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Enhanced visible-light photocatalytic degradation of methylene blue *via* a ternary $\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$ Z-scheme heterojunction with synergistic Fenton-like reactivity for sustainable water remediation

Huihui Fang,^a Jianhua Wang[✉]*^{ab} and Shuhui Shi^c

The rational design of heterostructured photocatalysts is critical for addressing organic dye pollution through solar-driven advanced oxidation processes. Herein, a ternary $\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$ nanocomposite was synthesized *via* a hydrothermal-precipitation route, integrating Z-scheme charge transfer and Fenton-like reactivity for enhanced visible-light photocatalysis. The composite's activity was evaluated *via* methylene blue (MB) degradation under a 300 W metal-halide lamp ($\lambda \geq 420$ nm), achieving 99.93% removal within 35 min—2.1× and 1.4× faster than pristine $\text{g-C}_3\text{N}_4$ and binary counterparts, respectively. This performance stems from synergistic effects: (i) a Z-scheme heterojunction between Ag_2O , CuFe_2O_4 , and $\text{g-C}_3\text{N}_4$, which spatially separates electron–hole pairs *via* tailored band alignment; (ii) Fenton-like $\text{Fe}^{3+}/\text{Cu}^{2+}$ redox cycles amplifying hydroxyl radical ($\cdot\text{OH}$) generation; and (iii) uniform dispersion of Ag_2O and CuFe_2O_4 on $\text{g-C}_3\text{N}_4$ nanosheets, maximizing active sites. Radical trapping confirmed $\cdot\text{OH}$ as the dominant species, with supplementary roles of h^+ and $\cdot\text{O}_2^-$. The composite retained >93% activity over ten cycles, demonstrating robust stability due to interfacial coupling and inhibited agglomeration. This work advances the development of multifunctional heterostructures for sustainable water remediation, bridging photocatalysis and Fenton chemistry.

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Introduction

The rapid global industrial expansion, coupled with the escalating use of antibiotics, pesticides, and organic dyes, has intensified the contamination of freshwater resources, exacerbating the energy and water crisis.¹ Organic pollutants such as synthetic dyes, pharmaceutical metabolites, and pesticide residues—many of which contain carcinogenic heterocyclic aromatic or azo groups—pose severe threats to ecosystems and human health due to their persistence, toxicity, and resistance to conventional physical or chemical degradation methods.² Addressing these challenges necessitates advanced oxidation processes (AOPs) capable of efficiently degrading recalcitrant organic compounds. Among these, visible-light-driven photocatalysis has emerged as a sustainable solution, leveraging solar energy to activate catalytic materials for pollutant degradation and hydrogen production.³

While titanium dioxide (TiO_2) has been widely studied for photocatalysis, its large bandgap (~ 3.2 eV) restricts light absorption to the ultraviolet range ($\sim 5\%$ of solar energy), limiting practical applications.⁴ To overcome this, narrow-bandgap semiconductors such as ferrites and graphitic carbon nitride ($\text{g-C}_3\text{N}_4$) have gained attention. Spinel ferrites like CuFe_2O_4 exhibit structural stability, visible-light absorption, and magnetic recoverability, enabling their dual role as photomagnetic Fenton catalysts.⁵ However, single-phase CuFe_2O_4 suffers from rapid charge recombination, low quantum yield, and limited surface area, hindering its photocatalytic efficiency.⁶ Similarly, $\text{g-C}_3\text{N}_4$ —a metal-free polymer with a moderate bandgap (~ 2.7 eV)—shows promise for visible-light-driven water splitting and pollutant degradation but is plagued by fast electron–hole (e^- – h^+) recombination.⁷

To address these limitations, heterostructured composites combining multiple semiconductors have been proposed. For instance, coupling $\text{g-C}_3\text{N}_4$ with CuFe_2O_4 enhances charge separation *via* Z-scheme mechanisms while enabling magnetic recovery.⁸ Further improvements can be achieved by integrating silver oxide (Ag_2O), a p-type semiconductor with a narrow bandgap (~ 1.2 eV), which extends light absorption into the visible spectrum and promotes reactive oxygen species (ROS) generation (*e.g.*, $\cdot\text{OH}$) through its redox-active surface.⁹ The

^aShenyang University, College of Science, Shenyang 110044, China. E-mail: jhwang1228@163.com

^bLiaoning Provincial Key Laboratory of Micro-nano Materials Research and Development, Shenyang 110044, China

^cShenyang Institute of Engineering, College of Science, Shenyang 110136, China


formation of p–n heterojunctions between Ag_2O , CuFe_2O_4 , and $\text{g-C}_3\text{N}_4$ can synergistically suppress $\text{e}^- - \text{h}^+$ recombination, enhance charge transfer, and increase active sites for catalytic reactions.¹⁰

Herein, we design a ternary $\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$ heterostructure to synergize the advantages of its components: (i) the visible-light responsiveness and porosity of $\text{g-C}_3\text{N}_4$, (ii) the magnetic Fenton activity of CuFe_2O_4 , and (iii) the narrow-bandgap sensitization of Ag_2O . The composite's performance is evaluated *via* methylene blue (MB) degradation under visible light, demonstrating enhanced efficiency through interfacial charge transfer and ROS generation. Mechanistic studies reveal a dual Z-scheme pathway facilitated by band alignment, while the magnetic properties of CuFe_2O_4 ensure facile catalyst recovery. This work provides a scalable strategy for designing multifunctional photocatalysts to address water pollution, with potential applications in energy conversion and environmental remediation.

Experimental

Synthesis of $\text{g-C}_3\text{N}_4$

Graphitic carbon nitride ($\text{g-C}_3\text{N}_4$) was synthesized *via* thermal polycondensation using urea ($\text{CH}_4\text{N}_2\text{O}$, $\geq 99.0\%$, Sigma-Aldrich) and melamine ($\text{C}_3\text{H}_6\text{N}_6$, $\geq 99.0\%$, Alfa Aesar) as precursors. Both reagents were of analytical grade and used without further purification.^{11–13} Urea and melamine were mixed in a 2 : 1 molar ratio and homogenized by grinding in an agate mortar for 30 min to ensure intimate contact between precursors. The homogeneous mixture was transferred to a covered alumina crucible and subjected to pyrolysis in a muffle furnace under static air. The temperature was ramped to 823 K at a heating rate of 2 K min^{-1} , maintained isothermally for 4 h to facilitate complete polycondensation, and then cooled naturally to ambient temperature. The resulting light-yellow bulk $\text{g-C}_3\text{N}_4$ was manually milled into a fine powder using an agate mortar, yielding a porous, layered structure.

Synthesis of $\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$

The ternary $\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$ composites were synthesized *via* a sequential hydrothermal and precipitation protocol, adapted from literature with modifications.^{11–13}

Preparation of $\text{g-C}_3\text{N}_4$ dispersion (Solution A). Pre-synthesized $\text{g-C}_3\text{N}_4$ (0.05 g) was dispersed in 60 mL ethylene glycol ($\text{C}_2\text{H}_6\text{O}_2$, $\geq 99.8\%$, Sigma-Aldrich) *via* ultrasonication (40 kHz, 30 min) to achieve a homogeneous suspension.

Precursor mixing (Solution B). In a light-shielded environment (< 5 lux) stoichiometric amounts of $\text{FeCl}_3 \cdot 6\text{H}_2\text{O}$, and $\text{Cu}(\text{CH}_3\text{COO})_2 \cdot \text{H}_2\text{O}$ were dissolved in 25 mL deionized water. The molar ratio of Fe : Cu was fixed at 2 : 1. To obtain the ternary $\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$ composites, AgNO_3 was loaded varied at 0–4 wt% relative to $\text{g-C}_3\text{N}_4$, $\text{FeCl}_3 \cdot 6\text{H}_2\text{O}$, and $\text{Cu}(\text{CH}_3\text{COO})_2 \cdot \text{H}_2\text{O}$ (see Table 1).

Hydrothermal assembly. Solution B was gradually added to Solution A under magnetic stirring (500 rpm, 30 min), followed by the addition of 1 g urea ($\text{CH}_4\text{N}_2\text{O}$, $\geq 99.0\%$) and sodium citrate ($\text{Na}_3\text{C}_6\text{H}_5\text{O}_7$, $\geq 99.0\%$) in a 1 : 1 mass ratio. The mixture was stirred vigorously to ensure precursor homogeneity, then transferred into a 100 mL Teflon-lined stainless-steel autoclave. Hydrothermal treatment was conducted at 453 K for 12 h under autogenous pressure.

Post-synthesis processing. The cooled product was centrifuged (5000 rpm, 10 min) and washed sequentially with deionized water and absolute ethanol ($\text{C}_2\text{H}_5\text{OH}$, $\geq 99.9\%$) to remove unreacted ions. The resulting $\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$ powder was dried in a vacuum oven (333 K, 12 h) to minimize aggregation.

Characterization

The structural and optical properties of the $\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$ composites were systematically characterized. X-ray diffraction (XRD) patterns were acquired using a Rigaku Ultima IV diffractometer ($\text{Cu-K}\alpha$ radiation, $\lambda = 1.5418 \text{ \AA}$, 40 kV, 40 mA) over a 2θ range of $10\text{--}80^\circ$ to identify crystalline phases. Chemical states and surface composition were analyzed *via* X-ray photoelectron spectroscopy (XPS, Thermo Scientific K-Alpha) with $\text{Al-K}\alpha$ excitation (1486.6 eV), and binding energies were calibrated against the C 1s peak (284.8 eV). In this study, a Thermo Scientific K-Alpha X-ray photoelectron spectrometer was used for testing. Strict adherence to the instrument's standard calibration specifications was maintained before and during the test to ensure data accuracy. The $\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$ composite in this study has a semiconductor-polymer heterostructure, and the sample surface was prone to positive charge accumulation due to photoelectron emission, leading to

Table 1 The feeding ratio of $\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$ composites

	$\text{g-C}_3\text{N}_4/\text{g}$	FeCl_3/g	$\text{Cu}(\text{AC})_2/\text{g}$	AgNO_3/g	Sodium citrate and urea/g	Water/ml	Ethylene glycol/ml
$\text{Ag}_2\text{O}@ \text{g-C}_3\text{N}_4$	0.050	0.00	0.00	0.005	1.00	25	60
$\text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$	0.050	0.012	0.05	0.00	1.00	25	60
$\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4\text{-0}$	0.050	0.00	0.00	0.00	1.00	25	60
$\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4\text{-1}$	0.050	0.028	0.014	0.0008	1.00	25	60
$\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4\text{-2}$	0.050	0.028	0.014	0.0016	1.00	25	60
$\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4\text{-3}$	0.050	0.028	0.014	0.0021	1.00	25	60
$\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4\text{-4}$	0.050	0.028	0.014	0.0029	1.00	25	60



a shift in the binding energy peak position. To eliminate the charge effect, we adopted a composite charge compensation strategy of “low-energy electron neutralization gun + sample grounding”, with the specific operations. Morphological features, including particle dispersion and heterojunction interfaces, were examined by field-emission scanning electron microscopy (FE-SEM, Hitachi SU8010) and high-resolution transmission electron microscopy (HR-TEM, JEOL JEM-2100F) operated at 200 kV. Optical absorption spectra were recorded using a PerkinElmer LAMBDA750 UV-Vis-NIR spectrophotometer equipped with a diffuse reflectance accessory (DRS, 200–800 nm), and bandgap energies were derived from Tauc plots using the Kubelka–Munk function. Photoluminescence (PL) spectra were measured at room temperature with a Horiba Fluorolog-3 spectrofluorometer (excitation $\lambda = 380$ nm) to assess charge carrier recombination dynamics.

Photocatalytic performance was evaluated by monitoring methylene blue (MB) degradation under visible light (300 W Xe lamp, $\lambda \geq 420$ nm), with absorbance at 664 nm tracked using a Shimadzu UV-2600 spectrophotometer. Reaction kinetics were modeled using the pseudo-first-order rate equation $-\ln(C/C_0) = kt$, where k is the rate constant. For degradation cycle, the catalyst was separated from the MB solution using high speed centrifugation method—this method avoids catalyst loss compared to magnetic separation and ensures high recovery efficiency (>98%). The recovered catalyst was sequentially washed with deionized water and absolute ethanol to remove residual MB molecules or intermediate products adsorbed on the surface, then dried in a vacuum oven at 333 K for 2 h to restore its dry state. For each subsequent cycle, a fresh MB solution (100 mL, 30 mg L⁻¹) was prepared, and the regenerated catalyst (0.05 g) was re-dispersed into the solution. The adsorption–desorption equilibrium step (dark stirring for 60 min) and irradiation conditions (lamp power, distance from reaction vessel) were strictly identical to the first cycle.

Photocatalytic activity

The photocatalytic performance of the Ag₂O@CuFe₂O₄@g-C₃N₄ composites was evaluated by degrading methylene blue (MB) under visible-light irradiation. A 300 W metal-halide lamp (Philips, $\lambda \geq 420$ nm with a UV cutoff filter) served as the light source, positioned 20 cm above the reaction vessel. Prior to illumination, 0.05 g of catalyst was dispersed in 100 mL of MB solution (30 mg L⁻¹) and magnetically stirred in the dark for 60 min to establish adsorption–desorption equilibrium. During irradiation, 3 mL aliquots were extracted at 10 min intervals, centrifuged (8000 rpm, 5 min) to remove catalyst particles, and analyzed using a UV-Vis spectrophotometer (Shimadzu UV-2600) by monitoring the absorbance at 664 nm. The degradation efficiency (η) and pseudo-first-order rate constant (k) were calculated as:

$$\eta (\%) = (1 - C_t/C_0) \times 100, \ln(C_t/C_0) = -kt \quad (1)$$

where C_0 and C_t represent the initial and time-dependent MB concentrations, respectively.

To identify reactive species, radical trapping experiments were conducted by introducing 1 mM scavengers: isopropanol (IPA, $\cdot\text{OH}$ quencher), ethylenediaminetetraacetic acid (EDTA, h^+ quencher), and benzoquinone (BQ, $\cdot\text{O}_2^-$ quencher). Control experiments without catalysts or light were performed to confirm the absence of non-photocatalytic degradation.

Results and discussion

X-ray diffraction analysis

The crystallographic phases of the synthesized materials were characterized by X-ray diffraction (XRD). As shown in Fig. 1a, pristine g-C₃N₄ exhibits a distinct peak at 27.8°, corresponding to the (002) interlayer stacking plane of graphitic carbon nitride (JCPDS 87-1526).¹⁴ For the Ag₂O@g-C₃N₄ composite (Fig. 1b), diffraction peaks at 38.0°, 44.2°, 64.5°, and 77.3° are indexed to the (200), (220), (311), and (222) planes of cubic-phase Ag₂O (JCPDS 41-1104), confirming its successful integration.¹⁵ Similarly, CuFe₂O₄@g-C₃N₄ (Fig. 1c) displays peaks at 24.18°, 33.08°, 35.9°, 40.9°, 43.64°, 49.68°, 54.22°, 57.68°, 62.46°, 64.22°, 71.98°, and 75.64°, which align with the (012), (104), (110), (113), (202), (024), (116), (122), (214), (300), (1010), and (220) planes of tetragonal spinel CuFe₂O₄ (JCPDS 34-0425).¹⁶

For the ternary Ag₂O@CuFe₂O₄@g-C₃N₄ composite (Fig. 1d, blue curve), the characteristic peaks of Ag₂O (38.0° for (200), 44.2° for (220), JCPDS 41-1104) and CuFe₂O₄ (35.9° for (110), 40.9° for (113), JCPDS 34-0425) are clearly observed alongside the (002) peak of g-C₃N₄ (27.8°). The absence of impurity peaks (e.g., α -Fe₂O₃ at 33.2°, CuO at 35.5°) indicates that the hydrothermal-precipitation process successfully synthesizes pure Ag₂O and CuFe₂O₄, which are uniformly loaded onto the g-C₃N₄ matrix without secondary phase formation. This further confirms the successful construction of the ternary heterostructure, as the intrinsic crystalline phases of all components are retained.

To evaluate the effect of Ag₂O/CuFe₂O₄ loading on the crystalline structure of g-C₃N₄, the full width at half maximum (FWHM) of the (002) peak was compared across all samples. The FWHM of the (002) peak in pure g-C₃N₄ is 2.19°; in Ag₂O@g-C₃N₄ and CuFe₂O₄@g-C₃N₄, it slightly decreases to 1.14° and 1.36°, respectively. For the ternary composite, the FWHM of the (002) peak is 0.76°, which is marginally narrowed than that of

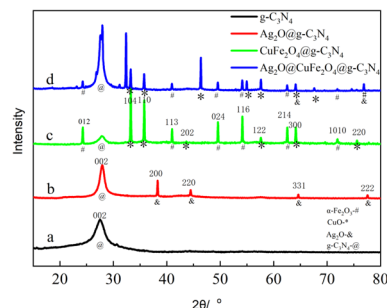


Fig. 1 XRD patterns of (a) g-C₃N₄, (b) Ag₂O@g-C₃N₄, (c) CuFe₂O₄@g-C₃N₄ and (d) Ag₂O@CuFe₂O₄@g-C₃N₄.



pure g-C₃N₄ but without significant broadening. Crystallite sizes were calculated using the Scherrer equation:

$$D = \beta \cos \theta / k\lambda \quad (2)$$

where D is the crystallite diameter, $k = 0.89$ (shape factor), $\lambda = 1.5418 \text{ \AA}$ (Cu-K α wavelength), β is the full width at half maximum (FWHM), and θ is the Bragg angle. According to the Scherrer equation, the crystallite size of g-C₃N₄ in the ternary composite is calculated to be 10.5 nm, comparable to that of pure g-C₃N₄ (26.5 nm). This minor FWHM variation may arise from the interfacial interaction between g-C₃N₄ and the loaded Ag₂O/CuFe₂O₄ nanoparticles, but it does not indicate obvious lattice distortion or crystallinity degradation of g-C₃N₄. This result is consistent with the earlier conclusion that the heterostructure synthesis preserves the intrinsic crystallinity of each component, which is critical for maintaining efficient charge transfer at the interfaces. Based on the dominant (200) and (113) peaks, Ag₂O and CuFe₂O₄ crystallites averaged $51.44 \pm 1.2 \text{ nm}$ and $48.14 \pm 0.9 \text{ nm}$, respectively, confirming nanoscale dimensions (Fig. 1d).

Morphological and structural characterization (SEM/TEM)

The microstructural evolution of the synthesized composites was systematically investigated using scanning and transmission electron microscopy. With the unified scale in Fig. 2, the following features are more clearly comparable: (1) pristine g-C₃N₄ exhibits curled lamellar nanosheets with a lateral size of 1–2 μm , consistent with its layered graphitic structure; (2) Ag₂O nanoparticles (35–45 nm) in Ag₂O@g-C₃N₄ are uniformly dispersed on the g-C₃N₄ nanosheets without obvious agglomeration; (3) CuFe₂O₄ particles (50–85 nm) in CuFe₂O₄@g-C₃N₄ are embedded in the interlayer gaps of g-C₃N₄, with a slightly larger particle size than Ag₂O; (4) in the ternary composite, Ag₂O (40 \pm 5 nm) and CuFe₂O₄ (60 \pm 8 nm) nanoparticles maintain heterogeneous distribution on g-C₃N₄ nanosheets, with no significant change in the lamellar structure of g-C₃N₄ compared to the pristine sample. This unified scale not only clarifies the morphological evolution of the composite system but also confirms the effective interfacial coupling between each component during synthesis. This competitive nucleation and growth process, coupled with the surface energy heterogeneity of g-C₃N₄, directly results in the observed selective attachment of Ag₂O and CuFe₂O₄, which further supports the formation of a well-dispersed ternary heterostructure.

High-resolution TEM (Fig. 3) provides crystallographic validation of the heterostructure. Pristine g-C₃N₄ (Fig. 3a) shows amorphous domains interspersed with ordered lattice fringes, consistent with its (002) interlayer stacking. In Ag₂O@g-C₃N₄ (Fig. 3b), discrete Ag₂O nanoparticles (30–50 nm) are anchored to the g-C₃N₄ surface. CuFe₂O₄@g-C₃N₄ (Fig. 3c) reveals tetragonal CuFe₂O₄ particles (50–85 nm) are anchored to the g-C₃N₄ surface.

For the ternary composite (Fig. 3d), cubic-phase Ag₂O and tetragonal CuFe₂O₄ nanoparticles (20–50 nm) are co-located on g-C₃N₄ nanosheets, forming intimate heterojunctions. The distinct lattice fringes (Ag₂O: 0.2372 nm (2 0 0), AgO: 0.2287 nm

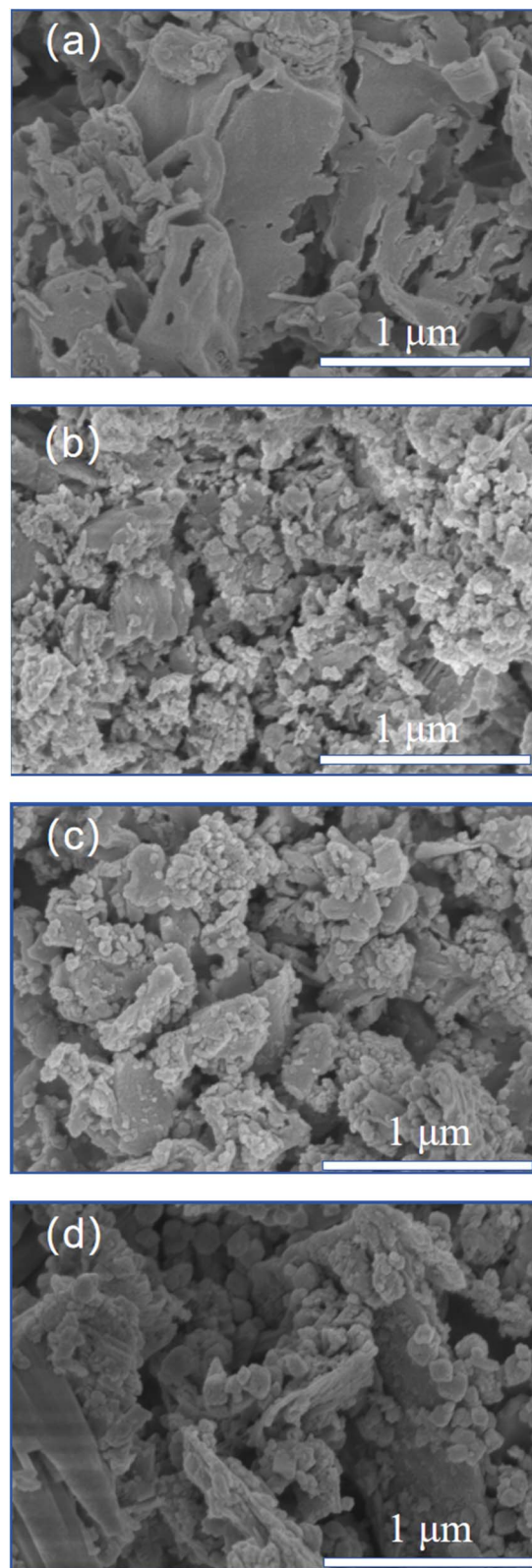


Fig. 2 SEM images of (a) g-C₃N₄, (b) Ag₂O@g-C₃N₄, (c) CuFe₂O₄@g-C₃N₄ and (d) Ag₂O@CuFe₂O₄@g-C₃N₄.

(2 0 1), 0.2236 nm(1 1 1), CuFe₂O₄: 0.219 nm (4 0 0)) and absence of interfacial amorphous layers confirm crystallographic coherence, critical for efficient charge transfer.



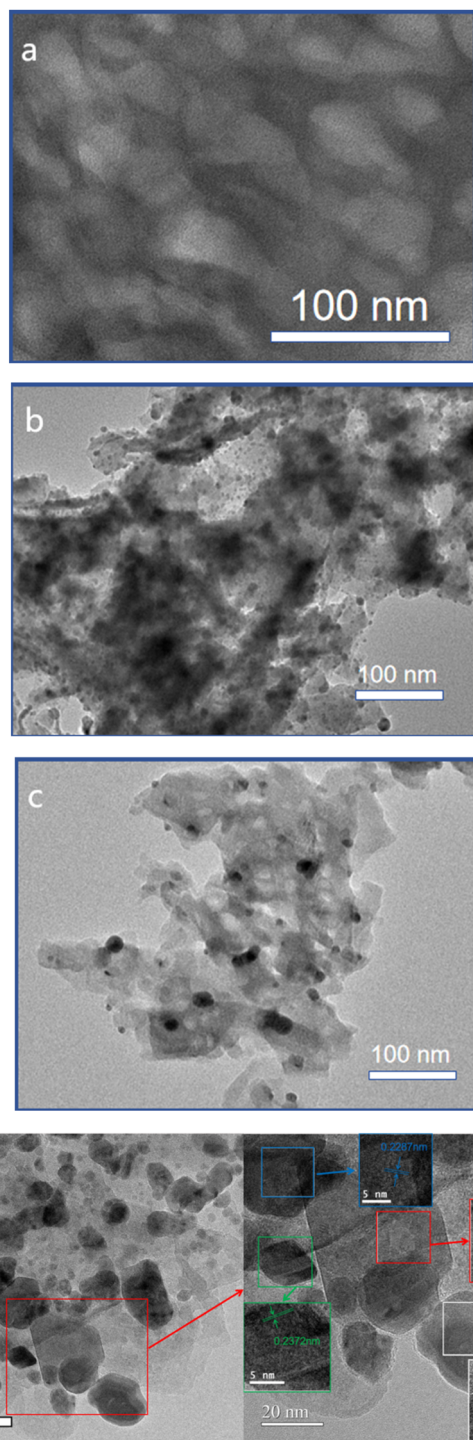


Fig. 3 TEM images of (a) g-C₃N₄, (b) Ag₂O@g-C₃N₄, (c) CuFe₂O₄@g-C₃N₄ and (d) Ag₂O@CuFe₂O₄@g-C₃N₄.

XPS analysis

X-ray photoelectron spectroscopy (XPS) was employed to probe the chemical states and surface composition of the Ag₂O@CuFe₂O₄@g-C₃N₄ heterostructure (Fig. 4a). The survey spectrum confirms the presence of C, O, N, Ag, Fe, and Cu, corroborating the successful integration of Ag₂O and CuFe₂O₄ into the g-C₃N₄ matrix (Yao *et al.*, 2015). High-resolution Ag 3d

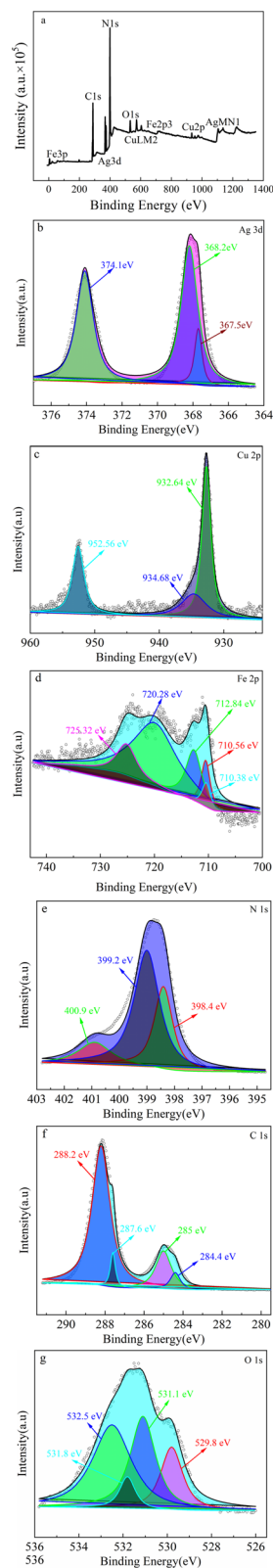


Fig. 4 XPS analysis of Ag₂O@CuFe₂O₄@g-C₃N₄ heterojunction emerging from the emissions of the Ag, Cu, Fe, N, C, and O elements; (a) survey spectrum of Ag₂O@CuFe₂O₄@g-C₃N₄, (b) Ag 3d, (c) Cu 2p, (d) Fe 2p, (e) N 1s, (f) C 1s and (g) O 1s.



spectra (Fig. 4b) and $\Delta E = E_{3d_{3/2}} - E_{3d_{5/2}} \approx 6.0$ eV reveal peaks at 368.3 eV (Ag $3d_{5/2}$) and 374.3 eV (Ag $3d_{3/2}$), deconvoluted into three components at 367.5 eV (Ag⁰), 368.2 eV (Ag⁺ in Ag₂O), and 374.1 eV (Ag²⁺ in AgO), indicating partial oxidation. The Cu 2p spectrum (Fig. 4c) displays peaks at 932.4 eV (Cu $2p_{3/2}$) and 952.2 eV (Cu $2p_{1/2}$), resolved into Cu⁺ (932.64 eV) and Cu²⁺ (934.68 eV), confirming Cu₂O and CuO coexistence. Similarly, the Fe $2p_{3/2}$ peak resolves into contributions at 710.38 eV, 710.56 eV (Fe²⁺ in FeO) and 712.84 eV (Fe³⁺ in Fe₂O₃), while Fe $2p_{1/2}$ peaks at 725.32 eV and 720.28 eV further validate mixed Fe oxidation states. N 1s spectra (Fig. 4e) resolve into three peaks at 398.4 eV (C=N-C), 399.2 eV (N-(C)₃), and 400.9 eV (C-N-H), characteristic of g-C₃N₄'s triazine structure. C 1s signals (Fig. 4f) at 284.4 eV (C-C), 285.0 eV (C-O/N), 287.6 eV (sp² C in g-C₃N₄), and 288.2 eV (O-C=O) reflect graphitic and functionalized carbon. O 1s deconvolution (Fig. 4g) identifies contributions from lattice oxygen in CuO/Ag₂O/Fe₂O₃ (529.7–531.1 eV), surface hydroxyl groups (531.8 eV), and adsorbed H₂O (532.5 eV). Collectively, the XPS data confirm the coexistence of Ag⁰/Ag⁺/Ag²⁺, Cu⁺/Cu²⁺, and Fe²⁺/Fe³⁺ redox pairs, anchored as Ag₂O, Cu₂O/CuO, and FeO/Fe₂O₃ on g-C₃N₄ nanosheets. Combined with TEM evidence, this confirms the formation of a multifunctional heterostructure with interfacial charge-transfer pathways, critical for visible-light-driven photocatalytic activity.

Optical properties

The light-harvesting capabilities and charge-carrier dynamics of the synthesized materials were investigated through UV-Vis diffuse reflectance spectroscopy (DRS) and photoluminescence (PL) spectroscopy. As shown in Fig. 5a, the Ag₂O@CuFe₂O₄@g-C₃N₄ heterostructure exhibits a pronounced red shift in absorption edge and enhanced visible-light absorption compared to pristine g-C₃N₄, attributed to synergistic interactions between the constituent phases. The optical bandgap energies (E_g), calculated *via* Tauc plots (Fig. 5b), decrease sequentially from 2.74 eV (g-C₃N₄) to 1.82 eV (ternary composite), confirming extended light absorption into the visible-NIR range. This broadening aligns with the formation of a p-n heterojunction, which enhances solar energy utilization by minimizing charge recombination. PL spectra (Fig. 5c) further reveal a marked quenching of emission intensity in the heterostructure, signifying suppressed electron-hole recombination due to efficient interfacial charge transfer between Ag₂O, CuFe₂O₄, and g-C₃N₄. These findings collectively demonstrate that the heterostructure design optimizes both light absorption and charge separation, critical for high-performance photocatalysis under visible irradiation.

As shown in Fig. 5, the interfacial charge transfer and separation of the Ag₂O@CuFe₂O₄@g-C₃N₄ ternary composite follow a dual Z-scheme heterojunction mechanism, whose thermodynamic feasibility and kinetic processes are jointly verified by band alignment, charge migration pathways, and experimental characterizations. Firstly, the band structures of the three components were determined *via* UV-Vis diffuse reflectance spectroscopy (Fig. 5) and X-ray photoelectron

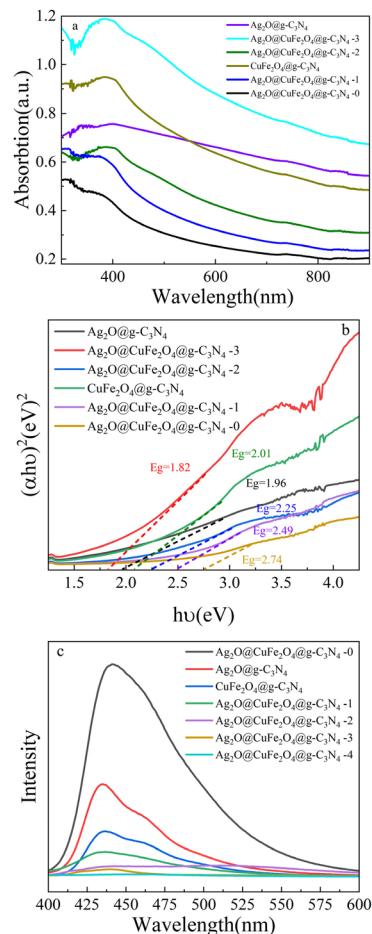


Fig. 5 (a) Diffuse reflectance spectra of Ag₂O@g-C₃N₄, CuFe₂O₄@g-C₃N₄ and Ag₂O@CuFe₂O₄@g-C₃N₄; (b) plot of transferred Kubelka–Munk versus the energy of the light absorbed for of Ag₂O@g-C₃N₄, CuFe₂O₄@g-C₃N₄ and Ag₂O@CuFe₂O₄@g-C₃N₄; (c) PL spectra of Ag₂O@g-C₃N₄, CuFe₂O₄@g-C₃N₄ and Ag₂O@CuFe₂O₄@g-C₃N₄.

spectroscopy (Fig. 4) valence band (VB) spectra:^{2,7} for g-C₃N₄, the optical bandgap (E_g) is 2.74 eV, the VB potential (E_{VB}) is +2.46 eV vs. normal hydrogen electrode (NHE), and the conduction band (CB) potential (E_{CB}) is −0.43 eV; for CuFe₂O₄, $E_g = 1.42$ eV, $E_{VB} = +0.21$ eV, and $E_{CB} = -1.21$ eV; for Ag₂O, $E_g = 1.20$ eV, $E_{VB} = +1.35$ eV, and $E_{CB} = +0.15$ eV. This band alignment satisfies the thermodynamic requirements for a dual Z-scheme heterojunction. The CB potential of g-C₃N₄ (−0.43 eV) is more negative than the VB potential of CuFe₂O₄ (+1.42 eV), which enables photogenerated electrons in the CB of g-C₃N₄ to transfer to the VB of CuFe₂O₄ and recombine with holes; the CB potential of Ag₂O (+0.15 eV) is more positive than the VB potential of g-C₃N₄, driving electrons in the CB of Ag₂O to migrate to the VB of g-C₃N₄ and recombine with holes. In contrast, the traditional type-II heterojunction is thermodynamically disadvantageous, as the CB of Ag₂O cannot reduce O₂ to 'O₂[−] and the VB of CuFe₂O₄ cannot oxidize H₂O to 'OH. Therefore, the dual Z-scheme structure is the only effective configuration for retaining redox-active electron-hole pairs.

Under visible-light irradiation, the charge transfer process proceeds in three steps. In the charge generation step, g-C₃N₄



and Ag_2O absorb visible light to generate electron-hole pairs, while CuFe_2O_4 only produces a small number of electron-hole pairs due to weak visible-light absorption. In the Z-scheme charge transfer step (mediated by Ag^0), in the first Z-scheme pathway, electrons in the CB of $\text{g-C}_3\text{N}_4$ migrate to the VB of CuFe_2O_4 and recombine with holes; XPS results show an increased content of Fe^{2+} (at 710.38 eV) in the ternary composite, confirming that electron injection into CuFe_2O_4 triggers the conversion of Fe^{3+} to Fe^{2+} , while retaining holes at +1.4 eV in the VB of $\text{g-C}_3\text{N}_4$ and electrons at -1.28 eV in the CB of CuFe_2O_4 . In the second Z-scheme pathway, electrons in the CB of Ag_2O cannot directly transfer to the VB of $\text{g-C}_3\text{N}_4$ due to an energy barrier; however, the partial reduction of Ag_2O to Ag^0 (XPS Ag 3d_{5/2} peak at 367.5 eV) forms an electron bridge, mediating electron transfer from the CB of Ag_2O to the VB of $\text{g-C}_3\text{N}_4$ *via* Ag^0 for recombination with holes, and retaining strongly oxidizing holes at +2.46 eV in the VB of Ag_2O and electrons at -1.34 eV in the CB of $\text{g-C}_3\text{N}_4$.

In the charge separation and reactive oxygen species (ROS) generation step, the retained electrons reduce O_2 to $\cdot\text{O}_2^-$, and the retained holes oxidize $\text{H}_2\text{O}/\text{OH}^-$ to $\cdot\text{OH}$. Additionally, $\text{Fe}^{2+}/\text{Cu}^+$ in CuFe_2O_4 react with H_2O_2 (generated by the protonation of $\cdot\text{O}_2^-$) through Fenton-like cycles, further enhancing $\cdot\text{OH}$ generation. Moreover, photoluminescence spectroscopy (Fig. 5c) confirm that the dual Z-scheme structure not only achieves spatial separation of electrons and holes but also retains their strong redox capabilities, effectively addressing the key limitations of single/binary catalysts.

Photocatalytic performance

Recent advances in 2D heterostructures have further pushed the boundary of visible-light photocatalysis, as their ultra-thin planar structure provides large specific surface area and short charge-transfer paths.^{17,18} These studies confirm the potential of 2D heterostructures but also highlight their limitations. To address these, we design a ternary $\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$ 2D heterostructure that integrates Z-scheme charge transfer and Fenton-like reactivity—filling the gap in multi-functional 2D heterostructure design. The visible-light-driven photocatalytic activity of the $\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$ composites was evaluated through methylene blue (MB) degradation (30 mg L⁻¹) under irradiation from a 300 W metal-halide lamp ($\lambda \geq 420$ nm). Prior to illumination, 0.05 g of catalyst was dispersed in 100 mL of MB solution and magnetically stirred in the dark for 60 min to establish adsorption-desorption equilibrium. Aliquots (3 mL) were extracted at 5 min intervals, centrifuged to remove catalyst particles, and analyzed *via* UV-Vis spectrophotometry by monitoring the MB absorbance at 664 nm (Fig. 6a). Fig. 6a shows the time-dependent UV-Vis spectra of MB during degradation, where the characteristic peak of MB at 664 nm (attributed to the $\pi \rightarrow \pi^*$ transition of the conjugated chromophore in MB molecules) decreases significantly with irradiation time, indicating efficient decolorization. Notably, the minor absorption peak near 300 nm also depresses synchronously, and this peak is assigned to the $\pi \rightarrow \pi^*$ transition of the aromatic benzene/heterocyclic rings in the MB molecular skeleton. The

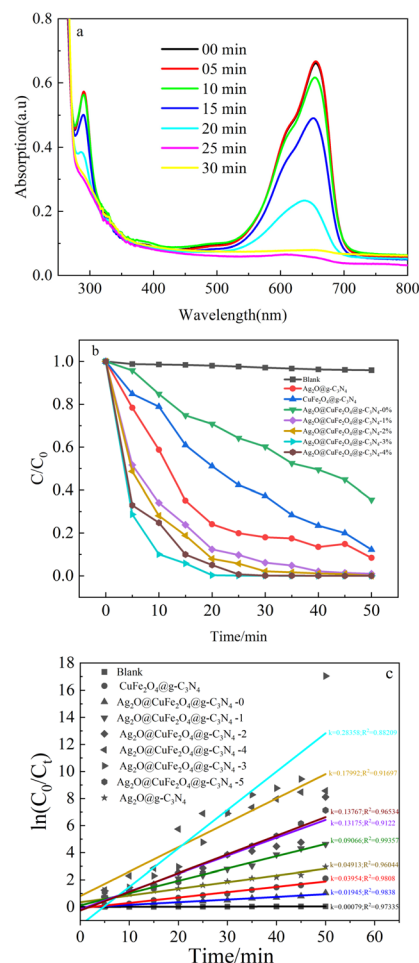


Fig. 6 The degradation curves of MB under visible lights change of (a) character peak, (b) degradation rate, (c) kinetic plots with irradiated time.

synchronous reduction of both peaks confirms that the $\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$ catalyst not only breaks the conjugated chromophore (responsible for MB's blue color) but also degrades the aromatic/heterocyclic backbone of MB—avoiding the accumulation of toxic intermediate products with intact aromatic rings. This is further supported by the absence of new absorption peaks in the 250–400 nm range (Fig. 6a), which rules out the formation of stable intermediate species and confirms the thorough mineralization of MB.

As shown in Fig. 6b, the ternary $\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$ -3% composite achieved 99.94% MB degradation within 35 min, outperforming pristine $\text{g-C}_3\text{N}_4$ (47.55%), $\text{Ag}_2\text{O}@ \text{g-C}_3\text{N}_4$ (82.41%), and $\text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$ (71.56%). Control experiments confirmed negligible photolysis (<5%) in the absence of catalysts. The degradation kinetics followed pseudo-first-order behavior, with the ternary composite exhibiting a rate constant $k = 0.132 \text{ min}^{-1}$, 2.10-, 1.21-, and 1.40-fold higher than $\text{g-C}_3\text{N}_4$, $\text{Ag}_2\text{O}@ \text{g-C}_3\text{N}_4$, and $\text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$, respectively. Notably, 90% MB degradation occurred within 17–20 min, highlighting rapid reaction kinetics (Fig. 6c). As shown in Fig. S1, the ternary $\text{Ag}_2\text{O}@ \text{CuFe}_2\text{O}_4 @ \text{g-C}_3\text{N}_4$ -3% have a better

Table 2 Comparison of catalytic efficiency of Ag₂O@CuFe₂O₄@g-C₃N₄ catalyst with previous literature for the reduction of MB

Organic dye	Catalyst	k_{app} (s ⁻¹)	Ref.
MB	g-C ₃ N ₄ /NiFe ₂ O ₄ (1 g l ⁻¹ , r.t.)	7.33×10^{-3}	16
MB	NiO-g-C ₃ N ₄ (0.5 g l ⁻¹ , r.t.)	8.33×10^{-3}	16
MB	g-C ₃ N ₄ /ZnO (0.5 g l ⁻¹ , r.t.)	3.10×10^{-3}	16
MB	CaFe ₂ O ₄ /g-C ₃ N ₄ (20 wt% CaFe ₂ O ₄ , 2 g l ⁻¹ , r.t.)	0.288×10^{-3}	16
MB	g-C ₃ N ₄ quantum dots/ZnO nanosheets (0.125 g l ⁻¹ , r.t.)	0.855×10^{-3}	16
MB	Graphite carbon coating hollow CuFe ₂ O ₄ spheres (0.1 g l ⁻¹ , r.t.)	0.633×10^{-3}	16
MB	Ag ₂ O@CuFe ₂ O ₄ @g-C ₃ N ₄ -1 (0.5 g l ⁻¹ , r.t.)	1.511×10^{-3}	This work
	Ag ₂ O@CuFe ₂ O ₄ @g-C ₃ N ₄ -2 (0.5 g l ⁻¹ , r.t.)	2.196×10^{-3}	
	Ag ₂ O@CuFe ₂ O ₄ @g-C ₃ N ₄ -3 (0.5 g l ⁻¹ , r.t.)	4.726×10^{-3}	
	Ag ₂ O@CuFe ₂ O ₄ @g-C ₃ N ₄ -4 (0.5 g l ⁻¹ , r.t.)	2.998×10^{-3}	
	Ag ₂ O@CuFe ₂ O ₄ @g-C ₃ N ₄ -5 (0.5 g l ⁻¹ , r.t.) time	2.295×10^{-3}	

catalysis performance than pristine Ag₂O, pristine CuFe₂O₄ and pristine g-C₃N₄.

Optimal performance was observed at a 3% Ag₂O@CuFe₂O₄ loading, where the heterojunction structure facilitated efficient charge separation while minimizing agglomeration. Increasing the loading to 4% reduced degradation efficiency to 99.91% due to particle aggregation, which restricted charge carrier mobility and increased electron-hole recombination. This underscores the critical balance between heterojunction density and dispersion: at 3% loading, uniformly distributed Ag₂O and CuFe₂O₄ nanoparticles on g-C₃N₄ nanosheets enhanced light absorption, active site availability, and interfacial charge transfer, whereas excessive loading (4%) disrupted these synergies. To highlight the advantages of the Ag₂O@CuFe₂O₄@g-C₃N₄ ternary composite, its MB degradation performance was compared with that of similar g-C₃N₄-based photocatalysts reported in recent literature (Table 2).¹⁶

As shown in Table 2, the ternary Ag₂O@CuFe₂O₄@g-C₃N₄ composite outperforms most literature-reported counterparts. Firstly, its rate constant 1.511×10^{-3} to 4.726×10^{-3} are higher than most g-C₃N₄ coupled catalyst; Secondly, it achieves nearly complete MB removal in 35 min, which higher than most g-C₃N₄ coupled catalyst under the same MB concentration and catalyst dosage; Finally, it retains >95% activity after 10 cycles, outperforming catalysts with fewer cycles or lower retention.

This enhanced performance originates from the synergistic effect of the Z-scheme heterojunction and Fenton-like reactivity, which addresses the limitations of single/binary catalysts reported in literature. These results confirm that

Ag₂O@CuFe₂O₄@g-C₃N₄ is a competitive candidate for sustainable MB-contaminated water remediation.

The recyclability of the Ag₂O@CuFe₂O₄@g-C₃N₄ composite was evaluated *via* ten consecutive MB degradation cycles, following a standardized protocol to ensure consistent experimental conditions across cycles as description in Characterization. The recyclability of the Ag₂O@CuFe₂O₄@g-C₃N₄ composite was evaluated over ten consecutive photocatalytic cycles (Fig. 7). The material retained >95% of its initial degradation efficiency, with no significant decline in performance, demonstrating exceptional cyclic stability. This robust retention of activity underscores the structural integrity of the Ag₂O@CuFe₂O₄@g-C₃N₄ heterojunction, confirming that the interfacial coupling between Ag₂O, CuFe₂O₄, and g-C₃N₄ remains stable under prolonged photocatalytic operation.

Active species analysis

To elucidate the dominant reactive species governing the photocatalytic mechanism, radical trapping experiments were conducted using tert-butanol (TBA) as a hydroxyl radical ([•]OH) scavenger. Under visible-light irradiation (300 W metal-halide lamp, $\lambda \geq 420$ nm), the degradation of MB by Ag₂O@CuFe₂O₄@g-C₃N₄ was evaluated in the presence and absence of TBA (Fig. 8). The introduction of TBA drastically suppressed MB degradation, reducing efficiency from 99.93% to 26.68% within 25 min, confirming [•]OH as the primary oxidative species. Notably, residual degradation in TBA-containing systems

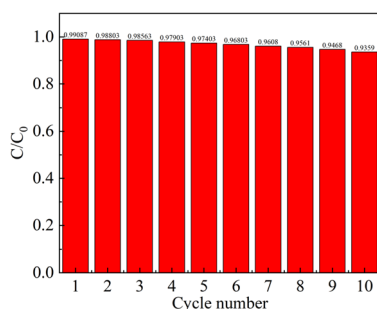


Fig. 7 Recycling test of MB using Ag₂O@CuFe₂O₄@g-C₃N₄ catalysts.

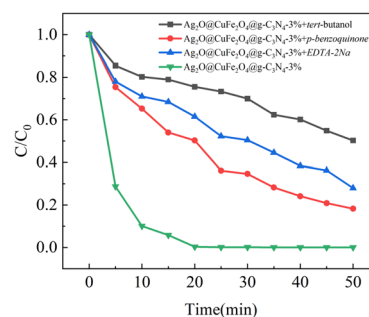


Fig. 8 Influences of different scavengers on photocatalytic degradation of MB in the presence of Ag₂O@CuFe₂O₄@g-C₃N₄-3 catalysts.



(Fig. 8) suggests minor contributions from alternate pathways (e.g., direct h^+ oxidation or $\cdot O_2^-$ participation). However, the near-complete quenching of activity underscores the pivotal role of $\cdot OH$, generated *via* H_2O oxidation by photogenerated holes and Fenton-like reactions involving $CuFe_2O_4$. These results align with the heterostructure's band alignment, which facilitates spatially separated redox reactions, optimizing $\cdot OH$ yield for rapid dye mineralization.

The reactive species generated during visible-light photocatalysis were systematically probed through radical trapping experiments. As depicted in Fig. 8, the degradation profiles of MB were compared across four systems: (i) pristine $Ag_2O@CuFe_2O_4@g-C_3N_4$, (ii) composite with *p*-benzoquinone (BQ, $\cdot O_2^-$ scavenger), (iii) composite with tert-butanol (TBA, $\cdot OH$ scavenger), and (iv) composite with disodium ethylenediaminetetraacetate dihydrate (EDTA-2Na, h^+ scavenger). Under 50 min of visible-light irradiation, the degradation efficiency declined from 98.06% (pristine catalyst) to 53.08% (BQ), 20.15% (TBA), and 38.60% (EDTA-2Na) within 20 min. This hierarchical suppression—most pronounced with TBA—confirmed hydroxyl radicals ($\cdot OH$) as the dominant active species, with supplementary roles of superoxide radicals ($\cdot O_2^-$) and holes (h^+). The results underscore a reaction mechanism where $\cdot OH$, produced *via* H_2O oxidation by h^+ and Fenton-like Fe^{2+}/Cu^+ cycles, drives MB mineralization, while $\cdot O_2^-$ and direct h^+ oxidation contribute secondary pathways.

Proposed activation mechanism

Fig. 9 schematically illustrates the Z-scheme charge transfer and photo-Fenton synergistic mechanism for MB degradation by $Ag_2O@CuFe_2O_4@g-C_3N_4$, with band positions of $g-C_3N_4$, $CuFe_2O_4$, and Ag_2O referenced to literature, the dual Z-scheme adapted from $g-C_3N_4$ -based heterostructure principles, and photo-Fenton reactions ($Fe^{2+}/Cu^+ + H_2O_2 \rightarrow \cdot OH + Fe^{3+}/Cu^{2+}$) aligning with $CuFe_2O_4$'s reported redox behavior. Based on discussed in interfacial charge transfer and separation^{2,7,17,18} of the $Ag_2O@CuFe_2O_4@g-C_3N_4$ ternary composite. Under visible light ($\lambda \geq 420$ nm), three steps occur: (1) charge generation; (2) Z-scheme transfer; (3) retained e^- reduce O_2 to $\cdot O_2^-$, retained h^+ oxidize OH^- to $\cdot OH$. The equations in the Fig. 9 show photo-Fenton ROS amplification: $\cdot O_2^-$ protonates to H_2O_2 , $CuFe_2O_4$'s Fe^{3+}/Cu^{2+} reduce to Fe^{2+}/Cu^+ , and $\cdot OH/\cdot O_2^-/h^+$ mineralize MB.

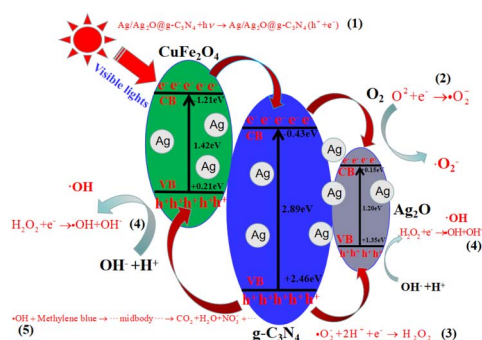


Fig. 9 Improvement photocatalytic oxidation of MB over $Ag_2O@CuFe_2O_4@g-C_3N_4$ heterostructures.

XPS, quenched PL, and TBA-suppressed efficiency validate the mechanism.

Conclusions

In this study, a highly efficient $Ag_2O@CuFe_2O_4@g-C_3N_4$ heterostructured composite was successfully synthesized, integrating Z-scheme charge transfer and Fenton-like reactivity for visible-light-driven organic dye degradation. Structural and compositional analyses confirmed the incorporation of Fe^{3+} , Cu^{2+} , and Cu^+ species into the $g-C_3N_4$ framework, while the introduction of Ag_2O and $CuFe_2O_4$ enhanced specific surface area and adsorption capacity for methylene blue (MB). The Z-scheme heterojunction, governed by the band alignment of Ag_2O (-0.5 eV CB), $CuFe_2O_4$ ($+0.7$ eV VB), and $g-C_3N_4$ (-1.3 eV CB/ $+1.4$ eV VB), facilitated spatial separation of photogenerated electron-hole pairs, suppressing recombination and prolonging carrier lifetimes. Recyclability tests demonstrated exceptional stability, with $>95\%$ MB degradation retained over ten cycles, underscoring the catalyst's durability and surface-dominated reaction mechanism. Radical trapping experiments identified hydroxyl radicals ($\cdot OH$) as the primary active species, generated *via* h^+ -mediated H_2O oxidation and Fenton-like Fe^{3+}/Cu^{2+} redox cycles, with supplementary contributions from superoxide radicals ($\cdot O_2^-$). The synergistic interplay between the Z-scheme heterojunction (enhancing light absorption and charge separation) and Fenton-like activity (amplifying ROS generation) endowed the composite with superior photocatalytic performance, achieving 99.94% MB degradation within 35 min. This work establishes $Ag_2O@CuFe_2O_4@g-C_3N_4$ as a robust, multifunctional catalyst for sustainable wastewater treatment, offering a scalable strategy to address organic pollution through solar energy utilization.

Author contributions

Jianhua Wang – funding acquisition, methodology, material synthesis, article writing, chart output, grammar correction, *etc.* Huihui Fang – material preparation, degradation experiments, grammar correction, *etc.* Shuhui Shi – project administration, supervision, grammar correction, article writing, chart output, *etc.*

Conflicts of interest

There are no conflicts to declare.

Data availability

The data supporting this article have been included as part of the supplementary information (SI). Supplementary information is available. See DOI: <https://doi.org/10.1039/d5ra02882e>.

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