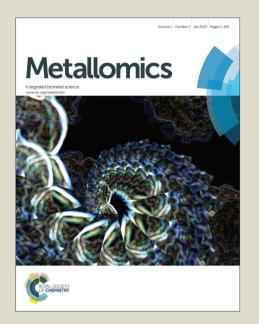
Metallomics

Accepted Manuscript



This is an *Accepted Manuscript*, which has been through the Royal Society of Chemistry peer review process and has been accepted for publication.

Accepted Manuscripts are published online shortly after acceptance, before technical editing, formatting and proof reading. Using this free service, authors can make their results available to the community, in citable form, before we publish the edited article. We will replace this Accepted Manuscript with the edited and formatted Advance Article as soon as it is available.

You can find more information about *Accepted Manuscripts* in the **Information for Authors**.

Please note that technical editing may introduce minor changes to the text and/or graphics, which may alter content. The journal's standard <u>Terms & Conditions</u> and the <u>Ethical guidelines</u> still apply. In no event shall the Royal Society of Chemistry be held responsible for any errors or omissions in this *Accepted Manuscript* or any consequences arising from the use of any information it contains.



Journal Name

RSCPublishing

ARTICLE

In vitro toxicological characterisation of three arsenic-containing hydrocarbons

Cite this: DOI: 10.1039/x0xx00000x

S. Meyer, a,b M. Matissek, c S. M. Müller, c M. S. Taleshi, d F. Ebert, b K. A. Francesconi d and T. Schwerdtle b

Received ooth January 2012, Accepted ooth January 2012

DOI: 10.1039/x0xx00000x

www.rsc.org/

Arsenic-containing hydrocarbons are one group of fat-soluble organic arsenic compounds (arsenolipids) in marine fish and other seafood. A risk assessment for arsenolipids is urgently needed, but has not been possible because of the total lack of toxicological data. In this study the cellular toxicity of three arsenic-containing hydrocarbons was investigated in cultured human bladder (UROtsa) and liver (HepG2) cells. Cytotoxicity of the arsenic-containing hydrocarbons was comparable to that for arsenite, which was applied as the toxic reference arsenical. Thereby, a large cellular accumulation of arsenic, as measured by ICP-MS/MS, was observed after incubation of both cell lines with the arsenolipids. Moreover, the toxic mode of action shown by the three arsenic-containing hydrocarbons seemed to differ from that observed for arsenite. Evidence suggests that the high cytotoxic potential of the lipophilic arsenicals results from a decrease in the cellular energy level. This first *in vitro* based risk assessment cannot exclude a risk to human health related to the presence of arsenolipids in seafood, and indicates the urgent need for further toxicity studies in experimental animals to fully assess this possible risk.

Introduction

In the environment arsenic occurs ubiquitously in many inorganic or organic species. Up to now more than 50 As species have been identified ¹.

For the general population food is the major source for arsenic. However, the arsenic content in terrestrial food (excluding rice) is quiet low, and inorganic arsenic (iAs) is the predominant arsenic form. iAs has been classified as a human carcinogen (group 1) by the International Agency for Research on Cancer (IARC). After chronic ingestion as well as inhalative exposure, iAs causes tumours of the lung, skin and bladder ²⁻³. To date the underlying molecular mechanisms of iAs-induced carcinogenicity are still not fully understood.

Toxicological evaluation of arsenobetaine indicates that this arsenic compound is non-toxic to humans. In humans after oral intake, arsenobetaine is excreted unchanged via urine ⁷⁻⁸.

In contrast to arsenobetaine, arsenolipids seem to be efficiently metabolised by humans. Thus, after ingestion of arsenolipids contained in cod liver oil by two volunteers, dimethylarsinic acid (DMA^V) was identified as major metabolite (up to 70%) ^{6, 9}. DMA^V represents also the major metabolite of iAs^{III}. DMA^V has been shown to exert genotoxicity in cultured mammalian cells (e.g. ¹⁰⁻¹²), to induce bladder cancer in rats ¹³ and has been classified by the IARC in 2012 as possibly carcinogenic to

In marine fish and other seafood the total arsenic content is in general much higher than that in terrestrial food. Furthermore, in seafood organic arsenic compounds are the major arsenicals, including the water-soluble arsenobetaine or arsenosugars as well as the lipid-soluble arsenolipids 4 . In general the fat fraction of livers and other organs of marine fish is rich in arsenic. Oils from these marine fish generally contain between 1 and 50 mg As/kg oil. Thereby the lipid-soluble arsenic compounds are around 10-30% of the total arsenic present in marine organisms 5 . In cod liver sampels up to 77% of the total arsenic (3.3 mg/kg dry mass) 6 and in tuna around 50% of the total arsenic (5.9 mg/kg dry weight) has been identified as lipid-extractable 4 .

^a Graduate School of Chemistry, University of Münster, Wilhelm-Klemm-Straße 10, 48149 Münster, Germany.

b Institute of Nutritional Science, University of Potsdam, Arthur-Scheunert-Allee 114-116, 14558 Nuthetal, Germany.

^c Institute of Food Chemistry, University of Münster, Corrensstraße 45, 48149 Münster, Germany.

^d Institute of Chemistry – Analytical Chemistry, University of Graz, Universitätsplatz 1, 8010 Graz, Austria.

2

3

4

5

6

7

8

9

10

11

12

13

14

15

16

17

18

19

20

21

22

23

24

25

26

27

28

29

30

31

32

33

34

35

36

37

38

39

40

41

42

43

44

45

46

47

48

49

50

51

52

53

54

55

56

Figure 1. Chemical structures and abbreviations of three arsenic-containing hydrocarbons investigated in this study.

humans (group 2B) 14. The toxicity of arsenolipids has not been investigated, and hence there are no toxicological data to assess the risks to human health related to the presence of arsenolipids in seafood 3.

Although the occurrence of arsenolipids in fish and other types of seafood was first reported in the 1960s 15, little was known about the structures of these compounds for many years afterwards. The first arsenolipid was identified in 1988 as an arsenosugar bound to a phospholipid 16. Through the improvement of analytical techniques more arsenolipids have been found in the last couple of years, especially in fish and algae. Two other groups of arsenolipids could be identified in fish oils, which were categorised into arsenic-containing hydrocarbons (AsHC) 17 and arsenic-containing long-chain fatty acids (AsFA) 18. Recently, further structures of arsenosugar-phospholipids (AsPL) have been reported in brown algae 19.

This study focuses on the cellular toxicity of three arseniccontaining hydrocarbons (AsHC 332, AsHC 360, and AsHC 444 (Fig. 1)). These arsenicals have been identified in fish oil 17, 20-21, but also in several fish meat samples, including tuna 4, cod 22-23 and herring 24, as well as in some brown algae 19. For the in vitro cellular system, immortalised human urothelium cells (UROtsa) were applied, since the bladder has been identified as target tissue for iAs and DMAV induced toxicity. For the second model system, liver cells (HepG2) were used, since after intestinal absorption the arsenicals are likely to be metabolised in the liver. In a first step the cytotoxic profiles of the arsenolipids were compared to the profile of the toxic reference compound iAs^{III}. Since all three arsenic-containing hydrocarbons exerted strong cytotoxic effects, further studies were carried out to identify the toxic mode of action of the respective arsenolipids. Endpoints included among others cellular bioavailability and distribution of the arsenicals, genotoxicity as well as effects on the cellular energy level.

Experimental

Material

Minimal essential medium Eagle (MEM), nonessential amino acids and the culture dishes were supplied by Biochrom (Berlin, Germany). Fetal calf serum (FCS), penicillin-streptomycin solutions and trypsin were products of PAA Laboratories GmbH (Pasching, Austria). Triton X-100 was purchased from Thermo Scientific (Rockford, USA), hydroxyapatite (high resolution) and potassium dihydrogen phosphate from Calbiochem (Bad Soden, Germany) and Giemsa stain, hydrogen peroxide solution (30%, suprapur), nitric acid (65% suprapur) and tetra-n-butylammonium bromide (TBAB) from Merck (Darmstadt, Germany). Sodium(meta)-arsenite (≥99% purity) and Alcian Blue were from Fluka Biochemika (Buchs, Germany). The cell-counting kit-8 (CCK-8®) was obtained from Dojindo molecular technologies (Munich, Germany) and the ICP-MS elemental standard (As, 1 mg/L) from SPECTEC (Erding, Germany). AMP and ADP-ribose as well as hexadecane (HC 332), octadecane (HC 360) and tetracosane (HC 444) were obtained from Sigma (Deisenhofen, Germany). ATP and ADP were products of Gerbu (Gaiberg, Germany). Sodium hydroxide and hydrochloric acid were purchased from Grüssing (Filsum, Germany), acetonitrile from VWR (Darmstadt, Germany), dipotassium hydrogen phosphate was purchased from Roth (Karlsruhe, Germany), and 23-gauge needles were obtained from Braun (Melsungen, Germany). All other chemicals were of p.a. grade and were from Merck (Darmstadt, Germany) or Fluka Chemie (Buchs, Germany). Prof. Dr. B. Epe (University of Mainz, Germany) kindly provided the Fpg protein. The urothelial cell line UROtsa was derived from a primary culture of a normal human urothelium through immortalisation with the SV-40 large T antigen. This cell line was kindly provided by Prof. M. Stýblo (University of North Carolina, USA). HepG2 cells were obtained from the European Collection of Cell Cultures (ECACC; number 85011430, Salisbury, UK).

Synthesis and preparation of arsenic-containing hydrocarbons for cytotoxicity studies

The arsenic-containing hydrocarbons were synthesised and purified in Graz; full details have been reported elsewhere ²⁵. In brief, iododimethylarsine was added to concentrated NaOH to form bis-(dimethylarsenic) oxide [(Me₂As)₂O], which was then heated with the appropriate 1-bromo-alkane. The product was extracted into chloroform, and crystallised from ethylacetate. The purity of the compounds was >99% as assessed by NMR spectroscopy and HPLC coupled with molecular (electrospray) and elemental (ICP-MS) mass spectrometry.

A portion (ca 15 mg) of the arsenolipid was transferred to a test tube and dissolved in ca 10 mL methanol. The arsenic purity of the compound was checked by HPLC/ICP-MS, and the precise arsenic concentration of this methanolic solution was determined by ICP-MS (performed on replicate portions of this solution that had been subjected to an acid digestion procedure). Replicates (10 - 15) of an appropriate aliquot of the methanolic solution of arsenolipid (each containing 250 µg As = or $3.33 \mu mol$ of compound) were then transferred to small

 Journal Name

screw-capped vials; methanol was allowed to evaporate and the vials were stored at 4 °C before use in the cytotoxicity studies.

Table 1. ICP-MS/MS parameters

Forward power	1550 W
Cool gas flow	15 L/min
Auxiliary gas flow	0.9 L/min
Nebulizer gas flow	1 L/min
Nebulizer type	MicroMist
Quadrupole 1	m/z 75
Reaction gas flow	O ₂ : 0.3 mL/min (purity 99.9999%)
Quadrupole 2	m/z 91
LOD*	3.0 ng/L

12.0 ng/L

* DIN 32645

LOQ*

Cell culture and incubation with arsenicals and alkanes

Cells were grown in culture dishes as monolayer using MEM containing FCS (10%, v/v), penicillin (100 U/mL) and streptomycin (100 µg/mL) for UROtsa cells. For HepG2 cells MEM containing FCS (10%, v/v), penicillin (100 U/mL) and streptomycin (100 µg/mL) was used, which was supplemented with non-essential amino acids (1%, v/v). The cultures were incubated at 37 °C with 5% CO₂ in the air with 100% humidity. For each experiment cells were seeded in a defined density (17000 cells/cm²). After 24 h, logarithmically growing cells were incubated with the respective compounds for 48 h.

Arsenical and alkane stock solutions (10 mM) were prepared in 100% EtOH and stored at 4 °C. For incubation the stock solution was diluted shortly before the experiment with EtOH. The EtOH concentration in all experiments was set to 0.5% for UROtsa cells and 1% for HepG2 cells. These concentrations did not induce any cytotoxic effects compared to untreated control cells (data not shown).

Cytotoxicity testing

The cytotoxicity of the arsenic-containing hydrocarbons and their respective non-arsenic-containing forms was elucidated by quantifying their effects on cell number, colony forming ability, lysosomal integrity as well as dehydrogenase activity.

Cell number and colony forming ability

Cell number and colony forming ability testing were performed as described before ¹¹. Briefly, after 48 h incubation with the arsenicals or alkanes, UROtsa or HepG2 cells were washed with phosphate buffered saline (PBS) and trypsinised. Subsequently, cell number and cell volume were measured by an automatic cell counter (Casy TTC®, Roche Innovatis AG). To evaluate the impact of the compounds on colony forming ability in UROtsa cells, after cell counting of each sample, 500 cells/dish were seeded again and after 6 – 7 days, colonies were fixed with EtOH, stained with Giemsa (25% in EtOH), counted and calculated as percent of control. Colony forming ability could not be carried out in HepG2 cells, since these cells cannot be fully singularised.

Lysosomal integrity (neutral red uptake assay)

The neutral red uptake assay is based on the ability of viable cells to incorporate and bind the supravital dye neutral red in their lysosomes ²⁶. UROtsa and HepG2 cells were cultured in 96-well culture plates and after 48 h incubation with the respective compound the medium was replaced by neutral red (3-amino-7-dimethylamino-2-methylphenazine hydrochloride) containing medium (UROtsa: 66.7 mg/L and HepG2 55.6 mg/L neutral red in MEM). After dye loading (3 h, 37 °C), cells were washed with PBS containing 0.5% formaldehyde and the incorporated dye was solubilised in 100 μL of acidified EtOH solution (50% EtOH, 1% acetic acid in PBS). Finally, the absorbance in each well was measured by a plate reader (Tecan Infinite M200[®] PRO, Tecan Deutschland GmbH, Crailsheim, Germany) at 540 nm.

Dehydrogenase activity (CCK-8 assay)

Cell viability was additionally assessed colorimetrically applying the cell-counting kit-8 (CCK-8)²⁷. Briefly, UROtsa and HepG2 cells were cultured in 96-well culture plates and after 48 h incubation with the respective compound, WST-8 (2-(2-methoxy-4-nitrophenyl)-3-(4-nitrophenyl)-5-(2,4-disulfophenyl)-2H-tetrazolium) solution was added and cells were incubated for 1 h. Absorbance in each well was determined by a plate reader (Tecan Infinite M200® PRO, Tecan Deutschland GmbH, Crailsheim, Germany) at 450 nm. By the activity of cellular dehydrogenases the water-soluble tetrazolium salt (WST-8) is reduced to a formazan dye. The amount of generated yellow-coloured formazan in the cells is directly

proportional to the number of viable cells per well.

LDH release

ARTICLE

1

2

3

4

5

6

7

8

9

10

11

12

13

14

15

16

17

18

19

20

21

22

23

24

25

26

27

28

29

30

LDH release was observed in both the cell lysates and the dosing media as described before ²⁸. Briefly, 40 µL of culture medium or 10 µL of cell lysates were mixed in a 96-well culture plate with reaction buffer (100 mM HEPES, 0.14 g/L NADH, 1.1 g/L sodium pyruvate, pH 7) to reach a total volume of 200 µL each. Absorbance was detected kinetically at 355 nm every 1.5 min at 37 °C. LDH release was calculated as percentage of untreated control cells.

Caspase-3 activity

Apoptosis was monitored by Caspase-3 activity as previously reported ²⁸. Briefly, after lysis of cells, the lysates were mixed with an equal amount of reaction buffer (50 mM PIPES, 10 mM EDTA, 0.5% CHAPS, 10 mM DTT, 80 µM DEVD-AFC) in a black 96-well plate. After 4 h of incubation at 37 °C, fluorescence of cleaved 7-amino-4-trifluoromethylcumarin (AFC) was monitored (ex. 400 nm, em. 505 nm). Caspase-3 activity was normalised by the respective protein contents, which were quantified by bicinchoninic acid (BCA) assay.

Cellular bioavailability of arsenic

Cellular bioavailability of the arsenic-containing hydrocarbons was studied after wet-ashing the cells (acid digestion) and measuring the arsenic content of the digest by inductively coupled plasma triple quadrupole mass spectrometry (ICP-MS/MS). Briefly, after 48 h incubation with the arsenicals, cells were trypsinised, collected by centrifugation and washed with ice-cold PBS. For calculation of cellular arsenic concentrations, volumes of cells and nuclei were measured by an automatic cell counter (Casy TTC®, Roche Innovatis AG) in each sample; these measurements are based on non-invasive (dye-free) electrical current exclusion with signal evaluation via pulse area analysis. Mean (± SD) volumes of non-incubated UROtsa and HepG2 cells were 1.92 (\pm 0.26) * 10⁻¹² L and 1.98 (\pm 0.29) * 10⁻¹² L, respectively. All arsenicals showed no significant effects on cell volumes at non-cytotoxic concentrations. After incubation with the ashing mixture (65% HNO₃/30% H₂O₂ (1/1, v/v)) at 95°C for at least 12 h, samples were diluted with 0.15 N HNO₃. Total arsenic was measured by ICP-MS/MS (Agilent 8800 ICP-QQQ, Agilent Technologies Deutschland GmbH, Böblingen, Germany) in the mass-shift mode using oxygen as reaction gas to eliminate interferences. Further ICP-MS/MS conditions are listed in Table 1.

Journal Name

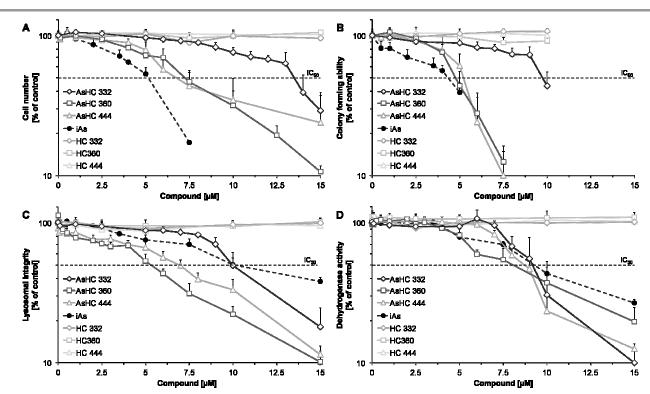


Figure 2. Cytotoxicity of three arsenic-containing hydrocarbons (AsHC 332, AsHC 360, AsHC 444), their respective non-arsenic-containing forms (HC 332, HC 360, HC 444) and arsenite in UROtsa cells after 48 h incubation. Data were normalized to the solvent control 0.5% EtOH (100 %), which exerted no significant cytotoxicity itself. Cytotoxicity was determined by impact on cell number (A), colony forming ability (B), lysosomal integrity as measured by neutral red uptake (C) and dehydrogenase activity as measured by the CCK-8 assay (D). Shown are mean values of at least three independent determinations +SD. Colony forming ability of control cells was app. 80%.

Cellular distribution of arsenic

To assess cellular arsenic distribution after 48 h incubation with the respective arsenic-containing hydrocarbons, cells were trypsinised, collected by centrifugation, washed with ice-cold PBS. Cell number and cell volume were determined in each sample as described above. After centrifugation the cell pellet was lysed by addition of bi-distilled water and sonicated (15 s,

100%, 0.8 cycles). Subsequently, the membrane-associated parts were separated from the cytosol (and nuclei plasma) by centrifugation (5 min, 23600 x g, 4 $^{\circ}$ C). ICP-MS/MS based analysis of total arsenic was carried out in the cytosol as well as in the wet-ashed membrane-containing fraction.

48 h incubation		Cell number	Colony forming ability	Lysosomal integrity	Dehydrogenase activity
AsHC	UROtsa	13.5	9.8	10.0	9.2
332	HepG2	17	-	17	52
AsHC	UROtsa	7.4	4.8	5.2	8.0
360	HepG2	8	-	21	30
AsHC	UROtsa	7.0	5.2	7.0	8.8
444	HepG2	21	-	40	63
Arsenite	UROtsa	5.2	4.3	10.1	9.2
	HepG2	17	-	44	38

Metallomics Accepted Manuscrip

ARTICLE

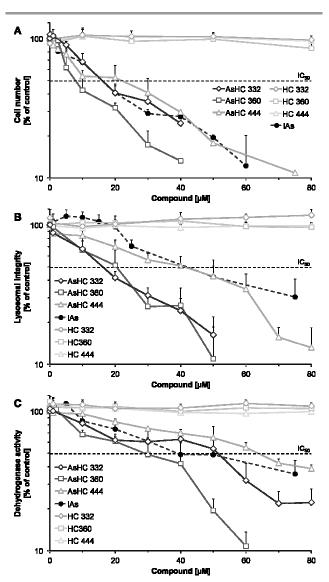


Figure 3. Cytotoxicity of three arsenic-containing hydrocarbons (AsHC 332, AsHC 360, AsHC 444), their respective non-arsenic-containing forms (HC 332, HC 360, HC 444) and arsenite in HepG2 cells after 48 h incubation. Data were normalized to the solvent control 1% EtOH (100%), which exerted no cytotoxicity itself. Cytotoxicity was determined by impact on cell number (A), lysosomal integrity as measured by neutral red uptake (B) and dehydrogenase activity as measured by the CCK-8 assay (C). Shown are mean values of at least three independent determinations +SD.

Alkaline unwinding

DNA strand breaks were quantified by the alkaline unwinding technique ²⁷. Briefly, 150000 HepG2 cells were seeded and after 24 h cells were incubated with the arsenicals for 48 h. Subsequently, the medium was removed, cells were washed with PBS and an alkaline solution (0.03 M NaOH, 0.02 M Na₂HPO₄, 0.9 M NaCl) was added. After neutralisation and sonication, separation of single- and double-stranded DNA was performed on 0.5 mL hydroxyapatite columns at 60 °C. Singleand double-stranded DNA were eluted with 1.5 mL of 0.15 M and 0.35 M potassium phosphate buffer, respectively. The DNA content of both fractions was determined by adding Hoechst 33258 dye to a final concentration of 0.77 µM to 1 mL of each sample and measuring the fluorescence with a microtiter fluorescence reader (Tecan Infinite M200® PRO, Tecan Deutschland GmbH, Crailsheim, Germany) at an excitation wavelength of 360 nm and an emission wavelength of 455 nm. DNA strand breaks were quantified by calibration with X-rays as described previously 29.

Micronuclei formation

Micronuclei formation was studied as previously reported ³⁰. Since our earlier studies indicated that several arsenicals interact with actin and/or the effect of cytochalasin B, we omitted the application of cytochalasin B ¹⁰. To ensure mitosis, cell proliferation was monitored by means of cell number quantification. An incubation time of 48 h was chosen, which is in these cell lines in accordance with around 2 cell cycles of untreated control cells. Briefly, UROtsa and HepG2 cells were seeded in 12-well plates on Alcian blue-coated glass coverslips. Cells were incubated with the respective arsenicals, fixed with an ice-cold fixation solution (90% methanol/10% PBS, -20 °C) for 10 min, stained with acridine orange (125 mg/L in PBS), and finally evaluated by fluorescence microscopy after coding of slides. Per coverslip, at least 1000 mononucleated cells were counted and categorised in mononucleated, binucleated, and multinucleated cells as well as cells with and without micronuclei.

Level of energy related nucleotides

The cellular levels of energy related nucleotides were quantified as described previously 31 with slight modifications. Shortly, incubated cells were trypsinised, respective cell volumes were determined in each sample and cells were pelletised. Cell pellets were lysed by 300 µL of 0.5 M KOH, pulled 10 times through a 23-gauge needle and the extract was neutralised by adding 60 µL phosphoric acid (10%). After centrifugation at 20630 x g (30 min, 4 °C), for separation of the nucleotides, 20 µL of the supernatant were injected into a LC system, consisting of an autosampler, a binary pump and a photodiode array detector (DAD) (Agilent 1200 series, Agilent Technologies Deutschland GmbH, Böblingen, Germany). The total adenosine nucleotides (TAN) and energy charge values (AEC) were calculated according to the following formulas for each sample:

$$TAN = [ATP] + [ADP] + [AMP]$$

$$AEC = \frac{[ATP] + 0.5 \cdot [ADP]}{TAN}$$

Statistical analysis

All experiments were carried out at least three times, each time on a different day. As indicated in the respective figure captions, from the raw data the mean standard deviation (SD) was calculated and a statistical analysis was performed by using the ANOVA-OneWay-test. Significance levels are p < 0.05, **p < 0.01 and ***p < 0.001.

Page 7 of 12 Metallomics

Journal Name ARTICLE

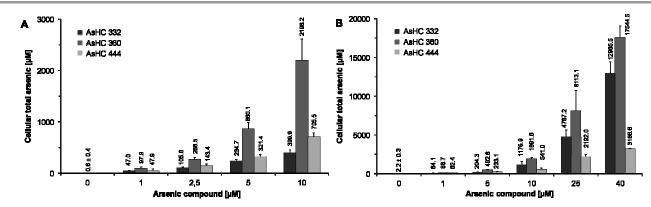


Figure 4. Cellular bioavailability of three arsenic-containing hydrocarbons (AsHC 332, AsHC 360, AsHC 444) in UROtsa (A) and HepG2 (B) cells after 48 h incubation. Cellular arsenic concentration was determined by ICP-MS/MS. The figure represents mean values of at least three independent determinations +SD.

Results

Cytotoxicity

To get a first idea about the *in vitro* toxicological profile of the respective arsenolipids, several cytotoxicity markers were examined after 48 h incubation with the three arsenic-containing hydrocarbons (AsHC 332, AsHC 360 and AsHC 444). In parallel arsenite (iAs^{III}) was studied as the toxic reference arsenical. Thus, effects on cell number and colony forming ability as well as on lysosomal integrity and dehydrogenase activity were quantified.

In UROtsa cells, cytotoxicity caused by the arsenic-containing hydrocarbons was in the same concentration range as that shown by iAs^{III} (Fig. 2). The two longer C-chain arsenic-containing hydrocarbons (AsHC 360 and AsHC 444) were comparable in their toxic behaviour, whereas the smaller AsHC 332 was less cytotoxic in all investigated endpoints than AsHC 360 and AsHC 444. iAs^{III} exerted slightly stronger effects than did the arsenolipids regarding the endpoints cell number and colony forming ability, but showed weaker effects for the viability markers lysosomal integrity and dehydrogenase activity.

HepG2 cells were less sensitive towards all applied arsenicals than UROtsa cells (Fig. 3); IC_{50} values of the arsenic-containing hydrocarbons were around 5-fold higher as compared to the respective IC_{50} values in UROtsa cells (Tab. 2). Toxicity order of the arsenolipids was similar in UROtsa and HepG2 cells, with AsHC 360 being the most and

Table 3. Total cellular arsenic after iAs incubation for 48 h in UROtsa and HepG2 cells $[\mu M]$

Incubation concentration [µM]	1	5	10	25
UROtsa:	10.3±1.7	53.0±10.9	-	-
HepG2:	11.5±6.1	34.9±10.9	57.7±10.9	547±60

AsHC 332 the less cytotoxic arsenolipid applied. Regarding all endpoints, iAs^{III} was similar or less cytotoxic as compared to the respective arsenic-containing hydrocarbons.

To verify that the toxic effects are associated with the dimethylarsinoyl group of the arsenic-containing hydrocarbons the corresponding hydrocarbons (HC 332, HC 360, HC 444) were studied in parallel. Cytotoxic effects were not observed in the applied respective concentrations of the alkanes, indicating that the dimethylarsinoyl group of the molecules is likely responsible for the cytotoxicity.

Cellular bioavailability and distribution

The cellular arsenic concentrations were determined after 48 h incubation with the three arsenic-containing hydrocarbons in both UROtsa and HepG2 cells. All three arsenic-containing hydrocarbons were strongly bioavailable in both UROtsa and HepG2 cells (Fig. 4). In UROtsa cells after incubation with a low sub-cytotoxic concentration (1 μ M) of the arsenolipids, cellular arsenic accumulated by a factor of up to 100 in relation to the incubated extracellular arsenic concentration. The most cytotoxic arsenolipid AsHC 360 exerted the highest cellular bioavailability, reaching a cellular accumulation factor of 200. Similar accumulation behaviour was observed in HepG2 cells. For AsHC 360 even an accumulation of more than 400-fold was observable.

Although iAs^{III} was cellular bioavailable as well, cellular accumulation was much lower in both cell lines. Thus, after incubation with iAs^{III}, cellular arsenic content was merely up to 10-fold higher than the respective extracellular incubation

Table 4. Cellular distribution of arsenic after incubation with 5 μM for 48 hof the arsenic-containing hydrocarbons

	Membrane- associated fraction	Cytosol- fraction	Total	% of total As*
AsHC 332:	56.1%	43.9%	165.7 μΜ	81.1%
AsHC 360:	36.1%	63.9%	637.9 μM	129.5%
AsHC 444:	30.4%	69.6%	213.2 μM	91.4%
*cellular bioa	ıvailability			

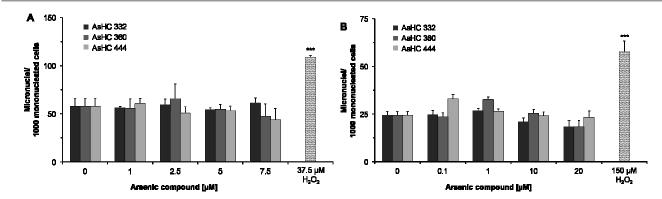


Figure 5. Formation of micronuclei in UROtsa (A) and HepG2 (B) cells after 48 h incubation with three arsenic-containing hydrocarbons (AsHC 332, AsHC 360, AsHC 444). Displayed are mean values of at least three independent determinations +SD. ***p < 0.001.

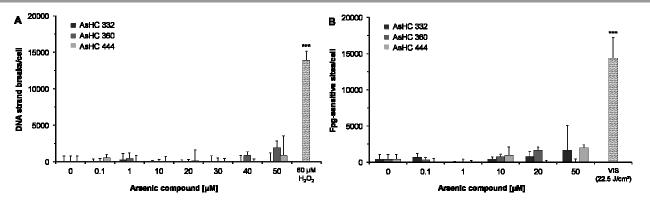


Figure 6. Induction of DNA strand breaks (A) and Fpg-sensitive sites (B) in HepG2 cells after incubation with three arsenic-containing hydrocarbons (AsHC 332, AsHC 360, AsHC 444) for 48 h quantified by the alkaline unwinding technique. Hydrogenperoxide (5 min, 60 μ M) and radiation with visible light (22.5 J/cm²) in combination with the photosensitiser RO-19-8022 (50 nM) served as positive controls. Displayed are mean values of at least three independent determinations +SD. ***p < 0.001.

concentration (Tab. 3).

Distribution analysis demonstrated that around 30-56% of the total cellular arsenic is localised in membrane-associated parts (Tab. 4) after incubation with the arsenolipids. When taking into account that the membrane fraction captures only a small portion of the cell, it becomes clear that the absolute arsenic concentration in the membrane fraction is much higher than in the cytosol.

Formation of micronuclei and induction of DNA damage

Genotoxicity of three arsenic-containing hydrocarbons was assessed at both the chromosomal and the DNA level, by quantifying the formation of micronuclei, DNA strand breaks and oxidative base modifications.

In both cell lines the arsenolipids increased neither the number of micronuclei (Fig. 5a,b), nor the amount of bi- or multinucleated cells (data not shown). Additionally the arsenolipids did not induce DNA strand breaks (Fig. 6a) and Fpg-sensitive sites (Fig. 6b) in HepG2 cells.

Lactate dehydrogenase (LDH) release and caspase-3 activity

LDH release was monitored in HepG2 cells to assess effects of the arsenicals on cell membrane integrity. In comparison to other studied endpoints, LDH release was the most insensitive viability marker, showing no significant effects up to 10 μM of the respective arsenolipids. After incubation with 40 μM of the arsenicals, strong LDH release was observable for all three arsenic-containing hydrocarbons (Fig. 7a). This indicates that in the high cytotoxic concentration range, the arsenolipids also disturb cell membrane integrity.

Apoptosis was monitored via caspase-3 activity in HepG2 cells. AsHC 360 did not show any significant caspase-3 activity up to an incubation concentration of 40 μ M (Fig. 7b). AsHC 332 and AsHC 444 exerted a significant increase of capase-3 activity only after incubation with 40 μ M. These data strongly suggest that the respective arsenolipids do not cause an apoptotic cell death in the applied concentration range. This indication is supported by the fact that no apoptotic bodies were visible in the microscopic slides prepared for the micronuclei test after incubation with the arsenic-containing hydrocarbons (data not shown).

Level of energy related nucleotides

The observed ATP and total adenosine nucleotide (TAN) levels of the investigated HepG2 cells are in the range of published data. In tumour cells the ATP concentration is

Page 9 of 12 Metallomics

 Journal Name ARTICLE

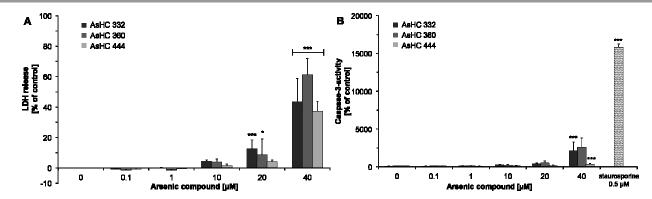


Figure 7. LDH release (A) and caspase-3 activity (B) after incubation with three arsenic-containing hydrocarbons (AsHC 332, AsHC 360, AsHC 444) for 48 h in HepG2 cells. Staurosporine (0.5 μ M) served as positive control. Shown are mean values of at least three independent determinations +SD. *p < 0.05 and ***p < 0.001.

 $3134\pm2135~\mu M^{32}$ and in the present work the ATP level of HepG2 cells were $1620\pm50~\mu M$ (Fig. 8). This represents around 90% of the TAN. This result is comparable with measured ATP levels in A549 cells (1595 $\pm291~\mu M$, 90% of the TAN) in a former study 31 . The energy charge value (AEC) of untreated HepG2 cells was calculated with 0.95 (data not shown), which indicates healthy cells, where the AEC is normally between 0.7 and 1.0 33 .

Incubation with beginning cytotoxic effects of the respective arsenolipids caused a significant decrease of the TAN levels (Fig. 9). The cellular ATP levels were strongly affected, leaving only 44.9%, 31.2% and 53.3% of cellular ATP in the case of incubation with 20 μM of AsHC 332 or 10 μM of AsHC 360 or AsHC 444, respectively (Fig. 8). Corresponding to the decrease of ATP levels, the cellular levels AMP, ADP and ADP-ribose increased (Fig. 9). Compared to control cells the amount of ADP and ADP-ribose is up to 2-fold higher and the AMP concentration even around 3 to 6-fold higher in cells incubated with the arsenolipids. This indicates that ATP recycling as well as ATP synthesis is disturbed.

Changes in the distribution of the levels of the adenosine nucleotides are also evident for the AEC values, which were

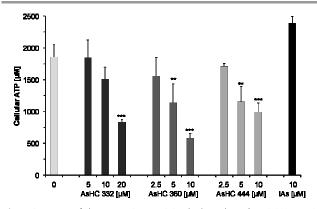


Figure 8. Impact of three arsenic-containing hydrocarbons (AsHC 332, AsHC 360, AsHC 444) and arsenite on the cellular ATP level in HepG2 cells after 48 h incubation. Shown are mean values of at least three independent determinations +SD. **p < 0.01 and ***p < 0.001.

decreased by the arsenolipids. Thus, for $20 \mu M$ of AsHC 332 the AEC value was 0.82, and for $10 \mu M$ of AsHC 360 and AsHC 444 the values were 0.72 and 0.75 respectively (data not shown)

In contrast to the arsenolipids, iAs^{III} did not significantly decrease the cellular ATP level up to an incubation concentration of 10 μ M. These data suggest that iAs^{III} causes cell death *in vitro* via a different mode of action than do the arsenic-containing hydrocarbons.

Discussion

This study investigated for the first time the cellular toxicity and bioavailability of three food-relevant arsenic-containing hydrocarbons. Arsenite was studied in parallel as the toxic arsenic reference species, to facilitate a first *in vitro* based risk assessment for this class of arsenolipids.

In both cultured human bladder (UROtsa) and liver (HepG2) cells, the three arsenic-containing hydrocarbons caused significant cytotoxicity, which occurred in the same concentration range as that observed after incubation with iAs^{III}. In comparison to other seafood-relevant arsenicals *in vitro*, the arsenolipids are at least 600-fold more toxic than a glycerol arsenosugar ³⁴⁻³⁵, and about 20 to 25-fold more toxic than their major metabolite DMA^{V 30}.

This high cellular toxicity of the arsenolipids is likely to result at least partly from their cellular bioavailability. The highly lipophilic arsenic-containing hydrocarbons seem to be able to easily pass the cell membranes, presumably by passive diffusion, enter the cell and accumulate inside the cell especially in lipophilic membranes. This assumption was supported by the cellular distribution studies, which indicated a strong affinity of the arsenolipids to the membrane-associated cellular fraction.

From a structural point of view, the arsenic-containing hydrocarbons are amphiphilic, consisting of a hydrophobic hydrocarbon tail and a hydrophilic dimethylarsinoyl head group. Therefore, these molecules might interact with biomembranes and attach to these membranes comparable to

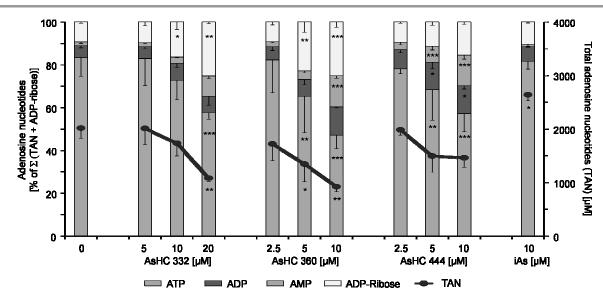


Figure 9. Impact of three arsenic-containing hydrocarbons (AsHC 332, AsHC 360, AsHC 444) and arsenite on adenosine nucleotides (ATP, ADP, AMP, ADP-risbose and total adenosine nucleotides (TAN)) in HepG2 cells after 48 h incubation. The nucleotides were measured with HPLC/DAD. Shown are mean values of at least three independent determinations +SD. *p < 0.05, **p < 0.01 and ***p < 0.001.

phospholipids. In this context it has already been discussed that AsPL might be used in membrane chemistry by algae ^{19, 36}. Furthermore, recently it has been demonstrated that phytoplankton can use non-phosphorus lipids to form membrane lipids ³⁷. In the case of phosphorus scarcity in oceanic waters, nitrogen or sulfur instead of phosphorus was used to build-up membrane lipids by the investigated algae. Arsenic is a member of the same group in the periodic table as phosphorus, and hence it is similar in some of its physical and chemical characteristics. Possibly, arsenic may also be used to form membrane lipids by algae living under phosphorus-deficient conditions, especially when non-selective enzymes are involved.

To elucidate the toxic modes of action of the arsenic-containing hydrocarbons, their genotoxic potential was investigated. In previous studies an increase in micronuclei induction by iAs^{III} was observed (e.g. 10, 38-39). Nevertheless, the three arseniccontaining hydrocarbons did not increase the number of micronuclei in both UROtsa and HepG2 cells. Likewise, the frequency of bi- and multinucleated cells was not affected by the arsenolipids. Whereas iAsIII has been shown before to generate DNA lesions, including Fpg-sensitive sites (e.g. 40-41), the arsenolipids failed to increase the cellular amount of DNA strand breaks and Fpg-sensitive sites. These data suggest that the observed cellular toxicity of the arsenic-containing hydrocarbons does not result from a genotoxic mode of action. Further studies also excluded the possibility that cell death was triggered by apoptosis or loss of cell membrane integrity. These endpoints were affected by the arsenic-containing hydrocarbons only at highly cytotoxic concentrations.

Lysosomal integrity has been the most sensitive viability marker investigated after arsenolipid exposure in this study. It is well known that lysosomal integrity strongly depends on the cellular ATP level. Thus, ATP is necessary to keep the membrane potential of these organelles, which means they have a pH value of approximately $4.5-5^{26}$. Accordingly, all three applied arsenic-containing hydrocarbons caused a concentration dependent decrease in the cellular amount of ATP. Beside its role as an energy carrier, ATP is also very important in signalling pathways and in development and regeneration processes ⁴². Consequently, a low ATP level can cause a number of cellular disorders, which are likely to contribute to the observed arsenolipid induced cell death in the present study. A possible underlying mechanism for the decrease in cellular ATP levels is that the arsenolipids attack mitochondrial membranes, thereby disturbing mitochondrial function and cellular ATP production. This hypothesis will be tested in our future studies.

Conclusion

Whereas arsenobetaine and arsenosugars have been shown to exert no or only very low toxicity in cultured cells $^{7-8, 30, 35}$, the three arsenic-containing hydrocarbons showed a strong cytotoxic potential in the low μ M concentration range, which is comparable to the cytotoxic potential found for arsenite. However, the toxic modes of action seem to differ between arsenite and the arsenic-containing hydrocarbons highlighting the need for further *in vitro* studies to understand the toxic mode of action of these potentially highly toxic class of organic arsenicals

To finally assess the risk to human health, toxicity studies in experimental animals should urgently be carried out.

Acknowledgment

2

3

4

5

6

7

8

9

10

11

12 13

14

15

16

17

18

19

20

21

22

23

24

25

26 27

28 29

30

31

32

33

34

35

36

37

38

39

40

41

42

43

44

45

46

47

48

49

50

51

52

53

54

55

56

Journal Name

This work was supported by the DFG grant number SCHW903/4-1, the Austrian Science Fund (FWF), project number 1550-N17, and the Graduate School of Chemistry (WWU Münster, Germany).

Reference

- K. A. Francesconi, Arsenic species in seafood: Origin and human health implications. *Pure Appl. Chem.* 2010, 85, 373-381.
- K. Straif, L. Benbrahim-Tallaa, R. Baan, Y. Grosse, B. Secretan, F. El Ghissassi, V. Bouvard, N. Guha, C. Freeman, L. Galichet, V. Cogliano, A review of human carcinogens: metals, arsenic, dusts, and fibres. *Lancet Oncol* 2009, 10. 453-4.
- EFSA, EFSA Panel on Contaminants in the Food Chain (CONTAM);
 Scientific Opinion on Arsenic in Food. EFSA Journal 2009, 7. 1351-1355.
- M. S. Taleshi, J. S. Edmonds, W. Goessler, M. J. Ruiz-Chancho, G. Raber, K. B. Jensen, K. A. Francesconi, Arsenic-containing lipids are natural constituents of sashimi tuna. *Environ Sci Technol* 2010, 44. 1478-1483, DOI: 10.1021/es9030358.
- V. Sele, J. J. Sloth, A. K. Lundebye, E. H. Larsen, M. H. G. Berntssen, H. Amlund, Arsenolipids in marine oils and fats: A review of occurrence, chemistry and future research needs. *Food Chemistry* 2012, *133*, 618-630, DOI: DOI 10.1016/j.foodchem.2012.02.004.
- E. Schmeisser, W. Goessler, K. A. Francesconi, Human metabolism of arsenolipids present in cod liver. *Anal Bioanal Chem* 2006, 385. 367-376, DOI: 10.1007/s00216-006-0401-x.
- T. Kaise, S. Watanabe, K. Itoh, The Acute Toxicity of Arsenobetaine. Chemosphere 1985, 14. 1327-1332, DOI: Doi 10.1016/0045-6535(85)90153-5.
- C. Newcombe, A. Raab, P. N. Williams, C. Deacon, P. I. Haris, A. A. Meharg, J. Feldmann, Accumulation or production of arsenobetaine in humans? *J Environ Monit* 2010, *12*. 832-7, DOI: 10.1039/b921588c.
- E. Schmeisser, A. Rumpler, M. Kollroser, G. Rechberger, W. Goessler, K. A. Francesconi, Arsenic fatty acids are human urinary metabolites of arsenolipids present in cod liver. *Angew Chem Int Ed Engl* 2006, 45. 150-154, DOI: 10.1002/anie.200502706.
- M. Bartel, F. Ebert, L. Leffers, U. Karst, T. Schwerdtle, Toxicological Characterization of the Inorganic and Organic Arsenic Metabolite Thio-DMA in Cultured Human Lung Cells. *J Toxicol* 2011, 2011. 1-9, DOI: 10.1155/2011/373141.
- F. Ebert, A. Weiss, M. Bultemeyer, I. Hamann, A. Hartwig, T. Schwerdtle, Arsenicals affect base excision repair by several mechanisms. *Mutat Res* 2011, 715. 32-41, DOI: S0027-5107(11)00180-1 [pii] 10.1016/j.mrfmmm.2011.07.004.
- M. J. Mass, A. Tennant, B. C. Roop, W. R. Cullen, M. Styblo, D. J. Thomas, A. D. Kligerman, Methylated trivalent arsenic species are genotoxic. *Chem Res Toxicol* 2001, *14*. 355-61.
- 13. US-EPA, Advisory on EPAs assessment of carcinogenic effects of organic and inorganic arsenic. *U. S. E. P. A. S. A. Board*, Washington DC, USA, EPA-SAB-07-008 2007.
- 14. IARC, A review of human carcinogens. Part C: Arsenic, metals, fibres, and dusts. *IARC Monographs* 2012. 196-211.
- G. Lunde, Analysis of arsenic in marine oils by neutron activation. Evidence of arseno organic compounds. *J Am Oil Chem Soc* 1968, 45. 331-332.
- 16. M. Morita, Y. Shibata, Isolation and identification of arseno-lipid from a brown alga. *Chemosphere* 1988, *17*. 1147-1152.

Metallomics Accepted Manuscrip

38

39

40

41

42

43

44

45

46

47

48

49

50

51 52

53

54

55

56

57

58

59 60

ARTICLE Journal Name

- 17. M. S. Taleshi, K. B. Jensen, G. Raber, J. S. Edmonds, H. A. Gunnlaugsdottir K Francesconi, Arsenic-containing hydrocarbons: natural compounds in oil from the fish capelin, Mallotus villosus. Chem Commun 2008. 4706-4707, DOI: 10.1039/b808049f.
- 18. A. Rumpler, J. S. Edmonds, M. Katsu, K. B. Jensen, W. Goessler, G. Raber, H. Gunnlaugsdottir, K. A. Francesconi, Arsenic-containing long-chain fatty acids in cod-liver oil: a result of biosynthetic infidelity? Angew Chem Int Ed Engl 2008, 47. 2665-2667, DOI: 10.1002/anie.200705405.
- 19. S. Garcia-Salgado, G. Raber, R. Raml, C. Magnes, K. A. Francesconi, Arsenosugar phospholipids and arsenic hydrocarbons in two species of brown macroalgae. Environ. Chem. 2012, 9. 63-66.
- 20. K. O. Amayo, A. Petursdottir, C. Newcombe, H. Gunnlaugsdottir, A. Raab, E. M. Krupp, J. Feldmann, Identification and quantification of arsenolipids using reversed-phase HPLC coupled simultaneously to high-resolution ICP-MS and high-resolution electrospray MS without species-specific standards. Anal Chem 2011, 83. 3589-3595, DOI: 10.1021/ac2005873.
- 21. K. O. Amayo, A. Raab, E. M. Krupp, J. Feldmann, Novel identification of lipid-soluble arsenic compounds using chemical derivatisations in conjunction with RP-HPLC-ICPMS/ESMS. Anal Chem 2013. DOI: 10.1021/ac4020935.
- 22. U. Arroyo-Abad, S. Lischka, C. Piechotta, J. Mattusch, T. Reemtsma, Determination and identification of hydrophilic and hydrophobic arsenic species in methanol extract of fresh cod liver by RP-HPLC with simultaneous ICP-MS and ESI-Q-TOF-MS detection. Food Chem 2013, 141. 3093-102, DOI: 10.1016/j.foodchem.2013.05.152 S0308-8146(13)00760-7 [pii].
- 23. U. Arroyo-Abad, J. Mattusch, S. Mothes, M. Moder, R. Wennrich, M. P. Elizalde-Gonzalez, F. M. Matysik, Detection of arseniccontaining hydrocarbons in canned cod liver tissue. Talanta 2010, 82. 38-43, DOI: S0039-9140(10)00248-1 [pii] 10.1016/j.talanta.2010.03.054.
- 24. S. Lischka, U. Arroyo-Abad, J. Mattusch, A. Kuhn, C. Piechotta, The high diversity of arsenolipids in herring fillet (Clupea harengus). Talanta 2013, 110. 144-52, DOI: 10.1016/j.talanta.2013.02.051 S0039-9140(13)00125-2 [pii].
- 25. M. S. Taleshi, R. K. Seidler-Egdal, K. B. Jensen, T. Schwerdtle, K. A. Francesconi, Synthesis and Characterisation of Arsenolipids: Naturally Occurring Arsenic Compounds in Fish and Algae. Organometallics in press.
- 26. G. Repetto, A. del Peso, J. L. Zurita, Neutral red uptake assay for the estimation of cell viability/cytotoxicity. Nat Protoc 2008, 3. 1125-31, DOI: nprot.2008.75 [pii] 10.1038/nprot.2008.75.
- 27. J. Bornhorst, F. Ebert, A. Hartwig, B. Michalke, T. Schwerdtle, Manganese inhibits poly(ADP-ribosyl)ation in human cells: a possible mechanism behind manganese-induced toxicity? J Environ Monit 2010, 12. 2062-9, DOI: 10.1039/c0em00252f.
- 28. F. Ebert, L. Leffers, T. Weber, S. Berndt, A. Mangerich, S. Beneke, A. Burkle, T. Schwerdtle, Toxicological properties of the thiolated inorganic arsenic and arsenosugar metabolite thio-dimethylarsinic acid in human bladder cells. J Trace Elem Med Biol 2013. DOI: S0946-672X(13)00094-1 [pii] 10.1016/j.jtemb.2013.06.004.

- 29. A. Hartwig, U. D. Groblinghoff, D. Beyersmann, A. T. Natarajan, R. Filon, L. H. Mullenders, Interaction of arsenic(III) with nucleotide excision repair in UV-irradiated human fibroblasts. Carcinogenesis 1997, 18. 399-405.
- 30. L. Leffers, F. Ebert, M. S. Taleshi, K. A. Francesconi, T. Schwerdtle, In vitro toxicological characterization of two arsenosugars and their metabolites. Mol Nutr Food Res 2013. 10.1002/mnfr.201200821.
- 31. J. Bornhorst, F. Ebert, H. Lohren, H. U. Humpf, U. Karst, T. Schwerdtle, Effects of manganese and arsenic species on the level of energy related nucleotides in human cells. Metallomics 2012, 4. 297-306, DOI: 10.1039/c2mt00164k.
- 32. T. W. Traut, Physiological concentrations of purines and pyrimidines. Mol Cell Biochem 1994, 140. 1-22.
- 33. M. S. Yang, L. C. Yu, R. C. Gupta, Analysis of changes in energy and redox states in HepG2 hepatoma and C6 glioma cells upon exposure to cadmium. Toxicology 2004, 201. 105-13, DOI: 10.1016/j.tox.2004.04.007.
- 34. P. Andrewes, D. M. Demarini, K. Funasaka, K. Wallace, V. W. Lai, H. Sun, W. R. Cullen, K. T. Kitchin, Do arsenosugars pose a risk to human health? The comparative toxicities of a trivalent and pentavalent arsenosugar. Environ Sci Technol 2004, 38. 4140-8.
- 35. L. Leffers, M. Unterberg, M. Bartel, C. Hoppe, I. Pieper, J. Stertmann, F. Ebert, H. U. Humpf, T. Schwerdtle, In vitro toxicological characterisation of the S-containing arsenic metabolites thio-dimethylarsinic acid and dimethylarsinic glutathione. Toxicology 2013. DOI: S0300-483X(13)00012-7 [pii] 10.1016/j.tox.2013.01.007.
- 36. R. V. Cooney, R. O. Mumma, A. A. Benson, Arsoniumphospholipid in algae. Proc Natl Acad Sci USA 1978, 75. 4262-4.
- 37. B. A. Van Mooy, H. F. Fredricks, B. E. Pedler, S. T. Dyhrman, D. M. Karl, M. Koblizek, M. W. Lomas, T. J. Mincer, L. R. Moore, T. Moutin, M. S. Rappe, E. A. Webb, Phytoplankton in the ocean use non-phosphorus lipids in response to phosphorus scarcity. Nature 2009, 458. 69-72, DOI: nature07659 [pii] 10.1038/nature07659.
- 38. T. W. Gebel, Genotoxicity of arsenical compounds. Int J Hyg Environ Health 2001, 203. 249-62, DOI: S1438-4639(04)70036-X [pii] 10.1078/S1438-4639(04)70036-X.
- 39. T. S. Wang, H. Huang, Active oxygen species are involved in the induction of micronuclei by arsenite in XRS-5 cells. Mutagenesis 1994, 9. 253-7.
- 40. T. Schwerdtle, I. Walter, I. Mackiw, A. Hartwig, Induction of oxidative DNA damage by arsenite and its trivalent and pentavalent methylated metabolites in cultured human cells and isolated DNA. Carcinogenesis 2003, 24. 967-74.
- 41. T. S. Wang, T. Y. Hsu, C. H. Chung, A. S. Wang, D. T. Bau, K. Y. Jan, Arsenite induces oxidative DNA adducts and DNA-protein cross-links in mammalian cells. Free Radic Biol Med 2001, 31. 321-30, DOI: S0891584901005810 [pii].
- 42. G. Burnstock, Historical review: ATP as a neurotransmitter. Trends sciences 2006, 27. pharmacological 10.1016/j.tips.2006.01.005.