

PAPER

View Article Online
View Journal



Cite this: DOI: 10.1039/d5va00397k

Mechanistic insights into the base-catalyzed hydrolysis of PCDDs via QSAR and DFT approaches

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Polychlorinated dibenzo-*p*-dioxins (PCDDs) are persistent organic pollutants that pose considerable threats to ecological and human health owing to their high toxicity potential. Understanding the mechanisms for underlying the base-catalyzed hydrolysis of PCDDs in aquatic environments is essential for assessing their environmental behaviour and ecological risks. Herein, we combined quantitative structure–activity relationship (QSAR) models with density functional theory calculations to analyse the base-catalyzed hydrolysis mechanisms of PCDDs. Among the four developed QSAR models, the single-parameter QSAR model based on the lowest unoccupied molecular orbital energy (E_{LUMO}) demonstrated the best performance, achieving a coefficient of determination of 0.89 and a root mean square error of 0.49, indicating superior overall performance. Results indicate that the second-order rate constants for base-catalyzed hydrolysis (k_{OH}) of PCDDs are primarily influenced by E_{LUMO} , molecular polarizability (α), molecular volume (V_{m}), degree of chlorination (N_{Cl}), and chlorine position. Specifically, increases in the α and V_{m} values of PCDDs lead to higher $\log k_{\text{OH}}$ values, while an increase in the E_{LUMO} value results in a lower $\log k_{\text{OH}}$ value. This study investigates the relationship between the molecular structure and the rate of base-catalyzed hydrolysis of PCDDs, providing valuable insight into their environmental fate. Furthermore, this research offers a novel theoretical perspective on the base-catalyzed hydrolysis of PCDDs, which will aid in regulatory assessments and risk management.

Received 3rd November 2025
Accepted 26th November 2025

DOI: 10.1039/d5va00397k

rsc.li/esadvances

Environmental significance

Polychlorinated dibenzo-*p*-dioxins (PCDDs) are highly toxic and persistent pollutants widely detected in aquatic environments, yet their degradation behaviour under alkaline conditions remains poorly understood. Understanding their base-catalyzed hydrolysis is crucial for predicting environmental fate and guiding remediation strategies. This study reveals how key molecular descriptors—particularly E_{LUMO} , polarizability, and molecular volume—influence hydrolysis rates. Integration of QSAR modelling and quantum chemical analysis provides a predictive framework for assessing PCDDs degradability. These findings enhance mechanistic understanding of PCDDs reactivity and support environmental risk assessments and regulatory management of persistent organic pollutants.

1 Introduction

Polychlorinated dibenzo-*p*-dioxins (PCDDs) are a class of highly toxic organic compounds, known as persistent organic pollutants.¹ Their structures are bridged by two benzene rings connected by two oxygen atoms, with each benzene ring potentially substituted by 1 to 4 chlorine atoms (Fig. S1). Toxic PCDD congeners are distinguished by the presence of chlorine atoms at the 2, 3, 7, and 8 positions, with 2,3,7,8-T₄CDD being the most toxic.² These compounds primarily exert their toxicity by binding to the aryl hydrocarbon receptor (AhR) in cells, influencing gene expression and disrupting cellular functions, which can lead to various toxic responses.^{3,4} PCDDs are typically byproducts of industrial processes, including incineration processes,⁵ chemical manufacturing,⁶ and metal production.⁷

Owing to their high bioaccumulation potential, lipophilicity, low water solubility, and semi-volatility, PCDDs do not readily degrade in the environment.^{8–10} As a result, these compounds can persist in soil, water, and organisms for extended periods, and have also been detected in dairy products, meat, and other foods consumed by humans,¹¹ posing a potential threat to ecosystems and human health. Therefore, understanding the environmental behaviour of PCDDs is essential for assessing their persistence and potential ecological risks.

Organic compound hydrolysis is a crucial chemical process in the environment.¹² Our recent research has identified base-catalyzed hydrolysis as the primary degradation pathway of PCDDs in aquatic environments. Initial findings suggest that the reactivity of PCDDs hydrolysis is influenced by position and quantity of chlorine atoms on PCDD congeners.¹³ Given that PCDDs have been detected in aquatic environments at concentrations as low as picograms per litre,^{14,15} even trace amounts raise concerns regarding their persistence and potential ecological impacts. However, the relationship between

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the rate of base-catalyzed hydrolysis and the specific molecular structures of PCDDs remains unclear, necessitating further investigation. Due to the time-consuming and labour-intensive nature of experimentally determining this relationship, it is impractical to examine each PCDD congener individually. Therefore, the development of a high-throughput model to evaluate the hydrolysis mechanisms of various PCDDs is essential.

Quantitative structure–activity relationship (QSAR) models have primarily been utilised to investigate the toxicity,¹⁶ AhR binding affinity,¹⁷ bioconcentration and biodegradability of PCDDs.^{18,19} The robust scientific foundation of QSAR technology is predicated on the principle that similar chemical structures are likely to exhibit analogous chemical behaviours.^{20,21} Consequently, QSAR technology can be applied to examine the relationship between the molecular structure and the base-catalyzed hydrolysis rate of PCDDs, as well as to analyse the mechanisms underlying their base-catalyzed hydrolysis. However, to date, no QSAR model for the base-catalyzed hydrolysis of PCDDs has been reported to date.

Herein, based on the second-order rate constants for base-catalyzed hydrolysis (k_{OH}) of 75 PCDDs, calculated using quantum chemical methods, we utilised multiple linear regression (MLR) to develop four QSAR models. These models aim to explore the mechanisms underlying the base-catalyzed hydrolysis of PCDDs. Developed in accordance with the guidelines of the Organization for Economic Co-operation and Development (OECD),²⁰ and these models serve as a valuable tool for assessing the environmental persistence of organic chemicals.

2 Computational details

2.1 Data collection

The k_{OH} values for 75 PCDDs were derived from our previous DFT/TST calculations under aqueous standard conditions (298.15 K and 1 M), and their reliability has been demonstrated.¹³ In accordance with OECD guidelines (OECD, 2007), the dataset was divided into a training set consisting of 60 compounds and a validation set comprising 15 compounds, maintaining a 4 : 1 ratio. For analytical purposes, the data were logarithmically transformed ($\log k_{\text{OH}}$) and standardised to uniform units of $\text{mol}^{-1} \text{L h}^{-1}$. The $\log k_{\text{OH}}$ values vary from a minimum of $-4.28 \text{ mol}^{-1} \text{L h}^{-1}$ (2-M₁CDD) to a maximum of $3.39 \text{ mol}^{-1} \text{L h}^{-1}$ (1,2,3,4,6,7,8,9-O₈CDD).

2.2 Calculation of molecular descriptors

Three molecular descriptors were considered for model development, including constitutional, geometric, and quantum chemical descriptors, which have been widely applied in QSAR studies to describe molecular size, electronic structure, and reactivity.^{8,22,23} Among the constitutional descriptors,²⁴ the total number of chlorine atoms (N_{Cl}), the number of chlorine atoms at the C _{β} position (N_{β}), and the number of pairs of meta-position chlorine atoms (N_{m}) were examined. The selected geometric and quantum chemical descriptors, including

molecular volume (V_{m}), energy of the highest occupied molecular orbital (E_{HOMO}), energy of the lowest unoccupied molecular orbital (E_{LUMO}), molecular polarizability (α) and the most positive partial charge on a hydrogen atom (q_{H^+}) *etc.*, were calculated using the Gaussian 16 software package,²⁵ at the M062X/6–31 + G(d,p)/SMD level. This method was selected based on its proven efficacy in investigating hydrolysis mechanisms and kinetics.^{13,26} For descriptors not available in the output file of Gaussian 16, the wave function file input feature of Multiwfn software was employed for computation and output.²⁷ The meanings of the molecular descriptors involved in models development are provided in Table S1.

2.3 Development and evaluation of models

MLR is a conventional statistical method used in the development of QSAR models.²⁸ In this study, stepwise MLR analysis was employed for variable selection and model development using SPSS software (version 26.0). The goodness-of-fit of the models was evaluated using statistical parameters such as the adjusted determination coefficient (R_{adj}^2) and the root mean square error (RMSE). Internal validation techniques, including leave-one-out cross-validation (Q_{LOO}^2), and external explained variance (Q_{ext}^2) were employed to evaluate the predictive ability of the models. Additional details regarding model validation are provided in the Supplementary Material. The Q_{LOO}^2 value was calculated as follows:

$$Q_{\text{LOO}}^2 = 1 - \frac{\sum_{i=1}^n (y_i - \hat{y}_{(-i)})^2}{\sum_{i=1}^n (y_i - \bar{y}_{(-i)})^2} \quad (1)$$

where y_i is the calculated $\log k_{\text{OH}}$ value for the i -th data point, $\hat{y}_{(-i)}$ is the predicted $\log k_{\text{OH}}$ value for the i -th data point when the model is trained without this point, and $\bar{y}_{(-i)}$ is the mean calculated $\log k_{\text{OH}}$ value of the remaining $n - 1$ data points, excluding the i -th data point.

2.4 Applicability domain of the models

To clarify the scope and limitations of the models, the applicability domain (AD) was characterised using a Williams plot derived from standardised residuals (δ) and leverage values (h).²⁹ Plotting was performed using OriginPro software (version 26.0). PCDDs were considered outliers if $|\delta| > 3$. Additionally, predictions for PCDDs with high leverage values ($h > h^*$, where h^* is the warning leverage) were deemed unreliable. The δ , h , and h^* values were calculated using the following equations:

$$\delta = \frac{y_i - \hat{y}_i}{\sqrt{\sum_{i=1}^n (y_i - \hat{y}_i)^2 / (n - D - 1)}} \quad (2)$$

where y_i and \hat{y}_i represent the calculated and predicted values for the i -th data point, respectively; n is the number of data points and D is the number of descriptors.

$$h_i = x_i^T (X^T X)^{-1} x_i \quad (3)$$



where x_i is the descriptor value of the i -th data point; X is the descriptor matrix of the data points; and x_i^T and X^T are the transposes of x_i and X , respectively.

$$h^* = 3(D + 1)/n \quad (4)$$

where n is the number of data points in the training set, and D has the same meaning as above.

3 Results and discussion

3.1 Development and validation of the QSAR models

First, forward stepwise regression was performed to screen the 15 molecular descriptors of three different types. Then MLR analysis was performed using $\log k_{\text{OH}}$ as the dependent variable and the selected molecular structural parameters as predictor variables, resulting in four QSAR models, as shown in Table 1.

A single-parameter QSAR model (1) with E_{LUMO} was developed to investigate the relationship between k_{OH} and molecular structure of PCDDs. The values of the descriptors used in the QSAR model (1), along with the calculated and predicted k_{OH} values, were listed in Table S2. High R_{adj}^2 and Q_{LOO}^2 values of the training set indicated the goodness-of-fit and robustness of the model (1). The linear regression results of predicted *versus* calculated values for model (1) as well as the Williams plot representing the AD, are shown in Fig. 1. The data points from both the training and validation sets are evenly distributed on

both sides of the reference line $y = x$, indicating that model (1) fit both datasets well. The AD analysis, shown in Fig. 1B, demonstrated the lack of outliers in the validation set, $h < h^*$, and $|\delta| < 3$.

As shown in Table 1, QSAR model (2) was developed to explore the relationship between k_{OH} and α . The relevant data are presented in Table S3. The R_{adj}^2 and Q_{LOO}^2 values of this QSAR model were 0.83 and 0.82, respectively, indicating high goodness-of-fit and good robustness of model (2). The differences (0.01) between the R^2 and Q_{LOO}^2 values was < 0.3 , indicating no over-fitting of model (2).³⁰ Moreover, this model demonstrated acceptable predictability, with $Q_{\text{ext}}^2 = 0.88$ and $\text{RMSE}_{\text{ext}} = 0.55$. The results of the linear fit of the model predictions compared to the calculated values and the Williams plot characterizing AD are presented in Fig. S2. The results of the AD characterization revealed $|\delta| < 3$ and $h < h^*(0.1)$, indicating that all PCDDs were within the AD.

Among geometric descriptors, V_m is generally used to develop QSAR models for the physicochemical properties of PCDDs,^{8,31} suggesting that V_m could provide a convenient first estimator of k_{OH} values for PCDDs. According to Table S4, QSAR model (3) demonstrated strong predictive capabilities within the training and external validation sets. The close $R_{\text{adj}}^2 = 0.82$ and $Q_{\text{LOO}}^2 = 0.81$ values for the training set, along with high $Q_{\text{ext}}^2 = 0.91$ values and lower $\text{RMSE}_{\text{ext}} = 0.48$ values for the validation set, suggested that model (3) was well-fitted, robust,

Table 1 Quantitative structure–activity relationships between $\log k_{\text{OH}}$ and molecular descriptors for PCDDs

No.	Equation	R_{adj}^{2a}	$\text{RMSE}_{\text{tra}}^b$	Q_{LOO}^{2c}	R_{ext}^{2d}	$\text{RMSE}_{\text{ext}}^e$	Q_{ext}^{2f}
1	$\log k_{\text{OH}} = -10.15E_{\text{LUMO}} - 4.01$	0.89	0.63	0.88	0.92	0.49	0.89
2	$\log k_{\text{OH}} = 0.077\alpha - 20.75$	0.83	0.78	0.82	0.89	0.55	0.88
3	$\log k_{\text{OH}} = 0.114V_m - 17.61$	0.82	0.79	0.81	0.92	0.48	0.91
4	$\log k_{\text{OH}} = 1.32 N_{\text{Cl}} - 0.48 N_{\beta} + 0.37 N_{\text{m}} - 4.92$	0.87	0.69	0.86	0.89	0.56	0.77

^a Adjusted determination coefficient. ^b The root mean square error on the training set. ^c Leave-one-out. ^d External determination coefficient. ^e The root mean square error on the validation set. ^f External explained variance.

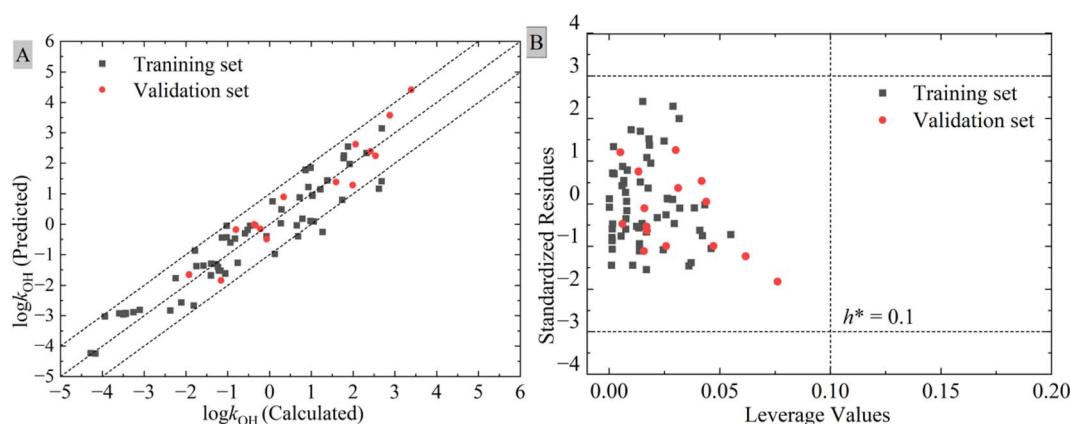


Fig. 1 (A) Plot of calculated *vs.* predicted $\log k_{\text{OH}}$ values of PCDDs for model (1); (B) Williams plot indicating the applicability domain of model (1). The vertical dotted line represents the warning leverage value ($h^* = 0.1$).



Table 2 Significant p , t -test and VIF values of the descriptors involved in model (4)

Descriptor	t^a	VIF	P^b
N_{Cl}	10.722	3.697	<0.001
N_{β}	-3.952	2.158	<0.001
N_m	2.155	2.531	<0.04

^a t represents the statistics obtained by the t -test. ^b P represents the significance level of the t -test.

and had good external predictive ability. Fig. S3A demonstrates the correlation between predicted and calculated values for model (3), indicating the model's good predictive performance. The AD characterization (Fig. S3B) revealed that all PCDDs were within the AD, with no outliers. In the validation sets, there is a 'good high leverage' point (1,2,3,4,6,7,8,9- O_8 CDD) with a higher than the warning h value of 0.1 and $|\delta| < 3$, implying that the model (3) possesses some degree of extrapolating ability.^{29,32}

PCDD congeners with the same number of chlorine atoms can display significantly different chemical and physical properties depending on the positions of the chlorine atoms attached to the parent structure.⁸ Therefore, different degrees of chlorination and chlorine atoms position is essential in understanding these differences. Three molecular descriptors (N_{Cl} , N_{β} , and N_m) were included in the QSAR model (4), which demonstrated good fitting, robustness, and predictive capability. N_{Cl} was the most significant descriptor that negatively contributed to $\log k_{OH}$ ($t = 10.722$, $P < 0.001$). The variance inflation factor (VIF) values for each descriptor were less than 5, indicating the absence of multicollinearity between the descriptors. The p -values for all descriptors less than 0.05, demonstrating statistically significant contributions to the predictive power of model (4). The fitting results between the predicted and calculated values of the model (4) were good, as shown in Table S5 and Fig. S4A. Additionally, Fig. S4B, illustrates the AD characterisation of model (4), revealing $|\delta| < 3$, which indicates no outliers (Table 2).

A comparison of the four models indicated that model (1), which utilised E_{LUMO} as a descriptor, demonstrated the best overall performance with the highest R_{adj}^2 (0.89) and Q_{ext}^2 (0.89) values and lowest RMSE_{ext} (0.49) value. This finding suggests that electronic properties, specifically E_{LUMO} , play a crucial role in the base-catalyzed hydrolysis mechanism of PCDDs. In practical applications, the choice of model should balance complexity, accuracy, and generalizability.^{33,34} Therefore, model (1) serves as a robust and reliable tool for the mechanism analysis of PCDDs, which is essential for understanding their environmental behaviour and potential impact.

3.2 Mechanism interpretation

Electrophilicity is related to the E_{LUMO} of an electrophile, where the LUMO represents the innermost orbital with available positions to accept electrons.^{35,36} A lower E_{LUMO} value implies that a molecule is more inclined to accept electrons,^{37,38}

meaning that PCDDs act as electrophiles, while OH^- ions serve as nucleophiles during the base-catalyzed hydrolysis of PCDDs. PCDDs primarily undergo hydrolysis *via* two pathways: dioxin ring-opening and hydrolytic dichlorination.¹³ In the ring-opening pathway, the electrophilic C-O bond, particularly in PCDDs with lower E_{LUMO} values, is more susceptible to nucleophilic attack by OH^- , resulting in the cleavage of the dioxin ring and the formation of open-ring products (chlorinated hydroxydiphenyl ethers). In the hydrolytic dechlorination pathway, lower E_{LUMO} values enhance the electrophilicity of C-Cl bonds, facilitating the substitution of chlorine atoms with hydroxyl groups to yield hydroxylated PCDDs.

Molecular polarizability is defined as the ratio of the induced dipole moment to the electric field that produces this dipole moment.³⁹ Molecules with high polarizability possess electrons that can move relatively easily compared to those with low polarizability.⁴⁰ Consequently, increased molecular polarizability enhances the likelihood of spatial electron distribution changes, which in turn increases reactivity towards nucleophiles or electrophiles.^{41,42} In base-catalyzed hydrolysis reactions, the electron clouds in PCDDs with higher α can move more easily. This phenomenon may contribute to a reduction in the stability of reactive sites, thereby decreasing the Gibbs free energies (ΔG^\ddagger) of reaction required for bond cleavage and ultimately enhancing the k_{OH} . For instance, 1,2- D_2 CDD exhibits a polarizability of 233.387 a.u. and a ΔG^\ddagger of 104.599 kJ mol⁻¹, while 1,2,6- T_3 CDD has a polarizability of 248.523 a.u. and a ΔG^\ddagger of 89.92 kJ mol⁻¹. Thus, $\log k_{OH}$ is positively correlated with α .

According to the molecular structure of PCDDs presented in Fig. S1, V_m was positively correlated with the number of chlorine atoms in PCDD congeners ($V_m = 10.559N_{Cl} + 109.414$, $R^2 = 0.983$). Therefore, a higher V_m value indicated a greater number of chlorine atoms, as the V_m value of PCDD congeners depends solely on the number of chlorine atoms.⁴³ Due to their strong electronegativity, the electron density at the reaction site will decrease with an increasing number of chlorine atoms, resulting in a faster attack by OH^- ions.^{42,44} Consequently, an increase in the V_m value is associated with an increase in the k_{OH} value for PCDDs.

The QSAR model (4) in Table 1 demonstrated the effects of different degrees of chlorination and chlorine positions on the base-catalyzed hydrolysis rate of PCDDs. For examples, the electrostatic potential (ESP) distributions on the molecular surfaces of 2,7- D_2 CDD, 1,2,3,4- T_4 CDD, 1,2,3,6- T_4 CDD, 1,2,3,7,8,9- H_6 CDD, 1,2,4,6,8,9- H_6 CDD and 1,2,3,4,6,7,8,9- O_8 CDD were calculated to evaluate the relationship between constitutional descriptors (N_{Cl} , N_{β} and N_m) and k_{OH} of PCDDs. The results are presented in Fig. 2, with the blue regions indicating positive ESP (low electronic density), which OH^- ions generally prefer to attack. According to Fig. 2, positive ESP values of PCDDs were primarily distributed at the C_γ position, increasing with the number of chlorine atoms. This finding aligned with the positive correlation between $\log k_{OH}$ and N_{Cl} in model (4), where a greater number of chlorine atoms enhanced the k_{OH} value. For the PCDD congeners with the same N_{Cl} and N_m values, such as 1,2,3,7,8,9- H_6 CDD and 1,2,4,6,8,9- H_6 CDD, the former with a higher N_{β} value exhibited a greater positive



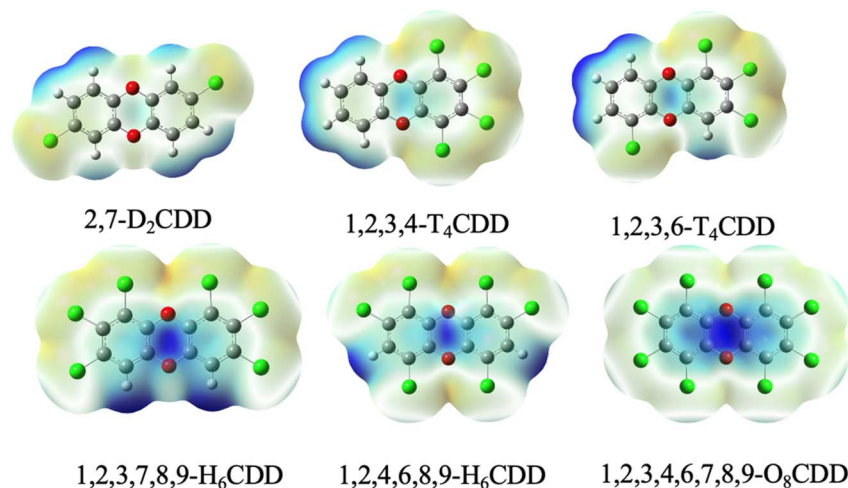


Fig. 2 Electrostatic potential (ESP) distribution of PCDDs calculated at the M062X/6–31+ G(d,p)/SMD level.

ESP than the latter with a lower N_{β} value. For PCDD congeners with the same N_{Cl} and N_{β} values, those with higher N_{m} values had a lower positive ESP, as seen with 1,2,3,4- $T_4\text{CDD}$ and 1,2,3,6- $T_4\text{CDD}$. This finding indicated that variations in the number of C_{β} position and paired meta-position chlorine atoms significantly affected the electron distribution of PCDDs with the same N_{Cl} values, leading to notable changes in their hydrolysis rates. Therefore, the size of the blue region was closely related to the number and position of chlorine atoms in PCDDs. This observation supported the relationship described in the model and explained the effects of the number of chlorine atoms, C_{β} position chlorine atoms and paired meta-position chlorine atoms on the base-catalyzed hydrolysis rate of PCDDs.

4 Model comparison

This study is the first to develop models for evaluating the hydrolysis reaction of PCDDs, addressing a gap in the literature, as no previous models have been reported in this area. To assess the accuracy and interpretability of these models, they were compared with two types of QSAR models: one for predicting the reaction of PCDDs with hydroxyl radicals and another for predicting the base-catalyzed hydrolysis rate of various organic compounds.

Previous researchers have developed QSAR models that investigate how structural and quantum chemical descriptors influence the reactivity of PCDDs with hydroxyl radicals. Yan *et al.* utilised partial least squares (PLS) regression to develop polyparameter lineal free energy models that concentrate on the rate constants for gas-phase reactions of hydroxyl radicals with PCDDs and dibenzofurans (PCDD/Fs).⁴⁵ Their findings indicated that electron-donating capacity, represented by descriptors such as E_{HOMO} and $q\text{H}^+$, was the primary factor affecting reaction rates. Likewise, Qi *et al.* employed MLR and found that E_{HOMO} significantly influenced the reaction rate constant, while N_{Cl} played a critical role, independent of chlorine positions.⁴⁶ Luo *et al.* used MLR and reported that the position of chlorination on PCDDs is a key determinant in the kinetics of hydroxyl radical oxidation kinetics.⁴⁷ Furthermore, Chen *et al.* developed a model using PLS regression with structural descriptors to predict the photolysis rate constants of PCDD/Fs on cherry leaf wax layers.⁴⁸ They discovered that PCDD/Fs with higher N_{Cl} and α values, along with lower E_{LUMO} , exhibited faster photodegradation, which aligns with our hydrolysis model findings and highlights the significance of these descriptors in predicting the reactivity and environmental persistence of PCDDs. Therefore, our models build upon these findings, further emphasising the critical role of key descriptors such as N_{Cl} and α values in the hydrolysis reaction.

Table 3 Comparison of different MLR models for aqueous k_{OH} values

Model $n^a p^b$	Training set			Validation set		
	R_{adj}^2	RMSE _{tra}	Q_{LOO}^2	R_{ext}^2	RMSE _{ext}	Q_{ext}^2
Bernhard <i>et al.</i> (1995)	6	1	0.838	—	—	—
Wang <i>et al.</i> (2018)	40	8	0.822	1.472	—	—
Xu <i>et al.</i> (2019)	23	3	0.865	0.389	0.801	0.925
Xu <i>et al.</i> (2019)	5	1	0.975	0.276	0.914	—
Xu <i>et al.</i> (2021)	24	3	0.842	—	0.729	0.919

^a n represents the number of chemicals in the data set. ^b p represents the total number of predictor variables.



As illustrated in Table 3, we conducted a comparison of the models developed in this study with QSAR models created through MLR for hydrolysis rates of various organic compounds.^{49–52} All models were built using rigorous statistical algorithms and well-defined descriptors, consistent with our findings in Model (1). Berger *et al.* observed that the acid-catalyzed hydrolysis rate constants of sulfonylureas decreased significantly with increasing E_{LUMO} values.⁴⁹ Similarly, Xu *et al.* reported that phthalate esters with higher α values displayed a significant increase in the k_{OH} of one side chain, corroborating with the findings of our model (2).⁵¹ Compared to these models, the current model offers more accessible descriptors, enhanced clarity and a larger dataset of 75 samples, thereby improving reliability and mitigating the risk of overfitting. This provides a simpler and more interpretable alternative to more complex models. Consequently, our model presents a more robust and interpretable framework for assessing the hydrolysis mechanism of PCDDs, utilising key descriptors while ensuring high predictive accuracy and simplicity. This makes it a valuable tool for evaluating the environmental persistence of PCDDs under base-catalyzed conditions.

5 Conclusion

As of 2025, the Chemical Abstracts Service Registry encompasses over 290 million unique chemical substances (Chemical Abstracts Service, 2025).⁵³ This extensive database relies exclusively on experimental measurements for risk assessment, underscoring the necessity for alternative approaches, such as QSAR models, to effectively evaluate the risks associated with both existing and newly identified chemicals. In this study, four QSAR models were developed for 75 PCDDs utilising molecular descriptors with the Gaussian 16 software package. Statistical diagnostics confirmed the quality of these models, while both internal and external validation demonstrated their robustness and predictive capability. The QSAR models derived from quantum chemical (E_{LUMO} and α), geometric (V_{m}), and constitutional (N_{Cl} , N_{Br} , and N_{m}) descriptors exhibited a strong correlation with k_{OH} and effectively portrayed the base-catalyzed hydrolysis of PCDDs. The developed models facilitate the exploration of the mechanisms responsible for the base-catalyzed hydrolysis of PCDDs. Additionally, these models provide molecular descriptors that can be utilised in the development of future machine learning and artificial intelligence models for the hydrolysis of aromatic heterocyclic compounds.

Author contributions

Kun Xie: writing – original draft, investigation, conceptualization. Haiqin Zhang: writing – review & editing, methodology, conceptualization.

Conflicts of interest

The authors have no conflict of interest to declare.

Data availability

The data supporting this article have been included as part of the supplementary information (SI). Supplementary information: details about QSAR model development, values of the descriptors used in the QSAR model, Williams plots, Fig. S1–S4 and Tables S1–S5. See DOI: <https://doi.org/10.1039/d5va00397k>.

Acknowledgements

We acknowledge funding from the Key Lab of Eco-restoration of Regional Contaminated Environment (Shenyang University), Ministry of Education (KF-23-03).

References

- 1 R. Lei, Z. Xu, Y. Xing, W. Liu, X. Wu, T. Jia, S. Sun and Y. He, *J. Hazard. Mater.*, 2021, **418**, 126265.
- 2 C. L. Fletcher and W. A. McKay, *Chemosphere*, 1993, **26**, 1041–1069.
- 3 P. S. Kulkarni, J. G. Crespo and C. A. M. Afonso, *Environ. Int.*, 2008, **34**, 139–153.
- 4 E. Q. S. Dossier, Polychlorinated dibenzo-p-dioxins (PCDDs), polychlorinated dibenzofurans (PCDFs), and dioxin-like polychlorinated biphenyls (DL-PCBs), Sub-group on Review of the Priority Substances List, Working Group E, *Common Implementation Strategy for the Water Framework Directive*, 2011, p. 35.
- 5 C.-M. Chen, *Chemosphere*, 2004, **54**, 1413–1420.
- 6 S. Sidhu and P. Edwards, *Int. J. Chem. Kinet.*, 2002, **34**, 531–541.
- 7 U. Quaß, M. Fermann and G. Bröker, *Chemosphere*, 2004, **54**, 1319–1327.
- 8 M. Kim, L. Y. Li and J. R. Grace, *Environ. Pollut.*, 2016, **213**, 99–111.
- 9 G. McKay, *Chem.-Eng. J.*, 2002, **86**, 343–368.
- 10 K. J. Friesen and G. R. B. Webster, *Environ. Sci. Technol.*, 1990, **24**, 97–101.
- 11 M. Lippmann and G. D. Leikauf, *Environmental Toxicants: Human Exposures and Their Health Effects*, Wiley & Sons, Limited, John, 2020.
- 12 W. Mabey and T. Mill, *J. Phys. Chem. Ref. Data*, 1978, **7**, 383–415.
- 13 H. Zhang, K. Xie, Q. Luo, J. Tang and Y. Zhang, *Environ. Sci. Technol.*, 2024, **58**, 5483–5490.
- 14 Y. Liu, P. Peng, X. Li, S. Zhang and M. Ren, *J. Hazard. Mater.*, 2008, **152**, 40–47.
- 15 Q. Zhang, L. Gao, M. Zheng, L. Liu and C. Li, *Bull. Environ. Contam. Toxicol.*, 2014, **92**, 585–589.
- 16 P. Kumar, A. Kumar and D. Singh, *Environ. Toxicol. Pharmacol.*, 2022, 103893.
- 17 F. Li, X. Li, X. Liu, L. Zhang, L. You, J. Zhao and H. Wu, *Environ. Toxicol. Pharmacol.*, 2011, **32**, 478–485.
- 18 P. de Voogt, D. C. G. Muir, G. R. B. Webster and H. Govers, *Chemosphere*, 1990, **21**, 1385–1396.



- 19 Q. Li, H. Yang, N. Hao, M. Du, Y. Zhao, Y. Li and X. Li, *J. Environ. Manag.*, 2023, **345**, 118898.
- 20 OECD, *Guidance Document on the Validation of (Quantitative) Structure-Activity Relationship [(Q)SAR] Models*, Organization for Economic Cooperation and Development, 2014.
- 21 Y. Gao, J. Zhang, S. Cui, Y. Wu, M. Huang and S. Zhuang, *Elsevier eBooks*, 2023, 89–99.
- 22 R. Todeschini and V. Consonni, *Methods and Principles in Medicinal Chemistry*, 2000.
- 23 K. Roy, S. Kar and R. N. Das, *Understanding the Basics of QSAR for Applications in Pharmaceutical Sciences and Risk Assessment*, 2015, 47–80.
- 24 A. R. Katritzky, V. S. Lobanov and M. Karelson, *Chem. Soc. Rev.*, 1995, **24**, 279.
- 25 M. J. Frisch, G. W. Trucks and H. B. Schlegel *et al.*, *Gaussian 16, Revision C.01*, Gaussian, Inc., Wallingford CT, 2016.
- 26 F. P. Pineda, J. Ortega-Castro, J. Raúl Álvarez-Idaboy, J. Frau, B. M. Cabrera, J. C. Ramírez, J. Donoso and F. Muñoz, *J. Phys. Chem. A*, 2011, **115**, 2359–2366.
- 27 T. Lu and F. Chen, *J. Comput. Chem.*, 2011, **33**, 580–592.
- 28 T. I. Netzeva, A. P. Worth, T. Aldenberg, R. Benigni, M. T. D. Cronin, P. Gramatica, J. S. Jaworska, S. Kahn, G. Klopman, C. A. Marchant, G. Myatt, N. Nikolova-Jeliazkova, G. Y. Patlewicz, R. Perkins, D. W. Roberts, T. W. Schultz, D. T. Stanton, J. J. M. van de Sandt, W. Tong and G. Veith, *Altern. Lab. Anim.*, 2005, **33**, 155–173.
- 29 P. Gramatica, E. Giani and E. Papa, *J. Mol. Graph. Model.*, 2007, **25**, 755–766.
- 30 A. Golbraikh and A. Tropsha, *J. Mol. Graph. Model.*, 2002, **20**, 269–276.
- 31 W. Y. Shiu, W. Doucette, F. A. P. C. Gobas, A. Andren and D. Mackay, *Environ. Sci. Technol.*, 1988, **22**, 651–658.
- 32 T. N. G. Borhani, M. Saniedanesh, M. Bagheri and J. S. Lim, *Water Res.*, 2016, **98**, 344–353.
- 33 M. Taghi Goodarzi, B. Dejaegher and Y. Vander Heyden, *J. AOAC Int.*, 2012, **95**, 636–651.
- 34 P. Gramatica, *Int. J. Quant. Struct.-Prop. Relat.*, 2020, **5**, 61–97.
- 35 L.-G. Zhuo, W. Liao and Z.-X. Yu, *Asian J. Org. Chem.*, 2012, **1**, 336–345.
- 36 A. Bendjeddou, T. Abbaz, A. Gouasmia and D. Villemain, *Int. Res. J. Pure Appl. Chem.*, 2016, **12**, 1–9.
- 37 K. F. Khaled, *Appl. Surf. Sci.*, 2008, **255**, 1811–1818.
- 38 J. Attarki, M. Khnifra, W. Boumya, H. Hajjaoui, A. Mahsoun, M. Sadiq, M. Achak, N. Barka and M. Abdennouri, *Surfaces*, 2023, **6**, 450–465.
- 39 Y. Wang, J. Chen, W. Tang, D. Xia, Y. Liang and X. Li, *Chemosphere*, 2019, **214**, 79–84.
- 40 J. Chen, X. Quan, K.-W. Schramm, A. Kettrup and F. Yang, *Chemosphere*, 2001, **45**, 151–159.
- 41 J. E. Edwards and R. G. Pearson, *J. Am. Chem. Soc.*, 1962, **84**, 16–24.
- 42 Z. Yang, S. Luo, S.-H. Ho, T. Ye, Z. Wei, D. Chen and M. Jin, *Environ. Pollut.*, 2016, **211**, 157–164.
- 43 K. R. Cox and W. G. Chapman, *J. Am. Chem. Soc.*, 2001, **123**, 6745.
- 44 M. B. Smith, *March's Advanced Organic Chemistry : Reactions, Mechanisms, and Structure*, WILEY, 2019.
- 45 C. Yan, J. Chen, L. Huang, G. Ding and X. Huang, *Chemosphere*, 2005, **61**, 1523–1528.
- 46 C. Qi, C. Zhang and X. Sun, *Int. J. Mol. Sci.*, 2015, **16**, 18812–18824.
- 47 S. Luo, S.-H. Ho, Z. Wei, Z. Yang, L. Chai and M. Jin, *Chemosphere*, 2017, **172**, 333–340.
- 48 J. Chen, X. Quan, F. Yang and W. J. G. M. Peijnenburg, *Sci. Total Environ.*, 2001, **269**, 163–170.
- 49 B. M. Berger and N. Lee Wolfe, *Environ. Toxicol. Chem.*, 1996, **15**, 1500–1507.
- 50 L. Wang, B. Chen and T. Zhang, *Chem. Eng. J.*, 2018, **342**, 372–385.
- 51 T. Xu, J. Chen, Z. Wang, W. Tang, D. Xia, Z. Fu and H. Xie, *Environ. Sci. Technol.*, 2019, **53**, 5828–5837.
- 52 T. Xu, J. Chen, X. Chen, H. Xie, Z. Wang, D. Xia, W. Tang and H. Xie, *Environ. Sci. Technol.*, 2021, **55**, 6022–6031.
- 53 Chemical Abstracts Service, 2025. <https://www.cas.org/cas-data>.

