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The existence of water vapor in the feed stream could led to a sharp decrease in the VOCs working capacity due to its strong competitive adsorption on the surfaces of MIL-101.

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| 1 | Competitive Adsorption of water vapor with |
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| 2 | VOCs Dichloroethane, Ethyl Acetate and |
| 3 | Benzene on MIL-101(Cr) in humid |
| 4 | atmosphere |
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| 9 | Abstract: It is well-known that water vapor is omnipresent. It would inevitably have |
| 10 | negative influence on VOCs adsorption on novel porous materials in the cases of |
| 11 | virtual situations. In this work, competitive adsorption behavior of water vapor with |
| 12 | three VOCs such as 1,2-dichloroethane(DCE), ethyl acetate(EA) and benzene on |
| 13 | MIL-101 in humid atmosphere were investigated by isotherm measurement, |
| 14 | breakthrough experiments and TPD experiments. Results showed that adsorption |
| 15 | capacities of MIL-101 for DCE, EA and benzene were separately up to 9.71, 5.79 and |
| 16 | 3.76 mmol/g, much higher than those of other conventional adsorbents. Breakthrough |
| 17 | experiments indicated that presence of water vapor in feed stream resulted in a sharp |
| 18 | decrease in the VOCs working capacities of MIL-101 due to competitive adsorption |
| 19 | of water vapor on MIL-101 surfaces. The breakthrough times and the working |
| 20 | capacities of these VOCs became smaller with an increase in the relative humidity. |
| 21 | TPD experiments indicated that the desorption activation energies of water vapor, |

| 22 | DCE, EA and benzene on MIL-101 were 72.9, 47.14, 41.9, and 38.16 kJ/mol, |
|----|---|
| 23 | respectively. The stronger interaction of water vapor with MIL-101 formed strong |
| 24 | competitive adsorption with VOCs on MIL-101, resulting in the sharp decrease of the |
| 25 | VOCs working capacities in humid atmosphere. |
| 26 | Key words: Adsorption, TPD, breakthrough, MIL-101, water vapor, VOCs |
| | |

27

28 1.Introduction

29 In recent decades, the release of anthropogenic toxic pollutants into the atmosphere 30 is a worldwide threat of growing concern. Air pollution is now recognized as a severe 31 problem and has become increasingly serious both nationally and worldwide¹. 32 Volatile organic compounds (VOCs) are the main pollutants in the ambient air 33 released from chemical, petrochemical, and related industries. VOCs may cause short-34 and long-term adverse health risk for humans, such as headaches and eve, nose and 35 throat irritation, dizziness, nausea, cancer, and even death even at very low 36 concentrations. More and more countries and regions have proposed stringent 37 legislations to impose stringent standards on VOC emissions from industries. 38 Therefore, it is urgently needed to develop more safe and efficient systems for the 39 removal of VOCs from polluted air.

There are many techniques available to abate the emission of VOCs, such as adsorption^{2,3}, catalytic oxidation⁴, condensation⁵, membrane separation⁶ and biological treatments. Among these, adsorption method has been considered as one of the most cost-effective and environmentally friendly technologies for the removal of VOCs, especially at low concentration. Adsorbents play a key role in the adsorption

technique. Up to now, much work has been conducted to investigate the adsorption of
VOCs on some traditional absorbents, such as activated carbon (AC), silica gel,
activated alumina and zeolites. Although activated carbon and zeolites were widely
applied for adsorption of VOCs, their disadvantages of the low capacity and difficulty
in regeneration have limited its more wide application⁷.

50 In recent years, a new class of porous materials assembled with metal ions and 51 organic linkers, known as metal-organic frameworks (MOFs), has been rapidly 52 developed as potential materials for VOCs adsorption due to their ultrahigh surface area, sturdy, open crystalline structure and adjustable chemical functionality ^{8,9}. The 53 presence of open metal sites (coordinatively unsaturated metal centers) or certain 54 55 functionalizations on the pore surfaces of MOFs could enhance the adsorption 56 selectivity/efficiency of MOFs towards certain toxic compounds via coordination 57 bonds, acid–base/electrostatic interactions, p-complex/H-bonding formation, etc⁸. It is 58 possible to take advantage of MOFs materials in order to develop new technologies for environmental remediation purposes. Barea et al.8 reviewed performances of 59 60 MOFs in environmental remediation processes, and their studies showed that some of 61 MOFs and modified MOFs exhibited remarkable adsorption capacities and good 62 selectivities of organic molecules. Among these MOFs, MIL-101(Matérial Institut 63 Lavoisier, chromium-terphthalate-based solid) is considered as one of the most 64 prominent representative in MOFs, which is basically built up from a hybrid 65 superterahedral (ST) building unit, which is formed by rigid terephthalate ligands and 66 trimeric chromium(III) octahedral clusters. It possesses a very large specific surface 67 area with ordered micro/mesoporous zeotype architecture and high chemical and thermal stability¹⁰. Therefore, MIL-101 may become a promising candidate as an 68 adsorbent for VOCs capture applications. Zhao et al.^{11, 12} measured the adsorption 69

| 70 | isotherms and kinetics of benzene and p-xylene on MIL-101, and reported that the |
|--|--|
| 71 | maximum capacities of MIL-101 were 16.5 and 10.9 mmol/g at 288 K, respectively, |
| 72 | much higher than activated carbons and zeolites. In addition, multiple cycle |
| 73 | experiments of these VOCs Adsorption-Desorption on MIL-101 were carried out, and |
| 74 | their results showed that efficiency of VOCs desorption can reach over 97%. Shi et |
| 75 | al. ¹³ reported that the adsorption capacity of MIL-101 for ethyl acetate is up to 10.5 |
| 76 | mmol/g at 288 K and 54 mbar. Diffusion coefficients of ethyl acetate within |
| 77 | MIL-101 are in the range of (1.617 to 2.264) \cdot 10 ⁻¹⁰ cm ² \cdot s ⁻¹ with a lower activation |
| 78 | energy of 8.361 kJ•mol ⁻¹ . In addition, multiple cycle experiments of these VOCs |
| 79 | Adsorption-Desorption on MIL-101 were carried out, and results showed that |
| 80 | desorption efficiency of VOCs can reach over 97% ^{11,13} . Thuyet al. ¹⁴ measured the |
| 81 | isotherms of C6-C9 on MIL-101, and then reported that its maximum capacities for |
| 82 | C6, C7, C8 and C9 reached about 9.95 , 8.82 , 8.75 and 6.17 mmol/g respectively. |
| 83 | Yang et al. ¹⁵ reported that MIL-101 had higher adsorption capacities for selected |
| | |
| 84 | VOCs than zeolite, activated carbon and other reported adsorbents, and thus it was |
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| 84 85 86 87 88 89 | VOCs than zeolite, activated carbon and other reported adsorbents, and thus it was suitable for the adsorptive removal of VOCs including polar acetone and nonpolar benzene, toluene, ethylbeznene, and xylenes. Huang et al. ¹⁶ measured the isotherms of n-hexane, toluene, methanol, butanone, dichloromethane, and n-butylamine on MIL-101, and reported that MIL-101 had much higher affinity and adsorption capacity to VOCs than activated carbon, and its affinity to n-butylamine was the |
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| 84 85 86 87 88 89 90 91 91 92 93 | VOCs than zeolite, activated carbon and other reported adsorbents, and thus it was suitable for the adsorptive removal of VOCs including polar acetone and nonpolar benzene, toluene, ethylbeznene, and xylenes. Huang et al. ¹⁶ measured the isotherms of n-hexane, toluene, methanol, butanone, dichloromethane, and n-butylamine on MIL-101, and reported that MIL-101 had much higher affinity and adsorption capacity to VOCs than activated carbon, and its affinity to n-butylamine was the strongest among these VOCs. These studies above showed that the capacities of MIL-101 for VOCs were several times as that of the ACs, showing a great application prospect in the fields of environmental remediation and protection. However, if MIL-101 is used as a novel adsorbent applied in the actual cases, it will face the |

95 It is well-known that water vapor is omnipresent. It often presents in various of 96 polluted gases in the cases of virtual situations, and would commonly have negative 97 influence on VOCs adsorption on adsorbents due to its strong competitive adsorption. 98 The study on VOCs adsorption on MIL-101 in the presence of water vapor had hardly 99 been done so far. Competitive adsorption behaviour and mechanism of water vapor 100 and VOCs on MIL-101(Cr) in humid atmosphere has not been revealed yet. Therefore, 101 it is necessary to investigate the effect of water vapor on VOCs adsorption on 102 MIL-101 and its competitive adsorption mechanism so that some new knowledge can 103 be obtained and adsorption process of MIL-101 for practical application of VOCs 104 removal can be optimized or designed reasonably.

105 The purpose of this work is to investigate competitive adsorption behavior of water 106 vapor and three VOCs EA, DCE and benzene on MIL-101 with the help of fixed bed 107 experiments. Adsorption isotherms of EA, DCE, benzene and water vapor were 108 separately measured by using a gravimetric method. The breakthrough curves of EA, 109 DCE and benzene were measured under the conditions of the absence and the 110 presence of water vapor, and then compared. TPD experiments were conducted to 111 estimate the interaction of water vapor and VOCs EA, DCE and benzene with the 112 surfaces of MIL-101. The competitive adsorption mechanism of water vapor with EA, 113 DCE and benzene on MIL-101 would be discussed and then reported here.

114 **2. Experimental section**

115 2.1. Materials. Chromium(III) nitrate nonahydrate ($Cr(NO_3)_3 \cdot 9H_2O_2$, > 99.0%, **116** Alfa), 1,4-benzene dicarboxylic acid ($C_8H_6O_4$, > 99.0%, Aldrich), hydrofluoric acid **117** (HF, 48.0%, Merck), *N*,*N*-dimethylformamide (DMF, 99.5%, Mallinckrodt), ethanol

| 118 | (99.7%, Tianjin), NH ₄ | $F (\geq 96.0\%, Tianjin)$, and 1,2-dichloroethane | (≥99.0%,Tianj | in) |
|-----|-----------------------------------|---|---------------|-----|
| | | | | |

119 All of above were used as received from vendors without further purification.

2.2. Synthesis of MIL-101 and Characterization. The MIL-101 used in this
work was synthesized by hydrothermal method and the specific are described in
supporting information (S1).

123 The surface morphology and particle size of MIL-101 samples were observed by 124 using a LEO 1530Vp scanning electron microscope (SEM) at an accelerating voltage 125 of 5.0 kV after gold deposition. The synthesized MIL-101 was characterized by X-ray 126 powder diffraction (XRD), which was performed on Bruker D8 Advance X-ray 127 diffractometer at 40 kV, 40 mA, with a scan speed of 2°/min and a step size of 0.02° in 128 $2 \sim 25$, using λ_{Cu} K α radiation. Specific surface area and pore texture of the sample 129 were measured using Micromeritics ASAP 2020.

130 2.3. Measurement of Adsorption Isotherms of DCE, EA and benzene Vapor. 131 Adsorption isotherms of DCE, EA and benzene vapor on MIL-101 sample were 132 measured by using a standard gravimetric technique (intelligent gravimetric analyzer, 133 IGA-003, Hiden) at 298-318 K. This intelligent gravimetric analyzer (IGA-003, 134 Hiden) is equipped with an ultra-sensitive balance of resolution 0.2 μ g. Details of the 135 measurement procedures can be available in the supporting information (S2). The 136 equilibrium and instantaneous uptakes of DCE, EA and benzene on the sample can be 137 calculated as follows:

$$Q_{e} = \frac{1000(W_{e} - W_{a})}{W_{a}M_{DCE}}$$
(1)

$$Q_{t} = \frac{1000(W_{t} - W_{a})}{W_{a}M_{DCE}}$$
(2)

138 where M_{DCE} (g/mol) is the molecular weight of VOC molecule; W_e (g) and W_t (g) 139 are the amount of adsorbent (the MIL-101) at equilibrium and time t (s); W_a (g) is 140 the initial weight of the sample (MIL-101); and Q_e (mmol/g) and Q_t (mmol/g) are 141 the VOC amount adsorbed per gram of adsorbent at equilibrium and at time t (s), 142 respectively(S2).

143 2.4. Measurement of Adsorption Isotherms of Water Vapor. Isotherm of water
144 was measured on Gravimetric water sorption analyzer (AQVADYNE DVS) equipped
145 with a microbalance with an accuracy of 1µg. Details of the measurement procedures
146 can be available in the supporting information (S2).

147 2.5. Temperature Programmed Desorption Experiments. Temperature 148 programmed desorption (TPD) is an effective technique of surface analysis¹¹. In this 149 work, TPD experiments were conducted to estimate the binding energy between an 150 adsorbate and an adsorbent. The detailed description of methods for TPD experiment 151 and estimation of desorption activation energy are given in supporting information 152 (S3).

153 2.6. Determination of Breakthrough Curves under Different Relative
154 Humidities. Fixed bed adsorption experiments were conducted at 308 K to measure
155 the breakthrough curves of DCE, EA and benzene on MIL-101 under the conditions
156 of relative humidities of 5%, 40%, and 80%, separately. The experimental setup is a
157 flow-typed fixed-bed adsorption system, which is shown in Supporting Information
158 (Fig. S1) and the detailed description of the method is given in Supporting
159 Information (S4).

161 **3. Result and Discussion**



Fig. 1. Adsorption- desorption isotherms of N₂ on MIL-101 at 77 K.

162 **3.1. Sample Characteristics.** Fig. 1 shows the nitrogen adsorption and desorption 163 isotherms of MIL-101 sample. It can be seen that the N₂ isotherm on MIL-101 was of 164 typical type-V profile with secondary uptakes, which is characteristic of the presence of the two kinds of microporous¹⁰. Textural properties of sample can be obtained from 165 166 the isotherm by analyzing the nitrogen adsorption and desorption isotherms with the 167 Micromeritics ASAP 2010 built in software. The BET specific surface area and total pore volume of MIL-101 synthesized in this work were 3360 m²/g and 1.75 cm³/g, 168 169 respectively. In addition, the SEM image and XRD pattern of the sample are also 170 obtained, which can be seen in Supporting Information (S5, Fig. S2 and S3).

3.2. Isotherms of Three VOCs on MIL-101. Fig. 2 presents the isotherms of DCE,
EA, and benzene on MIL-101 at 308 K. It shows that the isotherms of DCE and EA
were favorable ones, while the isotherm of benzene was an unfavorable isotherm. The
isotherms of DCE and EA were higher than that of benzene, suggesting that the
adsorption capacity of MIL-101 for DCE or EA was higher than that for benzene. The
amounts adsorbed of DCE, EA and benzene on MIL-101 followed the order: DCE >

177 EA > benzene. For example, at 8 mbar, the adsorption capacity of MIL-101 for DCE 178 was up to 9.7 mmol/g, that for DCE was about 5.5 mmol/g, and that for DCE was 179 about 3.3 mmol/g. This difference in adsorption capacity may be ascribed to the 180 different polarity of the three VOCs molecule. DCE and EA are polar molecules 181 whose dipole moments are 1.8D and 1.78D respectively. Benzene is non-polar 182 molecule with dipole moments of zero. The different polarities of the three VOCs 183 molecule result in the different interactions between VOC molecules and the 184 framework of MIL-101 sample. As a result, the isotherms of DCE and EA were 185 favorable ones due to their strong adsorption, while the isotherm of benzene was an 186 unfavorable isotherm due to its weaker adsorption compared to DCE and EA. In 187 addition, Table 1 gives the comparison between adsorption capacities of MIL-101 and 188 some other porous materials for DCE, EA and benzene at similar conditions. It 189 indicated that the adsorption capacities of MIL-101 were much higher than those of 190 the other materials.

Langmuir and Freundlich equations were applied to fit the experimental isotherm data in order to describe adsorption behavior of DCE, EA and benzene on MIL-101 sample, as shown in Fig. 2. The Langmuir equation seems to give a good fit to the experimental isotherm data of DCE and EA. However, the Langmuir equation was not fit for adsorption data of benzene, but the Freundlich equations can give a good fit.



Fig. 2. Adsorption isotherms of DCE, EA and Benzene on the MIL-101 at low pressure and 308 K (points, experimental data; solid curves, fitted isotherm with Langmuir/Freundlich equation).

197 Table 1. The Adsorption Capacities of MIL-101 and some other Materials for DCE, EA and

198 Benzene VOC T (K) Material Q (mmol/g) Pressure(mbar) Reference 9.7 308 8 MIL-101 Present work 17 DCE LC-1 3 303 8 18 **HC-MWCNTs** 8 0.05 323 MIL-101 5.5 308 8 Present work 19 EA 1 293 8 MCM-48 20 HDTMA clay 0.4 308 8 MIL-101 8 3.3 308 Present work 21 0.8 8 MOF-5 303 22 1 8 benzene silicalite-1 295 23 H-ZSM-5 8 1.3 303 23 303 8 SBA-15 0.83

199

200 The fitting parameters of Freundlich isotherm and Langmuir isotherm equations as 201 well as their linear correlation coefficients (R^2) were listed in Supporting Information

202 (Table S1). Examination of the data shows that the Langmuir and Freundlich 203 equations were able to fit the experimental adsorption data well since their correlation 204 coefficients R^2 were up to 0.99.

205

206 3.3. Effects of Water Vapor on Breakthrough Curves of VOCs through the 207 Fixed Bed of MIL-101. Figs. 3-5 show the adsorption breakthrough curves of DCE, 208 EA, and benzene through the fixed bed packed with MIL-101 at different relative 209 humidity, separately. It was clearly visible that the relative humidity of feed stream 210 containing VOCs had negative influence on the breakthrough curves of these VOCs in 211 the packed bed. The breakthrough times of these VOCs sharply decreased with an 212 increase in the humidity of the feed stream, implying that the working adsorption 213 capacity of MIL-101 sharply decreased due to presence of water vapor in the feed 214 stream. The higher relative humidity would make the breakthrough times of the VOCs 215 become shorter.



Fig. 3. Effect of relative humidity on the breakthrough curves of EA through the fixed bed of the MIL-101 (*T*=308 K, N₂ flow rate= 70 ml/min).

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Fig. 4. The effect of relative humidity on the breakthrough curves of DCE through the fixed bed of the MIL-101(*T*=308 K, N₂ flow rate= 70 ml/min).

218



Fig. 5. Effect of relative humidity on the breakthrough curves of Benzene through the fixed bed of the MIL-101 (*T*=308 K, N₂ flow rate= 70 ml/min).

219

220 Table 2. Effects of Relative Humidity on Breakthrough Times and Working

221 Capacities of DCE, EA and Benzene Adsorption on MIL-101 at 308 K.

| Adsorbate | C ₀ | RH (%) | Breakthrough | Working capacity |
|-----------|----------------|--------|--------------|------------------|
| | (mol/L) | | time (min) | (mmol/g) |

| Page | 14 | of | 24 |
|------|----|----|----|
|------|----|----|----|

| DCE | 5×10 ⁻⁵ | 5 | 16.3 | 2.282 |
|-----|--------------------|----|------|-------|
| | | 40 | 10.7 | 1.498 |
| | | 80 | 8.9 | 1.246 |
| EA | 2×10 ⁻⁵ | 5 | 27.9 | 1.562 |
| | | 40 | 12.8 | 0.717 |
| | | 80 | 10.6 | 0.594 |
| BE | 10 ⁻⁵ | 5 | 11.4 | 0.319 |
| | | 40 | 6.24 | 0.175 |
| | | 80 | 4.46 | 0.125 |
| | | | | |

222 Table 2 lists the breakthrough times and working adsorption capacities of EA, 223 DCE and Benzene on MIL-101 under the conditions of different relative humidities. It 224 indicated that the breakthrough times and the working capacities of these VOCs 225 became smaller with an increase in the relative humidity. When the relative humidity 226 of the feed stream increased from 0% to 80%, the VOCs breakthrough times or the 227 working adsorption capacities were decreased by about 45, 62 and 60.8% separately 228 for DCE, EA, and benzene. It could be attributed to the competition adsorption of 229 water vapor on the surfaces of MIL-101. MIL-101 is amphiphilic porous solids since 230 its framework is composed of inorganic (metal cations and oxygen anions) and organic moieties (nonpolar carbon atoms and benzene ring)^{23, 24}. Thus it has polar 231 232 sites due to the metal oxygen clusters and very non-polar regions due to the presence of organic and mostly aromatic linker²³. In dry condition, the unsaturated Cr^{3+} sites on 233 MIL-101 were also the strong adsorption site for VOCs¹¹. In the case of the presence 234 235 of water vapor in adsorption system, H₂O molecules preferably adsorbed on the hydrophilic centers such as trivalent metal cations²⁵, and then if the concentration of 236

| 237 | water vapor became higher further additional water molecules are bound by hydrogen |
|-----|--|
| 238 | bridges to these water nucleation sites resulting in small water clusters ^{23, 26} . As a result, |
| 239 | the part surface area of MIL-101 would be occupied by more $\mathrm{H_2O}$ molecules, and |
| 240 | thus the less adsorptive sites are available for adsorption VOCs. In fixed bed |
| 241 | adsorption experiment, when the gaseous mixture containing VOCs and water vapor |
| 242 | passed the fixed bed of MIL-101, some water molecules preferentially adsorbed on |
| 243 | the hydrophilic centers such as Cr^{3+} sites of MIL-101, and thus the surface active sites |
| 244 | of MIL-101 for adsorption VOCs became less. As a consequence of that, the working |
| 245 | adsorption capacity of the fixed bed of MIL-101 for VOCs would greatly decrease |
| 246 | due to competitive adsorption of water molecules. |

247 Figs. 6a-6c show a comparison of breakthrough curves of benzene at different 248 temperatures in the presence and absence of water vapor. It was observed from Fig. 6 249 that breakthrough times of benzene in the presence and absence of water vapor 250 decreased obviously as temperature rose. It can be attributed to a decrease in 251 adsorption capacity of MIL-101 for benzene due to rising of temperature, as shown in 252 Fig.S6. In addition, it was noticed that the breakthrough times in the presence of water 253 vapor was always shorter than in the absence of water vapor at different temperatures. 254 It meant that although the rise of temperature could weaken adsorption of water vapor 255 on MIL-101, there was still the negative effect of water vapor on benzene adsorption 256 on MIL-101 in the range of temperatures studied.



Fig. 6a. The breakthrough curves of benzene on MIL-101 at 308 K.



Fig. 6b. The breakthrough curves of benzene on MIL-101 at 318 K.



Fig. 6c. The breakthrough curves of benzene on MIL-101 at 328 K.

3.4. Isotherm of Water Vapor on MIL-101. Fig. 7 shows the adsorption isotherm
of water vapor on MIL-101 at 308K. It exhibited an S-shaped type of the isotherm
with maximum water vapor uptake of 83 mmol/g. The amount adsorbed of H₂O vapor
was low at the relative humidity below 30%, and then sharply increased when the
relative humidity reaching 40 RH%. As the relative humidity increased, the amount
adsorbed of water vapor water continued to rise.



Fig. 7. The adsorption isotherm of water vapor on the MIL-101.

268

3.5. Desorption Activation Energies of the VOCs and H₂O vapor on MIL-101.

Fig. 8 shows a series of TPD curves of EA, DCE and benzene on MIL-101 at different heating rates. Each of these TPD curves exhibited one peak due to EA, DCE or benzene desorption, and its peak temperature (T_p) increased with an increase in heating rate.

274



Fig. 8a. TPD spectrums of EA on MIL-101 at different heating rate from 4-12 K/min.



Fig. 8b. TPD spectrums of DCE on MIL-101 at different heating rate from 4-12 K/min.



Fig. 8c. TPD spectrums of benzene on MIL-101 at different heating rates from 4-12 K/min.

278 Knowing the values of T_p at the different heating rates employed (in Table 3),
279 desorption activation energy of EA, DCE or benzene can be estimated by using

Polanyi-Wigner equation (See Supporting Information S2 and Fig.S4). Calculation
indicated that desorption activation energies of the three VOCs on MIL-101 were
47.14, 41.9, and 38.16 kJ/mol, respectively, which followed the order: DCE > EA >
benzene.



Fig. 9. TPD spectrums of H_2O vapor on the MIL-101 at different heating rates from 4-12 K/min.

| 285 | Table 3 Desorption activation energies of EA, DCE, Benzene, and H_2O on the |
|-----|---|
| 286 | MIL-101. |
| | |

| | | | T _p (K) | | | | |
|------------------|------------|------------|--------------------|-------------|-------------|--------------------------------|----------------|
| VOCs | 4 K/min | 6 K/min | 8 K/min | 10 K/min | 12 K/min | - E _d (kJ/mol) | \mathbb{R}^2 |
| EA | 357 | 365 | 373.5 | 378 | 382.5 | 41.9 | 0.997 |
| DCE | 361 | 368 | 374.5 | 379 | 385 | 47.14 | 0.995 |
| Benzene | 355 | 363 | 369.5 | 377.5 | 382.5 | 38.16 | 0.993 |
| H ₂ O | 439 | 447 | 453 | 458 | 462 | 72.96 | 0.999 |

288 In similar manner, TPD experiments were also conducted for water vapor at 289 different heating rates. Fig. 9 exhibits the TPD spectrums of water vapor desorption from MIL-101 at different heating rates (Table 3). It was clearly visible that only one 290 peak appeared in each of their TPD curves, and the peak temperature (T_p) increased 291 292 with an increase in heating rate. In similar method mentioned above, the desorption 293 activation energy of water vapor on MIL-101 was estimated to be 72.9 kJ/mol (See 294 Supporting Infomation Fig. S5), which was higher than that of the three VOCs. The 295 higher desorption activation energy meant a stronger interaction between H_2O 296 molecules and the surfaces of MIL-101 compared to those three VOCs. This would 297 make the presence of water vapor in the feed stream result in a sharp decrease in the 298 VOCs working adsorption capacities of MIL-101 due to strong competitive 299 adsorption of water vapor on the surfaces of MIL-101.

300 4. Conclusion

301 Adsorption behavior of 1,2-dichloroethane(DCE), ethyl acetate(EA), and Benzene 302 in the fixed bed of MIL-101 was investigated in the presence of water vapor. The 303 results showed that the equilibrium amounts adsorbed of DCE, EA and benzene on 304 MIL-101 followed the order: DCE > EA > benzene. The maximum adsorption 305 capacities of MIL-101 for DCE, EA and benzene were separately up to 9.71, 5.79 and 306 3.76 mmol/g, being much higher than those of some other adsorbents such as zeolites, 307 activated carbons and clay. The presence of water vapor in the feed stream had 308 remarkable negative influence on the breakthrough behavior of VOCs in the fixed bed. 309 The breakthrough times and the working capacities of these VOCs became smaller 310 with an increase in the relative humidity. Isotherm of water vapor on MIL-101 at 308 311 K exhibited an S-shaped type of the isotherm, which can be expressed by DSLF

312 model. TPD experiments showed that the desorption activation energies of water 313 vapor, DCE, EA and benzene on MIL-101 followed the order: water vapor > DCE >314 EA > benzene, which were 72.9, 47.14, 41.9, and 38.16 kJ/mol, respectively. The 315 higher desorption activation energy of water vapor suggested the interaction between 316 water molecule and MIL-101 was stronger than that between DCE, EA, or benzene 317 and MIL-101. The stronger interaction of water vapor with MIL-101 formed stronger 318 competitive adsorption with VOCs on MIL-101. As a result, the existence of water 319 vapor in the feed stream could led to a sharp decrease in the working adsorption 320 capacity of MIL-101 for these VOCs due to its strong competitive adsorption on the 321 surfaces of MIL-101. Getting well-understanding of water vapor impact on the VOCs 322 adsorption on the MOFs under humid atmosphere would be helpful to design 323 reasonably adsorption process or study novel hydrophobic materials. Thus, an 324 interesting topic or a challenge topic will appear. Therefore, the preparation of 325 hydrophobic MOFs by surface modification for enhancing their adsorption of VOCs 326 in the practical application would be worthy of investigation in the future.

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