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Enhanced photocatalytic degradation of methylene blue dye via valorization of a polyethylene terephthalate plastic waste-derived metal–organic framework-based $\text{ZnO}@\text{Co-BDC}$ composite catalyst†

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The growth of industrialization contributes to the pollution of natural water bodies. Among these pollutants, organic dyes and plastic waste materials account for the majority of contaminants and are associated with health risks. This research explores a metal–organic framework (MOF)-based composite catalyst, $\text{ZnO}@\text{Co-BDC}$, for the degradation of methylene blue (MB). We developed a Co-BDC MOF synthesized through a solvothermal route using terephthalic acid (BDC) as a linker, which was extracted from polyethylene terephthalate plastic waste via alkaline hydrolysis, and using cobalt nitrate hexahydrate as a cobalt source. The $\text{ZnO}@\text{Co-BDC}$ composite catalyst was synthesized through a solvothermal route using cobalt, BDC and zinc precursors. The obtained products were characterized using powder X-ray diffraction, transmission electron microscopy, Fourier transform infrared spectroscopy, and ultraviolet diffuse reflectance spectroscopy (UV-DRS). The performance of $\text{ZnO}@\text{Co-BDC}$ was assessed for the degradation of MB and showed 87.5% under visible light irradiation for 80 min, surpassing pristine ZnO (74%) and Co-BDC MOF (39%) under the same conditions. The kinetic study indicated that the degradation followed first-order kinetics with a rate constant of $2.501 \times 10^{-2} \text{ min}^{-1}$. Furthermore, the effects of catalyst dose, irradiation time, pH, and MB concentration were optimized for the efficient composite catalyst $\text{ZnO}@\text{Co-BDC}$. The photodegradation mechanism was also investigated through UV-DRS and quenching experiments in the presence of different scavengers. Meanwhile, the developed composite demonstrated excellent recovery and reuse capabilities for up to six cycles under optimal conditions. The developed MOF-based composite catalyst enabled the simultaneous valorization of plastic waste and remediation of environmental pollutants by converting waste to wealth.

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1. Introduction

Water scarcity has become a pressing global issue due to the rapid increase in human population, industrialization, and pollution.¹ One significant contributor to water pollution is the presence of organic dyes, such as methylene blue (MB), which are commonly used in industries including dye manufacturing, leather production, and textile manufacturing.² These dyes are

known to have detrimental effects on both human health and the environment, as many of them are carcinogenic and can persist in water bodies for extended periods.³ As a result, there is an urgent need to develop effective and sustainable methods for the removal of organic dyes to mitigate their harmful impacts.

Traditional methods for the removal of organic dyes from water, such as filtration and chemical treatments, are often costly and inefficient. Moreover, these methods may generate secondary pollutants, posing additional environmental concerns. Therefore, there is a growing interest in exploring alternative technologies that are both cost-effective and environmentally friendly. Photocatalysis has emerged as a promising approach for the degradation of organic dyes, offering several advantages over traditional methods.^{4,5} Photocatalysts are materials that can initiate chemical reactions under light irradiation. They harness the energy from light to generate

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highly reactive species, such as hydroxyl radicals, which can effectively degrade organic pollutants.⁶ Photocatalysis offers a sustainable and efficient solution for the degradation of organic dyes, as it does not require the addition of harmful chemicals and does not produce secondary pollutants. Furthermore, photocatalytic processes can be tailored to be environmentally safe, as they can degrade organic dyes into harmless components that do not pose a risk to ecosystems.^{4,7,8}

Photocatalytic technology using semiconductors has gained more attention because it is very effective and has low toxicity.⁹ So far, many different semiconductor photocatalysts have been reported for converting solar energy and treating organic pollutants, *e.g.*, TiO_2 ,¹⁰ $\text{g-C}_3\text{N}_4$,¹¹ ZnO ¹² and BiOCl .¹³ Among these semiconductors, the ZnO and TiO_2 -based photocatalysts have garnered considerable scientific interest due to the favorable alignment of their valence and conduction band edges with the reduction and oxidation potential energies of water and oxygen.¹² This advantageous positioning enables the efficient generation of reactive oxygen species (ROS), including hydroxyl radicals and superoxide radicals, which are critical for the photocatalytic degradation processes. By effectively producing these ROS, ZnO and TiO_2 photocatalysts can degrade pollutants into less harmful compounds, significantly enhancing their performance in environmental remediation applications.^{13–15} However, ZnO has several drawbacks in practical applications, including poor utilization of visible light due to its wide band gap of 3.27 eV, quick recombination of photogenerated electron-hole pairs, and a low specific surface area. Many efforts have been made to address these drawbacks, including semiconductor coupling,¹⁴ doping with metals or non-metals,¹⁶ self-assembly,¹⁷ and using the template method.¹⁸ The template method using metal organic frameworks (MOFs) can solve the problem of low visible light utilization, rapid charge recombination and low specific surface area.¹³ The porous structure, characterized by a high specific surface area, enhances the adsorption of pollutant molecules onto the active sites of photocatalysts, improves light transmittance, and promotes the rapid transfer of photogenerated charge carriers to the photocatalyst surface. This facilitates efficient bulk charge separation.¹⁶

MOFs are porous materials formed by the connection of metal ions or clusters with organic ligands through robust chemical bonds, creating structures that can be one-dimensional, two-dimensional, or three-dimensional. These materials exhibit remarkable features, including high surface area, exceptional thermal stability, and unique porosity.¹⁹ Additionally, MOFs have a tunable topology and functionality, the ability to absorb light in the UV-Vis range, rich host-guest chemistry, well-developed pores for high accessibility of active sites, and Lewis acid sites that can be tuned by the formation of missing ligand defects.²⁰ These desirable properties make them one of the significant materials in the advancements across various fields in recent decades, including applications in luminescence sensors, adsorption/desorption, gas storage, molecular separation, drug delivery, and catalysis.^{21–23} In earlier reports, MOFs have been widely reported to be efficient photocatalysts, including in the photoreduction of CO_2 ,

photodegradation of organic dyes, and water splitting.^{24–26} Several representative MOFs, such as ZIF-8,²⁷ MOF-5,²⁸ MIL-125,²⁹ MIL-53,³⁰ MIL-101,³¹ HKUST-1,³² and UiO-66,³³ have been recognized for their high performance and diverse potential applications.

This study focuses on Co-BDC MOF, derived from polyethylene terephthalate (PET) plastic waste and integrated with a zinc source to form the $\text{ZnO}@\text{Co-BDC}$ composite catalyst, which serves as an efficient catalyst for the degradation of MB dye. PET plastic waste poses a significant environmental challenge due to its resistance to biodegradation, leading to its accumulation in landfills and water bodies, with detrimental effects on ecosystems, wildlife, and human health.³⁴ Transforming plastic waste into valuable materials like the Co-BDC MOF offers a sustainable solution for removing MB and other organic dyes while enhancing environmental protection. Repurposing plastic waste into functional materials helps decrease the amount of plastic in landfills and water bodies, reducing pollution and minimizing ecological harm. The $\text{ZnO}@\text{Co-BDC}$ composite catalyst exhibited superior performance, achieving a MB degradation rate of 87.5% using visible light irradiation for 80 minutes, surpassing pristine ZnO (74%) and Co-BDC MOF (39%). This research contributes to cost-effective and environmentally friendly solutions for water pollution remediation, emphasizing the significance of utilizing PET plastic waste and highlighting the potential of the composite in addressing dye contamination in water sources.

2. Experimental section

2.1. Chemicals and reagents

Zinc acetate hexahydrate ($\text{Zn}(\text{CH}_3\text{COOH})_2 \cdot 2\text{H}_2\text{O}$, 98%), cobalt nitrate hexahydrate ($\text{Co}(\text{NO}_3)_2 \cdot 6\text{H}_2\text{O}$, 99.5%), ethylene glycol (EG, 99.8%), ethanol (95%), *N,N*-dimethyl formamide (DMF, 99.8%), sodium hydroxide (NaOH, 99%), *p*-benzoquinone (BQ, 99.8%), isopropanol ($\text{C}_3\text{H}_7\text{OH}$, 98.7%), and disodium ethylenediamine-tetraacetate (EDTA, 99.8%) were utilized in this study without any additional treatment. The PET plastics used for the preparation of terephthalic acid (BDC) were gathered from household waste. Deionized water was used in all experiments conducted during this research.

2.2. Synthesis of BDC from PET plastic waste

The depolymerization of PET can be conducted *via* both acidic and alkaline hydrolysis methods, with some adjustments based on a previous study.³⁵ In the alkaline hydrolysis process, plastic bottles were first cut into small flakes using scissors. These flakes were then washed with deionized water and ethanol. A mixture of 10 g of NaOH, 50 mL of ethylene glycol, and 10 g of PET flakes was transferred into a round-bottom flask equipped with a reflux condenser and heated on a hot-plate stirrer. The mixture was boiled, and the heating process was stopped after a certain period. This cycle was repeated for 1 h at a temperature of 200 °C. Throughout this time, the plastic gradually decomposed, resulting in the liquid turning milky white. The flask was then removed from the hot plate and allowed to cool. After 30



minutes, 300 mL of water was added to the mixture with continuous stirring. The resulting solution was filtered using Whatman filter paper to eliminate undecomposed particles. Afterward, 5 M of sulfuric acid was added, and the mixture was allowed to stir vigorously for 30 min. Subsequently, the solution was centrifuged to separate the white precipitates and washed four times with deionized water. Then, the resulting product (BDC) was dried in an oven at 100 °C. The extracted BDC was saved for further use as a linker.

2.3. Synthesis of Co-BDC MOF

The Co-BDC MOF was synthesized using solvothermal methods and a Teflon-lined stainless-steel autoclave reactor. The synthesis process was conducted as reported in the literature with slight modifications.³⁶ Briefly, 15 mL of DMF, 15 mL of ethanol, and 15 mL of deionized water were combined in a beaker. Subsequently, 8 mmol of BDC was added to the mixture, and the solution was sonicated for 15 minutes to achieve a homogeneous suspension. Next, 3.78 mmol of $\text{Co}(\text{NO}_3)_2 \cdot 6\text{H}_2\text{O}$ was added to the solution and sonicated for an additional 15 min until complete dissolution occurred. The resulting solution was transferred to a 50 mL Teflon-lined stainless-steel autoclave and heated at 150 °C for 24 hours. Afterward, the product was collected *via* centrifugation and washed five times with DMF and ethanol. The final product was dried in a vacuum oven overnight at 60 °C.

2.4. Synthesis of ZnO

The direct precipitation method was used to prepare ZnO NPs, as previously reported in the literature.^{8,37} In a nutshell, 0.2 M $\text{Zn}(\text{CH}_3\text{COOH})_2 \cdot 6\text{H}_2\text{O}$ aqueous solution and 0.4 M NaOH aqueous solution were each made up to 50 mL in volume. The mixture was stirred vigorously at room temperature until a white suspension was formed. After centrifuging at 5000 rpm for 20 minutes, the white product was washed three times with distilled water and once with 100% ethanol. The final product was then calcined for 3 hours at 500 °C.

2.5. Synthesis of ZnO@Co-BDC composite

The ZnO@Co-BDC composite catalyst was prepared by dissolving 8 mmol of BDC, 3.78 mmol of $\text{CoCl}_2 \cdot 6\text{H}_2\text{O}$, and 5 mmol of $\text{Zn}(\text{CH}_3\text{COO})_2 \cdot 2\text{H}_2\text{O}$ in a mixture of 50 mL DMF and 50 mL ethanol. The resulting solution was transferred to a Teflon-lined autoclave reactor and subjected to a solvothermal reaction at 120 °C for 2 h. After the reaction was completed, it was allowed to cool to room temperature, and the resulting product was collected. Then, the ZnO@Co-BDC composite was washed three times with DMF and ethanol to remove impurities and dried in a vacuum oven at 70 °C for 24 h. The final product was stored at room temperature for further applications.

2.6. Characterization

The XRD investigation was conducted using a Shimadzu X-ray diffractometer (XRD, Shimadzu XRD-7000) to characterize the crystallites of the synthesized composite catalyst. The analysis

of morphologies was conducted using transmission electron microscopy (TEM; Tecnai F20 G2). The functional groups of the MOF-based composite catalyst were analyzed using a PerkinElmer FT-IR spectrometer 65 FT-IR (PerkinElmer). The optical absorbance, band gap energy and MB dye degradation performance of the material were evaluated using DRS UV-Vis spectroscopy (Shimadzu-3600 Plus).

2.7. Photocatalytic experiment

The degradation of MB was conducted following a previously reported method.³⁸ Initially, the prepared suspensions were stirred in the dark for 30 minutes to reach the adsorption/desorption equilibrium. In brief, a 250 mL beaker was filled with an aqueous solution of 125 mL of a 10 µg per mL MB at pH 7. Subsequently, 25 mg of the ZnO@Co-BDC catalyst was added, and the beaker was positioned under a 150 W tungsten lamp located 10 cm above the cell. The solution was continuously stirred with a magnetic stirrer and cooled with water. A control sample without the catalyst was also irradiated to investigate the direct photodegradation of MB. Within a 20-min interval, 5 mL of the aliquot was taken and centrifuged at 5000 rpm. The absorbance of the resulting clear solution was measured with a UV-visible spectrophotometer at the wavelength of maximum absorbance (λ_{max}) of 663 nm. Both the initial absorbance (A_0) and the absorbance after irradiation (A_t) were recorded. Then, other factors like light and adsorption, pH, catalyst dose, and initial MB concentration effects on the degradation of MB were studied for the most efficient catalyst, ZnO@Co-BDC composite.

After determining the optimal conditions for the ZnO@Co-BDC composite, reusability was also evaluated. A mixture of 75 mg of ZnO@Co-BDC was incubated with 375 mL of 20 µg per mL MB solution at pH 9 and stirred for 30 minutes, then exposed to visible light for 80 minutes. After each cycle, a 5 mL aliquot was taken to measure the absorbance of MB (663 nm), and the catalyst was cleaned with deionized water and acetone. Then, the catalyst was dried and reused for a second cycle with a fresh solution containing 20 µg per mL MB as in the first cycle. For the other cycles, the same protocol was followed.

2.8. Quenching experiment

To gain deeper insights into the photocatalytic mechanism, additional experiments were performed using specific scavengers (isopropanol, *p*-benzoquinone (BQ), and disodium ethylenediamine-tetraacetate (EDTA)) to examine the roles of hydroxyl radicals ($\cdot\text{OH}$), superoxide radicals ($\cdot\text{O}_2^-$) and holes (h^+), respectively. For each experiment, 20 mL of a 10 mM solution of scavengers was added to 125 mL of a 20 µg per mL MB solution containing 75 mg of the ZnO@Co-BDC catalyst at a pH of 9. Then, the mixture was exposed to visible light for 80 minutes. Finally, the absorbance of the resulting solution was measured at the λ_{max} of MB (663 nm). These experiments aimed to assess the influence of the scavengers on the photocatalytic degradation of MB and elucidate the specific roles of electrons, $\cdot\text{OH}$, and holes in the process.³⁹



3. Result and discussion

3.1. Characterization

The crystallinity, phase analysis, and structural properties of the PET plastic waste-derived BDC, Co-BDC and ZnO@Co-BDC composite catalyst were confirmed using powder X-ray diffraction (XRD) analysis. The XRD patterns shown in Fig. 1a confirm that the peaks of the PET-plastic waste-derived BDC are located at $2\theta = 17.4^\circ$, 25.3° , and 28° , which are characteristic and consistent with those reported in the literature.⁴⁰ As shown in Fig. 1b, the Co-BDC MOF crystals formed from the coordination of cobalt metals and BDC ligands display distinct peaks at the 2θ values of 8.8° , 11.5° , 12.0° , and 15.9° , which correspond to the diffraction planes of (100), (110), (101), and (200), respectively, with JCPDS card no. 96-710-3395. This confirms the successful synthesis of the Co-BDC MOF.² As illustrated in Fig. 1c, the diffraction peaks of as prepared ZnO are (100), (002), (101), (102), (110), (103), (200), (112), (201), and (202) with JCPDS no. of 00-036-1451, have been accurately identified as belonging to the hexagonal wurtzite phase of ZnO.⁴¹ Furthermore, the XRD spectra of the ZnO@Co-BDC composites, presented in Fig. 1d, showed diffraction peaks for both ZnO and Co-BDC MOF. All peaks associated with the individual nanoparticles were evident in the composite pattern, confirming their feasibility. The morphologies of the as-prepared Co-BDC and ZnO@Co-BDC composites were analyzed using TEM. In the TEM images of Co-BDC, irregularly shaped, dark contrast particles are dispersed over a lighter background, indicating a relatively aggregated state (Fig. S1a†). The particles appear to be mostly amorphous, lacking clear lattice fringes, which is typical for many MOFs due to their hybrid organic-inorganic structures. The contrast variation suggests different thicknesses or possibly compositional heterogeneity, with denser regions likely corresponding to the cobalt-rich areas. The TEM image of the ZnO/Co-BDC composite shows a clearly defined rhombic or diamond-shaped nanoparticle, which contrasts sharply with the irregular and loosely aggregated particles seen in

the pure Co-BDC MOF (Fig. S1b†). The composite structure displays a dense, dark central region surrounded by a lighter halo, indicating a possible core–shell or embedded architecture, where the ZnO core is either encapsulated by or closely associated with the Co-BDC matrix.

To further characterize the structural properties and composition of the nanocomposites, Fourier-transform infrared (FTIR) spectroscopy was also conducted (Fig. 2a). In Fig. 2a(I), the FTIR spectra of BDC displayed the in-plane and out-of-plane bending vibrations for the aromatic ring at 734 and 664 cm^{-1} .⁴² The band at 1689 cm^{-1} is attributed to the C=O stretching vibration of carbonyl groups, indicating the formation of BDC. In Fig. 2a(II), the FTIR spectrum shows peaks at around 559 cm^{-1} , which confirms the formation of ZnO.⁴³ The FTIR spectrum of Co-BDC MOF (Fig. 2a(III)) shows a broad peak around 3400 cm^{-1} , assigned to the stretching vibrations of hydroxyl groups and water molecules.⁴⁴ The reaction of $1,4\text{-H}_2\text{BDC}$ with Co(II) leads to complete deprotonation, evident from the disappearance of the characteristic peaks at 1689 cm^{-1} (stretching vibrations of C=O). New bands are observed in the range of $1672\text{--}1558\text{ cm}^{-1}$, attributed to the asymmetric stretching vibrations of the carboxylate group.⁴⁵ These new bands indicate that Co(II) has successfully interacted with the $1,4\text{-H}_2\text{BDC}$ ligands. Similarly, the FTIR pattern of the ZnO@Co-BDC composites (Fig. 2a(IV)) showed peaks corresponding to both ZnO and Co-BDC MOF. The presence of all peaks associated with the individual nanoparticles in the composite pattern confirms their successful integration and viability.

The Raman spectra of the ZnO@Co-BDC composite were recorded in the range of $130\text{--}2300\text{ cm}^{-1}$, as illustrated in Fig. 2b, and provide valuable insights into its structural composition. The spectrum exhibits a prominent peak at approximately 1618 cm^{-1} , attributed to the C=C stretching mode of the BDC linker in the Co-BDC MOF, alongside smaller peaks at $\sim 1455\text{ cm}^{-1}$ (symmetric COO^- stretching) and $\sim 1000\text{--}1200\text{ cm}^{-1}$ (C-H bending), confirming the presence of the MOF structure.⁴⁶ For ZnO, a characteristic peak at $\sim 438\text{ cm}^{-1}$ corresponds to the E_2 (high) mode of its wurtzite structure, with an additional feature at $\sim 595\text{ cm}^{-1}$ indicative of the E_1 (longitudinal optical) mode, suggesting the presence of defects such as oxygen vacancies.^{47,48} These defects are beneficial for photocatalytic applications, enhancing charge separation and ROS generation. The relative intensity of the Co-BDC peaks suggests a dominant contribution from the MOF, while the ZnO peaks indicate its crystalline retention within the composite. Minimal peak shifts or new features suggest that the interaction between ZnO and Co-BDC is primarily physical, likely involving ZnO NPs embedded in the MOF framework, supporting their synergistic role in environmental remediation applications.

3.2. Optical properties of ZnO@Co-BDC composite catalysts

UV-Vis DRS were employed to assess the light absorption properties of ZnO, Co-BDC MOF, and ZnO@Co-BDC composite catalyst, with the results presented in Fig. 3. The data reveal that the absorption edges of ZnO and Co-BDC MOF are approximately 292 nm and 415 nm , respectively, while the absorption

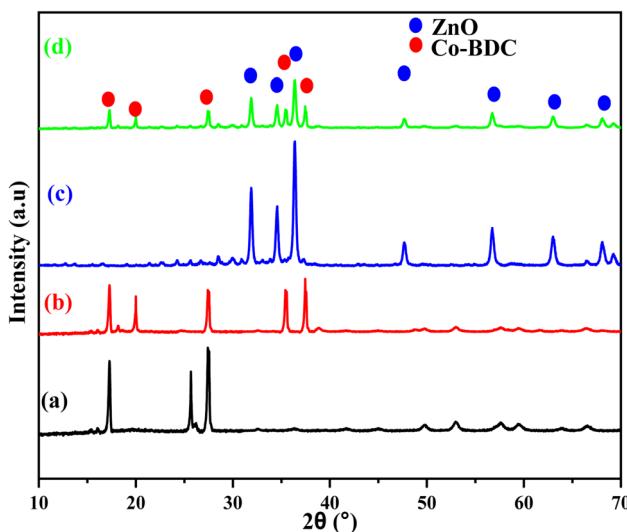


Fig. 1 XRD spectra of (a) BDC, (b) Co-BDC MOF (c) ZnO and (d) ZnO@Co-BDC.



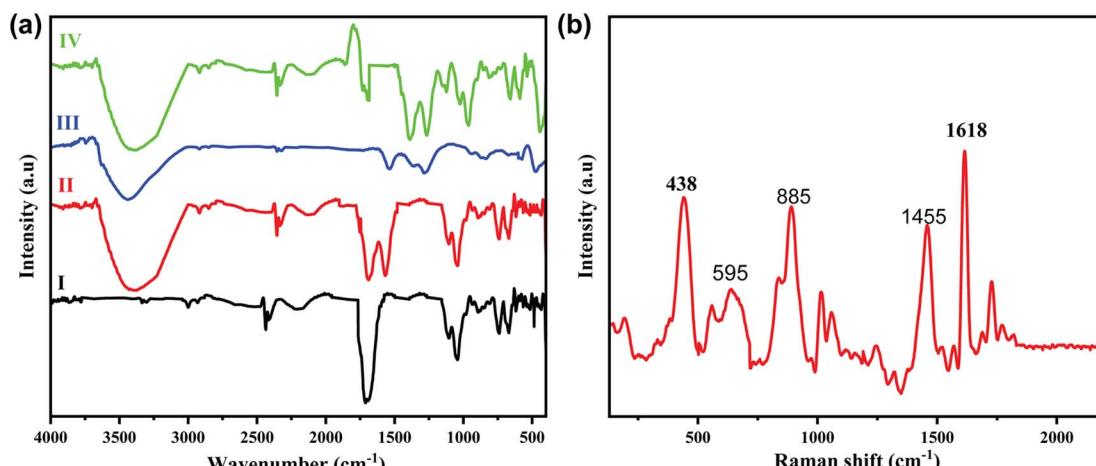


Fig. 2 (a) FT-IR spectra of (I) BDC, (II) ZnO, (III) Co-BDC MOF and (IV) ZnO@Co-BDC composite and (b) Raman spectra of ZnO@Co-BDC MOF.

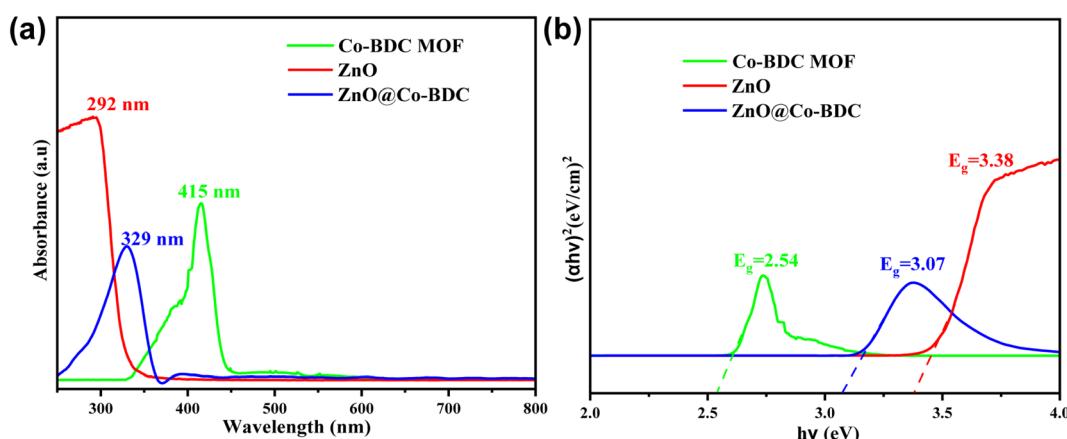


Fig. 3 (a) UV-Vis absorption spectrum of the as-prepared samples and (b) the $(\alpha h\nu)^2 - h\nu$ plot derived from the UV-Vis absorption spectrum.

edge of the ZnO@Co-BDC composite is around 329 nm (Fig. 3a). This indicates that ZnO nanoparticles demonstrate broad absorption in the UV region due to their wide bandgap, whereas pure Co-BDC MOF absorbs in the visible range. Consequently, the combination of ZnO and Co-BDC MOF effectively enhances light utilization and improves the dye degradation efficiency. The optical bandgap energy of the synthesized photocatalyst samples was determined using Tauc plots (Fig. 3b) by extrapolating the absorption edge, applying the following equation (eqn (1)) for direct allowed transitions.⁴⁹

$$(\alpha h\nu)^s = A(h\nu - E_g) \quad (1)$$

where α represents the absorption coefficient, h denotes Planck's constant, k is the absorption constant for a direct transition, and ν indicates the wave number.

3.3. Photocatalytic activities of photocatalysts

Recently, photocatalysts activated by visible light have proven to be effective candidates for the photodegradation of organic dyes in aqueous solutions.⁵⁰ The ZnO@Co-BDC composite catalyst demonstrates significantly higher photocatalytic efficiency

compared to ZnO and Cu-BDC MOF. The photocatalytic activity of the ZnO@Co-BDC composite was evaluated through the degradation of MB. To achieve this, the ZnO@Co-BDC composite photocatalyst was incubated with MB dye under visible light irradiation.

ZnO, Co-BDC MOF, and ZnO@Co-BDC composite successfully degraded MB, with its concentration declining over time. The degradation process occurred under visible light treatment for 80 min using a tungsten halogen lamp, with a 20 min interval between each exposure time. As compared to individual ZnO (Fig. 4a) and Co-BDC MOF (Fig. 4b), the ZnO@Co-BDC composite (Fig. 4c) showed the highest efficiency for the degradation of MB after 80 min under visible light treatments. In Fig. 4c, after 80 min, the MB solution incubated with the ZnO@Co-BDC composite catalyst exhibited the lowest absorbance as compared to MB solutions incubated with ZnO and Co-BDC MOF. The photographs in Fig. 4d depict the ZnO@Co-BDC composite before and after incubation in an MB solution. Upon irradiation, the blue color of the MB solution gradually disappears over time, aligning with the results observed in the absorbance spectra.

The percentage degradation (%D) of MB is calculated using eqn (2).



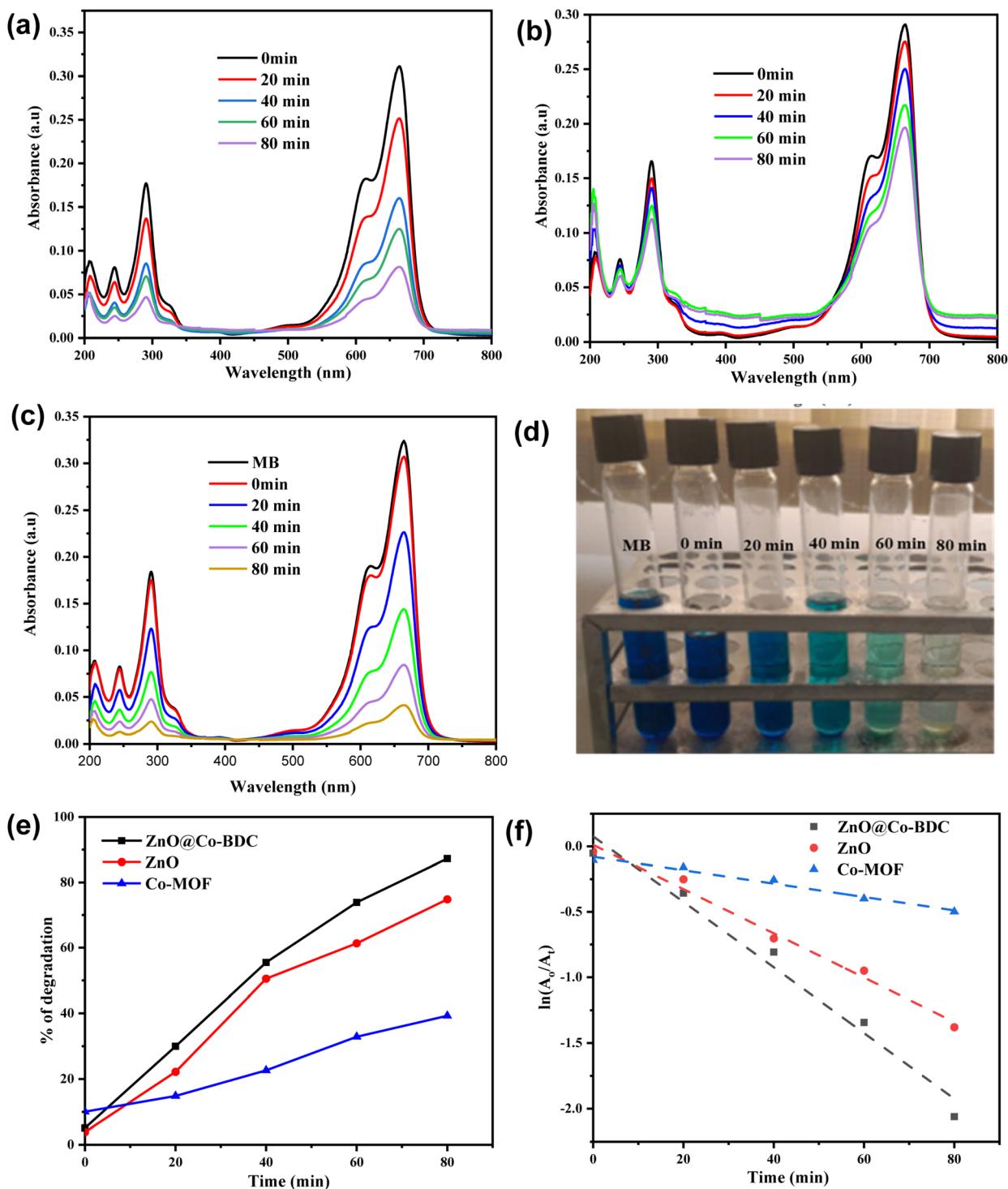


Fig. 4 UV-visible absorption spectra of MB at various irradiation times using (a) ZnO, (b) Co-BDC MOF, and (c) ZnO@Co-BDC composite. (d) Photographs of MB before and after incubation with the ZnO@Co-BDC composite at different time intervals. (e) Degradation efficiencies (%) of MB. (f) First-order kinetics of the photocatalyst under visible light irradiation.

$$\%D = \frac{A_0 - A_t}{A_0} \times 100\% \quad (2)$$

Here, A_0 is the absorbance at $t = 0$ min and A_t is the absorbance at t min.

As shown in Fig. 4e, the percentage degradation of MB using the ZnO@Co-BDC composite is higher (87.5%), as compared to

ZnO (74%) and Co-BDC (39%). ZnO shows lower degradation of MB under visible light because of its suboptimal utilization of visible light due to its wider band gap energy. In contrast, the Co-BDC MOF exhibits a significantly lower degradation rate of 39%, suggesting that it is less effective in photocatalytic applications on its own, likely due to inefficient charge separation.

Remarkably, the $\text{ZnO}@\text{Co-BDC}$ composite shows the highest degradation efficiency, indicating a synergistic effect that enhances photocatalytic activity. This improvement highlights the benefits of combining ZnO with Co-BDC MOF, as the composite takes advantage of the high surface area and porosity of the MOF to facilitate better dye absorption and efficient charge carrier separation. In addition, the moderate band gap (3.07 eV), which is low enough for the electrons to be excited with the low-energy visible light and high enough to reduce the

recombination rate, enhances the activity.⁵¹ Furthermore, the degradation efficiency of the $\text{ZnO}@\text{Co-BDC}$ composite is superior and similar to those of formerly reported photocatalysts, as indicated in Table S1.†

3.4. Kinetic analysis of MB dye degradation

To predict the degradation mechanism, a kinetics study is essential in wastewater treatment.⁵² In many studies, the photodegradation of pollutants follows a pseudo-first-order kinetic

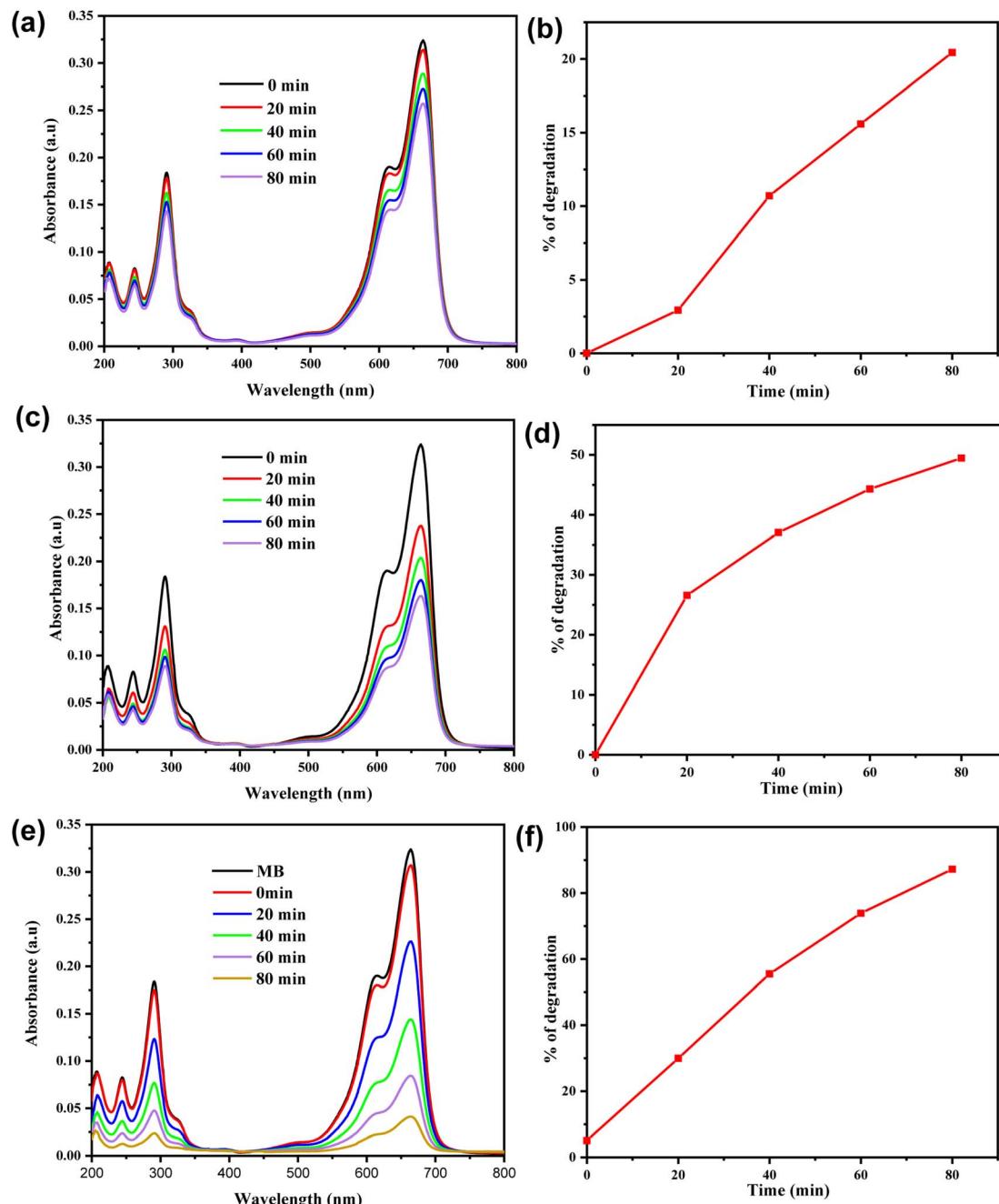


Fig. 5 (a) UV-Vis spectra of MB with light treatment and without any catalyst. (b) Degradation of MB (%) with light treatment and without any catalyst. (c) UV-Vis spectra of MB incubated with the $\text{ZnO}@\text{Co-BDC}$ composite catalyst under dark conditions. (d) Degradation of MB (%) incubated with the $\text{ZnO}@\text{Co-BDC}$ composite catalyst under dark conditions. (e) UV-Vis spectra of MB incubated with the $\text{ZnO}@\text{Co-BDC}$ composite catalyst under light irradiation. (f) Degradation of MB (%) incubated with the $\text{ZnO}@\text{Co-BDC}$ composite catalyst under light irradiation.



model. This model illustrates the relationship between the decomposition rate and time under particular reaction conditions.⁵³ Therefore, in this study, pseudo-first-order kinetics were used to assess the equilibrium data. The photodegradation experiments were conducted using the UV-Vis spectra by recording the absorbance of MB solutions at various exposure times (Fig. 4). The (A/A_0) values were measured to indicate the amount of MB molecules degraded using the maximum absorption spectra. Then, the kinetics of MB degradation was analyzed by a graph of $\ln(A/A_0)$ versus time, as shown in Fig. 4f. The rate of MB dye removal was confirmed using eqn (3).

$$\ln \frac{A_t}{A_0} = -k_t \quad (3)$$

A_0 and A_t represent the initial absorbance and absorbance at time (t), respectively, while k indicates the rate constant. As shown in Fig. 4f, the degradation of MB follows the pseudo-first-order kinetics.⁵⁴ The calculated rate constants for ZnO, Co-BDC MOF and ZnO@Co-BDC composites were 0.0169, 0.00512 and $2.501 \times 10^{-2} \text{ min}^{-1}$, respectively. The results indicated that the highest rate constant was observed with the ZnO@Co-BDC composite catalysts.

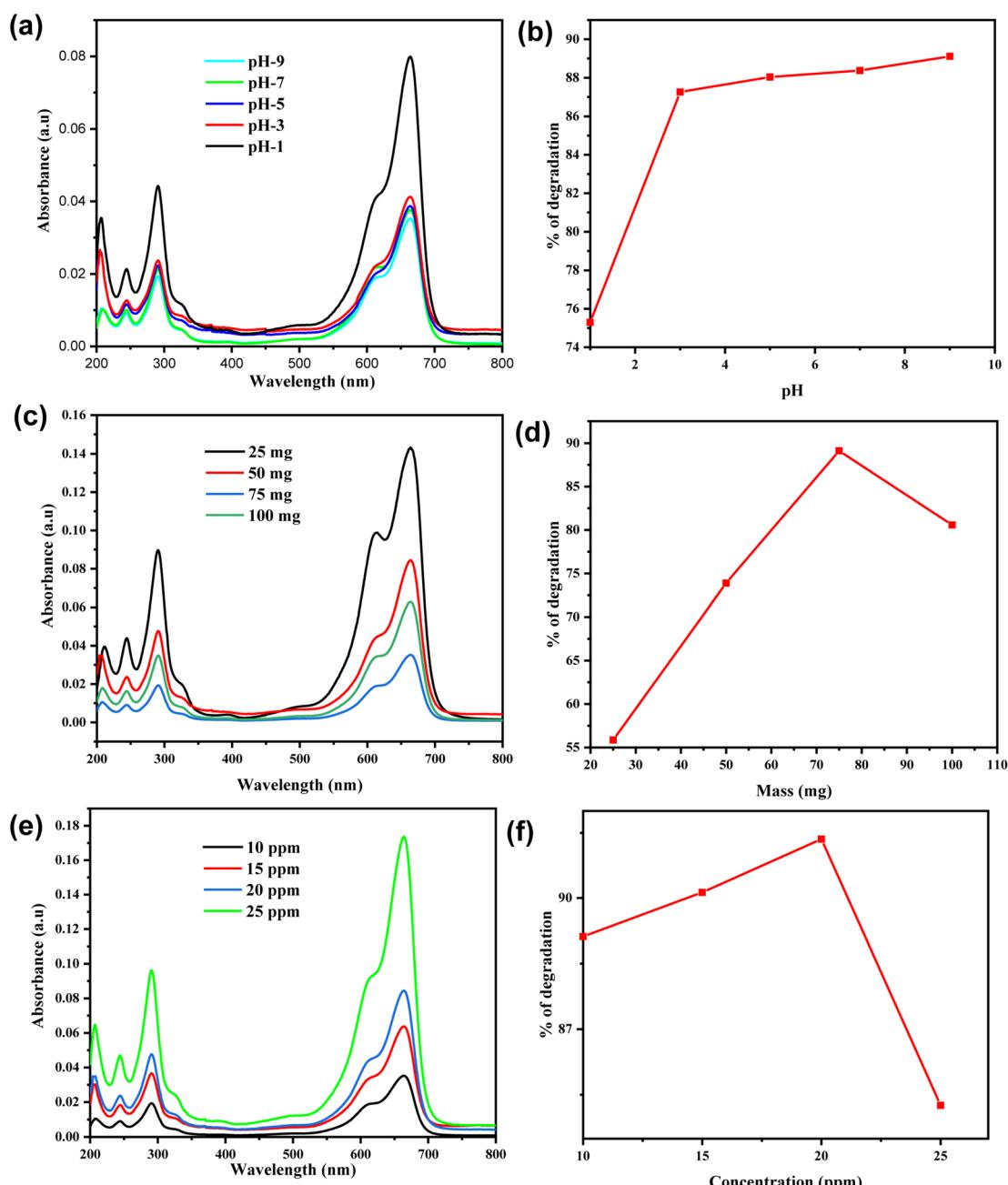


Fig. 6 (a) UV-Vis spectra for the degradation of MB at different pH values. (b) Degradation of MB (%) at different pH values. (c) UV-Vis spectra for the degradation of MB at different ZnO@Co-BDC doses. (d) Degradation of MB (%) at different ZnO@Co-BDC doses. (e) UV-Vis spectra for the degradation of MB (%) at different MB concentrations. (f) Degradation of MB (%) at different MB concentrations.



3.5. Effect of the operational parameters

3.5.1. Effect of light and adsorption. To evaluate the MB degradation efficiency of $\text{ZnO}@\text{Co-BDC}$, various conditions were tested: (1) exposing MB alone to visible light irradiation, (2) incubating MB with the $\text{ZnO}@\text{Co-BDC}$ composite in the dark, and (3) incubating MB with $\text{ZnO}@\text{Co-BDC}$ under visible light irradiation. The degradation percentage for exposing MB alone to visible light irradiation is only 20% (Fig. 5a and b), 49% for incubating MB with the $\text{ZnO}@\text{Co-BDC}$ composite in the dark (Fig. 5c and d), and MB degradation by incubating $\text{ZnO}@\text{Co-BDC}$ under visible light irradiation is 87.5% (Fig. 5e and f) after 80 min. The highest degradation of MB occurred with the $\text{ZnO}@\text{Co-BDC}$ catalyst under visible light treatment, followed by the dark condition and then light alone. The MB molecules adhered to the surface of the $\text{ZnO}@\text{Co-BDC}$ composite through chemical bonding, van der Waals forces, and electrostatic attraction, facilitating some dye removal.⁵⁵ Thus, the study demonstrated that the $\text{ZnO}@\text{Co-BDC}$ composite catalyst significantly enhanced the MB dye degradation process, even in the absence of light, while the degradation efficiency increased over time. Notably, light alone had minimal impact on the degradation of MB without the $\text{ZnO}@\text{Co-BDC}$ catalyst.

3.5.2. Effect of pH. The pH of a solution is a key factor in the photocatalytic degradation of wastewater pollutants. Changes in pH alter the charge on the catalyst surface, impacting the adsorption rate of dye molecules and consequently the degradation rate.⁵⁶ A study investigating the effect of pH within a range of 1 to 9 on the photocatalytic degradation of MB was conducted, with other conditions held constant (catalyst dose 75 mg, MB concentration $10 \mu\text{g mL}^{-1}$, and contact time 80 min) (Fig. 6a and b). The results shown in Fig. 6b indicate that the degradation of MB increased from 75% at pH 1 to 89% at pH 9, with the highest efficiency under basic conditions. In acidic environments, the catalyst surface becomes positively charged, repelling the cationic dye molecules and decreasing the adsorption and degradation efficiency. Conversely, in basic conditions, the negative charge on the catalyst enhances the electrostatic attraction with cationic dyes, improving adsorption and degradation.⁵⁷

3.5.3. Effect of catalyst dose. The impact of the $\text{ZnO}@\text{Co-BDC}$ composite catalyst dose was investigated using amounts ranging from 25 to 100 mg, with a 50 mL solution of $10 \mu\text{g}$ per mL MB at a pH of 9 over 80 minutes (Fig. 6c and d). The results indicate that the percentage of degradation increased from 25 to 75 mg and then decreased at 100 mg (Fig. 6d). The rise in catalyst dose led to an increased turbidity in the solution, which likely caused light scattering. This scattering reduces the amount of light absorbed by the catalyst, thereby limiting its photocatalytic degradation efficiency of MB.⁵⁸ Therefore, the optimal catalyst dose was found to be 75 mg, resulting in a 89% degradation efficiency of the MB solution.

3.5.4. Effect of initial MB concentration. The initial MB concentration in water influences how light interacts with the photocatalyst, which activates electron pairs and generates radicals. The effect of concentration was studied using an optimal dose of 75 mg, pH at 9 in 50 mL for 80 min (Fig. 6e and f). Concentrations were varied in $5 \mu\text{g mL}^{-1}$ increments (10, 15, 20,

and $25 \mu\text{g mL}^{-1}$). The result shows that the degradation efficiency increased from 89% at $10 \mu\text{g mL}^{-1}$ to 91% at $20 \mu\text{g mL}^{-1}$, but decreased at a higher concentration of $25 \mu\text{g mL}^{-1}$ to 85% (Fig. 6f); this is because an excessive MB concentration restricted light penetration, in turn limiting its interaction with the photocatalyst.⁵⁹ Additionally, maintaining a constant dose of 75 mg of catalyst produces a fixed amount of hydroxyl radicals, which may not be sufficient to effectively attack the increased number of MB molecules at higher concentrations.⁵⁷ Overall, the optimal conditions for achieving higher photocatalytic efficiency with the $\text{ZnO}@\text{Co-BDC}$ composite catalyst (91%) were evaluated to be pH 9, a catalyst dose of 75 mg, and an initial MB concentration of $20 \mu\text{g mL}^{-1}$ under visible light treatment for 80 min.

3.6. Stability and reusability

The stability of the $\text{ZnO}@\text{Co-BDC}$ composite photocatalyst was investigated by recycling the catalyst for six cycles. The photodegradation of MB was conducted at optimal conditions, and the photocatalyst was regenerated with ethanol, then centrifuged and dried at 80°C before being used in the next cycle. The results, as displayed in Fig. 7a, indicated that the photocatalyst could be easily regenerated with only a minimal change in intensity after being recycled for six cycles. During the recycling process, the photodegradation of the $\text{ZnO}@\text{Co-BDC}$ composite catalyst resulted in a MB degradation of 91%, 90.5%, 88.1%, 86.4%, 83.2%, and 80.5% (Fig. 7a), from the first cycle to the six cycles, respectively, indicating a slight reduction in the photocatalytic efficiency after six cycles. This slight decrease in activity was attributed to the inactivation of certain adsorption sites on the surface of the photocatalyst or a partial phase alteration of the dye solution activated by the photocatalyst in the presence of light. After completing each cycle (1st–6th) of photocatalytic degradation, the $\text{ZnO}@\text{Co-BDC}$ MOF was subjected to structural evaluation using XRD (Fig. S2†), revealing no noticeable changes in the phases of the composite catalyst at any stage, indicating its phase integrity throughout the process. Additionally, the XRD spectrum of Co-BDC MOF after the first cycle (Fig. S3†) showed no deviation from its original spectrum before degradation, further confirming the stability of the MOF component. To complement this, FTIR analysis of the $\text{ZnO}@\text{Co-BDC}$ composite was conducted after six cycles (Fig. S4†), verifying the retention of both Co-BDC MOF and ZnO phases, thus confirming the structural integrity of the composite. However, the XRD spectra of both $\text{ZnO}@\text{Co-BDC}$ revealed a decrease in peak intensity, suggesting that MB molecules were adsorbed onto the catalyst surface during the degradation process. Despite this adsorption, the unchanged peak positions and widths in the XRD patterns indicate that the crystallinity of the photocatalyst remained intact, demonstrating the robustness of the $\text{ZnO}@\text{Co-BDC}$ MOF under repeated photocatalytic cycles. These results indicate that it can be an effective photocatalyst for the photodegradation of MB and has recyclable properties.

3.7. Scavenger test

A mechanism for the photocatalytic degradation of MB using $\text{ZnO}@\text{Co-BDC}$ composite was explored by using various



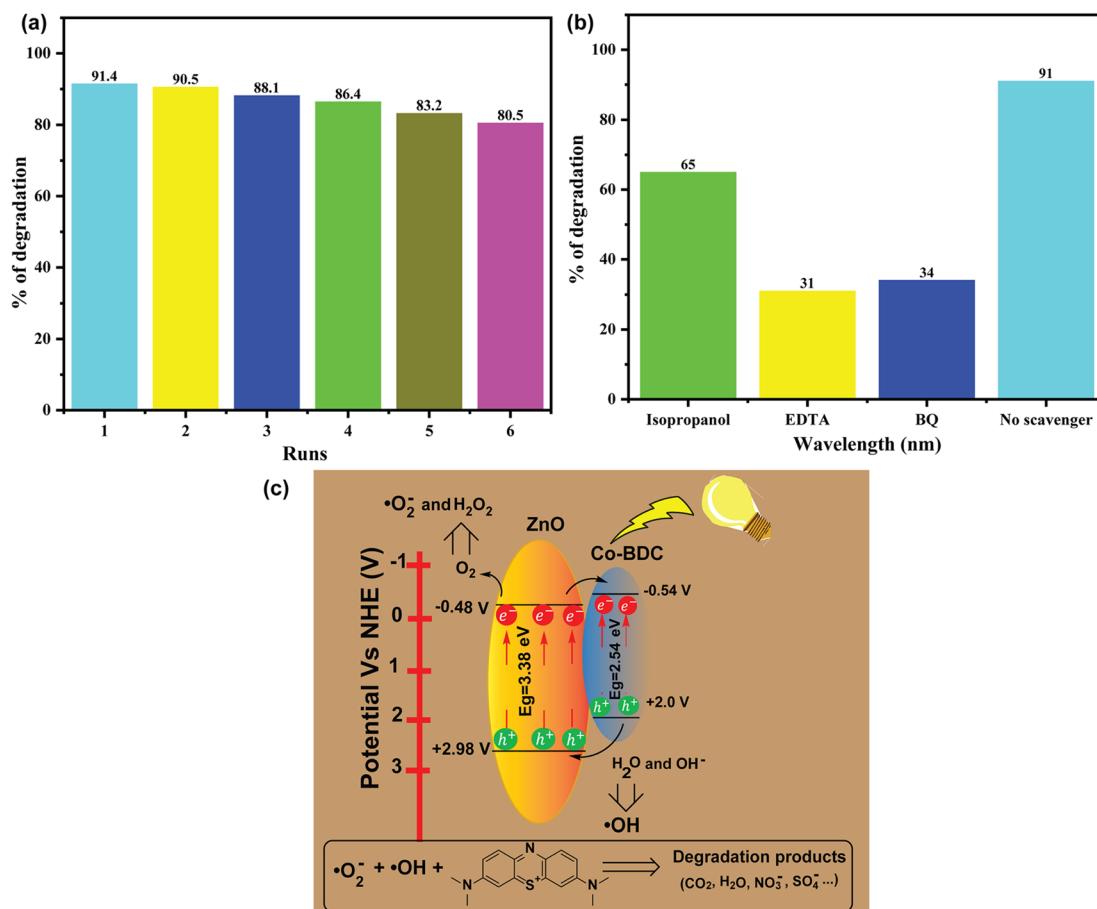


Fig. 7 Degradation percentage of MB (a) by reuse of ZnO@Co-BDC over six cycles (b) in the presence of different scavengers and (c) proposed photocatalytic degradation mechanisms of MB dye using the ZnO@Co-BDC composite catalyst.

scavengers to capture free radicals. Isopropanol was utilized as a scavenger for hydroxyl radicals (·OH), BQ for superoxide radicals (·O₂⁻), and EDTA for holes (h⁺).⁶⁰ In this study, different scavengers were used to clarify the role of the reactive oxygen species in the degradation of MB by trapping the free radicals. As shown in Fig. 7b, different scavengers, such as isopropanol, EDTA, and BQ, were used to assess the impact of the reactive oxygen species (·OH, h⁺ and ·O₂⁻) generated by the ZnO@Co-BDC composite catalyst in degrading MB under visible light. The results indicate that the photocatalytic efficiency of the ZnO@Co-BDC composite catalyst decreased significantly when EDTA-2Na was used, compared to the effects of BQ or isopropanol. The ZnO@Co-BDC catalyst removed 65%, 31%, 34%, and 91% of the MB dye using of isopropanol, EDTA, BQ, and without the catalyst, respectively (Fig. 7b). This suggests that positive holes (h⁺) and ·O₂⁻ radicals play a major role in breaking down the MB dye, while ·OH contributes less to this process.

3.8. Degradation mechanism

Fig. 7c presents a schematic representation of the mechanism for the photodegradation of MB using the ZnO@Co-BDC composite catalyst. It can be seen from the results that the

photocatalytic degradation ability of the ZnO@Co-BDC composite catalyst is higher than that of the Co-BDC MOF and pristine ZnO. On this basis, a mechanism for the photocatalytic degradation of MB by the ZnO@Co-BDC composite sample is proposed. The ZnO@Co-BDC composite sample, with a lower gap width than Co-BDC MOF, is more likely to accept lower photon energy intensity. The electrons in the valence band (VB) (e⁻) jump into the conduction band (CB) through the gap and form a photogenerated carrier hole (e⁻, h⁺).⁶¹ The photogenerated electrons on CB are transferred to VB on the material surface, e⁻ reacts with O₂ in the solution to generate ·O₂⁻, and ·O₂⁻ reacts with MB to generate intermediate substances and then generates small molecular substances. On the other hand, h⁺ on VB directly reacts with MB, and then catalyzes MB degradation to increase the capacity of the composite sample for photocatalysis.

The VB energy (E_{VB}) and CB energy (E_{CB}) of ZnO and Co-BDC MOF were determined using the absolute electronegativity (x) of each catalyst using eqn (4) and (5).⁶²

$$E_{VB} = x - E_e - 0.5E_g \quad (4)$$

$$E_{CB} = E_{VB} - E_g \quad (5)$$

E_e represents the energy of free electrons on the hydrogen scale (about 4.5 eV), and E_g refers to the energy bandgap of the semiconductor.^{63,64} The bandgap energy of the Co-BDC MOF catalyst is 2.54 eV, making it easily excitable to generate the electron–hole pairs, while ZnO has a bandgap of 3.38 eV, which makes it less easily excitable (Table S2†). When these two catalysts are combined, the bandgap reduces to 3.07 eV (Fig. 3b), forming a uniform lattice fringe. Upon irradiation with a tungsten halogen lamp, the CB electrons are transferred from the Co-BDC MOF to the ZnO, thereby inhibiting electron–hole recombination and enhancing photocatalytic performance.

4. Conclusion

In conclusion, we have reported a ZnO@Co-BDC composite prepared using the solvothermal method for MB degradation. The ZnO@Co-BDC composite exhibited significantly enhanced photocatalytic efficiency compared to pristine ZnO and Co-BDC MOF for the degradation of MB dye in aqueous solution under visible light irradiation. The incorporation of the Co-BDC MOF with ZnO nanoparticles altered the band structure of the composite, leading to improved photodegradation performance. The ZnO@Co-BDC composite achieved the highest degradation efficiency of MB, reaching 87.5% within 80 minutes under visible light exposure. In comparison, pristine ZnO and Co-BDC MOF demonstrated degradation efficiencies of 74% and 39%, respectively. Furthermore, the kinetic study also indicated that the degradation of MB followed first-order kinetics with a rate constant of $2.501 \times 10^{-2} \text{ min}^{-1}$. Factors that influence the degradation efficiencies, such as laser irradiation, catalyst dose, solution pH, and dye concentration, were also demonstrated. The stability and reusability of the ZnO@Co-BDC composite were evaluated and found to be effective for up to six cycles, which demonstrated better stability and reusability. Overall, the synthesized ZnO@Co-BDC composite catalyst has the potential to act as an alternative photocatalyst for MB degradation under visible light. Additionally, this MOF-based composite catalyst enabled the simultaneous valorization of plastic waste and remediation of environmental pollutants by converting waste to wealth.

Data availability

The datasets used in this study are accessible in the publication and can be obtained upon request from the corresponding author.

Conflicts of interest

The authors declare no conflict of interest.

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References

- W. Zhao, M. Adeel, P. Zhang, P. Zhou, L. Huang, Y. Zhao, M. A. Ahmad, N. Shakoor, B. Lou and Y. Jiang, *Environ. Sci.: Nano*, 2022, **9**, 61–80.
- M. F. Hanafi and N. Sapawe, *Mater. Today: Proc.*, 2020, **31**, A141–A150.
- K. Maheshwari, M. Agrawal and A. Gupta, in *Novel Materials for Dye-Containing Wastewater Treatment*, ed. A. K. Subramanian Senthilkannan Muthu, Springer, Singapore, 2021, pp. 1–25.
- R. Al-Tohamy, S. S. Ali, F. Li, K. M. Okasha, Y. A. Mahmoud, T. Elsamahy, H. Jiao, Y. Fu and J. Sun, *Ecotoxicol. Environ. Saf.*, 2022, **231**, 113160.
- W. M. Alamier, N. Hasan, I. S. Syed, A. M. Bakry, K. S. Ismail, G. Gedda and W. M. Girma, *Catalysts*, 2023, **13**, 1290.
- M. B. Tahir, T. Iqbal, M. Rafique, M. S. Rafique, T. Nawaz and M. Sagir, in *Nanotechnology and Photocatalysis for Environmental Applications*, ed. M. S. R. Muhammad Bilal Tahir and M. Rafique, Elsevier, 2020, pp. 65–76.
- R. Gusain, K. Gupta, P. Joshi and O. P. Khatri, *Adv. Colloid Interface Sci.*, 2019, **272**, 102009.
- B. A. Berehe, A. H. Assen, A. S. K. Kumar, H. Ulla, A. D. Duma, J. Y. Chang, G. Gedda and W. M. Girma, *Sci. Rep.*, 2023, **13**, 16454.
- F. Zhang, X. Wang, H. Liu, C. Liu, Y. Wan, Y. Long and Z. Cai, *Appl. Sci.*, 2019, **9**, 2489.
- D. Chen, Y. Cheng, N. Zhou, P. Chen, Y. Wang, K. Li, S. Huo, P. Cheng, P. Peng and R. Zhang, *J. Cleaner Prod.*, 2020, **268**, 121725.
- X. Liu, R. Ma, L. Zhuang, B. Hu, J. Chen, X. Liu and X. Wang, *Crit. Rev. Environ. Sci. Technol.*, 2021, **51**, 751–790.
- S. Sakthivel, B. Neppolian, M. Shankar, B. Arabindoo, M. Palanichamy and V. Murugesan, *Sol. Energy Mater. Sol. Cells*, 2003, **77**, 65–82.
- L. Yao, H. Yang, Z. Chen, M. Qiu, B. Hu and X. Wang, *Chemosphere*, 2021, **273**, 128576.
- C. B. Ong, L. Y. Ng and A. W. Mohammad, *Renewable Sustainable Energy Rev.*, 2018, **81**, 536–551.
- J. Z. Hassan, A. Raza, U. Qumar and G. Li, *Sustainable Mater. Technol.*, 2022, **33**, e00478.
- F. Sanakousar, C. Vidyasagar, V. Jiménez-Pérez and K. Prakash, *Mater. Sci. Semicond. Process.*, 2022, **140**, 106390.
- F. Wang, Y. Zhou, X. Pan, B. Lu, J. Huang and Z. Ye, *Phys. Chem. Chem. Phys.*, 2018, **20**, 6959–6969.
- S. J. Yang, J. H. Im, T. Kim, K. Lee and C. R. Park, *J. Hazard. Mater.*, 2011, **186**, 376–382.
- A. H. Chughtai, N. Ahmad, H. A. Younus, A. Laypkov and F. Verpoort, *Chem. Soc. Rev.*, 2015, **44**, 6804–6849.
- L. Jiao, J. Y. R. Seow, W. S. Skinner, Z. U. Wang and H.-L. Jiang, *Mater. Today*, 2019, **27**, 43–68.



21 J. Buasakun, P. Srilaong, K. Chainok, C. Raksakoon, R. Rattanakram and T. Duangthongyou, *Inorg. Chim. Acta*, 2020, **511**, 119839.

22 Z. Zhang, C. J. R. Wells, A. M. King, J. C. Bear, G. L. Davies and G. R. Williams, *J. Mater. Chem. B*, 2020, **8**, 7264–7274.

23 S. Chen, Y. Li and L. Mi, *Inorg. Chem. Commun.*, 2020, **118**, 107999.

24 C. V. Reddy, K. R. Reddy, V. a. Harish, J. Shim, M. Shankar, N. P. Shetti and T. M. Aminabhavi, *Int. J. Hydrogen Energy*, 2020, **45**, 7656–7679.

25 R. R. Ikreedeegh and M. Tahir, *J. CO₂ Util.*, 2021, **43**, 101381.

26 I. I. Alkhatab, C. Garlisi, M. Pagliaro, K. Al-Ali and G. Palmisano, *Catal. Today*, 2020, **340**, 209–224.

27 H.-P. Jing, C.-C. Wang, Y.-W. Zhang, P. Wang and R. Li, *RSC Adv.*, 2014, **4**, 54454–54462.

28 X. Qin, T. Qiang, L. Chen and S. Wang, *Microporous Mesoporous Mater.*, 2021, **315**, 110889.

29 D. Sun, L. Ye and Z. Li, *Appl. Catal., B*, 2015, **164**, 428–432.

30 Y. Gao, S. Li, Y. Li, L. Yao and H. Zhang, *Appl. Catal., B*, 2017, **202**, 165–174.

31 M. Wen, K. Mori, T. Kamegawa and H. Yamashita, *Chem. Commun.*, 2014, **50**, 11645–11648.

32 Y. Wu, X. Li, H. Zhao, F. Yao, J. Cao, Z. Chen, D. Wang and Q. Yang, *Chem. Eng. J.*, 2021, **426**, 131255.

33 D. Sun, Y. Fu, W. Liu, L. Ye, D. Wang, L. Yang, X. Fu and Z. Li, *Chem.-Eur. J.*, 2013, **19**, 14279–14285.

34 M. G. Kibria, N. I. Masuk, R. Safayet, H. Q. Nguyen and M. Mourshed, *Int. J. Environ. Res.*, 2023, **17**, 20.

35 R. Kaur, A. Marwaha, V. A. Chhabra, K. Kaushal, K.-H. Kim and S. Tripathi, *J. Cleaner Prod.*, 2020, **263**, 121492.

36 T. Sivam, N. Gowthaman, H. Lim, Y. Andou, P. Arul, E. Narayananamoothi and S. A. John, *Colloids Surf., A*, 2021, **622**, 126667.

37 S. Bagheri, K. Chandrappa and S. Hamid, *Pharma Chem.*, 2013, **5**, 265–270.

38 O. A. Zelekew, P. A. Fufa, F. K. Sabir and A. D. Duma, *Heliyon*, 2021, **7**, e07652.

39 A. Dehdar, G. Asgari, M. Leili, T. Madrakian and A. Seid-Mohammadi, *J. Environ. Manage.*, 2021, **297**, 113338.

40 Y. Wang, H. Wang, S. Li and S. Sun, *ACS Omega*, 2022, **7**, 35180–35190.

41 A. Norouzi and A. Nezamzadeh-Ejhieh, *Chem. Phys. Lett.*, 2020, **752**, 137587.

42 Y. Rao, G. Inwati and M. Singh, *Future Sci. OA*, 2017, FSO239.

43 D. D. Thongam, J. Gupta and N. K. Sahu, *Appl. Sci.*, 2019, **1**, 1030.

44 R. Yuniasari, F. Amri, S. A. Abrori, N. L. W. Septiani, M. Rezki, M. Fahmi and B. Yuliarto, *IOP Conf. Ser.: Mater. Sci. Eng.*, 2021, **1045**, 012010.

45 X. Hu, H. Hu, C. Li, T. Li, X. Lou, Q. Chen and B. Hu, *J. Solid State Chem.*, 2016, **242**, 71–76.

46 S. Ramalingam and S. Periandy, *Spectrochim. Acta, Part A*, 2011, **78**, 1149–1161.

47 L. Shi, *Raman Spectroscopy Study of Wurtzite Crystals: Gallium Nitride and Zinc Oxide*, Arizona State University, 2003.

48 M. Šćepanović, M. Grujić-Brojčin, K. Vojisavljević, S. Bernik and T. Srećković, *J. Raman Spectrosc.*, 2010, **41**, 914–921.

49 M. Fiaz, M. Kashif, S. Majeed, M. N. Ashiq, M. A. Farid and M. Athar, *ChemistrySelect*, 2019, **4**, 6996–7002.

50 S. Li, Z. Wang, X. Zhao, X. Yang, G. Liang and X. Xie, *Chem. Eng. J.*, 2019, **360**, 600–611.

51 E. Amdeha and R. S. Mohamed, *Environ. Technol.*, 2021, **42**, 842–859.

52 H. Yu, E. Nie, J. Xu, S. Yan, W. J. Cooper and W. Song, *Water Res.*, 2013, **47**, 1909–1918.

53 R. Ezzati, *Chem. Eng. J.*, 2020, **392**, 123705.

54 A. K. Sajjad, S. Shamaila, B. Tian, F. Chen and J. Zhang, *J. Hazard. Mater.*, 2010, **177**, 781–791.

55 K. G. Akpomie, J. Conradi, K. A. Adegoke, K. O. Oyedotun, J. O. Ighalo, J. F. Amaku, C. Olisah, A. O. Adeola and K. O. Iwuoazor, *Appl. Water Sci.*, 2023, **13**, 20.

56 K. M. Reza, A. Kurny and F. Gulshan, *Appl. Water Sci.*, 2017, **7**, 1569–1578.

57 T. B. Adamu, A. M. Mengesha, M. A. Kebede, B. L. Bogale and T. W. Kassa, *Results Chem.*, 2024, **10**, 101724.

58 I. Groeneveld, M. Kanelli, F. Ariese and M. R. van Bommel, *Dyes Pigm.*, 2023, **210**, 110999.

59 A. Ajmal, I. Majeed, R. N. Malik, H. Idriss and M. A. Nadeem, *RSC Adv.*, 2014, **4**, 37003–37026.

60 R. Guo, Y. Wang, J. Li, X. Cheng and D. D. Dionysiou, *Appl. Catal., B*, 2020, **278**, 119297.

61 W. Sun, S. Meng, S. Zhang, X. Zheng, X. Ye, X. Fu and S. Chen, *J. Phys. Chem. C*, 2018, **122**, 15409–15420.

62 A. Naseri, M. Samadi, N. M. Mahmoodi, A. Pourjavadi, H. Mehdipour and A. Moshfegh, *J. Phys. Chem.*, 2017, **121**, 3327–3338.

63 Z. Yu, H. Moussa, M. Liu, R. Schneider, M. Moliere and H. Liao, *Surf. Coat. Technol.*, 2019, **375**, 670–680.

64 N. K. Sompalli, Y. Li, J. Li and S. Kuppusamy, *Environ. Res.*, 2024, **259**, 119532.

