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Biases in Ketone Measurements using DNPH-coated Solid Sorbent Cartridges

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42 **Abstract**

43 Biases associated with carbonyl measurement using active air sampling through a 2,4-
44 dinitrophenylhydrazine (DNPH)-coated solid sorbent cartridge following the U.S. EPA Method TO-
45 11A are known but have not been fully investigated. Ketones are less reactive than aldehydes in the
46 derivatization with DNPH, resulting in poor collection efficiency. Field studies and laboratory
47 experiments demonstrate the uncertainties associated with two ketones (i.e., acetone and methyl ethyl
48 ketone [MEK]). Ketone collection efficiencies are inversely related to relative humidity (RH), sample
49 flow rate, and sample duration. Since water is a product in the bidirectional derivatization of
50 carbonyls, the reverse reaction competes with the forward reaction as RH increases. Laboratory
51 experiments demonstrate that ~35–80% of the ketones can be lost for $RH > 50\%$ with a single DNPH
52 cartridge at a temperature of 22 ± 2 °C. Optimal sampling flow rates and sampling durations under
53 high RH need to be determined in various environments to ensure tolerable collection efficiencies.

54 *Keywords:*

55 *Ketones, DNPH, relative humidity, and collection efficiencies*

56

57 1. Introduction

58 Acetone and methyl ethyl ketone (MEK) are the two most abundant atmospheric ketones in
59 carbonyls. Acetone, a precursor of methylglyoxal that forms secondary organic aerosols,¹⁻³ has been
60 used as a solvent for paints, varnishes, lacquers, fats, oils, waxes, resins, printing inks, plastics, and
61 glues.⁴ Acetone levels of $5.1 \mu\text{g m}^{-3}$ were reported in Guiyang⁵ and $17.8 \mu\text{g m}^{-3}$ in Guangzhou,
62 China.⁶⁻⁷ Excessive acetone exposure can cause eye irritation, respiratory distress (e.g., nose, throat,
63 and lung), and ultimately unconsciousness, seizures, coma, and even death. MEK is naturally emitted
64 by volcanoes, forest fires, and biological degradation; and is also a natural component of food.⁸ In
65 2005, the United States Environmental Protection Agency (U.S. EPA) removed MEK from the list of
66 hazardous air pollutants (HAPs), since there was insufficient evidence that manmade MEK caused
67 adverse health or other environmental effects⁹. However, MEK is still being monitored due to its
68 participation in tropospheric ozone (O_3) chemistry. Ambient MEK concentrations range from 0.16–
69 $3.45 \mu\text{g m}^{-3}$ in Guangzhou, China¹⁰⁻¹¹ to $20 \mu\text{g m}^{-3}$ in the industrial city of Gumi, Korea.¹²

70 Real-time analyzers have been developed to measure ambient concentrations of the most
71 abundant carbonyl species such as formaldehyde.^{13,14} Solid phase approaches combine ambient
72 sampling and derivatization into a single step to determine carbonyls, including aldehydes and
73 ketones.¹⁵⁻¹⁷ The most commonly used method for simultaneous determination of multiple carbonyls
74 is U.S. EPA Method TO-11A which calls for sampling onto a 2,4-dinitrophenylhydrazine (DNPH)-
75 coated solid sorbent (i.e., silicon gel) cartridge at a flow rate of 0.7 L min^{-1} , followed by solvent
76 elution and high-pressure liquid chromatographic (HPLC) analysis.¹⁸ The DNPH-coated cartridge
77 method has been globally applied for research and in compliance networks. This method has potential
78 interferences, since oxidants (e.g., nitrogen oxide [NO], nitrogen dioxide [NO_2], and O_3) react with
79 the DNPH, forming side-products that bias carbonyl quantification.¹⁹⁻²¹ Uncertainties in the
80 determination of unsaturated carbonyls such as acrolein and crotonaldehyde have also been

81 reported,^{22,23} owing to double-bond-containing carbonyls that react further with DNPH to form larger
82 molecules.²³

83 Collection efficiencies (CEs) for aldehydes have been reported by Herrington et al.²⁴ and
84 Herrington and Hays.²⁵ Ketones are less reactive than aldehydes in the derivatization with DNPH and
85 they are affected by sampling conditions. This paper examines the CEs of acetone and MEK under
86 different relative humidities (RHs), flow rates, and sample durations, for ambient sampling and
87 laboratory experiments.

88 2. Materials and Methods

89 2.1 Field and Laboratory Sample Collection

90 Twenty-four hour duration (midnight to midnight) samples were collected at three sites
91 representing urban (Tsim Sha Tsui), suburban (Sai Kung West Country Park), and coastal (Tai Tam
92 Bay) areas in Hong Kong from 10th – 23rd January 2010 (winter) and from 13th – 26th July, 2011
93 (summer). Ambient samples were collected on silica gel cartridges impregnated with acidified 2,4-
94 DNPH (Sep-Pak DNPH-silica, 55-105 μm particle size, 125Å pore size; Waters Corporation,
95 Milford, MA, USA) using an automatic carbonyl sampler (Model 8000, ATEC, Malibu, CA, USA) at
96 a flow rate of 0.7 L min⁻¹.¹⁸ The sampling system employs a heated inlet maintained at 50 °C to
97 minimize liquid water interferences with the DNPH-coated cartridge. The selected 24-hr sample
98 duration ensures that the collected carbonyls do not consume >30% of the derivatizing agent coated
99 on the cartridge. Two DNPH cartridges in series were collected at the three sites with an inlet height
100 of 1.2–1.5 m above ground level. Past studies show no appreciable breakthrough at such sampling
101 flow rates and durations.^{18,26,27}

102 Flow rates were verified in the field at the beginning and end of each sampling period using a
103 calibrated flow meter (Gilibrator Calibrator; Gilian Instruments, W. Caldwell, NJ, USA). A Teflon-
104 membrane filter assembly (Whatman, Clifton, NJ, USA) and an O₃ scrubber (Sep-Pak; Waters
105 Corporation, Milford, MA, USA) were installed in front of the DNPH-coated cartridge in order to

106 remove particulate matter and prevent possible O₃ interference, respectively²⁷. The recovery of
107 carbonyls collected in the process should not be affected by the O₃ trap²⁸. Collocated samples were
108 acquired to determine reproducibility, with correlation coefficient (r) >0.98 in the field. One cartridge
109 was designated as a field blank on each sampling trip for the three sites and was handled the same
110 way as the sample cartridges. Fourteen field blanks were collected at each site during each season.
111 The amounts of carbonyls detected in each cartridge were corrected by subtracting averaged field
112 blanks. The DNPH-coated cartridges were stored in a refrigerator at <4 °C after sampling and before
113 chemical analyses. Samples were analyzed within two weeks after sampling to minimize sample
114 degradation during cold storage. Meteorological parameters, including temperature, RH, air pressure,
115 and rainfall, were recorded during each sampling period.

116 Liquid vaporization to a Tedlar® gas sampling bag (Sigma-Aldrich, St. Louis, MO, USA)
117 was used to examine the stability of gaseous carbonyl standards in methanol under a vented fume
118 hood for different RH.^{15,16} High CEs (>93±5%) were achieved, suggesting negligible wall losses and
119 high stabilities, similar to those found in prior experiments^{15,16}. As RH inside the bag was <1%
120 during calibration, carbonyl concentrations varied by <10% after 24 hr storage. Various amounts of
121 water were injected into the bag to simulate different atmospheric RH.

122 2.2 Chemical Analysis

123 Target carbonyls were quantified as shown in Table 1. Since unsaturated carbonyls may react
124 with excess reagent to form adducts, these compounds were not accurately quantified due to co-
125 elution and changing response factors^{22,23}.

126 Each DNPH-coated cartridge was eluted with 2.0 mL of acetone-free acetonitrile (HPLC-
127 grade; Mallinckrodt Laboratory Chemicals, Phillipsburg, NJ, USA) to a volumetric flask. Prior tests
128 demonstrate that neither DNPH nor its derivatives remaining in the cartridge are detectable after the
129 2.0 mL elution²⁹. Certified calibration standards of monocarbonyl DNP-hydrazones (Supelco;
130 Bellefonte, PA, USA) were diluted into concentration ranges of 0.015-3.0 mg mL⁻¹ for instrument

131 calibration. The final volume of each calibration mixture was set at 2.0 mL with 8:2 (volume/volume)
132 of acetonitrile/pyridine (HPLC grade; Sigma-Aldrich, St. Louis, MO, USA). Concentrations of the
133 dicarbonyl DNP-hydrazones in the calibration standards ranged from 0.01-2.0 $\mu\text{g mL}^{-1}$. Instrument
134 response was linear with concentration ($r > 0.999$). The cartridge extracts and calibration standards
135 were analyzed by injecting 20 μL of the extract into an HPLC system (Series 1200; Agilent
136 Technology, Santa Clara, CA, USA) equipped with a photodiode array detector (PAD). The
137 separation column (i.e., 4.6 \times 250 mm Spheri-5 ODS 5 μm C-18 reverse-phase column; PerkinElmer,
138 Norwalk, CT, USA) was operated at room temperature (22 \pm 2 $^{\circ}\text{C}$).

139 The mobile phase consisted of three solvent mixtures: (A) 6:3:1 (v/v) of water, acetonitrile,
140 and tetrahydrofuran, respectively; (B) 4:6 (v/v) of water and acetonitrile, respectively; and (C)
141 acetonitrile. The gradient program was 80% A/20% B for 1 minute, followed by linear gradients to
142 50% A/50% B for 8 minutes, 100% B for 10 minutes, transition from 100% B to 100% C over 6
143 minutes, and 100% C for 5 minutes. The flow rate was 2.0 mL min^{-1} throughout the run. Absorbances
144 at 360 nm and 390 nm were used to identify the aliphatic and aromatic (e.g., benzaldehyde,
145 tolualdehydes, and 2,5-dimethylbenzaldehyde) carbonyls, respectively.

146 Identification and quantification of carbonyl compounds were based on retention times and
147 peak areas compared with the corresponding calibration standards, respectively. The minimum
148 detection limits (MDLs) in Table 1 were obtained by analyzing ≥ 7 replicates for each standard
149 solution containing the analytes at a concentration of 0.015 $\mu\text{g mL}^{-1}$. The MDLs of the target
150 carbonyls ranged from 0.002–0.010 $\mu\text{g mL}^{-1}$, which can be converted to 0.032–0.18 $\mu\text{g m}^{-3}$ based on
151 a sampling volume of 1.01 m^3 . Measurement precision, derived from replicate analyses, ranges from
152 0.5–3.2% for the target carbonyls. The chromatographic peaks were further confirmed as an
153 independent quality assurance/quality control (QA/QC) by a liquid chromatograph (LC)/mass
154 spectrometry (MS) system (Series 6100, Agilent Technology, Santa Clara, CA, USA) for
155 electrospray ionization (ESI) analysis.

156 2.3 *Determination of Collection Efficiency (CE)*

157 Carbonyl CEs were determined by connecting two or three cartridges in series for field and
158 laboratory samples. CEs were calculated as $(1-A_b/A_f) \times 100\%$ where A_b and A_f were the amounts of a
159 carbonyl collected on the back and front cartridges, respectively. Additional uncertainties may be
160 introduced owing to differences between standards generated by vaporization to the Tedlar® bag and
161 liquid injection to the HPLC. Laboratory experiments show a standard recovery of 87–98% with the
162 liquid vaporization method; this was normalized to obtain absolute CEs.

163 3. Results and Discussion

164 3.1 *Ambient Collection Efficiencies*

165 Average concentrations and CEs for field measurements are shown in Table 2. Carbonyl
166 concentrations were ~3 to 10 times higher at the urban than the coastal sites, indicating that
167 anthropogenic sources (e.g., vehicular and cooking emissions) dominated urban carbonyl production.
168 Formaldehyde was the most abundant carbonyl, on average ranging from 1.46 to 8.77 $\mu\text{g m}^{-3}$.
169 Acetaldehyde (0.5–3.37 $\mu\text{g m}^{-3}$) and acetone (0.06–1.09 $\mu\text{g m}^{-3}$) were the next most abundant
170 carbonyls.

171 High CEs (96–99%) were achieved for formaldehyde, consistent with >95% CE specified by
172 U.S. EPA.¹⁸ More variable CEs (83–98%) were found for acetaldehyde, deviating from the >95%
173 CEs specified by the cartridge supplier,²¹ but comparable with the 80% CE reported by Lazarus.³⁰
174 The CEs for other mono- and di-carbonyls were either close to >99% or undetectable (below the
175 MDLs in Table 1), as their concentration levels were one to two orders of magnitude lower than those
176 of formaldehyde, acetaldehyde, and acetone.

177 Lower CEs were found for acetone and MEK, ranging from -240% to 23% and -273% to
178 18%, respectively. More negative CEs were found for samples collected at the more humid coastal
179 site. Negative CEs indicate higher ketone concentrations on the back compared to the front cartridge.
180 Fig. 1 compares front and back chromatograms from the urban Tsim Sha Tsui site, showing acetone

181 and MEK peaks (ID# 3 and 7, respectively) were higher on the back than front cartridges. Similar
182 results are observed for aldehyde compounds. Gaseous carbonyl molecules diffuse onto the solid
183 sorbent surface where DNPH is immobilized (i.e., the addition of the -NH_2 group to the -C=O group)
184 and form a tetrahedral carbinolamine intermediate.³¹ The reaction rate for this reversible step is
185 expected to play a key role in determining the CEs of carbonyls. In the second step, the
186 carbinolamine intermediate loses a molecule of water (H_2O) to form the hydrazone derivative. Water
187 is a product of the reaction and, when the water mixing ratio is high (corresponding to RH at typical
188 ambient temperatures), the backward (i.e., reverse) reaction becomes prominent and competes with
189 the forward reaction for the carbonyls of lower reactivity.

190 For carbonyls, the $\text{>C}^{\delta+}=\text{O}^{\delta-}$ bond is highly polarized because of the differences in
191 electronegativity between carbon and oxygen. However, such polarization is much stronger for
192 aldehydes than ketones, since aldehydes consist of a strong electron withdrawing group of -H
193 attached to the polarized carbon. The nucleophilic -NH_2 tends to attack the aldehydes quickly,
194 resulting in a faster reaction rate than that for ketones. Therefore, less reactive ketones re-entrain
195 from the front cartridge and re-react on the back cartridges. Since $>90\%$ of the aldehydes have been
196 collected by the front cartridge, re-entrained ketones can be retrieved from the back cartridge. Fujita
197 et al.³² reported “unsatisfactory” CEs for acetone, but no values were provided for comparison.

198 3.2 *Relative Humidity Effect*

199 Table 2 shows larger seasonal variations in CEs for ketones than aldehydes. Ketone CEs
200 varied from -67 to -273% during summer under higher RH ($87\text{--}91\%$) and temperatures ($32\text{--}33\text{ }^\circ\text{C}$),
201 with higher CEs (-29 to 23%) during winter under lower RH ($56\text{--}80\%$) and temperatures ($14\text{--}15\text{ }^\circ\text{C}$).
202 Poorer efficiencies during summer could be caused by water vapor interferences, as H_2O mixing
203 ratios increase with temperature for a given RH. Grosjean and Grosjean³³ observed opposite effects
204 with a C_{18} -based cartridge. Aldehydes and ketones had good agreements ($r > 0.8$) between measured
205 and nominal concentrations at $55 \pm 10\%$ RH for temperatures similar to those of Hong Kong, while

206 poor agreement ($r < 0.8$) was found for heavier aldehydes (C_4 - C_9) at 3-7% RH. It is expected that
207 moisture can have dissimilar influences on C_{18} - and silica gel-based cartridges. Pires and Carvalho³⁴
208 also demonstrated that oxidants in air can react differently with DNPH on C_{18} -based cartridges in
209 generating a wide variety of compounds.

210 Laboratory experiments were carried out to examine CEs at five RH levels with temperatures
211 of 22 ± 2 °C. Experimental CEs in the laboratory are expected to be higher than those determined in
212 ambient air because there are fewer competitors (e.g., organic acids or nitrogen oxides) for reactions
213 with DNPH. Fig. 2 illustrates the relative distribution of acetone and MEK collected from a standard
214 gas stream with a series of three cartridges. The CEs for the front cartridge decreased from 90% at
215 $< 10\%$ RH to $\sim 20\%$ at $\sim 100\%$ RH. At low ($< 10\%$) RH, average CEs for acetone and MEK were
216 comparable, at $91 \pm 8\%$ and $89 \pm 9\%$, respectively, and nondetectable (0%) on the third cartridge. CEs
217 decreased to $-47 \pm 15\%$ for acetone and $-79 \pm 18\%$ for MEK at 75% RH (representing typical RH in
218 Hong Kong¹⁵). These results demonstrate that under high RH ($> 50\%$) and at temperature of 22 ± 2 °C,
219 ketone concentrations from single-cartridge sampling could be underestimated by ~ 35 – 80% . At $1 \mu\text{g}$
220 m^{-3} concentrations, a dual-cartridge sampler (i.e., single front-and-back) may not be adequate for
221 ambient RH $> 50\%$, as 2–10% of the ketones were found on the third cartridge as shown in Fig. 2.
222 The sum of ketones collected by the three cartridges in series was close to unity ($> 97\%$) as compared
223 to the laboratory-generated standards. However, it may not be practical or cost-effective to sample
224 three cartridges in series in the field.

225 3.3 *Sampling Flow Rate Effect*

226 The CEs of the two ketones collected at flow rates ranging from 0.1 to 1.0 L min^{-1} were
227 examined. Although the concentrations were correlated ($r > 0.91$), a progressive reduction in CEs for
228 ketones by two- to three-fold was found as flow rate increases (Fig. 3). With longer residence times at
229 lower flow rates, most of the carbonyls are expected to be retained in the cartridge and react with

230 DNPH. However, the two ketones still exhibited negative CEs at the lowest flow rate of 0.1 L min⁻¹,
231 reconfirming the inadequacy of using single-cartridge sampling at 90% RH.

232 3.4 *Sampling Duration Effect*

233 Fig. 4 shows that higher CEs (>55±12%) were obtained for 6 hr as compared to 24 hr sample
234 durations. At 75% RH, water molecules in the sampled air stream can accumulate and retain on the
235 polar-based silica gel, influencing the kinetics of derivatization and leading to decreases in CE. Apel
236 et al.³⁵ reported an average CE of 79% for acetone using C₁₈-based cartridge measurements for six hr
237 duration at an average of ~50% RH; lower CEs (45–65%) were found for nighttime samples.³⁵ The
238 suitability of quantifying daily or diurnal ketone variations using DNPH-coated cartridges under high
239 RH needs to be further investigated.

240 4. **Conclusion**

241 Both field observations and laboratory experiments demonstrate CE changes for carbonyls
242 depending on RH, sample flow rates, sample durations, and concentration levels.³⁶ Commercially
243 available carbonyl samplers with inlets heated to 50 °C can remove water droplets (i.e., rainfall) from
244 the air stream, but elevated RH affects CEs. Table 3 summarizes the interferences and solutions
245 related to the DNPH-coated cartridge method. Even though the potential effects from NO, NO₂, and
246 O₃ can be minimized by upstream denuders/absorbents, the method shows negative biases for the
247 determination of unsaturated aldehydes and ketones. The current DNPH method is adequate for
248 determination of species such as formaldehyde by HPLC. Larger molecular weight carbonyls (e.g.,
249 MEK) can be measured more reliably by gas chromatography with flame ionization detector (GC-
250 FID) than HPLC with PAD.^{37,38}

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317 **Table 1. Minimum detection limit (MDL) for the target carbonyls.**

318

Anonym	Carbonyl	CAS#	MW ^b	Class	MDL ($\mu\text{g m}^{-3}$) ^c
C1	formaldehyde	50-00-0	30	aliphatic	0.045
C2	acetaldehyde	75-07-0	44	aliphatic	0.075
ACE	acetone	67-64-1	58	aliphatic	0.086
ACRO	acrolein	107-02-8	56	aliphatic	0.091
nC3	propionaldehyde	123-38-6	58	aliphatic	0.032
CROT	crotonaldehyde	123-73-9	70	aliphatic	0.087
MEK	methyl ethyl ketone	78-93-3	72	aliphatic	0.092
i-C4	iso-butylaldehyde ^a	78-84-2	72	aliphatic	0.11
nC4	n-butylaldehyde ^a	123-72-8	72	aliphatic	0.11
benz	benzaldehyde	100-52-7	106	aromatic	0.081
iC5	iso-valeraldehyde	590-86-3	86	aliphatic	0.11
nC5	n-valeraldehyde	110-62-3	86	aliphatic	0.14
o-tol	o-tolualdehyde	529-20-4	120	aromatic	0.12
m-tol	m-tolualdehyde	620-23-5	120	aromatic	0.13
p-tol	p-tolualdehyde	104-87-0	120	aromatic	0.13
C6	hexaldehyde	66-25-1	100	aliphatic	0.15
2,5-DB	2,5-dimethylbenzaldehyde	5779-94-2	134	aromatic	0.18

319 ^a iso-Butylaldehyde and n-butylaldehyde were co-eluted in the HPLC analysis.320 ^b Molecular Weight321 ^c The MDL is the minimum detection limit of a carbonyl determined by analyzing ≥ 7 replicates of a standard solution containing an analyte at $0.015 \mu\text{g mL}^{-1}$. MDL is
322 expressed as $\mu\text{g m}^{-3}$, calculated using a sampled air volume of 1.01 m^3 (at a flow rate of 0.7 L min^{-1} for 24 hr).

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325**Table 2. Average concentrations and collection efficiency for 16 carbonyls measured by U.S. EPA method TO-11A.**

Symbol	Carbonyls	Urban (Tsim Sha Tsui)		Suburban (Sai Kung West Country Park)				Coastal (Tai Tam Bay)					
		Winter		Summer		Winter		Summer		Winter		Summer	
		Temp ^a = 14°C (9-16°C)		Temp = 32°C (30-33°C)		Temp = 15°C (10-16°C)		Temp = 33°C (29-34°C)		Temp = 15°C (10-17°C)		Temp = 33°C (28-34°C)	
		RH ^b = 56% (44-65%)		RH = 87% (81-98%)		RH = 67% (48-71%)		RH = 88% (78-97%)		RH = 80% (62-85%)		RH = 91% (85-99%)	
		Conc ^c	CE ^d	Conc	CE	Conc	CE	Conc	CE	Conc	CE	Conc	CE
aliphatic aldehydes													
C1	formaldehyde	7.36	99%	8.77	97%	2.52	97%	2.71	97%	2.02	98%	1.46	96%
C2	acetaldehyde	2.97	97%	3.37	90%	1.28	98%	0.91	87%	1.04	96%	0.50	83%
nC3	propionaldehyde	0.36	>99% ^e	0.61	99%	0.18	>99%	0.19	99%	0.19	98%	0.12	99%
C4n+1	n-butyraldehyde/isobutyraldehyde	0.23	>99%	0.45	>99%	0.10	>99%	0.22	>99%	0.05	>99%	0.23	>99%
i-C5	isovaleraldehyde	0.73	>99%	0.82	>99%	0.31	>99%	0.23	>99%	0.25	>99%	0.06	>99%
n-C5	valeraldehyde	0.13	>99%	0.11	>99%	0.04	>99%	0.06	>99%	0.04	>99%	0.05	>99%
C6	hexaldehyde	0.38	>99%	0.33	>99%	0.06	>99%	0.15	>99%	0.06	>99%	0.07	>99%
aromatic aldehydes													
benz	benzaldehyde	0.35	>99%	0.49	>99%	0.19	>99%	0.16	>99%	0.13	>99%	0.15	>99%
o-tol	o-tolualdehyde	0.08	>99%	0.06	>99%	ND	-	ND	-	ND	-	ND	-
m-tol	m-tolualdehyde	0.18	>99%	0.14	>99%	0.06	>99%	ND	-	0.03	>99%	0.03	>99%
p-tol	p-tolualdehyde	ND ^f	- ^g	ND	-	ND	-	ND	-	ND	-	ND	-
2,5-C6	2,5-dimethylbenzaldehyde	0.07	>99%	ND	-	ND	-	ND	-	ND	-	ND	-
unsaturated aldehydes													
ACRO	acrolein ^h	0.05	>99%	0.67	>99%	0.04	>99%	0.45	>99%	3.49	>99%	1.89	>99%
CROT	crotonaldehyde ^h	0.14	>99%	0.23	>99%	0.12	>99%	0.27	>99%	0.11	>99%	0.29	>99%
ketones													
ACET	acetone	1.09	23%	0.78	-50%	0.63	-12%	0.05	-133%	0.38	-10%	0.06	-240%
MEK	methyl ethyl ketone	0.38	18%	0.31	-67%	0.34	-20%	0.22	-244%	0.18	-29%	0.20	-273%

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^a Average temperature during winter (10th to 23rd January 2010) and summer (13th to 26th July 2011); ^b Average relative humidity; ^c Average concentration in $\mu\text{g m}^{-3}$ derived from front cartridge only, for 14 samples each during winter and summer; ^d Average collection efficiency, defined as $(1 - A_b/A_f) \times 100\%$ where A_b and A_f represent carbonyls collected on backup and front cartridges, respectively; ^e Carbonyl might exist slight amount in the backup cartridge but was below minimum detection limit (MDL), see MDLs in Table 1, expressed CE as >99%; ^f Not detectable, below MDL; ^g No collection efficiency is reported as carbonyl concentration was below MDLs; ^h Unsaturated carbonyl DNP-hydrazones may react with excess reagent to form adducts, which is subject to high uncertainties due to chromatographic and response factor issues^{10,11}.

332 **Table 3. Summary of interferences and remedies on DNPH-coated solid sorbent cartridge**
 333 **method in determination of carbonyls**
 334

Parameters	Influenced species	Interference	Remedy	References
O ₃	all carbonyls	positive and negative artifacts on the carbonyl derivatives	sampling with an upstream ozone scrubber	19,39
NO and NO ₂	formaldehyde and acetaldehyde	NO and NO ₂ react with DNPH, forming side-products which overlap with formaldehyde and acetaldehyde derivatives peaks	better chromatographic separation	20,21
-	unsaturated carbonyls	derivatives undergo polymerization	None	22,23
Relative Humidity (RH)	mostly ketones	Poor collection efficiencies at nominal sampling flow rates, leading to large underestimation of ketone concentrations	Use alternative derivatization agent; quantify ketones with GC-FID ^a	This study

335 ^a GC-FID: Gas chromatography-flame ionization detection
 336

337 **FIGURE CAPTIONS**

338 **Fig. 1.** Chromatograms for a pair of front (upper) and back (lower) cartridges for a sample collected
339 at the urban Tsim Sha Tsui site in Hong Kong on 15th January 2010.

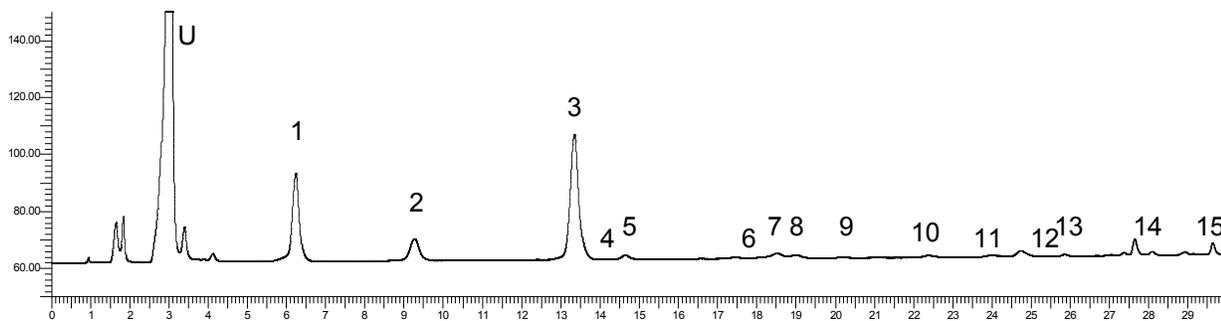
340 **Fig. 2.** Three-stage collection efficiencies (CEs) for acetone and methyl ethyl ketone (MEK) under
341 five different relative humidities (RHs) at room temperature ($22\pm 2^\circ\text{C}$). Distribution of the standard
342 concentrations is expressed in percent of total ketone injected (Gaseous concentrations were $1\ \mu\text{g m}^{-3}$
343 for acetone and MEK. Averages are shown based on a total of 30 sample-sets tested.)

344 **Fig. 3.** Collection efficiencies at laboratory temperatures of $22\pm 2^\circ\text{C}$ for 24-hr ketones at sampling
345 flow rates ranging from 0.1 to $1.0\ \text{L min}^{-1}$ in 90% relative humidity (RH), representing summertime
346 RH in Hong Kong. (Gaseous concentrations were $2\ \mu\text{g m}^{-3}$ for acetone and methyl ethyl ketone
347 (MEK). Averages are shown on a total of 22 dual-cartridge sample-sets tested.)

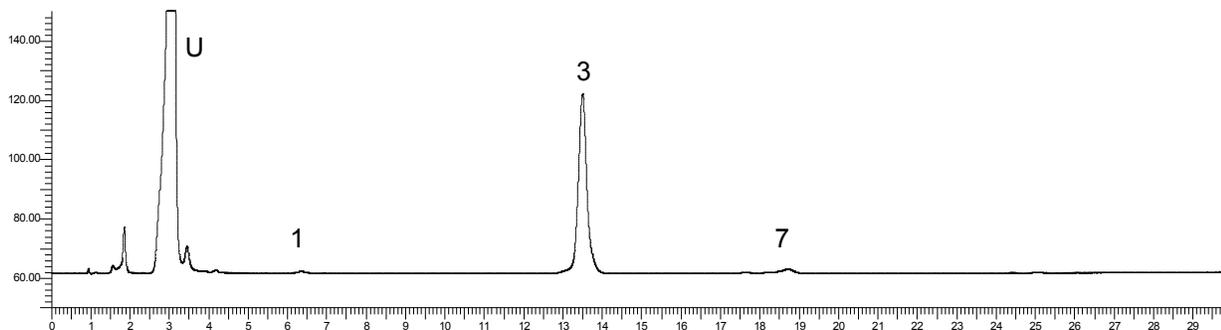
348 **Fig. 4.** Collection efficiencies at laboratory temperatures of $22\pm 2^\circ\text{C}$ for ketones at a flow rate of 0.7
349 L min^{-1} for sample durations ranging from 6 to 24 hr at a relative humidity (RH) of 75%, representing
350 typical RH in Hong Kong. Vertical uncertainty bars represent the measurement precision based on
351 collocated sampling. (Gas concentrations were $2\ \mu\text{g m}^{-3}$ for acetone and MEK. Averages are shown
352 on a total of 20 dual-cartridge sample-sets tested.)

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Peak Identification

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Carbonyl-DNP-hydrazone: 1: formaldehyde; 2: acetaldehyde; 3: acetone; 4: acrolein; 5: propionaldehyde;

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6: crotonaldehyde; 7: methyl ethyl ketone (MEK); 8: *iso*+ *n*-butyraldehyde; 9: benzaldehyde; 10: isovaleraldehyde;

360

11: *n*-valeraldehyde; 12: *o*-tolualdehyde; 13: *m*-tolualdehyde; 14: *n*-hexaldehyde; 15: 2,5-dimethylbezaldehyde

361

U: Unreacted DNPH

362

363

Fig. 1. Chromatograms for a pair of front (upper) and back (lower) cartridges for a sample collected

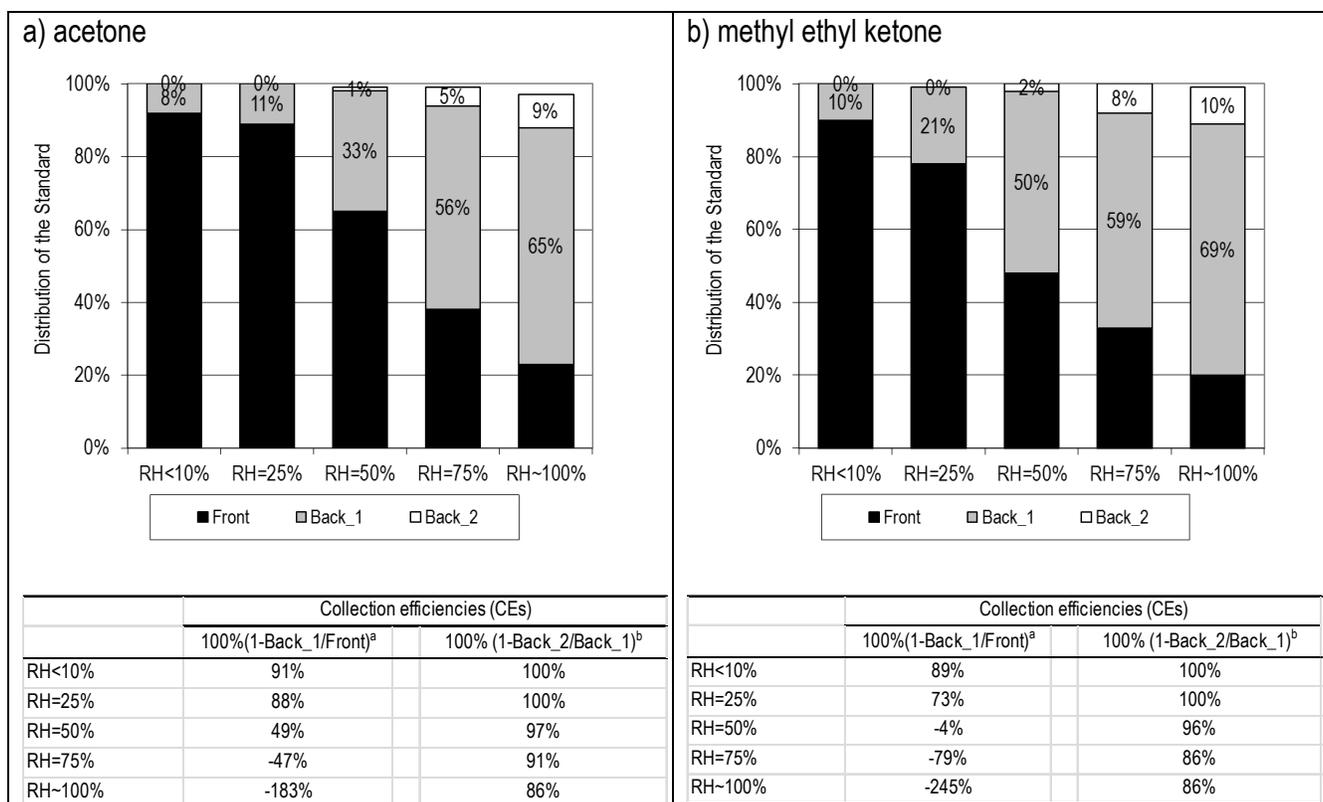
364

at the urban Tsim Sha Tsui site in Hong Kong on 15th January 2010.

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368 ^aThe collection efficiencies are calculated as 100% (1- $A_{\text{Back}_1}/A_{\text{Front}}$), where A_{Back_1} and A_{Front} were the amounts of a
 369 carbonyl collected on the Back_1 (i.e., second) and Front cartridges, respectively.

370
 371 ^bThe collection efficiencies are calculated as 100% (1- $A_{\text{Back}_2}/A_{\text{Back}_1}$), where A_{Back_2} and A_{Back_1} were the amounts of a
 372 carbonyl collected on Back_2 (i.e., third) and Back 1 (i.e., second) cartridges, respectively.

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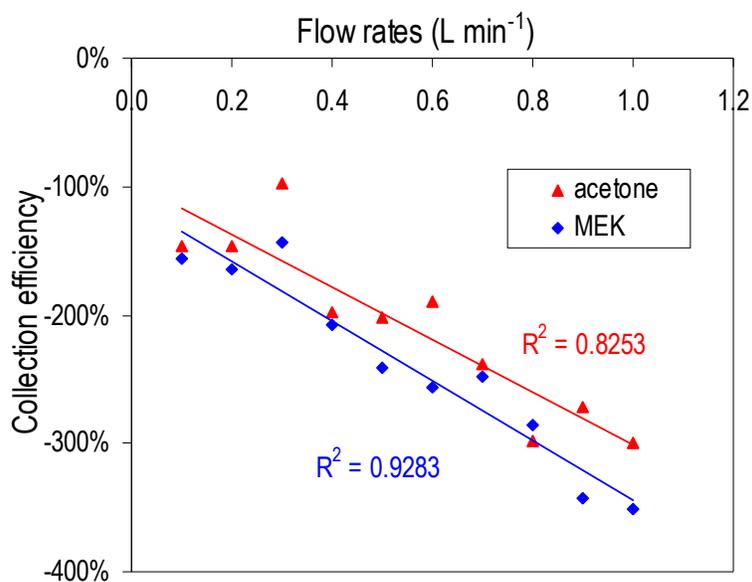
374 **Fig. 2.** Three-stage collection efficiencies (CEs) for acetone and methyl ethyl ketone (MEK) under

375 five different relative humidities (RHs) at room temperature (22±2°C). Distribution of the standard

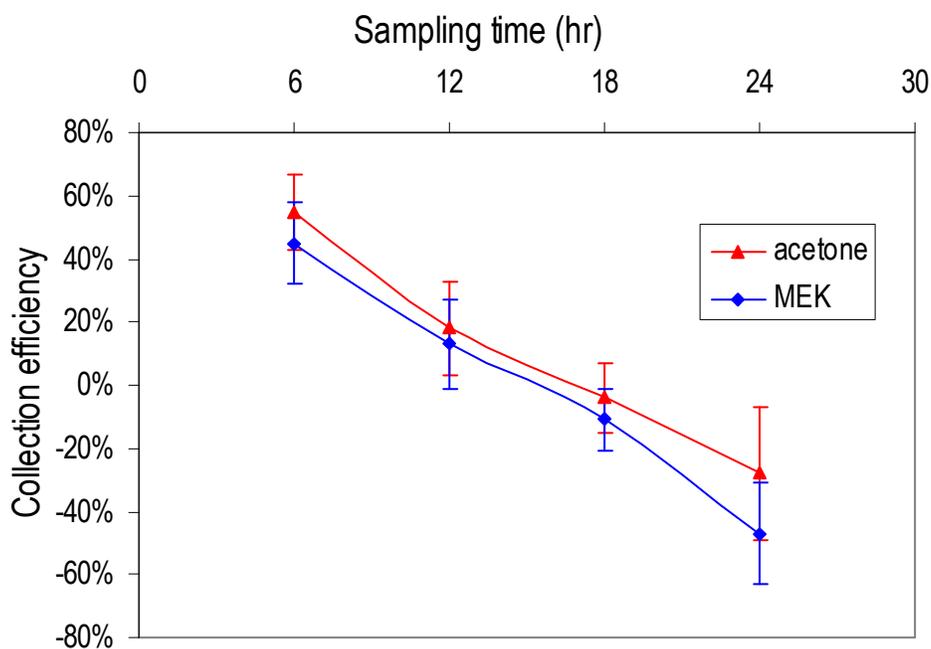
376 concentrations is expressed in percent of total ketone injected (Gaseous concentrations were 1 µg m⁻³

377 for acetone and MEK. Averages are shown based on a total of 30 sample-sets tested.)

378



379
380 **Fig. 3.** Collection efficiencies at laboratory temperatures of 22 ± 2 °C for 24-hr ketones at sampling
381 flow rates ranging from 0.1 to 1.0 L min⁻¹ in 90% relative humidity (RH), representing summertime
382 RH in Hong Kong. (Gaseous concentrations were 2 $\mu\text{g m}^{-3}$ for acetone and methyl ethyl ketone
383 (MEK). Averages are shown on a total of 22 dual-cartridge sample-sets tested.)
384



385
386 **Fig. 4.** Collection efficiencies at laboratory temperatures of 22 ± 2 °C for ketones at a flow rate of 0.7
387 L min^{-1} for sample durations ranging from 6 to 24 hr at a relative humidity (RH) of 75%, representing
388 typical RH in Hong Kong. Vertical uncertainty bars represent the measurement precision based on
389 collocated sampling. (Gas concentrations were $2 \mu\text{g m}^{-3}$ for acetone and MEK. Averages are shown
390 on a total of 20 dual-cartridge sample-sets tested.)