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2	Sensitive and selective determination of aqueous triclosan based on gold nanoparticles on
3	polyoxometalate/reduced graphene oxide nanohybrid
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34 ABSTRACT

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The widespread use of triclosan (TCS) in cleaning products for household, medical devices and 36 37 personal care poses a potential risk to the ecological system and human health due to its release into sediments, surface water and ground water resources and chronicle toxicity to aquatic organisms. A 38 39 novel molecular-imprinted electrochemical sensor based on gold nanoparticles decorating 40 polyoxometalate $(H_3PW_{12}O_{40})$ /reduced graphene oxide was developed for determination of trace 41 TCS in wastewater. Reduced graphene oxide (rGO) was functionalized by the polyoxometalate (POM) through electrostatic interaction between the POM and rGO nanosheets to produce a 42 43 photocatalyst (POM/rGO) in aqueous solution. Gold nanoparticles (AuNPs) were further deposited on the POM/rGO without using any reducing agent and the prepared nanomaterial 44 45 (AuNPs/POM/rGO) was employed modify glass carbon (GC)electrode to а (AuNPs/POM/rGO/GC) under infrared light. Several techniques, X-ray photoelectron spectroscopy 46 47 (XPS), reflection-absorption infrared spectroscopy (RAIRS), scanning electron microscope (SEM), and transmission electron microscope (TEM), were used for electrode characterization. TCS 48 49 imprinted film was generated on AuNPs/POM/rGO/GC via polymerization of phenol and TCS and characterized by cyclic voltammetry (CV) and electrochemical impedance spectroscopy (EIS). The 50 sensor was found to have linear detection range and limit of TCS as 0.5-50.0 nM and 0.15 nM, 51 52 respectively. The molecular imprinted sensor was applied to wastewater and lakewater samples and 53 demonstrated effective performance as compared to other complicated methods.

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55 *Keywords*: Triclosan; Wastewater, Reduced graphene oxide; Molecular imprinting; Sensor

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62 **1. Introduction**

TCS is a typical chemical used in pharmaceuticals and personal products ¹ such as surgical 63 suture materials or hand soaps, deodorants, toothpastes, antiseptic-creams, plastics, foodstuffs and 64 functional clothing for over 40 years.² Its widespread use has led to the release of TCS into 65 wastewater, sediments and many water sources.³ TCS is chronically toxic to aquatic organisms and 66 its presence in wastewaters may affect the ecosystem and human health.² Several analytical 67 methods have been reported to detect TCS using gas chromatography-tandem mass spectrometry, 68 liquid chromatography-mass spectrometry (LC-MS), liquid chromatography/electrospray ionization 69 tandem mass spectrometry and voltametrics. ⁴⁻¹⁰ But these techniques have some disadvantages 70 such as expensive apparatus and complicate operation.^{11, 12} A rapid and sensitive method to detect 71 72 triclosan is thus important to ensure human and environment safety. In recent years, various 73 nanosensors have been reported for selective, sensitive and rapid determination of toxic compounds, biomolecules and drugs. ¹²⁻¹⁵ 74

In the past few years, graphene has become an intensive interest of scientists all over the 75 world due to its stability and high surface area. ^{16, 17} Graphene has honeycomb-like structure via sp² 76 hybridization in one-atom-thickness. ¹⁸ Currently, graphene oxide (GO) is widely produced by 77 chemical oxidation of graphite and used as a precursor to graphene. GO can be reduced by thermal 78 treatment or chemical reduction to form rGO, ¹⁹ and rGO has been used for fuel cells, drug 79 detection and sensors.²⁰⁻²² Many papers have also reported the chemical, ²³ optical, ²⁴ adsorption²⁵ 80 and electronic properties ²⁶ of various nanostructured metals. AuNPs are utilized as electrode 81 surface for sensors ^{11, 27} to increase the surface area and rate of electron transfer. In addition, 82 polyoxometalates (POMs) are polyatomic anionic ion clusters composing of d-block transitional 83 metal-oxides, and they have multiple redox behavior and photo-electrochemical properties. ^{28, 29} 84 POMs are a class of photoactive materials used in homogeneous reactions or heterogeneous 85 86 processes. In reduced forms, their electron and proton transfer and/or storage abilities make them

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act as efficient donors or acceptors of several electrons without structural change. POMs also have
been shown to serve as reducing and capping agents for metal nanostructures.¹⁷ The molecular
imprinting technique is widely used for molecular recognition ³⁰ via the polymerization of target
molecules, forming specific cavities. ^{24, 27} From those above materials and molecular imprinting
technique, various sensors can be fabricated.

There is no report about determination of TCS by using a molecular imprinting method 92 based on the nanomaterials including rGO and AuNPs. Jiang et al prepared AuNPs on GO surface 93 94 by using polyethylenimine as a reducing reagent, followed by L-cysteine immobilization through an 95 Au–S bond. After the preparation, the nanocomposite was applied as a novel ZIC-HILIC material to achieve highly selective enrichment of glycopeptides from biological samples.³¹ In this study, 96 97 AuNPs were synthesized under the UV light on rGO surface with POM as reducing and stabilizing 98 reagent. We then prepared a TCS imprinted electrochemical sensor based on AuNPs deposition on 99 POM functionalized rGO. The developed imprinted electrochemical sensor shows high sensitivity 100 and selectivity in wastewater measurement.

101 **2.** Experimental

102 **2.1. Chemicals and materials**

103 TCS, methyltriclosan (MTC), triclocarbon (TCC), p-chlorophenol (PCP) and 2,4,6-trichlorophenol 104 (TCP) were purchased from Sigma-Aldrich (USA). A stock solution of TCS (1.0 mM) was 105 prepared in 5 mL of ultra pure water and then diluted to 25 mL. The working solutions were then 106 prepared with 0.10 M phosphate buffer solution (pH 7.0). Phenol, potassium ferricyanide 107 (K₃[Fe(CN)₆]), hydrogen tetra-chloroaurate hydrate (HAuCl₄), acetonitrile (MeCN), H₃PW₁₂O₄₀, 108 and isopropyl alcohol (IPA) were purchased from Sigma–Aldrich, USA. Potassium chloride (KCl) 109 and potassium ferrocyanide (K₄[Fe(CN)₆]) in analytic grade were supplied by Merck, Germany.

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112 **2.2. Instrumentation**

Differential pulse voltammetry (DPV) and CV were carried out on an electrochemical station
(IviumStat, U.S) equipped with a C3 cell stand. Electrochemical impedance spectroscopy (EIS) data
were acquired at 10 mV wave amplitude from 0.1 to 100 kHz and at an electrode potential of 0.195
V. The infrared spectra were obtained from a Bruker Tensor 27 FT-IR. XPS analysis was performed
on a PHI 5000 Versa Probe X-ray photoelectron spectrometer (Φ ULVAC-PHI, Inc., Japan/USA).
TEM images were obtained on a JEOL 2100 HRTEM instrument (JEOL Ltd., Tokyo, Japan) and
SEM images were obtained on a ZEISS EVO 50 analytic microscope (Germany).

120 **2.3.** Cleaning of glass carbon (GC) electrodes

121 All GC electrodes were first polished by 0.1 and 0.05 μ m alumina successively and then the 122 electrodes were sonicated in pure water and IPA + MeCN solution (50:50 by v/v) to remove 123 unreacted materials from the surface. The reference electrode was a Ag/AgCl/KCl_(sat) and the 124 counter electrode was a Pt wire.

125 **2.4. Preparation of rGO**

GO was prepared according to the protocol in our previous papers. ¹⁴ The as-prepared GO was dispersed into water (200 mL) with addition of hydrazine hydrate (4 mL, 80 wt%) and was heated at 100 °C for 24 h in an oil bath. The rGO was collected by vacuum filtration.

129 2.5. Fabrication of AuNPs on POM/rGO and preparation of AuNPs/POM/rGO/GC

130 The as-synthesized rGO was dissolved in an ethanol solution (2 mg mL⁻¹) for 1 h ultrasonic 131 agitation. A $H_3PW_{12}O_{40}$ solution (1 mL, 1 mM) was reduced using a ultra-violet (UV) light source. 132 The rGO suspension was then well mixed with the reduced POM at a 1:1 (0.5:0.5 by v/v) volume 133 ratio for 2 h to get POM/rGO. A HAuCl₄ solution (1 mM) was added to the POM/rGO solution (0.4 134 mg mL⁻¹) at a 1:1 (2.0:2.0 by v/v) volume ratio. The solution was sonicated to form a homogeneous 135 suspension. The prepared solution was stirred under the UV light for 40 min. Finally, 20 µL of

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136 AuNPs/POM/rGO (0.5 mg mL⁻¹) was dropped on the GC electrode and then the modified electrode

137 (AuNPs/POM/rGO/GC) was dried under an infrared heat lamp.

138 **2.6. Preparation of imprinted TCS sensors**

139 The preparation of TCS imprinted sensors is illustrated in Scheme 1. Firstly, TCS molecular imprinted polymer (MIP) film on AuNPs/POM/rGO/GC electrode (MIP/AuNPs/POM/rGO/GC) 140 was prepared by CV for 20 cycles using 80 mM phenol as a monomer in a phosphate buffer 141 solution (pH 7.0) containing 20 mM TCS at a scan rate of 100 mV s⁻¹ between 0.0 V and +1.0 V. 142 143 After electropolymerization, the electrode was dried at room temperature. For comparison, 144 MIP/GC, MIP/rGO/GC and MIP/POM/rGO/GC electrodes were also prepared with same way. A 145 non-polymer imprinted electrode (NIP) was prepared without using TCS for a control experiment 146 like the preparation of MIP. To break up the electrostatic interactions between phenol monomer and 147 polar groups of the TCS, we used 1.0 M NaCl as desorption agent in a batch system. A TCS 148 imprinted electrode was dipped into 25 mL of the 1.0 M NaCl aqueous solution and was swung in a 149 bath (200 rpm) at room temperature for 20 min. After that, the electrode was washed with ultra pure 150 water and dried in nitrogen gas under vacuum (200 mmHg, 25 °C). The MIP electrodes were stored 151 in a closed box without fluctuations of temperature and pressure. In addition, the voltammograms 152 were obtained in an insulation cabinet for avoiding temperature and pressure fluctuation to affect 153 the sensor response.

154 **2.7. Preparation of wastewater samples**

Wastewater samples were collected from an industrial wastewater pool in Izmir, Turkey, using precleaned amber glass bottles. Lakewater samples were collected from Van Lake in Turkey. The sample bottles were filled without headspace and immediately placed in coolers filled with icepacks and transferred to the laboratory for storage at 4 °C and analysis within one week. Before analysis, the collected wastewater and lakewater samples were centrifuged again at 4500 rpm for 5 min and

160 filtrated by a 0.45-µm syringe filter. The filtrates were then diluted with 0.1 M phosphate buffer



solution (pH 7.0) for analysis.



163

Scheme 1. Fabrication procedure of the MIP/AuNPs/POM/rGO/GC sensor

164 **3. Results and Discussion**

165 **3.1. Characterization of electrode surface**

166 TEM image of AuNPs/POM/rGO shows that the particle sizes of AuNPs are very similar at 167 the mean diameter of 8-9 nm (Fig. 1A). The AuNPs are presented in dark dots on a lighter-shaded substrate of planar POM/rGO sheets. The creased nature of rGO is highly beneficial in providing a 168 high surface area on GC electrodes. In addition, C, Au, O, W and P peaks have been observed in 169 EDX analysis (Fig. 1B), confirming the formation of AuNPs/POM/rGO nanohybrid. The IR spectra 170 171 of the AuNPs/POM/rGO also show the formation of the nanohybrid (Fig. 1C). The bands around 3200 cm⁻¹ and 1600 cm⁻¹ suggested the oxygen-containing functional groups of rGO. The peaks 172 173 around 1580 cm⁻¹ can be attributed to the stretching vibrations of C=O groups of the rGO sheets. Fig. 1C confirms the POM attached on rGO planes. The bands around 1050 cm⁻¹ and 1350 cm⁻¹ are 174 referred to metal-oxygen groups of POM/rGO. The formation of POM/rGO may be explained with 175 the electrostatic interaction between POM and rGO via strong adsorption.^{28, 32} The formation of 176 AuNPs/POM/rGO was further examined by XPS. The peaks of C_{1s}, P_{2p}, Au_{4f} and W_{4f} confirmed the 177

- formation of AuNPs/POM/rGO nanohybrid (Fig. 1D). The $Au_{4f}^{7/2}$ peak at 82.5 eV confirms the
- presence of AuNPs and the signal at 87.2 eV can be attributed to free gold nanoparticles.¹⁴



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Figure 1. (A) TEM image of the AuNPs/POM/rGO, (B) EDX analysis of AuNPs/POM/rGO, (C)
RAIRS spectra of rGO, POM/rGO and AuNPs/POM/rGO, (D) XPS spectra of AuNPs/POM/rGO
surface.

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185 SEM characterization was performed to evaluate the morphologies of the electrode surfaces 186 in step by step modification. Fig. 2A displays that GC electrode has smooth surface. Fig 2B shows 187 the layers of rGO indicating high surface area of modified GC surface while Fig 2C presents the 188 POM/rGO/GC electrode surface. For AuNPs/POM/rGO on GC electrode, an intensive layer was 189 observed covering the surface (Fig. 2D). An electrodeposition layer by electro polymerization of 190 phenol covered the MIP/AuNPs/POM/rGO/GC electrode. These images indicate that the imprinted 191 electrochemical sensor is accomplished (Fig. 2E). Moreover, AuNPs/POM/rGO/GC was regular 192 spheres while the surface was rough. Compared with POM/rGO/GC, the existence of AuNPs could 193 not only enhance the adsorption capacity but also conducive to the formation of 194 MIP/AuNPs/POM/rGO/GC, playing the role of a framework for the formation of 195 MIP/AuNPs/POM/rGO/GC. Some granular substances were attached on the surface of 196 MIP/AuNPs/POM/rGO/GC, indicating the formation of MIPs through electrochemical 197 polymerization.



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Figure 2. SEM images of bare GC (A), rGO/GC (B), POM/rGO/GC (C), AuNPs/POM/rGO/GC
(D) and MIP/AuNPs/POM/rGO/GC (E) surfaces

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Electro polymerization was performed by CV in a phosphate buffer solution (0.1 M, pH 7.0) and the voltammograms are presented in Fig. 3. It was clearly demonstrated that the currents decreased with number of the cycles. The oxidation of phenol was recorded as the irreversible peak at the potential of 0.65 V on the first scan. During continuous scanning, the current of the reduction peak decreased and then disappeared. This showed MIP film formation on the AuNPs/POM/rGO/GC electrode.



Figure 3. Cyclic voltammogram for the electrochemical polymerization of phenol with TCS in phosphate buffer solution. Phenol (80 mM), TCS (20 mM), pH 7.0, and scan rate of 100 mV s⁻¹ for 20 cycles

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214 **3.2.** Characterization of electrode impedance.

215 EIS of bare GC electrode displays a small semicircle at high frequencies. The value of 216 charge transfer resistance (R_{ct}) of the bare GC electrode was calculated to be 100 ohm (curve a of Fig. 4A). When the rGO was coating on the bare GC electrode, the value of R_{ct} was calculated as 70 217 ohm (curve b of Fig. 4A). This is clearly indicative that the rGO layer increases the electron transfer 218 219 rate. When the POM was coating on rGO/GC electrode, the value of R_{ct} was found to be 58 ohm (curve c of Fig. 4A). These performances were attributed to the large surface area and the 220 221 synergistic effect of POM and rGO. The EIS of AuNPs/POM/rGO/GC electrode presents in a 222 straight line, the characteristic of a diffusional limiting step (curve d of Fig. 4A). Hence, it is clear that AuNPs/POM/rGO nanocomposite effectively increased the electrode active area. Because 223

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224 AuNPs/POM/rGO nanocomposite has such a large surface area and the synergistic effect, it improved electro-oxidation of 1.0 mM [Fe(CN)₆]^{3-/4-} solution in 0.1 M KCl. Hence, this 225 226 nanocomposite facilitated the electron transfer and electro-oxidation of TCS.

227 In addition, after the electrochemical polymerization of phenol monomer on AuNPs/POM/rGO/GC electrode, the MIP/AuNPs/POM/rGO/GC electrode shows a large R_{ct} (610 228 ohm) (curve b of Fig. 4B), indicating that the MIP film displays a strong obstruction effect. After 229 230 removal of TCS molecules from the electrode surface, the recognition sites appear and the R_{ct} 231 decreases to about 270 ohm (curve c of Fig. 4B). After rebinding of TCS (10.0 nM), the R_{ct} 232 increases to 410 ohm (curve d of Fig. 4B), demonstrating that TCS adsorption can prevent the 233 electrochemical activity of 1.0 mM $[Fe(CN)_6]^{3-/4-}$.



234

235 Figure 4. (A) EIS of (a) bare GC; (b) rGO/GC; (c) POM/rGO/GC; (d) AuNPs/POM/rGO/GC electrodes in 1.0 mM [Fe(CN)6]^{3-/4-} solution in 0.1 M KCl, (B) EIS of (a) AuNPs/POM/rGO/GC; 236

(b) MIP/AuNPs/POM/rGO/GC (with template molecule); (c) MIP/AuNPs/POM/rGO/GC
(removing template); (d) after rebinding of TCS (10.0 nM) in 1.0 mM [Fe(CN)₆]^{3-/4-} solution in 0.1
M KCl, (C) DPVs of different electrodes in 0.1 M phosphate buffer (pH 7.0) (a)
MIP/AuNPs/POM/rGO/GC in blank buffer solution, (b) NIP/AuNPs/POM/rG/GC after rebinding
of 10.0 nM TCS, (c) MIP/AuNPs/POM/rGO/GC after rebinding of 10.0 nM TCS, (D) DPV curves
of different MIP electrodes in 0.1 M phosphate buffer (pH 7.0) after rebinding of 10.0 nM TCS (a)
GC; (b) rGO/GC; (c) POM/rGO/GC; (d) AuNPs/POM/rGO/GC electrodes

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245 **3.3.** Characterization of voltammetrics of electrodes

DPV showed the responses of TCS at different electrodes (Fig. 4C and 4D). The 246 247 MIP/AuNPs/POM/rGO/GC electrode shows no background current signal in 0.1 M phosphate 248 buffer (pH 7.0) (curve a of Fig. 4C). After rebinding of TCS (10.0 nM TCS), it shows a much 249 higher peak at about 0.65 V (curve c of Fig. 4C). However, the NIP/AuNPs/POM/rGO/GC 250 electrode shows a small current signal (curve b of Fig. 4C). This indicates that the non-specific 251 interaction of TCS is weak and the response after MIP is very strong. The performances of 252 different MIP sensors were also compared by DPV (Fig. 4D). It is shown that the performance of 253 MIP/AuNPs/POM/rGO/GC electrode (curve d of Fig. 4D) is better than that of MIP/POM/rGO/GC, 254 MIP/rGO/GC and MIP/GC electrodes (curves c, b and a of Fig. 4D) due to more effective surface 255 area.

3.4. Optimization of fabrication and analytical conditions

The effects of the concentration of AuNPs/POM/rGO on MIP/AuNPs/POM/rGO/GC electrode were first tested. Initially, with the increasing concentration of AuNPs/POM/rGO up to 0.5 mg mL⁻¹, the peak current of TCS increased and reached a maximum at 7 μ A. However, after the concentration exceeded 0.5 mg mL⁻¹, the peak current of TCS (10.0 nM) is decreased (Fig. 5A). Hence, 0.5 mg mL⁻¹ of AuNPs/POM/rGO was selected as the optimum amount.

The pH of the medium also produces a significant influence on the polymeric film. ^{11, 14} Fig. 5B demonstrates the DPV peak current in the pH range of 5.0-9.0. The maximum signal was appearing at pH 7.0. TCS molecules show different electrochemical oxidation behaviors to the polymeric film at different pHs. The DPV response of TCS increased with solution pH up to 7.0 and decreased subsequently. After the solution pH exceeded 7.0, the decrease of the peak current may be owing to the dissociation of the phenolic moiety.

The influence of TCS to phenol monomer molar ratio was also studied (Fig. 5C). The peak current of TCS achieved a maximum at the ratio of 1:4. This was linked with the available binding sites. At low amount of phenol monomer, the available binding sites were less. According to the results, the signal of TCS increased when the amount of monomer increased to 80.0 mM. The increase was resulted from increase of the number of binding site. However, at a high concentration of phenol monomer, the non-specific interactions of TCS-monomer could occur, reducing the specific response.

Fig. 5D shows the variation of DPV responses at different elution time. The TCS peak current shows increasing with the elution time, reaching a maximum at 20 min, and then it remained stable after 20 min, indicating that the elution of TCS was completed during 20 min. Thus, the optimal elution time at 20 min was taken.

Fig. 5E shows the effect of temperature on DPV responses in the range of 5-30 °C. As shown in Fig. 5E, the highest peak current occurred at 20 °C. After that, it remained stable. Therefore, the experiment temperature was chosen as 20 °C.



Figure. 5. Effects of the concentration of AuNPs/POM/rGO (A), pH (B), molar ratio TCS molecule
to phenol monomer (C), elution time (D) and temperature (E).

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3.5. The linear detection range of TCS

The differential pulse voltammograms at varying TCS concentrations (Fig. 6A) show that the peak currents increased with increasing TCS concentration. For each point of the calibration graph, six independent measurements were obtained and the mean value was used. The linear regression equation of TCS (Fig. 6B) was obtained as y = 0.471x + 1.593. From the equation, limits of TCS quantification (LOQ) and its detection (LOD) were found to be 5.0×10^{-10} M and 1.5×10^{-10}

292 M, respectively. ²⁷



TCS Concentration, nM

Figure 6. DPV profiles of the electrochemical sensor at different TCS concentrations in phosphate solution pH 7.0 from background without TCS to 50.0 nM TCS (A), and linear calibration curve of TCS (B).

Moreover, the recovery experiments in wastewater and lakewater samples were conducted using different TCS concentrations (Table 1). The recovery rate of 98.9-100% shows excellent recovery of the developed TCS imprinted electrochemical sensor. For a comparison, LC–MS as a sensitive method was further performed ¹⁴ and no significant difference between the LC–MS and DPV was found based on the Wilcoxon test ($T_{calculated} > T_{tabulated}$, p > 0.05) (Table 2).

302

Sample	Added TCS	Found TCS	Recovery
	(nM)	(nM)	(%)
	-	3.10 ± 0.03	-
	3.0	6.03 ± 0.02	98.9 ± 0.5
Wastewater	6.0	9.02 ± 0.04	99.1 ± 0.4
	9.0	12.1 ± 0.02	99.8 ± 0.2
	-	1.66 ± 0.06	-
Lakewater	3.0	4.65 ± 0.04	99.8 ± 0.6
	6.0	9.67 ± 0.02	100.1 ± 0.2
	9.0	10.6 ± 0.05	99.8 ± 0.3

Table 1. The TCS recoveries in wastewater and lakewater samples (n = 6)

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Table 2. Comparison of the analytic results from DPV and LC–MS methods in determination of TCS (n = 6) (Added TCS = 6.0 nM)

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	Found	I TCS
Sample	DPV	LC-MS
Wastewater (nM)	9.02 ± 0.04	8.95 ± 0.04
SD	0.09	0.09
RSD	1.0	1.0
Lakewater (nM)	9.67 ± 0.02	9.64 ± 0.04
SD	0.08	0.07
RSD	0.8	0.7

309 X: Mean ± Standard Error, SD: Standard Deviation, RSD: % Relative Standard Deviation

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311 3.6. Reproducibility, Stability and Selectivity of the MIP/AuNPs/POM/rGO/GC sensors

312 For the sensor, TCS selectivity was investigated by detection of other chemicals with the similar structure, TCC, MTC, PCP and TCP. TCS imprinted sensor presents 5.0, 6.6, 9.0 and 18.0 313 314 times as high as the signal for TCC, MTC, PCP and TCP, respectively (Fig. 7A and 7B), indicating 315 its good selectivity. The interference values of TCC, MTC, PCP and TCP are 22.0%, 15.1%, 11.1% 316 and 5.5%, respectively. In addition, the developed MIP sensor was applied to blank solution, 15.0 317 nM standard TCS and wastewater sample containing 15.0 nM TCS to investigate matrix effect. The 318 voltammogram obtained from wastewater sample containing 15.0 nM TCS was identical with the 319 voltammogram obtained from standard solution containing an equivalent of TCS (Fig. 7C). Thus, 320 we produced a highly selective sensor via creating binding sites that are specific to the target 321 molecule.



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323	Figure 7. (A) DPVs of imprinted electrochemical sensor towards 15.0 nM TCS, TCC, MTC, PCP
324	and TCP (B) The values of peak current of TCS, TCC, MTC, PCP and TCP (C) DPVs of imprinted
325	electrochemical sensor towards (a) blank solution, (b) 15.0 nM standard TCS solution and (c)
326	wastewater sample containing 15.0 nM TCS.

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For reproducibility study, six different MIP/AuNPs/POM/rGO/GC electrodes were prepared under the same condition and tested in TCS detection and analysis. After that, each MIP electrode was applied to wastewater samples for TCS analysis. According to the obtained results, the relative standard deviation (RSD) is 0.3% in 10.0 nM TCS.

The stability of MIP/AuNPs/POM/rGO/GC electrode was also checked. After 30 days, the signal was found to be approximate 98.8% of the original value which suggests its excellent longterm stability. Table 3 presents a comparison of the sensor performance in terms of linear range and LOD with other analytical methods. It is seen that the developed sensor showed a much lower limit of detection.

Method	Linear Range	LOD	Reference
Improved electro– oxidation	$5.1 \times 10^{-9} - 6.9 \times 10^{-6} \text{ mol } \text{L}^{-1}$	$4.5 \times 10^{-9} \text{ mol } \text{L}^{-1}$	1
UV–vis micro- spectrophotometry	$4.8 \times 10^{-8} - 5.9 \times 10^{-6} \text{ mol } \text{L}^{-1}$	$3.0 \times 10^{-9} \text{ mol } \text{L}^{-1}$	33
Spectrophotometry	$0 - 1.0 \times 10^{-4} \text{ mol } \text{L}^{-1}$	$2.7 \times 10^{-4} \text{ mol } \text{L}^{-1}$	34
Voltammetric	$8.6 \times 10^{-9} - 2.0 \times 10^{-7} \text{ mol } \text{L}^{-1}$	$6.5 \times 10^{-9} \text{ mol } \text{L}^{-1}$	35
HPLC	$2.7 \times 10^{-8} - 2.0 \times 10^{-7} \text{ mol } \text{L}^{-1}$	$6.9 \times 10^{-9} \text{ mol } \text{L}^{-1}$	36
ISE/MIP	$1.0 \times 10^{-8} - 1.0 \times 10^{-5} \text{ mol } \text{L}^{-1}$	$1.9 \times 10^{-9} \text{ mol } \text{L}^{-1}$	37
CNTs@TCS-MIP	$3.4 \times 10^{-9} - 1.4 \times 10^{-7} \text{ mol } \text{L}^{-1}$	$3.4 \times 10^{-9} \text{ mol } \text{L}^{-1}$	38
MIP/amperometric	$2.0 \times 10^{-7} - 3.0 \times 10^{-6} \text{ mol } \text{L}^{-1}$	$8.0 \times 10^{-8} \text{ mol } \text{L}^{-1}$	5
MIP/AuNPs/POM/rGO	$5.1 \times 10^{-10} - 5.0 \times 10^{-8} \text{ mol } \text{L}^{-1}$	$1.5 \times 10^{-10} \text{ mol } \text{L}^{-1}$	This study

337 T	able 3. Co	omparison	of the	performances	of different	methods for	TCS	analysis
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339 4. CONCLUSION

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340	A new	TCS imprinted electrochemical sensor based on AuNPs/POM/rGO modified GC electrode	
341	was pi	repared and tested for determination of trace TCS in aqueous solution. The prepared sensor	
342	exhibi	ts high selectivity and sensitivity in TCS detection with a detection limit of 0.15 nM. It	
343	demonstrates analytic capability comparable to other complicated methods but it offers simple and		
344	efficie	nt application in target detection from wastewater and lakewater samples.	
345			
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- 415 **Figure Caption**
- 416 Scheme 1. Fabrication procedure of the MIP/AuNPs/POM/rGO/GC sensor.
- 417 Fig. 1. (A) TEM image of the AuNPs/POM/rGO, (B) EDX analysis of AuNPs/POM/rGO, (C)
- RAIRS spectra of rGO, POM/rGO and AuNPs/POM/rGO, (D) XPS spectra of AuNPs/POM/rGO
 surface.
- 420 Fig. 2. SEM images of bare GC (A), rGO/GC (B), POM/rGO/GC (C), AuNPs/POM/rGO/GC (D)
 421 and MIP/AuNPs/POM/rGO/GC (E) surfaces.
- **Fig. 3.** Cyclic voltammogram for the electrochemical polymerization of phenol with TCS in phosphate buffer solution. Phenol (80 mM), TCS (20 nM), pH 7.0, and scan rate of 100 mV s⁻¹ for 20 cycles.
- 425 Fig. 4. (A) EIS of (a) bare GC; (b) rGO/GC; (c) POM/rGO/GC; (d) AuNPs/POM/rGO/GC electrodes in 1.0 mM [Fe(CN)₆]^{3-/4-} solution in 0.1 M KCl, (B) EIS of (a) AuNPs/POM/rGO/GC; (b) 426 427 MIP/AuNPs/POM/rGO/GC (with template molecule); (c) MIP/AuNPs/POM/rGO/GC (removing template); (d) after rebinding of TCS (10.0 nM) in 1.0 mM $[Fe(CN)_6]^{3-/4-}$ solution in 0.1 M KCl, (C) 428 429 DPVs of different electrodes in 0.1 M phosphate buffer (pH 7.0) (a) MIP/AuNPs/POM/rGO/GC in 430 blank buffer solution, (b) NIP/AuNPs/POM/rG/GC after rebinding of 10.0 nM TCS, (c) 431 MIP/AuNPs/POM/rGO/GC after rebinding of 10.0 nM TCS, (D) DPV curves of different MIP 432 electrodes in 0.1 M phosphate buffer (pH 7.0) after rebinding of 10.0 nM TCS (a) GC; (b) rGO/GC; 433 (c) POM/rGO/GC; (d) AuNPs/POM/rGO/GC electrodes.
- Fig. 5. Effects of the concentration of AuNPs/POM/rGO (A), pH (B), molar ratio TCS molecule to
 phenol monomer (C), elution time (D) and temperature (E).
- Figure 6. DPV profiles of the electrochemical sensor at different TCS concentrations in phosphate
 solution pH 7.0 from background without TCS to 50.0 nM TCS (A), and linear calibration curve of
 TCS (B).

- 439 Figure 7. (A) DPVs of imprinted electrochemical sensor towards 15.0 nM TCS, TCC, MTC, PCP
- 440 and TCP (B) The values of peak current of TCS, TCC, MTC, PCP and TCP (C) DPVs of imprinted
- 441 electrochemical sensor towards (a) blank solution, (b) 15.0 nM standard TCS solution and (c)
- 442 wastewater sample containing 15.0 nM TCS.