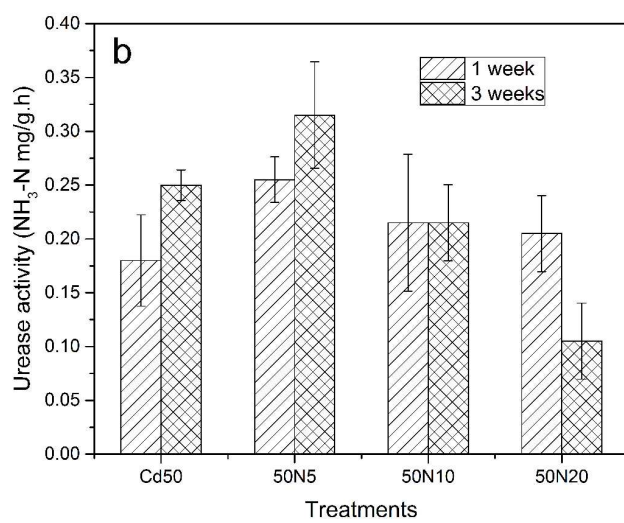
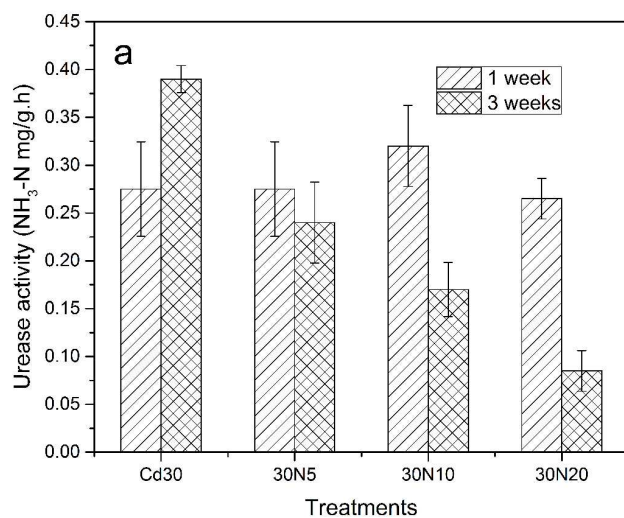


Fig. 4 Changes in Cd amounts in the leaf, stem and root of ramie under different treatments. All the values are mean of triplicates \pm SD



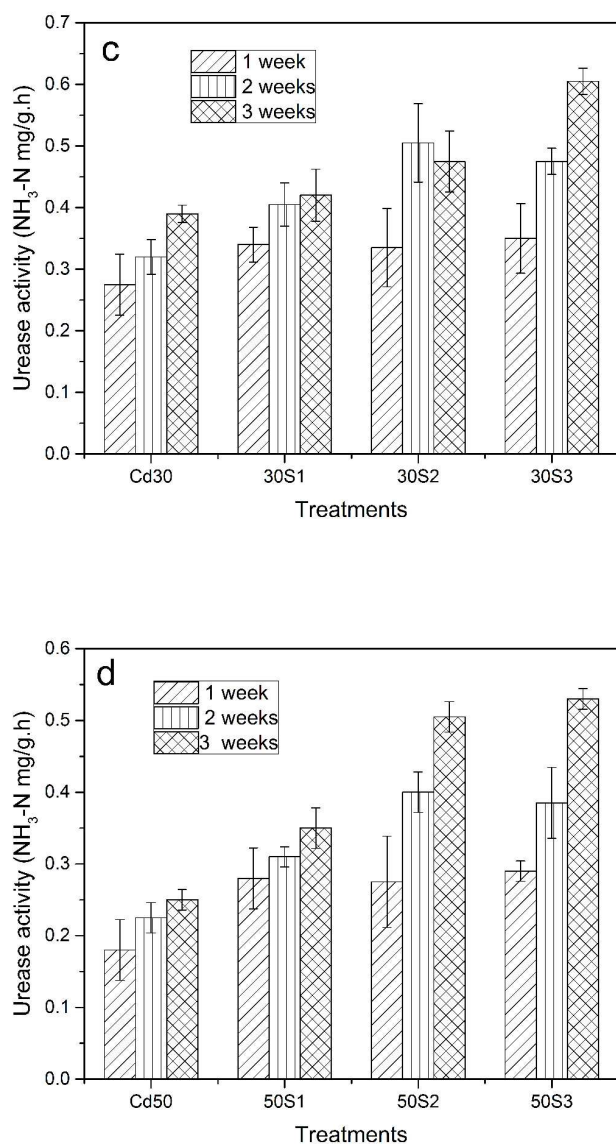


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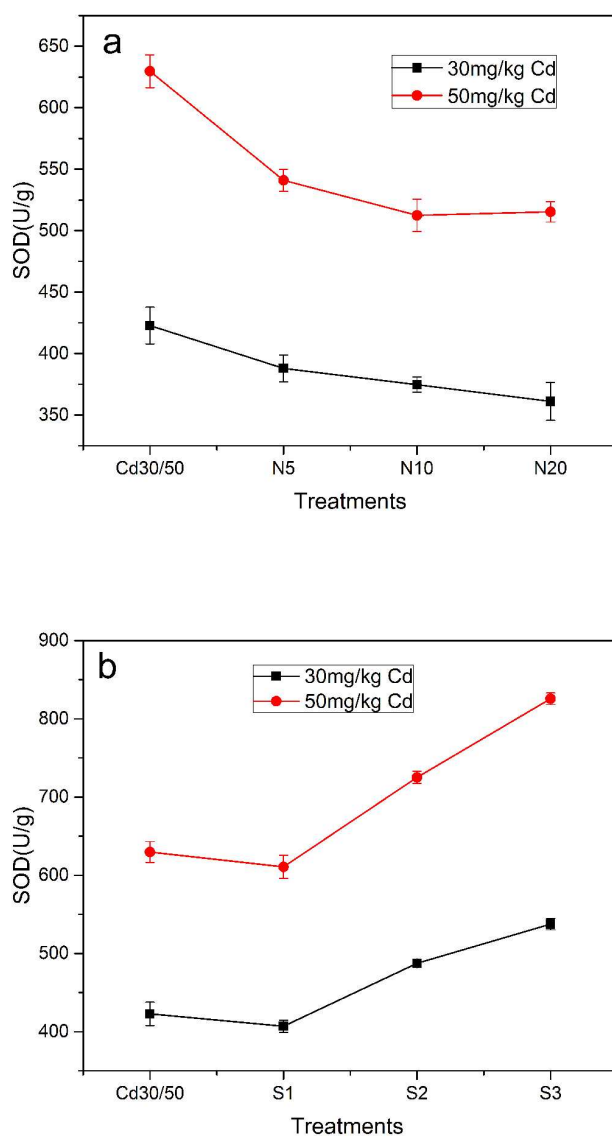


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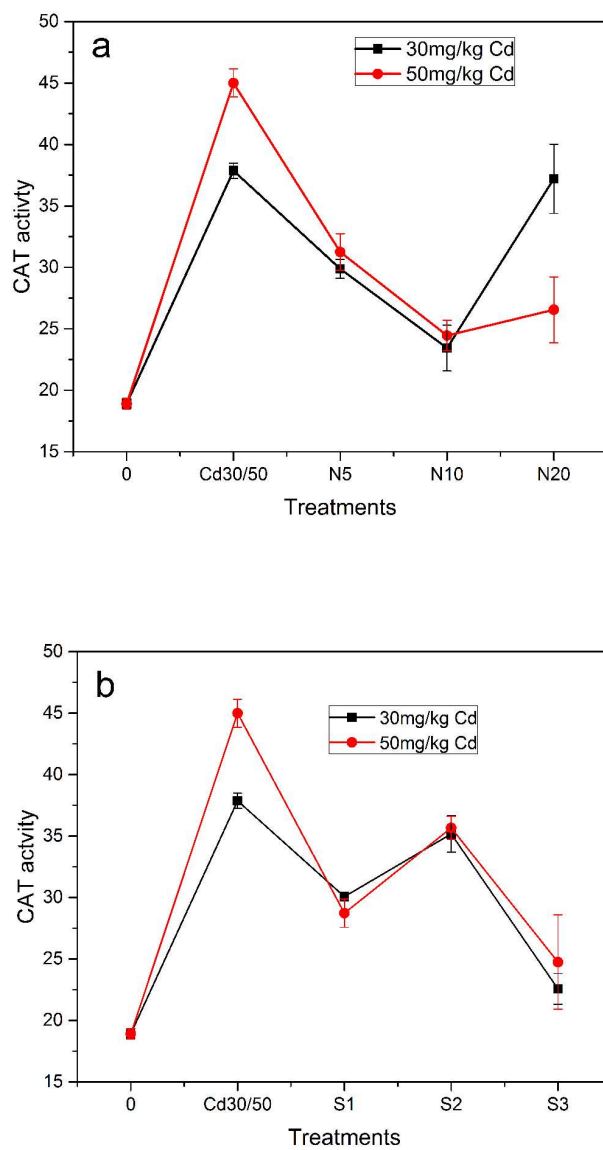


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Table 1

Physic-chemical properties of soil.

pH	Organic matter	Total N	Total P	CEC	Cd
6.6	19.3g kg ⁻¹	0.890g kg ⁻¹	0.236 g kg ⁻¹	16.7cmolkg ⁻¹	undetected

Table 2

Number of treatments

Number	Treatment
0	Control
Cd30 (Control)	30 mg kg ⁻¹ Cd
30N5	30 mg kg ⁻¹ Cd Cd +5 mmol kg ⁻¹ NTA
30N10	30 mg kg ⁻¹ Cd +10 mmo kg ⁻¹ 1NTA
30N 20	30 mg kg ⁻¹ Cd +20 mmol kg ⁻¹ NTA
30S1	30 mg kg ⁻¹ Cd + strain 1
30S2	30 mg kg ⁻¹ Cd + strain 2
30S3	30 mg kg ⁻¹ Cd + strain 3
Cd50 (Control)	50 mg kg ⁻¹ Cd
50N5	50 mg kg ⁻¹ Cd Cd +5 mmol kg ⁻¹ NTA
50N10	50 mg kg ⁻¹ Cd +10 mmo kg ⁻¹ 1NTA
50N 20	50 mg kg ⁻¹ Cd +20 mmol kg ⁻¹ NTA
50S1	50 mg kg ⁻¹ Cd + strain 1
50S2	50 mg kg ⁻¹ Cd + strain 2
50S3	50 mg kg ⁻¹ Cd + strain 3

NTA represented nitrilotriacetic acid; strain 1 represented the addition of strain one time, strain 2 represented twice and strain 3 represented three times. Each added for one week apart.

Table 3 The Cd accumulation in different ramie tissues and the TF value after the addition of various concentrations of *P. aeruginosa* ACCT 9027 or NTA. Cd30 and Cd50 represent the control groups with treatment of 30 mg kg⁻¹ Cd and 30 mg kg⁻¹ Cd. Data represent means± SD of three replicates.

Treatments	Cd content (mg kg ⁻¹ DW)			TF value
	Leaves	Stems	Roots	
0	UD	UD	UD	/
Cd30	242.93±7.92	136.87±7.78	29.76±2.12	0.272
30N5	43.85±2.33	69.50±2.12	107.85±5.87	0.420
30N10	35.62±1.70	108.46±2.68	167.38±3.11	0.345
30N20	67.52±1.96	206.57±4.94	267.53±7.78	0.41
30S1	29.45±1.91	255±4.24	393.5±6.36	0.292
30S2	26.85±0.64	272±15.55	373.5±8.49	0.321
30S3	35.25±2.47	236.25±13.08	476.75±6.01	0.228
Cd50	55.12±1.57	185.75±4.74	333.95±7.85	0.288
50N5	95.01±6.47	190.75±7.42	195.31±8.49	0.586
50N10	85.95±2.19	311.52±19.79	257.75±5.30	0.615
50N20	128.25±4.59	276.75±6.72	267.35±5.44	0.451
50S1	17.905±0.74	358.25±6.01	613.5±6.36	0.245
50S2	25.355±0.69	405.5±13.43	680.75±18.74	0.254
50S3	28.69±0.08	279.5±9.192	378±9.89	0.327