

**Spray deposited  $\beta$ -Bi<sub>2</sub>O<sub>3</sub> nanostructured films with visible photocatalytic activity for solar water treatment**

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## Spray deposited $\beta$ - $\text{Bi}_2\text{O}_3$ nanostructured films with visible photocatalytic activity for solar water treatment

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### Abstract

Bismuth oxide thin films were obtained by spray pyrolysis method using bismuth acetate as precursor salt. The films were characterized by X-ray diffraction (XRD), Fourier Transform Infrared (FTIR), UV-vis diffuse reflectance, X-ray Photoelectron Spectroscopy (XPS) and Scanning Electron Microscopy (SEM). The XRD patterns indicated that the pure  $\beta$  phase is obtained at 450 °C and was also confirmed by FTIR. This phase presents a nanoplate morphology which is adequate for the photocatalytic reactions. Moreover, the band gap value was of 2.6 eV indicating a good capacity of visible light absorption. The photocatalytic degradation of the Methyl Orange (MO) dye was pH dependent, being the acid solution easier to degrade. However, the  $\text{Bi}_2\text{O}_3$  films were easily converted into  $\text{BiOCl}$  when they were in contact with a solution containing  $\text{HCl}$ . In order to preserve the  $\beta$ - $\text{Bi}_2\text{O}_3$  phase, the Acid Blue 113 dye with its natural pH of 8 was used to evaluate the stability of the photocatalytic activity after five degradation cycles. The photoactivity

was practically stable indicating a good performance of the material. This encouraged us to test the films in a continuous flow solar reactor prototype for the degradation of the dye solution using sunlight radiation exclusively. The good performance of the  $\beta$ -Bi<sub>2</sub>O<sub>3</sub> films indicates that they can be used for sustainable water treatment applications.

Keywords: Bismuth oxide; thin films; photocatalysis; photoreactor; sunlight.

## 1. Introduction

Photocatalytic water treatment technologies are of special interest nowadays, where environmental friendly and sustainable processes need to be implemented in order to keep the modern life style. The real capabilities of the process have been widely demonstrated for TiO<sub>2</sub> based materials<sup>1, 2</sup>, with the disadvantage of the requirement of continuous UV light illumination that does not allow the use of natural sunlight energy. The scientific community is therefore looking for either doping/modifying the TiO<sub>2</sub> materials in order to enhance the visible-light response or to find other semiconductor materials with lower-than-TiO<sub>2</sub> band-gap energies which present comparable photocatalytic activities. One of such materials is Bismuth oxide, for which it has been shown that micrometric and nanometric powders and nanostructures have a strong potential for photocatalytic degradation of dyes (rhodamine B<sup>3</sup>, methyl orange<sup>4, 5</sup>, indigo carmine, methylene blue<sup>6</sup>, malachite green<sup>7</sup>, etc...), drugs (acetaminophen<sup>8</sup>, tetracycline<sup>9</sup>), 4-chlorophenol<sup>3, 4</sup>, acetaldehyde<sup>10</sup> as well as gases (toluene, NO and HCHO<sup>11</sup>). The use of slurry photocatalytic systems is very efficient due to the large surface active sites of the powders/nanoparticles. However, for a photoreactor a second process has to be implemented in order to recover the slurry, which increases the cost. The other option is the immobilization of the photocatalyst into an adequate support, where the difficulty resides on the attachment of the nanometric catalysts to the support. On the other hand, thin film photocatalyst could be used as a coating directly applied into the photoreactor walls, or deposited on grids through which the water can go through. The lower effective surface area obtained for thin films demands better and more efficient photocatalytic semiconductors, thus recent efforts are focused on obtaining higher surface area photocatalysts as well as visible-light active materials. One of these new promising materials in thin film form is bismuth oxide (Bi<sub>2</sub>O<sub>3</sub>).

Bi<sub>2</sub>O<sub>3</sub> possesses six polymorphisms:  $\alpha$ -Bi<sub>2</sub>O<sub>3</sub> (monoclinic),  $\beta$ -Bi<sub>2</sub>O<sub>3</sub> (tetragonal),  $\gamma$ -Bi<sub>2</sub>O<sub>3</sub> (BCC),  $\delta$ -Bi<sub>2</sub>O<sub>3</sub> (cubic),  $\varepsilon$ -Bi<sub>2</sub>O<sub>3</sub> (triclinic) and  $\omega$ -Bi<sub>2</sub>O<sub>3</sub> (triclinic), being this last one recently synthesized by Cornei et al.<sup>12, 13</sup>. The  $\alpha$  and  $\delta$  phases are stable, while  $\beta$ ,  $\gamma$ ,  $\varepsilon$  and  $\omega$  are metastable. Among all these polymorphic forms,  $\beta$ -Bi<sub>2</sub>O<sub>3</sub> has the strongest absorption in the visible light because it has the smallest band-gap ( $\sim 2.58$  eV)<sup>14</sup>, making it a good candidate as visible light-activated photocatalyst. However, this tetragonal phase is metastable and its synthesis is not easy, especially on the nanoscale<sup>15</sup>. The formation of non-stoichiometric bismuth oxide phases has been also frequently reported, like the Bi<sub>2</sub>O<sub>3</sub>/Bi<sub>2</sub>O<sub>3+x</sub> system produced by thermal vacuum treatment reported by Sajjad et al.<sup>16</sup> which produced new forms of Bi creating new defect states, or the Bi<sub>2</sub>O<sub>3</sub>/Bi<sub>2</sub>O<sub>4-x</sub> nanocomposite reported by Hameed et al. which showed a good photocatalytic degradation of methylene blue,

methyl orange and phenol<sup>17</sup>. Furthermore, Bi<sub>2</sub>O<sub>3</sub> is a p-type semiconductor with suitable band edge potentials for water oxidation<sup>17-19</sup>. As thin films, the  $\delta$ -Bi<sub>2</sub>O<sub>3</sub> has been obtained by electrodeposition by Koza et al.<sup>20</sup> and Helfen et al.<sup>21</sup> Oxidation processes of Bi-deposited films were applied by Leontie et al.<sup>14</sup> and Patil et al.<sup>22</sup> The Sol-Gel route has been used by Fruth et al.<sup>23</sup>, obtaining an oxygen deficient phase, and by Xiaohong et al.<sup>24</sup> who showed that monoclinic and tetragonal phases could be obtained after annealing above 550 °C. Spray pyrolysis has been used by Lockhande and Bhosale<sup>25</sup> and Rico-Fuentes et al.<sup>26</sup> The latter used bismuth acetate as the source solution and obtained two bismuth oxide phases:  $\beta$ -Bi<sub>2</sub>O<sub>3</sub> and the substoichiometry Bi<sub>2</sub>O<sub>2.33</sub>; this latter was converted into the  $\beta$ -phase after annealing. Soitah et al.<sup>27</sup> studied the optical and electrical properties of bismuth oxide films produced by a Pechini modified route and annealing at temperatures ranging between 400 °C and 700 °C. Physical deposition methods, such as pulsed laser deposition and magnetron sputtering have also been used<sup>28-32</sup>. A strong influence of the synthesis method and synthesis parameters on the Bi<sub>2</sub>O<sub>3</sub> phases and their properties has been observed, and still there are many questions to answer in order to understand this compound better. In this paper, we demonstrate that it is possible to obtain a pure  $\beta$ -phase Bi<sub>2</sub>O<sub>3</sub> film by selecting the appropriate temperature and deposition parameters of the spray pyrolysis method. We also show that this  $\beta$ -phase is photocatalytic active not only under UV but also using visible light; this fact motivated us to incorporate this material into a solar reactor prototype to prove its possible use in water treatment applications.

## 2. Experimental details

### 2.1 Synthesis of Bi<sub>2</sub>O<sub>3</sub> films

The precursor solution was prepared using Bismuth(III) acetate Bi(CH<sub>3</sub>CO<sub>2</sub>)<sub>3</sub> from Sigma-Aldrich in a molar concentration of 0.05, dissolved in acetic acid 25%v and deionized water 75%v at 45 °C under constant stirring in order to obtain a transparent homogeneous solution. This solution was deposited by the pneumatic spray pyrolysis technique at temperatures of 350, 375, 400, 425 and 450 °C, in order to identify the optimum deposition temperature. For each deposition 100 mL of the solution were sprayed at a flow rate of 4 mL/min and an air flow rate of 1.13 L/min. The substrates used in this work were Corning glass slides of 2.5 cm x 1.25 cm and c-Si substrates of 1 cm x 1 cm. The former were cleaned with trichloroethylene, acetone and methanol consecutively in an ultrasonic bath and dried with compressed nitrogen. Silicon substrates were cleaned with P solution, rinsed with deionized water and dried with compressed nitrogen.

## 2.2 Characterization

The films were characterized by X-ray diffraction (D8 Advance Bruker) using the  $\text{CuK}_\alpha$  wavelength (1.54056 Å). The morphology of the films was studied by Scanning Electron Microscopy (JSM-7600F Jeol). The thickness and roughness were measured with a profilometer (Sloan Dektak IIA) and Fourier Transformed Infrared (FTIR) spectra were acquired using a Nicolet 210 FTIR. The band gap of the material was determined from the diffuse reflectance spectra (DRS) acquired using a Perkin Elmer Lambda 35 UV-vis spectrophotometer equipped with an integration sphere; Spectralon® was used as reference blank. A PHI 5000 Versaprobe II Scanning X-ray Photoelectron Spectrometer (monochromatic Al K-alpha X-ray source with 1486.6 eV energy, 15 kV voltage and 1 mA anode current), was used to investigate the surface chemical composition. A spot size of 100  $\mu\text{m}$  was used in order to collect the photoelectron signal for both the high resolution (HR) and the survey spectra. Different pass energy values were exploited: 187.85 eV for survey spectra and 23.5 eV for HR peaks.

## 2.3 Photocatalytic tests

The photocatalytic activity of the  $\text{Bi}_2\text{O}_3$  films was evaluated by the degradation of a typical organic dye solution: either Methyl Orange (MO) dye ( $\text{C}_{14}\text{H}_{14}\text{N}_3\text{NaO}_3\text{S}$ ) or Acid Blue 113 ( $\text{C}_{32}\text{H}_{21}\text{N}_5\text{Na}_2\text{O}_6\text{S}_2$ ) both from Sigma-Aldrich, with a concentration of  $10^{-6}$  M (their absorption spectra are shown in Fig. A of the Electronic Supplementary Information). The films were placed in pairs inside a vial with 10 mL of the dye solution and exposed to different light sources in order to study the effect of the energy radiation. Three light sources with different wavelength ranges were used: a UV light lamp (26  $\text{W}/\text{m}^2$ ), a white light lamp (33  $\text{W}/\text{m}^2$ ) and simulated sunlight radiation (Oriel 96000 Xenon lamp with an intensity of 397  $\text{W}/\text{m}^2$ ). The degradation was measured by following the absorbance spectrum of the dye solution each 30 min. Total Organic Carbon (TOC) measurements were obtained using a Shimadzu TOC-V CSH and High Performance Liquid Chromatography (HPLC) Perkin Elmer 600 series Link; chromatography was used to analyze the dye decomposition.

## 3. Results

### 3.1 Microstructure, morphology and composition

Fig. 1a shows the evolution of the crystalline structure of the bismuth oxide films with the deposition temperature. At low temperatures, both BiO and  $\beta\text{-Bi}_2\text{O}_3$

phases were observed, as well as one peak corresponding to bismuth acetate (BiAc); this indicates that at these temperatures residues of the precursor are still present. When the temperature was increased to 450 °C, the pure  $\beta$ - $\text{Bi}_2\text{O}_3$  phase was obtained, as it is confirmed by the FTIR spectra of Fig. 1b. FTIR analysis shows the evolution of  $\text{BiO}_6$ ,  $\text{BiO}_3$  and  $\text{BiO}$  compounds with increasing temperature leading to the  $\beta$ - $\text{Bi}_2\text{O}_3$  phase at 450 °C. This was confirmed by XRD on several repetitions of the growth of the material at 450 °C, as shown in Fig. 1d, where the most prominent peaks correspond to the (201) and (220) families, according to the ICDD file No. 78-1793. Fig. 1c, shows the morphology of the  $\beta$ - $\text{Bi}_2\text{O}_3$  films, composed by mostly oriented nanoplates with rounded borders, with approximate dimensions of 30 x 200 nm. Such morphology is ideal for photocatalytic processes as it provides higher surface area than a flat surface.

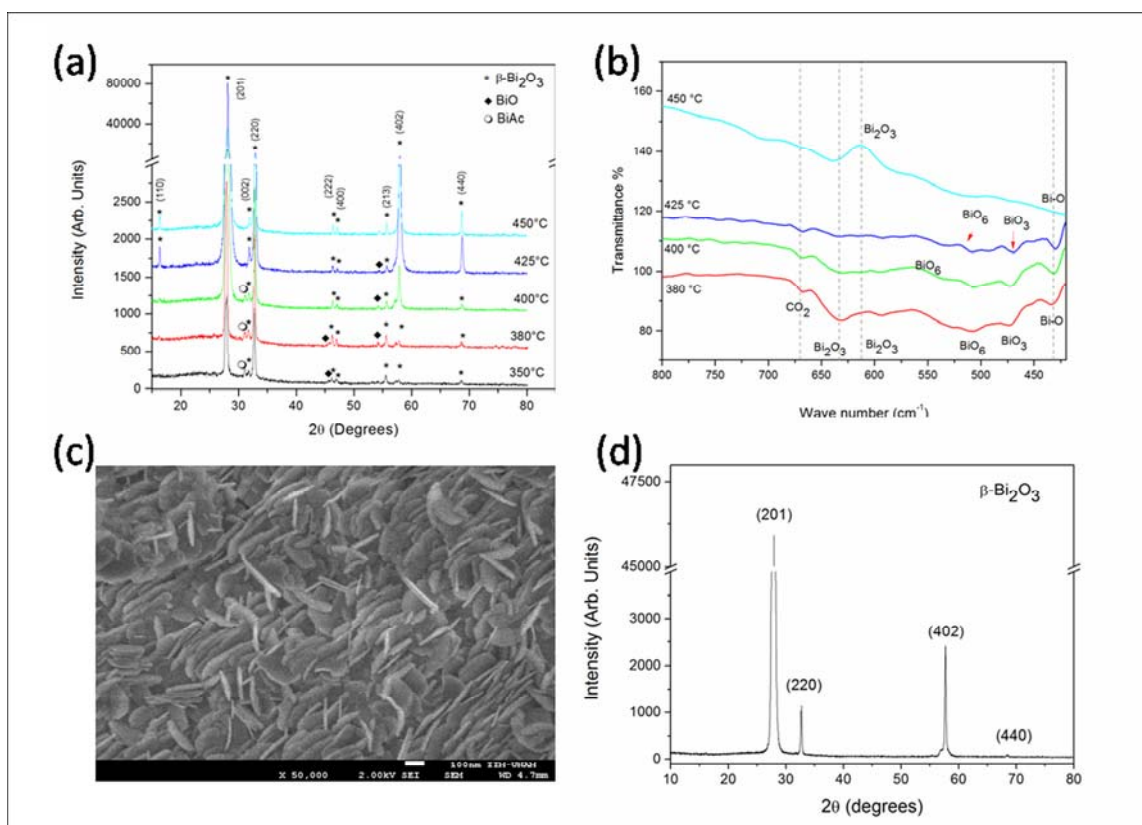


Fig. 1 (a) XRD patterns of films deposited at different temperatures, it is appreciated that  $\text{BiO}$  and  $\text{BiAc}$  peaks disappear at 450 °C. (b) FTIR spectra of the samples grown at different temperatures. (c) SEM image of the  $\beta$ - $\text{Bi}_2\text{O}_3$  film grown at 450°C with a nanoplate structure. (d) XRD pattern of the sample grown at 450 °C, where the  $\beta$ - $\text{Bi}_2\text{O}_3$  phase is exclusively obtained.

To evaluate the composition of the films, XPS analyses of the films were performed without argon cleaning. The XPS survey spectrum (Fig. 2a) shows three main photo-electronic peaks: O1s, C1s and Bi4f; their relative atomic concentrations were evaluated after subtracting the background using a Shirley function<sup>33</sup> obtaining a relative atomic composition of 49.2% of C, 37.7% of O and 13.1% of Bi. The carbon signal is coming from adventitious carbon, *i.e.* carbon species adsorbed on the semiconductor surface as a consequence of the exposition of the sample to atmospheric conditions. High resolution (HR) spectra of these peaks were acquired in order to evaluate the chemical bonding state. After an exhaustive research in literature<sup>34,35</sup> the peaks were attributed to different chemical species. The C1s peak is made up by three bonds: a first peak at lower binding energy due to C-C/C-H, a small shoulder due to C-O and a well-separated third peak due to C=O<sup>36</sup>. The O1s peak is also made up by three components: two of them due to the bonds with C and a bigger one at higher energy due to Bi oxide. Bi4f doublet peak (Bi4f<sub>7/2</sub> and Bi4f<sub>5/2</sub>) is made up by the superposition of two peaks each: the first at 156.8 eV assigned to metallic Bi<sup>37</sup> and the second one at 158.5 eV assigned to Bi<sub>2</sub>O<sub>3</sub> (the same two peaks are present also in the Bi4f<sub>5/2</sub> separated by a  $\Delta=5.4$  eV).



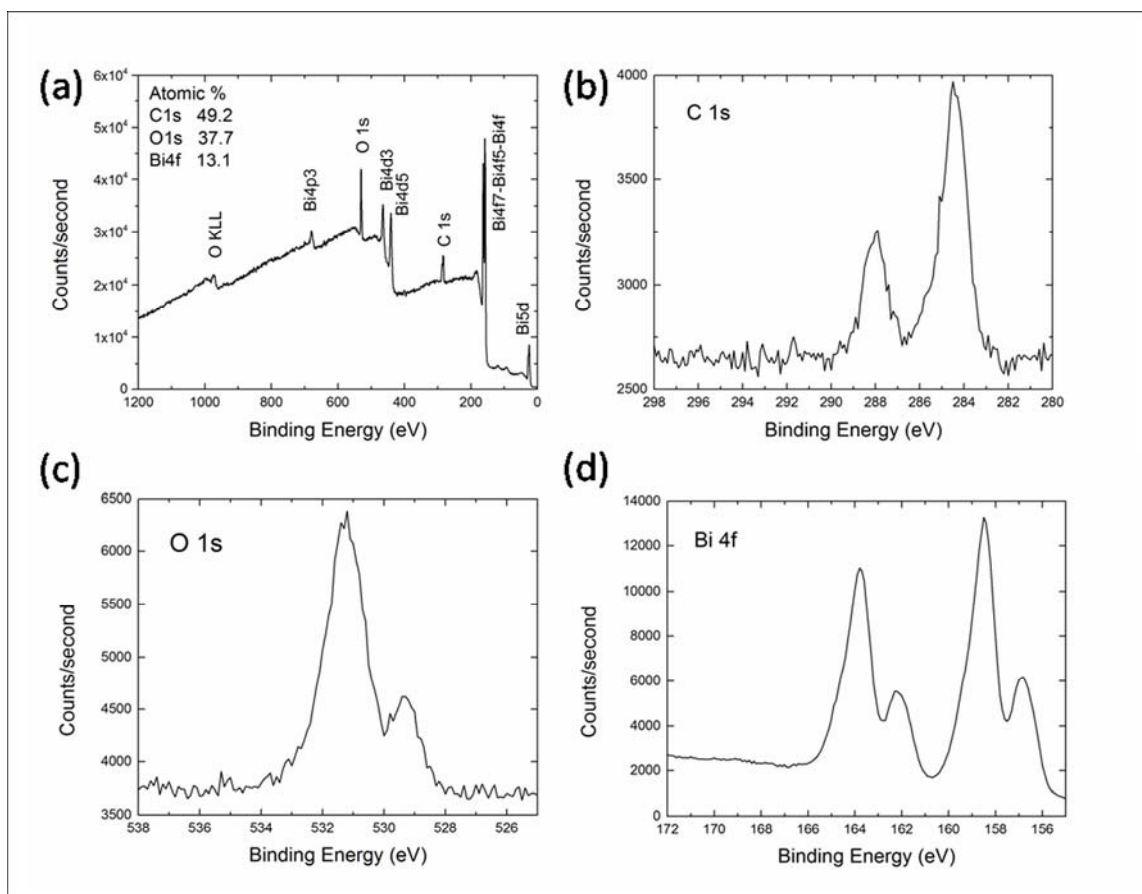


Fig. 2 XPS analysis of  $\text{Bi}_2\text{O}_3$  films. (a) survey spectrum, (b) C1s, (c) O1s, (d) Bi 4f high resolution spectra.

### 3.2 Optical properties

The optical properties of the films were evaluated by their UV-vis reflectance spectra. To convert the diffuse reflectance to the absorption coefficient  $\alpha$  the Kubelka-Munk function  $F(R)$  was calculated as shown in eq. 1; then  $[F(R) \cdot E^2]$  was plotted against the energy  $E$  and the extrapolation of the linear fitting to its intersection with the abscissas determined the band gap energy  $E_g$ , as can be seen in Fig. 3.

$$\alpha \approx \frac{(1-R_\infty)^2}{2R_\infty} \equiv F(R_\infty) \quad (\text{Eq. 1})$$

The band gap of all the  $\beta\text{-Bi}_2\text{O}_3$  films was determined giving an average value of  $2.60 \pm 0.07$  eV, indicating a good agreement with the reported values for this

phase<sup>38</sup>. This band gap value causes the material to absorb light of wavelengths down to 477 nm or in the blue region of the visible spectrum.

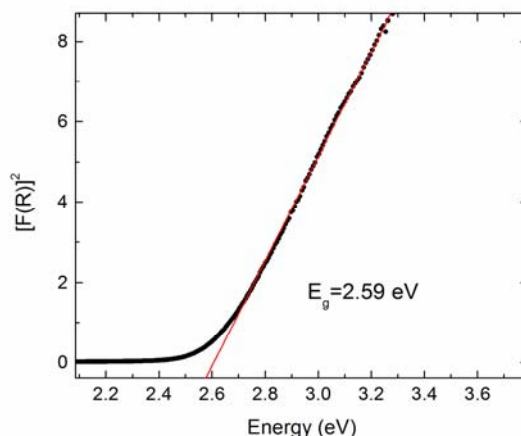


Fig. 3 Example of the determination of the energy band gap by means of the Kubelka-Munk function.

### 3.3 Photocatalysis

The photocatalytic activity of the  $\beta$ - $\text{Bi}_2\text{O}_3$  films was initially tested on the degradation of the methyl orange dye. The first tests indicated that the photocatalytic activity was highly pH-dependent, showing the highest degradation for the lowest pH values, as presented in figure 4a. This may be due to the specific form of the MO molecule in acid or basic environment<sup>39</sup> (insets of fig. 4a). As the degradation of the dye solution at neutral pH was very low we decided to evaluate the photocatalytic activity of the films under acidic conditions (pH=3). In this case, a good photocatalytic performance of the  $\beta$ - $\text{Bi}_2\text{O}_3$  was observed, achieving more than 80% of discoloration of the dye after 180 min (Fig. 4b), as it can be seen in the change of the absorption spectrum of the dye (bottom inset of figure 4b). The kinetics analysis revealed a first order reaction (top inset of fig. 4b), with an apparent rate of reaction  $k=9.5 \times 10^{-3} \text{ min}^{-1}$ . Some reports related to the use of immobilized catalysts as thin layers, describe the degradation reaction of organic pollutants with the Langmuir-Hinshelwood kinetic model<sup>40, 41</sup> and the mass transfer effect has been found to be negligible in some cases<sup>42</sup>, but this has to be verified in each specific reactor configuration. As it was previously mentioned, the band gap of the  $\beta$ - $\text{Bi}_2\text{O}_3$  films was of 2.6 eV, which means that it is able to absorb visible light starting from wavelengths of 477 nm. In order to prove this, three equivalent films were tested under different illumination sources: (i) simulated sunlight, (ii) white

light, and (iii) UV light again to compare; the emission spectra of each lamp was reported previously<sup>43</sup>. Figure 4c shows the performance of the films under these three irradiation sources, where is remarkable the highest efficiency of the films under sunlight achieving the complete discoloration in 180 minutes. The sample exposed to UV light degraded 80% of the dye, whereas the sample exposed to white light only reached about 60% in the same time. The high degradation efficiency of the sample exposed to simulated sunlight is attributed to the intensity of the lamp, which was higher than that used for the other two lamps. However, a degradation of 60% of the dye using white light is particularly noteworthy when compared to TiO<sub>2</sub> –the widest used photocatalyst– which is only activated by UV light. Just as a reference for comparison purposes, in a previous work TiO<sub>2</sub> films with different surface morphologies were produced by spray pyrolysis and tested under similar conditions, illuminating only with UV light<sup>44</sup>. The TiO<sub>2</sub> films modified with P25 powder supported on the surface presented a rate of reaction of 0.0069 min<sup>-1</sup>, while the pure TiO<sub>2</sub> P25 powder had a value of 0.02 min<sup>-1</sup>. In the present case, the β-Bi<sub>2</sub>O<sub>3</sub> films showed a rate of reaction of 0.0095 min<sup>-1</sup> under UV light, which is still nearly half the value of the P25 powder. Nonetheless compared to the TiO<sub>2</sub> films, the β-Bi<sub>2</sub>O<sub>3</sub> films presented a higher reaction rate. This result is encouraging especially because bismuth oxide is able to absorb visible light.

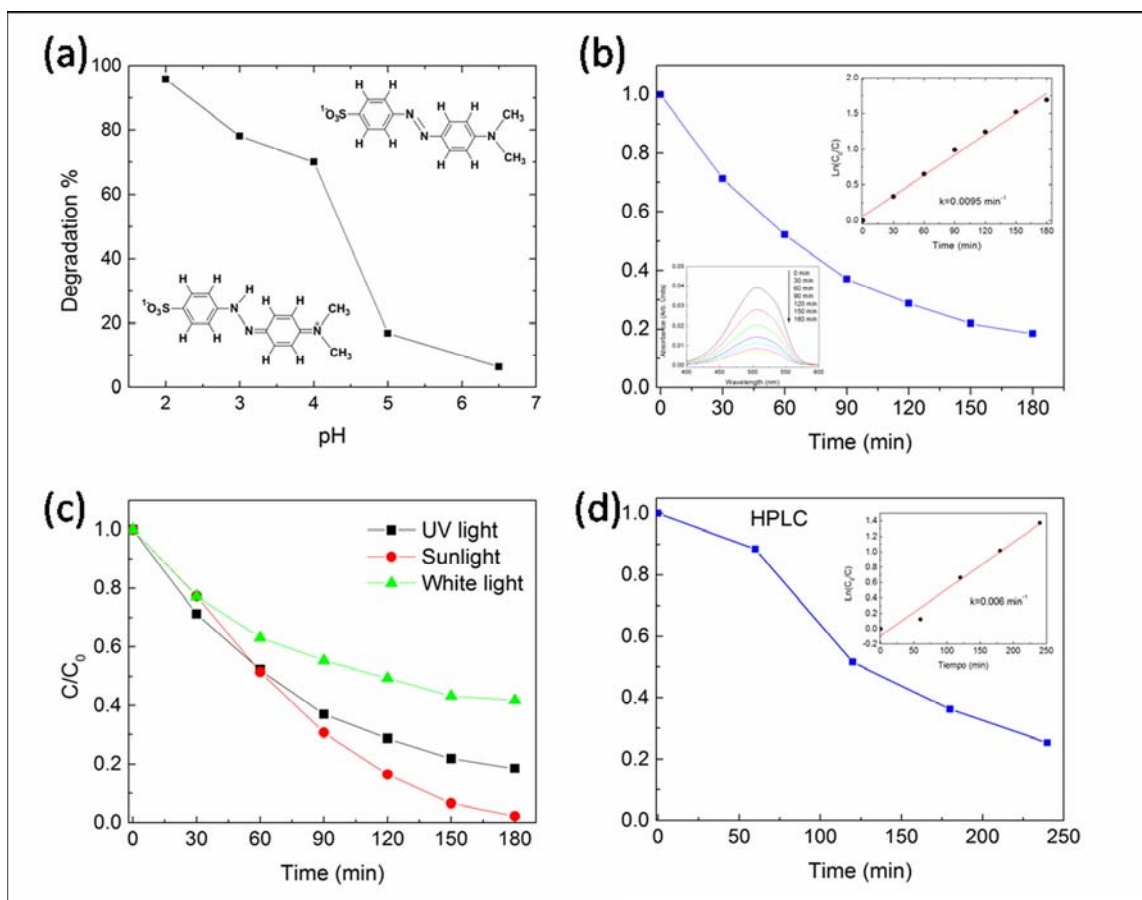
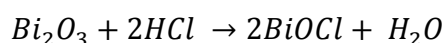


Figure 4. (a) pH dependence of the MO degradation with  $\beta$ - $\text{Bi}_2\text{O}_3$  films. (b) Photocatalytic degradation of the dye during time. The inset at the bottom shows the change of the absorption spectrum of the dye; the inset at the top shows a first order reaction rate. (c) Effect of different illumination sources on the photocatalytic degradation of the MO dye. (d) Degradation of the MO dye followed by HPLC technique; the inset shows a linear kinetic of the reaction.

HPLC is a very precise technique to measure the concentration of the dye solution and determine its degradation. This technique was used to analyze the degradation of the MO solution with a  $10^{-5} \text{ M}$  concentration exposed to direct sunlight and using the  $\beta$ - $\text{Bi}_2\text{O}_3$  films. It is important to remark that the higher the concentration, the slower the reaction occurs; however, the evolution of the concentration of the dye solution was good, reaching a 70% of degradation in 4 hours (with a  $k=6 \times 10^{-3} \text{ min}^{-1}$ ) as shown in figure 4b. The total organic carbon (TOC) was also measured before and after the degradation test. The obtained values were  $\text{TOC}_{\text{initial}}=2.599 \text{ mg/L}$ ,  $\text{TOC}_{\text{final}}=1.680 \text{ mg/L}$ , which means that the photocatalyst mineralized 35% of the dye (0.919 mg/L). It is important to remark

this issue, because even though we can have a transparent solution after the degradation process, the complete mineralization is not yet achieved. More time is required to complete the necessary reactions to get water free of contaminants.

The stability of the photocatalytic activity of the films was evaluated by repeating the degradation experiments reusing the same films. Fig. 5a indicates that the photocatalytic efficiency remained practically constant after five degradation cycles of the MO dye under simulated sunlight. However, after the fifth cycle, a change in the color of the film was noticed (from a light yellow to a white color), indicating a modification of the composition of the film. XRD analysis was performed on this sample after the treatment, revealing the presence of only BiOCl, as shown in Fig. 5b. The reason of this transformation is that the  $\beta$ -Bi<sub>2</sub>O<sub>3</sub> film was in contact with a solution that contained HCl to reduce the pH, reacting as follows:



Considering the Gibbs energy of the products minus the Gibbs energy of the reactants at 298 K, the energy of the reaction is:

$$\begin{aligned} \Delta E &= 2\Delta G_f^0(\text{BiOCl}) + \Delta G_f^0(\text{H}_2\text{O}) - \Delta G_f^0(\text{Bi}_2\text{O}_3) - 2G_f^0(\text{HCl}) \\ &= 2(-322.17 \text{ kJ/mol}) + (-228.59 \text{ kJ/mol}) - (-493.71 \text{ kJ/mol}) - 2(-95.31 \text{ kJ/mol}) \\ &= -569.84 \text{ kJ/mol} \end{aligned}$$

This indicates that the reaction is favorable. In order to avoid the transformation of  $\beta$ -Bi<sub>2</sub>O<sub>3</sub> to BiOCl, a different dye was tested without the addition of acid. For this experiment Acid Blue 113 dye was used leaving the solution with its original pH value of 8. Two different concentrations ( $10^{-6}$  and  $10^{-5}$  M) were tested under simulated sunlight, achieving 71.7 and 59.2% of dye degradation, respectively (Fig. 5c). The decrease of the degradation rate is due to the higher number of dye molecules in the solution that must adsorb into the film's surface, then react, desorb and incorporate again into the liquid in order to continue the process. Nonetheless, the degradation is only slightly lower although the concentration is 10 times larger. The total organic carbon of the Acid Blue 113 dye was measured before and after the photocatalytic process and the values were  $\text{TOC}_{\text{initial}}=3.429$  mg/L and  $\text{TOC}_{\text{final}}=2.425$  mg/L, indicating a mineralization of 29.3% of the dye. This response suggested that the  $\beta$ -Bi<sub>2</sub>O<sub>3</sub> films could be used in a solar photocatalytic reactor prototype in order to scale up the system for possible use in water treatment. The analysis of the adsorption of the dye onto the film's surface before the illumination showed that the Acid Blue 113 dye adsorbed more onto the Bi<sub>2</sub>O<sub>3</sub> film (~16%) than the methyl orange dye (negligible), as can be seen in Fig. B of the Electronic Supplementary Information.

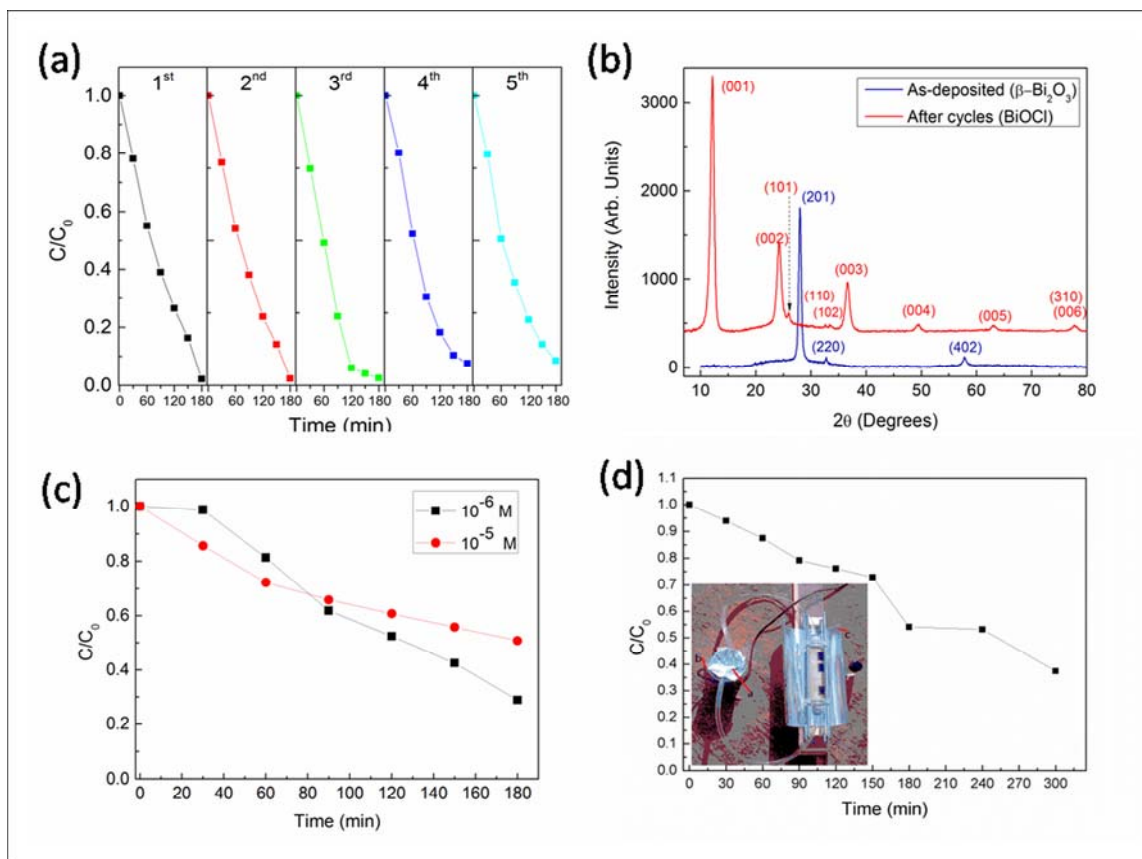


Fig. 5 (a) Degradation test under simulated sunlight after five cycles; the efficiency of the films remains practically constant. (b) XRD of as-deposited  $\beta$ - $\text{Bi}_2\text{O}_3$  film and after 5 degradation cycles, where it transformed to  $\text{BiOCl}$ . (c) Degradation of acid blue 113 at two different concentrations. (d) Dye degradation performance of the  $\beta$ - $\text{Bi}_2\text{O}_3$  films using a CPC solar reactor.

For this purpose, a small scale home-made Compound Parabolic Collector (CPC) solar reactor was implemented, and it is shown in Fig. 6; Fig. 6a shows the diagram of a cross section view of the reactor, and Fig. 6b is a picture of the prototype. Ten photocatalytic films with a total area of  $62.5 \text{ cm}^2$  were placed along the axis of the reactor tube and the dye solution was pumped through it at a rate of  $10 \text{ mL/s}$ . The CPC has an acceptance angle of  $90^\circ$  and the total volume of treated solution was  $230 \text{ mL}$ . With this configuration a 46% of dye discoloration was obtained in 3 hours, and 62% letting the reaction for a total of 5 hours, as shown in Fig. 5d. This result indicates a good performance of the material under ambient conditions. The smaller degradation percentage observed in the solar reactor

compared to the experiments performed in the glass vials is due to the larger solution volume to film's surface ratio ( $V_{sol}/A_{film}$ ), as can be seen in Table 1. Although an improvement of the  $V_{sol}/A_{film}$  as well as of the whole experimental setup can be done, these results show a promising performance of the  $\beta\text{-Bi}_2\text{O}_3$  films for sustainable and environmentally friendly water treatment processes.

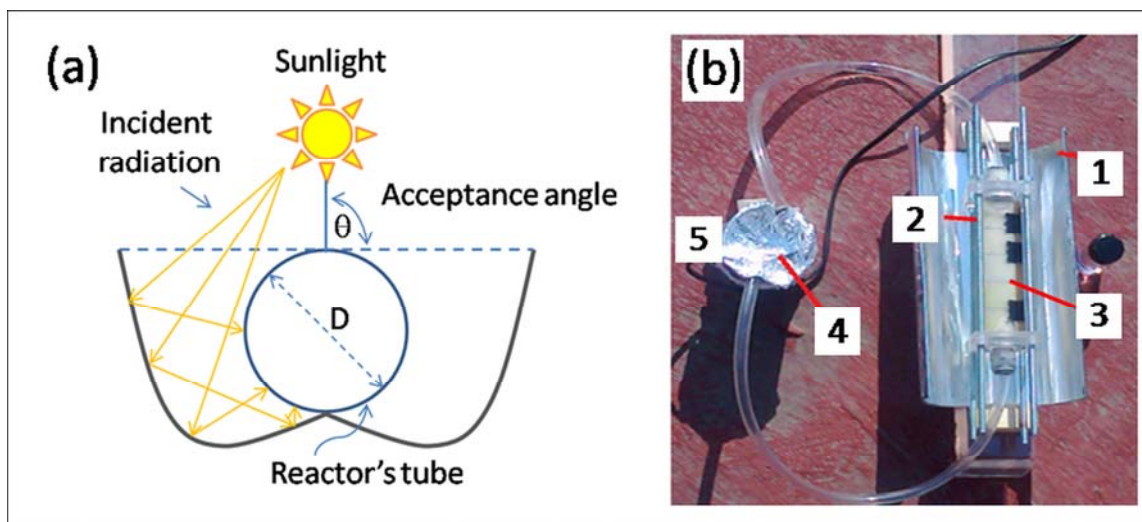


Fig. 6 (a) Diagram of the cross section of the CPC solar photocatalytic reactor, where the acceptance angle is  $\theta=90^\circ$  and the diameter is  $D=3.05$  cm. (b) Prototype of the solar reactor, where (1) is the collector, (2) is the reactor's tube, (3) are the thin films, (4) mixing tank, (5) water pump.

Table 1. Comparison of the photocatalytic performance of the  $\beta\text{-Bi}_2\text{O}_3$  films in the lab scale reactor and the CPC solar reactor, using Acid Blue 113 dye at  $10^{-6}\text{M}$ .

Type of reactor	Degradation %	Reaction time (min)	Solution volume (mL)	Area of photocatalyst ( $\text{cm}^2$ )	$V_{sol}/A_{film}$
CPC solar reactor	45.9	180	230	62.5	3.7
Glass vials	71.7	180	10	3.6	2.8

#### 4. Discussion

The spray pyrolysis technique allowed the growth of pure  $\beta$ -phase  $\text{Bi}_2\text{O}_3$  films at 450 °C that presented good photocatalytic activity in specific conditions. However, at the present stage it would be crude and misleading to compare the performance of  $\text{Bi}_2\text{O}_3$  films directly with  $\text{TiO}_2$ . More studies are needed to understand the processes involved with  $\beta$ - $\text{Bi}_2\text{O}_3$  photocatalyst. As it is often claimed, there are several factors that affect the photocatalytic reactions. Friedmann et al. classified these parameters into intrinsic –related directly with the semiconductor, such as crystallographic phase, crystallite size, impurities, etc.– and extrinsic –that involve the surrounding environment and conditions, including pH of the solution, structure of the pollutant, its initial concentration, light intensity, among others<sup>45</sup>; however, all of these parameters may affect in different way for different systems (defining a system as the couple of photocatalyst and pollutant). It is not evident that one specific parameter will give the same contribution in all systems. Certain correlations have been found in the well known  $\text{TiO}_2$ , but for the new photocatalysts such as  $\text{Bi}_2\text{O}_3$  these contributions may not be the same. In our work, the extrinsic parameters were evident in the case of the photocatalytic treatment of methyl orange because pH was decisive in the degradation percentage achieved by the films, as exposed in figure 4a. The other parameter that could affect is the light intensity in the case of the simulated sunlight, giving a higher discoloration percentage compared to the UV light while keeping the rest of the parameters constant (Fig. 4c). Here is worth to comment that for a photocatalyst in the thin film form the photoactive area is the one that is completely exposed on the surface. This means that even though the material is porous or rough, most of the photons cannot reach the deeper or more hidden parts of it and they will not contribute to the reactions. Roughness becomes important in the micrometer regime, while very low values of it do not play such a relevant role. In our  $\text{Bi}_2\text{O}_3$  films the surface is formed by compacted nanoplates, with an average roughness of 60 nm. This roughness increases the surface area compared to a smooth film, but their compact structure avoid the formation of pores. These morphological characteristics favored the photocatalytic reactions in our system ( $\beta$ - $\text{Bi}_2\text{O}_3$  – dye). On the other hand, diffusion of the species also control the process if the rate of adsorption of the dye molecules from fluid phase is slower than the rate of electron-hole reactions with the dye molecules<sup>46</sup>. In addition, when the surface coverage increases, there is a slowing down of the adsorption, as observed in Fig. 5c. For these reasons, in thin film photocatalysts the degradation percentage is not linearly proportional to the surface area, mainly because the internal structure remains in dark.



Dye discoloration tests provide a quick way to evaluate the photocatalytic activity of a semiconductor by the use of a UV-vis spectrophotometer; however they cannot be used as standards because the complete discoloration of the dye does not mean that the dye was degraded into innocuous compounds<sup>47</sup>. Then, additional tests, such as TOC, are required to verify the mineralization of the dye. This was reported by Vautier et al. for indigo carmine dye, which was mineralized by UV-irradiated TiO<sub>2</sub>, but the color also disappeared when only using visible light; however, its TOC remained intact<sup>48</sup>. The dyes used in the present work (methyl orange and acid blue 113) are very stable upon illumination, so the photolysis was very low (less than 10%, as it is shown in Fig. A of the Electronic Supplementary Information), but the TOC at the end of the photocatalytic test was lower than the initial value, indicating not only a discoloration of the dye but also a true degradation, 35% in the case of MO and 29% in the case of AB113. Other aspect that sometimes discourages the test with dyes is that their degradation mechanism is complicated and the real quantum efficiency of the photocatalytic reaction cannot be easily determined. Even so, other authors are devoted to find the degradation mechanisms of this kind of compounds and their degradation byproducts<sup>49, 50</sup>, which are important to know due to their environmental impact.

Beyond these technical issues, the degradation tests performed in this work demonstrated the effectiveness of  $\beta$ -Bi<sub>2</sub>O<sub>3</sub> films for the degradation of diluted dyes, which is still a global major concern. Moreover, this semiconductor can be used with visible light, making it suitable for solar applications as demonstrated with our small scale CPC solar reactor.

#### 4. Conclusions

Single  $\beta$ -phase Bi<sub>2</sub>O<sub>3</sub> films with good photocatalytic activity can be obtained by the spray pyrolysis technique at 450 °C. The energy band gap of 2.6 eV allows the absorption of visible light for the photocatalytic reactions and the platelet morphology is also helpful for this purpose. The activity of the films remains constant after 5 degradation cycles; however, if the solution contains HCl the chlorine reacts with the Bi<sub>2</sub>O<sub>3</sub> forming BiOCl, which indeed is also a photocatalyst. The good performance of the  $\beta$ -Bi<sub>2</sub>O<sub>3</sub> films under direct sunlight exposure is strong evidence that this material can be used for water treatment in solar reactors.

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