



Cite this: *RSC Adv.*, 2024, 14, 20585

Development of an aptasensor for dibutyl phthalate detection and the elucidation of assay inhibition factors†

Hyerin Song,^{ab} Hyun Jeong Lim^{ab} and Ahjeong Son  ^{*ab}

We developed a fluorescence aptasensor (hereafter 'SG-aptasensor') using SYBR Green I, a newly truncated 20-mer aptamer, and probe DNA to detect dibutyl phthalate (DBP). The detection range of DBP was 0.1–100 ng L⁻¹ with 0.08 ng L⁻¹ as the limit of detection. To adapt the assay to environmental samples in the near future, possible inhibition factors (experimental and environmental) have been tested and reported. The experimental inhibitors included the incubation time, temperature, pH, and ionic strength. Consequently, temperature (2–25 °C) and pH (7.0–9.0) ranges did not significantly inhibit the assay. The incubation time required for sufficient reaction was at least 4 h, and a relative humidity <20% may have induced fluorescence quenching. Tris–HCl-based incubation buffer with excess ionic strength (more than 0.2 M NaCl) demonstrated an abnormal increase in fluorescence. Environmental inhibitors including cations (Mg²⁺, Ca²⁺, and Cu²⁺) and humic acids were tested. The fluorescence signal was significantly reduced (~99%) by 100 mM Cu²⁺ compared to that by 0 mM Cu²⁺. In contrast, the reduction in fluorescence signal was marginal (<15%) when Mg²⁺ or Ca²⁺ ions were present. Inhibition of the assay was observed (~28%) in the presence of 100 mg L⁻¹ humic acids.

Received 24th April 2024
Accepted 17th June 2024

DOI: 10.1039/d4ra03045a

rsc.li/rsc-advances

1. Introduction

Phthalic acid esters (PAEs) enhance flexibility and transparency as plasticizers for various industrial products such as electronics, plastics, and food packaging.¹ Because PAEs can be easily released into the environment by leaching out from the products, their exposure to humans has increased. Subsequently, the endocrine disruption to humans by PAEs has been observed in various forms, such as developmental malformation, interference with reproduction in humans, and disturbances in the immune and nervous systems.² At this juncture, the US EPA set the action plan to regulate PAEs including dibutyl phthalate (DBP), diisobutyl phthalate (DIBP), butyl benzyl phthalate (BBP), di(2-ethylhexyl) phthalate (DEHP), diisononyl phthalate (DINP), and diisodecyl phthalate (DIDP), di-*n*-pentyl phthalate, di-*n*-octyl phthalate.³ European Chemicals Agency of the European Commission added endocrine disrupting properties to DBP, DEHP, BBP, and DIBP in 2021.⁴

Establishing PAE detection methods is important for preventing and mitigating potential hazards from chemicals that are

not yet replaceable. In addition to reliable instrumental analysis, biosensor technologies have been developed to provide screening tools for these chemicals. Biosensors for environmental monitoring have demonstrated their advantages, including specificity, fast response times, low cost, and ease of use.⁵ Among the biological receptors, aptamers are short, single-stranded nucleotides. Several aptamers and related aptasensors have been developed for detecting PAEs (Table 1).

Analyzing environmental samples using biosensors can be challenging due to the complexity of their nature. Environmental samples contain a wide range of potentially interfering substances, such as organic and inorganic compounds, microbes, and particulate matter, making it difficult to detect specific targets.¹⁷ Experimental conditions, such as temperature and pH, also change dynamically, causing fluctuations in biosensor signals.¹⁸

As summarized in Table 2, previous studies have indicated that inhibition occurs during biosensor-based analyses. Wang *et al.*¹⁹ and Jin *et al.*²⁰ described that the assay was inhibited by Mg²⁺, possibly through the mechanism of DNA aggregation, followed by disruption of DNA hybridization. Zhou *et al.*²¹ reported a reduction in the fluorescence signal due to water hardness from Ca²⁺ and Mg²⁺ ions. Zhan *et al.*²² indicated the interference of fluorescence resonance energy transfer (FRET) process by sodium vanadate. Wu *et al.*²³ showed that an assay using nanomaterials was inhibited by butyrylcholinesterase, causing the aggregation of AuNPs. Kim *et al.*²⁴ and Jin *et al.*²⁰ developed an inhibition resistance assay based on DNA hybridization and quantum dot nanoparticles.

^aDepartment of Environmental Science and Engineering, Ewha Womans University, 52 Ewhayeodae-gil, Seodaemun-gu, Seoul 03760, Republic of Korea. E-mail: ason@ewha.ac.kr; ahjeong.son@gmail.com; Tel: +82(2)3277-3339

^bCenter of SEBIS (Strategic Solutions for Environmental Blindspots in the Interests of Society), 52 Ewhayeodae-gil, Seodaemun-gu, Seoul 03760, Republic of Korea

† Electronic supplementary information (ESI) available. See DOI: <https://doi.org/10.1039/d4ra03045a>



Table 1 List of aptasensor studies conducted for detecting phthalic acid esters (PAEs)^a

Transducer type	Target	Characteristics of assay	Sensitivity of assay (LOQ & linearity range)	Selectivity of assay	References
Optics-based (fluorescence and colorimetry) transducer	DBP	Aptamer-SYBR Green I (SG-aptasensor)	0.0001–0.1 $\mu\text{g L}^{-1}$	DBP, nonylphenol ethoxylate, triclosan, bisphenol A (BPA), bisphenol S (BPS)	This study
	DMP, DEP, DBP, DHP, DIBP, DINP, DPP, BBP, MEHP, DEHP, PA	AuNP-gQD aptasensor	0.001–50 $\mu\text{g L}^{-1}$	DMP, DEP, DBP, DHP, DIBP, DINP, DPP, BBP, MEHP, DEHP, PA Nonylphenol, benzoic acid, BPA, BPS, bisphenol F, DES, beta-estradiol	Lim <i>et al.</i> (2022) ⁶
	PA, DMP, DEP, DBP, DIBP, BBP, DEHP	Non-equilibrium rapid replacement aptamer (NERRA) assay using aptamer and PoPo 3 dye	0.1–200 $\mu\text{g L}^{-1}$ (30 min) 1–100 $\mu\text{g L}^{-1}$ (30 s)	7 PAEs (PA, DMP, DEP, DBP, DIBP, BBP, DEHP), BPA, 4-nonylphenol	Kim <i>et al.</i> (2020) ⁷
	DEHP, DBP, BBP	Aptamer-AuNP-based colorimetric assay	0.003–10 $\mu\text{g L}^{-1}$ (mixture)	Mixture (DEHP, DBP, BBP), Cd^{2+} , atrazine, PCB77, PCB126, estrone, estradiol, ethinylestradiol, glucose, L-histidine, humic acids	Chen <i>et al.</i> (2021) ⁸
	DEP, DBP, DEHP	DNA-modified AuNPs based colorimetric sensor	421–1661 $\mu\text{g L}^{-1}$ (DEP) 321–701 $\mu\text{g L}^{-1}$ (DBP) 841–3322 $\mu\text{g L}^{-1}$ (DEHP)	DEP, DBP, DEHP, Fe^{2+} , Ni^{2+} , Zn^{2+} , Na^{+} , K^{+} , Cu^{2+} , CO_3^{2-} , NO_3^{-} , PO_4^{3-} , $\text{CH}_3\text{COO}^{-}$	Guo <i>et al.</i> (2021) ⁹
Electrochemical-based transducer	DEHP	Signaling-probe displaced electrochemical aptamer-based biosensor (SD-EAB)	0.0039–39 $\mu\text{g L}^{-1}$	DEHP, Hg^{2+} , Cr^{3+} , Cd^{2+} , ethyl acetate, benzoic acid, PA, kanamycin, sulfadimethoxine	Han <i>et al.</i> (2017) ¹⁰
	DEHP	AuFs-methylene blue	0.0005–0.001 $\mu\text{g L}^{-1}$	DEHP, DMP, DEP, DINP, DIBP, BBP, DIBP	Lee <i>et al.</i> (2022) ¹¹
	DEHP	DNA junction-aptamer-MCH-capture probe-Au electrode	0.1–5000 $\mu\text{g L}^{-1}$	DEHP, DOP, DPHP, BBP, DBP, DNHP	Chen <i>et al.</i> (2022) ¹²
	DBP	Coating SMIPs on the surface of modified GCE	0.1–10 000 $\mu\text{g L}^{-1}$	DBP, BPA, DVB, PPD, L-TRP	Wang <i>et al.</i> (2022) ¹³
PEC-based transducer	DBP	MIT using metal organic framework and Cu_2O heterostructure	0.000028–0.278 $\mu\text{g L}^{-1}$	DBP, NH_4^{+} , K^{+} , Na^{+} , Ca^{2+} , Mg^{2+} , SO_4^{2-} , Cl^{-} , and NO_3^{-}	Yu <i>et al.</i> (2023) ¹⁴
SERS-based transducer	DEHP	$\text{AgNCs-SiO}_2\text{-NH}_2$	0.0032–72.8 $\mu\text{g L}^{-1}$	DEHP, DEP, DBP, DINP, DIDP, BBP, TOTM	Tu <i>et al.</i> (2019) ¹⁵
	DBP	UCNPs decorated with AuNPs-aptamer	0.001–100 $\mu\text{g L}^{-1}$	DBP, DEHP, BBP, ethyl acetate, PA, Na^{+} , Mg^{2+} , Ca^{2+} , K^{+} , Fe^{2+}	Rong <i>et al.</i> (2021) ¹⁶

^a DMP: dimethyl phthalate, DEP: diethyl phthalate, DHP: dihexyl phthalate, DPP: dipentyl phthalate, MEHP: mono-2-ethylhexyl phthalate, PA: phthalic acid, PCB: polychlorinated biphenyl, AuFs: gold-nanoflowers, MCH: 6-mercapto-1-hexanol, DOP: dioctyl phthalate, DPHP: di(2-propylheptyl) phthalate, DNHP: di-*n*-hexyl phthalate, SMIPs: surface molecularly imprinted polymers, GCE: glassy carbon electrode, DVB: divinyl benzene, PPD: *p*-phenylenediamine, L-TRP: L-tryptophan, PEC: photoelectrochemical, MIT: molecular imprinted technology, SERS: surface-enhanced Raman spectroscopy, AgNCs: silver nanoclusters, TOTM: trioctyl trimellitate, UCNPs: upconversion nanoparticles.

In this study, we developed an aptasensor to detect DBP (SG-aptasensor) to investigate the possible inhibition effects of various factors in environmental samples. The tested inhibition factors included experimental (incubation time, temperature, pH, and ionic strength) and environmental factors (divalent cations of Mg^{2+} , Ca^{2+} , Cu^{2+} , and humic acids).

2. Materials and methods

2.1. Preparation of aptamer and probe DNA

The schematic in Fig. 1A shows the interaction of the aptamer and complementary probe DNA with DBP and SYBR Green I. The aptamer, 5'-TCT GTC CTT CCG TCA CAG GT-3' (20-mer) was



Table 2 List of inhibition factors studied for the biosensor applications^a

Biosensor type	Components	Target	Inhibition factors	Inhibition type (effects)	References
Fluorescence aptasensor	NanoGene assay MB-QD-probe & signaling probe DNA	<i>E. coli</i> O157 : H7 (bacteria)	Humic acids, Ca ²⁺ , SDS, ethanol	Compared to PCR, MB-QD assay is resistant to the presence of inhibitors	Kim <i>et al.</i> (2011) ²⁴
		<i>Pseudomonas putida</i> (bacteria)	Mg ²⁺	DNA aggregation	Wang <i>et al.</i> (2018) ²⁵
		<i>Microcystis aeruginosa</i> (bacteria)	Mg ²⁺	Prevent disrupting DNA hybridization using electrical discharge treatment	Jin <i>et al.</i> (2020) ²⁰
	SYBR Green I & Ag ⁺ specific oligonucleotides	Ag ⁺	Ca ²⁺ & Mg ²⁺ water hardness, hypochlorite	The instability of silver hypochlorite formed by silver ions and hypochlorite and the oxidation of hypochlorite, which would cause an unstable DNA	Zhou <i>et al.</i> (2020) ²¹
Fluorescence immunosensor (FRET)	CQDs & AuNps	Paraoxon (insecticide)	Butyrylcholinesterase (BChE)	Causing the aggregation of AuNPs and the corresponding recovery of FRET-quenched fluorescence emission	Wu <i>et al.</i> (2017) ²⁶
	N/S-CDs & 2,3-DPA	Alkaline phosphatase (ALP)	Sodium vanadate (Na ₃ VO ₄)	Na ₃ VO ₄ inhibited the process of ALP hydrolysis of PPi. (PPi and free Cu ²⁺ form a stable complex, which cannot form DPA, in the absence of ALP)	Zhan <i>et al.</i> (2021) ²²
Electrochemical aptasensor	GFETs sensor with PDMS	17 β -estradiol (E2) (EDCs)	pH, ionic strength	pH & ionic strength value in the environment (tap water) could fluctuate with time	Li <i>et al.</i> , (2019) ²⁷
	Sonic Hedgehog/ aptamer complexes	Sonic Hedgehog (protein)	Exonuclease III	Inhibiting cleavage of aptamers by exonuclease III <i>via</i> the steric hindrance effect to yield the displacement strands	Chen <i>et al.</i> (2023) ²⁸

^a N/S-CDs: nitrogen/sulfur co-doped carbon dots, DPA: diaminophenazine, PPi: pyrophosphate, GFET: graphene field-effect transistors, PDMS: polydimethylsiloxane, XOD: xanthine oxidase.

truncated from the previous study's aptamer (27-mer),⁶ which has shown the specific binding to PAEs. The probe DNA, 5'-TGT GAC GGA A-3', was designed as complementary to the aptamer sequence in the previous study⁶ (Fig. 1B). All oligonucleotides used in this study were commercially synthesized and purified using high-performance liquid chromatography (Bioneer Co., Daejeon, Korea). The aptamer and probe DNA were mixed with TE buffer (Tris-HCl 10 mM of pH 8.0 and EDTA 1 mM, Bioneer Co.), making their concentrations 100 μ M and 200 μ M, respectively. The final concentration of the aptamer and probe DNA in the reaction was 2.5 μ M after dilution with Tris-HCl buffer. Tris-HCl buffer (pH 8.0) comprises 0.02 M of Tris-HCl (Dyne Bio Inc., Gyeonggi-do, Korea), 0.02 M of MgCl₂ · 6H₂O (Daejung, Gyeonggi-do, Korea), 0.04 M of KCl (Duksan, Gyeonggi-do, Korea), and

0.1 M of NaCl (Junsei, Tokyo, Japan). All experiments were performed in triplicates unless otherwise stated.

2.2. Fluorescence measurement

SYBR Green I (Invitrogen, Carlsbad, CA, USA) is a fluorescence dye that intercalates double-stranded nucleic acids. The stock solution (10 000 \times) was prepared with 1 mL of dimethyl sulfoxide, and the final SYBR Green I concentration used for the SG-aptasensor was 1 \times , which was serially diluted with Tris-HCl buffer.

The fluorescence was measured at λ_{ex} = 265 nm and λ_{em} = 525 nm using a SpectraMax M2 spectrofluorometer (Molecular Devices, San Jose, CA, USA). The fluorescence signal was converted to normalized fluorescence based on eqn (1) to minimize the background signal, which changed in every reaction.



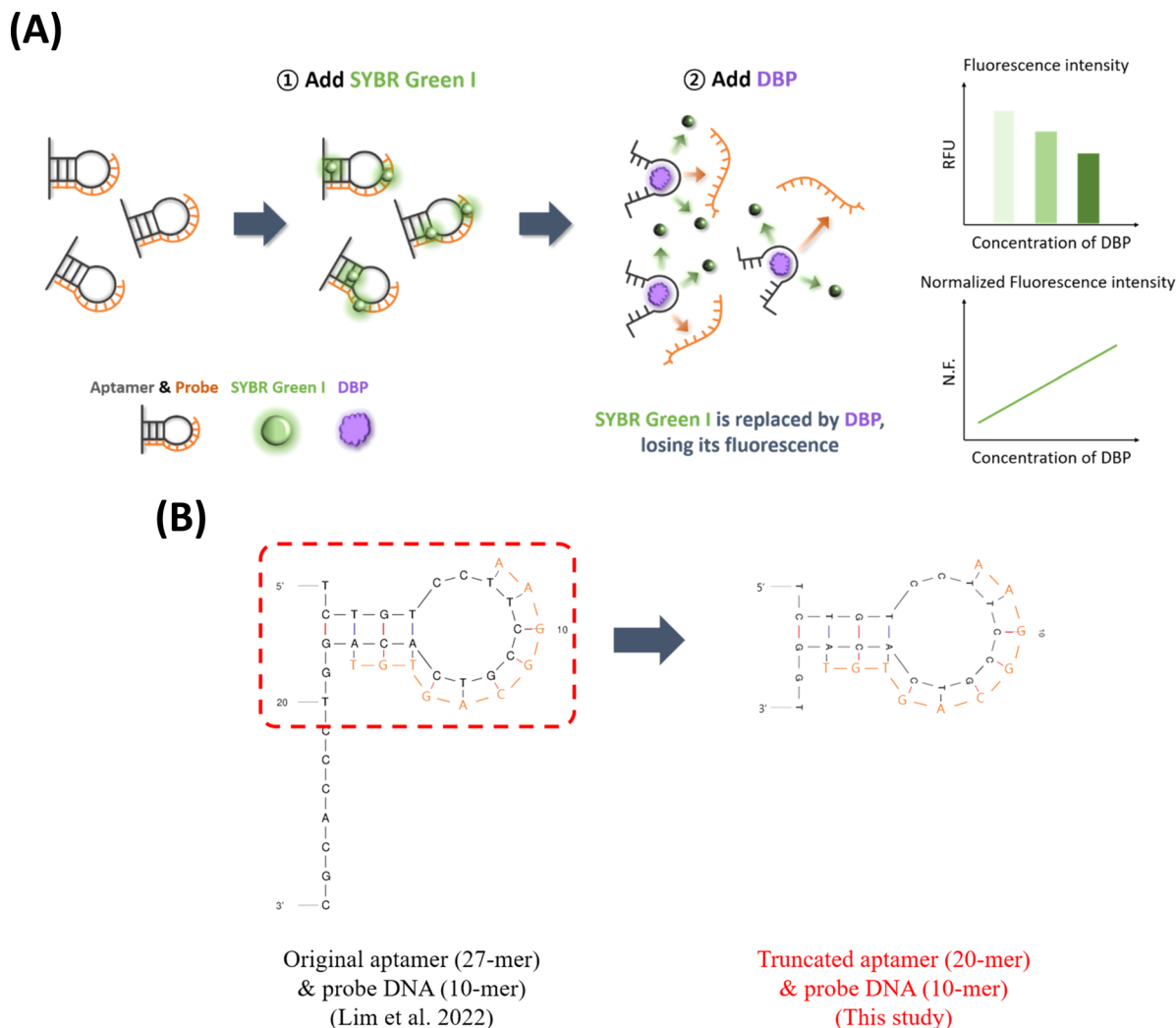


Fig. 1 (A) The schematic of SG-aptasensor for DBP detection and (B) secondary structures of the original aptamer, truncated aptamer, and probe DNA.

$$\text{Normalized fluorescence} = \frac{F_c - F_s}{F_c} \quad (1)$$

where F_c and F_s refer to the fluorescence signals of the negative control and sample, respectively.

2.3. Quantification of DBP using the SG-aptasensor

The reaction (total 200 μL) of the assay was prepared with Tris-HCl buffer (70 μL), the aptamer (100 μL , 5 μM), probe DNA (10 μL , 50 μM), and SYBR Green I (20 μL , 10 \times). Subsequently, target analyte DBP (20 μL , various concentrations) was added. The samples were then incubated for 4 h at ambient temperature (25 $^{\circ}\text{C}$) with gentle mixing at 300 rpm (MixMate Shaker, Eppendorf, Hamburg, Germany). For the quantification experiment, 100 mL DBP stock solution was prepared to 1000 mg L^{-1} by adding 0.1 g of DBP (99% purity, Junsei) to methanol (LC-MS Grade, Thermo Fisher Scientific, Waltham, MA, USA). DBP was serially diluted with deionized water to obtain final concentrations of 0.1, 0.5, 1, 5, 10, 50, 100, 500, 1000, 5000, and 10 000 ng L^{-1} .

For the selectivity experiment, four endocrine-disrupting or potentially endocrine-disrupting compounds were selected for

comparison with DBP.^{29,30} The details of four chemicals are listed in Table S1.† All compounds affect the human endocrine system. Nonylphenol ethoxylate (NPE) is commonly used as a surfactant in various products.³¹ Triclosan (TCS) has been used as an antibacterial agent for personal products and is known for disrupting the thyroid hormone.^{32,33} Bisphenol A (BPA) and bisphenol S (BPS) are chemical analogs and well-known endocrine-disrupting chemicals that are used in plastics, receipts, and food packaging.³⁴ DBP, NPE (70% in H_2O , Sigma-Aldrich, St Louis, MO, USA), TCS (>98%, TCI Co., Tokyo, Japan), BPA (>99%, Daejung), and BPS ($\geq 98\%$, Sigma-Aldrich) were first prepared to 1000 mg L^{-1} stock solution in methanol (LC-MS grade, Thermo Fisher Scientific) and subsequently diluted with Tris-HCl buffer. Each chemical (20 μL , 1 ng L^{-1}) was subjected to SG-aptasensor reaction, including the Tris-HCl buffer (70 μL), aptamer (100 μL), and probe (10 μL).

2.4. Assay inhibition experiments

To investigate the assay inhibition factors, four experimental factors (incubation time, temperature, pH, and ionic strength of



the Tris-HCl buffer) and four environmental factors (Mg^{2+} , Ca^{2+} , Cu^{2+} , and humic acids) were selected.

2.4.1. Experimental inhibitors. Experimental inhibitors refer to experimental conditions, such as temperature and pH. The assay inhibition ranges of these factors have been investigated to identify potential obstacles to aptasensor-based detection. For the incubation time experiment, each reaction included Tris-HCl buffer, the aptamer, probe DNA, SYBR Green I, and 1 ng L⁻¹ DBP or deionized water (negative control) and was incubated at ambient temperature for 0.5, 2, 4, 6, 8, 10, and 12 h.

Temperature experiments were conducted in a manner similar to that of the incubation time experiment, as described above. The temperature experiment occurred at 2 °C, 13 °C, 25 °C, and 37 °C using a refrigerator, incubator (Wise Cube, Daihan Scientific, Gangwondo, Korea), or oven (HB-500 Minidizer™, Ultra-Violet Products Ltd, Cambridge, UK). At each incubation, the humidity was measured to determine the effect of relative humidity.

The inhibition effect of the pH and ionic strength of the Tris-HCl buffer was examined. The tested pHs were 6.0, 7.0, 9.0, and 10.0. Solutions of various pH values were prepared with Tris-HCl (pH 8.0) buffer by adding HCl (0.02 M, Sigma-Aldrich) or NaOH (0.01 M, pH 12.0, Duksan). pH was measured using a PB-10 pH meter (Sartorius Co., Göttingen, Germany).

Tris-HCl buffer as an incubation buffer included 0.02 M $\text{MgCl}_2 \cdot 6\text{H}_2\text{O}$, 0.04 M KCl, and 0.1 M NaCl. To examine the effect of ionic strength, various NaCl concentrations (0.01, 0.05, 0.1, 0.2, and 0.5 M) were added to the Tris-HCl buffer, and DBP quantification was conducted at various ionic strengths. The total ionic strengths were 0.11, 0.15, 0.2, 0.3, and 0.6 M, respectively, based on eqn (2).³⁵

$$I = \frac{1}{2} \sum C_i Z_i^2 \quad (2)$$

where C_i is the concentration of each ion species in the solution and Z_i is the charge of the ion species.

2.4.2. Environmental inhibitors. Environmental inhibitors refer to the possible residuals carried from environmental samples. Cations and humic acids were selected as environmental inhibitors mainly because of the ubiquity of these constituents in the environmental samples. Several studies^{20,24,25} have indicated that DNA hybridization for bacterial quantification is inhibited by Mg^{2+} or humic acids. Tan *et al.*³⁶ developed a biosensor technology that overcomes fluorescence quenching caused by Cu^{2+} ion.

The 1 M stock solution of cations (Mg^{2+} , Ca^{2+} , and Cu^{2+} ions) was prepared by dissolving magnesium chloride hexahydrate ($\text{MgCl}_2 \cdot 6\text{H}_2\text{O}$, 10.165 g, Daejung), magnesium sulfate heptahydrate ($\text{MgSO}_4 \cdot 7\text{H}_2\text{O}$, 12.324 g, Daejung), calcium chloride dihydrate ($\text{CaCl}_2 \cdot 2\text{H}_2\text{O}$, 14.701 g, Junsei), and copper(II) sulfate pentahydrate ($\text{CuSO}_4 \cdot 5\text{H}_2\text{O}$, 24.968 g, Daejung) in 100 mL of deionized water. The stock solutions were serially diluted to 0.01, 0.1, 1, 10, and 100 mM using deionized water. Each Mg^{2+} , Ca^{2+} , and Cu^{2+} ion concentration in 20 μL was added to the reaction. When conducting cations experiments, the added

volumes of ion solution and Tris-HCl buffer were 20 μL and 50 μL , respectively.

Humic acids (Suwannee River Humic Acid Standard II 2S101H, 200 mg) was obtained from the International Humic Substances Society (Denver, CO, USA). The humic acids stock solution was prepared by dissolving humic acids (4 mg) in 10 mL of Tris-HCl buffer (pH 8.0). After shaking overnight (shaking incubator, Wise Cube) to ensure complete dissolution, the stock solution was serially diluted to 0.001, 0.01, 0.1, 1, 10, and 100 mg L⁻¹ in Tris-HCl buffer. The 20 μL of humic acids was subjected to the reaction (200 μL total) with 50 μL of Tris-HCl buffer.

3. Results and discussion

3.1. Assay configuration of SG-aptasensor

In the developed SG-aptasensor configuration (Fig. 1A), SYBR Green I was intercalated between the base pairs of double-stranded DNA, which were formed by the aptamer and probe DNA. Once the aptamer bound to the DBP, the probe DNA and SYBR Green I dissociate owing to the conformational change of the aptamer binding with the target. SYBR Green I loses its fluorescence owing to fluorescence quenching by surrounding water molecules.⁷ The reduced signal of SYBR Green I is inversely

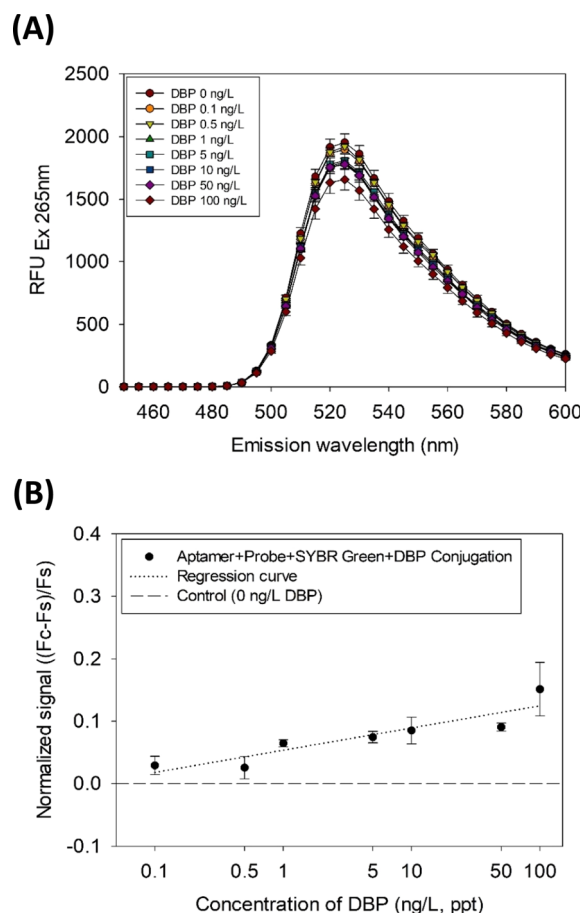


Fig. 2 (A) SYBR Green I emission spectra measured with DBP concentration variation and (B) DBP quantification results via the SG-aptasensor.



proportional to the amount of DBP. A truncated aptamer was used in this study instead of the original aptamer⁶ because of its smaller standard deviation for DBP detection (Fig. S1†).

3.2. Sensitivity and selectivity of the SG-aptasensor for DBP detection

The fluorescence intensity gradually decreased with increasing DBP concentration, as shown in the emission spectra (Fig. 2A). The range of quantification for DBP was three orders of magnitude (0.1–100 ng L⁻¹ or ppt) (Fig. 2B). In Fig. 2B, the linear regression equation is $y = 0.0357 \log x + 0.0535$ ($r^2 = 0.70$) and the limit of detection (LOD) is 0.08 ng L⁻¹ based on eqn (3a) and (3b).³⁷

$$\text{LOB} = \text{mean}_{\text{blank}} + 1.645 (\text{SD}_{\text{blank}}) \quad (3a)$$

$$\text{LOD} = \text{LOB} + 1.645 (\text{SD}_{\text{low concentration sample}}) \quad (3b)$$

Based on the limit of quantification of this study (LOQ = 0.0001 $\mu\text{g L}^{-1}$), the sensitivity of this assay is considered excellent as compared to the previous similar aptasensors for the detection of DBP, which ranges from 0.000028 to 321 $\mu\text{g L}^{-1}$ (Table 1). However, the limitation of this assay may be its lower linearity, because r^2 was similar to or lower than that in other studies, where the linearity ranged from 0.71 to 0.99.

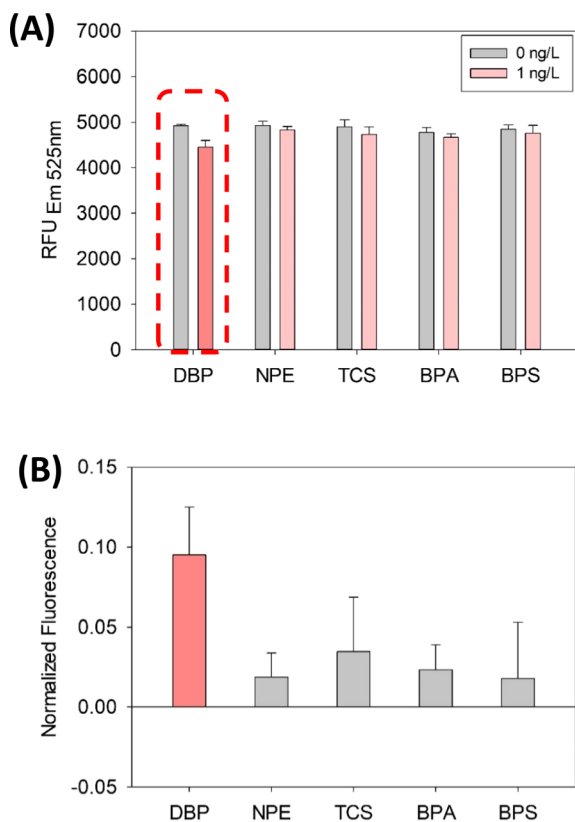


Fig. 3 Selectivity results of the SG-aptasensor for DBP detection with non-phthalate compounds: (A) fluorescence intensity and (B) normalized fluorescence.

As shown in Fig. 3, the selectivity of the SG-aptasensor is demonstrated in the presence of other endocrine-disrupting compounds (NPE, TCS, BPA, and BPS). DBP showed a significant decrease in fluorescence between 0 and 1 ng L⁻¹ DBP (dotted box in Fig. 3A, p -value = 0.0058) compared to the other four chemicals (p -values of 0.274, 0.204, 0.259, and 0.488 for NPE, TCS, BPA, and BPS, respectively) (Table S2†). The selective quantification of DBP was clearly demonstrated by normalized fluorescence (Fig. 3B, red bar for DBP).

3.3. Experimental inhibitors

The results pertaining to the experimental inhibitors (incubation time, temperature, pH, and ionic strength) are shown in Fig. 4. Various incubation times (0.5, 2, 4, 6, 8, 10, and 12 h) were tested for DBP detection using the SG-aptasensor and the results are presented in Fig. 4A and B. The fluorescence intensity for 0.5 h and 2 h are 5197 ± 1157 and 5300 ± 338 in the presence of DBP, respectively (Fig. 4A). The relatively higher standard deviations of 0.5 h and 2 h incubation indicate that the reaction might not have been completed. Afterwards, the standard deviations decreased after a 4 h incubation period, as depicted by the red arrow in Fig. 4A. Therefore, DBP detection by the SG-aptasensor requires an incubation time of at least 4 h. Additionally, a mild photobleaching effect of SYBR Green I was observed over an incubation period of 4 h (Fig. S2†). Therefore, the fluorescence decreased by ~10% after 12 h. The signal decrease over time might have been due to exposure of the fluorescent dye to light or air.³⁸

The effect of temperature on DBP detection by the SG-aptasensor was tested and is depicted in Fig. 4C and D. The temperature effect was somewhat interesting, as it did not follow the optimal conditions (*i.e.*, 37–42 °C) for general DNA hybridization. Unlike the typical pattern of higher temperatures providing better results for DNA hybridization, lower temperatures resulted in a higher normalized fluorescence in the SG-aptasensor platform (Fig. 4C). However, the individual comparison of 2 °C, 13 °C, and 25 °C each using a t -test indicates that they are not significantly different (all p -values > 0.05) (Table S3†). This indicates that the actual temperature change did not significantly influence the assay results.

Conversely, 37 °C provides the lowest normalized fluorescence as compared to the other three temperatures (Fig. 4D). After measuring the relative humidity of each temperature incubation setting, the 37 °C setting had a markedly lower relative humidity of <20%. In contrast, the humidity in the other three temperature settings ranged from 30% to 75%. As indicated in a previous study, microliter-scale volumes in microarrays are vulnerable to inadequate humidity, causing incomplete hybridization and degradation of the fluorescent dye.³⁹ The lower humidity can be the reason for the lower fluorescence signal. However, more studies may be required to elucidate the actual mechanism.

A pH range (6.0–10.0) of Tris-HCl buffer for DBP detection was tested and the results are presented in Fig. 4E and F. As shown in Fig. 4E, the fluorescence values are similar for all pH values. However, the Tris-HCl buffer with a pH lower than 7.0,



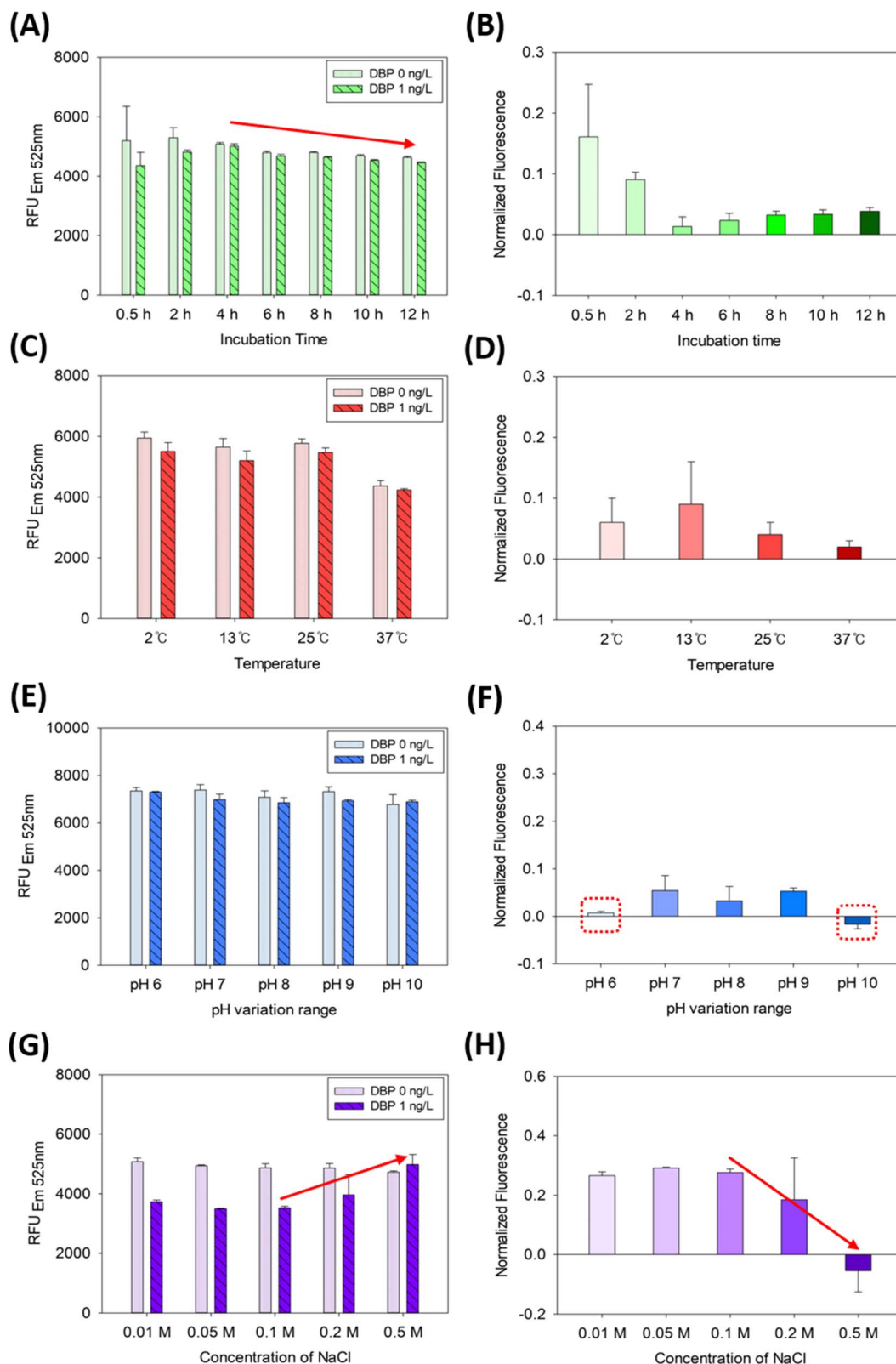


Fig. 4 DBP quantification using the SG-aptasensor in the presence of experimental inhibitors: (A) and (B) incubation time, (C) and (D) temperature, (E) and (F) pH, (G) and (H) ionic strength.

or higher than 9.0, may inhibit the assay, because the normalized fluorescence was smaller than other pH values (depicted by dotted red boxes in Fig. 4F). Acidic or basic buffers can affect the assay by either protonating or deprotonating SYBR Green I and DNA.^{40,41}

The effect of ionic strength of the Tris-HCl buffer was also investigated, as shown in Fig. 4G and H. As shown on the right side of the columns (DBP 1 ng L⁻¹, indicated by the red arrow) in Fig. 4G, the fluorescence increases corresponding to the excess NaCl concentrations added, whereas the fluorescence of the negative control (DBP 0 ng L⁻¹) is unchanged over NaCl concentrations. In the same manner, the normalized fluorescence at 0.5 M of NaCl was significantly reduced (-0.05 ± 0.07) as compared to 0.1 M of NaCl (0.27 ± 0.01) (red arrow in Fig. 4H). More than 0.2 M NaCl might have provided overstringency of the pH buffer for aptamer-DBP binding or an imbalance of charges in the solution. This result is in line with that of Hianik *et al.*⁴² DNA molecules have negatively charged phosphate groups in the backbone. At higher salt concentrations, the positively charged Na⁺ ion can preferentially bind with the negatively charged phosphate group, reducing the repulsive forces between DNA molecules and facilitating DNA hybridization.⁴³ Therefore, the effect may cause unnecessary binding between the aptamer and probe DNA, where the detachment of probe DNA is required for DBP detection.

3.4. Environmental inhibitors

The results pertaining to environmental inhibitors (divalent cations of Mg²⁺, Ca²⁺, Cu²⁺, and humic acids) are presented in Fig. 5. DBP (1 ng L⁻¹) detection in the presence of various Mg²⁺ ion showed a pattern similar to that of the negative control (0 ng L⁻¹ DBP), except for 100 mM Mg²⁺ (Fig. 5A). This is somewhat different from the results of several studies that have demonstrated a range of Mg²⁺ ion inhibition in DNA hybridization-based assays. For example, Jin *et al.*²⁰ indicated that the Mg²⁺ ion of 0.01–0.1 mM caused under-estimation of quantification, whereas the Mg²⁺ ion of 1–1000 mM caused over-estimation of quantification. Conversely, the SG-aptasensor did not demonstrate the serious inhibition in the presence of Mg²⁺. However, 100 mM Mg²⁺ ion (dotted red box in Fig. 5A) appeared to over-estimate the quantification (*i.e.*, by reducing the fluorescence signal [$\sim 13\%$] of 1 ng L⁻¹ DBP compared to the negative control of 0 mM Mg²⁺ ion). This probably is due to the specific role ('shield effect') of the Mg²⁺ ion, contributing to the secondary structure of the DNA.^{43,44} Based on the DNA folding form (mfold) calculation,⁴⁵ the Gibbs free energy (ΔG_f) of the aptamer secondary structure formation changed from -1.07 to -1.47 when 100 mM of the Mg²⁺ ion was added to the reaction as compared to the negative control (0 mM Mg²⁺ ion added). This indicates that a certain amount of Mg²⁺ ion can reinforce the formation of the secondary structure of the ssDNA aptamer, preventing it from reverting to its

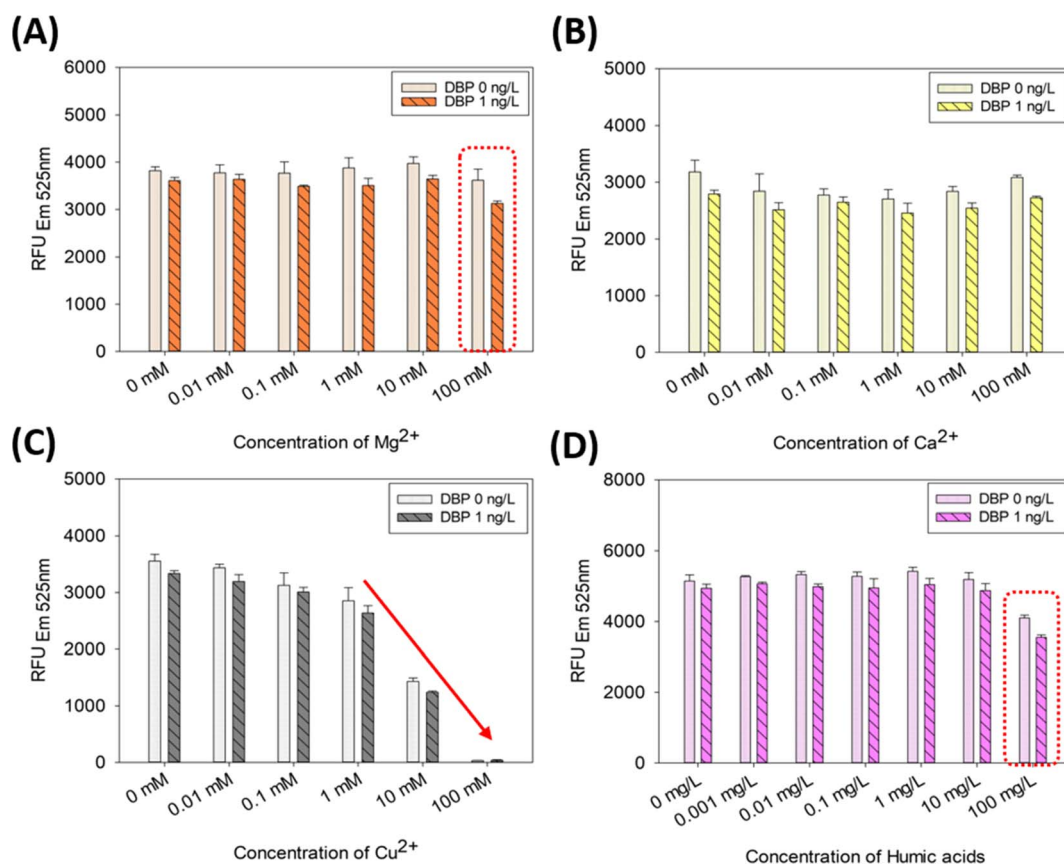


Fig. 5 DBP quantification using the SG-aptasensor in the presence of environmental inhibitors: (A) Mg²⁺ ion, (B) Ca²⁺ ion, (C) Cu²⁺ ion, and (D) humic acids.



original structure *via* electrostatic binding with negatively charged DNA. This may explain why more SYBR Green I dye was released from the aptamer-probe DNA hybrid in the presence of 100 mM Mg^{2+} . Because of the overestimation of the quantity compared to the negative control (0 mM Mg^{2+} ion), it can be considered as inhibition of the assay. However, the concentration of the Mg^{2+} ion in the environment is 4–528 mg L^{-1} , which is equivalent to 0.17–21.72 mM.^{19,46,47} Because 100 mM Mg^{2+} ion is beyond the environmentally relevant concentration, the inhibition by the Mg^{2+} ion may not be the major concern for the aptasensor-based applications.

DBP detection in the presence of various Ca^{2+} ion showed no remarkable inhibition (Fig. 5B). This result is consistent with that of Jin *et al.*²⁰ The ubiquity of Ca^{2+} ion is accentuated by its concentration in lake, river, or soil samples and ranges from 0.109 to 127 mg L^{-1} (0.0027 to 3.17 mM).^{46,47} Therefore, the concentration range of Ca^{2+} ion in the environment is acceptable for the SG-aptasensor.

DBP detection in the presence of various Cu^{2+} ion showed a dramatic inhibitive change at concentrations of 10 and 100 mM Cu^{2+} (Fig. 5C). In the presence of 100 mM Cu^{2+} , ~99% of the fluorescence intensity disappeared compared to that of the negative control (0 mM Cu^{2+}). This result can be deduced from the chemical nature of the Cu^{2+} ion. The Cu^{2+} ion is classified as a transition metal ion and an inherent fluorescence quenching ion because it suppresses the fluorescence emission by interfering with the process of the Jablonski diagram.^{36,48} In Zhao *et al.*,⁴⁹ the fluorescence was quenched approximately 88% in the presence of 100 μM Cu^{2+} ion. Furthermore, the environmentally relevant concentration range of Cu^{2+} ion was <0.033 mM in water and 0.11–64.7 mM in soils and sediments.^{50–52} Therefore, the aptasensor assay can exhibit the inhibition by the Cu^{2+} ion that is environmentally relevant, when working with real environmental samples (*e.g.*, soils and sediments).

DBP detection in the presence of various humic acids showed a significant decrease at 100 mg L^{-1} humic acids (red dotted box in Fig. 5D). The fluorescence signal decreased by 20% (without DBP) and 28% (with DBP) in the presence of 100 mg L^{-1} humic acids compared with that of the negative control (0 mg L^{-1} humic acids). The *t*-test also indicated a significant inhibition of the quantification results in the presence of 100 mg L^{-1} (*p*-values were 0.00071 without DBP and 0.00006 with DBP). In Kim *et al.*,⁵³ humic acids were found to interfere with DNA hybridization by causing random nonspecific binding between humic acids and DNA. The quantification capability of the assay was inhibited by approximately 50% by humic acids in the range of 0.001–1000 mg L^{-1} . The reduction in gene quantity was 20–50% in the presence of 100 mg L^{-1} humic acids. This result is in line with the present study, which showed a 20–28% decrease in the presence of 100 mg L^{-1} humic acids. In the previous studies regarding the occurrence of humic acids in the environment, the environmentally relevant concentration ranged from ~0.1 to 1970 mg L^{-1} .^{54–56} Therefore, 100 mg L^{-1} of humic acids is still environmentally relevant concentration and it may act as an inhibition factor of SG-aptasensor applications to environmental samples.

4. Conclusions

An SG-aptasensor was developed to detect DBP. The quantification range of DBP was 0.1–100 ng L^{-1} with a LOD of 0.08 ng L^{-1} . Environmental samples are complex and contain various constituents that can inhibit aptasensor experiments. Therefore, potential inhibition factors of the SG-aptasensor were investigated. The experimental inhibitors were tested for time, temperature, pH, and ionic strength. Our findings indicated that the assay could detect DBP in a more stable manner for at least 4 h. Various temperature ranges (2 °C, 13 °C, 25 °C) and pH buffers (7.0 to 9.0) had no significant influence. Excess ionic strength (above 0.2 M of NaCl) of the Tris-HCl buffer may cause an inhibitive increase in the fluorescence signal because SYBR Green I can bind to aggregated DNA. Environmental inhibitors such as divalent cations (Mg^{2+} , Ca^{2+} , and Cu^{2+}) and humic acids were tested. Cu^{2+} ion can significantly inhibit the assay, resulting in the reduction of 99% of the fluorescence signal in 100 mM Cu^{2+} ion, whereas the inhibition by Mg^{2+} and Ca^{2+} ion is marginal (<15%). A relatively high but environmentally relevant concentration of humic acids (100 mg L^{-1}) could also inhibit the assay. These findings underscore the importance of considering potential inhibitors when using SG-aptasensor for detecting DBP in environmental samples. The robustness of the proposed aptasensor with field samples will require further characterization in the future. This will allow us to observe its response in the presence of potential inhibitors as well as interference species.

Data availability

The data supporting this article have been included as part of the ESI.†

Author contributions

AS and HL performed the experimental design. HS performed the experimental setup and measurements. AS, HL, and HS analyzed the data. HS drafted the manuscript. AS and HL reviewed and revised the manuscript. All authors have approved the final manuscript.

Conflicts of interest

The authors declare that they have no competing financial interests or personal relationships that may have influenced this study.

Acknowledgements

This study was supported by the National Research Foundation of Korea (RS-2023-00217228) as part of the research conducted by the Convergence Research Center (CRC) at Ewha Womans University, Seoul, Korea.



Notes and references

- H. C. Erythropel, M. Maric, J. A. Nicell, R. L. Leask and V. Yargeau, *Appl. Microbiol. Biotechnol.*, 2014, **98**, 9967–9981.
- M. A. Kamrin, *J. Toxicol. Environ. Health, Part B*, 2009, **12**, 157–174.
- U. S. E. P. Agency, *Phthalates Action Plan*, 2012.
- E. C. (EC), Commission Regulation (EU) 2021/2045 of 23 November 2021 amending Annex XIV to Regulation (EC) No 1907/2006 of the European Parliament and of the Council concerning the Registration, Evaluation, Authorisation and Restriction of Chemicals (REACH) (Text with EEA relevance), European Commission, *Off. J. Eur. Union*, 2021, 6–10.
- M. J. Dennison and A. P. Turner, *Biotechnol. Adv.*, 1995, **13**, 1–12.
- H. J. Lim, H. Jin, B. Chua and A. Son, *ACS Appl. Mater. Interfaces*, 2022, **14**, 4186–4196.
- D. Kim, H. J. Lim, Y. G. Ahn, B. Chua and A. Son, *Talanta*, 2020, **219**, 121216.
- Y. Chen, Z. Wang, S. Liu and G. Zhao, *J. Hazard. Mater.*, 2021, **412**, 125174.
- R. H. Guo, C. C. Shu, K. J. Chuang and G. B. Hong, *Mater. Lett.*, 2021, **293**, 129756.
- Y. Han, D. Diao, Z. Lu, X. Li, Q. Guo, Y. Huo, Q. Xu, Y. Li, S. Cao, J. Wang, Y. Wang, J. Zhao, Z. Li, M. He, Z. Luo and X. Lou, *Anal. Chem.*, 2017, **89**, 5270–5277.
- K. Lee, N. G. Gurudatt, W. Heo, K. A. Hyun and H. I. Jung, *Sens. Actuators, B*, 2022, **357**, 131381.
- Q. Chen, M. J. Du and X. Q. Xu, *J. Electroanal. Chem.*, 2022, **914**, 116300.
- S. Wang, M. Pan, K. Liu, X. Xie, J. Yang, L. Hong and S. Wang, *Food Chem.*, 2022, **381**, 132225.
- L. Y. Yu, Y. Z. Shen, P. W. Gao, Q. Zhang, X. Y. Hu and Q. Xu, *Chem. Eng. J.*, 2023, **472**, 144925.
- D. Tu, J. T. Garza and G. L. Cote, *RSC Adv.*, 2019, **9**, 2618–2625.
- Y. W. Rong, S. Ali, Q. Ouyang, L. Wang, H. H. Li and Q. S. Chen, *J. Food Compos. Anal.*, 2021, **100**, 103929.
- H. Brammer and F. O. Nachtergaele, *Int. J. Environ. Stud.*, 2015, **72**, 56–73.
- K. R. Rogers, *Biosens. Bioelectron.*, 1995, **10**, 533–541.
- X. F. Wang, H. Kweon, S. Lee, H. Shin, B. Chua, M. R. Liles, M. K. Lee and A. Son, *Soil Biol. Biochem.*, 2018, **125**, 300–308.
- H. Jin, Y. Yoon, M. R. Liles, B. Chua and A. Son, *Analyst*, 2020, **145**, 6846–6858.
- X. Zhou, A. G. Memon, W. Sun, F. Fang and J. Guo, *Biosensors*, 2020, **11**, 6.
- Y. J. Zhan, S. T. Yang, L. F. Chen, Y. B. Zeng, L. Li, Z. Y. Lin, L. H. Guo and W. Xu, *ACS Sustainable Chem. Eng.*, 2021, **9**, 12922–12929.
- X. L. Wu, Y. Song, X. Yan, C. Z. Zhu, Y. Q. Ma, D. Du and Y. H. Lin, *Biosens. Bioelectron.*, 2017, **94**, 292–297.
- G. Y. Kim, X. Wang and A. Son, *J. Environ. Monit.*, 2011, **13**, 1344–1350.
- X. Wang, H. Kweon, S. Lee, H. Shin, B. Chua, M. R. Liles, M.-k. Lee and A. Son, *Soil Biol. Biochem.*, 2018, **125**, 300–308.
- X. Wu, Y. Song, X. Yan, C. Zhu, Y. Ma, D. Du and Y. Lin, *Biosens. Bioelectron.*, 2017, **94**, 292–297.
- Y. Li, Y. Zhu, C. Wang, M. He and Q. Lin, *Biosens. Bioelectron.*, 2019, **126**, 59–67.
- Q. R. Chen, H. H. Gao, J. L. Yao, B. Y. Jiang, R. Yuan and Y. Xiang, *Sens. Actuators, B*, 2023, **385**, 133702.
- A. Soares, B. Guieysse, B. Jefferson, E. Cartmell and J. N. Lester, *Environ. Int.*, 2008, **34**, 1033–1049.
- J. Xue, Q. Wu, S. Sakthivel, P. V. Pavithran, J. R. Vasukutty and K. Kannan, *Environ. Res.*, 2015, **137**, 120–128.
- A. R. Kim, S. H. Kim, D. Kim, S. W. Cho, A. Son and M. Y. Yoon, *Int. J. Mol. Sci.*, 2019, **21**, 208.
- A. Son, I. M. Kennedy, K. M. Scow and K. R. Hristova, *Presented in Part at the WEFTEC 2009*, 2009.
- K. B. Paul, J. M. Hedge, M. J. Devito, and K. M. Crofton, *Triclosan and Endocrine Disruption: Evidence for Alterations in Thyroid Hormone Homeostasis*, United States Environmental Protection Agency, 2007.
- H. J. Lim, E. H. Lee, S. D. Lee, Y. Yoon and A. Son, *Chemosphere*, 2018, **211**, 72–80.
- G. N. Lewis and M. Randall, *J. Am. Chem. Soc.*, 1921, **43**, 1112–1154.
- S. S. Tan, S. J. Kim and E. T. Kool, *J. Am. Chem. Soc.*, 2011, **133**, 2664–2671.
- D. A. Armbruster and T. Pry, *Clin. Biochem. Rev.*, 2008, **29**, S49–S52.
- C. L. Tiller and K. D. Jones, *Environ. Sci. Technol.*, 1997, **31**, 424–429.
- J. R. Choi, J. Hu, S. Feng, W. A. Wan Abas, B. Pingguan-Murphy and F. Xu, *Biosens. Bioelectron.*, 2016, **79**, 98–107.
- E. Tervola, K. N. Truong, J. S. Ward, A. Priimagi and K. Rissanen, *RSC Adv.*, 2020, **10**, 29385–29393.
- G. Jiang, Y. Ma, J. Ding, J. Liu, R. Liu and P. Zhou, *Chem.-Eur. J.*, 2023, **29**, e202300625.
- T. Hianik, V. Ostatna, M. Sonlajtnerova and I. Grman, *Bioelectrochemistry*, 2007, **70**, 127–133.
- K. L. Wong and J. Liu, *Biotechnol. J.*, 2021, **16**, e2000338.
- J. Anastassopoulou, *J. Mol. Struct.*, 2003, **651**, 19–26.
- M. Zuker, *Nucleic Acids Res.*, 2003, **31**, 3406–3415.
- A. Amadi, J. Yisa, I. Ogbonnaya, M. Dan-Hassan, J. Jacob and Y. Alkali, *J. Geogr. Geol.*, 2012, **4**, 13–21.
- A. Potasznik and S. Szymczyk, *J. Elem.*, 2015, **20**, 677–692.
- A. W. Varnes, R. B. Dodson and E. Wehry, *J. Am. Chem. Soc.*, 1972, **94**, 946–950.
- H. Zhao and M. L. Zastrow, *Biochemistry*, 2022, **61**, 494–504.
- Y. Liu, H. Wang, Y. Cui and N. Chen, *Int. J. Environ. Res. Public Health*, 2023, **20**, 3885.
- D. G. Heijerick, P. A. Van Sprang and A. D. Van Hyfte, *Environ. Toxicol. Chem.*, 2006, **25**, 858–864.
- T. F. Rozan and G. Benoit, *Geochim. Cosmochim. Acta*, 1999, **63**, 3311–3319.
- G. Y. Kim, X. F. Wang, H. Ahn and A. Son, *Environ. Sci. Technol.*, 2011, **45**, 8873–8880.
- M. Ali and W. Mindari, *Presented in Part at the MATEC Web of Conferences*, 2016.
- A. H. Jasim and K. A. Alghrebawi, *Dysona, Appl. sci.*, 2020, **1**, 88–95.
- L. Tripolskaja, A. Kazlauskaitė-Jadzevici, E. Baksienė and A. Razukas, *Agriculture*, 2022, **12**, 488.

