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## Environmental Impact Statement

Since their commercialization, electronic cigarettes (e-cigarettes) have been marketed and used as “safer” surrogates of conventional tobacco-containing cigarettes. The increasing popularity of these devices during the past decade necessitates accurate identification of the environmental impacts of e-cigarette consumption and its potential health endpoints. While most of the few previous research efforts in this area have focused on characterization of gas-phase e-cigarette emissions, the much-needed data regarding e-cigarette’s particulate emissions is scarce. This manuscript quantifies the level of exposure, as well as indoor emission rates of inorganic elements and organic compounds, providing insight regarding the toxico-chemical properties of e-cigarette’s secondhand emissions and potential needs for regulatory supervision on the manufacturing of these devices.

# Particulate Metals and Organic Compounds from Electronic and Tobacco-containing Cigarettes: Comparison of Emission Rates and Secondhand Exposure

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## 1 Abstract

2 In recent years, electronic cigarettes have gained increasing popularity as alternatives to normal  
3 (tobacco-containing) cigarettes. In the present study, particles generated by e-cigarettes and  
4 normal cigarettes have been analyzed and the degree of exposure to different chemical agents  
5 and their emission rates were quantified. Despite the 10-fold decrease in the total exposure to  
6 particulate elements in e-cigarettes compared to normal cigarettes, specific metals (e.g. Ni and  
7 Ag) still displayed a higher emission rate from e-cigarettes. Further analysis indicated that the  
8 contribution of e-liquid to the emission of these metals is rather minimal, implying that they  
9 likely originate from other components of the e-cigarette device or other indoor sources. Organic  
10 species had lower emission rates during e-cigarette consumption compared to normal cigarettes.  
11 Of particular note was the non-detectable emission of polycyclic aromatic hydrocarbons (PAHs)  
12 from e-cigarettes, while substantial emission of these species was observed from normal  
13 cigarettes. Overall, with the exception of Ni, Zn, and Ag, the consumption of e-cigarettes  
14 resulted in a remarkable decrease in secondhand exposure to all metals and organic compounds.  
15 Implementing quality control protocols on the manufacture of e-cigarettes would further  
16 minimize the emission of metals from these devices and improve their safety and associated  
17 health effects.

18 **Keywords:** Electronic cigarette, Particulate Matter, Metals, Organic compounds, Emission Rate

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## 22 1. Introduction

23 Use of electronic cigarettes (e-cigarettes) has rapidly increased worldwide during the past  
24 decade.<sup>1</sup> E-cigarettes are battery-operated electronic nicotine delivery devices (ENDD) that  
25 provide an aerosol from a liquid mixture of nicotine, propylene glycol, glycerol and flavorings  
26 (known as "e-liquid").<sup>2,3</sup> Considering that nicotine delivery by e-cigarettes is achieved by  
27 aerosolizing the "e-liquid" as opposed to the combustion of tobacco, e-cigarette consumption is  
28 presumed to carry with it lower health risks compared to normal cigarettes.<sup>1,4,5</sup> E-cigarettes,  
29 however, pose a significant regulatory challenge for the medical community and public health  
30 authorities. Although e-cigarettes are promoted as aids to reduce the use of tobacco-containing  
31 cigarettes,<sup>4</sup> there are still uncertainties regarding the degree to which they promote a clinically  
32 relevant cessation rate in smokers.<sup>6</sup> Moreover, unregulated production of e-cigarettes and e-  
33 liquids, in addition to the very limited scientific evidence regarding the chemical composition of  
34 the vapors and aerosols generated by e-cigarettes, have raised concerns about the potential  
35 adverse health effects of e-cigarette consumption.<sup>7-9</sup>

36 Based on recent studies, there is considerable inconsistencies in the levels of nicotine as well as  
37 organic compounds (such as propylene glycol and glycerol) in the vapors generated by different  
38 brands of e-cigarettes.<sup>10,11</sup> Another recent study reported generally lower levels of organic  
39 species, including volatile organic compounds (VOCs), carbonyls, polycyclic aromatic  
40 hydrocarbons (PAHs) and glycols, in the vapors emitted from e-cigarettes compared to tobacco  
41 smoke.<sup>12</sup> Williams et al.<sup>13</sup> analyzed 11 particulate elements in the aerosol generated by e-  
42 cigarettes and reported the presence of tin, silver, iron, nickel and aluminum in super-micron  
43 particles, and tin, chromium and nickel in sub-micron particles. Further analyses indicated that

44 these potentially toxic elements originate from the e-cigarette's cartomizer, alerting us to the  
45 need for improved quality control in e-cigarette manufacture and further investigations on the  
46 potential adverse health impacts of e-cigarette consumption.<sup>13</sup>

47 The majority of the few previous studies on the chemical characterization of e-cigarette  
48 emissions focused on vapor-phase emissions, emphasizing the need for fundamental data  
49 pertaining to aerosols emissions from e-cigarettes. In this study, particles generated by the  
50 consumption of e-cigarettes as well as normal cigarettes were collected in a room under  
51 controlled conditions and comprehensive chemical analyses (including the quantification of 42  
52 inorganic elements and various organic compounds) have been performed on the samples, in  
53 order to quantify the degree of secondhand exposure to particulate organic compounds and  
54 metals in a real-life setting. Moreover, emission rates of individual metals and organic  
55 compounds were determined using a single-compartment mass balance model and Monte-Carlo  
56 error estimation analysis.

57

## 58 **2. Methodology**

### 59 **2.1. Sampling protocol**

60 Samples of total suspended particles were collected indoors on quartz filters (Whatman  
61 International Ltd, Maidstone, England), using a high-volume PM sampler operating at a flow rate  
62 of 240 liters per minute (lpm), in a room having a volume of 48 m<sup>3</sup> on the fifth floor of the  
63 Fondazione IRCCS Istituto Nazionale dei Tumori, an Italian cancer research center in Milan.  
64 The room was furnished with typical office furniture. In order to ensure a homogenous  
65 distribution of the aerosol, the air inside the room was well-mixed by means of three fans

66 blowing into different directions during all experiments. The room's window faced a terrace,  
67 where outdoor particulate matter (PM) was simultaneously sampled and collected on similar  
68 quartz filters, using a similar high-volume PM sampler, operating at 240 lpm. The initial air  
69 exchange rate (AER) inside of the sampling room was estimated to range between 0.80 and 0.86  
70  $\text{hr}^{-1}$ , by measuring the decay rate of carbon monoxide (CO) generated by combustion of incense  
71 sticks, and further increased to about 1.1  $\text{hr}^{-1}$  during the sampling periods due to the additional  
72 AER caused by the sampler pump (operating at 240 lpm in a room of 48  $\text{m}^3$ ).

73 One of the most popular European brands of e-cigarette (Elips Serie C, Tank System, Ovale  
74 Europe Srl) and a typical, widely used brand of normal (i.e. tobacco-containing) cigarette was  
75 used for all experiments. E-cigarettes and normal cigarettes were smoked by three volunteer  
76 smoker subjects, including two males (aged 55 and 64 years old) and one female (aged 32 years  
77 old). Both the smoking procedure and indoor environment were designed in a way to simulate  
78 real-life conditions. Normal cigarettes were smoked '*ad libitum*' by the subjects, with an average  
79 rate of one puff every minute, leading to an average total time of seven minutes for completion  
80 of one tobacco-containing cigarette. A 3 minute pause was applied between two subsequent  
81 cigarettes. To enable a robust comparison between normal cigarette and e-cigarette emissions,  
82 the e-cigarette smoking procedure was designed to be similar to the smokers' conventional  
83 cigarette smoking habit. E-cigarettes were therefore vaped at a rate of one puff per minute,  
84 lasting for seven minutes, followed by three minutes of pause and continuing again for another  
85 seven minutes. During the vaping/smoking, two people were present in the room, one smoker  
86 and one person for instrument operation. Smokers were positioned on one side of the room while  
87 all instruments and samplers were positioned on the other side in order to avoid direct blowing of  
88 vapors into the inlets of the instruments and to ensure the highest possible mixing in the room.

89 The cartridge of the e-cigarette had a volume of 1.5 ml and was filled with the commercial e-  
90 liquid (consisting of propylene glycol, glycerol, aroma and water, based on the manufacturer). In  
91 order to investigate nicotine emission rate, e-cigarette samples were collected both without and  
92 with adding nicotine to the e-liquid (a concentration of 16 mg/ml nicotine was used in the e-  
93 liquid for the latter). Using the above-mentioned smoking protocol, an approximate e-liquid  
94 volume of 1.3 ml was consumed per hour and the e-liquid cartridge was refilled frequently  
95 throughout the experiments. A total of 6 e-cigarette samples (3 with nicotine and 3 without  
96 nicotine) and 3 normal cigarette samples were collected, each with parallel outdoor samples.  
97 With the exception of the nicotine analysis, e-cigarette samples with and without nicotine were  
98 combined for all other analyses, in order to increase the statistical power of the results and  
99 comparisons.

## 100 **2.2. Chemical analyses**

101 The concentration of black carbon (BC) was measured inside the smoking room, using an  
102 aethalometer (7-wavelength AE 31, Magee Scientific). Concentration of carbon monoxide (CO)  
103 was quantified by an indoor air quality analyzer (Graywolf Sensing Solutions, Shelton, CT,  
104 USA). Inorganic elements as well as organic compounds were measured by time-integrated PM  
105 samples collected on quartz filters. To measure the total concentration of elements in the PM,  
106 portions of the quartz filters were digested in an acid mixture (comprised of nitric acid,  
107 hydrochloric acid and hydrogen peroxide), inside of a microwave-assisted Teflon digestion bomb  
108 (Milestone ETHOS+), followed by analysis using a high-resolution magnetic sector inductively  
109 coupled plasma mass spectrometry (ICP-MS) (Thermo-Finnigan Element 2). Further details  
110 regarding this method can be found in Herner et al.<sup>14</sup> Concentration of individual organic species



111 was measured by gas chromatography mass spectrometry (GC-MS) method (GC-6980,  
112 quadrupoleMS-5973, Agilent Technologies), as described in Stone et al.<sup>15</sup> In brief, portions of  
113 quartz filters were extracted in a 1:1 solution of dichloromethane and acetone, using Soxhlets,  
114 followed by volume reduction using rotary evaporation, under high purity nitrogen and  
115 derivitization of carboxylic acids with diazomethane, and then analyzed by GC-MS.  
116 Concentration of elements and metals, as well as speciated organic compounds was also  
117 quantified in the e-liquid, using the same ICP-MS and GC-MS methods, respectively. For all of  
118 the analyses, detection limits of the measurements were calculated as 2 times the total analytical  
119 uncertainties, in the limit as the concentrations of the species approach zero (presented in Table  
120 S1).

121

### 122 2.3. Determination of emission rates

123 In order to calculate the emission rate of particle-bound species, a single-compartment mass  
124 balance model was applied to the smoking room for each species. The overall mass balance  
125 equation in the room is given in Equation 1:

$$126 \quad V \frac{dC_{in}}{dt} = S_i + C_{out} P(AER) V - C_{in}(AER)V - kC_{in}V \quad (\text{Eq. 1})$$

127 where  $C_{out}$  and  $C_{in}$  represent outdoor (i.e. ambient) concentration and indoor (i.e. inside of the  
128 smoking room) concentration, respectively (expressed in  $\text{ng}/\text{m}^3$ ).  $P$  represents the penetration  
129 efficiency of particles and is dimensionless.<sup>16</sup>  $AER$  and  $k$  are the air exchange rate and particle  
130 deposition loss, respectively, both measured in units of  $\text{hr}^{-1}$ .  $V$  represents the volume of the  
131 smoking room ( $48 \text{ m}^3$ ) and  $S_i$  is the indoor emission rate of the species ( $\text{ng}/\text{hr}$ ). Assuming steady

132 state conditions (i.e.  $dC_{in}/dt=0$ , achieved by rather long sampling periods as well as relatively  
133 high air exchange rate) and homogeneous distribution of compounds in the room (achieved by  
134 continuously mixing the air using three fans during all experiments), the mass balance for each  
135 compound is simplified to Equation 2:

$$136 \quad C_{in} = \frac{P(AER)C_{out}}{(AER+k)} + \frac{(S_i/v)}{(AER+k)} \quad (\text{Eq. 2})$$

137 Indoor and outdoor concentrations were determined following the chemical analyses on the filter  
138 samples and a total air exchange rate (AER) of 1.1 ( $\pm 0.06$ )  $\text{hr}^{-1}$  was determined for the  
139 experimental conditions (including 0.8  $\text{hr}^{-1}$  based on the decay rate of CO, in addition to 0.3  $\text{hr}^{-1}$   
140 to account for the AER caused by the pump, operating at 240 lpm in a 48  $\text{m}^3$  room). Rate of  
141 deposition loss (k) and penetration efficiency (P) were estimated to vary between 0.1-0.2  $\text{hr}^{-1}$  and  
142 0.7-0.85, respectively, based on the study by Long et al.<sup>16</sup> Based on the variables' uncertainties,  
143 the uncertainties of indoor emission rates ( $S_i$ ) were propagated using a Monte Carlo error  
144 estimation analysis,<sup>17</sup> by calculating the standard deviation of the mean emission rate values after  
145 performing the Monte Carlo based on 1000 iterations using randomly selected variables.

146

### 147 **3. Results and Discussion**

#### 148 **3.1. Mass concentration overview**

149 Figure 1 presents an overview of the total PM mass concentration indoors (i.e. inside of the  
150 smoking room) and outdoors, during e-cigarette and normal cigarette consumption. As seen in  
151 the figure, the overall increase in the indoor PM mass concentration level compared to the  
152 simultaneously-measured outdoor concentration is remarkable during normal cigarette smoking

153 (from about  $55 \mu\text{g}/\text{m}^3$  to more than  $250 \mu\text{g}/\text{m}^3$ ), while for e-cigarette vaping, similar levels are  
154 observed for indoor and outdoor PM concentrations. As will be shown later in the emission rate  
155 calculation section, despite similar levels of total PM concentrations indoor compared to outdoor  
156 during e-cigarette vaping, indoor emission rates of specific metals and organic compounds are  
157 significant. We should also note that presence of nicotine in the e-liquid had a very small effect  
158 (less than 0.1 %) on the e-cigarette's total PM emissions (as shown in Figure S1, supplementary  
159 information). Emission rate and indoor concentration of particle-phase nicotine during e-  
160 cigarette vaping are discussed separately, in section 3.4.

161 Figure 2 presents the mass fraction of aggregated elements, and groups of organic compounds  
162 (including PAHs, alkanes, organic acids, hopanes and levoglucosan). Concentration of these  
163 chemical groups in the air (i.e. normalized by the volume of air) is also presented in Figure S2  
164 (supplementary information). Specific chemical species in each group and their corresponding  
165 concentrations are given in Table S1 (supplementary information) for reference. As seen in  
166 Figure 2, the mass fraction of alkanes and organic acids is 5-10-fold greater for normal cigarette  
167 samples compared to both e-cigarette and outdoor samples. Hopanes, tracers of gasoline and  
168 diesel engine combustion emissions,<sup>18</sup> were only detected outdoors. PAHs were found in larger  
169 proportions in normal cigarette samples compared to outdoor ambient air (about 10 times), while  
170 they were undetected for indoor samples collected during e-cigarette vaping. Considering the  
171 potential carcinogenic effects of PAHs,<sup>19,20</sup> reduced emission of PAHs is one of the most  
172 desirable outcomes of e-cigarette vaping compared to normal cigarette consumption. The mass  
173 fraction of inorganic elements was highest for outdoor samples, driven by the considerably  
174 higher (10-60-fold) concentration of dust/soil dominated elements (such as Mg and Ca) in the

175 ambient air (Table S1-a). In the following sections, exposure concentrations as well as emission  
176 rates of specific chemical components are discussed.

177

## 178 **3.2. Level of exposure to chemical species**

### 179 **3.2.1. Black Carbon**

180 Real-time aethalometer-based black carbon (BC) measurements at 7 wavelengths (370, 470, 520,  
181 590, 660, 880 and 950 nanometers) inside of the smoking room are presented in Figure 3. BC  
182 concentrations in the aerosol are estimated from the absorption at the 880 nm wavelength (i.e.  
183 infrared).<sup>21</sup> The light absorption associated with shorter wavelengths may provide insights on  
184 possible composition of the aerosol in terms of organic species. Typically, aerosols with a high  
185 content of light-absorbing organic compounds would exhibit higher light absorbance at shorter  
186 wavelengths.<sup>21</sup> Therefore, for an aerosol with low organic content, the reported BC concentration  
187 at 880 nm (i.e. actual or “standard” BC) would be close to that at shorter wavelengths, whereas  
188 for an aerosol with a high organic content (such as PAHs in tobacco smoke), the BC  
189 concentration reported for shorter wavelengths will be higher than that of standard BC  
190 concentration. As shown in Figure 3, BC concentration during e-cigarette vaping is consistently  
191 close to zero, while during the period of normal cigarette consumption, a spectrum of distinct BC  
192 concentrations at different wavelengths is observed. The standard (i.e. actual) BC concentration  
193 during normal cigarette smoking (corresponding to 880 nm) peaks at  $4.1 \mu\text{g}/\text{m}^3$ , while the  
194 reported BC concentration at shorter wavelengths is much higher (e.g.  $24.8 \mu\text{g}/\text{m}^3$  at 470 nm  
195 (“blue” light) and  $65.4 \mu\text{g}/\text{m}^3$  at 370 nm (“ultraviolet” light)), likely due to the presence of

196 ultraviolet-absorbing organic compounds in the tobacco smoke.<sup>21</sup> As will be discussed in the  
197 following sections, results of our GC-MS analysis confirm the abundance of organic species  
198 (including, but not limited to, alkanes, organic acids and PAHs) in tobacco smoke.

199

### 200 **3.2.2. Inorganic Elements and Metals**

201 Indoor to outdoor concentration ratios (I/O) of the elements that were detected in e-cigarette  
202 samples are presented in Table 1, along with Pearson correlation coefficient (R) between indoor  
203 and outdoor concentrations. The actual air concentrations of 47 elements that were quantified by  
204 the ICP-MS analysis are also included in Table S1-a (supplementary information). I/O ratio and  
205 correlation coefficient (R) together provide insights on the origin of the indoor aerosol. An I/O  
206 ratio greater than 1 and a low R indicate that indoor PM is originating predominantly from  
207 indoor sources, whereas an I/O ratio smaller than 1 with moderate to high (i.e. greater than 0.7)  
208 R is an indication of significant infiltration of outdoor PM. As shown in Table 1, the average I/O  
209 ratio for 6 elements (including B, K, Ni, Zn, Ag and La) is greater than unity with a low  
210 correlation coefficient ( $-0.16 < R < 0.22$ ), indicating the presence of indoor emission source(s) for  
211 these elements during e-cigarette vaping. In order to investigate the potential influence of the e-  
212 liquid's elemental content on the observed indoor levels of the aforementioned 6 elements, ICP-  
213 MS analysis was performed on the e-liquid samples and concentration of these elements in the e-  
214 liquid was quantified (results of the e-liquid analysis are presented in Table S3, supplementary  
215 information). To segregate the contribution of e-liquid and other e-cigarette components to the  
216 emissions (which is of utmost importance from a regulatory standpoint), the single-compartment  
217 mass balance model (discussed in the section 2.3.) was employed to determine the portion of the

218 indoor concentrations that originate from the e-liquid. The results are presented in Figure 4. As  
219 seen in Figure 4, for all of these elements (specifically Ni, Zn, Ag and La), the estimated  
220 contribution of the e-liquid to the indoor concentration is lower than the actual indoor levels  
221 during e-cigarette vaping, suggesting that these elements originate from another indoor source. In  
222 a recent study by Williams et al.,<sup>13</sup> the cartomizer's elemental content was examined using  
223 electron microscopy and compared to the aerosol's elemental content. Their analysis confirmed  
224 the presence of metals such as Sn, Ag, Fe, Ni, Al and Cr in the cartomizer and their emission  
225 thereof in the aerosol generated by the e-cigarette, suggesting that the material used in the  
226 cartridge of the e-cigarette is a likely source contributing to the total emission of these metals in  
227 our study as well. Indoor concentration of B, K, Ni, Zn, Ag and La during e-cigarette vaping and  
228 normal cigarette smoking are compared in Figure 5. The concentrations of B, K and La were  
229 statistically significantly ( $p < 0.05$ ) higher in normal cigarettes. On the other hand, there was no  
230 statistically significant difference between e-cigarette and normal cigarette samples for Zn, Ni  
231 and Ag (potentially toxic and redox active species<sup>22,23</sup>). Considering the potential adverse health  
232 effects associated with the inhalation of these metals (particularly Ni and Zn, and the emission  
233 observed both in our analysis as well as the study by Williams et al.<sup>13</sup>), attention should be  
234 directed toward eliminating the use of these metals in the cartridges during the manufacturing  
235 process of e-cigarettes.

### 236 3.2.3. Organic Species

237 The concentration of individual organic species for both indoor and outdoor samples is shown in  
238 Table S1 (b-e). PAH concentrations, including benz(a)anthracene, chrysene,  
239 benzo(b)fluoranthene, benzo(k)fluoranthene, benzo(e)pyrene, benzo(a)pyrene, indeno(1,2,3-

240 cd)pyrene and benzo(g,h,i)perylene, were found to be substantially higher indoors than outdoors  
241 during normal cigarette smoking (ranging between 1.4 to 5.1 ng/m<sup>3</sup> for indoor and 0.19 ng/m<sup>3</sup> or  
242 less for outdoor), while they were not detected in the indoor samples that were collected during  
243 e-cigarette vaping. Hopanes and steranes, tracers of lubricating/engine oil in diesel and gasoline  
244 vehicles,<sup>24</sup> were non-detected for all indoor samples associated with e-cigarettes. Levoglucosan,  
245 a tracer of biomass combustion,<sup>25</sup> was detected in all indoor and outdoor samples, with indoor  
246 concentration during e-cigarette vaping about 80-fold lower than that during normal cigarette  
247 smoking (which was expected due to the presence of biomass in normal tobacco-containing  
248 cigarettes) and about 3-fold lower than ambient levels. N-alkanes and organic acids were the  
249 only groups of organic species that were found at higher concentrations indoor than outdoor  
250 during e-cigarette use.

251 For all of the organic species with detectable indoor concentrations during e-cigarette vaping, the  
252 I/O concentration ratios as well as Pearson correlation coefficients between indoor and outdoor  
253 concentrations were determined (presented in Table 2). As seen in the table, levoglucosan is the  
254 only organic compound with an I/O ratio smaller than unity. Considering the strong association  
255 between outdoor and indoor levoglucosan levels during e-cigarette vaping ( $R=0.99$ ), the indoor  
256 levels of levoglucosan for e-cigarette samples are most likely associated with ambient  
257 levoglucosan that infiltrated indoors. Alkanes and organic acids presented in Table 2 had average  
258 I/O ratios greater than one, indicating the presence of an indoor source for these species during e-  
259 cigarette consumption. This observation was further investigated by an analysis of the organics  
260 content of the e-liquid (presented in Table S3). Excluding hexadecanoic acid and octadecanoic  
261 acid, which respectively displayed concentrations of 511 ( $\pm 148$ ) and 247 ( $\pm 118$ )  $\mu\text{g/ml}$  in the e-

262 liquid, this analysis resulted in non-detectable levels for all measured compounds. Considering  
263 the non-detectable concentration of most n-alkanes and organic acids in the e-liquid, definitive  
264 conclusion regarding the sources of these organics in the indoor environment is not possible  
265 based on our data. The rather low concentration of these compounds during e-cigarette vaping  
266 (i.e. less than 100 ng/m<sup>3</sup> for most of the species, as shown in Tables S1-d and S1-e), is most  
267 likely indicative of a dominant background indoor level and is not representative of a specific  
268 source. It is important to note that additional quantification of the gas-phase organics in the  
269 indoor environment is needed to accurately determine the sources of indoor organic compounds,  
270 the emission of which may have disproportional partitioning in the gas-phase and particulate-  
271 phase. I/O ratio for alkanes and organic acids during normal cigarette smoking is also presented  
272 in Table S2 (supplementary information), with all of the compounds exhibiting I/O ratios greater  
273 than unity (ranging from a minimum of 4.3 for hexatriacontane, up to more than 300 for species  
274 such as tritriacontane), demonstrating the presence of strong indoor emission source(s) for these  
275 species during normal cigarette smoking. Indoor emission rate of these organic species are  
276 calculated and discussed in the next section.

### 277 **3.3. Emission rates of chemical species**

278 A single-compartment mass balance model was employed to determine and compare the indoor  
279 emission rates of chemical species during e-cigarette vaping and normal cigarette smoking. The  
280 results are presented in Tables 3 (inorganic elements) and 4 (organic species). A non-detectable  
281 (N.D.) emission rate indicates that there was no detectable indoor emission source for the species  
282 and the observed indoor concentrations were, therefore, mostly due to the penetration of outdoor  
283 aerosol to the indoor environment as a result of air exchange (AER=1.1 hr<sup>-1</sup>) between the indoor



284 and outdoor environments. The indoor emission rates of several elements with adverse health  
285 effects upon inhalation are substantially decreased in e-cigarette's aerosol compared to that from  
286 normal cigarette. Cadmium and Lead, for instance, are particularly toxic metals that have been  
287 consistently found in tobacco smoke in previous studies (e.g.<sup>26,27</sup>). Indoor emission rates  
288 associated with these metals during smoking of normal cigarettes were as high as 1012 ng/hr and  
289 657 ng/hr for Pb and Cd, respectively, while for e-cigarette vaping these rates are lower by 2-3  
290 orders of magnitude. Similarly, sulfur (another element commonly present in high concentrations  
291 in tobacco smoke<sup>28</sup>) had an indoor emission rate of about 34 µg/hr during normal cigarette  
292 smoking, while it had no detectable indoor emission during e-cigarette consumption. With the  
293 exception of Ni, Ti, Cr and Ag (elements also found in the e-cigarette's aerosol in a previous  
294 study by Williams et al.<sup>13</sup>), the indoor emission rate of all elements was higher for normal  
295 cigarette smoking compared to e-cigarette vaping. This observation suggests that even though e-  
296 cigarette's aerosol contains fewer deleterious elements, and at lower concentrations, it does  
297 contain several metals such as Ag, Cr, and Ni (which are toxic metals<sup>22,23</sup>) that are being emitted  
298 in higher rates from e-cigarettes. This is most likely due to the low quality and lack of  
299 supervision and control on the manufacturing process of e-cigarette's cartridges, as explained in  
300 the article by Williams et al.<sup>13</sup> Moreover, although the emission of Pb was substantially  
301 decreased in e-cigarette compared to normal cigarette, its presence in e-cigarette's aerosol can  
302 still be further reduced or eliminated by implementing better quality control procedures in the  
303 manufacturing process of e-cigarettes.

304 Tables 4 and 5 highlight the emission rates associated with organic species (n-alkanes and  
305 alkanolic acids in Table 4 and PAHs in Table 5). As seen in Table 4, the emission rate of all  
306 particulate alkanes and organic acids decreased by 2-3 orders of magnitude for e-cigarette

307 compared to normal cigarette. Seven PAHs were detected in normal cigarette samples (including  
308 chrysene, benzo(b)fluoranthene, benzo(k)fluoranthene, benzo(e)pyrene, benzo(a)pyrene,  
309 indeno(1,2,3-cd)pyrene and benzo(g,h,i)perylene) with indoor emission rates ranging from 105  
310 ng/hr benzo(e)pyrene to 307 ng/hr benzo(b)fluoranthene, while none of these species were  
311 detected indoors during e-cigarette vaping. We note here that although our analysis indicates that  
312 the particle-phase organic content of e-cigarette's aerosol is considerably lower than that of  
313 normal cigarette smoke, another portion of e-cigarette's emissions are expected to be in the  
314 vapor phase, the evaluation of which was not part of our analysis. Further analysis of vapor-  
315 phase e-cigarette emissions might be useful to uncover the vapor-phase levels of organic  
316 emissions from e-cigarettes.

#### 317 **3.4. Indoor concentration and emission rate of nicotine**

318 In order to examine particle-phase nicotine levels and emission rates, e-cigarette samples without  
319 the addition of nicotine (i.e. loaded with e-liquid only) and with a nicotine concentration of 16  
320 mg/ml (a typical dose of nicotine in commercially available nicotine-containing e-liquid) were  
321 collected and examined by GC-MS. The nicotine content of the e-liquid was also verified using  
322 our GC-MS methodology, which resulted in a nicotine concentration of 17.2 ( $\pm 0.5$ ) mg/ml,  
323 indicating a very good agreement between the commercial nicotine label (i.e. 16 mg/ml) and the  
324 actual nicotine content of the e-liquid. Indoor concentrations and emission rates, as well as I/O  
325 ratio of nicotine for normal cigarette as well as e-cigarette samples with and without the addition  
326 of nicotine are presented in Table 6. The indoor concentration and emission rate of particle-phase  
327 nicotine were both about 13 times higher during smoking of normal cigarette compared to the  
328 consumption of e-cigarette containing 16 mg/ml of nicotine in the e-liquid (1524 ng/m<sup>3</sup> versus

329 123 ng/m<sup>3</sup> and 91161 ng/hr versus 7103 ng/hr, respectively). Detectable levels of nicotine were  
330 found in particles collected during the experimental trials of e-cigarette without nicotine in the e-  
331 liquid. Nicotine is a semi-volatile compound, with strong affinity and persistence on indoor  
332 surfaces.<sup>29,30</sup> The observed (small) nicotine levels in e-cigarette samples without nicotine could,  
333 therefore, be due to particle-bound nicotine re-suspension in the room during sampling and can  
334 be considered as a “background” or “blank” emission level. Re-suspension of residual nicotine in  
335 indoor environments has been previously considered as one of the pathways of passive exposure  
336 to nicotine in houses with active smoker residents.<sup>30</sup> It should be mentioned that the  
337 concentration of residual and side-stream nicotine in an indoor environment can be significantly  
338 affected by gas-to-particle partitioning of nicotine, which is strongly dependent on the  
339 environmental conditions such as temperature, pressure and acidity of the aerosol.<sup>31,32</sup> In our  
340 study, the average consumption rate of the e-liquid was about 1.3 ml/hr, which, considering a  
341 nicotine concentration of 17.2 mg/ml in the e-liquid (measured by the GC-MS method), would  
342 result in a total emission rate of about 22.36 mg/hr nicotine from e-cigarettes (including both  
343 gas-phase and particle-phase nicotine). Based on our data, the indoor emission rate of particle-  
344 phase nicotine (after subtraction of nicotine level from e-cigarette without nicotine; which was  
345 considered as a “background” level) was about 4344 ng/hr, accounting for about 0.02 % of the  
346 total nicotine emission, implying that a trivial fraction of the total nicotine emissions from e-  
347 cigarettes is found in secondhand particulate emissions. Further investigation is required for  
348 accurate determination of nicotine’s gas-to-particle partitioning ratio in e-cigarette’s aerosol,  
349 which can provide useful insight about secondhand exposure to nicotine during e-cigarette  
350 vaping under different conditions.

351

#### 352 4. Summary and Conclusions

353 Analysis of secondhand emissions from a popular and widely-used e-cigarette brand indicated a  
354 very large reduction of particle-phase emissions compared to normal tobacco-containing  
355 cigarettes in a real-life setting. BC and particle-phase PAHs, deleterious chemical species present  
356 in high concentrations in tobacco smoke, were not detected in e-cigarette's aerosol. Emission  
357 rates of organic compounds (including alkanes and organic acids) as well as total emission of  
358 inorganic elements and metals were also significantly reduced (more than 100 times for organics  
359 and 10 times for elements) in e-cigarettes compared to normal cigarettes. Analysis of elemental  
360 emissions indicated the presence of toxic metals (such as Ni, Zn and Ag) in e-cigarette's aerosol,  
361 with Ni and Ag having higher indoor emission rates compared to normal cigarettes. Moreover,  
362 analysis of nicotine indicated that secondhand particle-phase nicotine accounted for about 21%  
363 of the total nicotine generation and emission during e-cigarette vaping. Based on our results, use  
364 of e-cigarettes from a public health perspective appears to be an improvement compared to  
365 normal tobacco-containing cigarettes, as exposure to most of the toxic and/or undesirable  
366 chemical species was found to be much lower than that for normal cigarettes. However,  
367 considering the lack of regulation on the manufacturing process of e-cigarettes, there appears to  
368 be a potential for utilization of toxic material (such as metals) in e-cigarettes, which could lead  
369 to their emission in e-cigarette's vapor and aerosol.<sup>1,13</sup> Implementing quality control regulations  
370 on the design and manufacturing process of e-cigarettes is therefore necessary to prevent  
371 potential utilization of non-desirable material in e-cigarettes and e-liquids.

372

373

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381 We wish to dedicate the work *in memoriam* of Giovanni Invernizzi.

382

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385

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435

## Figures and Tables

**Figure 1. Average outdoor and indoor (i.e. inside of the smoking room) mass concentration ( $\mu\text{g}/\text{m}^3$ ) during smoking and vaping of normal cigarettes and e-cigarettes. Error bars represent one standard error.**

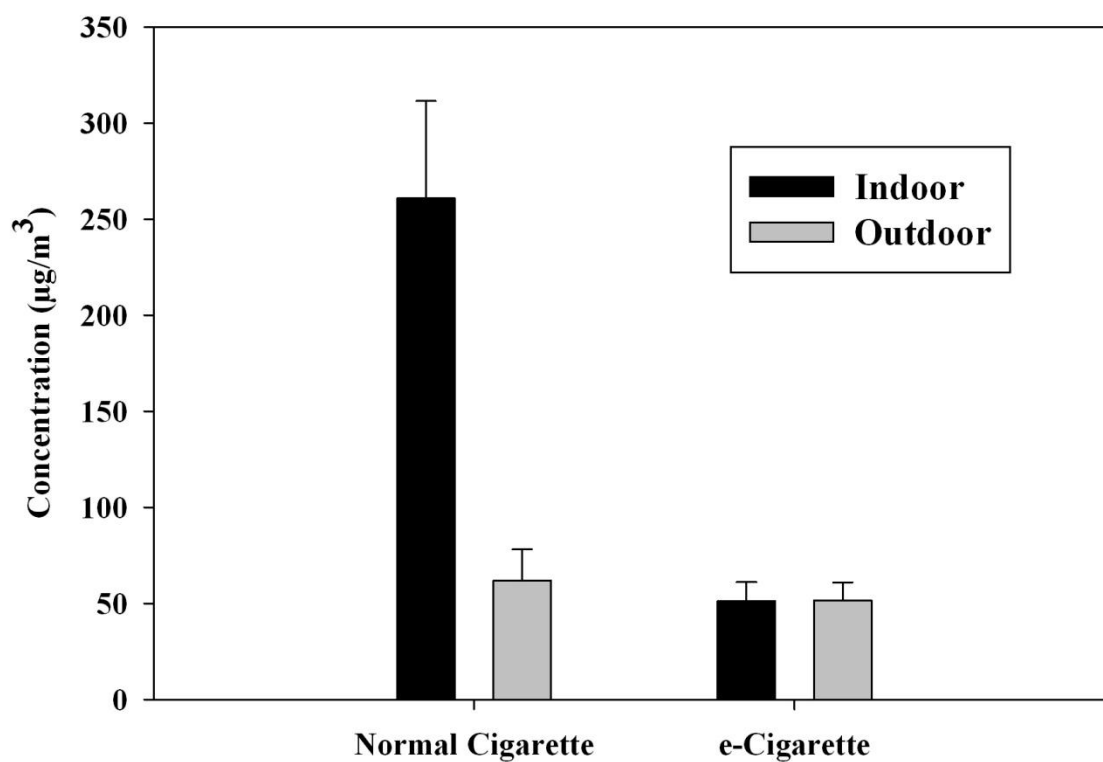


Figure 2. Average total concentration of elements as well as organic groups (polycyclic aromatic hydrocarbons (PAHs), hopanes, n-alkanes, organic acids and levoglucosan), normalized by PM mass concentration (ng/ $\mu$ g PM). Error bars represent analytical uncertainties.

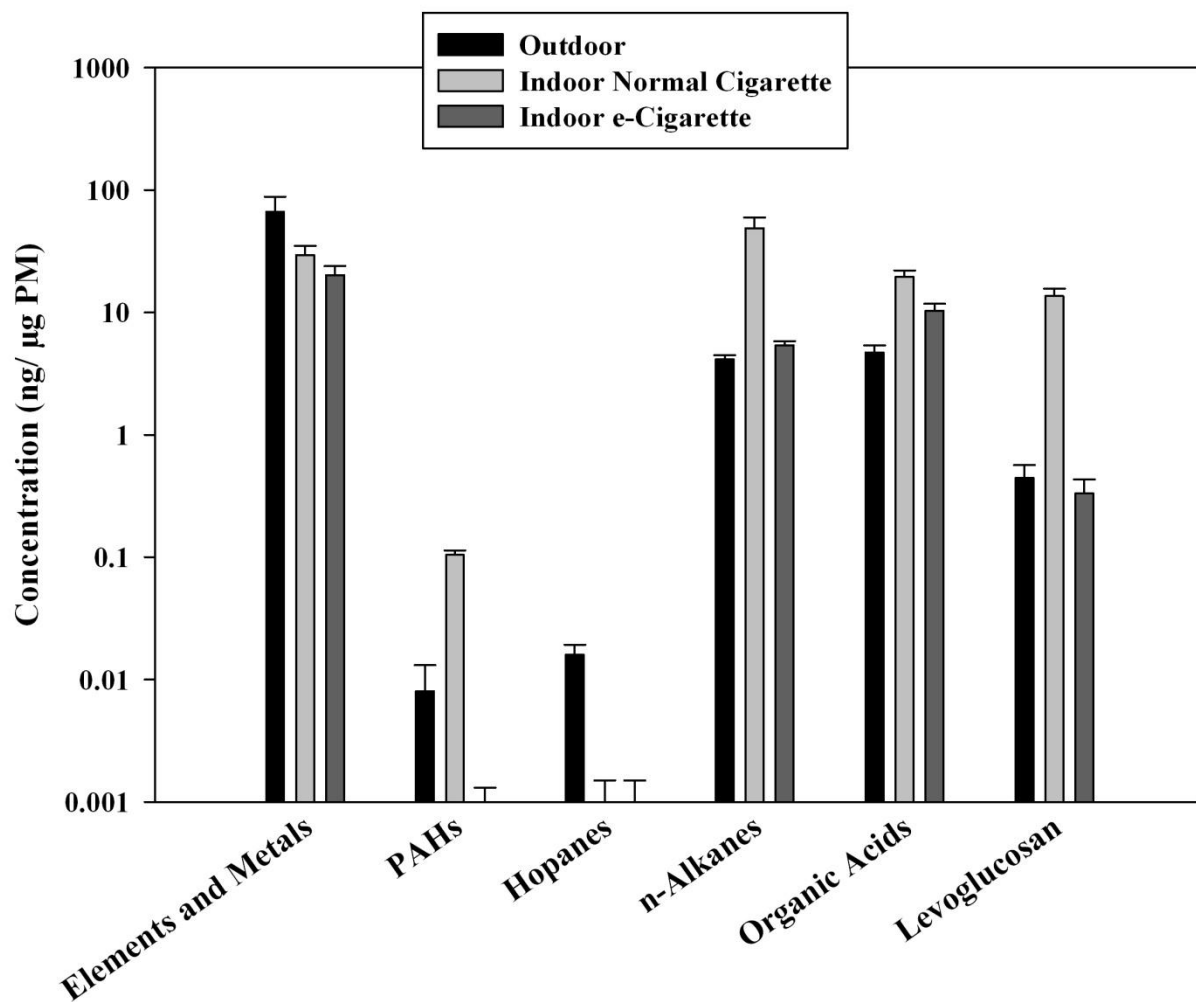
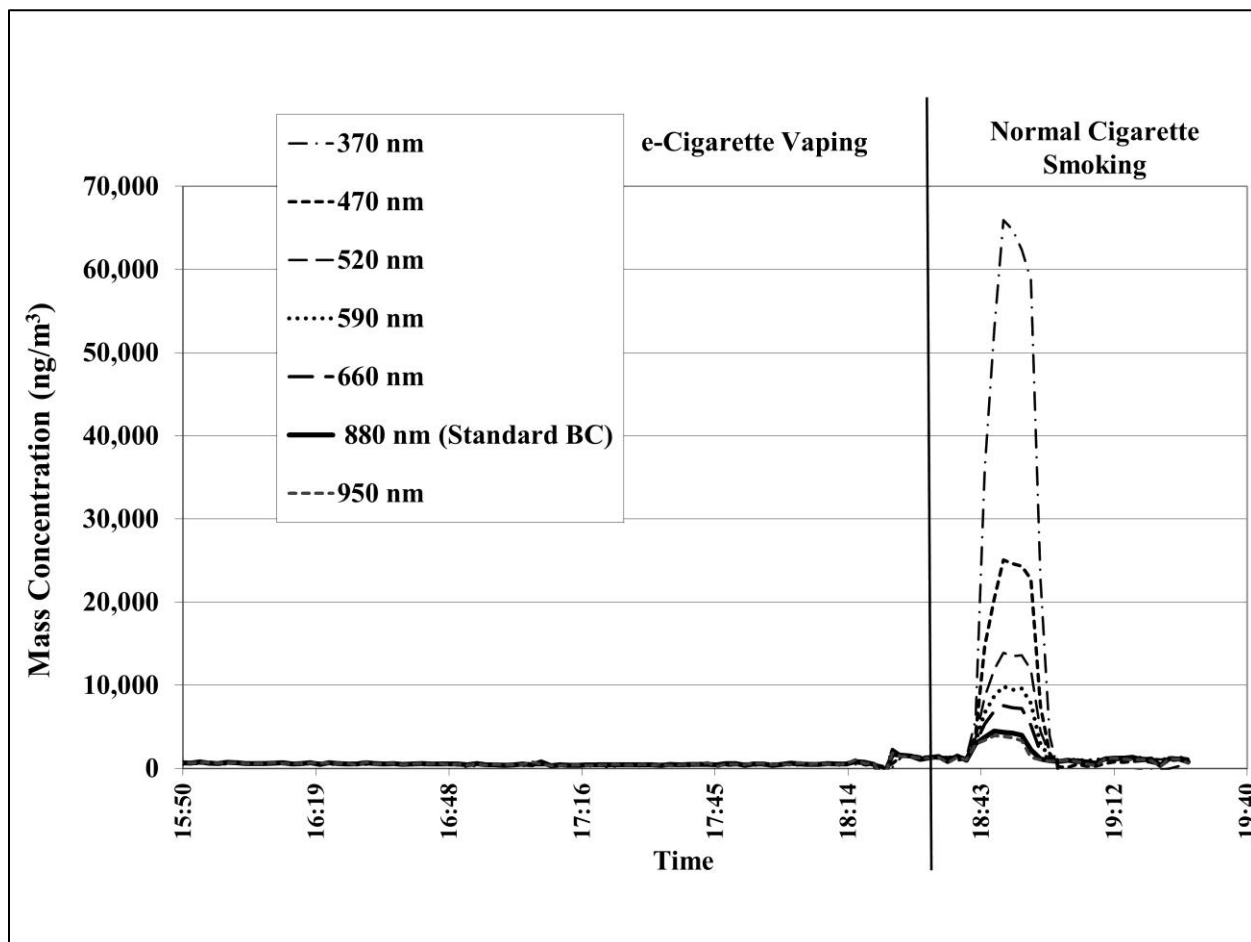




Figure 3. Comparison of indoor black carbon (BC) mass concentration ( $\text{ng}/\text{m}^3$ ) inside of the smoking room, between periods of normal cigarette and e-cigarette consumption.



**Table 1. Average indoor (i.e. inside of the smoking room) to outdoor mass concentration ratio and Pearson correlation coefficient (R) of elemental constituents of e-cigarette samples. Values correspond to average  $\pm$ standard error.**

<b>Element</b>	<b>In/Out Ratio</b>	<b>R</b>	<b>Element</b>	<b>In/Out Ratio</b>	<b>R</b>
<b>B</b>	<b>13.13 (<math>\pm</math>7.47)</b>	<b>0.17</b>	<b>Cu</b>	<b>0.43 (<math>\pm</math>0.2)</b>	<b>-0.49</b>
<b>Mg</b>	<b>0.10 (<math>\pm</math>0.06)</b>	<b>-0.35</b>	<b>Zn</b>	<b>1.26 (<math>\pm</math>0.33)</b>	<b>0.13</b>
<b>Al</b>	<b>0.48 (<math>\pm</math>0.11)</b>	<b>0.89</b>	<b>Rb</b>	<b>0.56 (<math>\pm</math>0.07)</b>	<b>0.26</b>
<b>S</b>	<b>0.81 (<math>\pm</math>0.27)</b>	<b>0.77</b>	<b>Sr</b>	<b>0.30 (<math>\pm</math>0.09)</b>	<b>0.82</b>
<b>K</b>	<b>1.53 (<math>\pm</math>0.56)</b>	<b>0.22</b>	<b>Mo</b>	<b>0.52 (<math>\pm</math>0.17)</b>	<b>0.61</b>
<b>Ca</b>	<b>0.18 (<math>\pm</math>0.16)</b>	<b>0.54</b>	<b>Ag</b>	<b>3.39 (<math>\pm</math>1.15)</b>	<b>-0.10</b>
<b>Ti</b>	<b>0.80 (<math>\pm</math>0.18)</b>	<b>0.45</b>	<b>Cd</b>	<b>0.91 (<math>\pm</math>0.2)</b>	<b>0.66</b>
<b>V</b>	<b>0.48 (<math>\pm</math>0.18)</b>	<b>0.37</b>	<b>Sn</b>	<b>0.37 (<math>\pm</math>0.33)</b>	<b>-0.18</b>
<b>Cr</b>	<b>0.92 (<math>\pm</math>0.44)</b>	<b>-0.12</b>	<b>Sb</b>	<b>0.30 (<math>\pm</math>0.05)</b>	<b>0.95</b>
<b>Mn</b>	<b>0.33 (<math>\pm</math>0.1)</b>	<b>-0.09</b>	<b>La</b>	<b>1.47 (<math>\pm</math>0.68)</b>	<b>0.01</b>
<b>Fe</b>	<b>0.26 (<math>\pm</math>0.1)</b>	<b>-0.27</b>	<b>W</b>	<b>0.85 (<math>\pm</math>0.38)</b>	<b>0.08</b>
<b>Co</b>	<b>0.44 (<math>\pm</math>0.23)</b>	<b>-0.07</b>	<b>Pb</b>	<b>0.89 (<math>\pm</math>0.26)</b>	<b>0.00</b>
<b>Ni</b>	<b>1.75 (<math>\pm</math>0.72)</b>	<b>-0.16</b>			

Figure 4. Average indoor (i.e. inside of the smoking room) concentration ( $\text{ng}/\text{m}^3$ ) of elements with average indoor to outdoor mass ratio greater than unity, compared to their estimated concentration caused by the e-liquid's elemental content. Error bars represent one standard error for the former and analytical uncertainties for the latter.

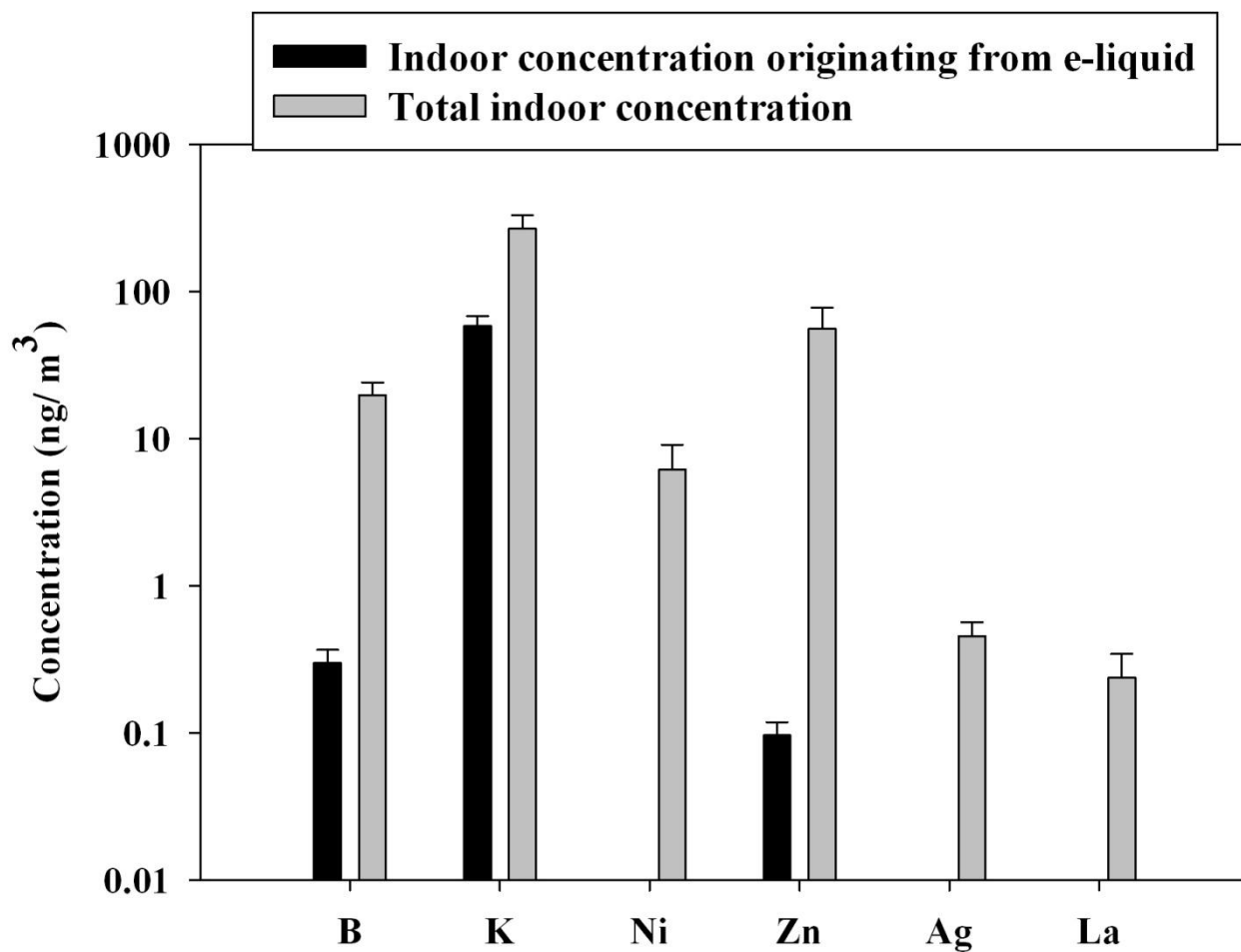
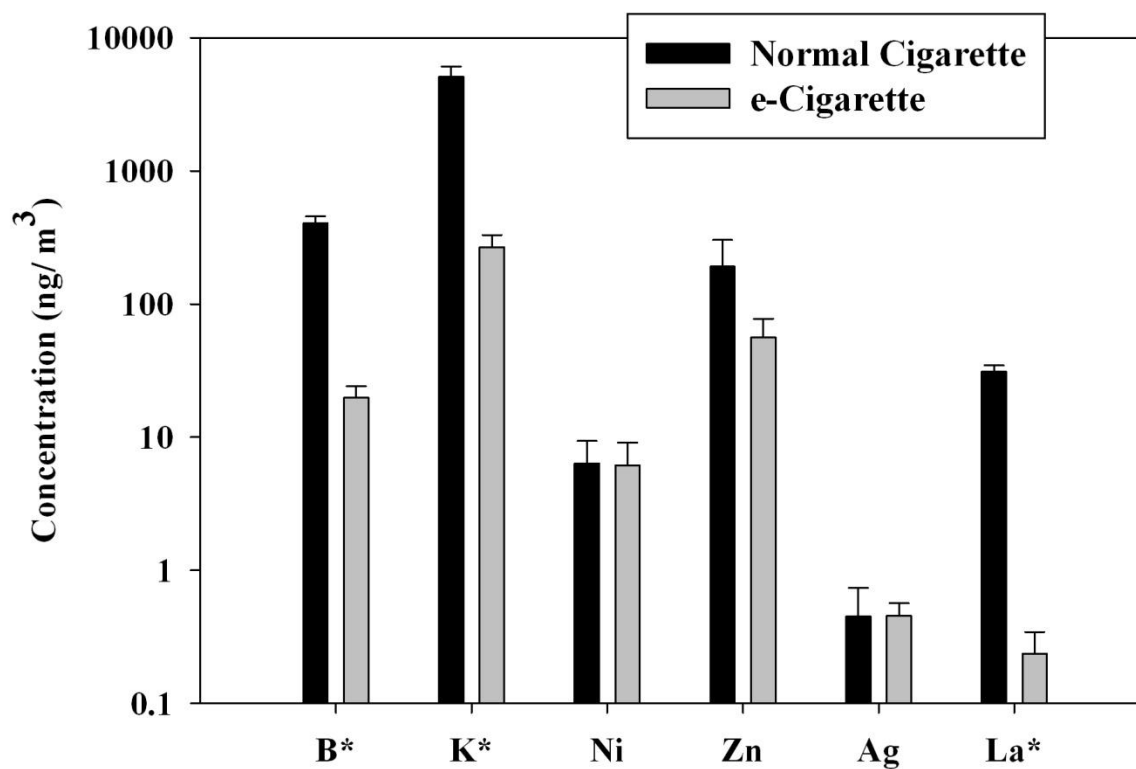


Figure 5. Average indoor (i.e. inside of the smoking room) levels ( $\text{ng}/\text{m}^3$ ) of e-cigarette versus normal cigarette emissions pertaining to the elements with average indoor to outdoor mass ratio greater than unity. Error bars represent one standard error.



\*Indicates elements with statistically significant difference ( $p < 0.05$ ) between regular cigarettes and e-cigarettes.

**Table 2. Average indoor (i.e. inside of the smoking room) to outdoor mass concentration ratio and Pearson correlation coefficient (R) of organic species for e-cigarette samples. Polycyclic aromatic hydrocarbons (PAHs) were undetected for indoor e-cigarette samples and therefore not included in this table. Values correspond to average  $\pm$ standard error.**

Species	I/O Ratio	R	Species	I/O Ratio	R
n-Eicosane	1.72 ( $\pm$ 0.33)	0.97	Tetradecanoic Acid	8.25 ( $\pm$ 1.46)	-0.41
n-Docosane	1.83 ( $\pm$ 0.33)	0.93	Pentadecanoic Acid	8.3 ( $\pm$ 1.34)	-0.4
n-Tetracosane	8.55 ( $\pm$ 3.06)	0.63	Hexadecanoic Acid	13.57 ( $\pm$ 5.5)	-0.3
n-Pentacosane	8.35 ( $\pm$ 4.08)	0.55	Heptadecanoic Acid	23.12 ( $\pm$ 5.55)	-0.62
n-Hexacosane	2.2 ( $\pm$ 0.3)	0.72	Nonadecanoic Acid	3.91 ( $\pm$ 2.26)	0.1
n-Heptacosane	3.15 ( $\pm$ 2.08)	0.25	Palmitoleic Acid	37.98 ( $\pm$ 8.05)	-0.86
Nonacosane	2.26 ( $\pm$ 1.14)	-0.26	Linoleic Acid	4.53 ( $\pm$ 1.35)	0.73
Triacontane	7.39 ( $\pm$ 4.4)	-0.16	Eicosanoic Acid	3.46 ( $\pm$ 2.26)	0.56
Hentriacontane	2.5 ( $\pm$ 1.25)	-0.07	Docosanoic Acid	1.52 ( $\pm$ 0.36)	0.9
Dotriacontane	2.55 ( $\pm$ 1.18)	-0.2	Tricosanoic Acid	1.65 ( $\pm$ 0.31)	0.16
Trtriacontane	1.52 ( $\pm$ 0.43)	-0.25	Tetracosanoic Acid	2.72 ( $\pm$ 0.56)	0.15
Tetratriacontane	1.49 ( $\pm$ 0.43)	-0.29	Pentacosanoic Acid	8.06 ( $\pm$ 2.64)	-0.65
Pentatriacontane	1.41 ( $\pm$ 0.34)	-0.25	Hexacosanoic Acid	1.87 ( $\pm$ 0.27)	-0.07
Hexatriacontane	1.45 ( $\pm$ 0.34)	-0.34	Octacosanoic Acid	1.31 ( $\pm$ 0.14)	-0.55
Heptatriacontane	1.3 ( $\pm$ 0.26)	-0.31	Triacontanoic Acid	5.3 ( $\pm$ 4.15)	-0.65
Octatriacontane	1.5 ( $\pm$ 0.33)	-0.48	Suberic Acid	1.66 ( $\pm$ 0.47)	-0.26
Decanoic Acid	5.91 ( $\pm$ 1.81)	0.82	Azelaic Acid	1.79 ( $\pm$ 0.47)	-0.07
Dodecanoic Acid	2.5 ( $\pm$ 0.6)	0.68	Levoglucosan	0.77 ( $\pm$ 0.31)	0.99

**Table 3. Average emission rates (ng/hr) of selected metals and elements during e-cigarette and normal cigarette consumption. Values correspond to average  $\pm$ propagated uncertainty. “N.D” represents the emission rates that were not detectable based on the mass balance model (i.e. indicating zero emission rate from indoor sources).**

Species	e-Cigarette (ng/hr)	Normal-Cigarette (ng/hr)	Species	e-Cigarette (ng/hr)	Normal-Cigarette (ng/hr)
<b>B</b>	<b>963.8 (<math>\pm</math>30.18)</b>	<b>23680 (<math>\pm</math>582.9)</b>	<b>Cu</b>	<b>N.D</b>	<b>1029 (<math>\pm</math>113.8)</b>
<b>Mg</b>	<b>N.D</b>	<b>N.D</b>	<b>Zn</b>	<b>1142 (<math>\pm</math>143.8)</b>	<b>8252 (<math>\pm</math>332.3)</b>
<b>Al</b>	<b>N.D</b>	<b>N.D</b>	<b>Rb</b>	<b>N.D</b>	<b>200.1 (<math>\pm</math>6.450)</b>
<b>S</b>	<b>N.D</b>	<b>34540 (<math>\pm</math>1580)</b>	<b>Sr</b>	<b>N.D</b>	<b>N.D</b>
<b>K</b>	<b>7765 (<math>\pm</math>560.3)</b>	<b>297500 (<math>\pm</math>7044)</b>	<b>Mo</b>	<b>N.D</b>	<b>N.D</b>
<b>Ca</b>	<b>N.D</b>	<b>N.D</b>	<b>Ag</b>	<b>20.91 (<math>\pm</math>0.730)</b>	<b>14.65 (<math>\pm</math>0.900)</b>
<b>Ti</b>	<b>50.16 (<math>\pm</math>26.29)</b>	<b>N.D</b>	<b>Cd</b>	<b>0.480 (<math>\pm</math>0.300)</b>	<b>657.3 (<math>\pm</math>15.10)</b>
<b>V</b>	<b>N.D</b>	<b>N.D</b>	<b>Sn</b>	<b>N.D</b>	<b>N.D</b>
<b>Cr</b>	<b>28.10 (<math>\pm</math>13.64)</b>	<b>N.D</b>	<b>Sb</b>	<b>N.D</b>	<b>N.D</b>
<b>Mn</b>	<b>N.D</b>	<b>N.D</b>	<b>La</b>	<b>3.210 (<math>\pm</math>0.690)</b>	<b>1846 (<math>\pm</math>45.36)</b>
<b>Fe</b>	<b>N.D</b>	<b>N.D</b>	<b>W</b>	<b>N.D</b>	<b>N.D</b>
<b>Co</b>	<b>N.D</b>	<b>N.D</b>	<b>Pb</b>	<b>96.16 (<math>\pm</math>29.93)</b>	<b>1012 (<math>\pm</math>248.7)</b>
<b>Ni</b>	<b>130.5 (<math>\pm</math>15.73)</b>	<b>36.39 (<math>\pm</math>10.42)</b>			

**Table 4. Average emission rates (ng/hr) of selected alkanes and organic acids during e-cigarette and normal cigarette consumption. Values correspond to average  $\pm$ propagated uncertainty**

Species	e-Cigarette (ng/hr)	Normal-Cigarette (ng/hr)	Species	e-Cigarette (ng/hr)	Normal-Cigarette (ng/hr)
n-Eicosane	529.3 ( $\pm$ 40.54)	11240 ( $\pm$ 287.3)	Tetradecanoic Acid	8308 ( $\pm$ 240.3)	16100 ( $\pm$ 397.6)
n-Docosane	477.3 ( $\pm$ 31.67)	9407 ( $\pm$ 240.1)	Pentadecanoic Acid	2289 ( $\pm$ 65.12)	7685 ( $\pm$ 183.2)
n-Tetracosane	604.9 ( $\pm$ 20.12)	5131 ( $\pm$ 127.1)	Hexadecanoic Acid	13960 ( $\pm$ 395.4)	129300 ( $\pm$ 3098)
n-Pentacosane	255.5 ( $\pm$ 9.143)	5765 ( $\pm$ 138.6)	Heptadecanoic Acid	572.4 ( $\pm$ 20.12)	8113 ( $\pm$ 189.6)
n-Hexacosane	125.5 ( $\pm$ 6.123)	3593 ( $\pm$ 83.49)	Palmitoleic Acid	1813 ( $\pm$ 46.77)	8308 ( $\pm$ 190.2)
Triacontane	241.4 ( $\pm$ 21.02)	23490 ( $\pm$ 545.4)	Linoleic Acid	444.1 ( $\pm$ 14.78)	65100 ( $\pm$ 1477)
Hentriacontane	317.2 ( $\pm$ 34.21)	165900 ( $\pm$ 3934)	Eicosanoic Acid	136.4 ( $\pm$ 6.983)	13900 ( $\pm$ 318.2)
Dotriacontane	312.6 ( $\pm$ 37.67)	35900 ( $\pm$ 839.4)	Docosanoic Acid	160.4 ( $\pm$ 10.36)	12570 ( $\pm$ 287.4)
Trtriacontane	274.3 ( $\pm$ 38.72)	94420 ( $\pm$ 2115)	Tricosanoic Acid	112.6 ( $\pm$ 7.978)	6479 ( $\pm$ 157.2)
Tetratriacontane	284.1 ( $\pm$ 39.78)	4706 ( $\pm$ 114.3)	Tetracosanoic Acid	449.3 ( $\pm$ 18.13)	10260 ( $\pm$ 239.9)
Pentatriacontane	220.1 ( $\pm$ 28.34)	3887 ( $\pm$ 98.88)	Pentacosanoic Acid	208.2 ( $\pm$ 8.873)	2842 ( $\pm$ 65.67)
Hexatriacontane	228.5 ( $\pm$ 26.35)	943.1 ( $\pm$ 29.34)	Hexacosanoic Acid	218.6 ( $\pm$ 12.84)	3505 ( $\pm$ 84.23)
Heptatriacontane	153.7 ( $\pm$ 19.55)	1110 ( $\pm$ 30.33)	Octacosanoic Acid	222.2 ( $\pm$ 18.93)	9880 ( $\pm$ 226.2)
Octatriacontane	208.8 ( $\pm$ 20.93)	776.6 ( $\pm$ 24.46)	Triacontanoic Acid	228.6 ( $\pm$ 18.95)	6720 ( $\pm$ 163.8)
Decanoic Acid	229.2 ( $\pm$ 8.981)	1368 ( $\pm$ 32.82)	Suberic Acid	282.3 ( $\pm$ 20.13)	2264 ( $\pm$ 60.23)
Dodecanoic Acid	2421 ( $\pm$ 102.2)	12270 ( $\pm$ 278.8)	Azelaic Acid	743.5 ( $\pm$ 48.80)	4979 ( $\pm$ 134.8)

**Table 5. Average emission rate (ng/hr) of polycyclic aromatic hydrocarbons (PAHs) during e-cigarette and normal cigarette consumption. Values correspond to average  $\pm$ propagated uncertainty. “N.D” represents the emission rates that were not detectable based on the mass balance model (i.e. indicating zero emission rate from indoor sources)**

Species	e-Cigarette (ng/hr)	Normal-Cigarette (ng/hr)
Chrysene	N.D	213.3 ( $\pm$ 5.983)
Benzo(b)fluoranthene	N.D	307.2 ( $\pm$ 7.237)
Benzo(k)fluoranthene	N.D	130.4 ( $\pm$ 3.235)
Benzo (e) pyrene	N.D	105.6 ( $\pm$ 3.982)
Benzo(a)pyrene	N.D	281.7 ( $\pm$ 6.873)
Indeno(1,2,3-cd)pyrene	N.D	270.2 ( $\pm$ 6.532)
Benzo(g,h,i)perylene	N.D	187.0 ( $\pm$ 4.231)



**Table 6. Indoor to outdoor mass concentration ratio and emission rates (ng/hr) of nicotine for e-cigarette samples with and without nicotine as well as normal cigarette. Values correspond to average  $\pm$ standard error for indoor to outdoor ratios and mass concentrations, and average  $\pm$ propagated uncertainty for emission rates.**

<b>Parameter</b>	<b>Sample</b>	<b>Nicotine</b>
<b>Concentration (ng/m<sup>3</sup>)</b>	<b>Normal Cigarette</b>	<b>1524 (<math>\pm</math>80.4)</b>
	<b>e-Cigarette</b>	<b>60.68 (<math>\pm</math>20.91)</b>
	<b>e-Cigarette+Nicotine</b>	<b>123.0 (<math>\pm</math>34.5)</b>
<b>Indoor/outdoor mass ratio</b>	<b>Normal Cigarette</b>	<b>254.3 (<math>\pm</math>13.4)</b>
	<b>e-Cigarette</b>	<b>6.7 (<math>\pm</math>3.5)</b>
	<b>e-Cigarette+Nicotine</b>	<b>18.6 (<math>\pm</math>7)</b>
<b>Emission Strength (ng/hr)</b>	<b>Normal Cigarette</b>	<b>91161 (<math>\pm</math>2170)</b>
	<b>e-Cigarette</b>	<b>2759 (<math>\pm</math>93)</b>
	<b>e-Cigarette+Nicotine</b>	<b>7103 (<math>\pm</math>172)</b>