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Elucidating gas–surface interactions relevant to atmospheric particle growth using combined temperature programmed desorption and temperature-dependent uptake[†]

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Understanding growth mechanisms for particles in air is fundamental to developing a predictive capability for their impacts on human health, visibility, and climate. In the case of highly viscous semi-solid or solid particles, the likelihood of impinging gases being taken up to grow the particle will be influenced by the initial uptake coefficient and by the residence time of the adsorbed gas on the surface. Here, a new approach that combines Knudsen cell capabilities for gas uptake measurements with temperature programmed desorption (TPD) for binding energy measurements of gases is described. The application of this unique capability to the uptake of organic gases on silica demonstrates its utility and the combination of thermodynamic and kinetic data that can be obtained. Lower limits to the initial net uptake coefficients at 170 K are $(3.0 \pm 0.6) \times 10^{-3}$, $(4.9 \pm 0.6) \times 10^{-3}$ and $(4.3 \pm 0.8) \times 10^{-3}$ for benzene, 1-chloropentane, and methanol, respectively, and are reported here for the first time. The uptake data demonstrated that the ideal gas lattice model was appropriate, which informed the analysis of the TPD data. From the thermal desorption measurements, desorption energies of 34.6 ± 2.5 , 45.8 ± 5.5 , and 40.0 ± 5.6 kJ mol⁻¹ (errors are 1σ) are obtained for benzene, 1-chloropentane, and methanol, respectively, and show good agreement with previously reported measurements. A multiphase kinetics model was applied to quantify uptake, desorption, and diffusion through the particle multilayers and hence extract desorption kinetics. Implications for uptake of organics on silica surfaces in the atmosphere and the utility of this system for determining relationships between residence times of organic gases and particle surfaces of varying composition are discussed in the context of developing quantitative predictions for growth of aerosol particles in air.

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Introduction

Exchange between the gas phase and the condensed phase of airborne particles is important as it determines concentrations of species in both phases.^{1,2} The growth of particles *via* net uptake from the gas phase is of particular importance. This is because particle size determines their impacts on light scattering and visibility,^{3–5} ability to alter cloud formation and properties,^{6–17} and the region of the lung they reach, which impacts health.^{18–25}

When a gas interacts with the surface of a particle in the atmosphere, its fate involves a number of potential pathways:

(1) inelastic collisions of the gas with the surface in which the gas adsorbs to the surface rather than elastically scattering; (2) desorption back to the gas phase; (3) reaction on the surface; (4) diffusion into the condensed phase; and (5) reaction in the bulk. A number of approaches have been developed that treat these interactions in different ways to quantify experimental measurements of net gas uptake and particle growth, and to provide predictive capabilities for these processes on scales from molecular to global.^{26–42}

Steps (1) and (2) depend on the interactions between the adsorbate and the surface which are determined by the intermolecular forces between the adsorbed gas molecule and the surface functional groups.^{43–45} Adsorbates are often described on a molecular basis as a 2D ideal gas or as a 2D ideal lattice gas.^{44,45} In the former case, the energy barrier for diffusion on the surface is much smaller than the thermal energy, $k_B T$ (k_B is the Boltzmann constant, T is the temperature), and molecules move relatively freely in the x, y directions parallel

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to the surface. In the case of the 2D ideal lattice gas, the energy barrier for diffusion is significantly greater than the available thermal energy so that the adsorbates are immobile and are treated as harmonic oscillators with no translational degrees of freedom. The intermediate case where the energy barrier and thermal energy are comparable is described as a hindered translator.⁴⁴

The stronger the adsorbate–surface interaction, the longer the residence time on the surface, providing increased opportunities for diffusion into the bulk, reaction on the surface, or the adsorbed species becoming “buried” by incoming molecules from the gas phase. While there is experimental evidence for such a “burying mechanism” of particle growth,^{46–48} quantifying its contribution requires a detailed molecular-level understanding of the residence time of species on the surface and hence the factors that contribute to the binding energies between the gas and surface.

Diffusion into the bulk, step (4), is dependent on the phase state of the particles and competes with desorption. The phase state of airborne particles is highly variable, depending on their source, reactions in air, temperature and relative humidity.^{49–59} Their phase can vary from low viscosity liquids to high viscosity semi-solids and solids, with the corresponding diffusion coefficients varying from high ($\sim 10^{-5} \text{ cm}^2 \text{ s}^{-1}$) to extremely low ($< 10^{-20} \text{ cm}^2 \text{ s}^{-1}$).⁶⁰ Furthermore, regions of very different viscosity can exist within one particle due to phase separation.⁶¹ While diffusion into liquids is relatively fast, this is not the case for solids and semi-solids,⁶⁰ and other processes such as desorption back into the gas phase or burying by incoming gases can be competitive.

We describe here an approach to study the interaction of gases with surfaces in non-reactive systems that combines temperature programmed desorption (TPD)⁶² measurements to obtain binding energies, with a Knudsen cell^{1,63} for uptake

measurements. The latter was very useful in identifying the appropriate gas adsorption model to use (ideal gas lattice model). The Knudsen cell portion of this apparatus was used to measure reactive gas uptake and has been recently described for the temperature dependent uptake of *n*-butylamine on various solid diacids.⁶⁴ In that case, an irreversible acid–base reaction occurs to form an ionic liquid layer, and desorption of the amine back into the gas phase is not important. Application to non-reactive systems is a more challenging problem since desorption competes with uptake. We report here adsorption and desorption from silica particles of benzene,⁶⁵ 1-chloropentane,⁶⁶ and methanol.⁶⁷ These gases have desorption energies spanning a range of values and different intermolecular forces such as hydrogen bonding, halogen bonding^{68,69} and dispersion forces. Silica is a well-studied surface due to its stability and widespread use. It is also relevant to the atmosphere as silica is the main component of suspended mineral dust particles,^{1,70} which can be transported over long distances. Organic coatings have been observed on such particles, making them a sink for volatile organic compounds, VOC.⁷¹ In addition, glass surfaces which have a high silica content are known to play a key role in the partitioning of organics indoors.^{72,73}

Experimental

Apparatus design

The apparatus design is based on the Knudsen cell designs of Caloz *et al.*⁷⁴ and of Mønster *et al.*⁷⁵ modified to incorporate TPD capabilities (Fig. 1). A detailed description is found in the ESI.† Briefly, the main chamber consists of a stainless steel six way cross with total volume of 5260 cm^3 and an estimated interior surface area of 2360 cm^2 . The bottom flange holds the removable sample cup and separates the cross from the cryogenic/heating parts of the apparatus. Temperature control of the sample from

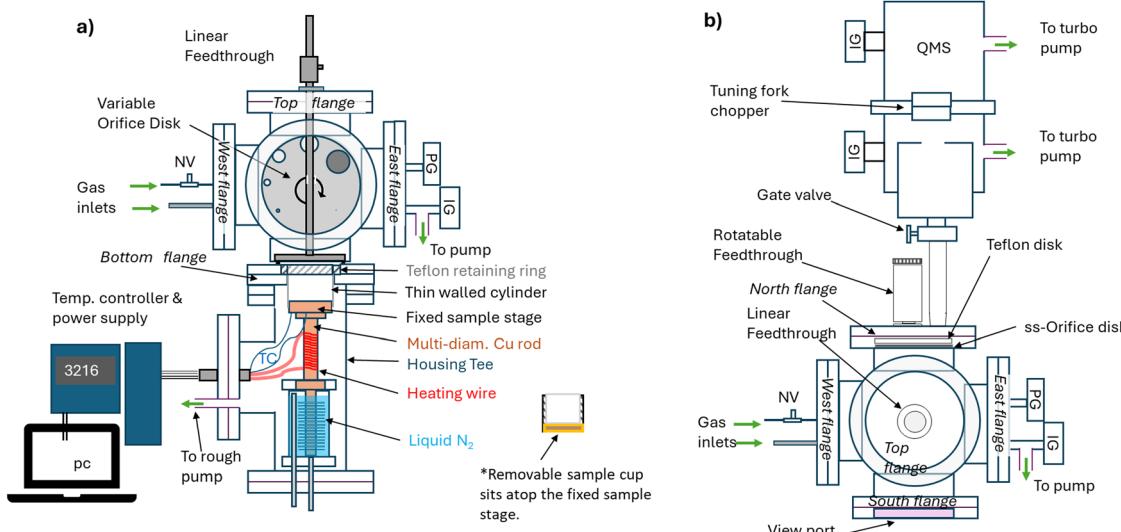


Fig. 1 Schematic of the Knudsen/TPD system: (a) side view looking through the view port (south flange) to reveal the rotatable orifice disk facing the north flange. (b) Top view of cell with line to mass spectrometer. PG = pressure gauge, IG = ionization gauge, TC = thermocouple, NV = needle valve, QMS = quadrupole mass spectrometer.

100 K to 400 K is provided by balancing heating and cooling with liquid nitrogen of a copper part that terminates in a fixed sample stage that is in contact with the bottom of the sample cup (Fig. S1, ESI[†]). The top flange of the main chamber holds a linear feedthrough (Fig. 1a) that controls the position of the sample cover for uptake experiments or the gas doser (Fig. S2, ESI[†]) for TPD experiments. A rotatable feedthrough (Fig. S3, ESI[†]) allows for the selection of orifices of different diameters (0.5, 1, 2, 4, 8, 11 and 14 mm) which control the gas flow rate to the quadrupole mass spectrometer (QMS, Extrel Core Mass Spectrometer equipped with an electron impact ionization source). One flange has a hinged easy access viewport to insert or remove the sample cup. It also provides a line-of-sight to view the selected orifice and ensure it is centered over the opening to the mass spectrometer. Gas escape rates were measured experimentally to define the effective orifice diameters (Table S2, ESI[†]).

Materials and methods

Sample preparation

A suspension of fumed silica powder (Cabosil[®] Grade 5, Cabot Corp.) (particle size ~ 100 nm) was made in either toluene or water, and 1–5 mg of silica was deposited on an argon plasma-cleaned silicon wafer. After the solvent had evaporated, the sample was placed in the vacuum chamber ($P = 1 \times 10^{-7}$ torr) and heated for at least 8 hours at 380 K to remove excess solvent and adsorbed water. The main gas phase contaminant at 10^{-7} torr is water vapor from the chamber walls. Considering the water impingement rate at 10^{-7} torr and 298 K is 3.6×10^{13} collisions $\text{cm}^{-2} \text{ s}^{-1}$, the time before the entire sample surface ($\sim 200 \text{ cm}^2$) is coated with a layer of water is on the order of hours, longer than the time for each sample run. Gases used in the experiments were prepared by adding liquid benzene (99.7%, EMD Chemicals), methanol (ACS Grade 99.8%, Fisher Chemical), or 1-chloropentane (99%, Sigma-Aldrich) to a 50 mL glass bulb attached to a vacuum manifold. Each underwent 3 freeze–pump–thaw cycles before use.

BET surface measurements. The surface area of the Cabosil both as a bulk powder and as deposited on the silicon wafers was determined using the traditional method of measuring the volume uptake of N_2 gas at 77 K. Briefly, N_2 at pressures from ~ 60 –230 torr were introduced into a calibrated volume and then expanded into a sample holder of known volume containing the Cabosil which had been heated and pumped to remove water. In one set of experiments, 0.13 g of loose powder was used. In the second set, measured amounts of an aqueous suspension of Cabosil was deposited on silicon wafers and dried in a manner similar to that used in the TPD and uptake experiments, with a total of 20 mg of Cabosil deposited. The measured surface areas of these two samples were within experimental error of each other, $196 \pm 66 \text{ m}^2 \text{ g}^{-1}$ (1σ) for the loose powder and $264 \pm 161 \text{ m}^2 \text{ g}^{-1}$ (1σ) for the powder deposited on the silicon wafers. Thus, the deposition of Cabosil on silicon discs does not change the available surface area for N_2 adsorption from that in the loose powder, within experimental error.

Uptake measurements. A series of Knudsen cell uptake experiments^{1,63} at 170 K were carried out to measure the non-reactive uptake of benzene, 1-chloropentane, and methanol on the silica (Cabosil) nanoparticles. For each experiment the sample was covered, and the temperature of the sample was adjusted to 170 K. The temperature was held at the desired value for ~ 10 minutes to ensure temperature equilibration throughout the sample. Once the silica was thermally equilibrated, the gas of interest was introduced into the cell to reach a stable signal in the QMS (denoted as I_0). The time required for the signal to stabilize, < 5 min, is short enough that the pressure inside the sample compartment remains at the background chamber pressure. After the signal (pressure) had stabilized, the sample lid was opened, exposing the silica particles to the gas of interest. As gases are adsorbed onto the surface, the gas signal decreases to a minimum value denoted as I_r . Because the gas is reversibly adsorbed, the intensity gradually recovers after the initial minimum but to a level slightly below the initial signal, where the net flux (adsorption – desorption) becomes constant. Initial net uptake coefficients (γ_0) were calculated using eqn (1),

$$\gamma_0 = \left(\frac{I_0}{I_r} - 1 \right) \left(\frac{A_o}{A_s} \right) \quad (1)$$

where I_0 and I_r are the signal before and immediately after exposing the sample to the molecules of interest. A_o and A_s are the areas of the variable orifice and sample, respectively.^{1,63} The ability to change A_o during an experiment using the rotatable orifice disc (Fig. 1b and Fig. S3, ESI[†]) facilitates measurement of a wide range of uptake coefficients by optimizing I_0/I_r in real time so that robust changes in signal on exposure of the gas molecule to the surface occur.

Temperature programmed desorption. For the TPD measurements, the multichannel effusive array doser (Fig. S2, ESI[†]) was installed on the linear feedthrough on the top flange of the chamber. The sample temperature was reduced to < 120 K, and a known volume of gases was dosed onto the surface. The sample was then annealed at 150–160 K, depending on the gas. This annealing step was necessary to ensure adsorbates were able to distribute across the accessible sample surface. Annealing for different durations (~ 5 –20 min) provided a reproducible method to obtain different sub-monolayer coverages, with longer annealing time corresponding to smaller initial coverages. After cooling the sample back to 120 K, the sample was set to heat at a temperature ramp of 0.2 K s^{-1} up to 298–320 K. The desorption rate is directly proportional to the intensity of the mass spectrometer signal set to the most intense mass fragment of each molecule, m/z 78 for benzene, m/z 55 for 1-chloropentane and m/z 31 for methanol. At the end of the temperature ramp, the sample was heated to 298 K to completely desorb all the benzene and methanol, while heating to 320 K was necessary for 1-chloropentane. The sample could then be cooled for subsequent experiments.

Kinetic modeling. The multi-layer kinetic model used a flux-based approach to describe vertical diffusion through the pores between silica particles and reversible adsorption of compounds to the silica particle surfaces. Concentrations in

the chamber in the gas phase, in the pores between silica particles and on the particle surfaces as a function of depth and time were calculated. A detailed description is included in the ESI[†] (Fig. S5).^{76,77} A list of parameters used in the model are presented in Table S3 (ESI[†]) along with explanations of the values. The desorption rate of molecules from the silica surface and the diffusion coefficient through the gas-phase pores were the only unknown parameters and were varied until the model could replicate the experimental measurements. The pressure was reset in the model each time the lid was closed, but drifts in the signals that occurred while the lid was open have not been treated. Note that $\alpha_{s,0}$ (surface accommodation coefficient on an adsorbate free silica surface) for all molecules was set to 1 in the model based on molecular dynamic simulations of small organic molecules interacting with silica surfaces and other surfaces.^{78–81} These calculations are for room temperature, whereas our experiments are at low temperatures where mass accommodation should be even higher. The values for $\alpha_{s,0}$ of benzene, methanol and chloropentane on silica surfaces could potentially be lower than one. We performed a sensitivity test where $\alpha_{s,0}$ of these molecules was reduced by up to a factor of 10 and found that within this range, we could obtain very similar fittings to the measurements by also decreasing k_d by the same factor.

Results & discussion

Uptake of organics onto silica particles

Representative time profiles for uptake at 170 K are shown in Fig. 2. Concentration dependent initial uptake coefficients were measured from the minimum in the signal immediately after opening the sample cup lid and are shown in Fig. 2. Uptake on the sample cup alone was negligible and experiments with the silicon wafer but no added silica particles (surface area 18 cm²) showed less than 5% reduction in the I_0 after opening the lid. Initial values of the net uptake coefficients (γ_0) are summarized in Table 1. Although the data are somewhat scattered at the lowest pressures, an estimate for the zero-pressure limit can be obtained from linear fits to the data (Fig. 3). The decrease in the initial uptake coefficients at higher gas phase concentrations may be due

Table 1 Initial net uptake coefficients at 170 K from Knudsen cell uptake measurements of 3 separate experiments ($\pm 1\sigma$)

Gases	γ_0^a
Benzene	$(3.0 \pm 0.6) \times 10^{-3}$
Methanol	$(4.3 \pm 0.8) \times 10^{-3}$
1-Chloropentane	$(4.9 \pm 0.6) \times 10^{-3}$

^a Extrapolated to zero pressure (Fig. 3). As discussed in the text, these should be considered lower limit values.

to the gas-substrate collision rates at higher pressures being fast enough to quickly decrease the numbers of available surface sites, while simultaneously increasing desorption. Thus, the measured values are lower limits, since increased desorption rates can reduce apparent uptake coefficients even at the shortest times when it is typically assumed the desorption rate is zero. By extrapolating to the zero-pressure limit, the influences of desorption as well as limitations in the response time are minimized in the reported value of the initial uptake coefficient.

A significant source of uncertainty in uptake measurements is determination of the effective available sample area. One way to address this is to use a single layer (or less) of particles evenly distributed across the sample holder in such a way that the adjacent particles do not touch.^{64,82} In the case of Cabosil, the particles do not spread out evenly over the surface of the silicon wafer (Fig. S7, ESI[†]) so that this approach could not be used.

Consequently, an alternative approach was developed where we used TPD profiles that showed both the multilayer and monolayer peaks to estimate that there are ~ 5 –10 multilayers of the gas initially adsorbed. The multilayer desorption peak that occurs at lower temperatures has a zero-order line shape and is narrower than the monolayer peak that occurs at higher temperatures. The profiles could be deconvoluted to determine the total signal associated with the monolayer signal (Fig. S4, ESI[†]). By calibrating the mass spectrometer signal using known pressures and escape rates (Fig. S8 and Table S2, ESI[†]), the monolayer signal can be converted to a total number of molecules. From the cross-sectional areas of each of the three molecules,⁸³ the available surface area was found to be $\sim 10\%$ of the BET surface area of the same sample determined using

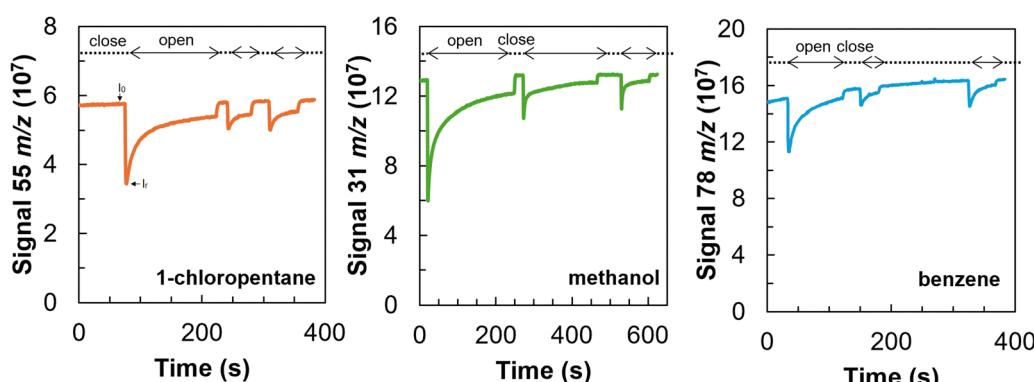


Fig. 2 Representative uptake profiles of 1-chloropentane, methanol, and benzene on silica particles at 170 K. Initial chamber pressures were 0.9×10^{-5} , 1.5×10^{-5} , and 1.6×10^{-5} torr for 1-chloropentane, methanol, and benzene, respectively.

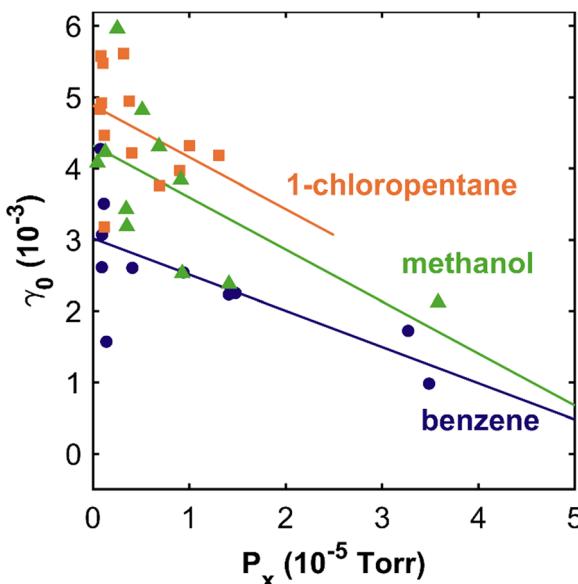


Fig. 3 Initial uptake coefficients as a function of the initial gas phase pressure are determined from the decrease in signal immediately after exposing the sample to various gas phase molecules: methanol (triangles), benzene (circles), and 1-chloropentane (squares). Solid curves are the best linear fit to the data.

N_2 adsorption. This is similar to the observation by Seisel *et al.*⁸⁴ who found that the effective surface area was roughly 1/20th of the BET surface area in uptake experiments of water on mineral dust. In that case they assumed that the total number of adsorbed molecules during the uptake experiments corresponded to one monolayer. A significant advantage of our approach is that the combination of uptake and desorption measurements on the same sample provides a way to experimentally determine the available surface area of powdered samples. The kinetic multilayer modeling results point to a

possible explanation for the apparent surface area of $\sim 10\%$ of the BET surface area. The model shows that during the first few seconds of exposure and uptake, the adsorbed molecules only reside in the top layers of the bed of particles (Fig. S6, ESI[†]).

In the uptake time profiles shown in Fig. 2, the signal does not recover to the initial signal (I_0) but rather to a smaller value where the rate of net uptake (*i.e.*, the difference of adsorption and desorption fluxes) is constant. We assume that the uptake has reached a steady-state at the end of the exposure. Note that this assumption may not be fully correct as the system can be limited by diffusion through silica particles and the time-dependent uptake may not be captured as discussed in the modeling section. The number of molecules that have been adsorbed are determined by numeric integration of the difference in the uptake profile signal with the lid closed and with the lid open. By measuring the number of adsorbed molecules at the steady state at various gas-phase concentrations, an effective adsorption isotherm can be obtained by neglecting bulk diffusion. These are shown in Fig. 4 with a Langmuir isotherm fit to the data. In the Langmuir model, it is assumed that each adsorbate is immobile on the surface and that they do not interact with neighboring molecules. The surface coverage can then be expressed as eqn (2),

$$\theta = \frac{KP}{1+KP} \quad (2)$$

where θ is the fractional surface coverage (the number of molecules adsorbed divided by the maximum number of molecules adsorbed), P is the concentration of the molecule in the gas phase (molecules cm^{-3}), and K is the equilibrium constant ($\text{cm}^3 \text{molecule}^{-1}$) for the reaction (I):



In reaction (I) A_g is the molecule in the gas phase, S is an available surface site, and A_{ads} is the adsorbed molecule.

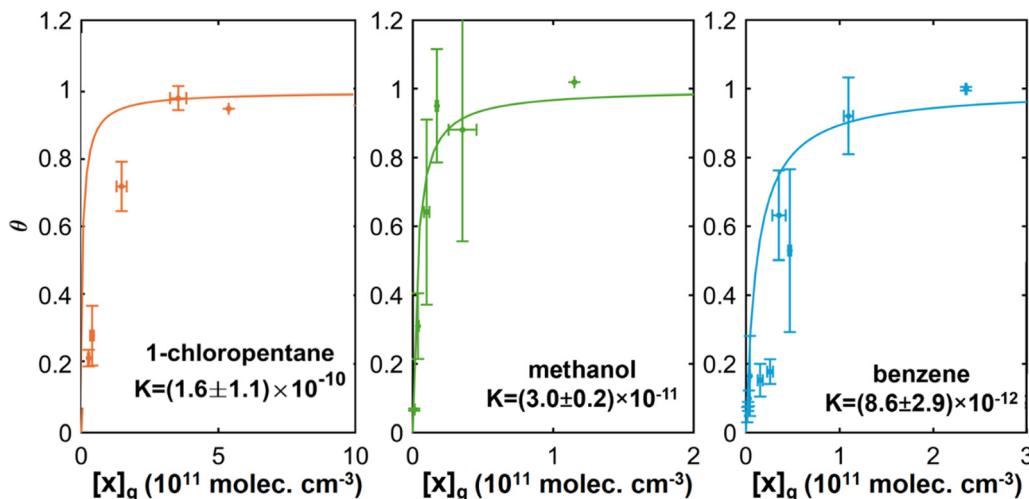


Fig. 4 Adsorption isotherms for 1-chloropentane, methanol, and benzene. The gas phase concentration is shown on the x-axis and the fractional surface coverage is shown on the y-axis. The fractional surface coverage was found by numeric integration of the uptake profiles from the point where the silica is first exposed to the gas until the system reaches equilibrium. The solid curves are the Langmuir isotherm fits to the data. Best fit K 's (unit of $\text{cm}^3 \text{molecule}^{-1}$) are shown in each chart where uncertainty is 1σ for the fitted parameter.

Langmuir isotherm plots as well as values of the equilibrium constants are shown in Fig. 4. The equilibrium constants were determined by linear regression of plots of $\theta/(1 - \theta)$ against the gas phase concentration which has a slope equal to K (Fig. S9, ESI†). It should be noted that K measured here is an effective value for the steady state Langmuir isotherm which may or may not be a true equilibrium due to possible gradients in the concentration of adsorbed molecules within the bed of particles. Nevertheless, the experimentally determined values of the Langmuir constant ($K_{\text{benzene}} = (8.6 \pm 2.9) \times 10^{-12} \text{ cm}^3 \text{ molecule}^{-1}$, $K_{\text{methanol}} = (3.0 \pm 0.2) \times 10^{-11} \text{ cm}^3 \text{ molecule}^{-1}$, and $K_{1\text{-chloropentane}} = (1.6 \pm 1.1) \times 10^{-10} \text{ cm}^3 \text{ molecule}^{-1}$) are in good agreement with those calculated with the kinetic model ($K_{\text{benzene}} = 7.7 \times 10^{-12} \text{ cm}^3 \text{ molecule}^{-1}$, $K_{\text{methanol}} = 4.4 \times 10^{-11} \text{ cm}^3 \text{ molecule}^{-1}$, and $K_{1\text{-chloropentane}} = 1.8 \times 10^{-10} \text{ cm}^3 \text{ molecule}^{-1}$). The K 's from the kinetic model are calculated from the expression $K = \sigma (k_a/k_d)$ taking $k_a = \alpha_{s,0} \cdot (v_T/4)$ where σ is the cross-sectional area of the molecule, $\alpha_{s,0}$ is the initial surface accommodation coefficient and v_T is the mean thermal velocity of the molecule in the gas phase (Table S3, ESI†).⁷⁶ Measurement of Langmuir isotherms along with uptake and desorption described below demonstrate the versatility of this apparatus in providing both kinetics and steady-state data simultaneously.

Temperature programmed desorption measurements of organics on silica. A typical TPD spectrum encompassing both multilayer and monolayer desorption and deconvolution of those curves is shown in Fig. S4 (ESI†). The integrated monolayer peak in Fig. S4 (ESI†) was used to determine the signal associated with one monolayer so that measured signal could

be converted to $d\theta/dt$. Examples of typical TPD curves are shown in Fig. 5a–c. Quantitative analysis of the desorption rate curves for the surface coverages of a monolayer or less was carried out using the inversion method developed by Kay and coworkers.⁸⁵ In this method the Polanyi-Wigner equation (eqn (3)) is applied,

$$-\frac{d\theta}{dt}(\theta, T) = \nu(\theta, T)\theta^n \cdot \exp\left(\frac{-E_d(\theta)}{RT}\right) \quad (3)$$

where n is the reaction rate order, $\nu(\theta, T)$ is the preexponential factor, and E_d is the desorption energy.^{62,86} In the analysis, it is assumed $\nu(\theta, T)$ is constant while E_d is a function of coverage and that n is 1. Briefly, through mathematical inversion of eqn (3) the desorption energies at various coverages can be determined from eqn (4):

$$E_d(\theta) = -RT \ln\left(\frac{-d\theta/dt}{\nu \cdot \theta}\right) \quad (4)$$

The coverage dependent desorption energy can be extracted from the experimental desorption curves. This desorption energy is then used in eqn (3) to simulate desorption curves at initial coverages smaller than the curve with the highest initial coverage that is used as a starting point. This involves integrating eqn (3) to determine the change in coverage with temperature (time). The simulated TPD curves are then compared with the experimentally measured desorption rates and the sum of the squared residuals is calculated. The process is then repeated, treating the preexponential factor as a variational parameter until the error between the experimental and simulated TPD curves is minimized as shown in Fig. 5d–f,

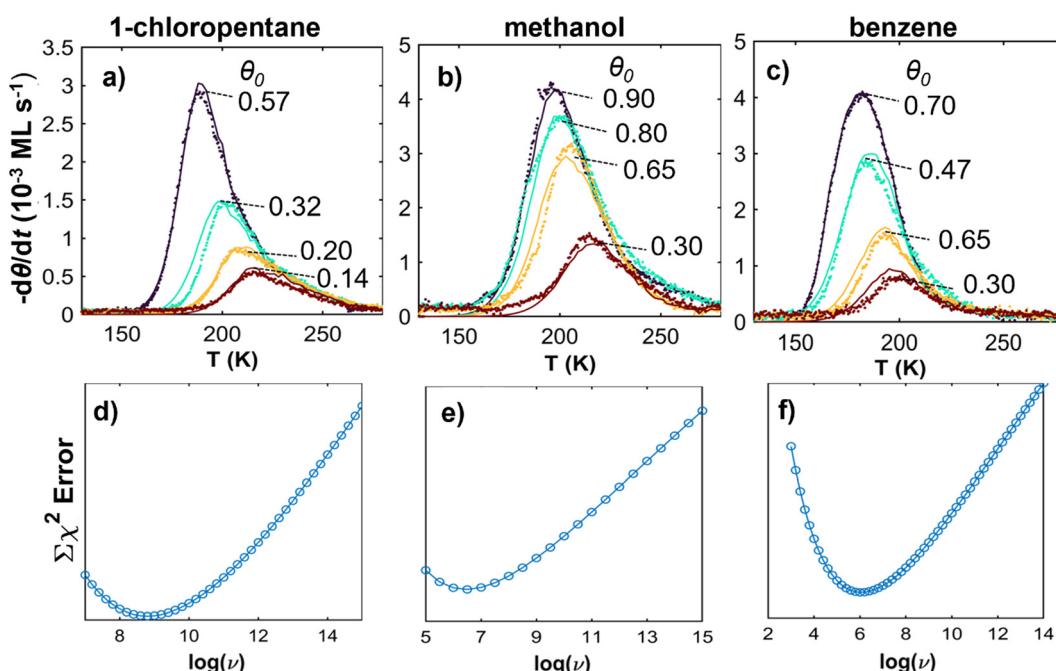


Fig. 5 TPD desorption rates versus temperature for typical experiments showing the data for decreasing initial molecular coverages (dots) along with best fits from inversion analysis (solid) for (a) 1-chloropentane, (b) methanol, (c) benzene. The total χ^2 error between the experimental and simulated desorption curves for all initial coverages versus \log_{10} of the preexponential factor used for the simulations is shown for (d) 1-chloropentane, (e) methanol, (f) benzene.

giving best-fit values for the preexponential factors. As discussed below, while the measured preexponential factor is affected by diffusion and adsorption/readsorption processes, these do not affect the desorption energies derived from this approach.⁸⁷

Fig. 6 shows desorption energies as a function of coverage from a typical TPD experiment. Extrapolation of the linear portion of the curves in Fig. 6 to the zero-coverage limit corresponds to the desorption energy of an isolated molecule that is not on a unique type of site *e.g.*, a defect site, or subject to adsorbate-adsorbate interactions.⁸⁵ These are summarized in Table 2. For comparison, Table 2 also shows results from Abelard *et al.*^{65,66} whose experimental approach and data analysis are similar to those in the present studies. Both the desorption energies and preexponential factors are in good agreement. However, the preexponential factors in both sets of studies (10^{6-9} s^{-1}) are much smaller than expected for desorption kinetics, where values $\sim 10^{13} \text{ s}^{-1}$ are often taken as typical.^{43,45,62}

Unexpectedly small preexponential factors have been observed in studies of desorption from porous catalysts and attributed to readsorption of the gas as it moves through the solid.⁸⁷⁻⁹² In the current experiments, the silica powder formed irregular mounds (Fig. S7, ESI†) across the surface of the silicon wafer. An estimate of ~ 22 layers of particles is obtained by assuming hexagonal close packing and uniform distribution of 1 mg of particles within a cylindrical bed of particles with 2 cm diameter. As discussed above, the number of molecules in a monolayer desorbed during TPD suggested that about 10% of the total particle surface area measured by BET adsorption of nitrogen was involved in the organic gas uptake, some of which will be subsurface layers. Therefore, the adsorbates are desorbing but have an opportunity to readsorb as they travel through the silica before ultimately escaping into the gas phase. While this affects the measured preexponential factor, it does not alter the measured desorption energies⁸⁷ which are

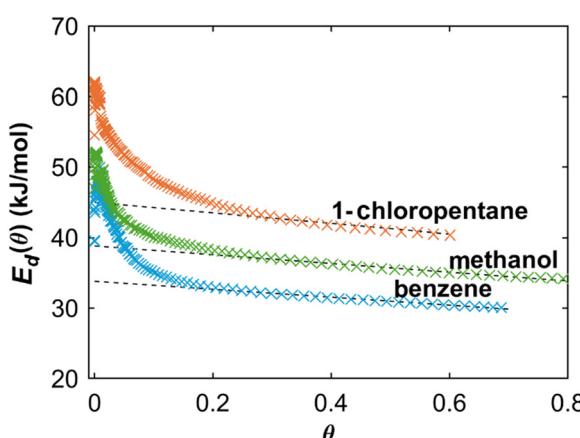


Fig. 6 Dependence of desorption energy on surface coverage for typical experiments by inversion of the Polanyi–Wigner equation, eqn (3), using a preexponential factor that corresponds to the minimum error for each molecule: benzene $\nu = 106.5$, 1-chloropentane $\nu = 109.4$, and methanol $\nu = 106.7$. The linear part of each curve was fit to a first order polynomial. The y -intercept corresponds to the desorption energy for an isolated molecule from a terrace site and is reported in Table 2.

Table 2 Desorption energies and pre-exponential factors from temperature programmed desorption inversion analysis of 3 separate experiments ($\pm 1\sigma$)

Gases	$E_d \text{ (kJ mol}^{-1}\text{)}$		$\log(\nu)$	
	This work	Previous work	This work	Previous work
Benzene	34.6 ± 3.0	31.0 ± 2.3^a	6.3 ± 0.6	6.0 ± 0.6^a
Methanol	40.0 ± 3.0	—	6.7 ± 0.5	—
1-Chloropentane	45.8 ± 5.5	43.7 ± 2.3^b	9.4 ± 0.6	9.1 ± 0.6^b

^a Ref. 65. ^b Ref. 66.

the parameters of interest here. Readsorption can also broaden the TPD peaks for powdered samples when compared to TPD from single crystal samples. Sneh and George reported methanol desorption from silicon wafers with the native SiO_2 layer and comparing the shape of the TPD curves in to those presented here shows that the monolayer peak broadens for the powdered sample and the maximum desorption rate shifts from 180 K to 200 K.⁶⁷ Using thin layers of powders has been shown to minimize the readsorption effects. For example, in the case of water desorption from TiO_2 , shifts of up to 40 K compared to the single crystal counterpart have been observed, but did not change the peak assignments.^{88,91}

It is clear from the gas uptake data and the Langmuir-type behavior that the 2D ideal lattice gas model is the most appropriate representation of adsorbates on silica. Evidence for the applicability of this model is the surface saturation observed in the gas uptake experiments (Fig. 2). In this case, adsorbates have higher rates of desorption at higher surface coverages. This can be seen experimentally in Fig. 7 where the desorption rate increases nonlinearly with coverage, especially at high surface coverage. From transition state theory, the desorption rate (R_{des}) for a 2D lattice gas can be expressed as eqn (5) (Campbell *et al.*⁴⁴ equation 70),

$$\frac{R_{\text{des}}}{A} = \left(\frac{kT}{h} \right) \left(\frac{q_{\text{TS}}^0/A}{(N_{\text{TS}}/A)^0 \cdot q_{\text{ad}}^0} \right) \left[\left(\frac{\theta}{1-\theta} \right) \left(\frac{N_{\text{TS}}}{A} \right)^0 \right] e^{-E_d/RT} \quad (5)$$

where the R_{des} is expressed in molecules s^{-1} , A is the surface area in cm^2 , k is Boltzmann's constant, h is Plank's constant, q_{TS}^0 and q_{ad}^0 are partition functions for the molecule in the transition and adsorbed state respectively, $(N_{\text{TS}}/A)^0$ is the standard surface concentration for molecules in the transition state, θ is the surface coverage, and E_d is the desorption energy. This equation holds for any choice of standard state. eqn (5) is similar in form to the Polanyi–Wigner equation used in the inversion analysis of the data (eqn (3)), where ν is defined as eqn (6),

$$\nu = \left(\frac{kT}{h} \right) \left(\frac{q_{\text{TS}}^0/A}{(N_{\text{TS}}/A)^0 \cdot q_{\text{ad}}^0} \right) \quad (6)$$

However, eqn (5) explicitly recognizes that for a 2D ideal lattice gas, the surface coverage dependence of the desorption rate depends on $\theta/(1-\theta)$, rather than θ^n (eqn (3)), so that desorption rates increase with coverage. This is consistent with the data in Fig. 7 that clearly show this strong dependence of desorption rates on surface coverage. At small surface

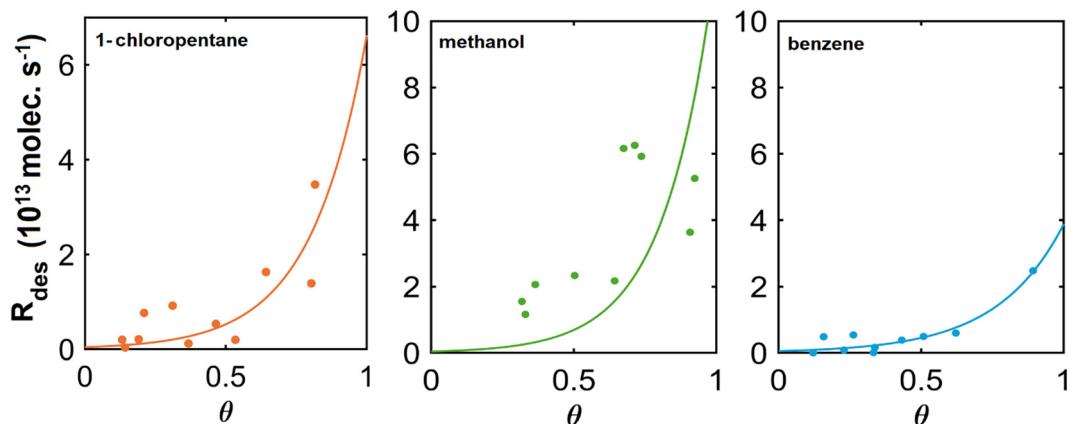


Fig. 7 Desorption rates at 170 K were determined from the TPD curves with varying initial coverages. The surface coverage at the specified temperature was determined by integration of the TPD curves. The solid curves are the best fit to the data.

coverages, $\theta \ll 1$ eqn (5) predicts desorption rates that are first order in θ . In terms of atmospheric relevance, it is these lower surface coverages that are of most interest.

From eqn (5), the slope of a line of $\ln[(R_{\text{des}})/(\theta/1 - \theta)]$ versus $1/T$ is equal to E_d/R if E_d is not a function of the coverage. The data in Fig. 6 show only a weak dependence on coverage for θ from 0.2 to 0.8. Therefore, assuming the coverage dependent contribution is small in this region, the slope was used to determine E_d . The plots for benzene, 1-chloropentane, and methanol are shown in Fig. 8. The desorption energies obtained in this manner (Table 3) agree within experimental error with those determined through inversion analysis (Table 2). Note that these desorption energies are similar to those calculated for benzene on silica by Fang *et al.*⁹³ (28 kJ mol⁻¹) or on quartz by Budi *et al.*⁹⁴ (34 kJ mol⁻¹). However, our measured methanol

value is significantly smaller than that calculated by Budi *et al.*⁹⁴ (61 kJ mol⁻¹) for adsorption on quartz.

The desorption energy is lowest for benzene, followed by methanol, and finally is highest for 1-chloropentane. Hydrogen bonding between the -OH groups of methanol and of the silica will be the major intermolecular force in that case, while for benzene, hydrogen bonding occurs between the -OH group and the π -system on the benzene ring.^{65,93} In the case of 1-chloropentane, dispersion forces and possibly halogen bonding⁶⁹ play a role.

Kinetic modeling studies. Fig. 9 shows that the kinetic model can replicate the measured time and concentration dependence of benzene uptake onto silica particles reasonably well for experiments performed at different pressures. This is also true for methanol (Fig. S10, ESI[†]) and 1-chloropentane (Fig. S11, ESI[†]). The model results suggest that adsorption onto

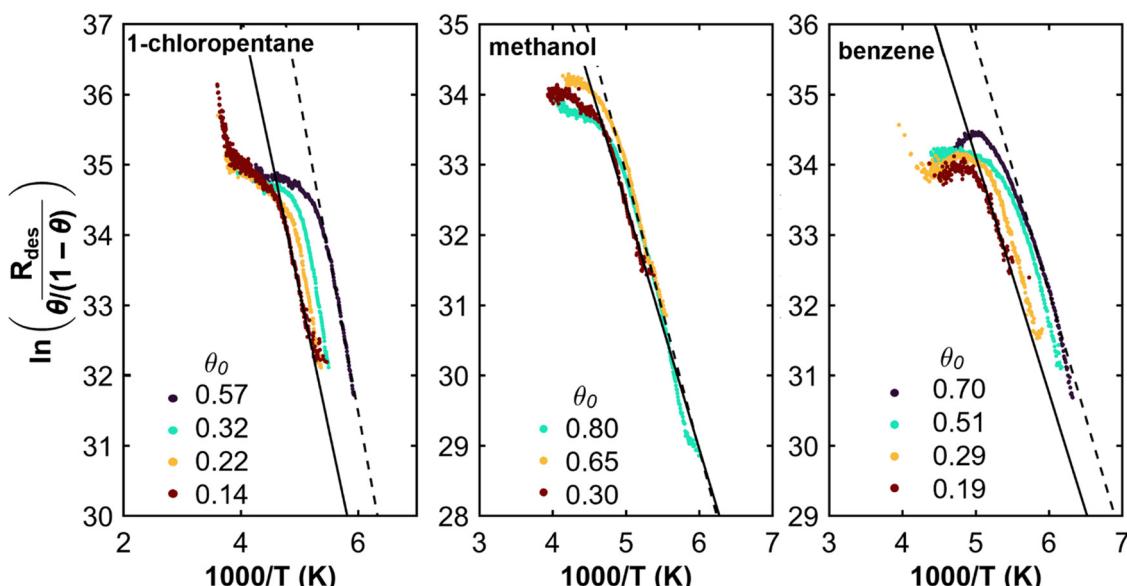


Fig. 8 Natural logarithms of desorption rates divided by $\theta/(1 - \theta)$ for 1-chloropentane, methanol, and benzene (from left to right). The slopes of the curves are equal to E_d/R . In each panel the dashed curve is the best fit to the largest initial coverage and the solid curve is the best fit to the lowest initial coverage.

Table 3 Desorption energies determined by Arrhenius fits to the TPD curves from 3 separate experiments ($\pm 1\sigma$)

Gases	E_d (kJ mol ⁻¹)
Benzene	34.5 \pm 5.0
Methanol	33.7 \pm 4.0
1-Chloropentane	40.1 \pm 5.6

the top layers of the silica particles will effectively slow down diffusion through the silica particles, resulting in a benzene concentration gradient as a function of depth through the silica particle layers (Fig. S6, ESI[†]). It also suggests that the net uptake is limited by diffusion and partitioning through silica particles at longer exposure time. This may explain the effectively lower surface area calculated on the timescale of our experiments. Note that in the absence of diffusion limitations the measured time dependence of the uptake cannot be replicated (Fig. S12, ESI[†]).

The best fits to the experimental data for different gases are compared in Fig. 10 and in Table 4. The model shows that the fit value for the diffusion coefficient for diffusion within the pores between particles is approximately the same for all the gases. The desorption rate constant (k_d) on the other hand is around 10 times greater for benzene (3 s^{-1}) than for 1-chloropentane (0.1 s^{-1}) or methanol (0.4 s^{-1}). These calculated desorption rate constants can be used to calculate pre-exponential factors, *via* $\nu = k_d/\exp(E_d/RT)$ which are summarized in Table 4. These are much closer to the typical values of $\sim 10^{13}$ expected for such processes.^{43,45,62}

Both the uptake and desorption of gases on silica surfaces are determined by the specific intermolecular forces involved in the particular gas-surface combination, as discussed in detail by Knopf *et al.*⁴³ These include hydrogen-bonding, dipole-dipole and London dispersion forces. Benzene has been shown in MD simulations to hydrogen-bond to the Si-OH group through its π electron cloud,⁹³ and methanol of course also hydrogen-bonds to the surface -OH groups.⁹⁵ In the case of the larger chloropentane molecule, both London dispersion forces as well as halogen bonding⁶⁹ between the chlorine and surface-OH groups contribute. It is therefore not surprising that the Langmuir equilibrium constants and desorption energies follow similar trends in that chloropentane is the most strongly bound, followed by methanol and benzene. An additional factor for uptake is that the portion of the gas phase molecule that binds to the surface -OH groups need to be appropriately oriented as it initially interacts with the surface if uptake is to occur, *i.e.* steric factors also play a role.

Summary and conclusions

Application of this new approach that combines Knudsen cell and TPD capabilities in the same system to study both adsorption and desorption processes, including for non-reactive gas-surface interactions, provides both thermodynamic and kinetics data. This approach provides desorption data and binding energies for benzene, 1-chloropentane and methanol on silica that are in good agreement with previous studies. The combination of

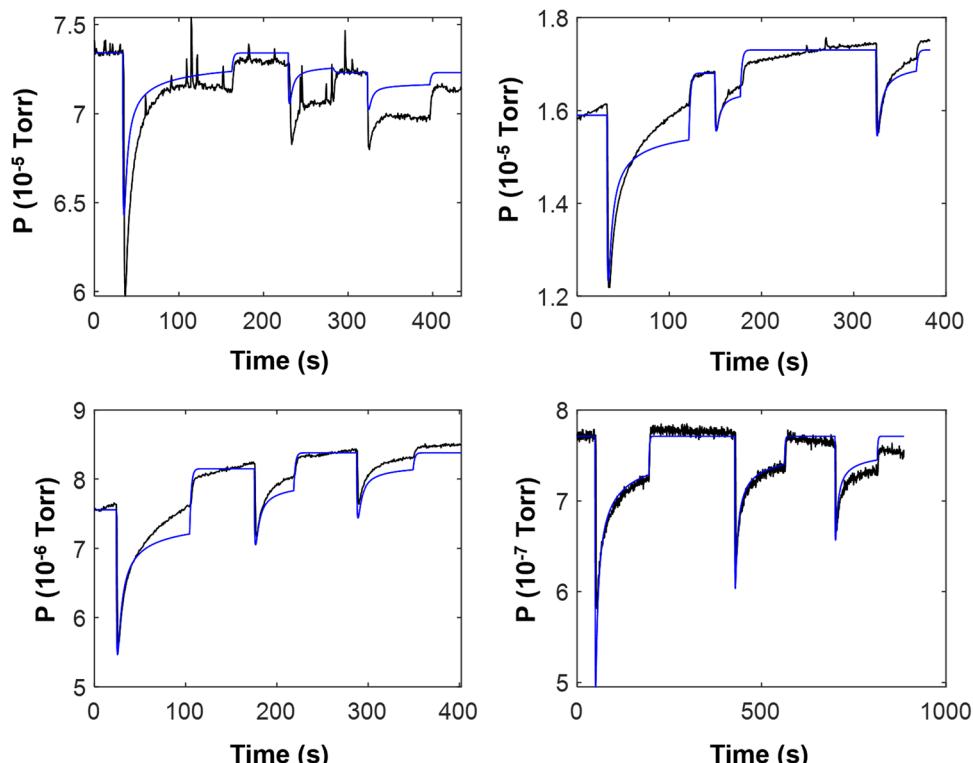


Fig. 9 Measurements (black) and kinetic multilayer model simulations (blue) of benzene uptake on silica particles for different initial benzene pressures.

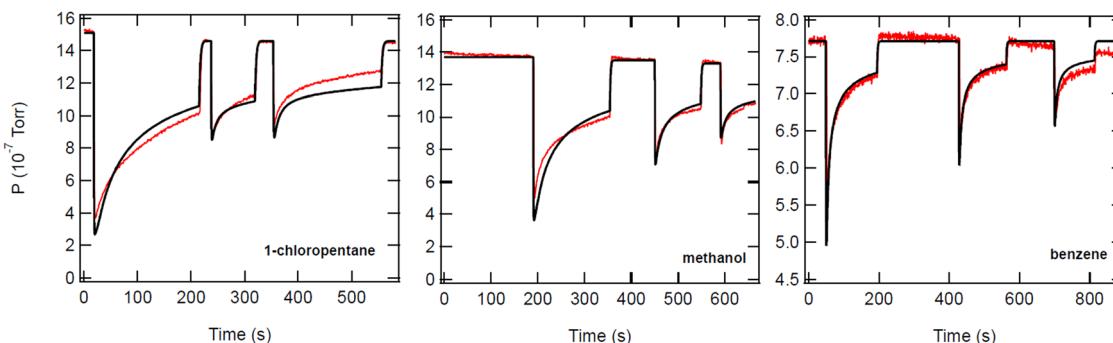


Fig. 10 Kinetic multilayer model simulations for uptake profiles of 1-chloropentane, methanol and benzene on silica at 170 K. The experimental data are shown in red, and the model fit is shown in black. The fitted parameters are $D_{g,p} = 0.0015 \text{ cm}^2 \text{ s}^{-1}$, 0.0012, 0.001 and $k_d = 0.1$, 0.4, and 3 s^{-1} for 1-chloropentane, methanol, and benzene respectively.

Table 4 Kinetic multilayer model results

Molecule	$D_p (\text{cm}^2 \text{ s}^{-1})^a$	$k_d (\text{s}^{-1})^b$	$\nu (\text{s}^{-1})^c$
Benzene	0.001	3	1.3×10^{11}
Methanol	0.0012	0.4	7.8×10^{11}
1-Chloropentane	0.0015	0.1	1.2×10^{13}

^a Diffusion coefficient through gas phase pores. ^b Desorption rate coefficient of molecules from the silica surface. ^c $\nu = k_d/\exp(E_d/RT)$, $T = 170 \text{ K}$ and E_d was determined by inversion analysis of TPD measurements.

uptake with TPD experiments clearly shows that sub-monolayer coverages of benzene, 1-chloropentane and methanol on silica can be described by Langmuir, 2D lattice gas, type adsorption. The combination of TPD and uptake measurements also allows for experimental determination of the effective available surface area of a powdered sample, which is a major source of uncertainty in typical uptake experiments. The kinetic modeling results provide a deeper understanding of the desorption mechanism and kinetics in this system where diffusion and readsorption onto the particles occurs. The combination of the Knudsen cell uptake measurements, TPD profiles, and kinetic multilayer modeling is more powerful than any of the individual parts.

In the atmosphere, reversible adsorption is a key process for gas condensed-phase interactions as it is a determining factor in aerosol particle growth. A large component of airborne particles is organic, formed by oxidation of organic gases that generate secondary organic aerosol (SOA) particles through new particle formation and growth.⁹⁶ Predicting growth rates requires knowledge of uptake coefficients as well as desorption energies, and how they are determined by the structural features of particle surfaces and the gases. Thus, based on an understanding of the relationship between binding energies and the gas and surface compositions, a predictive model can be developed that can be used to quantify particle growth mechanisms in the atmosphere. Future studies will use self-assembled monolayers having selected terminal groups in order to probe the reversible adsorption/desorption of gases of known structures on particles coated with various organic functionalities. This will provide insights into the relationships between gas-surface interactions and provide the basis of a

predictive capability for growth mechanisms of highly viscous organic particles in air.

Data availability

All relevant data are included in the manuscript or ESI.†

Conflicts of interest

There are no conflicts to declare.

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References

- 1 B. J. Finlayson-Pitts and J. N. Pitts, Jr., *Chemistry of the Upper and Lower Atmosphere - Theory, Experiments, and Applications*, Academic Press, San Diego, 2000.
- 2 J. H. Seinfeld and S. N. Pandis, *Atmospheric Chemistry and Physics: From Air Pollution to Climate Change*, Wiley, New York, 2006.
- 3 W. C. Hinds, *Aerosol Technology: Properties, Behavior and Measurement of Airborne Particles*, John Wiley & Sons Inc, New York, 1999.
- 4 M. Kanakidou, J. H. Seinfeld, S. N. Pandis, I. Barnes, F. J. Dentener, M. C. Facchini, R. Van Dingenen, B. Ervens, A. Nenes, C. J. Nielsen, E. Swietlicki, J. P. Putaud, Y. Balkanski, S. Fuzzi, J. Horth,

G. K. Moortgat, R. Winterhalter, C. E. L. Myhre, K. Tsigaridis, E. Vignati, E. G. Stephanou and J. Wilson, *Atmos. Chem. Phys.*, 2005, **5**, 1053–1123.

5 A. Singh, W. J. Bloss and F. D. Pope, *Atmos. Chem. Phys.*, 2017, **17**, 2085–2101.

6 A. S. Ackerman, O. B. Toon, J. P. Taylor, D. W. Johnson, P. V. Hobbs and R. J. Ferek, *J. Atmos. Sci.*, 2000, **57**, 2684–2695.

7 Y. C. Chen, M. W. Christensen, L. Xue, A. Sorooshian, G. L. Stephens, R. M. Rasmussen and J. H. Seinfeld, *Atmos. Chem. Phys.*, 2012, **12**, 8223–8235.

8 D. K. Farmer, C. D. Cappa and S. M. Kreidenweis, *Chem. Rev.*, 2015, **115**, 4199–4217.

9 E. A. Fofie, N. M. Donahue and A. Asa-Awuku, *Aerosol Sci. Technol.*, 2018, **52**, 242–251.

10 V. M. Kerminen, M. Paramonov, T. Anttila, I. Riipinen, C. Fountoukis, H. Korhonen, E. Asmi, L. Laakso, H. Lihavainen, E. Swietlicki, B. Svenningsson, A. Asmi, S. N. Pandis, M. Kulmala and T. Petaja, *Atmos. Chem. Phys.*, 2012, **12**, 12037–12059.

11 S. J. Lowe, D. G. Partridge, J. F. Davies, K. R. Wilson, D. Topping and I. Riipinen, *Nat. Commun.*, 2019, **10**.

12 A. Nenes, R. J. Charlson, M. C. Facchini, M. Kulmala, A. Laaksonen and J. H. Seinfeld, *Geophys. Res. Lett.*, 2002, **29**, 29–1–29–4.

13 D. Rosenfeld, *Science*, 2006, **312**, 1323–1324.

14 J. Schmale, S. Henning, S. Decesari, B. Henzing, H. Keskinen, K. Sellegri, J. Ovadnevaite, M. L. Pohlker, J. Brito, A. Bougiatioti, A. Kristensson, N. Kalivitis, I. Stavroulas, S. Carbone, A. Jefferson, M. Park, P. Schlag, Y. Iwamoto, P. Aalto, M. Aijala, N. Bukowiecki, M. Ehn, G. Frank, R. Frohlich, A. Frumau, E. Herrmann, H. Herrmann, R. Holzinger, G. Kos, M. Kulmala, N. Mihalopoulos, A. Nenes, C. O'Dowd, T. Petaja, D. Picard, C. Pohlker, U. Poschl, L. Poulain, A. S. H. Prevot, E. Swietlicki, M. O. Andreae, P. Artaxo, A. Wiedensohler, J. Ogren, A. Matsuki, S. S. Yum, F. Stratmann, U. Baltensperger and M. Gysel, *Atmos. Chem. Phys.*, 2018, **18**, 2853–2881.

15 S. Twomey, *Atmos. Environ.*, 1991, **25A**, 2435–2442.

16 D. M. Westervelt, J. R. Pierce, I. Riipinen, W. Trivitayanurak, A. Hamed, M. Kulmala, A. Laaksonen, S. Decesari and P. J. Adams, *Atmos. Chem. Phys.*, 2013, **13**, 7645–7663.

17 V. F. McNeill, *Annu. Rev. Chem. Biomol. Eng.*, 2017, **8**, 427–444.

18 D. W. Dockery, C. A. Pope, X. P. Xu, J. D. Spengler, J. H. Ware, M. E. Fay, B. G. Ferris and F. E. Speizer, *New Engl. J. Med.*, 1993, **329**, 1753–1759.

19 P. J. Landrigan, R. Fuller, N. J. R. Acosta, O. Adeyi, R. Arnold, N. Basu, A. B. Balde, R. Bertollini, S. Bose-O'Reilly, J. I. Boufford, P. N. Breysse, T. Chiles, C. Mahidol, A. M. Coll-Seck, M. L. Cropper, J. Fobil, V. Fuster, M. Greenstone, A. Haines, D. Hanrahan, D. Hunter, M. Khare, A. Krupnick, B. Lanphear, B. Lohani, K. Martin, K. V. Mathiasen, M. A. McTeer, C. J. L. Murray, J. D. Ndhimananjara, F. Perera, J. Potochnik, A. S. Preker, J. Ramesh, J. Rockstrom, C. Salinas, L. D. Samson, K. Sandilya, P. D. Sly, K. R. Smith, A. Steiner, R. B. Stewart, W. A. Suk, O. C. P. van Schayck, G. N. Yadama, K. Yumkella and M. Zhong, *Lancet*, 2018, **391**, 462–512.

20 P. J. Landrigan, R. Fuller, H. Hu, J. Caravanos, M. L. Cropper, D. Hanrahan, K. Sandilya, T. C. Chiles, P. Kumar and W. A. Suk, *Environ. Health Persp.*, 2018, **126**.

21 J. Lelieveld, J. S. Evans, M. Fnais, D. Giannadaki and A. Pozzer, *Nature*, 2015, **525**, 367–374.

22 P. M. Mannucci, S. Harari, I. Martinelli and M. Franchini, *Intern. Emerg. Med.*, 2015, **10**, 657–662.

23 C. A. Pope III and D. W. Dockery, *J. Air Waste Manage. Assoc.*, 2006, **56**, 709–742.

24 M. Shiraiwa, K. Ueda, A. Pozzer, G. Lammel, C. J. Kampf, A. Fushimi, S. Enami, A. M. Arangio, J. Frohlich-Nowoisky, Y. Fujitani, A. Furuyama, P. S. J. Lakey, J. Lelieveld, K. Lucas, Y. Morino, U. Poschl, S. Takahama, A. Takami, H. J. Tong, B. Weber, A. Yoshino and K. Sato, *Environ. Sci. Technol.*, 2017, **51**, 13545–13567.

25 J. E. Thompson, *J. Occup. Environ. Med.*, 2018, **60**, 392–423.

26 J. A. Faust and G. M. Nathanson, *Chem. Soc. Rev.*, 2016, **45**, 3609–3620.

27 J. F. Pankow, *Atmos. Environ.*, 1994, **28**, 185–188.

28 J. F. Pankow, *Atmos. Environ.*, 1987, **21**, 2275–2283.

29 J. L. Jimenez, M. R. Canagaratna, N. M. Donahue, A. S. H. Prevot, Q. Zhang, J. H. Kroll, P. F. DeCarlo, J. D. Allan, H. Coe, N. L. Ng, A. C. Aiken, K. S. Docherty, I. M. Ulbrich, A. P. Grieshop, A. L. Robinson, J. Duplissy, J. D. Smith, K. R. Wilson, V. A. Lanz, C. Hueglin, Y. L. Sun, J. Tian, A. Laaksonen, T. Raatikainen, J. Rautiainen, P. Vaattovaara, M. Ehn, M. Kulmala, J. M. Tomlinson, D. R. Collins, M. J. Cubison, E. J. Dunlea, J. A. Huffman, T. B. Onasch, M. R. Alfarra, P. I. Williams, K. Bower, Y. Kondo, J. Schneider, F. Drewnick, S. Borrmann, S. Weimer, K. Demerjian, D. Salcedo, L. Cottrell, R. Griffin, A. Takami, T. Miyoshi, S. Hatakeyama, A. Shimono, J. Y. Sun, Y. M. Zhang, K. Dzepina, J. R. Kimmel, D. Sueper, J. T. Jayne, S. C. Herndon, A. M. Trimborn, L. R. Williams, E. C. Wood, A. M. Middlebrook, C. E. Kolb, U. Baltensperger and D. R. Worsnop, *Science*, 2009, **326**, 1525–1529.

30 V. Perraud, E. A. Bruns, M. J. Ezell, S. N. Johnson, Y. Yu, M. L. Alexander, A. Zelenyuk, D. Imre, W. L. Chang, D. Dabdub, J. F. Pankow and B. J. Finlayson-Pitts, *Proc. Natl. Acad. Sci. U. S. A.*, 2012, **109**, 2836–2841.

31 M. Ammann and U. Pöschl, *Atmos. Chem. Phys.*, 2007, **7**, 6025–6045.

32 T. Berkemeier, A. J. Huisman, M. Ammann, M. Shiraiwa, T. Koop and U. Pöschl, *Atmos. Chem. Phys.*, 2013, **13**, 6663–6686.

33 J. P. Abbott, A. K. Lee and J. A. Thornton, *Chem. Soc. Rev.*, 2012, **41**, 6555–6581.

34 N. M. Donahue, W. Chuang, S. A. Epstein, J. H. Kroll, D. R. Worsnop, A. L. Robinson, P. J. Adams and S. N. Pandis, *Environ. Chem.*, 2013, **10**, 151–157.

35 I. Riipinen, J. R. Pierce, T. Yli-Juuti, T. Nieminen, S. Hakkinen, M. Ehn, H. Junninen, K. Lehtipalo, T. Petaja, J. Slowik, R. Chang, N. C. Shantz, J. Abbatt, W. R. Leaitch, V. M. Kerminen, D. R. Worsnop, S. N. Pandis, N. M. Donahue and M. Kulmala, *Atmos. Chem. Phys.*, 2011, **11**, 3865–3878.

36 P. Roldin, A. C. Eriksson, E. Z. Nordin, E. Hermansson, D. Mogensen, A. Rusanen, M. Boy, E. Swietlicki, B. Svenningsson, A. Zelenyuk and J. Pagels, *Atmos. Chem. Phys.*, 2014, **14**, 7953–7993.

37 M. Shiraiwa, C. Pfrang, T. Koop and U. Poschl, *Atmos. Chem. Phys.*, 2012, **12**, 2777–2794.

38 R. A. Zaveri, J. E. Shilling, A. Zelenyuk, J. M. Liu, D. M. Bell, E. L. D'Ambro, C. Gaston, J. A. Thornton, A. Laskin, P. Lin, J. Wilson, R. C. Easter, J. Wang, A. K. Bertram, S. T. Martin, J. H. Seinfeld and D. R. Worsnop, *Environ. Sci. Technol.*, 2018, **52**, 1191–1199.

39 V. Lannuque, M. Camredon, F. Couvidat, A. Hodzic, R. Valorso, S. Madronich, B. Bessagnet and B. Aumont, *Atmos. Chem. Phys.*, 2018, **18**, 13411–13428.

40 M. Shrivastava, C. D. Cappa, J. W. Fan, A. H. Goldstein, A. B. Guenther, J. L. Jimenez, C. Kuang, A. Laskin, S. T. Martin, N. L. Ng, T. Petaja, J. R. Pierce, P. J. Rasch, P. Roldin, J. H. Seinfeld, J. Shilling, J. N. Smith, J. A. Thornton, R. Volkamer, J. Wang, D. R. Worsnop, R. A. Zaveri, A. Zelenyuk and Q. Zhang, *Rev. Geophys.*, 2017, **55**, 509–559.

41 S. Tuovinen, J. Kontkanen, R. Cai and M. Kulmala, *Environ. Sci. Atmos.*, 2021, **1**, 543–557.

42 C. E. Kolb, R. A. Cox, J. P. D. Abbatt, M. Ammann, E. J. Davis, D. J. Donaldson, B. C. Garrett, C. George, P. T. Griffiths, D. R. Hanson, M. Kulmala, G. McFiggans, U. Pöschl, I. Riipinen, M. J. Rossi, Y. Rudich, P. E. Wagner, P. M. Winkler, D. R. Worsnop and C. D. O'Dowd, *Atmos. Chem. Phys.*, 2010, **10**, 10561–10605.

43 D. A. Knopf, M. Ammann, T. Berkemeier, U. Pöschl and M. Shiraiwa, *Atmos. Chem. Phys.*, 2024, **24**, 3445–3528.

44 C. T. Campbell, L. H. Sprowl and L. Arnadottir, *J. Phys. Chem. C*, 2016, **120**, 10283–10297.

45 D. A. Knopf and M. Ammann, *Atmos. Chem. Phys.*, 2021, **21**, 15725–15753.

46 A. C. van der Waals, P. S. J. Lakey, E. Rossich Molina, V. Perraud, L. M. Wingen, J. Xu, D. Soulsby, R. B. Gerber, M. Shiraiwa and B. J. Finlayson-Pitts, *Environ. Sci.: Processes Impacts*, 2018, **20**, 1593–1610.

47 A. C. van der Waals, V. Perraud, L. M. Wingen and B. J. Finlayson-Pitts, *Environ. Sci.: Processes Impacts*, 2020, **22**, 66–83.

48 A. C. van der Waals, L. M. Wingen, V. Perraud, Y. Zhao and B. J. Finlayson-Pitts, *ACS Earth Space Chem.*, 2020, **4**, 1435–1447.

49 E. Saukko, A. T. Lambe, P. Massoli, T. Koop, J. P. Wright, D. R. Croasdale, D. A. Pedernera, T. B. Onasch, A. Laaksonen, P. Davidovits, D. R. Worsnop and A. Virtanen, *Atmos. Chem. Phys.*, 2012, **12**, 7517–7529.

50 A. Virtanen, J. Joutsensaari, T. Koop, J. Kannisto, P. Yli-Pirila, J. Leskinen, J. M. Makela, J. K. Holopainen, U. Poschl, M. Kulmala, D. R. Worsnop and A. Laaksonen, *Nature*, 2010, **467**, 824–827.

51 A. Virtanen, J. Kannisto, H. Kuuluvainen, A. Arffman, J. Joutsensaari, E. Saukko, L. Hao, P. Yli-Pirila, P. Tiitta, J. K. Holopainen, J. Keskinen, D. R. Worsnop, J. N. Smith and A. Laaksonen, *Atmos. Chem. Phys.*, 2011, **11**, 8759–8766.

52 C. Kidd, V. Perraud, L. M. Wingen and B. J. Finlayson-Pitts, *Proc. Natl. Acad. Sci. U. S. A.*, 2014, **111**, 7552–7557.

53 A. P. Bateman, A. K. Bertram and S. T. Martin, *J. Phys. Chem. A*, 2015, **119**, 4386–4395.

54 M. Kuwata and S. T. Martin, *Proc. Natl. Acad. Sci. U. S. A.*, 2012, **109**, 17354–17359.

55 L. Renbaum-Wolff, J. W. Grayson, A. P. Bateman, M. Kuwata, M. Sellier, B. J. Murray, J. E. Shilling, S. T. Martin and A. K. Bertram, *Proc. Natl. Acad. Sci. U. S. A.*, 2013, **110**, 8014–8019.

56 M. Shiraiwa, Y. Li, A. P. Tsimpidi, V. A. Karydis, T. Berkemeier, S. N. Pandis, J. Lelieveld, T. Koop and U. Poschl, *Nat. Commun.*, 2017, 15002.

57 J. P. Reid, A. K. Bertram, D. O. Topping, A. Laskin, S. T. Martin, M. D. Petters, F. D. Pope and G. Rovelli, *Nat. Commun.*, 2018, 9.

58 Y. M. Han, Z. H. Gong, J. H. Ye, P. F. Liu, K. A. McKinney and S. T. Martin, *Environ. Sci. Technol.*, 2019, **53**, 13209–13218.

59 A. M. Maclean, N. R. Smith, Y. Li, Y. Z. Huang, A. P. S. Hettiyadura, G. V. Crescenzo, M. Shiraiwa, A. Laskin, S. A. Nizkorodov and A. K. Bertram, *ACS Earth Space Chem.*, 2021, **5**, 305–318.

60 T. Koop, J. Bookhold, M. Shiraiwa and U. Poschl, *Phys. Chem. Chem. Phys.*, 2011, **13**, 19238–19255.

61 M. A. Freedman, *Acc. Chem. Res.*, 2020, **53**, 1102–1110.

62 R. I. Masel, *Principles of Adsorption and Reaction on Solid Surfaces*, Wiley-Interscience, 1st edn, 1996.

63 D. M. Golden, G. N. Spokes and S. W. Benson, *Angew. Chem., Int. Ed. Engl.*, 1973, **12**, 534–546.

64 Y. Li, P. S. J. Lakey, M. J. Ezell, K. N. Johnson, M. Shiraiwa and B. J. Finlayson-Pitts, *ACS ES&T Air*, 2023, **1**, 52–61.

65 J. Abelard, A. R. Wilmsmeyer, A. C. Edwards, W. O. Gordon, E. M. Durke, C. J. Karwacki, D. Troya and J. R. Morris, *J. Phys. Chem. C*, 2016, **120**, 13024–13031.

66 J. Abelard, A. R. Wilmsmeyer, A. C. Edwards, W. O. Gordon, E. M. Durke, C. J. Karwacki, D. Troya and J. R. Morris, *J. Phys. Chem. C*, 2014, **119**, 365–372.

67 O. Sneh and S. M. George, *J. Phys. Chem.*, 1995, **99**, 4639–4647.

68 J. A. Ganske, L. M. Wingen, V. Perraud and B. J. Finlayson-Pitts, *ACS Earth Space Chem.*, 2019, **3**, 1315–1328.

69 G. Cavallo, P. Metrangolo, R. Milani, T. Pilati, A. Priimagi, G. Resnati and G. Terraneo, *Chem. Rev.*, 2016, **116**, 2478–2601.

70 C. R. Usher, A. E. Michel and V. H. Grassian, *Chem. Rev.*, 2003, **103**, 4883–4940.

71 O. A. Choobari, P. Zawar-Reza and A. Sturman, *Atmos. Res.*, 2014, **138**, 152–165.

72 A. P. Ault, V. H. Grassian, N. Carslaw, D. B. Collins, H. Destaillats, D. J. Donaldson, D. K. Farmer,

J. L. Jimenez, V. F. McNeill, G. C. Morrison, R. E. O'Brien, M. Shiraiwa, M. E. Vance, J. R. Wells and W. Xiong, *Chem*, 2020, **6**, 3203–3218.

73 J. M. Mattila, C. Arata, A. Abeleira, Y. Zhou, C. Wang, E. F. Katz, A. H. Goldstein, J. P. D. Abbatt, P. F. DeCarlo, M. E. Vance and D. K. Farmer, *Environ. Sci. Technol.*, 2022, **56**, 109–118.

74 F. Caloz, F. F. Fenter, K. D. Tabor and M. J. Rossi, *Rev. Sci. Instrum.*, 1997, **68**, 3172–3179.

75 J. Mønster, T. Rosenørn, O. J. Nielsen and M. S. Johnson, *Environ. Sci. Pollut. Res. Int.*, 2002, **1**, 63–67.

76 M. Shiraiwa, R. M. Garland and U. Pöschl, *Atmos. Chem. Phys.*, 2009, **9**, 9571–9586.

77 P. S. J. Lakey, G. C. Morrison, Y. Won, K. M. Parry, M. von Domaros, D. J. Tobias, D. Rim and M. Shiraiwa, *Commun. Chem.*, 2019, **2**, 56–63.

78 H. Y. Fan, E. S. Frank, P. S. J. Lakey, M. Shiraiwa, D. J. Tobias and V. H. Grassian, *J. Phys. Chem. C*, 2022, **126**, 6267–6279.

79 Y. Fang, P. S. J. Lakey, S. Riahi, A. T. McDonald, M. Shrestha, D. J. Tobias, M. Shiraiwa and V. H. Grassian, *Chem. Sci.*, 2019, **10**, 2906–2914.

80 L. B. Huang, E. S. Frank, S. Riahi, D. J. Tobias and V. H. Grassian, *J. Chem. Phys.*, 2021, 154.

81 M. von Domaros, P. S. J. Lakey, M. Shiraiwa and D. J. Tobias, *J. Phys. Chem. B*, 2020, **124**, 3836–3843.

82 R. C. Hoffman, M. E. Gebel, B. S. Fox and B. J. Finlayson-Pitts, *Phys. Chem. Chem. Phys.*, 2003, **5**, 1780–1789.

83 K. S. W. Sing, in *Adsorption by Powders and Porous Solids*, ed. F. Rouquerol, J. Rouquerol, K. S. W. Sing, P. Llewellyn and G. Maurin, Academic Press, 2nd edn, 2014, pp. 237–268, DOI: [10.1016/b978-0-08-097035-6.00007-3](https://doi.org/10.1016/b978-0-08-097035-6.00007-3).

84 S. Seisel, Y. Lian, T. Keil, M. E. Trukhin and R. Zellner, *Phys. Chem. Chem. Phys.*, 2004, **6**, 1926–1932.

85 S. L. Tait, Z. Dohnalek, C. T. Campbell and B. D. Kay, *J. Chem. Phys.*, 2005, **122**, 164707.

86 P. A. Redhead, *Vacuum*, 1962, **12**, 203–211.

87 R. J. Gorte, *J. Catal.*, 1982, **75**, 164–174.

88 D. D. Beck, J. M. White and C. T. Ratcliffe, *J. Phys. Chem.*, 1986, **90**, 3123–3131.

89 R. J. Cvetanović and Y. Amenomiya, *Catal. Rev.*, 1972, **6**, 21–48.

90 J. L. Falconer and J. A. Schwarz, *Catal. Rev.*, 2007, **25**, 141–227.

91 M. B. Hugenschmidt, L. Gamble and C. T. Campbell, *Surf. Sci.*, 1994, **302**, 329–340.

92 A. R. Wilmsmeyer, J. Uzarski, P. J. Barrie and J. R. Morris, *Langmuir*, 2012, 10962–10967, DOI: [10.1021/la301938f](https://doi.org/10.1021/la301938f).

93 Y. Fang, S. Riahi, A. T. McDonald, M. Shrestha, D. J. Tobias and V. H. Grassian, *J. Phys. Chem. Lett.*, 2019, **10**, 468–473.

94 A. Budi, S. L. S. Stipp and M. P. Andersson, *J. Phys. Chem. C*, 2018, **122**, 8236–8243.

95 H. Noller and G. Ritter, *J. Chem. Soc., Faraday Trans.*, 1984, **80**, 275–283.

96 P. J. Ziemann and R. Atkinson, *Chem. Soc. Rev.*, 2012, **41**, 6582–6605.