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1 **Nb-doped VO_x/CeO₂ catalyst for NH₃-SCR of NO_x at low**
2 **temperatures**

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12

13 **Abstract**

14 The promotion effect of Nb addition to VO_x/CeO₂ catalyst for the selective catalytic
15 reduction of NO_x by NH₃ was fully investigated. VO_x/CeO₂ and NbO_x doped
16 VO_x/CeO₂ catalysts were characterized by N₂ physisorption, XRD, H₂-TPR and
17 NH₃-TPD. The results showed that the addition of Nb could significantly promote the
18 SCR activity of the VO_x/CeO₂ catalyst, especially in the low temperature range.
19 VO_x/CeO₂ with 30 wt.% NbO_x catalyst showed the best catalytic performance and
20 better SO₂/H₂O tolerance than VO_x/CeO₂ catalyst. 30Nb-1VO_x/CeO₂ also exhibited
21 higher NH₃-SCR activity than 3V₂O₅-WO₃/TiO₂. The lower crystallinity, the stronger
22 redox capability and the more Brønsted acid sites of the Nb-VO_x/CeO₂ catalyst were
23 all responsible for the more excellent NH₃-SCR performance. Based on kinetic
24 experiments and *in situ* DRIFTS results, it was concluded that the
25 Langmuir-Hinshelwood mechanism existed for selective catalytic reduction of NO
26 over Nb-VO_x/CeO₂, in which adsorbed NO_x species reacted with adsorbed NH₃ to
27 finally form N₂ and H₂O.

28 **Keywords**

29 Selective catalytic reduction of NO_x with NH₃; Low temperature activity;
30 Vanadium-cerium catalyst; Niobium oxide

31

32 1 Introduction

33 Nitrogen oxides (NO and NO₂), which result from automobile exhaust gas and
34 industrial combustion of fossil fuels, have been major pollutants for air pollution.¹
35 They contribute to a variety of environmentally harmful effects such as
36 photochemical smog, acid rain, and haze formation.² The selective catalytic reduction
37 of NO_x with NH₃ (NH₃-SCR) in the presence of excess oxygen is now the most
38 efficient technology for the removal of nitrogen oxides from stationary sources.^{2,3}
39 V₂O₅-WO₃(MoO₃)/TiO₂ has been widely applied as an industrial catalyst for several
40 decades.^{4,5} However, some problems still remain for V₂O₅-WO₃(MoO₃)/TiO₂, such
41 as the relatively narrow operating temperature window of 300-400 °C, low N₂
42 selectivity and high conversion of SO₂ to SO₃ at high temperatures.^{4,6,7} Besides, the
43 high concentration of ash containing K₂O, CaO, As₂O₃ in the flue gas reduces the
44 performance and longevity of V₂O₅-WO₃(MoO₃)/TiO₂ catalysts.^{8,9} These could be
45 avoided by locating the SCR unit downstream of the electrostatic precipitator unit and
46 even downstream of the desulfurizer, through the development of a highly efficient
47 low temperature SCR system.

48 In our previous study, we have developed a VO_x/CeO₂ catalyst, prepared by a
49 homogeneous precipitation method, showing excellent NH₃-SCR activity, N₂
50 selectivity and SO₂ durability.¹⁰ However, the catalytic activity was not high enough
51 for the application in the deNO_x process of exhaust gas with low temperature, such as
52 the flue gas after dust removal and desulfurization from coal-fired power plants.

53 Therefore, it is very necessary to modify this vanadium-cerium catalyst to improve
54 the low temperature activity, which is crucial for its practical utilization.

55 Niobium compound materials are of current interest as important catalysts for
56 various reactions, such as the removal of nitrogen oxides, the hydrogenation and
57 oxidative dehydrogenation of alkanes, acting as a catalyst promoter, catalyst support
58 or solid acid catalyst.¹¹ It was reported that when the niobium oxides were introduced
59 into V₂O₅/TiO₂ catalysts, the conversion of NO in NH₃-SCR reaction increased 2-4
60 times at low temperatures.¹² The addition of Nb to MnO_x-CeO₂ was also found to be
61 very effective in improving the NH₃-SCR activity and N₂ selectivity.^{13, 14} Mn₂Nb₁O_x
62 catalyst exhibited higher NH₃-SCR activity than MnO_x catalyst.¹⁵ In addition, it was
63 reported that the introduction of the third main group element (such as Mn, Fe, Co,
64 Mo) could also improve the activity, stability or SO₂ durability of the SCR catalysts.
65¹⁶⁻¹⁹ Therefore, based on our V-Ce catalyst, we can add other elements to adjust its
66 physicochemical properties, expecting to enhance the low temperature SCR activity.

67 In this study, a series of M-VO_x/CeO₂ catalysts (M = Mn, Fe, Co, Nb, Mo) were
68 prepared by the homogeneous precipitation method and were applied to the
69 low-temperature NH₃-SCR reaction. The addition of Nb could significantly promote
70 the SCR activity over the VO_x/CeO₂ catalyst. Among the catalysts with different
71 NbO_x contents, VO_x/CeO₂ with 30 wt.% NbO_x catalyst showed the best catalytic
72 performance and better SO₂/H₂O tolerance compared to VO_x/CeO₂ catalyst. The
73 lower crystallinity, the stronger redox capability and the more Brønsted acid sites of
74 the Nb-VO_x/CeO₂ catalyst were all responsible for its higher SCR activity.

75 2 Experiments

76 2.1 Catalyst synthesis and activity tests

77 The VO_x/CeO_2 oxide catalysts were prepared by a homogeneous precipitation
78 method. Aqueous solutions of NH_4VO_3 ($\text{H}_2\text{C}_2\text{O}_4$ was added to facilitate the
79 dissolution of NH_4VO_3) and $\text{Ce}(\text{NO}_3)_2$ were mixed with the desired molar ratios (the
80 mass ratio of vanadium oxide was controlled at 1 wt.%). An excess of aqueous urea
81 solution was then added to the mixed solution. The solution was heated to 90 °C and
82 held there for 12 h under vigorous stirring. After filtration and washing with deionized
83 water, the resulting precipitate was dried at 100 °C overnight and subsequently
84 calcined at 350 °C for 3 h in air. M- VO_x/CeO_2 catalysts (the mass ratio of MO_x were
85 controlled at 30 wt.%) were also prepared by homogeneous precipitation methods
86 using $\text{Mn}(\text{NO}_3)_2$, $\text{Co}(\text{NO}_3)_3$, $\text{Fe}(\text{NO}_3)_3$, NbCl_5 , and $(\text{NH}_4)_6\text{Mo}_7\text{O}_{24}$ as precursors,
87 respectively. Nb- VO_x/CeO_2 catalysts with different Nb contents (10, 30, 50 wt.%) and
88 the unpromoted 3 wt.% VO_x/CeO_2 catalyst were also prepared by the same method.
89 For comparison, the conventional 3 wt.% V_2O_5 -10 wt.% WO_3/TiO_2 and 1 wt.%
90 V_2O_5 -10 wt.% WO_3/TiO_2 were prepared by impregnation method using NH_4VO_3 ,
91 $(\text{NH}_4)_{10}\text{W}_{12}\text{O}_{41}$, $\text{H}_2\text{C}_2\text{O}_4 \cdot 2\text{H}_2\text{O}$ as precursors and anatase TiO_2 as support. After
92 impregnation, the excess water was removed in a rotary evaporator at 60 °C. The
93 samples dried at 100 °C overnight and then calcined at 500 °C for 3h in air condition.

94 Before NH_3 -SCR activity tests, the catalysts were pressed, crushed and sieved to
95 40-60 mesh. The SCR activity tests were carried out in a fixed-bed quartz flow reactor

106 at atmospheric pressure. The reaction conditions were controlled as follows: 500 ppm
107 NO, 500 ppm NH₃, 5 vol.% O₂, 100 ppm SO₂ (when used), N₂ balance. Under
108 ambient conditions, the total flow rate was 500 ml/min and the gas hourly space
109 velocity (GHSV) was 50 000 h⁻¹. The amount of catalysts used in activity tests was
110 0.6 ml (about 0.7 g). The effluent gas including NO, NH₃, NO₂ and N₂O was
111 continuously analyzed by an FTIR spectrometer (Nicole Nexus 670) equipped with a
112 heated, low-volume multiple-path gas cell (2m). The FTIR spectra were collected
113 after the SCR reaction reached a steady state, and the NO_x conversion and N₂
114 selectivity were calculated as follows:

$$NO_x \text{ conversion} = \left(1 - \frac{[NO]_{out} + [NO_2]_{out}}{[NO]_{in} + [NO_2]_{in}} \right) \times 100\%$$
$$N_2 \text{ selectivity} = \frac{[NO]_{in} + [NH_3]_{in} - [NO_2]_{out} - 2[N_2O]_{out}}{[NO]_{in} + [NH_3]_{in}} \times 100\%$$

115 2.2 Characterization of catalysts

116 The surface area and pore characterization of the catalysts were obtained from N₂
117 adsorption/desorption analysis at -196 °C using a Quantachrome Quadrasorb SI-MP.
118 Prior to the N₂ physisorption, the catalysts were degassed at 300 °C for 5 h. Surface
119 areas were determined by the BET equation in the 0.05-0.35 partial pressure range.
120 Pore volumes and average pore diameters were determined by the
121 Barrett-Joyner-Halenda (BJH) method from the desorption branches of the isotherms.

122 Powder X-ray diffraction (XRD) measurements of the catalysts were carried out on
123 a computerized PANalytical X'Pert Pro diffractometer with Cu Kα (λ = 0.15406 nm)

114 radiation. The data of 2θ from 10 to 80° were collected at $8^\circ/\text{min}$ with the step size of
115 0.07° .

116 The H_2 -TPR experiments were carried out on a Micromeritics Auto Chem 2920
117 chemisorption analyzer. The samples (30 mg) were pretreated at 300°C in a flow of
118 20 vol.% O_2/Ar (50 ml/min) for 0.5 h in a quartz reactor and cooled down to room
119 temperature (30°C) followed by Ar purging for 0.5 h. A 50 mL/min gas flow of 10%
120 H_2 in Ar was then passed over the samples through a cold trap to the detector. The
121 reduction temperature was raised at $10^\circ\text{C min}^{-1}$ from 30 to 1000°C .

122 2.3 NH_3 -TPD studies

123 NH_3 -TPD experiments were performed in the same instrument as the H_2 -TPR,
124 equipped with a quadrupole mass spectrometer (MKS Cirrus) to record the signals of
125 NH_3 ($m/z = 17$ for NH_3 , the interference of H_2O was eliminated by using a cold trap
126 before the detector). Prior to TPD experiments, the samples (100 mg) were pretreated
127 at 300°C in a flow of 20 vol.% O_2/Ar (50 ml/min) for 0.5 h and cooled down to room
128 temperature (30°C). The samples were then exposed to a flow of 2500 ppm NH_3/Ar
129 (50 ml/min) at 30°C for 1 h, followed by Ar purging for another 1 h. Finally, the
130 temperature was raised to 600°C in Ar flow at the rate of $10^\circ\text{C min}^{-1}$.

131 2.4 *In situ* DRIFTS studies

132 *In situ* DRIFTS experiments were performed on an FTIR spectrometer (Nicolet
133 Nexus 670) equipped with a smart collector and an MCT/A detector cooled by liquid
134 nitrogen. The reaction temperature was controlled precisely by an Omega

135 programmable temperature controller. Prior to each experiment, the sample was
136 pretreated at 300 °C for 0.5 h in a flow of 20 vol.% O₂/N₂ and then cooled down to
137 175 °C. The background spectra were collected in flowing N₂ and automatically
138 subtracted from the sample spectrum. The reaction conditions were controlled as
139 follows: 200 ml/min total flow rate, 500 ppm NH₃ or/and 500 ppm NO + 5 vol.% O₂,
140 and N₂ balance. All spectra were recorded by accumulating 100 scans with a
141 resolution of 4 cm⁻¹.

142 **3 Results and Discussion**

143 3.1 Catalytic performance

144 3.1.1 SCR activity over 30M-1VO_x/CeO₂

145 Fig. 1 shows the NO_x conversion and N₂ selectivity over VO_x/CeO₂ and doped
146 VO_x/CeO₂ catalysts. The VO_x/CeO₂ catalyst presented 80% NO_x conversion at 200 °C.
147 Mn doped VO_x/CeO₂ exhibited nearly 100% NO_x conversion at 100-200 °C, but the
148 NO_x conversion decreased rapidly above 200 °C. It also showed lower N₂ selectivity
149 in the whole temperature that we investigated. NH₃-SCR activities over Fe, Co and
150 Mo doped VO_x/CeO₂ catalysts were all lower than that over VO_x/CeO₂. Among these
151 catalysts, Co doped VO_x/CeO₂ catalyst presented the lowest NO_x conversion and the
152 maximal NO_x conversion was only 50%. Furthermore, the N₂ selectivity over
153 Co-VO_x/CeO₂ was rather low. Contrarily, Nb doped VO_x/CeO₂ catalyst exhibited
154 higher catalytic performance than the VO_x/CeO₂ catalyst. The addition of Nb
155 enhanced the NO_x conversion and N₂ selectivity over VO_x/CeO₂ simultaneously. It

156 showed 60% and 90% NO_x conversion at 150 and 175 °C, respectively, and 100% N₂
157 selectivity was obtained in the temperature range of 100-350 °C. The best NH₃-SCR
158 activity was obtained over the Nb doped VO_x/CeO₂ catalyst, therefore we chose Nb to
159 further investigate the influence of doping amount on the catalyst structure and
160 catalytic performance.

161 3.1.2 SCR activity over Nb-1VO_x/CeO₂ catalysts

162 The effect of NbO_x addition amount to VO_x/CeO₂ catalyst on NO_x conversion was
163 shown in Fig. 2. 10 wt.% NbO_x doped VO_x/CeO₂ catalyst showed higher NH₃-SCR
164 activity than unpromoted VO_x/CeO₂ catalyst. The 30Nb-VO_x/CeO₂ catalyst presented
165 the best catalytic activity, over which NO_x was completely reduced at about 175 °C.
166 Any further increase in Nb content resulted in a decline in activity. These indicated
167 that Nb content had a significant effect on the SCR activity over VO_x/CeO₂ catalysts,
168 and 30 wt.% NbO_x was optimal. Although, for the 30Nb-VO_x/CeO₂ catalyst, the
169 content of 1% V seems to be negligible comparing with 30% Nb, V indeed played an
170 important role in NH₃-SCR reaction and 30Nb-VO_x/CeO₂ exhibited much higher
171 catalytic activity than 30Nb/CeO₂ (as shown in Fig. S1). At lower than 175 °C, the
172 reaction rates normalized by surface area over 30Nb-VO_x/CeO₂ was higher than that
173 over VO_x/CeO₂ (as shown in Fig. S2). The highest reaction rate over 30Nb-VO_x/CeO₂
174 was obtained at 175-350 °C. The lower reaction rates over 30Nb-VO_x/CeO₂ than
175 VO_x/CeO₂ above 200 °C were due to its larger specific surface area and similar NO_x
176 conversion. 30Nb-1VO_x/CeO₂ and 1VO_x/CeO₂ catalyst were chosen as the model
177 catalysts to carry out further investigation, such as the effect of Nb on H₂O/SO₂

178 tolerance and the relationship between catalyst structure and catalytic activity.

179 To better evaluate the NH_3 -SCR activity over $\text{Nb-VO}_x/\text{CeO}_2$ catalyst, we also
180 carried out the comparative SCR activity test over $\text{V}_2\text{O}_5\text{-WO}_3/\text{TiO}_2$ (Fig. 3). 30 wt.%
181 $\text{Nb-1VO}_x/\text{CeO}_2$ exhibited higher NO_x conversion than $1\text{VO}_x/\text{CeO}_2$ and $3\text{VO}_x/\text{CeO}_2$.
182 However, there was not notable enhancement over $30\text{Nb-3VO}_x/\text{CeO}_2$ in contrast to
183 $3\text{VO}_x/\text{CeO}_2$ (as shown in Fig. S3). Compared to $3\text{VO}_x/\text{CeO}_2$, the $30\text{Nb-1VO}_x/\text{CeO}_2$
184 catalyst not only decreased the content of vanadium oxide, but also enhanced the
185 catalytic activity. The NH_3 -SCR performance over $30\text{Nb-1VO}_x/\text{CeO}_2$ catalyst was
186 also better than that over $3\text{V}_2\text{O}_5\text{-10WO}_3/\text{TiO}_2$ and $1\text{V}_2\text{O}_5\text{-10WO}_3/\text{TiO}_2$. The NO_x
187 conversion over three catalysts at 175 °C was 96%, 63% and 15%, respectively.
188 Therefore, $30\text{Nb-1VO}_x/\text{CeO}_2$ showed excellent NH_3 -SCR performance.

189 3.1.3 Influence of H_2O and SO_2 on SCR activity

190 Fig. 4 shows the effect of SO_2 and H_2O on the catalytic activity over $1\text{VO}_x/\text{CeO}_2$
191 and $30\text{Nb-1VO}_x/\text{CeO}_2$ catalysts at 250 °C. When 100 ppm SO_2 was introduced to the
192 inlet gas, the NO_x conversion over VO_x/CeO_2 decreased to 24% in 48 h and could not
193 recover to the initial activity after the removal of SO_2 , which indicates that the
194 inhibiting effect of SO_2 on the SCR activity over the VO_x/CeO_2 catalyst was severe
195 and irreversible. However, the SO_2 inhibiting effect over $\text{Nb-VO}_x/\text{CeO}_2$ was quite
196 different. The NO_x conversion decreased slightly, and nearly 90% NO_x conversion
197 was obtained in the presence of 100 ppm SO_2 for a 48 h test. The NH_3 -SCR
198 performance over VO_x/CeO_2 catalysts after SO_2 poisoning for 48 h is shown in Fig.

199 S4. The activity over Nb-VO_x/CeO₂ was still higher than that over VO_x/CeO₂. 95%
200 NO_x conversion could be obtained on the Nb-VO_x/CeO₂ catalyst at 250 °C and only
201 26% NO_x conversion on VO_x/CeO₂. This proved again that the Nb-VO_x/CeO₂ catalyst
202 showed higher SO₂ resistance than VO_x/CeO₂.

203 When 5% H₂O was introduced to the inlet gas, the NO_x conversion over VO_x/CeO₂
204 decreased rapidly to 56% and kept in 56% in 48 h test. The catalytic activity could
205 recover to the original level after the removal of H₂O, indicating that the poison of
206 H₂O was reversible. Meanwhile, H₂O had no influence on the catalytic activity over
207 Nb-VO_x/CeO₂ catalyst and 100% NO_x conversion was maintained all the time.
208 Nb-VO_x/CeO₂ exhibited much higher catalytic activity and stronger resistance to
209 SO₂/H₂O than VO_x/CeO₂.

210 3.2 Catalyst Characterization

211 3.2.1 N₂ physisorption

212 Table 1 shows the N₂ physisorption results of 1VO_x/CeO₂ and 30Nb-1VO_x/CeO₂
213 catalysts. Nb-VO_x/CeO₂ catalyst presented larger specific surface area and pore
214 volume than VO_x/CeO₂. The addition of Nb to VO_x/CeO₂ resulted in bigger specific
215 surface area and pore volume, which is beneficial to the enhancement of SCR activity.

216 3.2.2 XRD

217 The XRD patterns of 1VO_x/CeO₂ and 30Nb-1VO_x/CeO₂ catalysts are shown in Fig.
218 5. For both catalysts, all the peaks in the diffraction profiles attributed to CeO₂ of a
219 cubic fluorite structure (43-1002). No vanadium species and niobium species were

220 detected, suggesting that V and Nb species were highly dispersed on the catalysts.
221 The intensity of peaks of Nb doped catalyst was weaker than that of VO_x/CeO_2 ,
222 indicating a loss of crystallinity. The decreasing crystallinity may contribute to the
223 increase of surface area.

224 3.2.3 H_2 -TPR

225 The redox properties of $1\text{VO}_x/\text{CeO}_2$ and $30\text{Nb-}1\text{VO}_x/\text{CeO}_2$ catalysts were
226 investigated by H_2 -TPR and the spectra are shown in Fig. 6. There are three peaks at
227 380, 460 and 710 °C (740 °C) over $1\text{VO}_x/\text{CeO}_2$ and $30\text{Nb-}1\text{VO}_x/\text{CeO}_2$. According to
228 the literatures^{10,20}, the reduction peaks of surface Ce^{4+} to Ce^{3+} and the bulk Ce^{4+} to
229 Ce^{3+} were centered at 509 and 812 °C, respectively. The reduction peaks at 399, 523
230 and 766 °C over CeO_2 could be assigned to the reduction of surface oxygen species,
231 that of oxygen in deeper interior and that of oxygen in bulk, respectively.²¹ The
232 reduction peak around 400-500 °C is due to the surface vanadium oxide, while high
233 temperature peak over 700 °C is assigned to the reduction of bulk vanadium oxide.²²
234 Niobium oxide could be reduced at much higher temperature.¹⁵ Therefore, the low
235 temperature peak at 380 °C and 460 °C could be due to the reduction of surface Ce^{4+}
236 and V^{5+} and that in deeper interior, respectively. The TPR peak at 710 °C (740 °C)
237 could be attributed to the reduction of bulk Ce and V. Both V and Ce possessed redox
238 capability and acted as reactive sites to catalyze NH_3 -SCR reaction. The amount of H_2
239 consumption of Nb- VO_x/CeO_2 catalyst was higher than that of VO_x/CeO_2 (1.15:1).
240 The intensity of the low temperature (380 °C) reduction peak of Nb- VO_x/CeO_2 was
241 much stronger than that of VO_x/CeO_2 (1.44:1). This could be due to the interaction of

242 V, Ce and Nb over Nb-VO_x/CeO₂ catalyst resulting in better dispersion of vanadium
243 species and stronger redox capability. The active temperature window in NH₃-SCR
244 reaction was between 150-400 °C. Stronger redox capability in this temperature range
245 could enhance NH₃-SCR performance. Therefore, Nb doped VO_x/CeO₂ showed
246 higher NH₃-SCR activity.

247 3.2.4 NH₃-TPD

248 Fig. 7 shows NH₃-TPD results over 1VO_x/CeO₂ and 30Nb-1VO_x/CeO₂ catalysts
249 using the fragment of $m/z = 17$ to identify NH₃. There were two NH₃ desorption peaks
250 around 100 and 250 °C on both catalysts. It is generally accepted that NH₄⁺ ions
251 bound to Brønsted acid sites are less thermally stable than coordinated NH₃ molecules
252 bound to Lewis acid sites and desorb at lower temperatures.^{18, 23, 24} Therefore, the
253 desorption peak at 100 °C could be ascribed to the desorption of physisorbed NH₃ and
254 the partial ionic NH₄⁺ bound to the weak Brønsted acid sites, and the peak at 250 °C
255 could be assigned to the desorption of ionic NH₄⁺ bound to strong Brønsted acid sites
256 and coordinated NH₃ bound to the Lewis acid sites. The amount of acid sites over
257 Nb-VO_x/CeO₂ was remarkably larger than that of VO_x/CeO₂ (1.68:1). It indicates that
258 Nb-VO_x/CeO₂ catalyst presented more acid sites which could facilitate the adsorption
259 and activation of NH₃ during catalytic reaction even in the presence of H₂O and SO₂
260 thus enhances the catalytic activity in NH₃-SCR. In the presence of SO₂, more acid
261 sites over Nb-VO_x/CeO₂ catalyst inhibited the adsorption of SO₂ and the deposition of
262 sulfate thus enhanced the resistance to SO₂. More acid sites could also promote the
263 adsorption of NH₃ in the presence of H₂O and weaken the competitive adsorption of

264 H₂O with NH₃. Furthermore, larger specific surface area was obtained over
265 Nb-VO_x/CeO₂. Therefore, Nb-VO_x/CeO₂ catalyst showed stronger resistance to
266 H₂O/SO₂.

267 3.3 *In situ* DRIFTS studies

268 To investigate NH₃/NO_x adsorption on 1VO_x/CeO₂ and 30Nb-1VO_x/CeO₂ catalysts
269 together with the SCR reaction mechanism, *in situ* DRIFTS were conducted at 175 °C
270 and the results of NH₃ adsorption on VO_x/CeO₂ and Nb-VO_x/CeO₂ catalysts are
271 shown in Fig. 8A. After NH₃ adsorption and N₂ purge, both catalysts were covered by
272 different NH₃ species. The bands at 1423 and 1670 cm⁻¹ were assigned to asymmetric
273 and symmetric bending vibrations of ionic NH₄⁺ on Brønsted acid sites and the bands
274 at 1596 (1604) and 1147 (1200) cm⁻¹ were attributed to asymmetric and symmetric
275 bending vibrations of the N-H bonds in NH₃ coordinately linked to the Lewis acid
276 sites.²⁵⁻²⁷ Nb-VO_x/CeO₂ catalyst provided more acid sites than VO_x/CeO₂ (1.71:1),
277 especially the Brønsted acid sites, than VO_x/CeO₂, which was in well harmony with
278 the NH₃-TPD results. As shown in our previous paper (Chemical Engineering Journal
279 250 (2014) 390-398)¹⁵, the NbO_x itself did not show any NH₃-SCR activity in the
280 whole temperature range that we investigated. However, the addition of Nb to
281 1VO_x/CeO₂ increased the surface acidity. Nb mainly played as acid sites for the
282 promotion of NH₃ adsorption in NH₃-SCR reaction.

283 The *in situ* DRIFT spectra of NO + O₂ adsorption on 1VO_x/CeO₂ and
284 30Nb-1VO_x/CeO₂ catalysts at 175 °C were also conducted, and the results are shown

285 in Fig. 8B. When VO_x/CeO_2 catalyst was exposed to $\text{NO} + \text{O}_2$, bands assigned to
286 nitrate species were observed including monodentate nitrate (1540 and 1270 cm^{-1}),
287 bridging nitrate (1215 and 1592 cm^{-1}) and bidentate nitrate (1570 and 1243 cm^{-1}).²⁸⁻³⁰
288 On the $\text{Nb-VO}_x/\text{CeO}_2$ catalyst, the adsorption amount of NO_x was larger than that on
289 the VO_x/CeO_2 catalyst (1.67:1). The addition of Nb to the VO_x/CeO_2 catalyst
290 increased the acidity but did not inhibit the adsorption of NO_x on the catalyst surface
291 simultaneously. It could be due to its stronger oxidation at low temperature and its
292 larger specific surface area. Therefore, the $\text{Nb-VO}_x/\text{CeO}_2$ catalyst produced more
293 nitrate species than VO_x/CeO_2 .

294 According to the Arrhenius equation, the activation energy over VO_x/CeO_2 and
295 $\text{Nb-VO}_x/\text{CeO}_2$ catalysts was calculated as 47 and 38 kJ/mol (as shown in Fig. S5). The
296 interaction of V, Ce and Nb of the $\text{Nb-VO}_x/\text{CeO}_2$ catalyst decreased the activation
297 energy of NO reduction and promote the NH_3 -SCR reaction.

298 To investigate the reactivity of adsorbed NH_3 species in the SCR reaction on
299 $30\text{Nb-1VO}_x/\text{CeO}_2$ catalysts, the *in situ* DRIFTS of reaction between pre-adsorbed
300 NH_3 and $\text{NO} + \text{O}_2$ at $175\text{ }^\circ\text{C}$ were recorded as a function of time (Fig. 9A). After NH_3
301 pre-adsorption and N_2 purge, the catalyst surface was covered by the adsorbed NH_3
302 species. When $\text{NO} + \text{O}_2$ was introduced, the intensity of the bands attributed to NH_3
303 species decreased and disappeared after 10 min. At the same time, the bands assigned
304 to nitrate species (monodentate nitrate at 1540 cm^{-1} , bridging nitrate at 1215 , 1592
305 cm^{-1} and bidentate nitrate at 1570 , 1243 cm^{-1}) appeared. This result suggested that
306 both ionic NH_4^+ and coordinated NH_3 could react with NO_x and participate in the

307 NH₃-SCR reactions. Fig. S6 showed the band intensities of adsorbed NH₃ species
308 over VO_x/CeO₂ and 30Nb-VO_x/CeO₂ pretreated by exposure to NH₃ followed by
309 exposure to NO + O₂ at 175 °C. The reactive rate of adsorbed NH₃ species with gas
310 NO and O₂ over Nb-VO_x/CeO₂ was much higher than that over VO_x/CeO₂. Therefore,
311 more adsorbed NH₃ species on Nb-VO_x/CeO₂ contributed to its better SCR activity.

312 The reaction between the pre-adsorbed NO_x and NH₃ on 30Nb-1VO_x/CeO₂
313 catalysts was also investigated by the *in situ* DRIFTS, and the results were shown in
314 Fig. 9B. After NO + O₂ pre-adsorption and N₂ purge, Nb-VO_x/CeO₂ catalyst surface
315 was covered by various nitrate species. When NH₃ was introduced, the intensity of
316 bridging nitrate decreased, which indicated that adsorbed NO_x species could also react
317 with NH₃. NH₃ adsorbed species were observed at the band of 1427 cm⁻¹.

318 The kinetic experiments were also carried out to investigate the reaction order and
319 the tests met the condition of differential reactor model. The NO conversion data in
320 the kinetics test were in differential regime (conversion less than 20%) (as shown in
321 Fig. S7). A relative small particles size (40-60 mesh) and the volume hourly space
322 velocity (about 500 000 h⁻¹)³¹ ensured the elimination of internal and external
323 diffusion, respectively. The rates of NO conversion increased linearly with NO
324 concentration over 30Nb-1VO_x/CeO₂ catalyst (as shown in Fig. S8), and the reaction
325 order for NO was calculated as 0.555. The reaction order for NO was lower than 1,
326 indicating the presence of Langmuir-Hinshelwood mechanism, which was in
327 agreement with DRIFTS results. In Fig. 9B, the intensity of the bands attributed to
328 nitrate species weakened after exposure to NH₃ indicating that adsorbed NO_x species

329 could react with adsorbed NH_3 to finally form N_2 and H_2O . In our previous study ¹⁰,
330 for the $3\text{VO}_x/\text{CeO}_2$ catalysts pre-adsorbed NO_x species, when NH_3 was introduced,
331 the intensity of the bands attributed to monodentate nitrate and bridging nitrate
332 species decreased slightly. The amount of bidentate nitrate species increased markedly,
333 which may be due to the transformation of monodentate and bridging nitrate to
334 bidentate nitrate (Fig. 10 ¹⁰ and Fig. S5 ¹⁰). The adsorbed nitrate species were mostly
335 inactive in the NH_3 -SCR reaction and therefore the $3\text{VO}_x/\text{CeO}_2$ catalysts mainly
336 followed the Eley-Rideal mechanism.

337 **4 Conclusions**

338 A systematic study on the effect of Nb addition to the VO_x/CeO_2 catalyst for the
339 low-temperature NH_3 -SCR reaction was carried out. VO_x/CeO_2 and Nb doped
340 VO_x/CeO_2 catalysts were prepared by homogeneous precipitation method and the
341 SCR activity at low temperature was enhanced by the addition of Nb. The NH_3 -SCR
342 activity over 1% VO_x/CeO_2 with 30 wt.% NbO_x catalyst were higher than that over the
343 3% $\text{V}_2\text{O}_5\text{-WO}_3/\text{CeO}_2$ catalyst. 30Nb-1 VO_x/CeO_2 showed higher catalytic activity than
344 1 VO_x/CeO_2 catalyst, due to the weaker crystallinity, the stronger redox capability and
345 the more Brønsted acid sites. The excellent $\text{SO}_2/\text{H}_2\text{O}$ tolerance was also obtained over
346 30Nb-1 VO_x/CeO_2 catalyst. The Langmuir-Hinshelwood mechanism existed for
347 selective catalytic reduction of NO Over 30Nb-1 VO_x/CeO_2 , in which adsorbed NO_x
348 species reacted with adsorbed NH_3 to finally form N_2 and H_2O .

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Table 1 N₂ physisorption of VO_x/CeO₂ and Nb-VO_x/CeO₂ catalysts

Catalysts	Specific surface area (m ² /g)	Pore diameter (nm)	Pore volume (cc/g)
1VO _x /CeO ₂	131.3	3.50	0.11
30Nb-1VO _x /CeO ₂	168.2	3.48	0.15

395

Figure Captions

396

397 Fig. 1 The NO_x conversion (A) and N₂ selectivity (B) over M-VO_x/CeO₂ catalysts.

398 Reaction conditions: [NO] = [NH₃] = 500 ppm, [O₂] = 5 vol. %, N₂ balance, total flow

399 rate 500 ml/min and GHSV = 50 000 h⁻¹.

400 Fig. 2 NH₃-SCR activity over Nb-VO_x/CeO₂ catalysts. Reaction conditions: [NO] =

401 [NH₃] = 500 ppm, [O₂] = 5 vol. %, N₂ balance, total flow rate 500 ml/min and GHSV

402 = 50 000 h⁻¹.

403 Fig. 3 NH₃-SCR activity over VO_x/CeO₂ and V₂O₅-WO₃/TiO₂ catalysts. Reaction

404 conditions: [NO] = [NH₃] = 500 ppm, [O₂] = 5 vol. %, N₂ balance, total flow rate 500

405 ml/min and GHSV = 50 000 h⁻¹.

406 Fig. 4 Effect of SO₂ and H₂O on NH₃-SCR activity over VO_x/CeO₂ and Nb-VO_x/CeO₂

407 catalysts at 250 °C. Reaction conditions: [NO] = [NH₃] = 500 ppm, [SO₂] = 100 ppm

408 (when used), [H₂O] = 5% (when used), [O₂] = 5 vol. %, N₂ balance, total flow rate

409 500 ml/min and GHSV = 50 000 h⁻¹.

410 Fig. 5 XRD patterns of VO_x/CeO₂ and Nb-VO_x/CeO₂ catalysts.

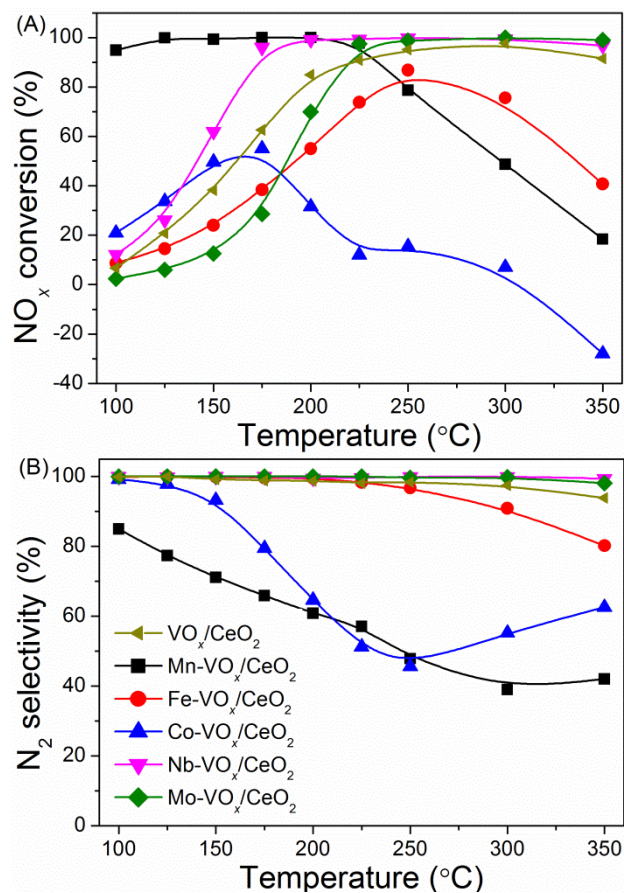
411 Fig. 6 H₂-TPR results over VO_x/CeO₂ and Nb-VO_x/CeO₂ catalysts.

412 Fig. 7 NH₃-TPD results over VO_x/CeO₂ and Nb-VO_x/CeO₂ catalysts.

413 Fig. 8 DRIFT spectra of 500 ppm NH₃ adsorption (A) and 500 ppm NO + 5 vol.% O₂

414 adsorption (B) on VO_x/CeO₂ and Nb-VO_x/CeO₂ catalysts.

415 Fig. 9 *In situ* DRIFT spectra over Nb-VO_x/CeO₂ pretreated by exposure to NO + O₂
416 followed by exposure to NH₃ at 175 °C (A), and by exposure to NH₃ followed by
417 exposure to NO + O₂ at 175 °C (B).
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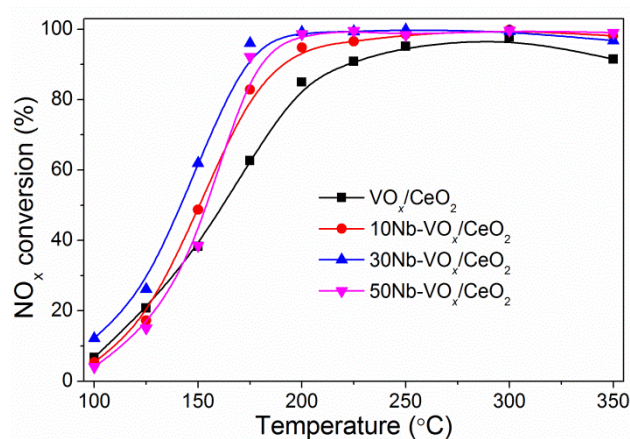


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420 Fig. 1 The NO_x conversion (A) and N₂ selectivity (B) over M-VO_x/CeO₂ catalysts.421 Reaction conditions: [NO] = [NH₃] = 500 ppm, [O₂] = 5 vol. %, N₂ balance, total flow422 rate 500 ml/min and GHSV = 50 000 h⁻¹.

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Fig. 2 NH₃-SCR activity over Nb-VO_x/CeO₂ catalysts.

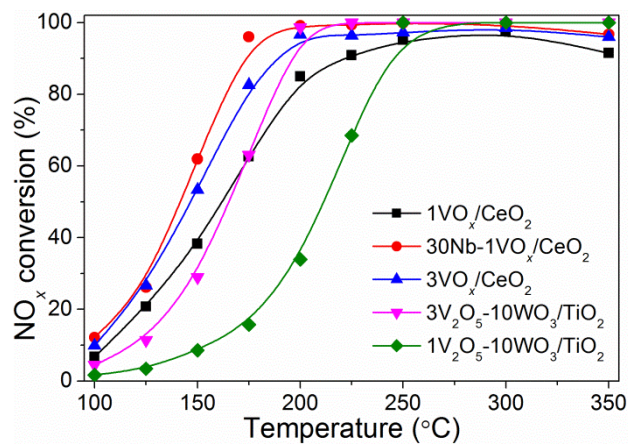
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Reaction conditions: [NO] = [NH₃] = 500 ppm, [O₂] = 5 vol. %, N₂ balance, total flow

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rate 500 ml/min and GHSV = 50 000 h⁻¹.

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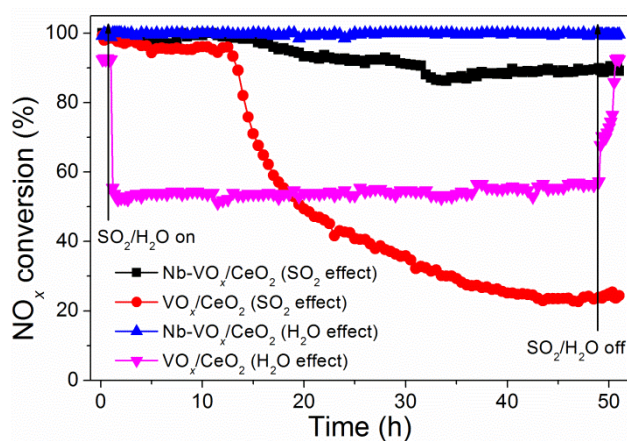
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431 Fig. 3 NH₃-SCR activity over VO_x/CeO₂ and V₂O₅-WO₃/TiO₂ catalysts.432 Reaction conditions: [NO] = [NH₃] = 500 ppm, [O₂] = 5 vol. %, N₂ balance, total flow

433

rate 500 ml/min and GHSV = 50 000 h⁻¹.

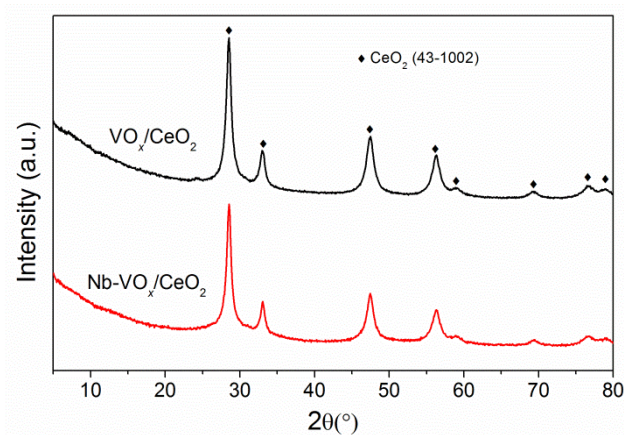
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436 Fig. 4 Effect of SO₂ and H₂O on NH₃-SCR activity over VO_x/CeO₂ and Nb-VO_x/CeO₂437 catalysts at 250 °C. Reaction conditions: [NO] = [NH₃] = 500 ppm, [SO₂] = 100 ppm438 (when used), [H₂O] = 5% (when used), [O₂] = 5 vol. %, N₂ balance, total flow rate439 500 ml/min and GHSV = 50 000 h⁻¹.

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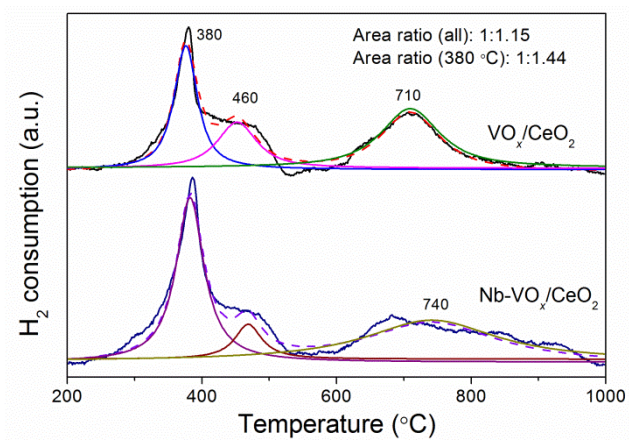


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Fig. 5 XRD patterns of VO_x/CeO_2 and $\text{Nb-VO}_x/\text{CeO}_2$ catalysts.

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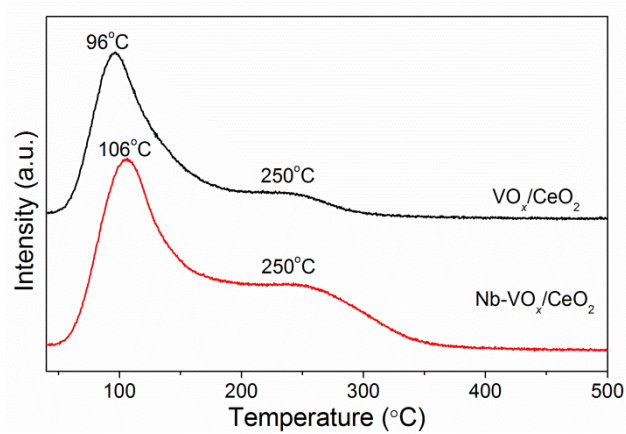


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Fig. 6 H₂-TPR results over VO_x/CeO₂ and Nb-VO_x/CeO₂ catalysts.

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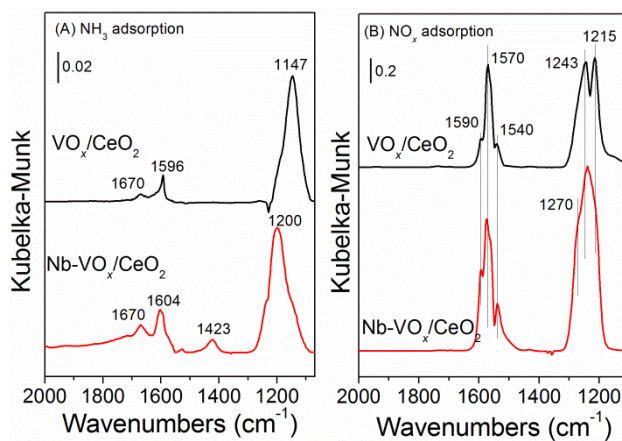


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Fig. 7 NH₃-TPD results over VO_x/CeO₂ and Nb-VO_x/CeO₂ catalysts.

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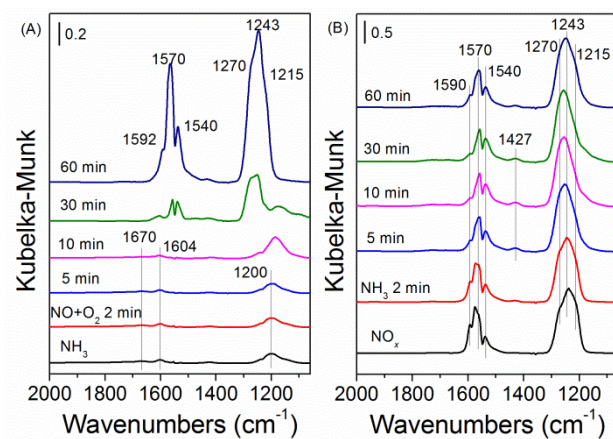
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451 Fig. 8 DRIFT spectra of 500 ppm NH₃ adsorption (A) and 500 ppm NO + 5 vol.% O₂

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adsorption (B) on VO_x/CeO₂ and Nb-VO_x/CeO₂ catalysts.

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455 Fig. 9 *In situ* DRIFT spectra over Nb-VO_x/CeO₂ pretreated by exposure to NO + O₂456 followed by exposure to NH₃ at 175 °C (A), and by exposure to NH₃ followed by457 exposure to NO + O₂ at 175 °C (B).

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