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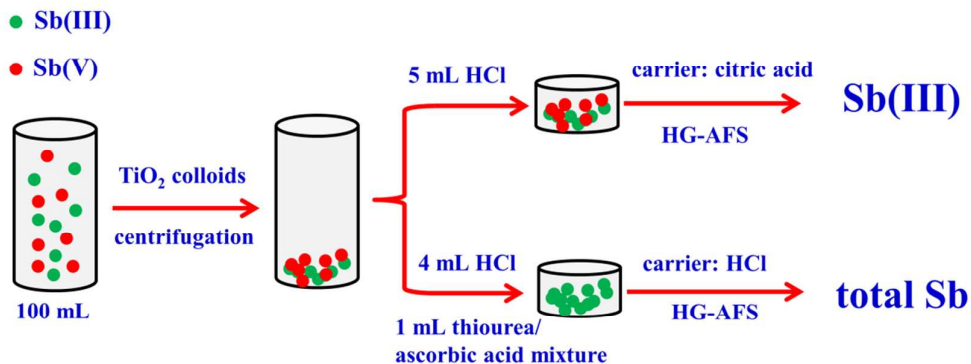
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# Speciation Analysis of Antimony in Water Samples via Combined Nano-Sized TiO<sub>2</sub> Colloid Preconcentration and AFS Analysis

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A method for the speciation of Sb was developed using nano-sized TiO<sub>2</sub> colloids preconcentration and HG-AFS as a determination means.



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1                   **Speciation Analysis of Antimony in Water Samples via**  
2                   **Combined Nano-Sized TiO<sub>2</sub> Colloid Preconcentration and AFS**  
3                   **Analysis**

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8                   **Abstract:** A method for the speciation of Sb(III) and Sb(V) in water samples was  
9                   developed using nano-sized TiO<sub>2</sub> colloids as an adsorbent and hydride generation  
10                  atomic fluorescence spectrometry (HG-AFS) as a determination means. The  
11                  adsorption efficiency for Sb(III) and Sb(V) exceeded 98.0% within a short time when  
12                  the pH value was 6.0. Two carrier acid solutions, citric acid (1%) for Sb(III) and HCl  
13                  (2%) for total Sb, were used. After the determinations of Sb(III) and total Sb, Sb(V)  
14                  could be calculated by subtraction. The effects of such conditions as the Sb(III)/Sb(V)  
15                  ratio and the presence of coexisting ions were studied. The enrichment factors for  
16                  Sb(III) and total Sb were both 20. The detection limits of this method were 10 ng/L  
17                  for Sb(III) and 13 ng/L for total Sb. The relative standard deviations ( $n=6$ ) were 6.7%  
18                  for Sb(III) and 4.4% for total Sb, demonstrating good analytical precision. The  
19                  accuracy of the proposed method was checked with an environmental water sample  
20                  certified reference material (GSB 07-1376-2001) with a Sb(III) concentration of  
21                  1.24±0.26 mg/L. The obtained Sb(III) concentration was 1.22±0.03 mg/L, which was  
22                  in good agreement with the certified value. The developed method was simple, fast  
23                  and sensitive with no filtration or desorption steps and provided low detection limits  
24                  for Sb(III) and Sb(V).

25                  **Keywords:** Antimony speciation; Nano-sized TiO<sub>2</sub> colloids; HG-AFS

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## 27 1. Introduction

28 Water is the source of life. The determination and analysis of the contents of  
29 aquatic chemical substrates are important for research on environmental change and  
30 the influence of human activities on water quality as well as the prevention of water  
31 pollution. Antimony (Sb) exists as  $\text{Sb}^{3+}$  and  $\text{Sb}^{5+}$  in the natural aquatic environment,  
32 but the toxicity of  $\text{Sb}^{5+}$  is 10 times higher than that of  $\text{Sb}^{3+}$ .<sup>1</sup> Therefore, it is more  
33 meaningful to analyse antimony speciation than to determine the total concentration  
34 of antimony in water samples. The analysis of the speciation of  $\text{Sb}^{3+}$  and  $\text{Sb}^{5+}$  is much  
35 more difficult than that of total antimony because of the quite low concentrations of  
36 antimony in natural water samples. Thus, speciation methods with high sensitivity and  
37 selectivity for antimony need to be established.

38 Antimony speciation in water samples is usually investigated by on-line  
39 instrumental hyphenated techniques or a combination of preconcentration and  
40 instrumental analysis.<sup>2-4</sup> S  $\text{b}^{3+}$  and  $\text{Sb}^{5+}$  in water using high-pressure  
41 liquid chromatography–inductively coupled plasma–mass spectrometry  
42 (HPLC-ICP-MS), achieving detection limits of 0.1  $\mu\text{g/L}$  for  $\text{Sb}^{3+}$  and  $\text{Sb}^{5+}$ .<sup>5</sup> High  
43 pressure liquid chromatography–inductively coupled plasma–atomic emission  
44 spectrometry (HPLC-ICP-AES),<sup>6</sup> high pressure liquid chromatography–atomic  
45 fluorescence spectrometry (HPLC-AFS),<sup>7,8</sup> gas chromatography–inductively coupled  
46 plasma–mass spectrometry (GC-ICP-MS), GC-AFS, GC-AAS<sup>2</sup> and graphite  
47 furnace–atomic absorption spectrometry (GF-AAS)<sup>9</sup> were reported to provide good  
48 performance in the analysis of trace  $\text{Sb}^{3+}$  and  $\text{Sb}^{5+}$  in water samples. In recent years,  
49 several studies have reported antimony speciation utilising the combination of  
50 preconcentration and instrumental analysis. A 2 ng/L determination limit was obtained  
51 using liquid-phase micro-extraction combined with ETAAS.<sup>10</sup> Solid-phase extraction  
52 using carbon nano-tubes combined with ETAAS was also proposed by López-García  
53 et al. for antimony speciation, with a determination limit of 50 ng/L.<sup>11</sup> In most cases,

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4 54 the concentrations of  $\text{Sb}^{3+}$  and total Sb were analysed first due to the different  
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6 55 complexation or hydrogenation behaviours of  $\text{Sb}^{3+}$  and  $\text{Sb}^{5+}$ , after which the  
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8 56 concentration of  $\text{Sb}^{5+}$  can be calculated by subtraction.  
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11 57 In recent years, our research group has been working on the analysis of  
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13 58 ultra-trace metal ions using nano-sized  $\text{TiO}_2$  colloids as an adsorbent paired with  
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15 59 highly sensitive HG-AFS and GFAAS, yielding some promising results.<sup>12-16</sup>  
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17 60 Compared with nano-sized  $\text{TiO}_2$  powder, the colloids possess the advantage of easily  
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19 61 changing from the solid phase to the colloid phase after adsorption and centrifugation.  
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21 62 The dosage of colloids is low because of their high adsorption capacity, and an  
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23 63 autosampler can be used because the viscosity of resulting solution is close to that of  
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25 64 aqueous solution. Moreover, the concentrated analyte samples in colloidal form can  
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27 65 be directly analysed by AAS or AFS without desorption. In this work, nano-sized  
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29 66  $\text{TiO}_2$  colloids were used for the preconcentration of  $\text{Sb}^{3+}$  and  $\text{Sb}^{5+}$  in environmental  
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31 67 water samples because the adsorption could reach 98% in a pH range of 5.0 to 8.0.  
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34 68 HG-AFS is one of the most sensitive methods for the determination of antimony,  
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36 69 and it is also the Chinese standard method for the determination of total antimony in  
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38 70 mineral water. The measurement conditions were explored to apply this method to the  
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40 71 antimony speciation ( $\text{Sb}^{3+}$ ,  $\text{Sb}^{5+}$ ) after pre-concentration. The results showed that  $\text{Sb}^{5+}$   
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42 72 has only a slight influence on the determination of  $\text{Sb}^{3+}$  when 1% citric acid buffer  
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44 73 solution was used as the carrier liquid, and the  $\text{Sb}^{5+}$  concentration can be calculated as  
45  
46 74 the difference between the total antimony and  $\text{Sb}^{3+}$  concentrations.  
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49 75 A combination method using nano-sized  $\text{TiO}_2$  colloids as an adsorbent for  
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51 76 preconcentration and HG-AFS as a determination means was suggested by this work  
52  
53 77 for the speciation of  $\text{Sb}^{3+}$  and  $\text{Sb}^{5+}$ , which was simple and fast (without desorption  
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55 78 after preconcentration). The detection limits of our method were 10 ng/L for  $\text{Sb}^{3+}$  and  
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57 79 13 ng/L for total Sb. The relative standard deviations ( $n=6$ ) were 6.7% for  $\text{Sb}^{3+}$  and  
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59 80 4.4% for total Sb. This method was also applied to the speciation of  $\text{Sb}^{3+}$  and  $\text{Sb}^{5+}$  in  
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81 environmental water samples, and the results were satisfactory.

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4 825 83 **2. Materials and Methods**6  
7 84 **2.1 Reagents**

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9 85 All chemicals and reagents used throughout this study were of analytical grade or  
10 86 higher purity. Nanometre-size TiO<sub>2</sub> colloids (2.0% w/w) were supplied by Infrared  
11 87 Academy of Wuhan University (Wuhan, China). The characterisation of the used TiO<sub>2</sub>  
12 88 colloids was presented in our previous work.<sup>15</sup> Sb(III) standard solutions (1000 mg/L)  
13 89 were prepared using antimony trioxide and diluting to 1000 mL. A total of 100 mg L<sup>-1</sup>  
14 90 Sb(V) was oxidised from Sb(III) standard solutions (1000 mg/L) by potassium  
15 91 permanganate and sulphuric acid. A series of standard working solutions were  
16 92 prepared by diluting the standard stock solution with 2% (v/v) HCl. Thiourea-ascorbic  
17 93 acid solution (5% w/w) was dissolved with high-purity deionised water. The AFS  
18 94 carrier liquid for total Sb was prepared as 1.5% (w/w) KBH<sub>4</sub> and 2.0% (v/v) HCl. The  
19 95 AFS carrier liquid for Sb(III) was prepared as 1.5% (w/w) KBH<sub>4</sub> and 1% (w/w) citric  
20 96 acid. HCl and NaOH aqueous solutions were used to adjust the pH of the solutions. Ar  
21 97 (99.99%) was applied as a carrier gas. Water samples were drawn from Donghu Lake  
22 98 (Wuhan, Hubei Province, China) and Yangtze River (Wuhan, Hubei Province, China).  
23 99 The environmental reference material (GSB 07-1376-2001) was purchased from the  
24 100 Institution for Environmental Reference Materials of the Ministry of Environmental  
25 101 Protection (Beijing, China). All of the solutions were prepared with ultra-pure water  
26 102 (resistivity >18.0 MΩ/cm).

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47 103 **2.2 Instruments**

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49 104 A PF6-2 non-dispersive atomic fluorescence spectrometer (Beijing Purkinje  
50 105 General Instruments Co, Ltd, Beijing, China) equipped with a HAF-2 antimony  
51 106 hollow cathode lamp was used for antimony determination. The principle for the  
52 107 analysis of ions was described in our previous work.<sup>15</sup> The instrumental conditions for  
53 108 the analysis are listed in **Table 1**. The pH values were measured using a  
54 109 Mettler-Toledo 320-S pH meter (Mettler-Toledo, Greifensee, Switzerland). An  
55 110 LD5-2A centrifuge (Beijing Medical Centrifuge Factory, Beijing, China) was used to

111 separate the suspensions. The zeta potentials of TiO<sub>2</sub> colloids in the presence and  
 112 absence of Sb(III) or Sb(V) were determined using a Zetasizer 3600 (Malvern  
 113 Instrument, UK).

114 **Table 1 Instrument conditions for the analysis of Sb**

Negative high voltage /V	Flow rate of Ar /mL min <sup>-1</sup>	Lamp current /mA		Atomization T / °C
		Main	Auxiliary	
280	400	30	30	170

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## 116 2.3 Methods

### 117 2.3.1 Determination of the adsorption ratio for Sb(III) and Sb(V)

118 First, 10-mL standard solutions of Sb(III) or Sb(V) (100 µg/L) were placed in a  
 119 centrifuge tube (100 mL) containing 70 mL of ultra-pure water and then  
 120 supplemented with 0.5 mL of nano-sized TiO<sub>2</sub> colloids. The pH of the mixture was  
 121 adjusted to a fixed value. The mixed solution was shaken for 1 min by hand and then  
 122 centrifuged for 20 min at 4,500 rpm. The supernatant was transferred to a volumetric  
 123 flask (100 mL), after which 10 mL of thiourea-ascorbic acid solution (5% w/w) and 2  
 124 mL of conc. HCl were added. The solution was subsequently diluted to 100 mL. The  
 125 Sb(III) or Sb(V) concentration of the supernatant was determined by AFS with a 2%  
 126 HCl carrier solution. The fluorescence obtained was compared to that of Sb(III) or  
 127 Sb(V) standard solution (10 µg/L) without adsorption to calculate the adsorption ratio  
 128 of nano-sized TiO<sub>2</sub> colloids.

### 129 2.3.2 Preconcentration and determination of Sb(III) in water samples

130 First, 100-mL water samples were added to a centrifuge tube (120 mL) after air  
 131 pumping filtration using a 0.45-µm filtration membrane. Next, 0.5 mL of nano-sized  
 132 TiO<sub>2</sub> colloids was added to the centrifuge tube. The pH of the solution was adjusted to  
 133 6.0. The solution was centrifuged for 20 min at 4,500 rpm.

134 After centrifugation, the TiO<sub>2</sub> precipitates containing Sb were reconverted to  
 135 colloids with 2% HCl to a total volume of 5 mL. The concentration of Sb(III) in the

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4 136 resulting colloid samples was determined using a citric acid buffer carrier solution (1%  
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6 137 w/w, pH=4.38).

### 138 **2.3.3 Preconcentration and determination of Sb(V) in water samples**

139 After centrifugation, the TiO<sub>2</sub> precipitates containing Sb were reconverted to  
140 colloids to a total volume of 5 mL by the addition of 1 mL of thiourea-ascorbic acid  
141 solution and 4 mL of 2% HCl. The concentration of total Sb was determined with  
142 carrier HCl (2% v/v). The concentration of Sb(V) in the water samples could be  
143 calculated as the difference between the total Sb and Sb(III). A 20-fold  
144 preconcentration was used in this work because it is convenient for centrifugation, as  
145 the capacity of the centrifugation tube was 100 mL.

146

## 147 **3. Results and Discussion**

### 148 **3.1 Effects of HCl, citric acid and KBH<sub>4</sub> on the speciation analysis**

149 Under the acidic environment provided by the carrier solutions, KBH<sub>4</sub> can  
150 reduce Sb(III) to produce corresponding the hydride (H<sub>3</sub>Sb). H<sub>3</sub>Sb is atomised into  
151 hydrogen and antimony atoms in the atomiser cell and then analysed by hydride  
152 generation atomic fluorescence spectrometry (HG-AFS). The acid type and  
153 concentration as well as the concentration of KBH<sub>4</sub> could affect the speciation  
154 analysis of Sb. In total antimony mineral water national standard methods, the acidity  
155 of the sample solution system is 1% HCl, and the carrier liquid is 2% HCl. As shown  
156 in **Fig. 1**, under these conditions, Sb(V) can be partially reduced to Sb(III), even in the  
157 absence of ascorbic acid. Therefore, Sb(III) and Sb(V) cannot be distinguished.

158 However, when acetic acid (HAc) or citric acid (CA) buffer solution were used  
159 as the carrier liquids, the fluorescence signals of Sb(V) and Sb(III) were absent and  
160 strong, respectively, and the fluorescence of the former had little effect on that of  
161 latter. Although increasing the citric acid and HCl concentrations could improve the  
162 sensitivity toward both Sb(III) and Sb(V), the effect of Sb(V) on the signal of Sb(III)  
163 could be no longer be ignored. Moreover, the TiO<sub>2</sub> colloids become unstable with



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4 164 increasing citric acid concentration.

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6 165  $\text{KBH}_4$  concentrations in the range of 1% to 4% have little effect on the speciation  
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8 166 analysis of Sb. Therefore, the concentrations of  $\text{KBH}_4$ , citric acid and HCl were  
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10 167 selected as 1.5% (w/w), 1% (w/w) and 2.0% (v/v), respectively, in this work.

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### 13 14 169 **3.2 Influence of Sb(V) on Sb(III) fluorescence intensity**

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16 170 Different proportions of Sb(III) (10  $\mu\text{g/L}$ ) and Sb(V) (0 to 90  $\mu\text{g/L}$ ) were prepared.  
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18 171 The mixtures were determined according to 2.3.2 to investigate the effect of the ratio  
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20 172 of Sb(V) to Sb(III) on the determination. The relative fluorescence signals of the  
21  
22 173 mixtures were presented in **Fig. 2**.

23  
24 174 As shown in **Fig. 2**, in this method, within the range of Sb(III)/Sb(V) from 10/0 to  
25  
26 175 10/90, the presence of Sb(V) had little effect on the determination of Sb(III). The  
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28 176 concentration of Sb(III) could be determined alone when 1% CA was used as the  
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30 177 carrier liquid.

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### 33 34 179 **3.3 Effect of pH on the adsorption of Sb(III) and Sb(V) over $\text{TiO}_2$ colloids**

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36 180 When pH is lower than 4.0, the nano-sized  $\text{TiO}_2$  colloids disperse well and cannot  
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38 181 be separated from the bulk solution effectively. Therefore, the pH value should be set  
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40 182 higher than 4.0 to separate the colloidal particles from the solution. A series of pH  
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42 183 values were chosen to determine the adsorption efficiencies for Sb(III) and Sb(V)  
43  
44 184 according to method 2.3.1. The effect of pH on the adsorption of antimony species is  
45  
46 185 illustrated in **Fig. 3**.

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48 186 As shown in **Fig. 3**, the adsorption of Sb(III) and Sb(V) is slightly dependent on  
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50 187 pH. The adsorption efficiency for Sb(III) and Sb(V) could exceed 98.0% when the pH  
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52 188 was between 4.0 and 7.0. The high adsorption ratio indicated that both Sb(III) and  
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54 189 Sb(V) showed high adsorption affinity onto  $\text{TiO}_2$  colloids.

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56 190 To verify the electrostatic attraction between  $\text{TiO}_2$  colloid surface and Sb species,  
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58 191 the zeta potential of the  $\text{TiO}_2$  samples with adsorbed Sb(III) and Sb(V) was measured,  
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4 192 and the results are shown in **Fig. 4**. The results showed noticeable changes in the  
5  
6 193 surface potential of TiO<sub>2</sub> colloids in the presence of either 50 µg/L Sb(III) or 50 µg/L  
7  
8 194 Sb(V). The clear change in the point of zero charge (pH<sub>pzc</sub>) indicated that both Sb(III)  
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10 195 and Sb(V) formed inner-sphere complexes with TiO<sub>2</sub> colloids. The same zeta potential  
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12 196 patterns were also observed in the adsorption of Ge(IV) over TiO<sub>2</sub> colloids<sup>16</sup> and the  
13  
14 197 adsorption of As(III)/As(V) onto TiO<sub>2</sub> nano-powder,<sup>17</sup> whereas a different pattern was  
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16 198 observed for the adsorption of Se(IV)/Se(VI) onto TiO<sub>2</sub> colloids.<sup>15</sup>  
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### 200 **3.4 Effect of nano-sized TiO<sub>2</sub> colloid concentration on the adsorption efficiency**

201 Different volumes of nano-sized TiO<sub>2</sub> colloids (0.05 mL to 0.6 mL) were,  
202 respectively added to 100 mL 10 µg/L Sb(III) and Sb(V) solutions. The sorption  
203 efficiency of Sb(III) was relatively low when the TiO<sub>2</sub> colloid amount was less than  
204 0.4 mL. Quantitative sorption (>98.0%) of Sb(III) occurred with the increase of TiO<sub>2</sub>  
205 colloid amount (higher than 0.4 mL in this work). 0.5 mL TiO<sub>2</sub> colloids was chosen  
206 for the latter study, which resulted in complete adsorption of Sb(III) and Sb(V) and  
207 facilitated the experiments operation.  
208

### 209 **3.5 Adsorption isotherms of Sb(III) and Sb(V) on TiO<sub>2</sub>**

210 To study the adsorption isotherms, a series of Sb(III) or Sb(V) working solutions  
211 were prepared with concentrations of 0.1, 0.5, 0.6, 0.8, 1.0, 1.5 and 2.0 mg/L,  
212 respectively. The relationship between  $q_e$  and  $c_e$  was simulated using the Langmuir  
213 isotherm model (Eq. (1)):

$$214 \quad q_e = \frac{Kq_{max}c_e}{1+Kc_e} \quad (1)$$

215 where  $q_e$  is the amount of solute adsorbed per gram of TiO<sub>2</sub> colloids (µg/g).  $c_e$  is the  
216 equilibrium concentration (µg/L).  $K$  and  $q_{max}$  (µg/g) represent the equilibrium  
217 constant and maximum adsorption capacity of the solute on the TiO<sub>2</sub> surface,  
218 respectively.

219 As shown in **Fig. 5**, the amount of both Sb(III) and Sb(V) adsorbed onto the TiO<sub>2</sub>

220 colloids increased sharply and then plateaued as the corresponding equilibrium  
221 concentration increased. The experimental data could be well simulated by the  
222 Langmuir isotherm model, which indicated both Sb(III) and Sb(V) could form a  
223 single adsorption layer over the TiO<sub>2</sub> colloids. The calculated maximum adsorption  
224 amount was 46.4 mg/g and 70.8 mg/g for Sb(III) and Sb(V), respectively.

### 226 3.6 Influence of coexisting ions on the adsorption and preconcentration of Sb(III)

227 The effect of coexisting ions in real water on the adsorption and preconcentration  
228 of Sb(III) was studied. Various amounts of ions were added to a 100-mL solution  
229 containing 0.25 µg/L Sb(III). The resulting solution contained K<sup>+</sup> (1.8 mg), Na<sup>+</sup> (4  
230 mg), Ca<sup>2+</sup> (2 mg), Mg<sup>2+</sup> (2.4 mg), Cl<sup>-</sup> (74.55 mg), SO<sub>4</sub><sup>2-</sup> (9.9 mg), NO<sub>3</sub><sup>-</sup> (3.1 mg) and  
231 F<sup>-</sup> (0.005 mg). Although there are many other ions co-existing in the environmental  
232 samples, the concentration of the ions used above were 200 (F<sup>-</sup>)-3 × 10<sup>6</sup> (Cl<sup>-</sup>) times that  
233 of Sb(III), and the extraordinarily high concentration of the chosen ions could  
234 compensate for the effects of other coexisting ions. The enrichment procedure was  
235 then followed according to 2.3.2. The recovery of Sb(III) was 99.5%, which showed  
236 that coexistent ions had little effect on the proposed method.

### 238 3.7 Detection limit and relative standard deviation

239 The detection limit (D.L.) of the analysis method was obtained by Eq. (2).

$$240 \text{ D.L.} = 3\sigma \times \frac{c}{\bar{X}} = 3c \times \frac{\sigma}{\bar{X}} = 3c \times RSD \quad (2)$$

241 where  $\bar{X}$  is the mean signal of the independent analysis of a sample with a known  
242 concentration ( $c$ ) close to the LOD.  $\sigma$  is the standard deviation, and  $RSD$  is the  
243 relative standard deviation.

#### 244 3.7.1 Detection limit (D.L.) of Sb(III)

245 A 50 ng/L Sb(III) solution was enriched 20 times according to 2.3.2 and subjected  
246 to AFS. The detection limit and the relative standard deviation ( $RSD$ ,  $n=6$ ) of Sb(III)  
247 were 10 ng/L and 6.7%, respectively.

### 248 3.7.2 Detection limit of total Sb

249 A 100 ng/L Sb solution (containing Sb(III) 50 ng/L, Sb(V) 50 ng/L) was enriched  
 250 20 times according to section 2.3.3 and subjected to AFS. The detection limit and the  
 251 *RSD* ( $n=6$ ) of total Sb were 13 ng/L and 4.4%, respectively.

### 252 3.7.3 Accuracy of the method

253 An environmental reference material (GSB 07-1376-2001) with a Sb(III)  
 254 concentration of  $1.24 \pm 0.26$  mg/L was analysed. The environmental reference material  
 255 was diluted by a certain factor gradually because it was too thick to be determined.  
 256 The reference material solution after diluting was determined by 2.3.2. The obtained  
 257 Sb(III) concentration was  $1.22 \pm 0.03$  mg/L, which was in good agreement with the  
 258 certified value. The  $t$  value was 1.63 ( $\alpha=0.05$ ,  $n=6$ ), which was lower than the critical  
 259 value (2.571). The  $t$ -test results indicated that the analysed value was not statistically  
 260 significantly different from the certified value. The *RSD* for the analysis of the  
 261 reference material was 2.5%, which was much lower than the *RSD* obtained in section  
 262 3.8.1 (6.7%). The analysis of samples with lower concentrations will most likely  
 263 result in higher *RSD* values, while the analysis of samples with higher concentrations  
 264 will most likely yield lower *RSD* values.

265

## 266 3.8 Analysis of natural water samples

267 The above-mentioned method was used for the analysis of Sb(III) and  $Sb_{total}$  in water  
 268 samples from Donghu Lake and the Yangtze River. The results were listed in Table 2.  
 269 The concentrations of Sb(V) were found to be 0.07 and 0.06  $\mu\text{g/L}$  for Donghu Lake  
 270 and the Yangtze River, respectively.

271 Table 2 Determination of Sb(III) and total antimony ( $Sb_{total}$ ) species in water samples  
 272 (sample volume: 100 mL, final volume: 5 mL,  $n=3$ ).

Sample	Added ( $\mu\text{g/L}$ )		Found ( $\mu\text{g/L}$ )		Recovery (%)	
	Sb (III)	$Sb_{total}$	Sb (III)	$Sb_{total}$	Sb (III)	$Sb_{total}$
Donghu	---	---	$0.13 \pm 0.02$	$0.20 \pm 0.04$	---	---

Lake	0.20	0.40	$0.32 \pm 0.02$	$0.57 \pm 0.02$	$95.0 \pm 6.0$	$92.5 \pm 3.2$
Yangtze	---	---	0.00	$0.06 \pm 0.01$	---	---
River	0.10	0.20	$0.10 \pm 0.01$	$0.26 \pm 0.02$	$100.0 \pm 10.0$	$100.0 \pm 7.7$

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274 **4. Conclusions**

275 The innovation of this method lies in the application of nano-sized TiO<sub>2</sub> colloids as  
 276 an adsorbent for the speciation of Sb(III) and Sb(V) to increase the detection limit.  
 277 Sb(III) and Sb(V) could be selectively analysed by the use of appropriate carrier  
 278 solutions. The results suggested that this method had many advantages, such as a  
 279 simple preconcentration process (without desorption), good selectivity and  
 280 anti-interference ability, low detection limit and good precision. The detection limit  
 281 (D.L.) and the relative standard deviation (RSD,  $n=6$ ) of this method were 10 ng/L  
 282 and 6.7% for Sb(III), respectively, and 13 ng/L and 4.4% for total Sb, respectively.  
 283 The proposed method can be applied to determine ultra-trace Sb(III) and Sb(V) in  
 284 environmental water samples with highly satisfactory results.

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 288 Water Pollution Control and Management (2013ZX07503-001).

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290 **References**

- 291 1. E. Russeva and I. Havezov, *Anal. Lab.*, 1998, 115-127.
- 292 2. R. Miravet, E. Hernández-Nataren, A. Sahuquillo, R. Rubio and J. F.  
 293 López-Sánchez, *TrAC Trends in Analytical Chemistry*, 2010, 29, 28-39.
- 294 3. Z. Ge and C. Wei, *Journal of Chromatographic Science*, 2013, 51, 391-399.
- 295 4. S. L. C. Ferreira, W. N. L. dos Santos, I. F. dos Santos, M. M. S. Junior, L. O.  
 296 B. Silva, U. A. Barbosa, F. A. de Santana and A. F. de S. Queiroz, *Microchem*  
 297 *J.*, 2014, 114, 22-31.
- 298 5. F. S. éy, C. Gleyzes, O. Grosso, B. Plau and O. F. X. Donard, *Analytical and*

- 1  
2  
3  
4 299 *Bioanalytical Chemistry*, 2012, 404, 2939-2948.  
5 300 6. N. Ulrich, *Analytica Chimica Acta*, 1998, 359, 245-253.  
6 301 7. I. De Gregori, W. Quiroz, H. Pinochet, F. Pannier and M. Potin-Gautier,  
7 302 *Talanta*, 2007, 73, 458-465.  
8  
9 303 8. I. De Gregori, W. Quiroz, H. Pinochet, F. Pannier and M. Potin-Gautier,  
10 304 *Journal of chromatography. A*, 2005, 1091, 94-101.  
11  
12 305 9. D. Mendil, H. Bardak, M. Tuzen and M. Soylak, *Talanta*, 2013, 107, 162-166.  
13 306 10. J. M. Serafimovska, S. Arpadjan and T. Stafilov, *Microchemical Journal*, 2011,  
14 307 99, 46-50.  
15  
16 308 11. I. López-García, R. E. Rivas and M. Hernández-Córdoba, *Talanta*, 2011, 86,  
17 309 52-57.  
18  
19 310 12. S. H. Qian, H. Lin, X. Q. Li, M. Xiao, H. B. Deng and L. J. Xiang, *Wuhan*  
20 311 *University Journal of Natural Sciences*, 2007, 2, 349-352.  
21  
22 312 13. S. Qian, S. Zhang, Z. Huang, M. Xiao and F. Huang, *Microchimica Acta*, 2009,  
23 313 166, 251-254.  
24  
25 314 14. S. Qian, Z. Huang, J. Fu, J. Kuang and C. Hu, in *Analytical Methods*, 2010,  
26 315 vol. 2, p. 1140.  
27  
28 316 15. J. Fu, X. Zhang, S. Qian and L. Zhang, *Talanta*, 2012, 94, 167-171.  
29 317 16. L. Xiang, X. Zhang, M. Lu, S. Qian and S. Mo, *Journal of Analytical Atomic*  
30 318 *Spectrometry*, 2012, 27, 359-363.  
31  
32 319 17. M. Pena, X. Meng, G. P. Korfiatis and C. Jing, *Environmental Science &*  
33 320 *Technology*, 2006, 40, 1257-1262.  
34 321  
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10 338 **Figure Captions**

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12 339 Fig. 1 Fluorescence of Sb(III) and Sb(V) with HCl, CA and HAc carrier solutions.

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14 340 Sb(III)=10  $\mu\text{g/L}$ , Sb(V)=10  $\mu\text{g/L}$ , mixture: 10  $\mu\text{g/L}$  Sb(III) and 10  $\mu\text{g/L}$  Sb(V)

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16 341 Fig. 2 Effect of Sb(V) on the fluorescence intensity of Sb(III)

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18 342 Fig. 3 Effect of pH on the adsorption ratio for Sb(III) and Sb(V)

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20 343 Fig. 4 Zeta potential of 0.1%  $\text{TiO}_2$  colloids as a function of pH for 50  $\mu\text{g/L}$  Sb(III) and

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22 344 50  $\mu\text{g/L}$  Sb(V)

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24 345 Fig. 5 Dependence of the adsorption amount on the equilibrium concentration of

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26 346 Sb(III) and Sb(V). Lines represent the fit of the data to the Langmuir adsorption

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28 347 isotherm model

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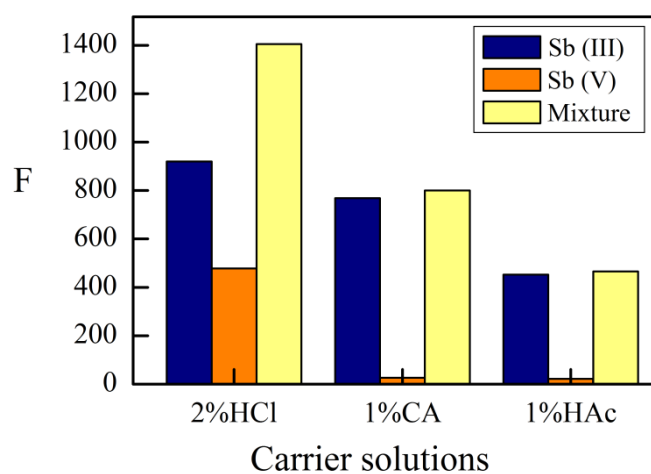
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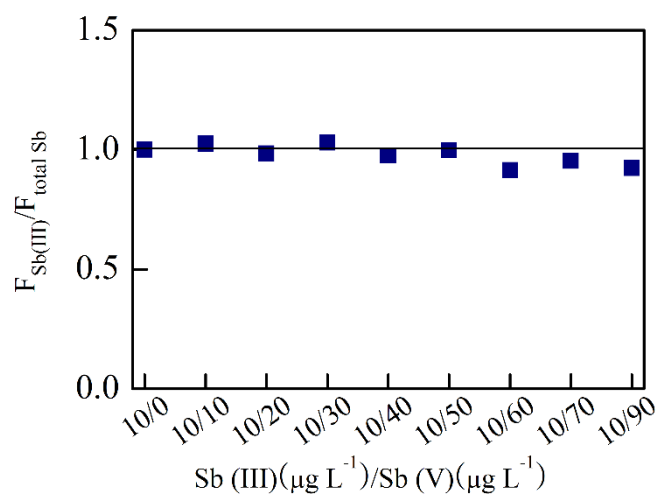
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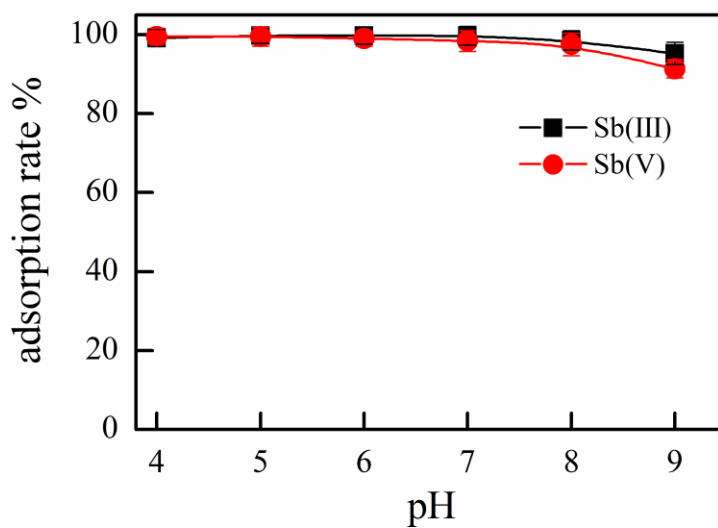
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370 **Figure 1**

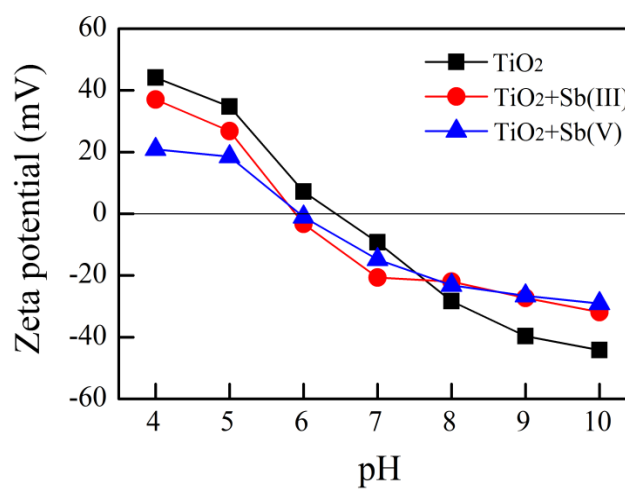
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372 **Figure 2**





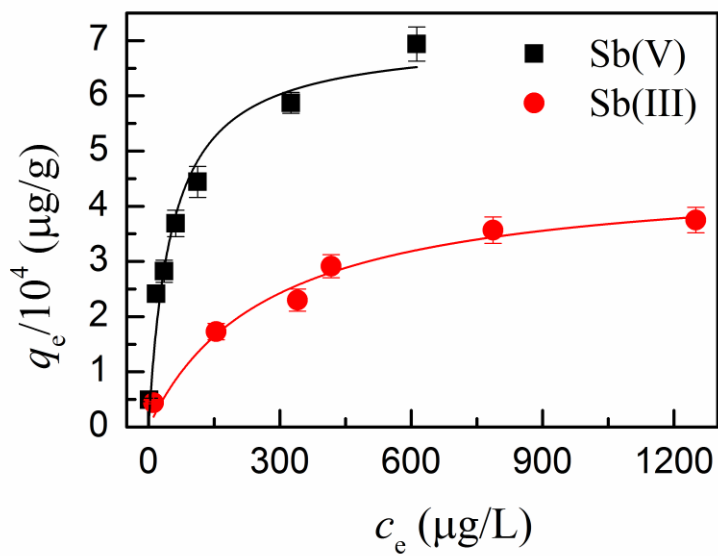
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374 **Figure 3**

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376 **Figure 4**

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379 **Figure 5**

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