



Cite this: *RSC Sustainability*, 2024, **2**, 3036

Validating the “greenness” of chemicals *via* life cycle assessment: the case of anisole as an anti-solvent in perovskite solar cells†

A. Kamal Kamali ^{*a} Nilanka M. Keppetipola, ^{‡b} Yuka Yoshihara, ^c Ajay Kumar Jena, ^{¶d} Satoshi Uchida, ^{¶e} Hiroshi Segawa, ^f Guido Sonnemann, ^{¶a} Thierry Toupane ^{¶b} and Ludmila Cojocaru ^{¶b}

Technological development is increasingly driven by environmental sustainability, with labels like ‘green’ gaining traction. However, the complex interactions in a product’s life cycle make the environmental impact of materials and chemicals highly context-dependent, highlighting the need for context-specific environmental assessments. Anisole has been popularly used as a “green” alternative to chlorobenzene for perovskite solar cell (PSC) fabrication. This work validates the technical and environmental performance of anisole in the fabrication of PSCs. PSCs exhibiting conversion efficiencies exceeding 22% can be attained by using anisole as an antisolvent. Moreover, around 50% reduction in the potential toxicity is obtained when substituting chlorobenzene for anisole embodied in the carcinogenic human and ecosystem toxicity impact categories. Nonetheless, anisole embodies higher impact in all 14 remaining impact categories. This is due to anisole’s multistep synthesis procedure that contributes to more than double the climate change impact of chlorobenzene, synthesized by a single-step method. To reduce the emissions several recommendations and strategies are proposed. Ultimately, it has been proved that context-specific and transparent environmental assessments are needed to make informed decisions in research and development leading to environmentally sustainable solutions.

Received 5th July 2024
Accepted 9th September 2024

DOI: 10.1039/d4su00361f
rsc.li/rscsus

Sustainability spotlight

Perovskite solar cells have rapidly emerged as promising alternatives to other established solar cell technologies and as they are currently at the trial stage of commercial development it is essential to verify and propose sustainable practices that ensure their widespread adoption without compromising the environment. In the antisolvent treatment of the perovskite layer, anisole has been proposed as a green alternative to chlorobenzene. Efficiencies over 22% and halved carcinogenic toxicity have been achieved by substituting chlorobenzene for anisole according to life cycle assessment. However, other environmental impacts have doubled, such as climate change impact. This highlights the need for transparent, context-specific environmental assessments. This work contributes to access to clean energy (SDG 7), industry innovation (SDG 9), and climate action (SDG 13) by contributing to the development of environmentally sustainable commercializable solar cell technology. Ultimately, anisole use in antisolvent treatment is one of many optimizations needed in PSC development.

Introduction

Terminologies such as “Green,” “Eco-efficient,” and “Environmentally friendly” are commonly employed to convey that

certain materials, products, or processes are environmentally sustainable. However, the results of environmental assessments are typically comparative and dependent on the conditions and assumptions specified,¹ rendering their outcomes

^aUniv. Bordeaux, CNRS, Bordeaux INP, ISM, UMR 5255, Groupe Analyse du Cycle de Vie et Chimie Durable (CyVi), 351 Cours de la Libération, F-33405 Talence, France. E-mail: ahmad-kamal.kamali@u-bordeaux.fr; guido.sonnemann@u-bordeaux.fr

^bUniv. Bordeaux, CNRS, Bordeaux INP, ISM, UMR 5255, Groupe Chimie Moléculaire et Matériaux (C2M), 351 Cours de la Libération, F-33405 Talence, France. E-mail: nilanka.keppetipola@college-de-france.fr; cojocaru@g.ecc.u-tokyo.ac.jp; thierry.toupane@u-bordeaux.fr

^cThe University of Tokyo, Department of Chemical System Engineering, Graduate School of Engineering, 7-3-1, Hongo, Bunkyo-ku, Tokyo 113-8656, Japan. E-mail: yoshihara_y@chemsys.t.u-tokyo.ac.jp

^dThe University of Tokyo, Komaba Institute for Science, Graduate School of Arts and Sciences, 4-6-1, Komaba, Meguro, Tokyo 153-8904, Japan. E-mail: ajayjena@g.ecc.u-tokyo.ac.jp; cojocaru@g.ecc.u-tokyo.ac.jp

[¶]The University of Tokyo, Research Center for Advanced Science and Technology, 4-6-1, Komaba, Meguro, Tokyo 153-8904, Japan. E-mail: uchida@rcast.u-tokyo.ac.jp

^fThe University of Tokyo, Department of General Systems Studies, Graduate School of Arts and Sciences, 3-8-1, Komaba, Meguro-ku, Tokyo 153-8902, Japan. E-mail: csegawa@mail.ecc.u-tokyo.ac.jp

† Electronic supplementary information (ESI) available. See DOI: <https://doi.org/10.1039/d4su00361f>

‡ Current position: Chimie du solide et de l’énergie (CSE), UMR 8260, Collège de France, 05, Paris 75231 CEDEX, France.



predominantly relative rather than absolute. Hence, the utilization of a “green” material or adherence to a “green” process does not automatically confer “green” status upon a product; rather, it may simply reduce its environmental impact compared to an alternative. Researchers across various fields prioritize the integration of “green” materials into their technologies to achieve environmental sustainability. This is evident in the clean energy generation sector, particularly in developing perovskite solar cells (PSCs).²⁻⁴ Recommendations made by solvent selection guides, namely CHEM21 and GlaxoSmithKline (GSK),^{5,6} are usually applied. These guides score solvents based on environmental, health, and safety criteria, including physical properties. Solvents are assigned scores, and those with the highest scores are color-coded green. Unfortunately, this color coding can be misinterpreted. Chemicals marked as green are sometimes mistakenly assumed to be environmentally sustainable. This misconception arises because the scoring system considers multiple criteria, with the environmental aspect being just one of them. A chemical could have excellent health and safety performance but poor environmental performance and still receive a high score. Moreover, reducing the environmental aspect to a single score hampers a thorough understanding of what the score actually represents. In other words, the green color-coding of these solvents does not inherently ensure a reduced environmental impact of the resulting product. The actual environmental impact depends on various factors, including the solvent’s life cycle, and the performance of the produced PSCs, among other aspects.

Researchers worldwide are working on developing photovoltaic cells exhibiting high power conversion efficiency (PCE) and low environmental footprint. Generally, PCE is defined as the ratio of energy output from solar cells (P_{\max}) to the energy input (P_{in}) of sunlight and is defined as:

$$\text{PCE} = \frac{P_{\max}}{P_{\text{in}}} \times 100\% = \frac{J_{\text{sc}} \times V_{\text{oc}} \times \text{FF}}{P_{\text{in}}} \times 100\%$$

where, P_{\max} is the maximum power output of the solar cell, J_{sc} is the short-circuit current density, V_{oc} is the open-circuit voltage, FF is the fill factor, and P_{in} is the input power (100 mW cm^{-2} in standard test conditions). Photovoltaic parameters of solar cells, defined as J_{sc} , V_{oc} , and FF are obtained from measuring $I-V$ curves.

PSCs have emerged as promising alternatives to other established solar cell technologies.⁷ The efficiency of PSCs has now reached 26.1% in single junction and 33.9% in tandem configuration with Si-solar cells.⁸ A major contribution to such a fast rise in PCE was the development of methods producing high-quality perovskite films exhibiting a pinholes-free and flat surface.⁹⁻¹² As the morphology of the perovskite film influences the performance of devices significantly a lot of efforts during the initial days have been devoted to controlling the crystallization of perovskite.¹³ Several techniques such as sequential deposition,¹⁴ antisolvent methods,¹⁵ solvent-assisted crystallization,¹⁶ additives included in the precursors, and co-evaporation,¹⁷ have been developed to enhance the quality of the perovskite film. The community’s dedication to advancing the environmental sustainability of PSCs is apparent through

environmental assessment studies published.^{18,19} These studies evaluate the environmental impact of PSCs across diverse fabrication methods and configurations, while also considering variations in PCEs. Furthermore, decent efficiencies have been achieved by integrating a variety of “green” solvents into the fabrication process of solar cells.²⁰⁻³⁰

The solvent-engineering method employing antisolvent treatment is one of the most popular and commonly used perovskite deposition methods,^{9,31} leads to high efficiencies but faces environmental challenges. A one-step solvent-assisted deposition approach for producing uniform $\text{CH}_3\text{NH}_3\text{PbI}_3$ thin films through rapid crystallization was developed.¹⁵ In this process, chlorobenzene, the antisolvent employed, is promptly introduced after applying a perovskite precursor solution in dimethylformamide (DMF) during spin coating. The role of chlorobenzene is to rapidly reduce the solubility of perovskite in mixed solvent and promote fast nucleation of the crystals in the film. During this antisolvent treatment, the density of nuclei increases resulting in a uniform and pinhole-free film with large perovskite grains up to a size of microns.³² The antisolvent step significantly improves solar cell efficiency, reduces hysteresis, and enhances the stability of the devices. However, chlorobenzene, widely used as an antisolvent, is known to be a toxic substance with suppressive and anesthetic impacts on the human central nervous system.³³ Owing to its high toxicity, the use of chlorobenzene constitutes a major concern²⁸ for industrial-scale manufacturing. Thus, alternative solvents have been explored to not only reduce toxicity but also achieve high solar cell efficiency. For example, claimed to be “green” antisolvents such as ethanol,²⁵ tetraethyl orthocarbonate,³⁴ anisole,^{24,35} etc., have been studied and applied as antisolvents to obtain high-efficiency PSCs. Nonetheless, the actual “greenness” of these solvents has yet to be verified using standardized environmental assessment methods.

In this work, we employ the solvent-engineering technique to fabricate PSCs, comparing the use of chlorobenzene and anisole as antisolvents. Our investigation delves into the technical and performance aspects of the fabrication of PSCs, focusing on the morphology, crystallography, and surface composition of the perovskite film. Additionally, we apply the life cycle assessment (LCA) methodology to uncover the environmental implications of substituting chlorobenzene for anisole during the antisolvent treatment phase of PSC production. In this context, various characterization techniques are employed to ensure that PSCs fabricated using antisolvents have the same characteristics and offer identical performance of over 22% with excellent reproducibility. The conducted experiments and detailed characterization are paramount for a valid LCA comparison, as they demonstrate that the PSCs fabricated using the two antisolvents are identical in structure and performance. Our objective is to ascertain whether adopting anisole compared to alternative antisolvents can reduce the environmental footprint attributed to the antisolvent treatment step. In accordance with the study objective, the LCA boundary focuses solely on the chemicals used in the antisolvent treatments, *i.e.*, anisole and chlorobenzene. It does not assess any components of the PSC, as these components have been proven to be identical in both



performance and composition, regardless of whether anisole or chlorobenzene is used in the antisolvent treatment step.

This case study shows the importance of conducting lab-scale LCAs to validate the potential positive impact associated with the adoption of “green” materials and processes. Our motivation stems from recognizing that the environmental footprint of materials can vary depending on their utilization and end-of-life management. Hence, while the integration of “green” materials holds promise for reducing environmental impact in emerging technologies, it does not inherently guarantee such an outcome. Moreover, several ideas are explored and discussed such as material efficiency, bio-based synthesis routes, and solvent recovery demonstrating the relative “greenness” of solvents.

Experimental

In this section, the employed materials and followed processes for the fabrication of PSCs are detailed. Next, the LCA, a leading method in environmental assessment, is introduced. LCA modelling parameters are transparently communicated along with the built life cycle inventories to facilitate the results’ replication.

Materials

Fluorine-doped tin oxide ($\text{SnO}_2\text{:F}$, FTO) glass substrates were purchased from Nippon Sheet Glass Co. Ltd (NSG), and VisionTek Systems Ltd. Acetone, ethanol, and propan-2-ol were provided by Fisher. SnO_2 (15 wt%) in H_2O colloidal dispersion was bought from Alfa Aesar while PbI_2 (99.99%) was obtained from Kojundo Chemical Laboratory Co. LTD. PbI_2 , FAI, and KI were obtained from Tokyo Chemical Industry (TCI), CsI from Sigma-Aldrich or TCI, *N,N*-dimethylformamide (DMF) and dimethyl sulfoxide (DMSO) from TCI. Titanium di-isopropoxide bis (acetylacetone) (75 wt% in isopropanol), lithium bis (trifluoro methane sulfonyl) imide (LiTFSI), and 4-*tert*-butylpyridine (*t*BP) were purchased from Sigma-Aldrich. TiO_2 (24 nm) was purchased from JGC C&C and spiro-OMeTAD ($(2,2',7,7'\text{-tetrakis}-(N,N\text{-di-4-methoxyphenylamino)-9,9'\text{-spiro-bifluorene})$) from Merck. Anisole and chlorobenzene were obtained from Wako Pure Chemical Industries, LTD. All chemicals were used without any further purification.

Fabrication of PSCs

Compact TiO_2 , mesoporous- TiO_2 , spiro-OMeTAD, and Au layers were prepared following a method earlier reported by Mariotti *et al.*³⁶ A precursor solution of perovskite of composition $\text{K}_{0.025}(\text{Cs}_{0.1}\text{FA}_{0.9})_{0.975}\text{PbI}_3$ was made by dissolving 54.6 mg of CsI, 325 mg of FAI, and 1016.6 mg of PbI_2 in 1.2 ml of DMF and 300 μl of DMSO. Then, 35 μl of KI (49.8 mg KI in 200 μl DMSO) was added to it. After that, 100 μl of this solution was spin-coated at 1000 r.p.m. for 10 seconds and at 3000 r.p.m. for 30 seconds. During the spin coating, 500 μl of chlorobenzene or anisole was dropped during the second spin coating step, 8 seconds before completing the spinning, which was followed by annealing the substrates at 155 °C for 15 min. According to the

industrial solvent selection guide, anisole is categorized as a minimal health hazardous material than chlorobenzene.³⁷

Material characterization and solar cells evaluation

The scanning electron microscopy (SEM) images of the perovskite layer surface and perovskite solar cells were recorded using a field emission scanning microscope (FE-SEM JSM6700F, Jeol 6700F). The *I*-*V* characteristics and photovoltaic parameters were measured under AM1.5G conditions at 100 mW cm^{-2} using a 300 W Xenon solar simulator (YSS-80A; Yamashita Denso Co., Ltd, Japan). The light source was calibrated using a Si photodiode of BS-520 (Bunkokeiki Co., Ltd, Japan). The active area of solar cells was 0.18 cm^2 . The incident photon to electron conversion efficiency (IPCE) was measured using Bunkokeiki, CEP-2000MLQ Instruments system.

Life cycle assessment (LCA)

LCA is a standardized method frequently used for evaluating the environmental performances of products, processes, or services over their full life cycle. This LCA follows the framework of ISO standards (14040 and 14044).^{38,39} LCA consists of four stages: goal and scope definition, life cycle inventory (LCI) analysis, life cycle impact assessment (LCIA), and interpretation. In this section, the first three stages are described, while the LCIA results are presented and interpreted in the Results and discussion section.

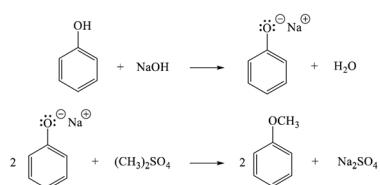
The goal of this LCA was to evaluate the environmental performance of anisole and compare it to chlorobenzene, thereby verifying whether the substitution of chlorobenzene for anisole can reduce the environmental impact of the antisolvent treatment step. This analysis focuses on the use of these two chemicals as antisolvents in the laboratory production process of PSCs. The functional unit selected is carrying an antisolvent treatment on the perovskite layer, resulting in a PSC with a PCE of 22%. For this goal, chlorobenzene and anisole were compared on a 1-to-1 basis, given that equal volumes of these two solvents are used to obtain identical results. The absolute environmental results correspond to 1 kg of chemicals. This mass-based functional unit is justified because both chemicals are used in equal quantities and produce PSCs of similar characteristics and efficiencies.

The scope of this LCA encompasses the entire life cycle of the antisolvents (anisole and chlorobenzene) noted as a cradle-to-grave study. It considers impacts arising from material extraction, chemical synthesis, transportation, use, and end-of-life stages. Notably, there are no impacts associated with the use phase, as the antisolvents evaporate during the thermal treatment of the perovskite layer. It is assumed that the entire amount of anisole or chlorobenzene is evaporated into the atmosphere. The LCA boundary focuses solely on the chemicals used in the antisolvent treatments. It does not assess any components of the PSC, as these components have been proven to be identical in both performance and composition, regardless of whether anisole or chlorobenzene is used in the antisolvent treatment step. The geographical scope of the study has been set to solvent production outside of Europe, and their use



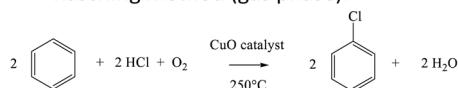
in southern France. Additionally, this is a contemporary study, where background and foreground data have been sourced from the latest available version of the ecoinvent database at the time of conducting the research. Life cycle inventories have been developed for anisole's most common production route, which relies on petrochemicals.⁴⁰ The foreground subsystem representing anisole^{40,41} and chlorobenzene production has been constructed based on the following chemical reactions.

(a) Synthesis of Anisole from Phenol:



(b) Synthesis of Chlorobenzene, industrial method:

Rasching method (gas phase):



The synthetic reactions of phenol, used as a precursor for anisole, is available in ESI† along with the LCI in Table S1.† The synthesis route has been optimized and scaled up to emulate industrial production conditions, by assuming that 80% of benzene is recycled and calculating the heating and distillation energy necessary during anisole production based on engineering equations.⁴² The energy calculation has been carried out according to eqn (S1) and (S2)† and according to information disclosed in Tables S2–S4.† The background subsystem including raw material acquisition and chlorobenzene production has been modelled based on generic datasets. An LCI has been also built to account for the impacts arising during the chlorobenzene life cycle (Table S5†).

The life cycle impact assessment method used to conduct this analysis is the environmental footprint method version 3.1 (adapted).⁴³ This LCIA method has been selected due to its wide coverage of chemical toxicity embodied in three impact categories namely, human carcinogenic and non-carcinogenic toxicity as well as ecotoxicity.⁴⁴ The ecoinvent version 3.9.1 APOS was used to model background inventory processes.⁴⁵ Modeling was carried out using SimaPro 9.5 software.

Results and discussion

Processing, morphology, and PSCs performances, perovskite film prepared using chlorobenzene and anisole as antisolvents

To prepare PSCs, the K_{0.025}(Cs_{0.1}FA_{0.9})_{0.975}PbI₃ composition was selected and spin-coated on the SnO₂ layer (electron transport layer) followed by a quick application of the antisolvents, chlorobenzene or anisole at a fixed time of 8 seconds before the end of the spin-coating procedure (total time of 30 seconds). In both cases, a volume of 500 μL was dripped quickly through a micropipette tip. The substrates coated with perovskite were annealed at 155 °C for 15 min followed by the coating of spiro-

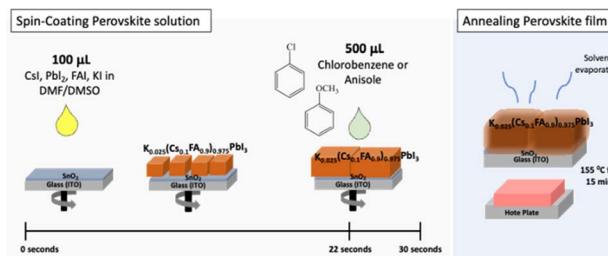


Fig. 1 Schematic representation of the processing method of K_{0.025}(Cs_{0.1}FA_{0.9})_{0.975}PbI₃ layer using chlorobenzene and anisole as antisolvents.

OMeTAD and Au layers³⁶ to complete the devices. The fabrication process of perovskite film is depicted in Fig. 1.

As shown in Fig. 2 the surface morphology of the untreated films (without antisolvent, Fig. 2a and d) shows uneven surfaces with aggregated particles and poor coverage while antisolvent-treated chlorobenzene (Fig. 2b and e) and anisole (Fig. 2c and f) films show homogeneous smooth surfaces with well-defined perovskite grain. It has been established that the antisolvent nature governs the perovskite crystallization, and thus the grain size and surface roughness of the perovskite films.⁴⁶ A similar trend was observed in our experiments.

According to SEM images, the average grain size was estimated to be 574 ± 131 nm for chlorobenzene-treated film while that for the anisole-treated film was found to be 649 ± 177 nm. Larger grains in the case of anisole can be ascribed to its higher boiling point, *i.e.* 153.8 °C, and higher viscosity, leading to slower crystallization resulting in larger grains. It has been also reported that anisole forms intermolecular interaction with DMF and DMSO *via* hydrogen bonding that can play a role in the formation of perovskite films.⁴⁷ According to the X-ray diffraction analysis (Fig. S1†), both prepared films show the expected diffraction lines for quasi-cubic structure K_{0.025}(Cs_{0.1}FA_{0.9})_{0.975}PbI₃ perovskite films. The surface composition of perovskite films after each antisolvent step was investigated by X-ray photoelectron spectroscopy. As depicted in Fig. S2(a) and (b),† the expected core level emission lines for Cs, C, N, Pb, and I

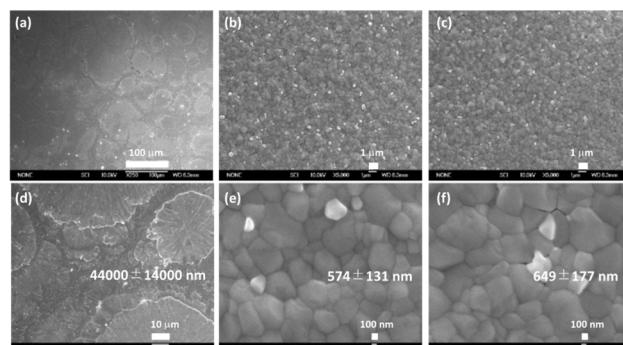


Fig. 2 SEM images on the surface of K_{0.025}(Cs_{0.1}FA_{0.9})_{0.975}PbI₃ film (a and d) without antisolvent treatment, (b and e) with chlorobenzene, and (c and f) with anisole (top (a–c) low-magnification and bottom (d–f) high-magnification SEM images).



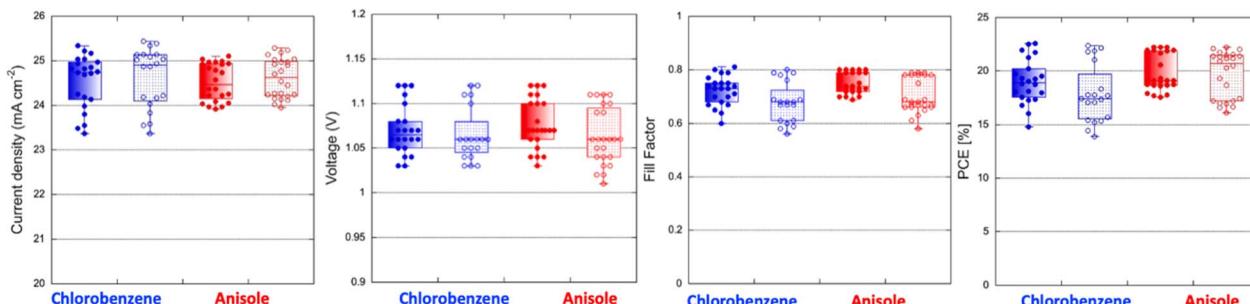


Fig. 3 Statistics of photovoltaic parameters (reverse scan full dots, forward scan open dots) of the perovskite solar cells, $K_{0.025}(Cs_{0.1}FA_{0.9})_{0.975}PbI_3$ layer treated with chlorobenzene and anisole.

Table 1 Photovoltaic parameters of $K_{0.025}(Cs_{0.1}FA_{0.9})_{0.975}PbI_3$ based solar cells using chlorobenzene and anisole as antisolvent

Antisolvent	Scan direction	J (mA cm^{-2})	V (V)	FF	PCE (average) (%)
Chlorobenzene	Reverse	24.9 (24.5 ± 0.6)	1.12 (1.06 ± 0.03)	0.81 (0.72 ± 0.06)	22.6 (19.0 ± 2.1)
	Forward	25.0 (24.6 ± 0.7)	1.12 (1.06 ± 0.03)	0.80 (0.67 ± 0.08)	22.4 (17.9 ± 2.7)
Anisole	Reverse	25.1 (24.5 ± 0.4)	1.12 (1.07 ± 0.03)	0.80 (0.75 ± 0.04)	22.5 (19.8 ± 1.7)
	Forward	25.2 (24.6 ± 0.4)	1.11 (1.06 ± 0.03)	0.79 (0.70 ± 0.07)	22.1 (19.7 ± 2.2)

elements were clearly detected. Moreover, a surface composition of $K_0(Cs_{0.12}FA_{0.88})PbI_{2.88}$ and $K_{0.04}(Cs_{0.18}FA_{0.81})_{0.95}PbI_{2.94}$ was determined for anisole and chlorobenzene-treated films, respectively (Table S6†). These compositions are not far behind the $K_{0.025}(Cs_{0.1}FA_{0.9})_{0.975}PbI_3$ perovskite nominal composition introduced by Mariotti *et al.*³⁶

Before assessing the environmental and technical impacts of the antisolvents used for fabrication of perovskite solar cells the power conversion efficiency (PCE) of solar cells which is crucial parameter for demonstrating comparable or superior efficiencies using anisole, a proposed green solvent, compared to the toxic chlorobenzene. In this context, the devices with perovskite layer treated with chlorobenzene and anisole, planar n-i-p devices with ITO/SnO₂/K_{0.025}(Cs_{0.1}FA_{0.9})_{0.975}PbI₃/spiro-OMeTAD/Au structure were fabricated. Fig. 3 and Table 1 display the statistical distribution of photovoltaic parameters for both cases. The J - V curves, incident photon-to-current efficiency (IPCE), and cross-section SEM images of the best-efficient devices for chlorobenzene and anisole-treated

perovskite films are shown in Fig. 4a-d. Devices made using chlorobenzene as antisolvent exhibited a PCE of 22.6% with an open circuit voltage (V_{oc}) of 1.12 V, a short circuit current (J_{sc}) of 24.9 mA cm^{-2} , and a fill factor (FF) of 0.81 under reverse scan and a PCE of 22.4% with a V_{oc} of 1.12 V, J_{sc} of 25.0 mA cm^{-2} and FF of 0.80 under forward scan (Table 1). The PSCs with $K_{0.025}(Cs_{0.1}FA_{0.9})_{0.975}PbI_3$ layer prepared using anisole as antisolvent yielded similar photovoltaic responses, PCE of 22.5% with a V_{oc} of 1.12 V, J_{sc} of 25.1 mA cm^{-2} , and FF of 0.80 under reverse scan and PCE of 22.1% with V_{oc} of 1.11 V, J_{sc} of 25.2 mA cm^{-2} , and FF of 0.79 under forward scan, but in average with smaller hysteresis index and narrow distribution of data. The integrated photocurrent values (from IPCE), were 23.6 and 24.0 mA cm^{-2} for the devices using chlorobenzene and anisole, respectively, showing quite a close match with the J_{sc} values extracted from J - V curves shown. We found that the efficiency of over 22%, obtained in the present work, is higher than others reported for slightly different perovskite absorbers and device architectures. Besides, IV hysteresis is minimal compared to others which can

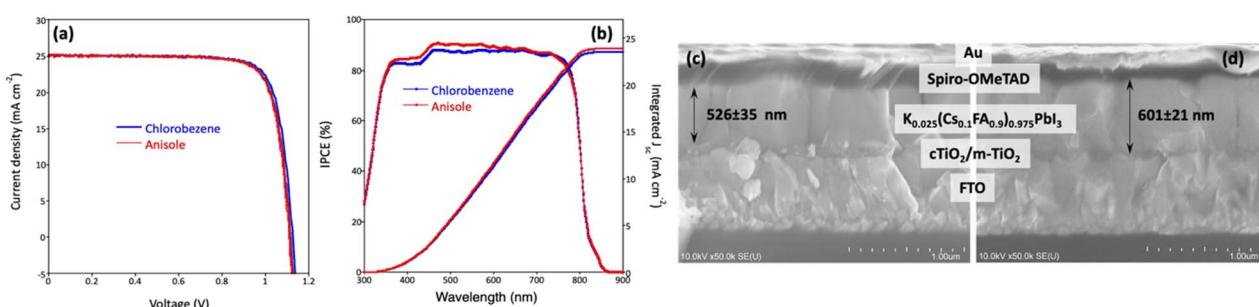


Fig. 4 (a) The best I - V curves (reverse and forward scans), and (b) IPCE of $K_{0.025}(Cs_{0.1}FA_{0.9})_{0.975}PbI_3$ based solar cells, perovskite layer treated with chlorobenzene and anisole, cross-sectional SEM images of $K_{0.025}(Cs_{0.1}FA_{0.9})_{0.975}PbI_3$ based solar cells, perovskite film treated by (c) chlorobenzene and (d) anisole.



be related to the device structure.^{24,35} The efficiencies of 20.01% (18.95%) and 19.0% (18.1%) reported by Wei *et al.*,²⁴ and 20.4% (19.04%) and 20.53% (19.11) reported by Saliba *et al.*³⁵ for chlorobenzene and anisole, respectively.

Life cycle impact assessment results

As shown above, anisole appeared to be a suitable alternative to chlorobenzene as an antisolvent to fabricate efficient PSCs. The question, therefore, arises about the environmental impact of the use of these solvents. In this context, LCIA results of anisole and chlorobenzene, normalized to 1 kg of the chemicals consumed during the fabrication of perovskite solar cells, have been produced and are shown in Table 2 and displayed in Fig. 5.

It is obvious that compared to chlorobenzene, anisole has reduced potential toxicity on human health and the environment. Expressed in comparative toxic units, anisole is responsible for carcinogenic human toxicity of 1.85×10^{-8} CTUh kg⁻¹ compared to 4.17×10^{-8} CTUh kg⁻¹ for chlorobenzene. Similarly, ecotoxicity of anisole was estimated to be 2.13×10^2 CTUe kg⁻¹ whereas the same for chlorobenzene was 3.74×10^2 CTUe kg⁻¹. Moreover, both chemicals have similar non-carcinogenic toxicity with chlorobenzene having a slight edge over anisole embodied in a 13.19% reduced non-carcinogenic toxicity. However, the impact of anisole on climate change was found to be substantially higher than that of chlorobenzene, emitting 6.20 kg CO₂ equivalent per kg of solvent, which was about twice the amount released for chlorobenzene (2.83 kg CO₂ equivalent per kg of solvent). All the remaining categories such as ionizing radiation, ozone depletion, particulate matter, photochemical ozone formation, and water use, that have an indirect effect on human health follow the trend of climate change.⁴⁸ A comparative chart with an embedded contribution analysis (Fig. 5) demonstrates the potential toxicity effects, both direct and indirect, of the two antisolvents. This pinpoints toxicity and environmental hotspots, including potential toxicity hotspots in the life cycle of anisole and chlorobenzene used for laboratory-scale PSC fabrication.

In the following sections, several ideas are explored to reduce the environmental impact of anisole: material efficiency, alternative synthetic routes (bio-sourced precursors), and antisolvent recovery.

Material efficiency, lowering the amount of antisolvents

The possibility of reducing the quantity of antisolvent was investigated by experimenting with various amounts: 50, 100, 300, and 500 μ L. The corresponding performances of PSCs were then evaluated (Fig. 6) and the resulting average photovoltaic parameters are listed in Table S7.[†] Initially, 500 μ L of anisole was used to treat the perovskite absorber on a 25 mm \times 25 mm substrate.

As shown in Fig. 6, decreasing the amount of anisole from 500 μ L to 100 μ L led to comparable performances, an average PCE of $20.1 \pm 0.5\%$ with J_{sc} of 25.6 ± 0.2 mA cm⁻², V_{oc} of 1.06 ± 0.01 V, and FF of 0.74 ± 0.02 and average PCE of $20.0 \pm 0.3\%$ with J_{sc} of 25.9 ± 0.1 mA cm⁻², V_{oc} of 1.06 ± 0.01 V, and FF of 0.73 ± 0.01 , respectively (Table S7[†]). However, at 50 μ L, the efficiency of the solar cells collapsed, and reproducibility became poor, resulting in a wide distribution of efficiencies ($9.8 \pm 8.5\%$). This volume was insufficient to cover the entire perovskite surface, leading to poor morphology on the 25 mm \times 25 mm substrate. For chlorobenzene, lower-performing solar cells were obtained when the quantity was reduced to 300 μ L, an average PCE of $17.1 \pm 3.8\%$ with J_{sc} of 24.9 ± 0.1 mA cm⁻², V_{oc} of 1.02 ± 0.12 V, and FF of 0.66 ± 0.1 were registered. A more significant distribution of data was observed with 100 μ L (average PCE of $12.2 \pm 8.6\%$) and 50 μ L of chlorobenzene (average PCE of $13.3 \pm 6.4\%$), likely due to its faster evaporation compared to anisole (Fig. 6, Table S7[†]). These results demonstrate that lower amounts of anisole (to 100 μ L) can be used to obtain higher-quality perovskite compared to chlorobenzene.

A sensitivity analysis was conducted to evaluate the environmental impact of variations in the use of chlorobenzene and anisole, as shown in Fig. 6. Initially, a 1-to-1 comparison was made using 500 μ L of each solvent. However, data displayed in

Table 2 Impact assessment results according to the environmental impact method version 3.1 (adapted) for two antisolvents, anisole, and chlorobenzene, considering mass (1 kg) as a functional unit

Impact category	Anisole	Chlorobenzene	Unit
Human toxicity, cancer	1.85×10^{-8}	4.17×10^{-8}	CTUh
Human toxicity, non-cancer	1.15×10^{-7}	1.00×10^{-7}	CTUh
Ecotoxicity, freshwater	2.13×10^2	3.74×10^2	CTUe
Climate change	6.20×10^0	2.83×10^0	kg CO ₂ eq.
Ionizing radiation	3.16×10^{-1}	1.23×10^{-1}	kBq U-235 eq.
Ozone depletion	6.93×10^{-7}	5.91×10^{-7}	kg CFC11 eq.
Particulate matter	3.56×10^{-7}	1.83×10^{-7}	Disease inc.
Photochemical ozone formation	2.78×10^{-2}	2.22×10^{-2}	kg NMVOC eq.
Water use	4.27×10^0	3.04×10^0	m ³ depriv.
Acidification	4.13×10^{-2}	1.81×10^{-2}	mol H ⁺ eq.
Eutrophication, marine	6.72×10^{-3}	3.52×10^{-3}	kg N eq.
Eutrophication, freshwater	1.71×10^{-3}	7.00×10^{-4}	kg P eq.
Eutrophication, terrestrial	7.09×10^{-2}	3.72×10^{-2}	mol N eq.
Land use	1.47×10^1	4.80×10^0	Pt
Resource use, fossils	1.37×10^2	6.84×10^1	MJ
Resource use, minerals and metals	1.38×10^{-4}	1.25×10^{-5}	kg Sb eq.



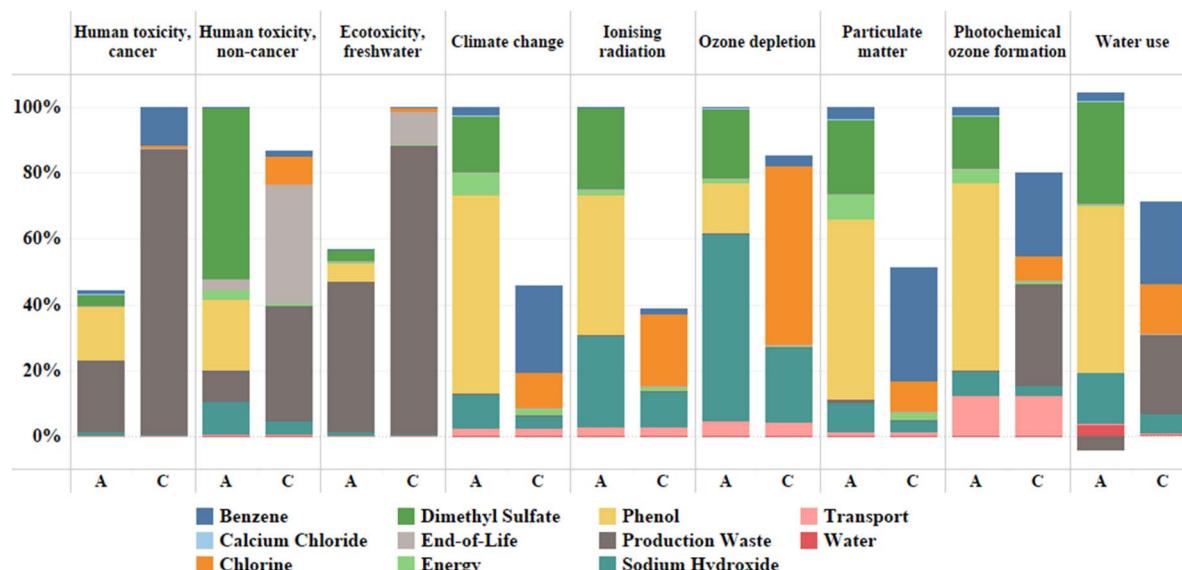


Fig. 5 Comparative and contribution analysis of the potential toxicity and other environmental impacts of anisole (A) and chlorobenzene (C).

Fig. 6 suggest that the amount of anisole can be reduced significantly, either by half or even to one-fifth of the volume of chlorobenzene. Consequently, comparisons were also made on 1-to-2 and 1-to-5 bases. The results, presented in Fig. 7 and Table S8,† demonstrate that reducing the amount of anisole by half makes it more advantageous than chlorobenzene in 9 out of 16 environmental indicators. However, anisole's impact on climate change remains higher compared to chlorobenzene. When the anisole volume is reduced to one-fifth, anisole becomes less environmentally impactful than chlorobenzene across all assessed indicators except for the resource use in the minerals and metals category (1.38×10^{-4} kg Sb eq. for anisole *versus* 6.24×10^{-5} kg Sb eq. for chlorobenzene), as shown in Table S7.†

Bio-based synthesis routes

The lower human carcinogenic and ecosystem toxicities of anisole are attributed to two main factors: the reduced potential toxicity of its production waste and its lower direct toxicity when released into the air, especially in comparison to chlorine-containing compounds like chlorobenzene. However, the

multistep synthesis procedure of anisole contributes to a higher environmental impact due to waste (*i.e.*, benzene used for anisole extraction) while the one-step synthesis of chlorobenzene from chlorine and benzene has a lower environmental impact. Therefore, the adoption of anisole in large-scale manufacturing of PSCs might be beneficial only when produced with higher efficiencies and/or from renewable sources.^{49–51}

Although the bio-based synthesis of anisole holds promise for reducing environmental impact, it remains an unexplored domain. As of the time of writing, no research articles have been published on the bio-based synthesis of anisole. However, some publications have investigated the bio-based synthesis of phenol,^{52–54} which is a key precursor for anisole production. Presently, bio-oil can be utilized to produce renewable phenol,⁵² but the process is highly inefficient. According to our calculations, 263.15 grams of bio-oil extracted from approximately 1.253 kilograms of corn cobs are required to produce just 1 gram of phenol. Aside from economic concerns, scaling this process to an industrial level could have severe environmental implications, especially in terms of land use. Additionally, the energy-intensive pyrolysis process reaches temperatures of up

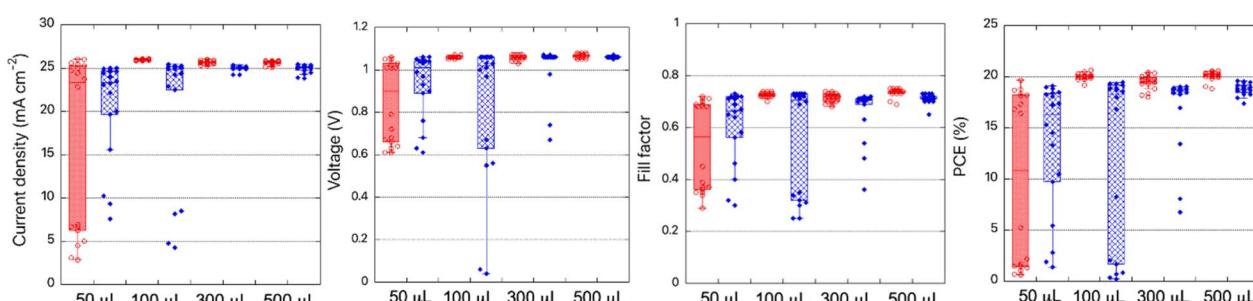


Fig. 6 Influence of anisole and chlorobenzene amount as antisolvent on the photovoltaic parameters of perovskite solar cells. The photovoltaic parameters are extracted from forward scan conditions, red (open dots) are for anisole and blue (square dots) for chlorobenzene.



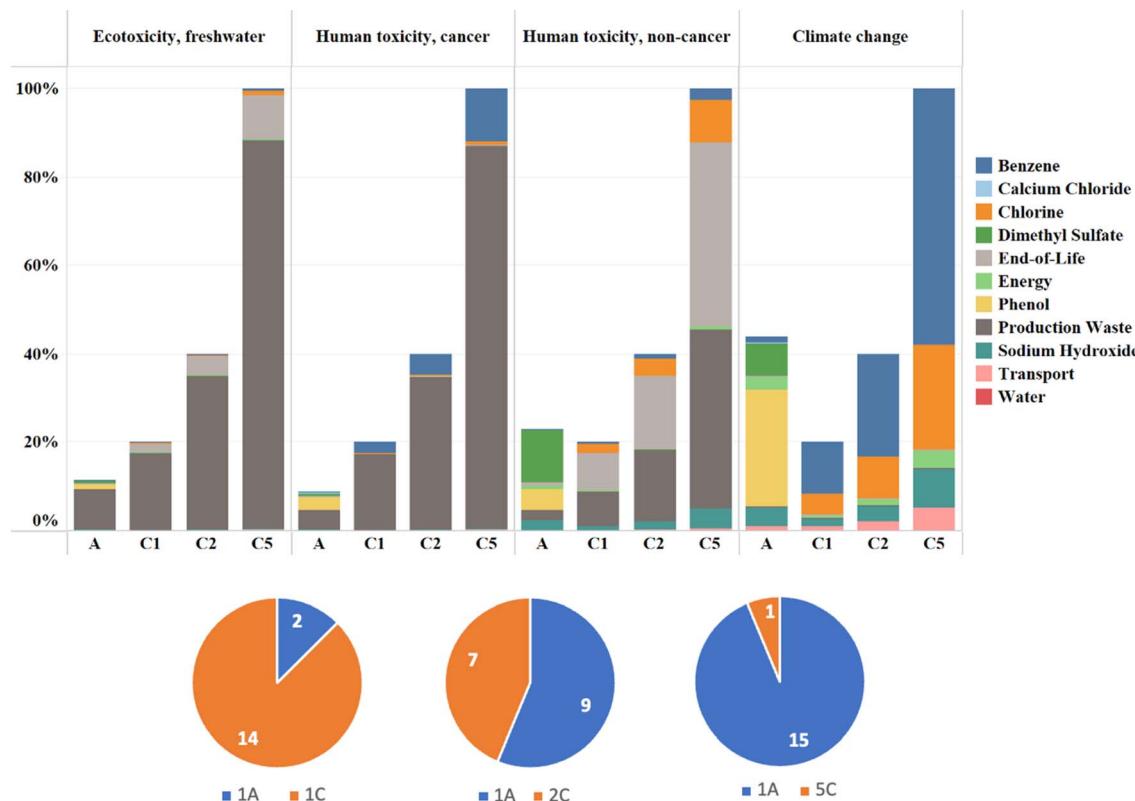


Fig. 7 Above: comparative and contribution analysis of the potential toxicity and climate change impacts of anisole (A) and chlorobenzene (C) on a 1-to-1, 1-to-2, and 1-to-5 basis. Below: pie charts illustrating the number of impact categories where each antisolvent demonstrates a reduced impact compared to its alternative.

to 800 °C,⁵² and catalysis further contributes to the environmental impact of this bio-based synthesis route. Another approach involves synthesizing anisole from lignin, but this method yields only 5% and relies on a $\text{ZrO}_2\text{-FeO}_x$ catalyst.⁵³ It's evident that the synthesis of anisole from biomass still requires significant optimization to become a viable substitute for the petroleum route.

Antisolvent recovery

Antisolvent treatment, like any other production process, can be made more environmentally friendly by incorporating recycling processes. In this case, the recycling of antisolvent could be facilitated through the compartmentalization of the PSC production step. Each material would be deposited on the substrate in a dedicated, closed compartment, allowing for the capture of chemical vapors and their subsequent recycling. For the antisolvents used in this study, this could be achieved *via* condensation, a well-known solvent recovery method.⁵⁵ Research on the vapor recovery of anisole and chlorobenzene in the context of PSC production is currently unexplored. Since these solvents interact with perovskite materials during antisolvent treatment, chemical mixing could occur in the vapors, resulting in impure solvent vapor. Techniques for the separation and recycling of such vapors exist and could be optimized for scaling up this treatment step.^{56,57} This would benefit the environmental performance of both anisole and chlorobenzene

treatments, with greater benefits for the solvent with a higher recovery rate. Therefore, future research needs to focus not only on achieving high-performance PSCs but also on the recovery of anisole, paving the way for industrial scale-up.

The relative “greenness” of chemicals

Labeling chemicals as “green” requires a thorough, context-specific assessment rather than relying on general guidelines. Achieving absolute “greenness” is difficult because chemicals can behave differently under various conditions, leading to diverse outcomes. In other words, a chemical that is environmentally friendly in one synthetic route might be harmful in another. This discrepancy often stems from factors such as its production method, the yield it produces when used as a reactant, the quality of its resulting products, and other factors such as the need for catalysis. For example, the production method can substantially impact the environmental footprint of a chemical. Petrol-based ethanol has around three times the carbon emissions of bio-sourced ethanol.⁴⁵ On the other hand, hydrogen peroxide illustrates how yield can vary across different processes. Thus, the oxidation of propylene with hydrogen peroxide, also known as the HPPO process, yielded high-yield and high-quality propylene oxide at the industrial scale,^{58,59} while wastewater treatment using hydrogen peroxide *via* the Fenton reaction still requires improvements in terms of efficiency to be economically viable.⁶⁰ Consequently, a context-



specific and transparent environmental assessment is essential for making informed decisions when developing new technologies and considering potential substitutes for chemicals. Transparent assessments enable researchers to tailor evaluations to their specific process conditions, enhancing the accuracy and relevance of their results.

Conclusions

The environmental viability of chlorobenzene substitution with anisole has been investigated, shedding light on the need for context-specific and transparent environmental assessment. Solvent selection guides showed anisole as a better alternative to chlorobenzene. Based on this recommendation, researchers have employed anisole as an antisolvent in the fabrication of PSC. A claim that anisole is a “green” antisolvent is made, lacking a scientific foundation. This work evaluates the technical and environmental performance of anisole in the fabrication of PSC.

The use of anisole as a less toxic alternative to chlorobenzene in the processing of PSCs based on $K_{0.025}(Cs_{0.1}FA_{0.9})_{0.975}PbI_3$ was assessed by comparison of solar cell performance. PSCs (ITO/SnO₂/K_{0.025}(Cs_{0.1}FA_{0.9})_{0.975}PbI₃/spiro-OMeTAD/Au) made using anisole led to slightly better performances, showing an efficiency over 22% with improved reproducibility and minimum hysteresis, than those fabricated with chlorobenzene. Performance improvement in the case of anisole was related to a quasi-cubic crystallite, larger grain sizes, *i.e.* 649 ± 177 nm, smooth even surface, resulting in higher V_{oc} and FF. Furthermore, experiments were conducted to determine if smaller volumes of anisole and chlorobenzene could achieve high-performance PSC. The results showed that the volume of anisole could be reduced by up to five times compared to chlorobenzene, while still producing high-performance PSC with excellent reproducibility.

A cradle-to-grave LCA study of anisole and chlorobenzene shows that anisole is less ecotoxic (2.13×10^2 CTUe kg⁻¹) than chlorobenzene (3.74×10^2 CTUe kg⁻¹). Anisole demonstrates significantly low carcinogenic human toxicity, (1.85×10^{-8} CTUh kg⁻¹) in comparison to chlorobenzene (4.17×10^{-8} CTUh kg⁻¹). However, the LCA results revealed other environmental impacts of anisole to be higher than chlorobenzene's, arising from anisole's multi-step production process which is more complex than chlorobenzene. Nonetheless, the ability to use lower quantities of anisole than chlorobenzene in the fabrication of PSC can lower its environmental impact. A sensitivity analysis shows that anisole amount should be lowered by five-fold to provide a higher environmental performance than chlorobenzene, with reservation to the resource use impact category.

This study illustrates that context-specific and transparent environmental assessments are needed to make informed decisions in research and development leading to environmentally sustainable solutions. The environmental friendliness or so-called “greenness” of anisole is investigated. As PSCs are currently at a trial stage of commercial development, we believe that this work of addressing the solvent potential toxicity during

perovskite film fabrication contributes to the development of environmentally conscious industrial production.

Data availability

Data discussing the results of this article are available as a part of the main text and the data supporting this article have been included as a part of ESI.†

Author contributions

N. M. K. and Y. Y. performed device fabrication, evaluation, and data analysis of the devices. A. K. K. and G. S. assess the potential toxicity among other environmental impacts using LCA. N. M. K. wrote the manuscript and revised it by Y. Y., A. K. K., A. K. J., T. T., G. S., H. S., S. U., L. C. All authors approved the final version of the manuscript. L. C. T. T. contributed to the conceptualization and project supervision. All authors contributed to manuscript development.

Conflicts of interest

There are no conflicts to declare.

Acknowledgements

This work benefited from State assistance managed by the National Research Agency under the “Programme d'Investissements d'Avenir” (ANR-19-MPGA-0006), supported by ELOR-PrintTec ANR-10-EQPX-28-01, and the New Energy and Industrial Technology Development Organization (NEDO) PV R&D programs. The authors would like to acknowledge IRP-NextPV-II a CNRS and RCAST/University Tokyo collaboration. Dr Christine Labrugère (PLACAMAT) is thanked for her precious assistance in recording XPS data and Philippe Legros (PLACAMAT) for SEM measurements.

References

- 1 Bjørn, M. Diamond, M. Owsiania, B. Verzat and M. Z. Hauschild, *Environ. Sci. Technol.*, 2015, **49**, 6370–6371.
- 2 C. Wu, K. Wang, J. Li, Z. Liang, J. Li, W. Li, L. Zhao, B. Chi and S. Wang, *Matter*, 2021, **4**(3), 775–793.
- 3 J. Doolin, R. G. Charles, C. S. P. De Castro, R. G. Rodriguez, E. V. Péan, R. Patidar, T. Dunlop, C. Charbonneau, T. Watson and M. L. Davies, *Green Chem.*, 2021, **23**(6), 2471–2486.
- 4 M. Zhang, D. Xin, X. Zheng, Q. Chen and W. H. Zhang, *ACS Sustain. Chem. Eng.*, 2020, **8**(35), 13126–13138.
- 5 D. Prat, A. Wells, J. Hayler, H. Sneddon, C. R. McElroy, S. Abou-Shehada, P. J. Dunn and P. CHEM21, *Green Chem.*, 2016, **18**(1), 288–296.
- 6 M. Alder, J. D. Hayler, R. K. Henderson, A. M. Redman, L. Shukla, L. E. Shuster and H. F. Sneddon, *Green Chem.*, 2016, **18**(13), 3879–3890.
- 7 J. Y. Kim, J. W. Lee, H. S. Jung, H. Shin and N. G. Park, *Chem. Rev.*, 2020, **120**, 7867–7918.



8 <https://www.nrel.gov/pv/assets/pdfs/best-research-cell-efficiencies.pdf>.

9 N. J. Jeon, J. H. Noh, Y. C. Kim, W. S. Yang, S. Ryu and S. I. Seok, *Nat. Mater.*, 2014, **13**, 897–903.

10 F. Ma, Y. Zhao, Z. Qu and J. You, *Acc. Mater. Res.*, 2023, **4**, 716–725.

11 N. Ahn, D. Y. Son, I. H. Jang, S. M. Kang, M. Choi and N. G. Park, *J. Am. Chem. Soc.*, 2015, **137**, 8696–8699.

12 W. S. Yang, J. Noh, N. J. Jeon, Y. C. Kim, S. Ryu, J. Seo and S. Seok, *Science*, 2015, **348**, 1234–1237.

13 A. Srivastava, J. A. K. Satrughna, M. K. Tiwari, A. Kanwade, S. C. Yadav, K. Bala and P. M. Shirage, *Mater. Today Commun.*, 2023, **35**, 105686.

14 J. Burschka, N. Pellet, S. J. Moon, R. Humphry-Baker, P. Gao, M. K. Nazeeruddin and M. Grätzel, *Nature*, 2013, **499**, 316–319.

15 M. Xiao, F. Huang, W. Huang, Y. Dkhissi, Y. Zhu, J. Etheridge, A. Gray-Weale, U. Bach, Y.-B. Cheng and L. Spiccia, *Angew. Chem., Int. Ed.*, 2014, **53**, 9898–9903.

16 S. Dinghan, S. Yu, X. Cai, M. Peng, Y. Ma, X. Su, L. Xiao and D. Zou, *J. Mater. Chem. A*, 2014, **2**, 20454–20461.

17 L. Cojocaru, K. Wienands, T. W. Kim, S. Uchida, A. J. Bett, S. Rafizadeh, J. C. Goldschmidt and S. Glunz, *ACS Appl. Mater. Interfaces*, 2018, **10**, 26293–26302.

18 R. Vidal, J. Alberola-Borràs, N. Sánchez-Pantoja and I. Mora-Seró, *Adv. Energy Sustainability Res.*, 2021, **2**(5), 2000088.

19 S. Weyand, C. Wittich and L. Schebek, *Energies*, 2019, **12**(22), 4228.

20 T. Bu, L. Wu, X. Liu, X. Yang, P. Zhou, X. Yu, T. Qin, J. Shi, S. Wang and S. Li, *Adv. Energy Mater.*, 2017, **7**(20), 1700576.

21 N.-G. Park, *Nat. Sustain.*, 2021, **4**(3), 192–193.

22 J. Lee, M. Malekshahi Byranvand, G. Kang, S. Y. Son, S. Song, G. W. Kim and T. Park, *J. Am. Chem. Soc.*, 2017, **139**(35), 12175–12181.

23 M. Wang, Q. Fu, L. Yan, J. Huang, Q. Ma, M. Humayun, W. Pi, X. Chen, Z. Zheng and W. Luo, *Chem. Eng. J.*, 2020, **387**, 123966.

24 X. Cao, L. Hao, G. Su, X. Li, T. Dong, P. Chao, D. Mo, Q. Zeng, X. He and J. Wei, *RSC Sustainability*, 2023, **1**(5), 1290–1297.

25 W. Xu, Y. Gao, W. Ming, F. He, J. Li, X. Zhu, F. Kang, J. Li and G. Wei, *Adv. Mater.*, 2020, **32**(38), 2003965.

26 G. Jang, H. Kwon, S. Ma, S. Yun, H. Yang and J. Moon, *Adv. Energy Mater.*, 2019, **9**(36), 1901719.

27 Y. Cui, S. Wang, C. Li, A. Wang, J. Ren, C. Yang, B. Chen, Z. Wang and F. Hao, *Green Chem.*, 2021, **23**(10), 3633–3641.

28 M. Zhang, Z. Wang, B. Zhou, X. Jia, Q. Ma, N. Yuan, X. Zheng, J. Ding and W. Zhang, *Sol. RRL*, 2018, **2**(2), 1700213.

29 M. T. Hoang, F. Ünlü, W. Martens, J. Bell, S. Mathur and H. Wang, *Green Chem.*, 2021, **23**(15), 5302–5336.

30 Z. Tang, S. Wang, W. Zhu, L. Ding and F. Hao, *Green Chem.*, 2023, **25**(3), 1150–1156.

31 D. Taylor, Q. Sun, K. P. Goetz, Q. An, T. Schramm, Y. Hofstetter, M. Litterst, F. Paulus and Y. Vaynzof, *Nat. Commun.*, 2021, **12**, 1–11.

32 S. Ghosh, S. Mishra and T. Singh, *Adv. Mater. Interfaces*, 2020, **7**, 2000950.

33 Y. Yun, F. Wang, H. Huang, Y. Fang, S. Liu, W. Huang, Z. Cheng, Y. Liu, Y. Cao, M. Gao, L. Zhu, L. Wang, T. Qin and W. Huang, *Adv. Mater.*, 2020, **32**, 1907123.

34 M. Wang, Q. Fu, L. Yan, J. Huang, Q. Ma, M. Humayun, W. Pi, X. Chen, Z. Zheng and W. Luo, *Chem. Eng. J.*, 2020, **387**, 123966.

35 M. Yavari, M. Mazloum-Ardakani, S. Gholipour, M. Mahdi Tavakoli, S.-H. Turren-Cruz, N. Taghavinia, M. Grätzel, A. Hagfeldt and M. Saliba, *Adv. Energy Mater.*, 2018, **8**, 1800177.

36 S. Mariotti, D. Mantione, S. Almosni, M. Ivanović, T. Bessho, M. Furue, H. Segawa, G. Hadzioannou, E. Cloutet and T. Toupance, *J. Mater. Chem. C*, 2022, **10**, 16583–16591.

37 R. K. Henderson, C. Jiménez-González, D. J. C. Constable, S. S. R. Alston, G. A. Inglis, G. Fisher, J. Sherwood, S. P. Binks and A. D. Curzons, *Green Chem.*, 2011, **13**, 854–862.

38 ISO – 14044 – Environmental Management – Life Cycle Assessment – Requirements and Guidelines, Geneva, 2006.

39 ISO – 14040 – Environmental Management – Life Cycle Assessment – Principles and Framework, Geneva, 2006.

40 Anisole, *Organic Syntheses, Coll.*, 1941, **1**, 58.

41 <https://www.essentialchemicalindustry.org/chemicals/phenol.html>.

42 F. Piccinno, R. Hischier, S. Seeger and C. Som, *J. Cleaner Prod.*, 2016, **135**, 1085–1097.

43 B. A. Andreasi, F. Biganzoli, N. Ferrara, A. Amadei, A. Valente, S. Sala and F. Ardente, *European Commission, Joint Research Centre*, Publications Office of the European Union, 2023.

44 M. Damiani, N. Ferrara and F. Ardente, *European Commission, Joint Research Centre*, Publications office of the European Union, Luxembourg, LUX, 2022.

45 G. Wernet, C. Bauer, B. Steubing, J. Reinhard, E. Moreno-Ruiz and B. Weidema, *Int. J. Life Cycle Assess.*, 2016, **21**, 1218–1230.

46 Y. Yu, S. Yang, L. Lei, Q. Cao, J. Shao, S. Zhang and Y. Liu, *ACS Appl. Mater. Interfaces*, 2017, **9**, 3667–3676.

47 P. Zhao, B. J. Kim, X. D. Ren, D. G. Lee, G. J. Bang, J. B. Jeon, W. B. Kim and H. S. Jung, *Adv. Mater.*, 2018, **30**, 1802763.

48 Joint Research Centre, Institute for Environment and Sustainability, *International Reference Life Cycle Data System (ILCD) Handbook – Framework and Requirements for Life Cycle Impact Assessment Models and Indicators*, Publications Office, 2010.

49 H. Taghvaei and M. R. Rahimpour, *RSC Adv.*, 2016, **6**, 98369–98380.

50 R. C. Runnebaum, T. Nimmanwudipong, D. E. Block and B. C. Gates, *Catal. Sci. Technol.*, 2011, **2**, 113–118.

51 A. Ould Hamou and J. B. Giorgi, *J. Phys. Chem. C*, 2019, **123**, 8122–8132.

52 L. Dai, Z. Zeng, Q. Yang, S. Yang, Y. Wang, Y. Liu, R. Ruan, C. He, Z. Yu and L. Jiang, *Fuel*, 2020, **279**, 118532.

53 H. Ishimaru, T. Yoshikawa, Y. Nakasaka, E. Fumoto, S. Sato and T. Masuda, *Catal. Today*, 2023, **407**, 194–200.

54 M. Wojcieszek, L. Knuutila, Y. Kroyan, M. balsemao, R. Tripathi, J. Keskitalo, A. Karvo, A. Santasalo-Aarnio,



O. Blomstedt and M. Larmi, *Sustainability*, 2021, **13**(16), 8729.

55 M. Smallwood, *Solvent Recovery Handbook*, CRC, 2002.

56 V. Nabok, A. K. Hassan and A. K. Ray, *J. Mater. Chem.*, 2000, **10**, 189–194.

57 M. Leemann, G. Eigenberger and H. Strathmann, *J. Membr. Sci.*, 1996, **113**, 313–322.

58 F. Schmidt, M. Bernhard, H. Morell and M. Pascaly, *Chim. Oggi – Chem. Today*, 2014, **32**, 31–35.

59 V. Russo, R. Tesser, E. Santacesaria and M. Di Serio, *Ind. Eng. Chem. Res.*, 2013, **52**, 1168–1178.

60 P. Vinosh Muthukumar, B. Gopalakrishnan and B. Bharathiraja, *J. Ind. Chem. Soc.*, 2022, **99**, 100622.

