

Analytical Methods

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Short communication**A highly sensitive liquid chromatography-tandem mass spectrometry method for the analysis of a toxic water disinfection by-product, N-nitrosomethylethylamine**

Yassine Kadmi^{1*}, Lidia Favier¹, Mouni Lotfi², Nouredine Nasrallah³, Dominique Wolbert¹

¹Ecole Nationale Supérieure de Chimie de Rennes, CNRS, UMR 6226, 11 Allée de Beaulieu, CS 50837, 35708 Rennes Cedex 7, France.

²Faculté des Sciences de la Nature et de la Vie et des Sciences de la Terre. Université Akli Mohand Oulhadj, Bouira, Algérie.

³Faculté de Génie Mécanique et Génie des Procédés, laboratoire de Génie de la Réaction Chimique, BP 32 El-Alia, Bab-Ezzouar, 16000 Alger, Algérie.

*Corresponding author:

Kadmi Yassine

Tel.: +33 223238134

Fax: +33 223238120

E-mail address: yassine.kadmi@gmail.com

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Abstract

Recently, among the emerging contaminants, N-nitrosomethylethylamine has become of special concern because it is a potent human mutagenic and carcinogenic contaminant has been detected chlorinated or chloraminated drinking waters and wastewaters. In this work a sensitive and robust method, which was based on solid-phase extraction followed by ultra-high-pressure liquid chromatography coupled with tandem mass spectrometry, was developed for the determination of N-nitrosomethylethylamine in water at ultra-trace levels. Chromatographic separation was performed on a C18 column. Quantification of the N-nitrosomethylethylamine was achieved by a triple quadrupole mass spectrometer that was equipped with an electrospray interface and was operating in positive ionization mode. Under optimized conditions, the calibration curve was linear from 0.1 to 100 $\mu\text{g/L}$ ($r^2 \geq 0.999$). The precision of the intra- and inter-day values was found to be less than 2.5%, and the accuracy of the method was within $\pm 3\%$. Moreover, an extraction efficiency greater than 86% was obtained at different concentration levels with relative standard deviation, $\text{RSD} < 4.2\%$. Therefore, the experimental results showed that the proposed analytical method can be used successfully to determine N-nitrosomethylethylamine at ultra-trace levels (ng/L) in aqueous samples.

63

1 Introduction

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66 A wide array of disinfection by-products, including N-nitrosamines, is formed during
67 water treatment using chlorination and chloramination processes.^{1,2} These compounds
68 comprise a group of mutagenic chemicals that have been classified as probable human
69 carcinogens.³ In recent years, N-nitrosomethylethylamine (NMEA), which is a non-
70 halogenated N-nitrosamine, has attracted considerable attention because it is frequently
71 detected in drinking water in many countries around the world.^{4,5}

72

73 The United States Environmental Protection Agency (U.S. EPA) has classified NMEA
74 into the B2 group (probable carcinogenic effects on humans) and indicated that this
75 compound produces an increased cancer risk at the 10^{-6} level at the very low concentration of
76 20 ng/L.⁶ Consequently, sensitive and reliable analytical techniques to determine ultra-trace
77 levels of NMEA in water are required. Due to the low concentration levels of this compound in
78 environmental samples, extraction and pre-concentration steps are necessary. Solid-phase
79 micro-extraction (SPME)^{7,8} and solid-phase extraction (SPE)^{9,10} have been used to
80 preconcentrate NMEA in water samples. However, the SPME method has some limitations
81 such as the possibility of the sample contamination, low extraction recoveries, low pre-
82 concentration factor, and high detection limits. Furthermore, it is especially used for the
83 extraction of volatile organic molecules and is particularly combined with gas
84 chromatography. As an alternative method, SPE was successfully applied to the extraction of
85 a wide variety of compounds such as volatile and non-volatile organic compounds from
86 environmental water samples.

87

88 Several analytical techniques have been developed for the quantification of NMEA. They are
89 based on liquid chromatography (LC) and gas chromatography (GC). The analyte has been
90 analyzed in water samples by using GC technique coupled with different types of detectors,
91 such as nitrogen chemiluminescence detection (NCD)¹¹, nitrogen phosphorous detection
92 (NPD)¹², mass spectrometry (MS)¹³⁻¹⁵, and tandem mass spectrometry (GC/MS/MS).^{16,17}
93 However, these techniques are limited to the analysis of volatile and thermally stable
94 compounds. Moreover, liquid chromatography-tandem mass spectrometry (LC/MS/MS)^{18,19}
95 methods have also been reported for the determination of NDMA or other N-nitrosamines in
96 water samples. To date, only a few and recent LC-MS/MS methods have been reported for the
analysis of N-nitrosamines in water samples. Plumlee *et al.* in 2008¹⁸ described a optimized

1
2 97 method only for NDMA determination. More recently, Ripolles *et al.* reported a SPE-LC-
3 98 MS/MS combined with a triple quadrupole analyzer using a atmospheric pressure chemical
4 ionization (APCI) mode for the analysis of NDMA and other N-nitrosamines in drinking
5 99 water samples. For NMEA the achieved recoveries were only of 64 - 88% and the estimated
6 limit of detection was found to be 5 ng/L.¹⁹ However, to the best of our knowledge, the
7 100 analytical method developed in this work is the first UHPLC/MS/MS method that has been
8 101 proposed for the determination of NMEA at ultra-trace concentrations providing high
9 102 sensitivity and high SPE recoveries (between 85% to 97%).
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12 105 The scope of the research reported in this paper was to develop a simple, rapid, sensitive,
13 106 and reliable analytical method for the analysis of ultra-trace levels of NMEA in water by
14 107 combining solid-phase extraction with ultra-high-pressure liquid chromatography–
15 108 electrospray ionization–tandem mass spectrometry (SPE-UHPLC-(ESI)-MS/MS).
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110 2 Materials and methods

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112 A NMEA standard solution (2000 mg/L in methanol) was purchased from LGC
113 Standards (Wesel, Germany). All chemical reagents used in this work (for SPE procedure,
114 solution preparation and LC/MS/MS measurements) were of the highest analytical purity
115 grade. Acetonitrile and formic acid were obtained from J.T. Baker (Deventer, Netherlands).
116 Methanol and dichloromethane, obtained from Fischer Scientific-Bioblock (Illkirch, France),
117 were of LC-MS grade. Acetic acid was supplied by Acros Organics (Noisy-le-Grand, France).
118 Stock solutions (100 mg/L in methanol) for NMEA were prepared and stored at -20°C for at
119 least three months. The working solutions were freshly prepared by a series of dilutions with
120 acetonitrile/ultrapure water (60:40, v/v). The ultrapure water was produced by an Elga
121 Option-Q DV-25 system (Antony, France). Surface waters samples were collected from a
122 river (Britanny region, France) and stored at 4°C until SPE extraction and analysis (within
123 one week of collection).
124

125 Chromatographic separation was performed on an Acquity™ UHPLC H-Class system
126 (Waters, Saint-Quentin en Yvelines, France), with a BEH C18 column (100 mm × 2.1 mm i.d.,
127 1.7 µm; Waters) maintained at 45°C. The mobile phase that was used consisted of formic acid
128 in acetonitrile and water (60:40:0.1, v/v/v). The flow rate was 0.4 mL/min and the injection
129 volume was 5 µL. The total run time was two minutes. The UHPLC system was connected to
130 a Waters Acquity™ triple quadrupole mass spectrometer (MS/MS). The positive ionization

1
2 131 tandem MS detection in multiple reaction monitoring (MRM) mode was used.
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5 133 The analyte was extracted, purified, and concentrated from water samples using a Sep-
6 Pak Plus[®] AC-2 cartridge (400 mg, 85 μ m; Waters, Guyancourt, France). Then, the extract was
7 134 evaporated to a final volume of approximately 100 μ L in an N-Evap system (Organomation,
8 Berlin, MA, USA) under a high-purity nitrogen stream. For the SPE procedure, several factors
9 135 were optimized in order to obtain high recoveries in ultrapure water, i.e., cartridge
10 136 conditioning, pH values of the samples, loading rates, washing conditions, and the elution
11 137 volumes. Then the selected SPE technique was used to determine the recovery of NMEA in a
12 138 real water samples.
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18 143 The performance and reliability of the proposed method were assessed by determining the
19 144 regression equation, linear range, analyte detectability, precision, accuracy, and extraction
20 145 recovery for the N-nitrosamine studied.
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23 147 Linearity of the proposed method was performed by direct injection of seven working
24 148 solutions, prepared in ultrapure water in the concentration range from 1 to 100 μ g/L. Each
25 149 solution was analyzed in triplicate. The calibration curves were constructed by a least squares
26 150 linear regression analysis. This method was used to determine the slope, intercept, and
27 151 correlation coefficient (r^2) of the linear regression equation. The LOD and the LOQ values
28 152 were determined at concentrations with a signal-to-noise ratio (S/N) of 3 and 10, respectively.
29 153 The instrument limit of detection (LOD) is the lowest concentration of analyte that the
30 154 analytical process can reliably differentiate from background levels, while the instrument
31 155 limit of quantification (LOQ) is the lowest concentration of analyte that can be quantified.
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34 157 The intra-day and inter-day precision of the analyses was estimated in terms of
35 158 repeatability. These parameters were expressed as relative standard deviation (RSD, %). The
36 159 accuracy (RE, %) was expressed by $100 - [(mean\ observed\ concentration)/(spiked\ concentration)] \times 100$. Moreover, the RSD calculated at each concentration level was not
37 160 allowed to exceed 15%, and the RE had to be within $\pm 15\%$ of the actual value.
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40 162 The extraction recovery (R, %) was calculated using the following procedure: a sample
41 163 spiked with the analyte was extracted using the developed solid phase extraction procedure
42 164 and the analysis result was compared to that of an unextracted standard which was prepared at
43 165 the equivalent final concentration. So, the extraction recovery was calculated as the ratio
44 166 between the resulting peak areas of the extracted and non extracted samples.
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47 168 The matrix effect (ME = A/B) was evaluated by calculating the ratio of the peak area in
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166 the presence of the matrix (A: samples spiked after extraction) to the peak area in absence of
167 the matrix (B: pure standard solution). In this work, the matrix effect was estimated by using
168 real environmental water samples.

169 170 **3 Results and discussion**

171
172 Different mobile phases (i.e., acetonitrile/water and methanol/water) containing acetic
173 acid or formic acid and mobile phase flow rates were tested and compared for the NMEA
174 analysis by UHPLC/MS/MS. Finally, acetonitrile/water containing 0.1% formic acid was used
175 as the mobile phase due to the good separation and high sensitivity to NMEA. The results
176 demonstrated that the flow rate of the mobile phase of 0.4 mL/min achieved satisfactory
177 separation, limited the dilution of the analyte chromatographic peak, and allowed a low
178 solvent requirement. Moreover, the effect of column temperature was also examined. Column
179 temperatures from 35 to 50°C were assayed, and 45°C was selected. Under these conditions,
180 the analysis time was two minutes. Fig. 1 shows the typical UHPLC/MS/MS chromatogram
181 obtained for ultrapure water spiked with NMEA obtained under optimized conditions.

182
183 For the MS/MS detector, the result showed that electrospray operation in positive
184 ionization mode (ESI) was better and had excellent signal sensitivity. In order to achieve the
185 quantification of NMEA, the mass spectrometric parameters, such as collision energy and
186 cone voltage, were optimized to attain the maximum sensitivity for the detection of the
187 analyte. The precursor ions and product ions were observed in the MS/MS spectra after
188 infusing the standard solution (1 mg/L) into the mass spectrometer. In this work, two sensitive
189 MRM transitions were selected for the N-nitrosamine that was studied. Different conditions of
190 the cone voltage, source temperature, and the collision energy were tested. The optimized
191 MS/MS transitions used for the UHPLC/MS/MS analysis, as well as specific cone voltage,
192 source temperature, collision energy, and segment periods, are provided in Table 1.

193
194 SPE extraction and concentration of the analyte was achieved with the Sep-Pak Plus[®] AC-
195 2 cartridge. To establish a SPE method for NMEA extraction the effects of several parameters
196 influencing the extraction efficiency, such as organic solvents and their volume, pH of the
197 samples, loading rates, washing conditions, and the elution volume, were investigated and
198 optimized in detail in this study. The selected SPE enrichment conditions were sample
199 conditioning with methanol, dichloromethane, acetonitrile (8 mL of each), and 5 mL of water.

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2 200 Then, 250 mL of the water sample spiked with the analyte and acidified with formic acid to pH
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4 201 2 was loaded at the optimum flow rate (approximately 3 mL/min). After the sample solution
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6 202 had passed through, the cartridge was washed with 5 mL of ultrapure water adjusted to pH 2
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8 203 to remove co-adsorbed matrix materials from the cartridge. Subsequently, the analyte retained
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10 204 on the SPE cartridge was eluted with 6 mL of dichloromethane, 4 mL of acetonitrile, and 2
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12 205 mL of methanol at a flow rate that ranged from 2 to 3 mL/min. Solvents are carefully
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14 206 evaporated (at 20-25°C) and concentrated under a high-purity nitrogen stream to a volume of
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16 207 50 µL. The obtained extracts are brought up to a final volume of 100 µL using
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18 208 acetonitrile/ultrapure water (60:40, v/v). Finally, the extract was stored at 4°C until further
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20 209 analysis was performed by UHPLC/MS/MS.
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211 The calibration curves showed good linearity ($r^2 \geq 0.999$) over the concentration range of
212 1 to 100 µg/L for NMEA in water samples. The linear regression equation of the calibration
213 curve was $y = 27.1761x + 23.4682$, where y represents the peak area and x represents the
214 concentration of the analyte. The instrumental limit of detection (LOD) and the instrumental
215 limit of quantification (LOQ) of NMEA were 1 and 2 µg/L, respectively.
216

217 The intra-day precision, inter-day precision, and accuracy of the method were evaluated
218 by spiking NMEA in ultrapure water at three quality control levels (1, 2 and 20 µg/L). The
219 intra-day precision and inter-day precision were less than 2.5% and 3% (RSD, %),
220 respectively. The accuracy ranged from 100 to 103%. The detailed values of intra-day, inter-
221 day precision, and accuracy are shown in Table 2. All the values were within the 15%
222 acceptable range. Therefore, the UHPLC/MS/MS method proved to be precise and accurate.
223

224 The SPE extraction recoveries were established by analyzing spiked ultrapure water
225 samples (N = 6) at three quality control concentration levels. The calculated extraction
226 recoveries of the NMEA were greater than 86%, and the relative standard deviations were less
227 than 4.3% (Table 3). Therefore, the SPE-UHPLC/MS/MS method that we developed allowed
228 quantification limits in the range of ng/L (considering that the pre-concentration factor of the
229 SPE method is 2500). Under these conditions the detection limit and the quantification limit
230 of the overall analytical procedure were 0.4 and 0.8 ng/L, respectively. As illustrated in Table
231 3, these values are lower than the ones reported in the literature^{1,19} for the LC-MS/MS
232 methods confirming the performance of the developed procedure.
233

234 For the calculations, the matrix effect of NMEA was evaluated by analyzing spiked
235 samples (N = 6) at three different concentration levels. Moreover, the presence of co-extracted
236 matrix components may severely affect the quantification of the analyte by UHPLC-ESI-
237 MS/MS. The matrix effect of NMEA was found to be within the acceptable range; all
238 recovery values ranged from 85% to 97% and the relative standard deviations were less than
239 4.5% in the river water samples. The results, as well as the satisfactory recoveries of NMEA
240 in river waters, are shown in Table 4.

241

242 4 Conclusion

243

244 In conclusion, the UHPLC-ESI-MS/MS method that was developed in this work showed
245 good linearity, precision, and accuracy for the determination of NMEA in water. Furthermore,
246 the SPE method using Sep-Pak Plus[®]AC-2 cartridges provided high recoveries for the
247 extraction and concentration of the analyte from environmental water samples. The SPE-
248 UHPLC-(ESI)-MS/MS analytical method can be considered to be a promising technique that
249 has obvious advantages over conventional analytical techniques in this field of application.
250 On the other hand UHPLC-ESI-MS/MS under positive mode of ionization provides high
251 sensitivity for the determination and quantification of NMEA in real water samples at ultra-
252 trace levels (ng/L).

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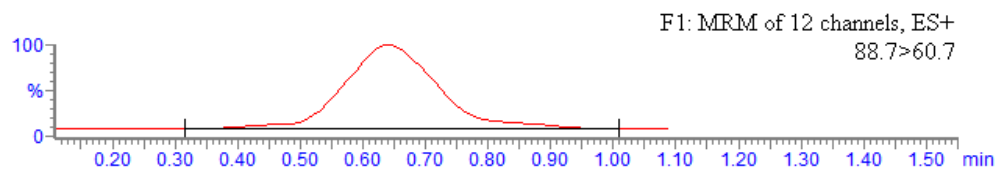
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16 **Fig. 1.** UHPLC-(ESI⁺)-MS/MS chromatograms obtained from the analysis of NMEA standard
17 at 5 µg/L (only quantification transition), retention time (min), and peak area (arbitrary units).
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1 - List of tables

2
3 **Table 1.** Optimized MS/MS parameters.
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Parameter	Value
Source temperature (°C)	120
Capillary voltage (kV)	3.0
Desolvation temperature (°C)	350
Desolvation gas flow (L/h)	750
Cone gas flow (L/h)	75
Quantification transition, m/z	88.7 > 60.7
Confirmation transition, m/z	88.7 > 42.8
Cone voltage (V)	25
Collision energy (eV)	10

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6
7 **Table 2.** Precision and accuracy of the method for the determination of NMEA using
8 UHPLC/MS/MS.

Spiked level (µg/L)	Intra-day ^{a)}		Inter-day ^{a)}	
	(RSD, %)	(RE, %)	(RSD, %)	(RE, %)
1	1.25	100.61	2.16	101.93
2	1.38	101.52	2.17	102.25
20	2.23	102.36	2.47	102.77

9 ^{a)} N = 6

10 All abbreviations as given in section 2.
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Table 3. Recoveries (R) and relative standard deviations (RSD) and detection limits of NMEA at different concentrations (N = 6) for the presently developed method compared with literature reported data obtained for the LC-MS/MS methods.

Spiked level (µg/L)	Detected level (µg/L)	R (%)	RSD (%)	LOQ method (ng/L)	LOQ (ng/L) [1]	LOQ (ng/L) [19]
1	0.43	86.00	3.23			
2	1.97	98.62	3.63	0.8	2.5	5
20	17.43	87.15	4.25			

25

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Table 4. Determination of the matrix effect (ME) and relative standard deviations (RSD) of NMEA using SPE/UHPLC/MS/MS method.

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Spiked level (µg/L)	matrix effect ^a			
	river water 1		river water 2	
	ME (%)	Precision (RSD, %)	ME (%)	Precision (RSD, %)
1	85.12	3.22	84.89	3.13
2	97.14	4.12	96.71	4.15
20	89.11	3.67	88.79	3.37

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^aN = 6