

Photochemical & Photobiological Sciences

Accepted Manuscript



This is an *Accepted Manuscript*, which has been through the Royal Society of Chemistry peer review process and has been accepted for publication.

Accepted Manuscripts are published online shortly after acceptance, before technical editing, formatting and proof reading. Using this free service, authors can make their results available to the community, in citable form, before we publish the edited article. We will replace this *Accepted Manuscript* with the edited and formatted *Advance Article* as soon as it is available.

You can find more information about *Accepted Manuscripts* in the [Information for Authors](#).

Please note that technical editing may introduce minor changes to the text and/or graphics, which may alter content. The journal's standard [Terms & Conditions](#) and the [Ethical guidelines](#) still apply. In no event shall the Royal Society of Chemistry be held responsible for any errors or omissions in this *Accepted Manuscript* or any consequences arising from the use of any information it contains.

1

2 **Catechin as a new improving agent for photo-Fenton-like system at near-neutral**
3 **pH for the removal of inderal**

4 Zongping Wang^a, Yizhou Guo^a, Zizheng Liu^{*ab}, Xiaonan Feng^a, Yiqun Chen^a, Tao
5 Tao^a

6 ^a*School of Environmental Science and Engineering, Huazhong University of Science*
7 *and Technology, Wuhan, 430074, China*

8 ^b*School of Civil Engineering, Wuhan University, Wuhan, 430074, China*

9

10

11 **Abstract**

12 Photodegradation of inderal in a photo-Fenton like system at near-neutral pH
13 modified with catechin, a natural catecholate siderophore, was investigated under
14 simulated sunlight. Main factors influencing the process, such as Fe(III)-catechin
15 complexes, pH, and catechin concentration were examined. Photodegradation of
16 inderal was strongly dependent on pH, following the order pH 3.0 < 5.0 < 7.0 < 6.0.
17 Formation of the Fe(III)-catechin complex resulted in the stabilization of ferric iron
18 and generation of [•]OH under irradiation at near-neutral pH values. The removal of
19 inderal was about 75% under optimal conditions (50 μmol/L Fe(III) and 200 μmol/L
20 catechin at pH 6.0). Photodegradation products of inderal were identified by
21 LC-ESI-MS and GC-MS, and the photodegradation pathway was proposed. Iron in

* Corresponding author. Tel.: +86-27-87792406; Fax: +86-27-87792101.
E-mail address: lzz@hust.edu.cn (Z.Z. Liu).

22 the Fe(III)–catechin system was reused by simple addition of catechin to the reaction
23 mixture. Results of the present work suggest that the Fe(III)–catechin complex may
24 be used in the photo-Fenton like process, an advanced approach to the removal of
25 organic pollutants at near-neutral pH.

26 *Keywords:* Fe(III)–catechin complex; Catecholate siderophore; Photo-Fenton; Neutral
27 pH; Inderal

28

29 **1. Introduction**

30 Processes based on hydroxyl radical ($\bullet\text{OH}$) chemistry are currently used for the
31 degradation of organic contaminants. Much is known about the nature of the Fenton
32 reaction, which has been widely used as a source of $\bullet\text{OH}$, since its first description in
33 a paper published in 1894¹. In recent decades, studies have been done to examine the
34 application of various types of Fenton reactions in wastewater treatment. These
35 reactions include homogeneous Fenton (and Fenton like) processes and photo-Fenton
36 (and photo-Fenton like) systems using various Fe(II) and/or Fe(III) salts and peroxide
37 in acidic media. One system that has been found to be most effective for removal of
38 various organic pollutants from aqueous media, soils, and sediments is the
39 photo-Fenton reaction²⁻⁶, which generates mainly the highly reactive radical $\bullet\text{OH}$ ⁷⁻⁹.

40 However, the traditional Fenton like system has several drawbacks, in particular, its
41 limited pH range and ferric iron precipitation during the process. To overcome these
42 deficiencies, different approaches have been used in related studies. In this regard, one
43 approach that holds promise is the use of organic and inorganic ligands for iron.

44 Ligands that have been frequently used in photo-Fenton like processes for removal of
45 organic substances in aqueous solutions are citrate ¹⁰ and oxalate ¹¹. Compared with
46 traditional Fenton systems, photo-Fenton-like processes could improve the removal
47 efficiency of substrates by using the aforementioned organic ligands at near-neutral
48 pH. The photo-Fenton like process strongly absorbs ultraviolet-visible (UV-vis)
49 radiation and stabilizes ferric ions over a wider pH range. The high quantum yield of
50 Fe(II) in the presence of ligands increases the amount of •OH generated, thereby
51 increasing the removal efficiency of organic pollutants ^{12,13}. Other ligands have also
52 been used to complex iron in photo-Fenton like systems. These include EDTA ¹⁴,
53 humic acids ^{15,16}, and EDDS ¹⁷. Furthermore, Solar photo Fenton was developed to
54 reduce the process time and energy consumption. ^{18,19}

55 Complexes of iron with strong organic ligands comprise >99.9% of dissolved iron
56 in natural environments ²⁰. This abundance has recently generated much interest in
57 siderophores, iron chelators in biological systems and natural waters as well. Various
58 fungi and heterotrophic bacteria cultured under conditions of limited iron levels have
59 been found to produce siderophores that form complexes with high stability constants
60 ²¹⁻²⁴. The photoreactivity of siderophores is primarily determined by the chemical
61 structure of their Fe(III)-binding groups, namely, hydroxamate and catecholate
62 moieties. Hydroxamate groups are photochemically resistant regardless of Fe(III)
63 complexation. Catecholate siderophores are susceptible to photooxidation and
64 undergo light-induced ligand oxidation and reduction of Fe(III) to Fe(II) ²⁵.

65 Most of the studies on siderophores were centered on identifying structures ²⁶,

66 determining dissociation constants for Fe(III)–siderophore complexes^{27, 28}, and
67 examining iron uptake in the ocean²⁹. However, few studies addressed the use of
68 photo-Fenton like systems with Fe(III)–siderophore complexes for remediation of
69 freshwater environments. In the present work, catechin (CAT), which is a natural
70 catecholate siderophores and can be easily obtained from green plants especially
71 Chinese tea, was used as a model catecholate siderophore. We examined the effect of
72 Fe(III)–siderophore complexes on photodegradation of pollutants dissolved in neutral
73 aqueous media. For this purpose, we used inderal, a popular amine drug, as a model
74 pollutant. The effects of Fe(III) complexes, pH values, and catechin concentration on
75 the degradation of inderal in the Fe(III)–catechin system were examined. Generations
76 of $\cdot\text{OH}$ and Fe(II) were detected for the photochemical mechanism investigation.
77 Photoproducts of the inderal were determined by LC-ESI-MS and GC-MS,
78 respectively, and the photodegradation pathway was proposed accordingly.

79

80 **2. Materials and methods**

81 *2.1. Materials*

82 Inderal (CAS: 318-98-9, 99%), catechin (CAS: 225937-10-0, 98%),
83 nordihydroguaiaretic acid, pyrocatechol violet, 2,3-dihydroxybenzoic acid, and
84 2-chloro-3',4'-dihydroxyacetophenone were purchased from Sigma. Inderal was
85 confirmed from its UV–vis spectra; its molecular structure is shown in [Figure A.1](#). A
86 stock solution of inderal in deionized water was prepared. Ferric chloride (99%),
87 hydrochloric acid, sodium hydroxide, and benzene were supplied by Wuhan

88 Chemicals Corporation. All chemicals were of analytical-reagent grade or of higher
89 purity. Doubly distilled deionized water (Milli-Q) was used to prepare all solutions.

90

91 *2.2. Photolysis experiments*

92 Photolysis experiments were performed in a 60 mL capped cylindrical Pyrex vessel
93 (40 mm i.d., containing 50 mL solution) under a 150 W xenon short-arc lamp.
94 Radiation of wavelengths less than 300 nm was filtered out by using Pyrex glass to
95 simulate sunlight. To measure the quantum yields, the monochromatic irradiations at
96 300 and 365 nm were carried out with the xenon short-arc lamp equipped with a LOT
97 Oriel grating monochromator. The light and the cells were kept parallel with a fixed
98 distance of 7.0 cm. Lamp output was monitored over time by ferrioxalate
99 actinometry³⁰ ($I_{o(300\text{ nm})} \approx 0.63 \times 10^{14} \text{ photon cm}^{-2} \text{ s}^{-1}$; $I_{o(365\text{ nm})} \approx 1.9 \times 10^{15}$
100 $\text{photon cm}^{-2} \text{ s}^{-1}$). All solutions for the photolysis experiments were freshly prepared
101 prior to irradiation. The pH was measured by using a pHs-3C meter, and adjusted as
102 needed. Aliquots of samples were withdrawn at various intervals, and substrate decay
103 was measured by high-performance liquid chromatography (HPLC).

104 *2.3. Analysis*

105 The concentration of inderal was analyzed on a Shimadzu Essentia LC-15C HPLC
106 system with an Agilent HC-C18 column (5 μm , 250 mm \times 4.6 mm). The mobile phase
107 was a solution of acetonitrile and KH_2PO_4 buffer (pH 2.5, 10 mmol/L) at a ratio of
108 35:65 (v/v). The detector wavelength was 213 nm and the flow rate was 1.0 mL/min.
109 Benzene was used to scavenge $\bullet\text{OH}$ in a reaction that produces phenol. The phenol

110 concentration was analyzed by using a mobile phase of 40:60 (v/v)acetonitrile–H₂O
111 solution, and the detector wavelength was set at 270 nm. The concentrations of
112 ferrous ions were determined using 1,10-phenanthroline spectrophotometry at a
113 wavelength of 510 nm. The concentration of Fe³⁺ was calculated by subtracting the
114 concentration of Fe²⁺ from the total iron concentration in solution. NO₃⁻ was
115 determined by ion chromatography. The calibration curve for the detection of
116 inderal, •OH, and Fe(III) is shown in [Figure A.2](#) (Supplementary Data).

117 *2.4. Photoproduct identification*

118 Photoproducts of inderal were obtained after about 80% of the parent compound
119 had been degraded by Fe(III)–catechin complexes at pH 6.0. The reaction solution
120 was evaporated to dryness in a rotary evaporator. The residue was trimethylsilylated
121 at room temperature with 0.1 mL of hexamethyldisilazane and 0.05 mL of
122 chlorotrimethylsilane. GC–MS was carried out on an Agilent 7890A/5975C system
123 with an HP-5 MS capillary column. The gas chromatograph was operated in
124 temperature-programming mode as follows: hold at initial temperature of 80 °C for 20
125 min, ramp at 5 °C/min to 280 °C and hold at this temperature for 3 min. The injection
126 port and MSD were held at 250 and 280 °C, respectively.

127 The LC-MS system used in the study was an Agilent LC/MSD SL ion trap mass
128 spectrometer equipped with an ESI source in the positive ion mode. The mobile phase
129 was (A) 98% H₂O + 2% CH₃CN + 0.2% formic acid and (B) CH₃CN + 0.2% formic
130 acid. Gradient elution was 2% of B for 1 min followed by a linear increase to 95% B
131 at 25 min, and then held constant for an additional 5 min. The mass spectral data were

132 obtained in the positive ion mode between m/z 100-300. The settings were: capillary
133 voltage, 3500 V; drying gas, 10 L/min; drying gas temperature, 350°C; capillary exit,
134 150 V; skimmer, 40 V; octopole RF amplitude, 160Vpp; ICC target, 100 000; trap
135 drive, 60; maximum accumulation time, 100 ms.

136

137 **3. Results and discussion**

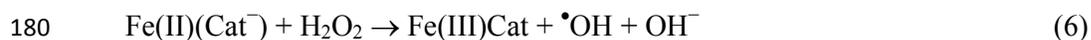
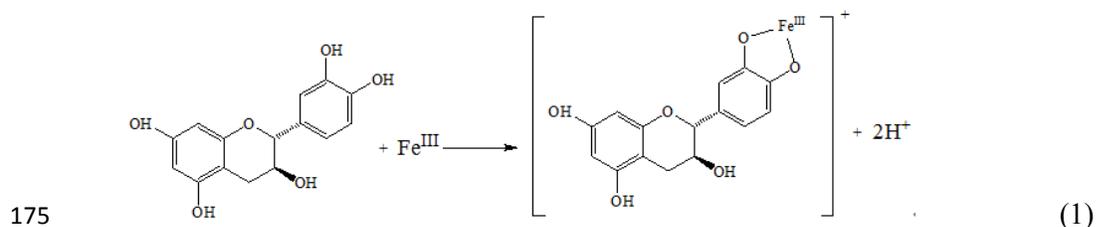
138 *3.1. Inderal photodegradation by five Fe(III) complexes*

139 Five kinds of organic ligands were used to complex with Fe(III) at near-neutral pH
140 (6.0) to modify the photo-Fenton like system. The ligands were catechin,
141 nordihydroguaiaretic acid, pyrocatechol violet, 2,3-dihydroxybenzoic acid, and
142 2-chloro-3',4'-dihydroxyacetophenone (molecular structures are shown in [Table A.1](#)).
143 Inderal degradation through the photo-Fenton like system with various iron
144 complexes was observed [Figure 1](#). The highest efficiency of inderal degradation at pH
145 6.0 was observed in the photo-Fenton like system with Fe(III)-catechin. About 75%
146 of inderal was removed after 120 min. Degradation of inderal in the presence of
147 Fe(III)-2,3-dihydroxybenzoic acid and Fe(III)-2-chloro-3',4'-dihydroxyacetophenone
148 complexes in this photo-Fenton like system was negligible (Supplementary Data for
149 nordihydroguaiaretic acid and pyrocatechol violet are shown in [Figure A.3](#)). This
150 result may be attributed to the stable complexes that catechin forms with Fe(III) at pH
151 6.0. By forming such complexes, catechin enhances the solubility and stabilizes Fe(III)
152 in aqueous solution at near-neutral pH ([Figure A.4](#)). Thus, catechin has an important
153 contribution to preventing the precipitation of Fe(III) at pH 6.0, and the

154 Fe(III)–catechin chelate broadens the pH range for iron solubility. These results
155 suggest that catechin may be an effective ligand in the photo-Fenton like system at
156 near-neutral pH.

157 3.2. Role of CAT and formation of photo-Fenton process

158 To understand the effect of CAT and irradiation on inderal degradation, experiments
159 using Fe(III)/light (with or without CAT) under the same conditions (50 $\mu\text{mol/L}$ Fe(III)
160 and 200 $\mu\text{mol/L}$ catechin at pH 6.0) were conducted. Results of these experiments
161 (Figure 2) show that direct photodegradation of inderal at near-neutral pH was
162 negligible. When Fe(III) and catechin were used simultaneously without irradiation
163 for 120 min, inderal degradation was negligible. Degradation of inderal in the
164 photo-Fenton like system was faster; $\sim 75\%$ inderal was removed after 120 min.
165 Therefore, irradiation markedly enhanced inderal degradation. This enhancement may
166 be attributed to the photoreactivity of the Fe(III)–catechin complex. The formation of
167 Fe(III)–catechin complex was confirmed by UV–vis spectroscopy of Fe(III), catechin,
168 and Fe(III)–catechin mixture (Figure A.4a). The Fe(III)–catechin mixture produced an
169 obvious spectral shift relative to catechin. There was evidence of formation of a 1:1
170 ligand–metal complex at near-neutral pH. The conditional stability constant for Fe(III)
171 binding of the Fe(III)–catechin complex was found to be $(1.6 \pm 0.1) \times 10^6$ by
172 continuous variation methods (Figure A.5). Stabilization of Fe(III) at near neutral pH
173 by CAT and formation of photo-Fenton like system are illustrated in Scheme 1. The
174 photo-Fenton like process could be explained by the following reactions^{3, 31, 32}:



181 Formation of Fe(III)–catechin complexes evidently increased the rate of $\text{HO}_2^{\cdot}/\text{O}_2^{\cdot-}$
 182 formation and of $\cdot\text{OH}$ formation³³. These radicals were detected in the reaction
 183 solutions (Figure A.6a). During rapid ligand-to-metal charge transfer (LMCT) of the
 184 Fe(III)–catechin complex, the species Fe^{2+} and $\text{Cat}^{\cdot-}$ are generated. In the presence of
 185 oxygen, $\text{O}_2^{\cdot-}$ forms and subsequently generates H_2O_2 which, in turn, produces $\cdot\text{OH}$
 186 through a reaction with Fe(II) in the Fenton process (Eqs. (3)–(6)). Under acidic
 187 conditions, Fe(III) and catechin are free species (Figure A.4b), which have weaker
 188 ability to produce $\cdot\text{OH}$ compared with the Fe(III)–catechin complex. At near-neutral
 189 pH, catechin may increase the solubility and stability of iron. Thus, the
 190 Fe(III)–catechin complex becomes a predominant species and has considerable
 191 photoreactivity at this pH. On the other hand, the amount of catechin for Fe(III)
 192 complexation was more than enough at increased catechin concentration. Thus, it
 193 outcompeted hydroxide ions for Fe(III) at higher pH. Therefore, amount of $\cdot\text{OH}$
 194 formation increased in the order $\text{pH } 3.0 < 5.0 < 7.0 < 6.0$.

195 Inderal photodegradation is mainly attributed to the reactive oxygen species, $\bullet\text{OH}$.
196 The presence of light played an important role in the production of $\bullet\text{OH}$. In the
197 absence of light, $\bullet\text{OH}$ formation could almost be neglected, in accordance with the
198 comparison between Fe(III)-CAT and photo-Fenton like systems used for inderal
199 degradation. This trend is supported by rate measurements of Fe(III) reduction (Figure
200 A.6b). Results in Figure A.6b show that ~20% of Fe(III) was reduced to Fe(II) at
201 near-neutral pH under irradiation. This result confirms that irradiation affected inderal
202 degradation by enhancing $\bullet\text{OH}$ generation in the system. Furthermore, the use of
203 irradiation approximating sunlight (main wavelength at 365 nm) makes the process
204 potentially economical and practical.

205 The effect of the catechin modified photo-Fenton like process on inderal
206 degradation at pH 6.0 was examined to obtain more insight into how $\bullet\text{OH}$ participates
207 in the process. For this purpose, 2-propanol was used as the scavenger for $\bullet\text{OH}$ in the
208 solution containing. 50 $\mu\text{mol/L}$ Fe(III), 200 $\mu\text{mol/L}$ catechin, and 10 $\mu\text{mol/L}$ inderal.
209 Results in Figure 3 show that addition of 2-propanol markedly inhibited inderal
210 degradation. 2-Propanol and inderal reacted competitively with $\bullet\text{OH}$. Use of 10
211 mmol/L 2-propanol resulted in removal of only 15% of inderal, demonstrating that
212 inderal degradation was mainly due to the reactivity of $\bullet\text{OH}$. When the 2-propanol
213 concentration was increased, the extent of inderal degradation decreased substantially.
214 Thus, these results also indicate that inderal degradation was mainly due to $\bullet\text{OH}$.

215 3.3. Effect of pH

216 To understand better the optimum pH for the photo-Fenton like process modified

217 with catechin, experiments at pH 3.0, 5.0, 6.0, and 7.0 were carried out respectively.
218 Initial concentrations of Fe(III), catechin, and inderal were 50, 200, and 10 $\mu\text{mol/L}$,
219 respectively. The rate of inderal degradation increased in the order $\text{pH } 3.0 < 5.0 < 7.0$
220 < 6.0 (Figure 4), which is different from the trend for the traditional Fe(III)/light
221 process. The photo-Fenton reaction with Fe(III) was ineffective at pH 3.0 because of
222 chemical equilibrium of Eq. (1) in the solution³¹. Data on $\bullet\text{OH}$ formation in solutions
223 at various pH values are consistent with this observation (Figure A.6a in
224 Supplementary Data). Therefore, complete complexation of Fe(III) and catechin was
225 essential to the photo-Fenton like system in this work. Under acidic conditions, the
226 chemical equilibrium shifted to the left, leading to the presence of free Fe(III) and
227 catechin. The efficiency of the photo-Fenton like process with Fe(III)-OH complex is
228 quite low compared with that of the process with the Fe(III)-catechin complex.
229 Therefore, the degradation efficiency was low at pH 3.0. The higher efficiency of
230 inderal degradation by the Fe(III)-catechin complex could be attributed to the
231 complexation of iron by catechin, which leads to greater stabilization of iron and
232 formation of hydroperoxyl/superoxide radicals at pH 6.0. We³⁴ and other
233 investigators³⁵ also reported similar observations on the photo-Fenton like systems
234 modified with natural or biological products, which are more complicated and not
235 easier to get than catechin.

236 *3.4. Effect of catechin concentration*

237 Inderal photodegradation at pH 6.0 in the presence of catechin at various
238 concentrations was determined. Inderal degradation and catechin concentration were

239 highly correlated (Figure 5). Use of 50 $\mu\text{mol/L}$ catechin resulted in a degradation rate
240 of <40% over 120 min, whereas addition of 200 $\mu\text{mol/L}$ catechin increased the rate to
241 ~75%. Increasing the catechin concentration from 200 to 500 $\mu\text{mol/L}$ led to a higher
242 rate of inderal removal (from 75% to 85%). This effect is attributed to the high
243 concentration of Fe(III)–catechin complex, which could enhance the formation of
244 $\text{HO}_2^\bullet/\text{O}_2^{\bullet-}$ and $^\bullet\text{OH}$ and, ultimately, inderal removal. This phenomenon was observed
245 during the degradation of alachlor photoinduced by Fe(III)–citrate complex¹⁰. On the
246 other hand, when the concentration of catechin is 800 $\mu\text{mol/L}$, $^\bullet\text{OH}$ radicals are
247 competitively sequestered by inderal and excess free catechin in aqueous solution;
248 such competition thus decrease the efficiency of inderal degradation.

249 3.5. Photodegradation products and pathway

250 Using LC–ESI-MS and GC–MS, we identified products of inderal
251 photodegradation in the catechin modified photo-Fenton system to examine further
252 the mechanism of inderal photodegradation at near-neutral pH. HPLC chromatograms
253 and (+)-ESI-MS spectra of inderal and its photodegradation products after 120 min of
254 photodegradation are presented in Figure A.7. Two products with retention times of
255 1.4 and 5.0 min were identified. The m/z ratio of the photoproduct at 1.4 min is 134
256 and that of its fragment ion is 92. At 5.0 min, the m/z ratio of the photoproduct was 161.
257 $^\bullet\text{OH}$ attack of side chains ending in their cleavage and oxidation might have led to the
258 formation of this photoproduct.

259 The low-molecular-weight photoproducts were volatile and detectable by GC–MS.
260 Using the NIST98 library database, we unambiguously identified the main

261 intermediates of the Fe(III)–catechin system (Table A.2). Mass spectra of standard
262 compounds and photoproducts could be compared in Figure A.8. Table A.2 also
263 indicates that most of the intermediates were hydroxylation products such as
264 3,4-dihydroxymandelic acid, pyrogallol, protocatechuic acid, *p*-phthalic acid, and
265 gallic acid. •OH may further attack these hydroxylation products, producing a series
266 of low-molecular-weight carboxylic acids. These results thus confirm that inderal
267 photodegradation in the Fe(III)–catechin system mainly involved the reactive oxygen
268 species, •OH. As inderal contains an amine group, tests were performed to determine
269 whether inorganic products were formed. NO₃⁻ and NO₂⁻ were determined by ion
270 chromatography. Only NO₃⁻ was detected in the solution. Hence, the amine group of
271 inderal was transformed to inorganic oxy-nitrogen ions. The proposed pathway for
272 inderal photodegradation according to the above results is illustrated in Figure 6. The
273 side chain and benzene ring of inderal first undergo attack by •OH, leading to
274 1,8-dihydroxynaphthalene and 1,2-dihydroxy-3-isopropylamino-propane. Subsequent
275 cleavage of one of the benzene rings produces benzene derivatives such as
276 3,4-dihydroxymandelic acid, pyrogallol, protocatechuic acid, *p*-phthalic acid, and
277 gallic acid. •OH further reacts with the intermediates and generates a series of
278 low-molecular-weight carboxylic acids.

279 3.6. Reuse of Fe(III)/Fe(II) in solution

280 In our photo-Fenton like system, the solubility and stability of iron in aqueous
281 solution at near-neutral pH is increased by the use of catechin to form stable
282 complexes with Fe(III). By alternating between the +2 and +3 oxidation states, iron

283 catalyzes the reaction in the solution. Fe(III)–catechin complexes subjected to
284 irradiation produce Fe(II) via LMCT. As the amount of iron used contributes to the
285 cost of processing and excess amounts discolor the water, a system for Fe(III)/Fe(II)
286 reuse must be in place. In this regard, simple addition of catechin to solutions treated
287 through the photo-Fenton reaction provides a convenient approach.

288 In our study, three runs were conducted to investigate inderal photodegradation.
289 Initially, the degradation was performed at pH 6.0 with 50 $\mu\text{mol/L}$ Fe(III) and 200
290 $\mu\text{mol/L}$ catechin. After 120 and 240 min respectively, 200 $\mu\text{mol/L}$ catechin was added
291 to the solution. As shown in [Figure 7](#), the rate constants for inderal photodegradation
292 were 2.3×10^{-4} , 9.5×10^{-5} , and $3.3 \times 10^{-5} \text{ s}^{-1}$, respectively. There was a large effect
293 on inderal removal within 120 min in all three runs. The rate of inderal removal
294 slowly decreased in subsequent runs. Accumulation of intermediates produced during
295 the earlier run may explain this observation. Such intermediates may compete
296 with $\bullet\text{OH}$ for reaction with inderal in the solutions.

297

298 **4. Conclusions**

299 We examined the use of the Fe(III)–catechin complex in the photo-Fenton like
300 process by subjecting inderal-contaminated samples to treatment by the process. The
301 degree and rate of inderal degradation at near-neutral pH was markedly increased
302 upon use of the Fe(III)–catechin complex. By preventing Fe(III) precipitation,
303 catechin contributed to stabilization of Fe(III) at near-neutral pH. Increased efficiency
304 of inderal degradation with increasing pH ($3.0 < 5.0 < 7.0 < 6.0$) provides evidence

305 that controlling the pH of the reaction was essential to the efficiency of the process.
306 The efficiency of inderal degradation increased with catechin concentration. The
307 reactive oxygen species $\bullet\text{OH}$ was mainly involved in inderal photodegradation in
308 solution. As reuse of iron in the Fe(III)–catechin solution simply involved adding
309 catechin to the reaction solution, recovery of Fe was convenient. Attack of $\bullet\text{OH}$ on
310 inderal led to 1,8-dihydroxynaphalene, 1,2-dihydroxy-3-isopropylamino-propane,
311 3,4-dihydroxymandelic acid, pyrogallol, protocatechuic acid, *p*-phthalic acid, gallic
312 acid, and low-molecular-weight carboxylic acids.

313

314 **Acknowledgments**

315 This work was supported by a grant from the National Natural Science Foundation
316 of China (51078161, 51409108) and Huazhong University of Science and Technology
317 Independent Innovation Fund (0118261028). Thanks for Huazhong University of
318 Science and Technology Analytical and Testing Center. The anonymous reviewers are
319 also gratefully acknowledged.

320

321 **Appendix A. Supplementary information**

322 Supplementary information associated with this article can be found in the online
323 version.

324

325

326

327

328

329 **References**

- 330 1. H. J. H. Fenton, LXXIII.—Oxidation of tartaric acid in presence of iron, *J. Chem.*
331 *Soc., Trans.*, 1894, **65**, 899-910.
- 332 2. W. Y. Huang, M. Brigante, F. Wu, C. Mousty, K. Hanna and G. Mailhot,
333 Assessment of the Fe(III)-EDDS complex in Fenton-like processes: from the radical
334 formation to the degradation of bisphenol A, *Environ. Sci. Technol.*, 2013, **47**,
335 1952-1959.
- 336 3. L. Chen, C. Y. Deng, F. Wu and N. S. Deng, Decolorization of the azo dye
337 Orange II in a montmorillonite/H₂O₂ system, *Desalination*, 2011, **281**, 306-311.
- 338 4. Y. G. Zuo and J. Zhan, Effects of oxalate on Fe-catalyzed photooxidation of
339 dissolved sulfur dioxide in atmospheric water, *Atmos. Environ.*, 2005, **39**, 27-37.
- 340 5. E. M. Glebov, I. P. Pozdnyakov, V. P. Grivin, V. F. Plyusnin, X. Zhang, F. Wu
341 and N.S. Deng, Intermediates in photochemistry of Fe (III) complexes with carboxylic
342 acids in aqueous solutions, *Photoch. Photobio. Sci.*, 2011, **10**, 425-430.
- 343 6. Y. Deng and J. D. Englehardt, Treatment of landfill leachate by the Fenton
344 process, *Water Res.*, 2006, **40**, 3683-3694.
- 345 7. S. Q. Liu, L. R. Feng, N. Xu, Z. G. Chen and X. M. Wang, Magnetic nickel
346 ferrite as a heterogeneous photo-Fenton catalyst for the degradation of rhodamine B in
347 the presence of oxalic acid, *Chem. Eng. J.*, 2012, **203**, 432-439.
- 348 8. L. Y. Ge, H. H. Deng, F. Wu and N. S. Deng, Microalgae-promoted
349 photodegradation of two endocrine disrupters in aqueous solutions, *J. Technol.*
350 *Biotechnol.*, 2009, **84**, 331-336.

- 351 9. A. Durán, J. M. Monteagudo and E. Amores, Solar photo-Fenton degradation of
352 Reactive Blue 4 in a CPC reactor, *Appl. Catal. B: Environ.*, 2008, **80**, 42-50.
- 353 10. H. Katsumata, S. Kaneco, T. Suzuki, K. Ohta and Y. Yobiko, Photo-Fenton
354 degradation of alachlor in the presence of citrate solution, *J. Photochem. Photobiol., A*,
355 2006, **180**, 38-45.
- 356 11. Y. Chen, F. Wu, Y. X. Lin, N. S. Deng, N. Bazhin and E. Glebov,
357 Photodegradation of glyphosate in the ferrioxalate system, *J. Hazard. Mater.*, 2007,
358 **148**, 360-365.
- 359 12. Y. G. Zuo and J. Hoigné, Evidence for photochemical formation of H₂O₂ and
360 oxidation of SO₂ in authentic fog water, *Science*, 1993, **260**, 71-73.
- 361 13. Y. G. Zuo and J. Hoigne, Formation of hydrogen peroxide and depletion of oxalic
362 acid in atmospheric water by photolysis of iron (III)-oxalato complexes, *Environ. Sci.*
363 *Technol.*, 1992, **26**, 1014-1022.
- 364 14. P. Kocot, A. Karocki and Z. Stasicka, Photochemistry of the Fe(III)-EDTA
365 complexes, *J. Photochem. Photobiol., A*, 2006, **179**, 176-183.
- 366 15. E. Lipczynska-Kochany and J. Kochany, Effect of humic substances on the
367 Fenton treatment of wastewater at acidic and neutral pH, *Chemosphere*, 2008, **73**,
368 745-750.
- 369 16. C. Fan, L. Tsui and M. Liao, Parathion degradation and its intermediate formation
370 by Fenton process in neutral environment, *Chemosphere*, 2011, **82**, 229-236.
- 371 17. W. Y. Huang, M. Brigante, F. Wu, K. Hanna and G. Mailhot, Development of a
372 new homogenous photo-Fenton process using Fe(III)-EDDS complexes, *J.*

- 373 *Photochem. Photobiol., A*, 2012, **239**, 17-23.
- 374 18. J. A. Sánchez Pérez, I. M. Román Sánchez, I. Carra, A. Cabrera Reina, J. L.
375 Casas López and S. Malato, Economic evaluation of a combined photo-Fenton/MBR
376 process using pesticides as model pollutant. Factors affecting costs, *J. Hazard. Mater.*,
377 2013, **244-245**, 195-203.
- 378 19. I. Carra, S. Malato, M. Jiménez, M. I. Maldonado and J. A. Sánchez Pérez,
379 Microcontaminant removal by solar photo-Fenton at natural pH run with sequential
380 and continuous iron additions, *Chem. Eng. J.*, 2014, **235**, 132-140.
- 381 20. J. Granger and N. M. Price, The importance of siderophores in iron nutrition of
382 heterotrophic marine bacteria, *Limnol. Oceanog.*, 1999, **44**, 541-555.
- 383 21. T. P. Murphy, D. R. Lean and C. Nalewajko, Blue-green algae: their excretion of
384 iron-selective chelators enables them to dominate other algae, *Science*, 1976, **192**,
385 900-902.
- 386 22. D. A. Hutchins, A. E. Witter, A. Butler and G. W. Luther, Competition among
387 marine phytoplankton for different chelated iron species, *Nature*, 1999, **400**, 858-861.
- 388 23. C. G. Trick, R. J. Andersen, A. Gillam and P. J. Harrison, Prorocentrin: an
389 extracellular siderophore produced by the marine dinoflagellate *Prorocentrum*
390 *minimum*, *Science*, 1983, **219**, 306-308.
- 391 24. S. A. Amin, D. H. Green, F. C. Küpper and C. J. Carrano, Vibrioferrin, an
392 Unusual Marine Siderophore: Iron Binding, Photochemistry, and Biological
393 Implications, *Inorg. Chem.*, 2009, **48**, 11451-11458.
- 394 25. K. Barbeau, E. L. Rue, C. G. Trick, K. W. Bruland and A. Butler, Photochemical

- 395 reactivity of siderophores produced by marine heterotrophic bacteria and
396 cyanobacteria based on characteristic Fe (III) binding groups, *Limnol. Oceanog.*, 2003,
397 **48**, 1069-1078.
- 398 26. J. D. Martin, Y. Ito, V. V. Homann, M. G. Haygood and A. Butler, Structure and
399 membrane affinity of new amphiphilic siderophores produced by *Ochrobactrum* sp.
400 SP18, *J. Biol. Inorg. Chem.*, 2006, **11**, 633-641.
- 401 27. R. T. Reid, D. H. Livet, D. J. Faulkner and A. Butler, A siderophore from a
402 marine bacterium with an exceptional ferric ion affinity constant, *Nature*, 1993, **366**,
403 455-458.
- 404 28. J. F. Wu and G. W. Luther III, Complexation of Fe (III) by natural organic
405 ligands in the Northwest Atlantic Ocean by a competitive ligand equilibration method
406 and a kinetic approach, *Mar. Chem.*, 1995, **50**, 159-177.
- 407 29. S. Pérez-Miranda, N. Cabirol, R. George-Téllez, L. S. Zamudio-Rivera and F. J.
408 Fernández, O-CAS, a fast and universal method for siderophore detection, *J.*
409 *Microbiol. Meth.*, 2007, **70**, 127-131.
- 410 30. C. G. Hatchard and C. A. Parker, A new sensitive chemical actinometer. II.
411 Potassium ferrioxalate as a standard chemical actinometer, *Proc. R. Soc. A*, 1956, **235**,
412 518-536.
- 413 31. M. E. Bodini, M. A. Del Valle, R. Tapia, F. Leighton and L. Gonzalez, Study of
414 the iron catechin complexes in dimethyl sulphoxide. Redox chemistry and interaction
415 with superoxide radical anion in this medium, *Bol. Soc. Chil. Quím.*, 2001, **46**,
416 309-317.

- 417 32. M. J. Hynes and M. Ó Coinceanainn, The kinetics and mechanisms of the
418 reaction of iron (III) with gallic acid, gallic acid methyl ester and catechin, *J. Inorg.*
419 *Biochem.*, 2001, **85**, 131-142.
- 420 33. A. S. Cornish and W. J. Page, The catecholate siderophores of *Azotobacter*
421 *vinelandii*: their affinity for iron and role in oxygen stress management, *Microbiology*,
422 1998, **144**, 1747-1754.
- 423 34. Z. P. Wang, Z. Z. Liu, F. Yu, J. P. Zhu, Y. Chen and T. Tao,
424 Siderophore-modified Fenton-like system for the degradation of propranolol in
425 aqueous solutions at near neutral pH values, *Chem. Eng. J.*, 2013, **229**, 177-182.
- 426 35. F. Baldi, D. Marchetto, D. Zanchettin, E. Sartorato, S. Paganelli and O. Piccolo,
427 A bio-generated Fe (III)-binding exopolysaccharide used as new catalyst for phenol
428 hydroxylation, *Green Chem.*, 2010, **12**, 1405-1409.
- 429
- 430

431

Captions for Figures

432 **Figure 1** – Inderal photodegradation by five Fe(III) complexes. Reaction conditions
433 were as follows: [Fe(III)] = 50 $\mu\text{mol/L}$, [ligand] = 200 $\mu\text{mol/L}$, [inderal]
434 =10 $\mu\text{mol/L}$, pH = 6.0.

435 **Figure 2** – Comparison of Fe(III)/light (with or without CAT) processes. Reaction
436 conditions were as follows: [Fe(III)] = 50 $\mu\text{mol/L}$, [catechin] = 200 $\mu\text{mol/L}$, [inderal]
437 = 10 $\mu\text{mol/L}$, pH = 6.0.

438 **Figure 3** – Effect of 2-propanol on inderal photodegradation. Reaction conditions
439 were as follows: [Fe(III)] = 50 $\mu\text{mol/L}$, [catechin] = 200 $\mu\text{mol/L}$, [inderal] = 10
440 $\mu\text{mol/L}$, pH = 6.0.

441 **Figure 4** – Effect of pH on inderal photodegradation. Reaction conditions were as
442 follows: [Fe(III)] = 50 $\mu\text{mol/L}$, [catechin] = 200 $\mu\text{mol/L}$, [inderal] =10 $\mu\text{mol/L}$.

443 **Figure 5** – Inderal degradation in the presence of catechins of various concentrations.
444 Reaction conditions were as follows: [Fe(III)] = 50 $\mu\text{mol/L}$, [inderal] =10 $\mu\text{mol/L}$, pH
445 = 6.0

446 **Figure 6** – The proposed pathways for indirect photodegradation of inderal in the
447 Fe(III)–catechin system.

448 **Figure 7** – Photodegradation of inderal after addition of catechin (200 μM) to the
449 Fe(III) solution (50 μM) in the three runs at pH 6.0.

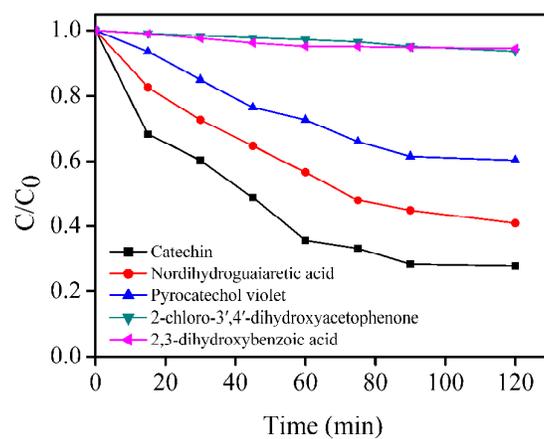
450

451 **Scheme 1** – Stabilization of Fe(III) at near neutral pH by CAT and formation of
452 photo-Fenton system.

453

454

455



456

457 **Figure 1** – Inderal photodegradation by five Fe(III) complexes. Reaction conditions458 were as follows: [Fe(III)] = 50 $\mu\text{mol/L}$, [ligand] = 200 $\mu\text{mol/L}$, [inderal]459 =10 $\mu\text{mol/L}$, pH = 6.0.

460

461

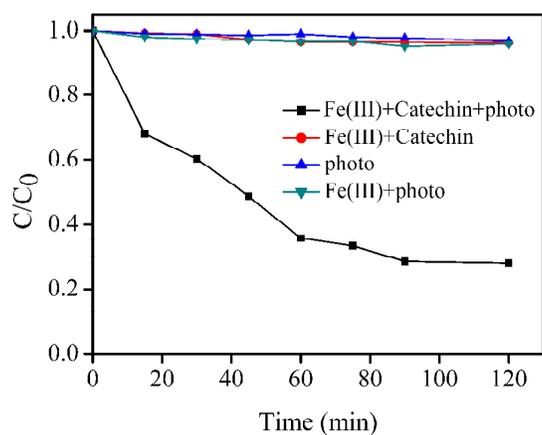
462

463

464

465

466



467

468 **Figure 2** – Comparison of Fe(III)/light (with or without CAT) processes. Reaction469 conditions were as follows: [Fe(III)] = 50 $\mu\text{mol/L}$, [catechin] = 200 $\mu\text{mol/L}$, [inderal]470 = 10 $\mu\text{mol/L}$, pH = 6.0.

471

472

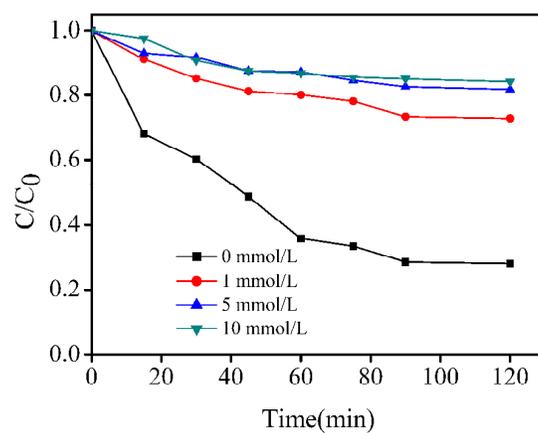
473

474

475

476

477



478

479 **Figure 3** – Effect of 2-propanol on inderal photodegradation. Reaction conditions480 were as follows: [Fe(III)] = 50 $\mu\text{mol/L}$, [catechin] = 200 $\mu\text{mol/L}$, [inderal] = 10481 $\mu\text{mol/L}$, pH = 6.0.

482

483

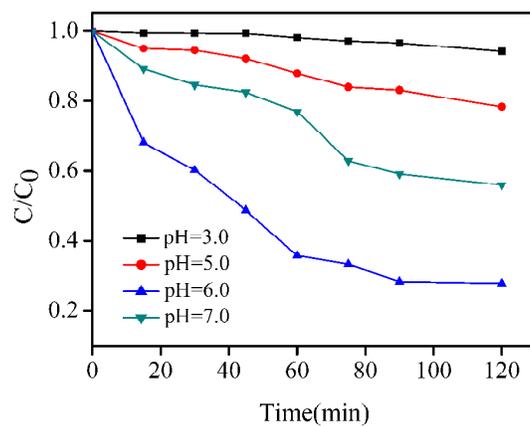
484

485

486

487

488



489

490 **Figure 4** – Effect of pH on inderal photodegradation. Reaction conditions were as491 follows: [Fe(III)] = 50 $\mu\text{mol/L}$, [catechin] = 200 $\mu\text{mol/L}$, [inderal] = 10 $\mu\text{mol/L}$.

492

493

494

495

496

497

498

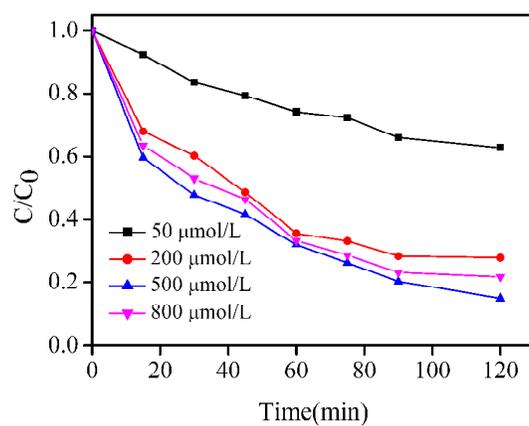
499

500

501

502

503



504

505 **Figure 5** – Inderal degradation in the presence of catechins of various concentrations.

506 Reaction conditions were as follows: [Fe(III)] = 50 µmol/L, [inderal] = 10 µmol/L, pH

507 = 6.0

508

509

510

511

512

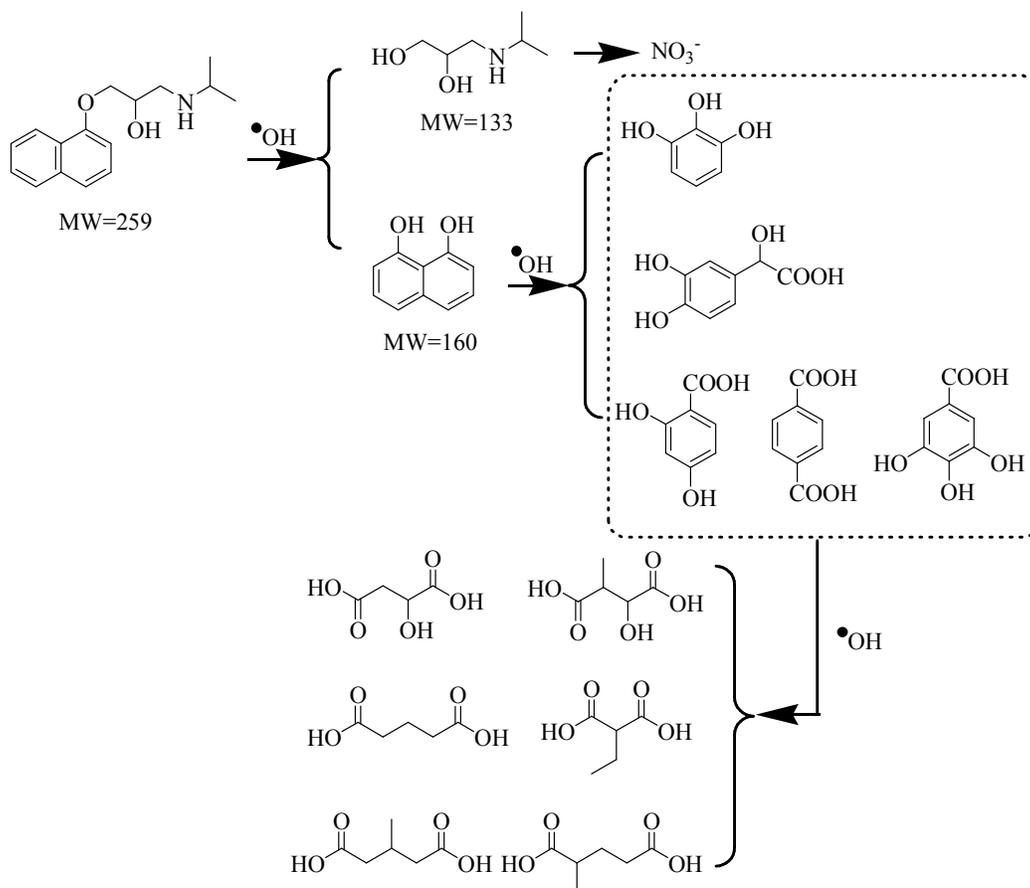
513

514

515

516

517

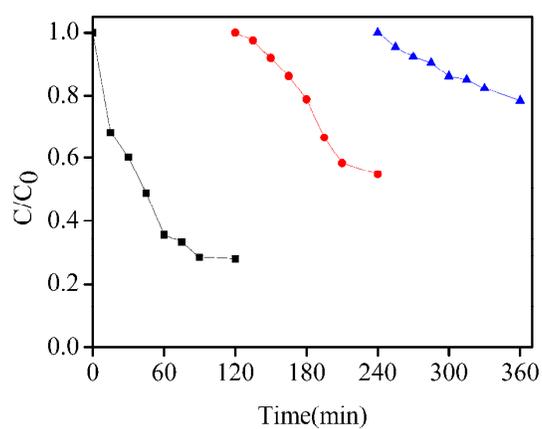


518

519 **Figure 6** – The proposed pathways for indirect photodegradation of inderal in the

520 Fe(III)–catechin system.

521



522

523 **Figure 7** – Photodegradation of inderal after addition of catechin (200 μM) to the524 Fe(III) solution (50 μM) in the three runs at pH 6.0.

525

526

527

528

529

530

531

532

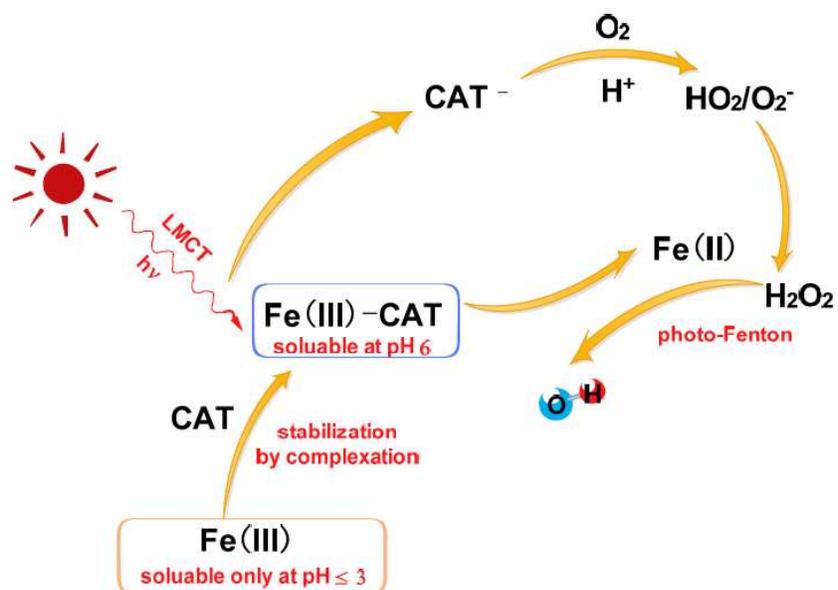
533

534

535

536

537



538

539 **Scheme 1** – Stabilization of Fe(III) at near neutral pH by CAT and formation of

540 photo-Fenton system.

541

542

543