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Reusable isotype heterojunction $g\text{-C}_3\text{N}_4$ /alginate hydrogel spheres for photocatalytic wastewater treatment†

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Various visible-light-driven photocatalysts have been studied for practical applications in photocatalytic wastewater treatment *via* solar irradiation. Among them, $g\text{-C}_3\text{N}_4$ has attractive features, including its metal-free and environmentally friendly nature; however, it is prone to charge recombination and has low photocatalytic activity. To solve these problems, isotype heterojunction $g\text{-C}_3\text{N}_4$ was recently developed; however, the methods employed for synthesis suffered from limited reproducibility and efficiency. In this study, isotype heterojunction $g\text{-C}_3\text{N}_4$ was synthesized from various combinations of precursor materials using a planetary ball mill. The isotype heterojunction $g\text{-C}_3\text{N}_4$ synthesized from urea and thiourea showed the highest photocatalytic activity and completely decolorized Rhodamine B (RhB; 10 ppm) in 15 min under visible-light irradiation. Furthermore, to improve recyclability, isotype heterojunction $g\text{-C}_3\text{N}_4$ was immobilized in alginate hydrogel spheres. The isotype heterojunction $g\text{-C}_3\text{N}_4$ /alginate hydrogel beads were used in 10 repeated RhB degradation experiments and were able to maintain their initial photocatalytic activity and mechanical strength. These achievements represent an advance towards practical, sustainable photocatalytic wastewater treatment.

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1. Introduction

Sustainable, sunlight-powered photocatalytic wastewater treatment technologies have undergone considerable recent development. However, their photocatalytic efficiencies and device lifetimes remain insufficient for practical applications. TiO_2 has been thoroughly studied as a typical photocatalyst because of its strong photocatalytic activity under UV irradiation. TiO_2 features band gaps for its anatase (3.2 eV) or rutile (3.0 eV) polymorphs that require UV irradiation ($\lambda < 380$ nm) for photoexcitation;¹ however, UV radiation comprises a relatively low proportion (4%) of solar radiation.² Extensive efforts to tune TiO_2 band gaps *via* incorporation of additives have been reported. Anionic dopants, such as nitrogen and sulfur,^{3,4} grafted metal cations that induce interfacial charge transfer, such as Cu(II), Ce(III), and Fe(III),^{5–7} and dye sensitizers^{8,9} each bear various disadvantages. Alternative semiconductors with narrow band gaps may offer simpler routes to improved photocatalysts.^{10–12}

Notably, undoped graphitic carbon nitride ($g\text{-C}_3\text{N}_4$) semiconductors have an appropriate band gap range (2.6–2.8 eV)^{13–16} for efficient solar photocatalysis. $g\text{-C}_3\text{N}_4$ is synthetically accessible from various solvent-free reactions of inexpensive precursors; furthermore, $g\text{-C}_3\text{N}_4$ is metal-free, non-toxic, chemically/thermally stable and environmentally benign. However, $g\text{-C}_3\text{N}_4$ device architectures must be designed to suppress otherwise rapid electron–hole recombination. Fortunately, heterojunction device designs may improve charge separation.^{17–19} In a typical charge transfer pathway through a heterojunction (Type II), excited electrons move to a semiconductor with a relatively lower conduction band, whereas holes move to another semiconductor with a higher valence band, thus enhancing charge separation.²⁰ Previous efforts to produce $g\text{-C}_3\text{N}_4$ heterojunctions have involved the complex, prolonged, and wasteful synthesis of composites that include metal oxides. More recent reports describe one-step synthetic routes to $g\text{-C}_3\text{N}_4$ heterojunction compositions that exclude metal oxides. Solehudin *et al.* and Liao *et al.* synthesized isotype heterojunction $g\text{-C}_3\text{N}_4$ by calcining a mortar-crushed mixture of urea and melamine;^{21,22} however, manual mixing methods are challenging to reproduce. Balakrishnan *et al.* developed citric acid-assisted chemically functionalized black $g\text{-C}_3\text{N}_4$ isotype heterojunction hydrogels for tetracycline degradation and H_2O_2 production and demonstrated excellent photocatalytic decomposition performance for tetracycline degradation.²³ Dong *et al.* reported isotype heterojunction $g\text{-C}_3\text{N}_4$ by calcining precursors deposited from a dried solution of urea and thiourea,²⁴

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unfortunately, solvent requirements restrict mass producibility in this case. Further investigation is required to elucidate the relationship between precursor composition and photocatalytic activity of isotype heterojunction $\text{g-C}_3\text{N}_4$. Balakrishnan *et al.* also developed phosphorylated $\text{g-C}_3\text{N}_4$ /sulfur-self-doped- $\text{g-C}_3\text{N}_4$ homojunction carboxymethyl cellulose beads for highly efficient photocatalytic generation of H_2O_2 under visible light.²⁵ The homo-junction $\text{g-C}_3\text{N}_4$ is characterized by a simple manufacturing process due to the use of the same material, high thermal and chemical stability for long-term use, and limited charge carrier separation efficiency compared to heterojunctions, leading to less effective recombination suppression. Therefore, we focused on isotype heterojunction $\text{g-C}_3\text{N}_4$ in this study.

For practical applications, photocatalysts must be sufficiently reusable; however, device designs have required trade-offs between reusability and performance. Increased photocatalyst surface area facilitates efficient contact with the degradation target, leading to increased photocatalytic reaction rates. Accordingly, recent reports have identified photocatalysts with high specific surface area nanostructures.^{11,26,27} Unfortunately, such photocatalyst particles have lower compatibility with solid-liquid separation methods, diminishing reusability and foreshadowing environmental nanopollution.²⁸ Photocatalysts may be recovered using laboratory-scale centrifugation; however, wastewater treatment facilities are not equipped for continuous centrifugation processes. Reassuringly, photocatalyst immobilization techniques may improve high surface area photocatalyst recovery.

Alginate is a natural, biodegradable and environmentally benign polysaccharide extractable from brown algae. Alginate materials have been applied as food additives,²⁹ drug delivery carriers,³⁰ dye or heavy metal ion adsorbents,³¹ and polymer flocculant.^{32,33} Alginate polymers consist of various proportions and distributions of 1,4-linked β -D-mannuronic acid and α -L-guluronic acid. Alginate carboxyl groups interact electrostatically with multivalent cations to instantaneously form hydrogels, which have been applied to immobilize biocatalysts or enzymes. Alginate biocatalyst immobilization has been demonstrated using two approaches that impart different macrostructures. Alginate beads may be formed by addition of a catalyst-containing alginate solution a crosslinker.³³ Alginate capsules with a hard outer shell and viscous core were synthesized *via* addition of a viscous solution containing the crosslinker and catalyst into an alginate solution.³⁴ By encapsulating $\text{g-C}_3\text{N}_4$ in hydrogel beads or capsule, the operability of $\text{g-C}_3\text{N}_4$ in practical and large-scale applications is improved and possible nanopollution is prevented, increasing the potential for commercialization.^{25,35} Although two such cases of $\text{g-C}_3\text{N}_4$ immobilization in alginate beads and capsules have been reported, their photocatalytic activities are low and further improvement is required to apply for practical application.^{36,37}

We hypothesized that reproducible $\text{g-C}_3\text{N}_4$ heterojunctions could be practically synthesized, immobilized in alginate, and applied as durable and efficient photocatalysts. We first combined various sets of solid precursor materials using planetary ball milling to form $\text{g-C}_3\text{N}_4$ heterojunctions. We then

evaluated their photophysical properties using X-ray diffraction (XRD), Fourier-transform infrared spectroscopy (FT-IR), and UV-vis diffuse reflectance spectroscopy. Additional properties, such as Brunauer–Emmett–Teller (BET) specific surface area, hydrodynamic diameter, and zeta potential, were also evaluated. Then, we evaluated their visible light-powered photocatalytic activity for the degradation of Rhodamine B (RhB). We then evaluated the activity and durability of solid bead or hollow capsule photocatalysts containing alginate hydrogel sphere-immobilized $\text{g-C}_3\text{N}_4$.

2. Material and methods

2.1. Materials

The following reagents were purchased from commercial sources and used without further purification. Guaranteed reagent-grade urea, thiourea, calcium chloride anhydrate (CaCl_2), and ammonium oxalate monohydrate (AO); Wako special-grade melamine and *p*-benzoquinone (BQ); and Wako 1st-grade sodium alginate (80–120 cP; 10 g L⁻¹, 20 °C) were purchased from FUJIFILM Wako Pure Chemical (Osaka, Japan). Polyvinylpyrrolidone (PVP; Ave. M.W. 1 300 000) was purchased from Sigma-Aldrich (St. Louis, MO, USA). Chemical usage-grade rhodamine B was purchased from JUNSEI CHEMICAL (Tokyo, Japan). Aerioxide® P25 was purchased from Evonic Industries AG (Essen, NRW, Germany). Liquid chromatography-grade 2-propanol (IPA) was purchased from Merck KGaA (Darmstadt, HE, Germany).

2.2. Synthesis of $\text{g-C}_3\text{N}_4$

Isotype heterojunction $\text{g-C}_3\text{N}_4$ was synthesized by a modified literature procedure.^{21,24} In a representative synthetic procedure, urea and thiourea (1 : 1 by wt.) were placed in a zirconia bowl (45 mL) with zirconia balls (15 mm × 7) and mixed using a planetary ball mill (Pulverisette 7 Classic Line, Fritsch, Idar-Oberstein, RP, Germany; 15 min; 300 rpm). The obtained powder was transferred to an alumina crucible with a lid and heated under air (5 °C min⁻¹) to the desired calcination temperature (550 °C; 2 h). The resultant $\text{g-C}_3\text{N}_4$ sample (CN-TU) was cooled (25 °C) and collected. Analogous syntheses using melamine and urea are denoted as CN-MU. Non-heterojunction $\text{g-C}_3\text{N}_4$ samples were synthesized separately from single components (urea, CN-U; thiourea, CN-T; and melamine, CN-M).

2.3. Synthesis of $\text{g-C}_3\text{N}_4/\text{Alg}$ beads

CN-TU (1% w/v) was added to ultrapure water and sonicated (400 W, 20 kHz, 30 min). Then, sodium alginate (1% w/v) was added, and the resultant heterogeneous mixture was vigorously stirred and placed in a syringe with a 26G (0.23 mm i.d.) needle. Using a syringe pump, the mixture was added dropwise into a stirring (500 rpm; 20 min) aqueous solution of CaCl_2 (0.1 M), after which the beads were fully crosslinked. Finally, the obtained $\text{g-C}_3\text{N}_4/\text{Alg}$ beads were recovered by straining through a colander and washed by stirring in ultrapure water (30 min).



2.4. Synthesis of g-C₃N₄/Alg capsules

CN-TU (2% w/v) was sonicated in aqueous in CaCl₂ (0.5 M) (30 min). PVP (8% w/v) was then added as a thickening agent and dissolved *via* vigorous stirring. Using a syringe pump, the resultant heterogeneous mixture was placed in a syringe with a 26G needle and added dropwise to a stirring (700 rpm; 5 min) aqueous solution of sodium alginate (0.5% w/v), after which the capsules were fully crosslinked. Finally, the obtained g-C₃N₄/Alg capsules were recovered by straining through a colander and washed by stirring in ultrapure water (30 min).

2.5. Characterization of g-C₃N₄

The phase structures of the powder samples were analyzed by XRD using a Bruker D8 ADVANCE diffractometer (Billerica, MA, USA) with a Cu K α ($\lambda = 1.5418 \text{ \AA}$) radiation source (40 kV; 40 mA). Diffraction patterns were obtained over a scan range ($2\theta = 10^\circ$ – 60° ; step width, 0.1° ; scan rate, 0.5° s^{-1}). Sample chemical compositions were identified by FT-IR using a Shimadzu (Kyoto, Japan) IRAffinity-1S spectrophotometer (resolution, 1 cm^{-1} ; cumulative number, 20). The samples were prepared as KBr tablets. Diffuse reflection spectra of the samples were measured using a JASCO (Tokyo, Japan) V-650 UV-vis spectrometer equipped with an integrating sphere. BaSO₄ was used as the standard. Tauc plots were analyzed to obtain sample band gaps. The photoluminescence spectra were measured using a Shimazu (Tokyo, Japan) RF-5300PC spectrophotometer to assess the recombination of the photo-generated holes and electrons of the different samples.

Nitrogen adsorption–desorption isotherms (-196°C) were measured using a Micromeritics (Norcross, GA, USA) ASAP 2020 after sample degassing (250°C ; 3 h). The specific surface areas of the samples were analyzed using the BET multi-point method. Sample hydrodynamic diameters were measured by dynamic light scattering (DLS) using a Beckman Coulter (Brea, CA, USA) DelsaMax Pro. Simultaneously, sample electrophoretic mobilities in ultrapure water were measured, and the zeta potential was determined based on the following Smoluchowski (eqn (1)):

$$v = \varepsilon \zeta E / \mu \quad (1)$$

where v is the electrophoretic mobility, ε is the dielectric constant of the solvent, ζ is the zeta potential, E is the electric field, and μ is the viscosity of the solvent.

2.6. Photocatalytic degradation experiments

The photocatalytic activity of the samples was evaluated based on the visible-light degradation of RhB. To achieve adsorption equilibrium between the photocatalyst and the degradation targets, the photocatalytic sample was added to aqueous RhB (10 ppm) and stirred (30 min) in the dark. Equivalent masses (0.1 g) of the as-prepared g-C₃N₄ powder (0.1 g), g-C₃N₄/Alg bead (1000 pieces), or g-C₃N₄/Alg capsule (500 pieces) photocatalysts were added. The samples were irradiated with visible light ($\lambda > 440 \text{ nm}$) using a USHIO (Tokyo, Japan) UXR-300BF xenon lamp (300 W) equipped with a UV cut-off filter. The RhB

concentration was monitored by measuring the absorbance ($\lambda = 553 \text{ nm}$) using a JASCO V-650 UV-vis spectrophotometer and plotted as the normalized concentration c_t/c_0 [%]. For powder samples, particles were removed by centrifugation (13 000 rpm; 30 min) prior to absorbance measurements. To evaluate photocatalytic activity, pseudo-first-order reaction rate constants were calculated using eqn (2):

$$-\ln(c_t/c_0) = kt \quad (2)$$

where c_t is the RhB concentration at irradiation time t , c_0 is the initial concentration of RhB, and k is the rate constant.

2.7. Active species trapping experiments

Active species trapping experiments were conducted to identify the active species contributing to RhB degradation. IPA, BQ, and AO (7.5 mM) were used as scavengers for $\cdot\text{OH}$, $\cdot\text{O}_2^-$, and the holes, respectively. These experiments were carried out using the same procedure as in Section 2.6, except for the addition of scavengers.

2.8. Recycling tests of g-C₃N₄/Alg beads

To evaluate the durability of recycled g-C₃N₄/Alg beads, RhB degradation experiments described in section 2.6 were repeated 10 times, with an additional bead washings step in between each cycle involving stirring (30 min) the beads in ultrapure water to remove residual adsorbed RhB. Scanning electron microscopy (FE-SEM) using a JEOL (Tokyo, Japan) JSM-7500M microscope (acceleration voltage, 5 kV; emission current, 10 μA) imaged the g-C₃N₄/Alg bead surfaces before and after the tests. Microscopy samples were vacuum freeze-dried and sputter-coated with osmium, an antistatic agent. In addition, the compressive strength of the beads was measured using an A&D (Tokyo, Japan) STB-1225S universal testing machine equipped with a load cell (50 N). The elastic modulus was determined based on the data within 0–20% of the strain range by using the following Hertz model (eqn (3)):

$$F = 4/3ED^{0.5}\Delta D^{1.5} \quad (3)$$

where F is the load, E is the modulus of elasticity, D is the bead diameter, and ΔD is the displacement.

3. Results and discussion

3.1. Characterization

Fig. 1 shows similar XRD patterns for each sample. A stronger high angle signal (27.4° , 002) is attributable to the interlayer stacking of the aromatic units of g-C₃N₄, whereas the smaller low angle signal (13.3° , 002) to the repetition of the triazine units in the plane.³⁸ These data are consistent with successful syntheses of g-C₃N₄ using each set of conditions. In addition, CN-U and CN-T exhibited smaller and broader peaks than CN-M, indicative of their lower crystallinity. Plausibly, their urea and thiourea precursor materials, which are known to generate various intermediates in the calcination process, disrupted the formation of uniform and well-connected heptazine-based g-



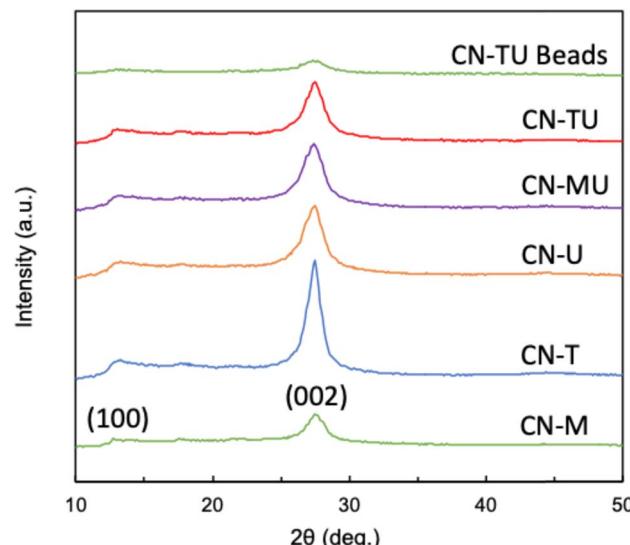


Fig. 1 X-ray diffraction patterns of $\text{g-C}_3\text{N}_4$ synthesized from different raw materials.

C_3N_4 .³⁹ In contrast, melamine contains a triazine subunit, which could promote polymerization, leading to the relatively higher crystallinity for CN-M. The result of XRD spectra for CN-TU encapsulated in alginate is also shown in Fig. 1. Although peak intensity is low, the peaks at 27.4° and 13.3° indicate the successful formation of $\text{g-C}_3\text{N}_4/\text{Alg}$ beads.

Fig. 2 shows similar FTIR spectra for each sample; material differences for each type of $\text{g-C}_3\text{N}_4$ did not impart spectral differences. Furthermore, the spectral data are typical for $\text{g-C}_3\text{N}_4$, suggesting successful syntheses. The sharpest peak (800 cm^{-1}) was attributed to the breathing vibration of the triazine unit.^{21–24,38} Two peaks ($1200\text{--}1350\text{ cm}^{-1}$) are attributed to the out-of-plane bending vibration of the heteroaromatic ring.⁴⁰ Multiple peaks ($1400\text{--}1700\text{ cm}^{-1}$) are attributed to the

$\text{C}=\text{N}$ stretching vibration of the heteroaromatic ring.²² A broad peak ($3000\text{--}3650\text{ cm}^{-1}$) is attributed to $\text{N}-\text{H}$ and $\text{O}-\text{H}$ stretching vibrations, suggesting the presence of uncondensed amine groups and adsorbed water molecules on the particle surface.^{24,40} In addition, no unique peaks were observed for CN-TU and CN-T synthesized from raw materials containing sulfur atoms, suggesting that sulfur was desorbed as SO_x during calcination. The result of FTIR spectra for CN-TU encapsulated in alginate is also shown in Fig. 2. Despite low absorption due to the dilution of $\text{g-C}_3\text{N}_4$ by alginate, successful encapsulation of $\text{g-C}_3\text{N}_4$ in alginate is confirmed.

Fig. 3 shows the UV-vis diffuse reflectance spectra of each sample. As shown in Table S.1†, the band gaps of each sample are varied but within the visible range. The absorption edges of all samples ($450\text{--}480\text{ nm}$) are appropriate for high reactivity under visible-light irradiation. Zhang *et al.* reported that they successfully improved the visible-light absorption of $\text{g-C}_3\text{N}_4$ by adding different amounts of barbituric acid (BA) to the raw dicyandiamide material.⁴¹ In their study, the C/N ratio of $\text{g-C}_3\text{N}_4$ increased with increasing BA content, narrowing the band gap. Therefore, differences in C/N ratio, plausibly originating from differences in precursor materials, may be responsible for the varied band gaps of the $\text{g-C}_3\text{N}_4$ samples. For the samples obtained from blends of two precursor materials, the band gap of CN-TU was located between those of CN-T and CN-U, and CN-MU, between CN-M and CN-U. The optical properties of the prepared $\text{g-C}_3\text{N}_4$ are consistent with previous reports,^{21,24} suggesting the formation of isotype heterojunctions in CN-TU and CN-MU.

The photoluminescence (PL) spectra of the samples were analyzed to examine the recombination of the photo-generated holes and electrons and the results are shown in Fig. 4.²³ The $\text{g-C}_3\text{N}_4$ samples prepared using a single starting material such as CN-T, CN-U, and CN-M show higher intensity meaning higher recombination. In contrast, the $\text{g-C}_3\text{N}_4$ samples prepared using two starting materials such as CN-TU and CN-MU show

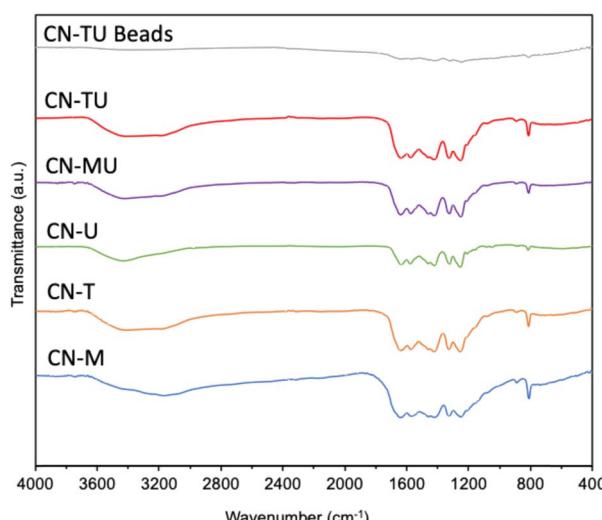


Fig. 2 FT-IR spectra of $\text{g-C}_3\text{N}_4$ synthesized from different raw materials.

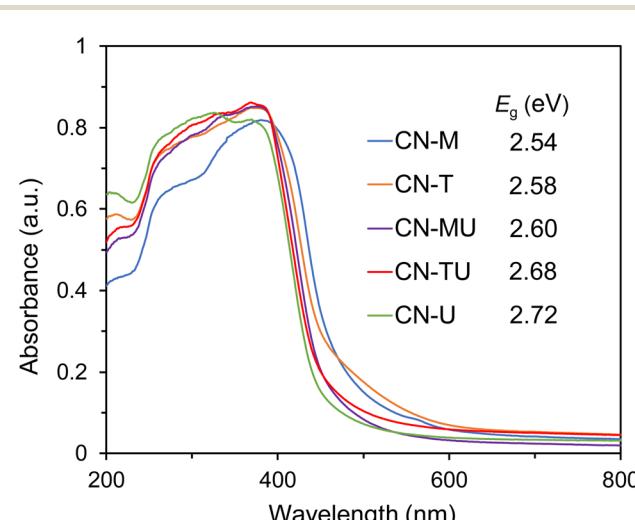


Fig. 3 UV-vis diffuse reflectance spectra and band gap energies of $\text{g-C}_3\text{N}_4$ synthesized from different raw materials.



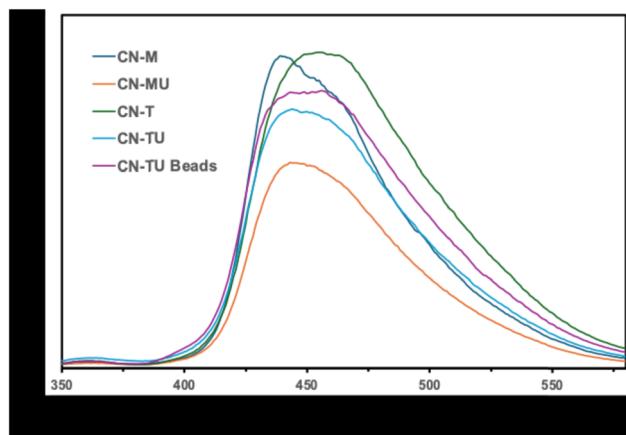


Fig. 4 Photoluminescence (PL) of $\text{g-C}_3\text{N}_4$ synthesized from different raw materials.

relatively lower intensity meaning suppressed recombination. The results also support the formation of hetero type.

Table S1† shows significant differences in BET specific surface areas for each set of precursor materials. Urea and thiourea precursors, which release soft bubbles of NH_3 , H_2S , and CO_2 gases during the polymerization process, produced more porous $\text{g-C}_3\text{N}_4$.⁴² Moreover, the relatively small particle size (hydrodynamic diameter) of CN-U suggests that urea generates $\text{g-C}_3\text{N}_4$ with a lower degree of polymerization.

The surface charge of photocatalytic particles affects the adsorption properties of the degradation targets. The isoelectric point of $\text{g-C}_3\text{N}_4$ has been reported to be between pH 4.4–5.1, with a negative zeta potential at higher pH.^{40,43} Table S1† shows zeta potentials in ultrapure water that are consistent with previous reports and suggest that $\text{g-C}_3\text{N}_4$ selectively absorbs cationic compounds.

3.2. Photocatalytic performance of $\text{g-C}_3\text{N}_4$

Photocatalytic degradation experiments were performed under visible-light irradiation using RhB (10 ppm) as a representative pollutant, and compared against P25, the best-known commercial TiO_2 photocatalyst. Fig. 5 shows the temporal changes in the RhB concentration; the gray area represents the time period prior to visible-light irradiation whilst RhB was adsorbed on the photocatalyst surface in the dark. The RhB concentration decreased during dark period before the beginning of irradiation ($t = 0$), indicating that $\text{g-C}_3\text{N}_4$ is highly adsorptive for RhB. The high adsorptivity may originate from electrostatic interactions between the negatively charged $\text{g-C}_3\text{N}_4$ surface and the RhB cationic dye. Furthermore, all $\text{g-C}_3\text{N}_4$ samples showed higher activity than P25. The CN-TU-forming isotype heterojunction exhibited the highest photocatalytic activity and completely decolorized RhB in 15 min.

Fig. 6 shows the relationship between sample BET specific surface area and photocatalytic activity. The photocatalytic activities of non-heterojunction CN-U, CN-T, and CN-M were proportional with BET specific surface area. The plots for CN-TU and CN-MU, which were formed with isotype heterojunctions,

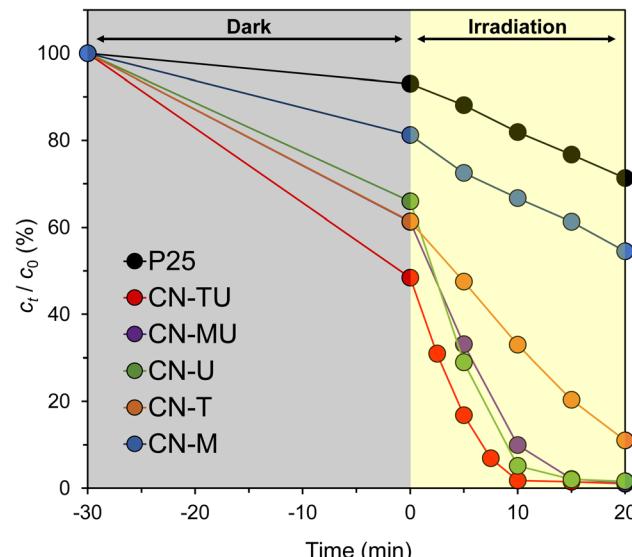


Fig. 5 RhB degradation by $\text{g-C}_3\text{N}_4$ synthesized from different raw materials under visible light irradiation.

were higher than would be predicted by the linear relationship for the non-heterojunction samples. Plausibly, these samples featured isotype heterojunctions that improved charge separation, offering improved activity beyond what their surface areas could offer otherwise. CN-MU showed lower photocatalytic activity than CN-U, possibly because the negative effect of the low specific surface area derived from melamine exceeded the positive effect of improved charge separation. These data suggest that urea and thiourea are the optimal precursor materials for the synthesis of isotype heterojunction $\text{g-C}_3\text{N}_4$ with large surface areas.

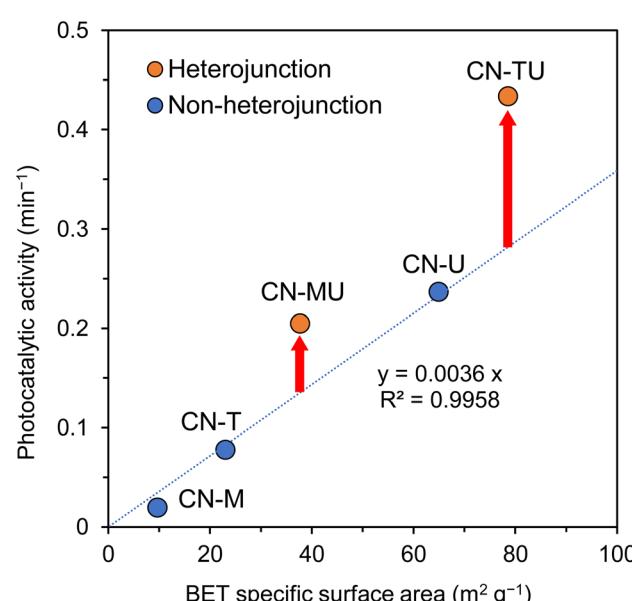
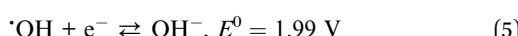
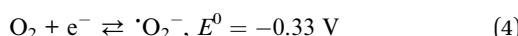


Fig. 6 Relationship between BET specific surface area and photocatalytic activities.



3.3. Active species trapping experiments

To identify the active species generated by $\text{g-C}_3\text{N}_4$, active species trapping experiments were performed using CN-TU and appropriate scavengers. Fig. 7 shows BQ, a $\cdot\text{O}_2^-$ scavenger, inhibited the degradation reaction to the greatest extent. AO, a hole scavenger, offered slightly decreased inhibition, whereas IPA, an $\cdot\text{OH}$ scavenger, provided minimal inhibition. These data suggest that $\cdot\text{O}_2^-$ and h^+ are involved in RhB degradation by CN-TU. Dong *et al.* reported that isotype heterojunction $\text{g-C}_3\text{N}_4$ synthesized from urea and thiourea had a valence band and conduction band levels of 1.76 and -0.82 eV, respectively.²⁴ Eqn (4) describes the generation of $\cdot\text{O}_2^-$ by the reduction of oxygen, whereas eqn (5) and (6) describe the generation of $\cdot\text{OH}$ by the oxidation of OH^- and H_2O , respectively.



Therefore, CN-TU has sufficient reduction potential to generate $\cdot\text{O}_2^-$, but insufficient oxidation potential to generate $\cdot\text{OH}$. Plausibly, $\cdot\text{O}_2^-$ generated by the reduction of dissolved oxygen by excited electrons or direct oxidation by holes is the source of RhB degradation.

3.4. Photocatalytic performance of $\text{g-C}_3\text{N}_4$ /Alginate hydrogel spheres

We attempted to improve the recovery and reusability of CN-TU, which had the highest photocatalytic activity in powder form, by immobilizing it in alginate hydrogel spherical capsules or beads. These hydrogel spheres are composed of environmentally benign natural polysaccharides and calcium ions. Alginate

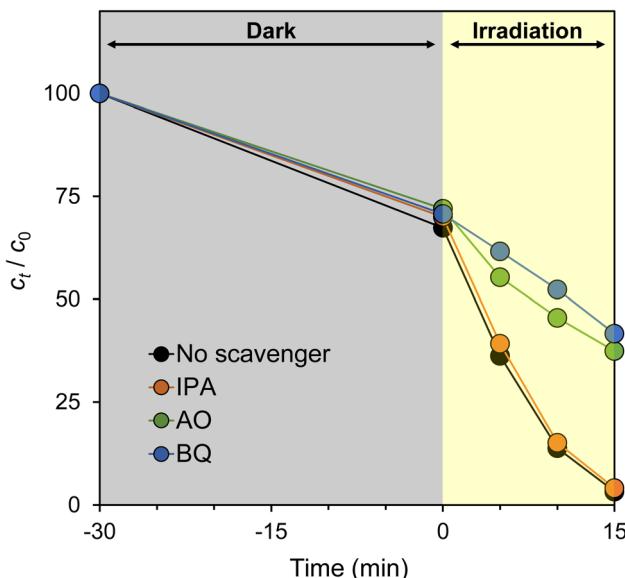


Fig. 7 Active species trapping experiment using different scavengers.

hydrogel spheres were synthesized by one-step methods yielding $\text{g-C}_3\text{N}_4$ /Alg beads or hollow $\text{g-C}_3\text{N}_4$ /Alg capsules. The microscopic images in Fig. 8 reveal that relative to the transparent alginate beads (Fig. 8a), $\text{g-C}_3\text{N}_4$ /Alg beads (Fig. 8b) contained uniformly dispersed $\text{g-C}_3\text{N}_4$ throughout the hydrogel. For the $\text{g-C}_3\text{N}_4$ /Alg capsules (Fig. 8c), $\text{g-C}_3\text{N}_4$ was encapsulated in the core surrounded by an alginate shell. The diameters of the $\text{g-C}_3\text{N}_4$ /Alg beads and $\text{g-C}_3\text{N}_4$ /Alg capsules were 2.43 ± 0.10 and 4.42 ± 0.06 mm, respectively. The thickness of the alginate shell in the $\text{g-C}_3\text{N}_4$ /Alg capsules was 0.33 ± 0.08 mm. In the capsules, Ca^{2+} in the droplets cross-linked the alginate while diffusing outward from the droplet owing to the concentration gradient. In contrast, in the case of beads, cross-linking occurs by Ca^{2+} entering into alginate droplets from the outer surface to the inside. Consequently, the diameter of the capsules was larger than that of the beads despite using the same injection needle. These hydrogel spheres were well-dispersed in stirring water and immediately settled when the stirring was halted.

To discuss recovery of photocatalyst in more detail, we compared the terminal settling velocity of $\text{g-C}_3\text{N}_4$ /Alg beads and powdered $\text{g-C}_3\text{N}_4$ in water. The former was determined by measuring the actual time taken for the beads to settle for a fixed depth, while the latter was calculated from the median diameter of CN-TU (1.2 μm , Table S1†) by using the following Stokes' (eqn (7)):

$$v_s = g(\rho_p - \rho_f)d_p^2/18\eta \quad (7)$$

where v_s is the terminal settling velocity, g is the gravitational acceleration, ρ_p is the density of the particles, ρ_f is the density of water, d_p is the particle size, and η is the viscosity of water. As a result, the terminal settling velocity of $\text{g-C}_3\text{N}_4$ /Alg beads was measured to be 1.36 cm s^{-1} , while that of powdered $\text{g-C}_3\text{N}_4$ was calculated to be 0.11 cm s^{-1} . Therefore, the terminal settling

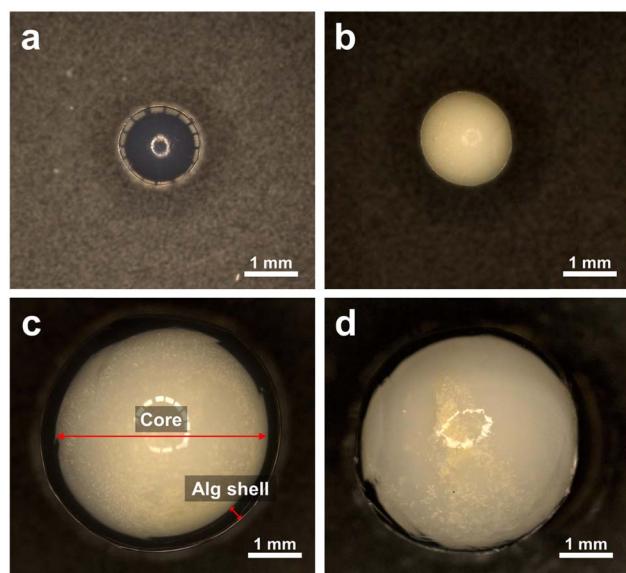


Fig. 8 Microscopic images of (a) an alginate bead, (b) a $\text{g-C}_3\text{N}_4$ /Alg bead, (c) a $\text{g-C}_3\text{N}_4$ /Alg capsule and (d) agglomeration of $\text{g-C}_3\text{N}_4$ in the capsule.



velocity of $\text{g-C}_3\text{N}_4/\text{Alg}$ beads is at least 12 times faster than that of the powder. At this velocity, in the separation process after pollutant degradation, the $\text{g-C}_3\text{N}_4/\text{Alg}$ beads can be easily separated from treated water in a very short period of standing time. Whilst since $1.2\ \mu\text{m}$ is merely the average particle diameter of powdered $\text{g-C}_3\text{N}_4$ and much smaller size particles are also included, it requires centrifugation with 13 000 rpm for 30 min to separate them from the media completely. These results demonstrate that immobilization of $\text{g-C}_3\text{N}_4$ in alginate beads improves recovery.

Fig. 9 summarizes the photocatalytic activity of powdered CN-TU, CN-TU immobilized in Alg beads ($\text{g-C}_3\text{N}_4/\text{Alg}$ beads), and CN-TU immobilized in Alg capsules ($\text{g-C}_3\text{N}_4/\text{Alg}$ capsules). A control experiment using Alg beads not containing $\text{g-C}_3\text{N}_4$ demonstrated a negligible decrease in RhB concentration, suggesting RhB is stable under visible light irradiation and that the alginate hydrogel negligibly adsorbs RhB. Therefore, the decrease in RhB concentration may be attributed to photocatalytic degradation in the experiments involving $\text{g-C}_3\text{N}_4/\text{Alg}$ beads and $\text{g-C}_3\text{N}_4/\text{Alg}$ capsules. As expected, the degradation rate of immobilized $\text{g-C}_3\text{N}_4$ was inferior to the powder because RhB diffusion is hindered inside the alginate hydrogel. Comparing the two types of immobilization methods, $\text{g-C}_3\text{N}_4/\text{Alg}$ beads completely decolorized RhB (10 ppm) in 60 min, whereas $\text{g-C}_3\text{N}_4/\text{Alg}$ capsules decolorized only 60%. Photocatalytic activities of the $\text{g-C}_3\text{N}_4/\text{Alg}$ beads and capsules were evaluated by a first-order kinetic constant, and were $0.049\ \text{min}^{-1}$ and $0.012\ \text{min}^{-1}$, respectively. The photocatalytic activity of the $\text{g-C}_3\text{N}_4/\text{Alg}$ beads is approximately 4 times higher than that of the $\text{g-C}_3\text{N}_4/\text{Alg}$ capsules. The capsules seemed to have a lower mass transfer resistance than the beads because the alginate layer is thinner for the $\text{g-C}_3\text{N}_4/\text{Alg}$ capsules than for the $\text{g-C}_3\text{N}_4/\text{Alg}$ beads (shell thickness, $0.3\ \text{mm}$; bead radius, $1.2\ \text{mm}$). However, increased mass transfer resistance in the capsule may originate from their larger diameters ($4.42\ \text{mm}$, vs. $2.43\ \text{mm}$ for beads)

and highly viscous, PVP solution-filled cores. Furthermore, Fig. 8d shows an example of agglomeration of $\text{g-C}_3\text{N}_4$ particles inside a capsule sample; this was observed in some but not all capsule samples. Thus, decreased contact area with RhB could decrease photocatalytic activity of $\text{g-C}_3\text{N}_4/\text{Alg}$ capsules. In contrast, the $\text{g-C}_3\text{N}_4/\text{Alg}$ beads showed higher photocatalytic activity because the immobilized particles did not agglomerate. These results confirmed that solid alginate hydrogel beads are more suitable as carriers for the immobilization of $\text{g-C}_3\text{N}_4$. In previous studies in which $\text{g-C}_3\text{N}_4$ was immobilized in alginate hydrogels, even under optimal conditions, the degradation of RhB (10 ppm) took $3\text{--}36\ \text{h}$.^{33,34} Neither study formed heterojunctions in $\text{g-C}_3\text{N}_4$, which did not improve the charge separation problem. Hao *et al.* added a higher concentration of $\text{g-C}_3\text{N}_4$ (15% w/v) to the alginate solution,³⁶ which would have resulted in a higher photocatalyst content per bead and poor light penetration inside the beads. This means that the self-shading effect of the photocatalyst could decrease in photocatalytic activity. Falletta *et al.* had a larger bead diameter ($5\ \text{mm}$),³⁷ which could decrease diffusion rate of the RhB solution. In contrast, isotype heterojunction $\text{g-C}_3\text{N}_4$ with improved charge separation were immobilized at a low concentration (1% w/v) in alginate beads with smaller diameter ($2.43\ \text{mm}$) in this study, suggesting high activity of $\text{g-C}_3\text{N}_4$ itself, high light penetration, and low mass transfer resistance may explain our improved photocatalytic performance. In addition, both previous studies were carried out using xenon lamp irradiation without eliminating the UV light range in their experiments. Nonetheless, the $\text{g-C}_3\text{N}_4/\text{Alg}$ beads in this study demonstrated improved photocatalytic activity, as evidenced by the complete decolorization of RhB (10 ppm) in 60 min using only visible light.

3.5. Recycling tests of $\text{g-C}_3\text{N}_4/\text{Alg}$ beads

Recycling tests were performed to prove the practicality of the $\text{g-C}_3\text{N}_4/\text{Alg}$ beads. Fig. 10a shows changes of RhB concentration over time during exposure to visible-light irradiation, and Fig. 10b shows the photocatalytic activity of $\text{g-C}_3\text{N}_4/\text{Alg}$ beads during 10 repeated degradation cycles. Interestingly, the photocatalytic activity improved in the 2nd run; in the subsequent runs, RhB was completely decolorized within 45 min. This enhancement could originate from slight photocatalytic degradation of the carrier alginate hydrogel, resulting in pore widening or channel formation that facilitated improved RhB diffusion rate inside the beads.⁴⁴ As shown in the FE-SEM images (Fig. 11a and b), the surface of the beads immediately after synthesis was smooth (Fig. 11a), whereas wrinkles were observed on the surface of the beads after recycling (Fig. 11b). This observation suggests that the degradation of the alginate hydrogel inside the beads created a slight cavity that caused nonuniform shrinkage during drying. Moreover, the consistent photocatalytic activity throughout the 10 repeated cycles suggests that the $\text{g-C}_3\text{N}_4$ particles did not leak from the beads. This represents an improvement from previous studies using *N,N*-dimethylacrylamide or chitosan hydrogels as carriers, which demonstrated decreased degradation rates over the course of recycling experiments. Those materials may have

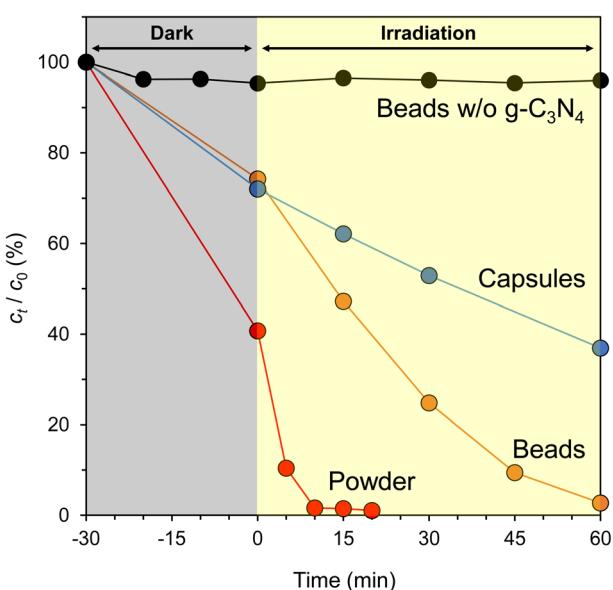


Fig. 9 RhB degradation by two types of $\text{g-C}_3\text{N}_4/\text{Alg}$ hydrogel spheres.



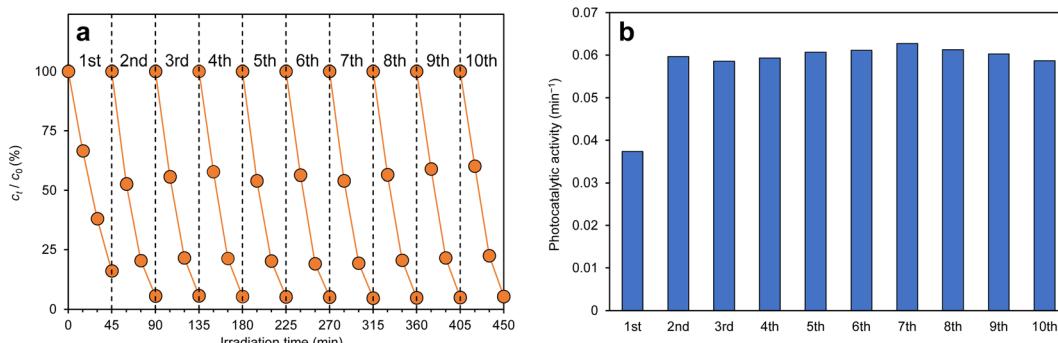


Fig. 10 (a) Temporal changes in RhB concentration and (b) photocatalytic activity in recycling tests of $\text{g-C}_3\text{N}_4/\text{Alg}$ beads.

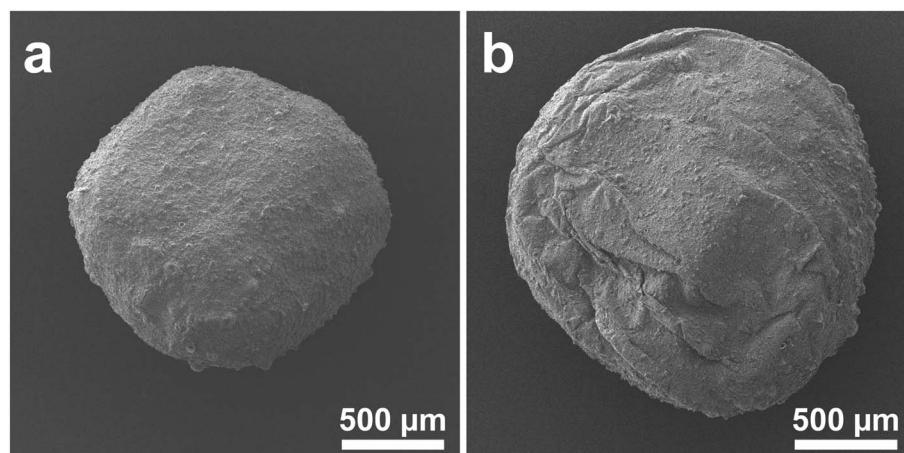


Fig. 11 SEM images of $\text{g-C}_3\text{N}_4/\text{Alg}$ bead (a) before and (b) after 10 repeated degradation experiments.

insufficiently cross-linked in the presence of photocatalytic particles, leading to gel collapse and photocatalytic particle loss following repeated degradation experiments.^{45,46}

Furthermore, the durability of $\text{g-C}_3\text{N}_4/\text{Alg}$ beads was evaluated during the 10 repeated cycles. Fig. S1† shows the bead elastic modulus with respect to irradiation time following each cycle. The $\text{g-C}_3\text{N}_4/\text{Alg}$ beads maintained a constant elastic modulus throughout the 10 cycles despite demonstrating activity for RhB degradation. The original bead strength could be further improved by adding poly(vinyl alcohol) (PVA), a biodegradable polymer, to the precursor material. Li *et al.* developed a PVA/alginate hydrogel with up to four times the elastic modulus relative to a simple alginate hydrogel.⁴⁷ These results prove that the $\text{g-C}_3\text{N}_4/\text{Alg}$ beads developed in this study are sufficiently durable and are useful photocatalysts that overcome previous difficulties involving recovery and reuse.

4. Conclusions

In this study, we demonstrated facile syntheses of isotype heterojunction $\text{g-C}_3\text{N}_4$ using a ball mill. $\text{g-C}_3\text{N}_4$ synthesized from urea and thiourea showed the highest photocatalytic activity owing to high specific surface area and effective charge separation. The results of the active species trapping experiments

suggest that $\cdot\text{O}_2^-$ and holes, rather than $\cdot\text{OH}$, contribute to RhB degradation. Moreover, the isotype heterojunction $\text{g-C}_3\text{N}_4$ was immobilized in alginate hydrogel spheres. A comparison of the bead shapes and hollow capsules showed that the $\text{g-C}_3\text{N}_4/\text{Alg}$ beads with well-dispersed $\text{g-C}_3\text{N}_4$ particles exhibited better photocatalytic activity. Repeated RhB degradation experiments using $\text{g-C}_3\text{N}_4/\text{Alg}$ beads showed no decrease in photocatalytic activity, and the beads maintained their elastic moduli throughout the entire cycle. This study demonstrates the synthesis of recyclable isotype heterojunction $\text{g-C}_3\text{N}_4$ that may circumvent present challenges to the practical application of sustainable photocatalytic wastewater treatment.

Data availability

The data supporting this article have been included as part of the ESI.†

Conflicts of interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.



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