


Cite this: *RSC Adv.*, 2023, 13, 15714

Enhanced volatile fatty acid production from waste activated sludge by urea hydrogen peroxide: performance and mechanisms

Siyi Wang,^a Tianbing Jiang,^a Xiaoguo Chen,^{ab} Kai Xiong^a and Yanzhe Wang^a

Anaerobic acidogenesis of waste activated sludge (WAS) presents significant potential for resource recovery and waste treatment. However, the slow hydrolysis of WAS limits the efficiency of this approach. In this study, we applied urea hydrogen peroxide (UHP) pretreatment to enhance WAS hydrolysis and investigated the effects of operating parameters on volatile fatty acid (VFA) production and the associated mechanisms. Results demonstrated that UHP significantly improved WAS hydrolysis and VFA production, with a three-fold increase in soluble chemical oxygen demand (SCOD) compared to the control group. UHP dosage emerged as the most critical factor for VFA production, with the maximum VFA concentration increasing from 1127.6 to 8800.9 mg COD per L as UHP dosage ranged from 0 to 6 mmol g⁻¹ VSS (Volatile suspended solids). At an optimal UHP dosage of 4 mmol g⁻¹ VSS, both the unit oxidant promotion efficiency ($\Delta\text{VFAs}/\Delta\text{UHP}$) and the maximum VFA concentration reached relatively high levels, at 35.3 mg COD per mmol and 7527.3 mg COD per L, respectively. UHP pretreatment generated alkaline conditions, H₂O₂, ·OH and free ammonia, which collectively disrupted the extracellular polymeric substances (EPS) structure, transforming unextractable EPS into extractable forms and promoting the release of organic matter during both the pretreatment and fermentation stages. Excitation-emission matrix (EEM) analysis revealed that UHP increased the concentration of easily utilizable organic matter, providing more substrates for acidogenic bacteria and enhancing VFA production. Furthermore, weak alkaline conditions and high free ammonia concentrations in the UHP group facilitated VFA accumulation by preventing rapid acidification and suppressing methanogen activity. This study offers valuable insights into the potential of UHP pretreatment for enhancing WAS hydrolysis and VFA production, with promising applications in wastewater treatment and resource recovery.

Received 17th April 2023

Accepted 17th May 2023

DOI: 10.1039/d3ra02538a

rsc.li/rsc-advances

1 Introduction

The activated sludge process is widely used in wastewater treatment plants (WWTP), during which a large amount of waste activated sludge (WAS) is generated. WAS usually contains plenty of unstable organic matter and harmful components such as heavy metals, organic micropollutants and pathogens¹ and therefore needs to be treated to prevent its environmental risk. Currently, the main treatment processes are sanitary landfill, incineration, composting and anaerobic digestion.² Due to the huge volume and high treatment cost, the disposal of WAS takes up to 60% of a WWTP total operation cost.³ WAS contains a large amount of organic carbon and inorganic nutrients, making it an important renewable resource. As anaerobic digestion (AD) can recover resources and energy through biogas production while stabilizing sludge, it is

considered a cost-effective approach for both stabilizing sludge and recovering resources.⁴ However, a large amount of biogas produced is flared directly because of moisture and impurities that make it difficult to utilize.⁵

During AD of WAS, volatile fatty acids (VFAs) are also produced as intermediate products, which can be utilized as alternative carbon sources for biological denitrification⁶ or as building blocks for high-value products such as biodiesel⁷ and polyhydroxyalkanoates.⁸ Compared to biogas, VFAs provide a higher commercial value and require less fermentation time.⁴ In addition, VFAs production reduce greenhouse gas emissions through recovery of carbon sources and contribute to achieving zero carbon emissions from WWTP. Therefore, recovering VFAs through anaerobic fermentation provides a promising WAS treatment approach to compensate operating costs and reduce greenhouse gas emissions.⁴

However, the slow hydrolysis rate of WAS reduces the yield and production rate of VFAs from anaerobic fermentation, since most of the organics are present in sludge cells or entangled in the extracellular polymeric substances (EPS) matrix, which cannot be directly used by VFAs-forming bacteria.⁴ To enhance

^aSchool of Resources and Environmental Engineering, Wuhan University of Technology, Wuhan 430070, China

^bHubei Key Laboratory of Mineral Resources Processing and Environment, Wuhan 430070, China. E-mail: xiaoguo_chen@whut.edu.cn


the hydrolysis of WAS, several pretreatments have been proposed, such as physical, chemical, and biological pretreatments.⁹ Oxidative pretreatments such as potassium permanganate,¹⁰ potassium permanganate/sodium sulfite,¹¹ persulfate,¹² and hydrogen peroxide,¹³ are proved to effectively promote VFAs production through enhancing the hydrolysis of WAS, indicating that oxidative pretreatment is a promising pretreatment technology for VFAs production. However, these traditional oxidizers have limitations such as transportation difficulties, safety hazards, and introducing exogenous elements, making post-treatment of fermentation broth and residues more difficult.

Urea hydrogen peroxide (UHP) is an inexpensive and environmentally friendly solid oxidizer that has been successfully used for the treatment of organic contaminants in water.¹⁴ Recent research indicates that UHP treatment can effectively enhance the dewatering performance of WAS¹⁵. UHP has a higher reactive oxygen content and greater oxidation capacity than other peroxide counterparts,¹⁶ which can facilitate the dissolution of organic matter in WAS¹⁵, making it a potentially promising pretreatment technology for VFAs production. However, there is no report on UHP pretreatment of WAS for VFAs production.

The aims of this paper were to: (a) explore whether UHP pretreatment enhance the hydrolysis of WAS and thus increase the accumulation of VFAs; (b) determine the factors affecting the pretreatment efficiency such as pretreatment time, UHP dosage and initial pH; (c) unravel the mechanism of UHP pretreatment enhancing WAS hydrolysis and VFAs production.

2 Materials and methods

2.1 WAS and reagents

The sludge was sampled from the sludge dewatering unit of a domestic WWTP located in Wuhan, China, which operates using a double effluent (DE) oxidation ditch system. Its main characteristics are as follows: total suspended solids (TSS) of $15.6 \pm 0.1\%$, volatile suspended solids (VSS) of $8.3 \pm 0.1\%$, total solid nitrogen of $16.0 \pm 1.0 \text{ mg g}^{-1}$ TSS, total solid phosphorus of $23.0 \pm 0.7 \text{ mg g}^{-1}$ TSS, and total organic carbon of $30.6 \pm 1.7 \text{ mg g}^{-1}$ TSS. The sludge suspension with 5% TSS was prepared by adding pure water to the WAS and its total chemical oxygen demand (TCOD) is $31\,373.3 \pm 1882.4 \text{ mg L}^{-1}$. The UHP ($\text{CH}_6\text{N}_2\text{O}_3$, $\geq 99.5\%$) used in this study was purchased from Condice (Wuhan, China), and all other reagents used were of analytical grade and obtained from Sinopharm Chemical Reagent Co., Ltd (Shanghai, China).

2.2 Experimental procedure

The experiment was conducted using serum bottles of 500 mL as reactors. The reactors were initially fed with 300 mL of 5% TSS WAS suspension and supplemented with 2 mmol UHP per g VSS. The reactors were then incubated at 25 °C and 120 rpm for 1 hour for pretreatment, followed by flushing with ultra-high purity nitrogen for 3 min and incubated at 35 °C and 120 rpm for anaerobic fermentation. No additional inoculation was

performed and the pH was not controlled during the entire experiment. Samples were taken before and after pretreatment and periodically during anaerobic fermentation. Sampling was carried out under the protection of high-purity nitrogen gas. After the sampling was completed, nitrogen gas was continuously introduced for an additional 3 minutes to ensure that the reactor maintains an anaerobic environment. The samples were centrifuged at 8000 rpm for 10 min and the supernatant was filtered through a $0.45 \mu\text{m}$ cellulose acetate membrane for VFAs, soluble chemical oxygen demand (SCOD), dissolved organic matter (DOM), and NH_4^+-N determination, while the precipitate was used for loosely bond EPS (LB-EPS) and tightly bond EPS (TB-EPS) analysis. The control experiments followed the same procedures as described above, with the exception that no UHP was added.

To examine the effects of various UHP pretreatment factors on WAS hydrolysis and VFAs production, we altered the pretreatment time, UHP dosage, and initial pH independently. The pretreatment time was assessed using incubation times of 1, 6, 12, 24 and 48 hours. Different dosages of UHP (0, 0.5, 1, 2, 4, and 6 mmol g^{-1} VSS) were added to the reaction system. Prior to UHP addition for pretreatment, the initial pH values of the WAS suspension were adjusted to 5.0, 7.0, 9.0 and 11.0 using NaOH or HCl, as needed. All experiments were carried out in triplicate, and the average values with standard deviations were reported.

To determine whether $\cdot\text{OH}$ is generated during the UHP pretreatment of WAS and its contribution to WAS hydrolysis, *tert*-butanol (TBA) was used as a quenching agent for the $\cdot\text{OH}$ radical scavenging experiment. In the experimental group, 1 mmol UHP per g VSS + 0.1 mmol TBA per g VSS was added, while the control group only received 1 mmol UHP per g VSS. The experimental procedure was the same as described earlier. Samples were taken before and after the pretreatment and the SCOD was measured.

2.3 Analytical methods

The SCOD, TCOD and NH_4^+-N were analyzed using the standard methods.¹⁷ The pH was measured in the reactors using a Sartorius PB-10 pH meter (Sartorius, Germany). The EPS extraction method was performed as previously reported.¹⁸ The EPS protein (PN) content was measured using the Coomassie Brilliant Blue G-250 method, with bovine serum albumin (BSA) as a standard.¹⁹ The polysaccharide (PS) content of EPS was determined using the anthrone-sulfuric acid method.²⁰

The excitation-emission matrix (EEM) spectra were determined using an EEM spectrofluorometer (F-380, Gangdong, Tianjin, China) with emission wavelength and excitation wavelength ranging from 200 nm to 600 nm at 0.5 nm increments. The band-pass of excitation and emission slits was set at 5 nm, and all measurements were carried out at a scanning speed of $12\,000 \text{ nm min}^{-1}$. The EEM fluorescence spectra data was processed according to ref. 21.

The composition of VFAs was analyzed using an Agilent 7820A gas chromatograph (Agilent, USA) equipped with a flame ionization detector and a DB-FFAP column ($30 \text{ m} \times 250 \mu\text{m} \times$



0.25 μm). A sample injection volume of 1.0 μL was used, and 2-ethylbutyric acid was used as an internal standard. The oven temperature program employed for separation of VFAs was as follows: started at 60 $^{\circ}\text{C}$, linearly increased by 12 $^{\circ}\text{C min}^{-1}$ to 160 $^{\circ}\text{C}$, and then hold at 160 $^{\circ}\text{C}$ for 3 min. The VFAs concentration was converted into the COD concentration using the conversion factors.²²

2.4 Data analysis

Statistical significance between experimental groups was determined using an independent samples *t*-test in IBM SPSS Statistics 26 software, with a significance level of $P < 0.05$.

The rate of sludge hydrolysis (V) was calculated using eqn (1).

$$V = \frac{\text{SCOD}_t - \text{SCOD}_0}{t} \quad (\text{mg L}^{-1} \text{ h}^{-1}) \quad (1)$$

where SCOD_0 and SCOD_t represent the concentrations of SCOD before and after pretreatment, respectively, and t is the pretreatment time.

The degree of sludge disintegration (DD_{COD}) was calculated using eqn (2) to express the increase of soluble organic matter in released organic matter.²³

$$\text{Disintegration degree } (\text{DD}_{\text{COD}}) = \frac{\text{SCOD}_t - \text{SCOD}_0}{\text{TCOD} - \text{SCOD}_0} \times 100\% \quad (2)$$

Here, SCOD_t represents the SCOD of the treated sludge, SCOD_0 represents the SCOD of the raw sludge, and TCOD represents the TCOD of the raw sludge.

3 Results and discussions

3.1 Enhancement of WAS hydrolysis and VFAs production by UHP pretreatment

3.1.1 Promotion of WAS hydrolysis by UHP. Previous studies have suggested that SCOD is a reliable indicator for evaluating the effectiveness of WAS pretreatment.²⁴ Therefore, SCOD concentrations before and after UHP pretreatment were measured, and the results are presented in Fig. 1A. After one hour of pretreatment, the SCOD of the UHP group increased from 468.5 to 2334.0 mg L^{-1} , higher than the control group, which only increased to 871.0 mg L^{-1} . The SCOD after UHP-pretreatment was 2.7 times that of the control group. Consistently, the DOM (including soluble PN and soluble PS) of the

UHP group after pretreatment was 2.8 times that of the control group (data not shown), indicating that UHP pretreatment can promote WAS hydrolysis, significantly increasing the release of PN and PS. The DD_{COD} of the UHP group was 6.04%, significantly higher than the control group (1.30%). Moreover, the rate of sludge hydrolysis for the UHP group reached 1865.5 $\text{mg (L}^{-1} \text{ h}^{-1})$, while the control group only achieved 402.5 $\text{mg (L}^{-1} \text{ h}^{-1})$. These results demonstrate that UHP pretreatment can directly enhance WAS hydrolysis during the pretreatment process by increasing the degree of hydrolysis and accelerating the hydrolysis rate.

As shown in Fig. 1A, the SCOD in UHP group further increased substantially during fermentation, reaching a maximum value of 8453.3 mg L^{-1} , while the control group remained consistently low, ranging from 1236.7 to 2103.3 mg L^{-1} . The maximum DD_{COD} value in the UHP group reached 25.84%, significantly higher than the control group's 5.29%. These results suggest that UHP pretreatment has a strong promoting effect on the hydrolysis of WAS during the fermentation phase, and this effect is even more significant than during the pretreatment phase. Zhang *et al.*²⁵ investigated the pretreatment efficiency of CaO_2 and found that the maximum SCOD of 0.3 g CaO_2 per g VSS treatment group was 1.7 times higher than that of control group. Wang *et al.*²⁶ treated WAS with 0.1 g sodium percarbonate (SPC) per g TSS, and showed that the SCOD of SPC-treated group was 1.7 times higher than that of control group. The oxidant dosage used in this study was only 0.19 g g^{-1} VSS (equal to 0.1 g g^{-1} TSS), which was lower than or comparable to those in the literature. Nonetheless, the SCOD increase was 3.0 times, higher than that in these references. These findings indicate that UHP is a more effective oxidant for enhancing WAS hydrolysis.

3.1.2 Promotion of VFAs production by UHP. The impact of UHP pretreatment on VFAs production during WAS fermentation was assessed by measuring the concentrations and composition of VFAs, and the results are shown in Fig. 1B. Consistent with the trends observed in SCOD, the UHP group demonstrated a significant increase in VFAs concentration, reaching a maximum of 5665.7 mg COD per L , which is five times the maximum concentration of 1127.6 mg COD per L in the control group. This result indicates that UHP pretreatment can effectively enhance VFAs production during WAS fermentation. This enhancement effect is noticeably superior to that of oxidizing agents such as ferrates and percarbonates, but not as effective as potassium permanganate and peroxymonosulfate (Table 1). In comparison to pretreatment methods using potassium permanganate and peroxymonosulfate, UHP does not introduce exogenous elements that could impose additional costs for treating fermentation broth and residue. Therefore, utilizing UHP for WAS pretreatment is a more economical and efficient approach in practice. The VFAs production rate in the UHP group was faster than that in the control group, with maximum VFAs production achieved on day 5, compared to day 7 for the control group. Upon reaching maximum VFAs production, the VFAs/SCOD ratio in the UHP group was 69.2%, higher than the 53.6% observed in the control group. This indicates that the organic matter released by WAS following

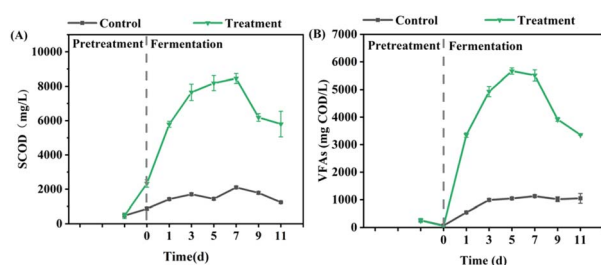


Fig. 1 Performances of pretreatment on (A) WAS hydrolysis and (B) VFAs production after 1 h pretreatment with 2 mmol UHP per g VSS.



Table 1 Effects of various oxidant pretreatments on VFAs production during sludge fermentation

Oxidants	Dosage (mg g ⁻¹ VSS)	VFAs yield (mg COD per g VSS)			References
		Control	Treatment	Treatment/Control	
KMnO ₄	299	34	295	8.7	10
K ₂ FeO ₄	500	135	254	1.9	28
K ₂ FeO ₄	500	46	177	3.8	29
Ca(ClO) ₂	16.4	48	174	3.6	30
K ₂ SO ₄ ·KHSO ₄ ·2KHSO ₅	147.1	29	311	10.7	12
Na ₂ CO ₃ ·1.5H ₂ O	197.5	155	445	2.9	26
Na ₂ CO ₃ ·1.5H ₂ O	161.3	24	64	2.7	27
CaO ₂	120	315	456	1.5	31
H ₂ O ₂	60.1	315	349	1.1	31
CO(NH ₂) ₂ ·H ₂ O ₂	188	43	215	5.0	This study

UHP treatment underwent a more complete transformation into VFAs. The increased VFAs yield and production rate could offer various benefits, such as reducing equipment volume and operational costs from an engineering perspective.²⁷

3.1.3 Impact of UHP on VFAs composition. UHP pretreatment significantly impacted the composition of VFAs during the fermentation period. In the control group, the primary components shifted from acetic acid (38.2%) and iso-valeric acid (38.2%) at the onset of fermentation to a predominance of propionic acid by day 5, with the proportion of acetic acid dropping to 7.3%. In contrast, the UHP group maintained a dominance of acetic acid throughout the first 7 days of fermentation. Although the proportion of acetic acid decreases as fermentation progresses, it still remained near 50% until the maximum VFAs production was reached on day 5 (Fig. 2A). During this process, the concentration of acetic acid consistently increased, reaching 2807.3 mg COD per L on day 5, which was significantly higher than the 69.0 mg COD per L observed in the control group (Fig. 2B). Acetic acid is considered the optimal additional carbon source for biological nitrogen and phosphorus removal.⁶ The increased proportion and production of acetic acid make it more advantageous as a carbon source for biological nitrogen and phosphorus removal.

In the later stages of fermentation, both the proportion and concentration of acetic acid in the UHP group decreased

substantially. During this period, the pH ranged from 7.67 to 7.91, and the free ammonia (FA) concentration varied between 87.5 and 163.7 mg L⁻¹. Given that high pH and FA levels may inhibit the methanogenesis process,^{32,33} the decline in acetic acid concentration is less likely to be caused by methanogenesis and more likely due to microbial utilization of acetic acid for polyhydroxyalkanoate synthesis.^{34–36}

3.2 Effect of UHP pretreatment conditions on anaerobic fermentation of WAS

3.2.1 Effect of pretreatment duration. The influence of pretreatment time on WAS hydrolysis and VFAs production is shown in Fig. 3. Within the experimental pretreatment time range, the effect of pretreatment time on sludge hydrolysis and VFAs production was not significant (Fig. 3A). Particularly, after 1–24 h UHP pretreatment, there was no significant difference in SCOD among the experimental groups, all in the range of 4850.0–5766.7 mg L⁻¹. In comparison, the SCOD of the 48 h experimental group was higher, reaching 6900.0 mg L⁻¹. The overall change pattern of each experimental group during the fermentation period was similar, with SCOD gradually increasing with the progress of fermentation, reaching the maximum on day 9, and no significant difference in SCOD among the experimental groups, ranging from 11 570.0 to 12 603.3 mg L⁻¹. The change in VFAs was similar to that of SCOD, increasing with the extension of fermentation time and reaching the maximum concentration on day 9, with concentrations ranging from 7063.3 to 7787.1 mg COD per L and no significant difference among the experimental groups (Fig. 3B).

Interestingly, a substantial amount of VFAs was generated after UHP pretreatment and prior to fermentation. The concentration of VFAs increased considerably as the pretreatment duration extended, with the 48 hour group showing an increase from 23.9 mg COD per L before treatment to 2411.2 mg COD per L after treatment. Notably, the primary component of the produced VFAs was acetic acid, accounting for over 50% of the total VFAs. This suggests that a high concentration of UHP (6 mmol UHP per g VSS) during the pretreatment stage not only enhances the release of organic matter in WAS but also directly

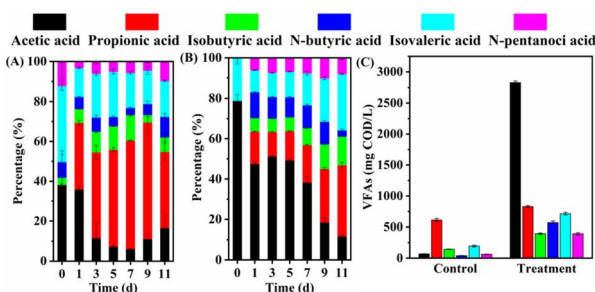


Fig. 2 Changes in VFAs composition after 1 h of 2 mmol UHP per g VSS pretreatment for (A) control group and (B) UHP group, and (C) the VFAs composition under the optimal acid production conditions for each respective group.

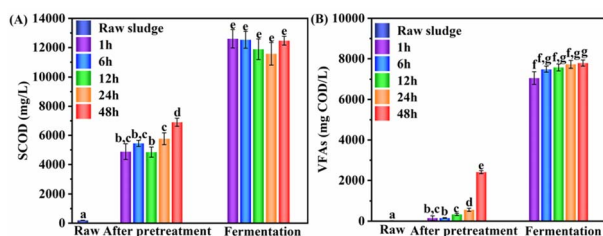


Fig. 3 Effect of pretreatment duration on (A) WAS hydrolysis and (B) VFAs production with 6 mmol UHP per g VSS. Data in fermentation phase were collected at their maximum concentration time (day 9).

oxidizes a portion of the organic matter into VFAs, predominantly acetic acid. However, a similar trend was not observed in Fig. 1B, where both the UHP group and the control group experienced a notable decrease in VFAs after pretreatment. This discrepancy could be attributed to the higher initial VFAs concentration (256.6 mg COD per L) and the lower UHP dosage (2 mmol UHP per g VSS) in this particular reaction system. The low concentration of UHP resulted in a smaller quantity of VFAs produced through the oxidation process. Simultaneously, the high initial VFAs concentration led to increased losses due to volatilization during the pretreatment process, ultimately causing a net decrease in VFAs concentration.

3.2.2 Effect of UHP dosage. In contrast to the effects of pretreatment time on WAS hydrolysis and VFAs production, the UHP dosage had a significant impact on both aspects, with hydrolysis and VFAs production increasing markedly as the UHP dosage increased (Fig. 4). When the UHP dosage was below 4 mmol g⁻¹ VSS, $\Delta\text{SCOD}/\Delta\text{UHP}$ increased as the dosage increased, reaching a maximum value of 48.2 mg mmol⁻¹ at a dosage of 4 mmol g⁻¹ VSS. However, further increasing the dosage to 6 mmol g⁻¹ VSS led to a substantial reduction in $\Delta\text{SCOD}/\Delta\text{UHP}$, dropping to 7.6 mg mmol⁻¹. A similar trend was observed during the fermentation stage, with $\Delta\text{SCOD}/\Delta\text{UHP}$ decreasing most rapidly when the UHP dosage increased from 4 mmol g⁻¹ VSS to 6 mmol g⁻¹ VSS, falling from 130.8 mg mmol⁻¹ to 19.6 mg mmol⁻¹. Comparable results were obtained in studies using potassium ferrate²⁹ and Ca(ClO)₂³⁰ for WAS pretreatment. Furthermore, when the Ca(ClO)₂ dosage exceeded a certain threshold (e.g., 0.2 g g⁻¹ TSS), the SCOD

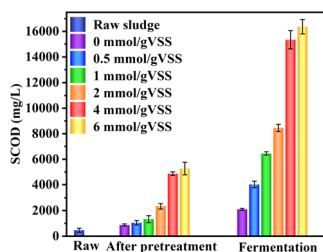


Fig. 4 Variations in SCOD for WAS with different UHP dosages after 1 h pretreatment, Data in fermentation phase were collected at their maximum concentration time.

concentration even dropped lower than that of the blank control group.³⁰ This might be attributed to the excessive oxidant reacting with the dissolved organic matter, oxidizing it, and subsequently causing a decrease in SCOD.²⁹

The maximum VFAs concentration during fermentation increased from 1127.6 to 8800.9 mg COD per L with the increase in UHP dosage from 0 to 6 mmol g⁻¹ VSS. A similar trend was observed in studies involving potassium monopersulfate (PMS)³⁷ and CaO₂.³⁸ However, as the UHP dosage increased, $\Delta\text{VFAs}/\Delta\text{UHP}$ initially increased and then decreased. At a UHP dosage of 4 mmol g⁻¹ VSS, both $\Delta\text{VFAs}/\Delta\text{UHP}$ and the maximum VFAs concentration reached relatively high levels, at 35.3 mg COD per mmol and 7527.4 mg COD per L, respectively. Further increasing the dosage to 6 mmol g⁻¹ VSS raised the VFAs to 8800.9 mg COD per L, but led to a significant decrease in $\Delta\text{VFAs}/\Delta\text{UHP}$, down to 24.2 mg COD per mmol. Taking into account the final VFAs yield and the efficiency of increasing VFAs per unit of UHP, the optimal UHP dosage was determined to be 4 mmol g⁻¹ VSS.

Fig. 5B demonstrates that UHP pretreatment significantly influenced the composition of VFAs. In the control group without UHP, propionic acid was the main component of VFAs, accounting for 54.3%, followed by isovaleric acid at 17.4%, and acetic acid at a mere 6.1% (69.0 mg COD per L). The introduction of 0.5 mmol UHP per g VSS had minimal impact on VFAs composition. However, as the UHP dosage increased, acetic acid's proportion rose substantially while propionic acid's proportion decreased. When the UHP addition exceeded 2 mmol g⁻¹ VSS, VFAs composition became relatively stable, with acetic acid having the highest proportion (49.3–56.2%), followed by isovaleric acid (12.5–14.0%) and propionic acid (11.5–14.8%). Acetic acid production increased with UHP addition, from 2807.3 mg COD per L to 4944.7 mg COD per L, when UHP addition ranged between 2 and 6 mmol g⁻¹ VSS. Concurrently, the time required for acetic acid and VFAs to reach their maximum production extended from 5 to 9 days.

Similar patterns were observed with other oxidants. He *et al.*³⁹ treated WAS with potassium ferrate (PF) and found that the proportion of acetic acid increased from 13.1% in the control to 18.4, 26.3, 48.2, 54.0, and 68.4% with PF dosages of 14, 28, 56, 84, and 140 mg Fe(vi) per g TSS, respectively. Yang *et al.*¹² reported that when PMS addition increased from 0 to 0.09 g g⁻¹ TSS, the proportion of acetic acid rose from 17.3% to 60.3%, and when the addition further increased to 0.24 g g⁻¹

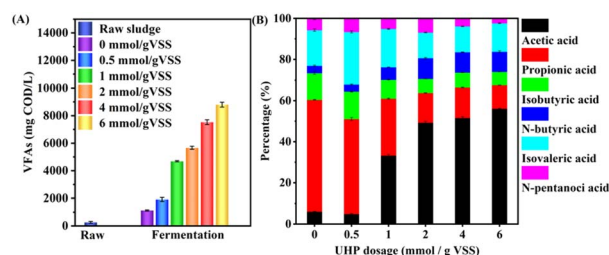


Fig. 5 Variations in (A) VFAs concentration and (B) VFAs composition at the time of maximum VFAs production with different UHP dosages.



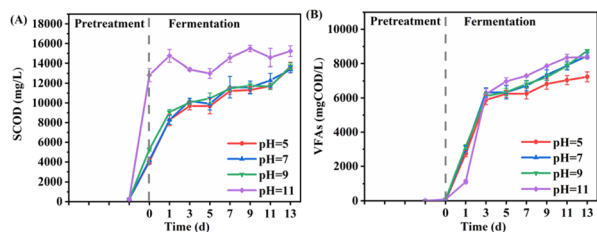


Fig. 6 Effect of initial pH on (A) WAS hydrolysis and (B) VFAs production after 1 h pretreatment with 6 mmol UHP per g VSS.

TSS, the proportion of acetic acid declined to 40.0%. These results indicate that increasing oxidant dosage indeed promotes the production and accumulation of acetic acid. This may be attributed to the conversion of higher molecular weight VFAs (e.g., butyric acid, propionic acid, and valeric acid) to acetic acid through oxidant treatment⁴⁰ or biodegradation to acetic acid in the anaerobic fermentation system by acetic acid-forming bacteria.^{41,42}

3.2.3 Effect of initial pH in pretreatment. Considering the potential impact of pH on UHP hydrolysis rate, and consequently, its pretreatment efficacy on WAS, the influence of initial pH during pretreatment on WAS hydrolysis and subsequent VFAs production was investigated. As shown in Fig. 6, under alkaline conditions, particularly at pH 11, the release of organic matter from WAS during pretreatment was substantially enhanced, resulting in an SCOD of 12 830.0 mg L⁻¹, significantly higher than that observed under neutral and acidic conditions (4165.0 mg L⁻¹ and 4031.7 mg L⁻¹). However, during fermentation, the SCOD increase for the group pretreated at pH 11 was less pronounced, while other groups experienced a marked increase. By the end of the experiment (day 13), the concentrations reached 13 380.0–13 713.3 mg L⁻¹, approaching the 15 226.7 mg L⁻¹ found in the pH 11 group. These results suggest that the initial pH during pretreatment has limited impact on the ultimate hydrolysis extent of WAS.

In contrast to SCOD trends, VFAs production in the strongly alkaline group (pH = 11) was notably suppressed at the onset of fermentation (day 1) but quickly rebounded. This suppression might be attributed to the high pH (10.37), which inhibited acid-producing bacteria.⁴³ Subsequently, the VFAs production in this group increased rapidly, even exceeding other groups, possibly due to the higher hydrolysis extent of WAS providing ample carbon sources for fermentation. However, by the end of the experiment, VFAs production in this group was comparable to other groups, except for the pH 5 group, with values ranging between 8375.9 and 8747.2 mg COD per L. This finding indicates that while strongly alkaline conditions promote WAS hydrolysis, they do not enhance VFAs production. Maspolim *et al.*⁴³ reported similar conclusions. The total VFAs production in the acidic treatment group (pH = 5) was only 7236.9 mg COD per L, significantly lower than that in neutral and alkaline treatment groups. As the pH of this group remained above 8.0 in the later stage of fermentation, it is less likely that inhibition of acid-producing bacteria caused the observed difference, warranting further investigation into the specific cause.

3.3 Mechanisms of enhanced VFAs production by UHP pretreatment

3.3.1 UHP-enhanced WAS hydrolysis process. The extent of WAS hydrolysis is a crucial factor in determining the efficiency of VFAs production during fermentation. EPS is considered the primary component maintaining WAS floc structure, and its compositional changes have a crucial impact on the stability of WAS structure and hydrolysis efficiency. To investigate the mechanisms underlying UHP-promoted WAS hydrolysis, we analyzed changes in the content and composition of three types of EPS (DOM, LB-EPS, and TB-EPS) in WAS before and after pretreatment (Fig. 7).

Following UHP pretreatment, DOM increased considerably, indicating that some LB-EPS was converted into a soluble form. The TB-EPS/LB-EPS ratio after UHP pretreatment was 2.5, significantly lower than the 5.4 before pretreatment and smaller than the control group's 6.0, suggesting that some TB-EPS was transformed into LB-EPS. TB-EPS, as tightly bound EPS, primarily maintains the stability of the WAS structure. A decrease in the TB-EPS/LB-EPS ratio negatively impacts WAS stability. Moreover, PS and PN analysis in LB-EPS and TB-EPS revealed that the PS/PN of LB-EPS increased only slightly after UHP pretreatment (from 1.29 to 1.38), while the PS/PN of TB-EPS dropped sharply from 2.71 to 1.93. As the primary substances responsible for maintaining the rigid structure of sludge,⁴⁴ a substantial decrease in the proportion of PS would inevitably compromise the sludge's stability.

Previous research has shown that extractable EPS accounts for only a small portion of EPS in WAS, with the majority being non-extractable EPS.²⁹ Our study discovered that after UHP treatment, although DOM increased significantly, the extractable EPS (the sum of LB-EPS and TB-EPS) did not decrease but rather increased by 0.6 times (from 172.8 mg COD per L to 279.6 mg COD per L), while the control group experienced a slight decrease to 168.7 mg COD per L. This finding suggests that UHP treatment can convert non-extractable EPS in WAS into extractable EPS. Similar result was also found for co-treatment by free nitrous acid and calcium peroxide of WAS.⁴ In comparison to extractable EPS, non-extractable EPS is more

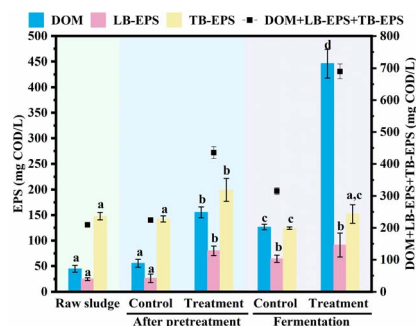


Fig. 7 Variations in DOM, LB-EPS, and TB-EPS after 1 h of 2 mmol UHP per g VSS pretreatment and on day 7 of fermentation. DOM, LB-EPS, and TB-EPS represent the sum of PN and PS concentrations. Distinct letters above bars denote significant differences ($p < 0.05$) for the same parameter.



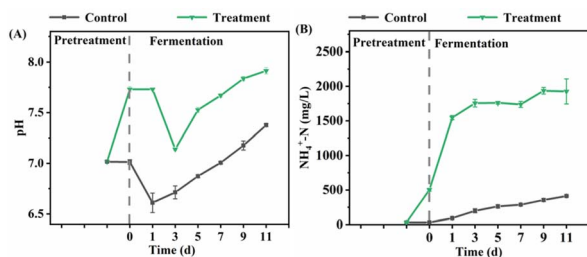


Fig. 8 Variations in (A) pH and (B) $\text{NH}_4^+\text{-N}$ after 1 h of pretreatment using 2 mmol UHP per g VSS.

tightly bound to WAS and contributes more significantly to WAS structure stability. By converting non-extractable EPS into extractable EPS, UHP further diminishes WAS stability. This destructive effect on WAS stability during the pretreatment process enhances further hydrolysis during the subsequent fermentation stage, aligning with the substantial increase in DOM observed in the UHP group during fermentation.

UHP can decompose in water, generating H_2O_2 and urea. Urea further hydrolyzes to produce ammonia nitrogen, leading to an increase in the system's pH. These intermediate substances may alter the composition of EPS, affecting the hydrolysis of WAS. To further determine how UHP changes the composition of EPS in WAS, the pH of the system before and after treatment was measured (Fig. 8A), and the concentrations of FA and H_2O_2 were theoretically calculated to analyze their possible impacts on EPS. The UHP used in this study can theoretically produce 52.8 mmol L^{-1} of H_2O_2 . In addition to its inherent oxidizing ability, H_2O_2 can generate highly active $\cdot\text{OH}$ under the catalysis of transition metals present in WAS¹³, although the amount of produced $\cdot\text{OH}$ is not very large¹⁵. To determine the contribution of $\cdot\text{OH}$ to the hydrolysis of WAS, we conducted a radical quenching experiment using the $\cdot\text{OH}$ scavenger TBA. The results showed that the addition of the quencher reduced the SCOD of UHP pretreated WAS by 11.3%, indicating that the $\cdot\text{OH}$ generated by the decomposition of UHP indeed had a certain promoting effect on the hydrolysis of WAS. These findings suggested that H_2O_2 and the produced $\cdot\text{OH}$ may break the bonds connecting the EPS skeleton through oxidation, thereby disrupting the EPS structure and accelerating its hydrolysis⁴⁵. This could be a crucial mechanism for UHP-promoted changes in EPS composition and organic matter release.

After UHP pretreatment, the pH value increased from 7.0 to 7.7, while the control group's pH remained around 7.0 (Fig. 8A). Alkaline conditions favor the dissociation of acidic groups in EPS, resulting in some EPS components carrying negative charges. The repulsive forces between these negatively charged components can promote the dissolution of PN and PS in EPS.⁴⁶ The significant increase in PN and PS concentrations in the DOM of the UHP group supports this hypothesis. Therefore, the elevation of pH promoting the dissolution of PN and PS in EPS may be another crucial mechanism for UHP pretreatment-enhanced EPS hydrolysis. The increased pH also converts

ammonium ions to FA. According to the reaction system's pH and ammonia nitrogen concentration,⁴⁷ the FA concentration after pretreatment in the UHP group was 14.9 mg L^{-1} , significantly higher than the 0.2 mg L^{-1} in the control group. Previous studies have shown that FA can attack tyrosine-like proteins and aromatic-like proteins in sludge EPS and thus destroy hydrogen bonding networks,⁴⁸ making the EPS structure loose or even dissolving.⁴⁹ This could be an essential reason for UHP causing non-extractable EPS to transform into extractable EPS and the conversion of TB-EPS to LB-EPS and DOM.

3.3.2 UHP-enhanced VFAs accumulation process. EEM analysis was conducted on liquid samples to assess the influence of UHP on the bioavailability of organic matter released from WAS, with results presented in Fig. 9. The fluorescence intensity (FI) changes across the five regions before and after pretreatment were relatively small, with region IV exhibiting the highest intensity, followed by region V (Fig. 9A and B). After UHP pretreatment, the FI of regions I and IV increased substantially, with their combined intensity increasing to 1.3 times that of the control group. During the fermentation period, the FI further increased, reaching 2.3 times that of the control group on the first day of fermentation (Fig. 9C). Regions IV and I represent dissolved microbial by-products and tyrosine-like proteins, respectively, both of which are considered easily utilizable by microbes.⁵⁰ Consequently, the increased concentration of these components provides more substrates for acidogenic bacteria, promoting VFAs production.

Although the proportion of regions IV + I in the total FI of the UHP group after pretreatment (74.1%) was slightly lower than that of the control group (75.2%), it was higher during the fermentation stage (80.3–80.6%) compared to the control group (79.1–79.8%). This finding indicates that UHP treatment did not negatively impact the bioavailability of hydrolysis products while enhancing WAS hydrolysis and even exhibited a promoting effect, which is advantageous for boosting VFAs

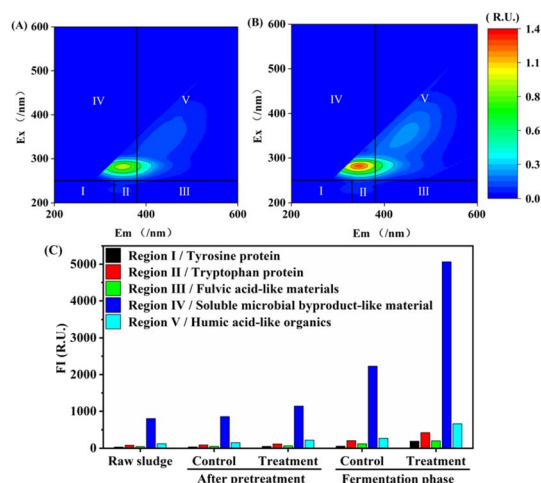


Fig. 9 EEM fluorescence spectra of WAS supernatant after 1 h pretreatment with 2 mmol UHP per g VSS in (A) control group and (B) UHP group, and (C) FRI after pretreatment and on day 1 of fermentation. All samples were diluted 15 times.



production. The higher VFAs/SCOD ratio in the UHP group compared to the control group supports this assertion.

Further investigation revealed that the $\Delta\text{PN}/\Delta\text{PS}$ of DOM in the UHP group was 1.1 after pretreatment, significantly higher than that in the control group (0.86) and higher than that in the initial WAS LB-EPS and TB-EPS (0.78 and 0.37, respectively). This result suggests that UHP pretreatment is more conducive to PN solubilization in WAS. The elevated PN proportion promotes acetic acid production during fermentation,⁵¹ resulting in a marked increase in both the concentration (2826.2 mg COD per L) and proportion (near 50%) of acetic acid in the UHP group, thereby enhancing overall VFAs production.

Since the pH range of the UHP group is 7.1–7.9 (Fig. 8A), it exhibits weak alkalinity, which can neutralize acids and prevent rapid acidification during fermentation from inhibiting acidogenic bacterial activity.⁵² This slightly alkaline environment is beneficial for the growth of acid-producing bacteria such as *Sedimentibacter*⁴³. Consequently, the elevated pH caused by the UHP treatment promotes the production of VFAs. Furthermore, the pH in the UHP group is higher than that of the control group (6.6–7.4) (Fig. 8A) and exceeds the optimal pH range for methanogenic bacteria (6.8–7.2),³² which is unfavorable for methanogen growth. Additionally, the high FA concentration (64.7–163.7 mg L⁻¹) in the UHP group during fermentation inhibits methanogenic enzyme activity⁵³. These UHP treatment characteristics restrict the conversion of VFAs to methane, thereby fostering the accumulation of VFAs.

4 Conclusions

This study demonstrated the effectiveness of UHP pretreatment in enhancing WAS hydrolysis and VFAs production. UHP pretreatment significantly increased the SCOD and VFAs concentration, with UHP dosage being the most crucial factor for optimal results. The mechanisms underlying the enhancement of WAS hydrolysis and VFAs production include the disruption of EPS structure, promotion of organic matter release, and increase in easily utilizable organic matter concentrations. Additionally, weak alkaline conditions and high free ammonia concentrations in the UHP group facilitated VFAs accumulation by preventing rapid acidification and inhibiting methanogen activity. The findings of this study provide valuable insights into the potential of UHP pretreatment for WAS hydrolysis and VFAs production enhancement, paving the way for innovative applications in wastewater treatment and resource recovery.

Conflicts of interest

There are no conflicts to declare.

Acknowledgements

We would like to express our gratitude to Jianmin Qiu, Zixun Qin and Yiting Cheng for their assistance in sample measurement.

References

- W. Fang, X. Zhang, P. Zhang, J. Wan, H. Guo, D. S. M. Ghasimi, X. C. Morera and T. Zhang, *J. Environ. Sci.*, 2020, **87**, 93–111.
- T. Liang, K. Elmaadawy, B. Liu, J. Hu, H. Hou and J. Yang, *Process Saf. Environ. Prot.*, 2021, **145**, 321–339.
- W. Q. Guo, S. S. Yang, W. S. Xiang, X. J. Wang and N. Q. Ren, *Biotechnol. Adv.*, 2013, **31**, 1386–1396.
- J. Wang, Y. Lou, K. Feng, H. Zhou, B. Liu, G. Xie and D. Xing, *J. Hazard. Mater.*, 2022, **423**, 127022.
- Q. Zeng, F. Zan, T. Hao, S. K. Khanal and G. Chen, *Water Res.*, 2022, **208**, 117839.
- Y. Wang, X. Liu, Y. Liu, D. Wang, Q. Xu, X. Li, Q. Yang, Q. Wang, B. J. Ni and H. Chen, *Bioresour. Technol.*, 2020, **318**, 124266.
- S. Dufreche, R. Hernandez, T. French, D. Sparks, M. Zappi and E. Alley, *J. Am. Oil Chem. Soc.*, 2007, **84**, 181–187.
- N. Frison, E. Katsou, S. Malamis, A. Oehmen and F. Fatone, *Environ. Sci. Technol.*, 2015, **49**, 10877–10885.
- Y. Yuan, X. Hu, H. Chen, Y. Zhou, Y. Zhou and D. Wang, *Sci. Total Environ.*, 2019, **694**, 133741.
- Q. Xu, Q. Fu, X. Liu, D. Wang, Y. Wu, Y. Li, J. Yang, Q. Yang, Y. Wang, H. Li and B. J. Ni, *Chem. Eng. J.*, 2021, **406**, 126797.
- W. Li, Y. Lou, A. Fang, K. Feng and D. Xing, *Chem. Eng. J.*, 2020, **394**, 124920.
- J. Yang, X. Liu, D. Wang, Q. Xu, Q. Yang, G. Zeng, X. Li, Y. Liu, J. Gong, J. Ye and H. Li, *Water Res.*, 2019, **148**, 239–249.
- Preethi, J. Rajesh Banu, S. Kavitha, R. Yukesh Kannah, S. Varjani and M. Gunasekaran, *Sci. Total Environ.*, 2022, **817**, 152873.
- Q. Chen, Z. Cheng, X. Li, C. Wang, L. Yan, G. Shen and Z. Shen, *Appl. Catal., B*, 2022, **302**, 120866.
- D. Wang, C. Pan, L. Chen, D. He, L. Yuan, Y. Li and Y. Wu, *Water Res.*, 2022, **225**, 119195.
- J. Zhou and H. Wang, *Chem. Eng. J.*, 2020, **390**, 124567.
- SEPA, *Water and Wastewater Monitoring Methods China*, Environmental Science Publishing House, Beijing, 2002.
- Y. Wang, X. Wang, K. Zheng, H. Guo, L. Tian, T. Zhu and Y. Liu, *Bioresour. Technol.*, 2022, **364**, 128024.
- R. W. Blakesley and J. A. Boezi, *Anal. Biochem.*, 1977, **82**, 580–582.
- B. Frølund, R. Palmgren, K. Keiding and P. H. Nielsen, *Water Res.*, 1996, **30**, 1749–1758.
- C. Yang, A. Zhou, Z. He, L. Jiang, Z. Guo, A. Wang and W. Liu, *Environ. Sci. Pollut. Res.*, 2015, **22**, 9100–9109.
- P. Jin, X. Wang, Q. Zhang, X. Wang, H. H. Ngo and L. Yang, *Bioresour. Technol.*, 2016, **200**, 722–730.
- H. Sari Erkan and G. Onkal Engin, *J. Environ. Chem. Eng.*, 2020, **8**, 103918.
- O. G. Apul and F. D. Sanin, *Bioresour. Technol.*, 2010, **101**, 8984–8992.
- Q. Zhang, X. Gao, Y. Jin, L. Zhao, H. Zhu and P. Zhang, *Bioresour. Technol.*, 2020, **309**, 123379.



- 26 Y. Wang, P. Sun, H. Guo, K. Zheng, T. Zhu and Y. Liu, *J. Environ. Manage.*, 2022, **313**, 115025.
- 27 Q. Zhang, X. Cheng, F. Wang, S. Fang, L. Zhang, W. Huang, F. Fang, J. Cao and J. Luo, *Sci. Total Environ.*, 2022, **838**, 156054.
- 28 L. Li, J. He, M. Wang, X. Xin, J. Xu and J. Zhang, *ACS Sustainable Chem. Eng.*, 2018, **6**, 16819–16827.
- 29 L. Li, J. He, X. Xin, M. Wang, J. Xu and J. Zhang, *Chem. Eng. J.*, 2018, **332**, 456–463.
- 30 Q. Zhang, Y. Wu, J. Luo, J. Cao, C. Kang, S. Wang, K. Li, J. Zhao, M. Aleem and D. Wang, *Sci. Total Environ.*, 2020, **712**, 136500.
- 31 J. Xu, X. Li, L. Gan and X. Li, *Sci. Total Environ.*, 2018, **644**, 547–555.
- 32 X. Wang and Y. C. Zhao, *Int. J. Hydrogen Energy*, 2009, **34**, 245–254.
- 33 C. Zhang, Y. Qin, Q. Xu, X. Liu, Y. Liu, B. J. Ni, Q. Yang, D. Wang, X. Li and Q. Wang, *ACS Sustainable Chem. Eng.*, 2018, **6**, 9120–9129.
- 34 H. Satoh, Y. Iwamoto, T. Mino and T. Matsuo, *Water Sci. Technol.*, 1998, **38**, 103–109.
- 35 H. Salehzadeh and M. C. M. Van Loosdrecht, *Biotechnol. Adv.*, 2004, **22**, 261–279.
- 36 K. Sudesh, H. Abe and Y. Doi, *Prog. Polym. Sci.*, 2000, **25**, 1503–1555.
- 37 B. Jin, J. Niu, J. Dai, N. Li, P. Zhou, J. Niu, J. Zhang, H. Tao, Z. Ma and Z. Zhang, *Bioresour. Technol.*, 2018, **265**, 8–16.
- 38 Y. Li, J. Wang, A. Zhang and L. Wang, *Water Res.*, 2015, **83**, 84–93.
- 39 Z. W. He, W. Z. Liu, Q. Gao, C. C. Tang, L. Wang, Z. C. Guo, A. J. Zhou and A. J. Wang, *Bioresour. Technol.*, 2018, **247**, 174–181.
- 40 A. Shanableh and S. Jones, *Water Sci. Technol.*, 2001, **44**, 129–135.
- 41 Z. W. He, C. X. Yang, L. Wang, Z. C. Guo, A. J. Wang and W. Z. Liu, *Chem. Eng. J.*, 2016, **290**, 125–135.
- 42 C. Yang, W. Liu, Z. He, S. Thangavel, L. Wang, A. Zhou and A. Wang, *Bioresour. Technol.*, 2015, **175**, 509–516.
- 43 Y. Maspolim, Y. Zhou, C. Guo, K. Xiao and W. J. Ng, *Bioresour. Technol.*, 2015, **190**, 289–298.
- 44 S. S. Adav, D. J. Lee and J. H. Tay, *Water Res.*, 2008, **42**, 1644–1650.
- 45 S. Pilli, S. Yan, R. Tyagi and R. Surampalli, *Rev. Environ. Sci. Biotechnol.*, 2015, **14**, 453–472.
- 46 J. Wingender, *Microbial Extracellular Polymeric Substances: Characterization, Structure and Function*, 2011.
- 47 D. Wang, Y. Huang, Q. Xu, X. Liu, Q. Yang and X. Li, *Bioresour. Technol.*, 2019, **275**, 163–171.
- 48 Q. Fu, S. Long, Y. Xu, Y. Wang, B. Yang, D. He, X. Li, X. Liu, Q. Lu and D. Wang, *J. Hazard. Mater.*, 2023, **452**, 131305.
- 49 J. Zhang, D. Wang, X. Li, X. Liu, Q. Yang, Q. Xu, G. Yang, A. Duan and Y. Fang, *J. Cleaner Prod.*, 2023, **389**, 135862.
- 50 X. Jia, C. Zhu, M. Li, B. Xi, L. Wang, X. Yang, X. Xia and J. Su, *Int. J. Hydrogen Energy*, 2013, **38**, 8691–8698.
- 51 Y. Chen, J. Luo, Y. Yan and L. Feng, *Appl. Energy*, 2013, **102**, 1197–1204.
- 52 Y. Lu, H. Yuan, B. Yan, X. Zuo and X. Li, *Biomass Bioenergy*, 2022, **164**, 106553.
- 53 H. Yuan and N. Zhu, *Renewable Sustainable Energy Rev.*, 2016, **58**, 429–438.

