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A novel dibenzo[*a,c*]phenazine-based fluorescent probe for fast and selective detection of thiophenols in environmental water†

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We herein report a novel dibenzo[*a,c*]phenazine-based fluorescent probe with fast response to thiophenols over a wide pH range from 5 to 13. The probe possesses a large Stokes shift (120 nm). More importantly, it displays a high selectivity and sensitivity for thiophenols in the presence of other analytes such as biothiols and common metal ions. A good linear relationship between the fluorescence intensity at 570 nm and the thiophenol concentration in the range of 0–20 μM was observed with a low detection limit at 40 nM. In addition, it has been successfully applied to detect thiophenols in environmental water (such as seawater, tap water and spring water) with high recovery, which confirmed its potential value in environmental monitoring.

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1. Introduction

Thiophenols are widely used in the production of pesticides, dyes, and medicines.^{1–3} However, they are also known as volatile toxic substances, and can cause serious damage to bioorganisms and environment, especially the water environment.⁴ The median lethal concentration (LC₅₀) of thiophenols to fish (0.01–0.04 mM) and mice is 0.01–0.04 mM and 46.2 mg kg^{−1}, respectively.^{5,6} Inhalation or ingestion of thiophenols can cause serious health problems in humans, such as muscle atrophy, vomiting, and even death.^{6,7} When thiophenols are incorporated into the water ecosystem, they can be absorbed by aquatic organisms, and ultimately harm human beings.⁸ Therefore, there is an urgent need to develop highly sensitive and selective methods to detect thiophenols in environmental samples.

Several approaches, such as high-performance liquid chromatography (HPLC), gas chromatography (GC), and capillary electrophoresis, have been developed separation and detection of thiophenols.^{9–13} Among numerous detecting methods of thiophenols, fluorescent method was considered as the most powerful tool for thiophenols detection, because of the intrinsic advantages such as high selectivity and high sensitivity.^{14–19} In these probes, 2,4-dinitrobenzenesulfonyl or 2,4-dinitrophenyl groups were employed as recognition moieties linked to various fluorophores.^{20–25} However, there are still some defects, such as

high pH-dependence of the fluorescence intensity, weak fluorescence enhancement after the addition of thiophenols, and relatively long response time, which limit the practical application of fluorescent method in environmental samples. Thus, it is still of great significance to explore novel applicable probes for rapid detection of thiophenols.

Dibenzo[*a,c*]phenazines is a kind of excellent organic fluorescent compounds with large Stokes shift, high quantum yields, good photostability, and easy synthesis, which were widely used in organic light-emitting diode (OLED) field.^{26–28} However, dibenzo[*a,c*]phenazines have so far received little attention in the field of fluorescent probe. Since the first report of 2,4-dinitrobenzenesulfonyl group as a fluorescence quencher,²⁹ a lot of fluorescent probes with 2,4-dinitrobenzenesulfonyl group as recognition site were synthesized and reported.^{30–33} Herein, we employed dibenzo[*a,c*]phenazin-11-amine as a fluorophore and 2,4-dinitrobenzenesulfonyl group as recognition site to develop a novel fluorescent probe for fast detecting of thiophenols in water. The probe exhibits high selectivity and sensitivity for thiophenols with a large Stokes shift (120 nm). It is noteworthy that this probe can work in a wide pH range from 5 to 13 with fast response. As far as we know, this is the first thiophenols fluorescent probe which can work in such a wide pH range. Moreover, the probe demonstrated its capability to monitor thiophenols in real water samples.

2. Experimental

2.1. Materials and instrumentation

All solvents and materials were purchased in analytical grade and used as received. ¹H NMR and ¹³C NMR spectra were based on a Bruker Avance-500 MHz NMR spectrometer (Bruker-

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Biospin, Billerica, MA, USA). High-resolution mass spectra were taken on WATERS MALDI SYNAPT G2 HDMS (MA, USA). The pH was measured by Ion analyzer (LEICI, China). UV-vis spectra were recorded on spectrophotometer (UV-1800, Shimadzu, Japan). Fluorescence spectra were captured on a Microplate Reader (Infinite M1000 Pro, Tecan, Switzerland). High performance liquid chromatographic (HPLC) analysis was performed on a Shimadzu HPLC system (LC-20, Shimadzu corp., Kyoto, Japan) equipped with a C18 column (Shimadzu InertSustain 4.6 I. D. \times 250 mm, 5 μ m, S/N 6LR98081).

2.2. Synthesis of probe-dbp

Probe-dbp was synthesized according to Scheme 1. The characterization data including ^1H NMR and ^{13}C NMR were given in the ESI.† To a solution of dbp-NH₂ (0.1 mmol) in dry pyridine (5 mL) at 0 °C, a solution of 2,4-dinitrobenzenesulfonyl chloride (0.15 mmol) in CH₂Cl₂ (5 mL) was added dropwise. The mixture was stirred at room temperature overnight. After solvent was evaporated, the crude product was purified by column chromatography to give the probe-dbp as a yellow solid (23.6 mg, yield 45%). ^1H NMR (500 MHz, DMSO-d₆) δ 9.20 (dd, J = 8.2 Hz, 1H), 9.18 (d, J = 8.2 Hz, 1H), 8.92 (d, J = 2.1 Hz, 1H), 8.77 (d, J = 7.2 Hz, 2H), 8.60 (dd, J = 8.8, 2.1 Hz, 1H), 8.40 (d, J = 8.8 Hz, 1H), 8.31 (d, J = 9.1 Hz, 1H), 7.95 (d, J = 2.1 Hz, 1H), 7.88 (d, J = 8.1 Hz, 1H), 7.85 (d, J = 8.1 Hz, 1H), 7.79 (m, 3H), 5.74 (s, 1H). ^{13}C NMR (125 MHz, DMSO-d₆) δ 150.73, 148.36, 142.62, 142.24, 141.45, 139.43, 136.36, 132.13, 132.04, 131.78, 131.51, 131.20, 129.74, 129.53, 128.77, 128.74, 128.03, 126.18, 125.84, 125.05, 124.10, 121.05, 116.18. HRMS (ESI) m/z calcd for C₂₆H₁₆N₅O₆S ([M + H]⁺): 526.0816, found 526.0810.

2.3. Spectral measurements

2.5 mM of probe-dbp stock solution and 5 mM *p*-thiocresol (a very common thiophenol) stock solution were prepared in analytical grade dimethyl sulfoxide (DMSO). The stock solutions were diluted with the mixed solvent (pH = 7.4, 10 mM PBS buffer/DMSO = 1 : 1, v/v) to desired concentration. The absorption and fluorescence spectra were carried out at room temperature (λ_{ex} = 450 nm and λ_{em} = 570 nm). The stock solutions (5 mM) of *p*-F-C₆H₄SH, *p*-CH₃O-C₆H₄SH, C₆H₅NH₂ and C₆H₅OH were also prepared in DMSO. Other analytes such as amino acids, common metal ions and aliphatic thiols were dissolved in ultrapure water. All stock solutions were used immediately after preparation and were diluted with the mixed solvent (pH = 7.4, 10 mM PBS buffer/DMSO = 1 : 1, v/v) to

desired concentration. The mixture of probe-dbp and analytes were incubated for 15 min at 37 °C before recording the fluorescence and absorption spectra. The results were reported as the mean \pm standard deviation of triplicate experiments.

2.4. Preparation of real water samples

The tap water of Qingdao, spring water of Laoshan Mountain (Qingdao, Shandong province, China), and seawater from Jiaozhou Bay (Qingdao, Shandong province, China) were collected as the crude water samples. The water samples were filtered and kept at 4 °C before use. These water samples were then spiked with *p*-thiocresol to desired concentration (0, 2.0, 4.0, 6.0, 8.0, 10.0, 15.0, 20.0 μ M). Then, the water samples were incubated with probe-dbp for 15 min before measurement, and the fluorescence intensity at 570 nm was recorded.

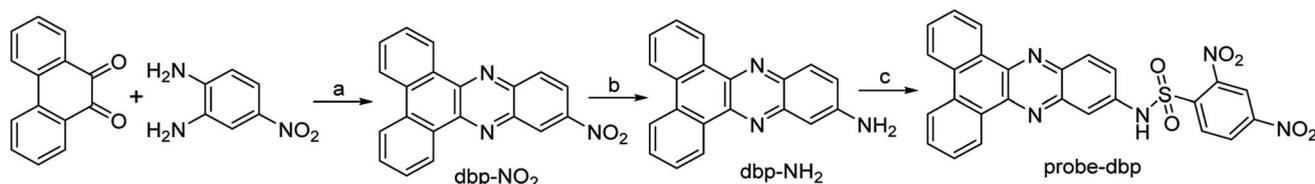
3. Results and discussion

3.1. Synthesis of probe-dbp

Dibenzo[*a,c*]phenazines is a kind of excellent organic fluorescent compounds with large Stokes shift and good photostability. On the other hand, 2,4-dinitrobenzenesulfonyl group is a well-known fluorescence quencher and an effective reaction site for thiophenols recognition. Base on this fact, we designed probe-dbp with dibenzo[*a,c*]phenazin-11-amine as fluorophore linked to 2,4-dinitrobenzenesulfonyl group as recognition site. The synthetic routes were shown in Scheme 1. 9,10-Phenanthrenequinone and 4-nitro-*o*-phenylenediamine were selected as starting material. A mixture of them were refluxed in ethanol to give compound dbp-NO₂. Dbp-NO₂ was then reduced by Pd-C and N₂H₄·H₂O to produce dbp-NH₂. Dbp-NH₂ reacted with 2,4-dinitrobenzenesulfonyl chloride to afford probe-dbp. The experimental part and spectral data can be found in the ESI.†

3.2. Photophysical properties and response towards thiophenols

The absorption and fluorescence spectroscopy of probe-dbp before and after addition of *p*-thiocresol were measured in aqueous medium (pH = 7.4, 10 mM PBS buffer/DMSO = 1 : 1, v/v) and depicted in Fig. S1.† The maximum absorption of probe-dbp was located at 450 nm. Before the addition of *p*-thiocresol, probe-dbp displayed no fluorescence. When *p*-thiocresol (50 μ M) was added into the solution containing probe-dbp (5 μ M), a strong fluorescence enhancement (~50 fold) was observed at 570 nm (λ_{ex} = 450 nm) (Fig. S2†).



Scheme 1 The synthetic routes of probe-dbp. Reagents and conditions: (a) C₂H₅OH, CH₃COOH, reflux; (b) C₂H₅OH, Pd-C, N₂H₄·H₂O, reflux; (c) 2,4-dinitrobenzenesulfonyl chloride, pyridine, CH₂Cl₂, RT.



The time-dependent response of probe-dbp to *p*-thiocresol was further evaluated. As depicted in Fig. 1A, fluorescence was observed immediately after the addition of *p*-thiocresol. The fluorescence intensity increased rapidly at 570 nm in the first 5 min upon the addition of the *p*-thiocresol, and reached equilibrium in 15 min. The results indicated that probe-dbp was a rapid and remarkable fluorescent probe for detecting thiophenols.

Fluorometric titration experiments of probe-dbp were carried out and the results were shown in Fig. 1B. As the thiophenol concentration increasing, the emission intensity of the probe-dbp solution grew and reached a saturation when 10 equivalents of *p*-thiocresol was added. In addition, a good linear relationship (Fig. 1C) between the fluorescence intensity at 570 nm and the thiophenol concentration in a range of 0–20 μM ($R^2 = 0.99545$) was observed with a detection limit of 40 nm ($3\delta/k$).

3.3. The effect of pH

Probe-dbp itself is pH stable in a broad pH range from 3 to 13 (Fig. 1D). However, with the addition of *p*-thiocresol (50 μM), the probe solution showed a large fluorescence enhancement at

570 nm between pH 4.0 and 13.0. Moreover, the fluorescence enhancement at 570 nm was stable towards pH ranging from 5 to 13. As far as we know, there is no thiophenols fluorescent probe has been reported possessing such stable fluorescence enhancement over such a wide pH range (Table S1†). The results suggested that probe-dbp was able to apply for thiophenols monitoring in a wide pH range, even in a strong alkaline condition.

3.4. Selectivity for thiophenols

To explore the selectivity of probe-dbp towards thiophenols, the fluorescence change of probe-dbp towards various relevant analytes including thiophenols (*p*-thiocresol, 4-methoxy thiophenol, 4-fluorothiophenol), aliphatic thiols such as ethanethiol, glutathione (GSH), cysteine(Cys), homocysteine (Hcy), amino acids (Lys, His, Val, Tyr, Ala, Leu, Trp, Ile, Ser, Met), nucleophilic substances such as Ph-NH₂, Ph-OH, NaSH, and other potential interfering species (Pb(AcO)₂, FeCl₃, NaHSO₃, CuSO₄, FeSO₄, KI, CuI, CaCl₂, MgSO₄, NaCl, SnCl₂, Na₂SO₃) were studied. As illustrated in Fig. 2A, addition of aliphatic thiols, amino acids, nucleophilic substances and other competitive

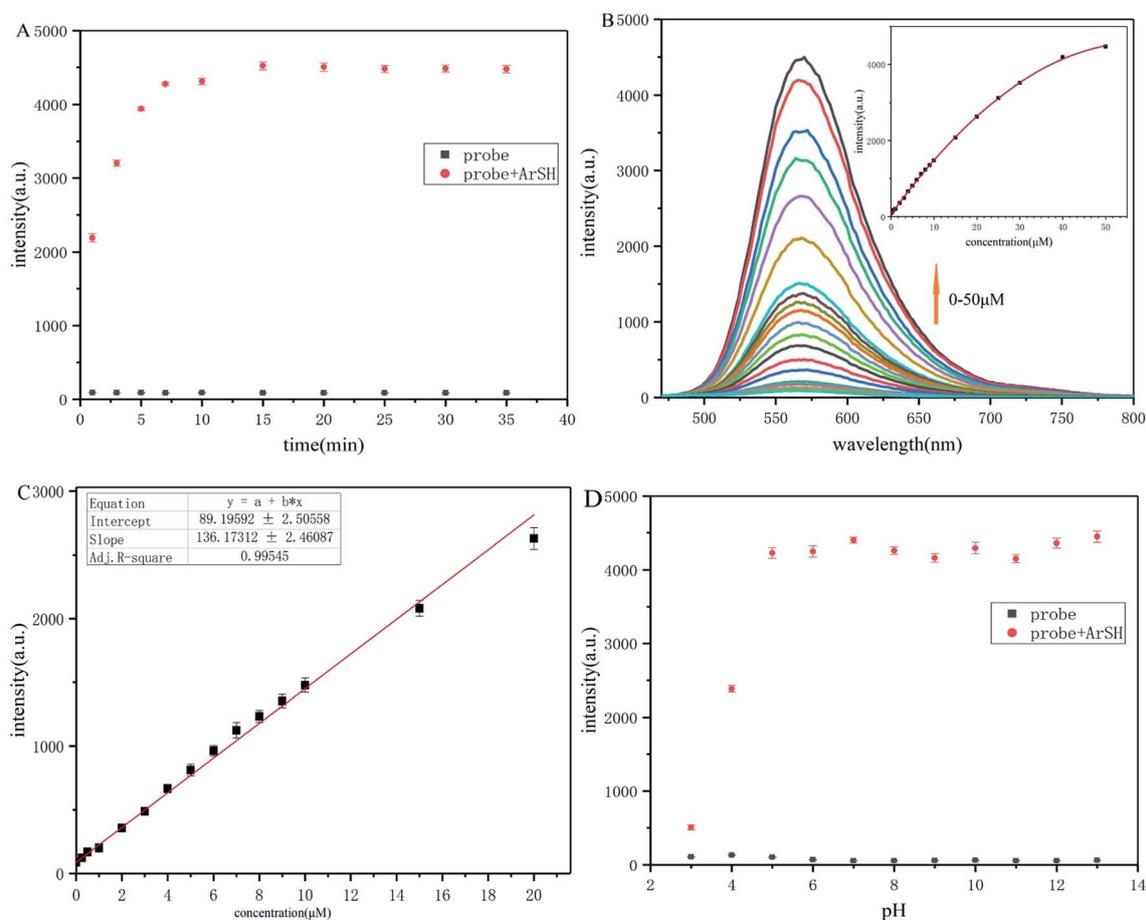


Fig. 1 (A) Fluorescence intensity at 570 nm for probe-dbp with/without *p*-thiocresol as a function of time. (B) Fluorescence emission spectra of probe-dbp upon the addition of different concentration of *p*-thiocresol (0, 0.25, 0.5, 1, 2, 3, 4, 5, 6, 7, 8, 9, 10, 15, 20, 25, 30, 40, 50 μM). (C) The linear relationship between the fluorescence intensity of probe-dbp at 570 nm and the concentration of *p*-thiocresol. (D) The effect of pH on the fluorescence intensity of probe-dbp at 570 nm. $\lambda_{\text{ex}} = 450 \text{ nm}$, $\lambda_{\text{em}} = 570 \text{ nm}$. Relative standard deviation (RSD), $n = 3$.

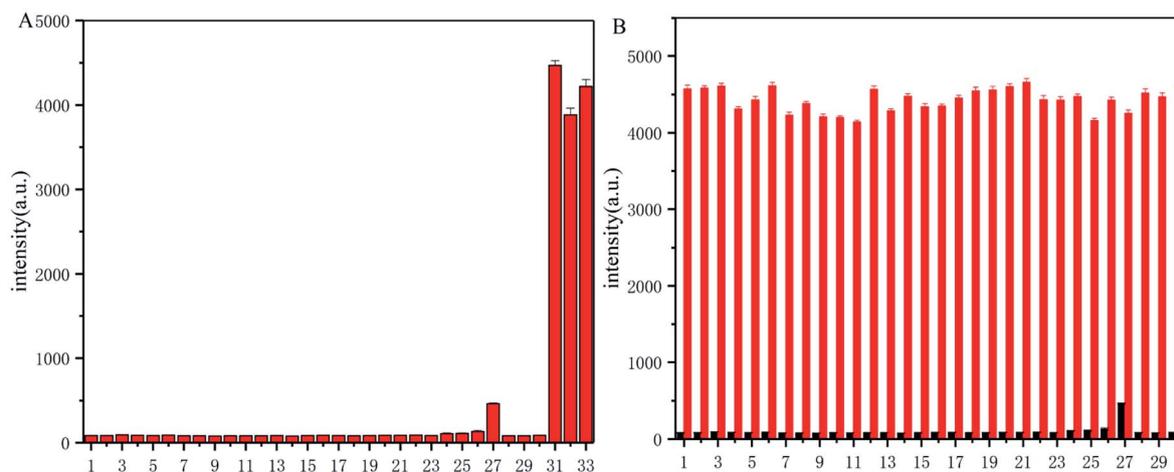


Fig. 2 (A) Fluorescence intensity responses of probe-dbp (5 μM) at 570 nm in the presence of 100 μM aliquots of various analytes. (B) Fluorescence intensity responses of probe-dbp (5 μM) at 570 nm toward *p*-thiocresol in the presence of various analytes. (1) blank, (2) Na₂SO₃, (3) Pb(AcO)₂, (4) FeCl₃, (5) NaHSO₃, (6) CuSO₄, (7) FeSO₄, (8) KI, (9) CuI, (10) CaCl₂, (11) MgSO₄, (12) NaCl, (13) SnCl₂, (14) Lys, (15) His, (16) Val, (17) Tyr, (18) Ala, (19) Leu, (20) Try, (21) *iso*-Leu, (22) Ser, (23) Met, (24) GSH, (25) homo-Cys, (26) Cys, (27) NaHS, (28) C₂H₅SH, (29) aniline, (30) phenol, (31) *p*-thiocresol, (32) 4-methoxythiophenol, (33) 4-fluorothiophenol. Black bar represent the addition of a single analyte (100 μM). Red bar represent the subsequent addition of *p*-thiocresol (100 μM) to the mixture. λ_{ex} = 450 nm, λ_{em} = 570 nm. Relative standard deviation (RSD), *n* = 3.

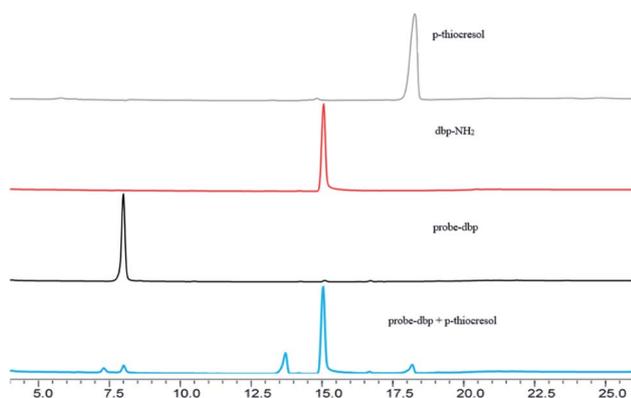


Fig. 3 HPLC chromatograms of *p*-thiocresol, dbp-NH₂, probe-dbp, probe-dbp + *p*-thiocresol.



Scheme 2 The proposed detection mechanism towards thiophenol.

fluorescence enhancement, the enhancement is negligible compared to thiophenols. Moreover, when thiophenol was added into the solution of probe-dbp and the above-mentioned analytes (Fig. 2B), remarkable fluorescence changes could also be observed, indicating that probe-dbp possesses good anti-interference ability. Thus, it can be concluded that probe-dbp is highly selective for thiophenols over other analytes.

species to the solution of probe-dbp did not induce remarkable fluorescence enhancements. In contrast, obvious fluorescence enhancements were observed when thiophenols were added into the solution of probe-dbp. Although NaHS caused a weak

3.5. Response mechanism

To understand the response mechanism of probe-dbp towards thiophenols, HPLC and mass spectrometry were employed to

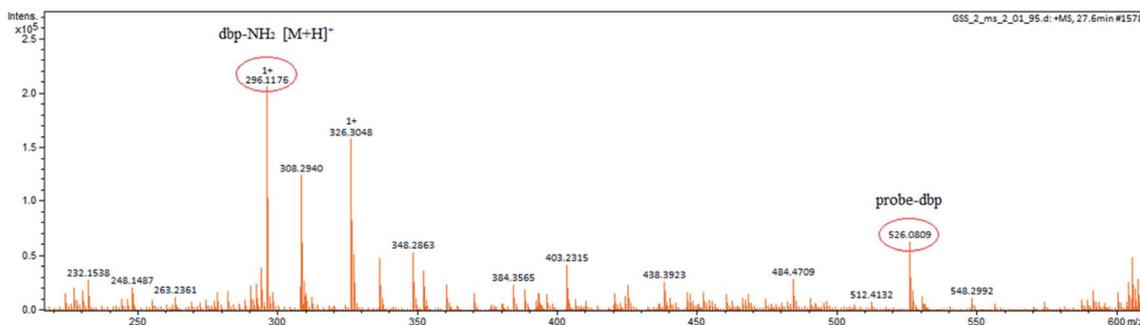


Fig. 4 Mass spectrum of the reaction solution of probe-dbp and *p*-thiocresol.



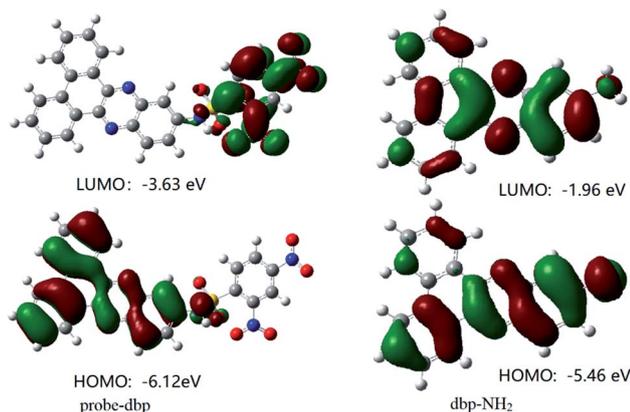


Fig. 5 Frontier orbital diagrams of probe-dbp and dbp-NH₂.

analyze probe-dbp, dbp-NH₂, and the products of probe-dbp reacted with *p*-thiocresol. As shown in Fig. 3, probe-dbp, dbp-NH₂ and *p*-thiocresol possess a retention time at 8.0 min, 15.1 min and 18.3 min, respectively. With the addition of *p*-thiocresol into probe-dbp, the peak of probe-dbp at 8.0 min

decreased. Meanwhile, the peak of dbp-NH₂ at 15.1 min appeared. Furthermore, we could see the $m/z = 296.1176$ in the mass spectrum (Fig. 4), which belong to the cleavage product dbp-NH₂. From the HPLC and HRMS analysis, we can conclude that the 2,4-dinitrobenzenesulfonyl group was removed by *p*-thiocresol to give the fluorophore molecular dbp-NH₂ (Scheme 2).

To further verify the response mechanism of probe-dbp for thiophenols, the density functional theory (DFT) was applied using the Gaussian 09 program. As shown in Fig. 5, the highest occupied molecular orbital (HOMO) of probe-dbp is primarily distributed in dibenzo[*a,c*]phenazin-11-amine group, while the lowest unoccupied molecular orbital (LUMO) is mainly distributed in 2,4-dinitrobenzenesulfonyl group. The calculated results demonstrate the possible photoinduced electron transfer (PET) processes in probe-dbp. Electron transfer from dibenzo[*a,c*]phenazin-11-amine group (PET donor) to 2,4-dinitrobenzenesulfonyl group (PET acceptor) attenuates the original fluorescence of dibenzo[*a,c*]phenazin-11-amine. When 2,4-dinitrobenzenesulfonyl group was removed, the fluorescence of dbp-NH₂ can be restored because of the removal of 2,4-

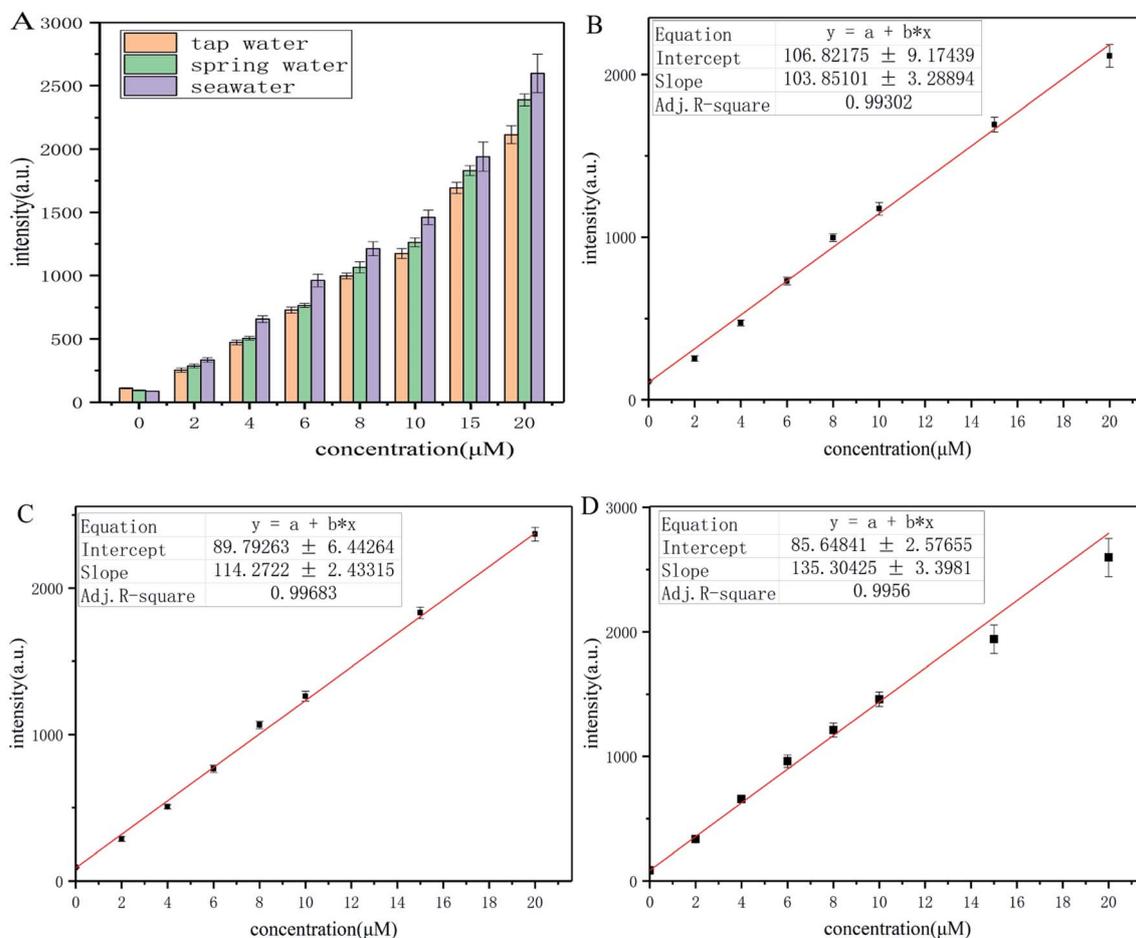


Fig. 6 (A) Fluorescence response of probe-dbp (5 μ M) towards various concentration of *p*-thiocresol in different water samples, including tap water, spring water and seawater. The linear relationship between the fluorescence intensity of probe-dbp at 570 nm and the concentration of *p*-thiocresol in tap water (B), spring water (C) and seawater (D).



dinitrobenzenesulfonyl group. The calculated results exhibit good consistency with the experimental results.

3.6. Application in real water samples

Considering the potential damage of thiophenols to the water ecosystem, probe-dbp was employed to evaluate thiophenols concentration in real water samples. Tap water of Qingdao city, spring water from Laoshan Mountain (Qingdao, Shandong province, China) and seawater from Jiaozhou Bay (Qingdao, Shandong province, China) were collected and filtered for test. A variety of concentrations of *p*-thiocresol (0–20 μ M) were spiked into these water samples. Then, these water samples were incubated with probe-dbp. The fluorescence intensity at 570 nm were recorded. It shows good relationship between the fluorescence intensity and thiophenol concentrations in a range of 0–20 μ M (Fig. 6) in all water samples. As shown in Table S2,[†] the calculated thiophenol recoveries for each water samples ranged from 93% to 108%, suggesting that probe-dbp is specific for thiophenols detection in real water samples. The method described above might provide a new way for quantitative detection of thiophenols in both freshwater and seawater samples.

4. Conclusion

In this work, we have designed and synthesized probe-dbp with dibenzo[*a,c*]phenazin-11-amine as fluorophore linked to 2,4-dinitrobenzenesulfonyl group as recognition site. Probe-dbp exhibited high selectivity and sensitivity towards thiophenols over a wide pH range from 5 to 13. Moreover, the probe can be used for the quantitative detection of thiophenols in both freshwater and seawater samples. It can be concluded that probe-dbp possesses a great potential for detection and quantification of thiophenols in environmental systems, even in a strong alkaline condition.

Conflicts of interest

There are no conflicts of interest to declare.

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References

- 1 A. Eychmuller and A. L. Rogach, *Pure Appl. Chem.*, 2000, **72**, 179–188.
- 2 K. Shimada and K. Mitamura, *J. Chromatogr. B: Biomed. Sci. Appl.*, 1994, **659**, 227–241.
- 3 J. C. Love, L. A. Estroff, J. K. Kriebel, R. G. Nuzzo and G. M. Whitesides, *Chem. Rev.*, 2005, **105**, 1103–1169.
- 4 I. C. Popoff, C. B. Thanawalla, R. A. Creager and J. R. Frank, *J. Agric. Food Chem.*, 1972, **20**, 80.
- 5 T. P. Heil and R. C. Lindsay, *J. Environ. Sci. Health - Part B Pesticides, Food Contam. Agric. Wastes*, 1989, **24**, 349–360.
- 6 P. Amrolia, S. G. Sullivan, A. Stern and R. Munday, *J. Appl. Toxicol.*, 1989, **9**, 113–118.
- 7 T. R. Juneja, R. L. Gupta and S. Samanta, *Toxicol. Lett.*, 1984, **21**, 185–189.
- 8 P. V. R. Snelgrove, *Bioscience*, 1999, **49**, 129–138.
- 9 Y. Sun, Z. Lv, Z. Sun, C. Wu, Z. Ji and J. You, *Anal. Bioanal. Chem.*, 2016, **408**, 3527–3536.
- 10 K. Beiner, P. Popp and R. Wennrich, *J. Chromatogr. A*, 2002, **968**, 171–176.
- 11 D. R. Dreyer, H.-P. Jia, A. D. Todd, J. Geng and C. W. Bielawski, *Org. Biomol. Chem.*, 2011, **9**, 7292–7295.
- 12 H. Li, J. Fan, J. Wang, M. Tian, J. Du, S. Sun, P. Sun and X. Peng, *Chem. Commun.*, 2009, 5904–5906, DOI: 10.1039/b907511a.
- 13 O. Pluchery, C. Humbert, M. Valamanesh, E. Lacaze and B. Busson, *Phys. Chem. Chem. Phys.*, 2009, **11**, 7729–7737.
- 14 H. Liu, C. Guo, S. Guo, J. Fan, L. Wang and D. Shi, *Talanta*, 2019, **201**, 301–308.
- 15 F. Wu, H. Wang, J. Xu, H.-Q. Yuan, L. Zeng and G.-M. Bao, *Sens. Actuators, B*, 2018, **254**, 21–29.
- 16 Y. Zhang, Y. Hao, X. Ma, S. Chen and M. Xu, *Environ. Pollut.*, 2020, **265**, 114958.
- 17 S. Chen, H. Li and P. Hou, *Anal. Chim. Acta*, 2017, **993**, 63–70.
- 18 X. Xie, M. Li, F. Tang, Y. Li, L. Zhang, X. Jiao, X. Wang and B. Tang, *Anal. Chem.*, 2017, **89**, 3015–3020.
- 19 L. Xiong, J. Ma, Y. Huang, Z. Wang and Z. Lu, *ACS Sens.*, 2017, **2**, 599–605.
- 20 L. Dong, G.-R. Chen, X.-P. He and S. Vidal, *Org. Biomol. Chem.*, 2019, **17**, 9251–9256.
- 21 J. Gai, C. Chen, J. Huang, J. Sheng, W. Chen and X. Song, *Tetrahedron Lett.*, 2020, **61**, 152038.
- 22 Y. Geng, H. Tian, L. Yang, X. Liu and X. Song, *Sens. Actuators, B*, 2018, **273**, 1670–1675.
- 23 Y. Guo, D. Zhang, J. Wang, H. Lu and S. Pu, *Dyes Pigm.*, 2020, **175**, 108191.
- 24 K.-R. Kim, H. J. Kim and J.-I. Hong, *Anal. Chem.*, 2019, **91**, 1353–1359.
- 25 J. Niu, Y. Li, R. Zhang, J. Zhang, X. Wang, X. Xie, X. Jiao and B. Tang, *Chem. Commun.*, 2021, **57**, 2800–2803.
- 26 T. Yamamoto, K. Sugiyama, T. Kushida, T. Inoue and T. Kanbara, *J. Am. Chem. Soc.*, 1996, **118**, 3930–3937.
- 27 A. Samadi-Maybodi, R. Akhoondi and M. J. Chaichi, *J. Fluoresc.*, 2010, **20**, 671–679.
- 28 T. Yamamoto, *Macromol. Rapid Commun.*, 2002, **23**, 583–606.
- 29 W. Jiang, Q. Q. Fu, H. Y. Fan, J. Ho and W. Wang, *Angew. Chem. Int. Ed.*, 2007, **46**, 8445–8448.
- 30 W. Lv, Y. Chen, L. Bian and X. Chen, *Talanta*, 2019, **197**, 204–210.
- 31 J. J. Wu, D. D. Su, C. Q. Qin, W. Li, J. Rodrigues, R. L. Sheng and L. T. Zeng, *Talanta*, 2019, **201**, 111–118.
- 32 W. Y. Zhang, X. J. Liu, H. Y. Zhang, C. C. Feng, C. X. Liu, M. M. Yu, L. H. Wei and Z. X. Li, *J. Mater. Chem. C*, 2015, **3**, 8248–8254.
- 33 S. Pagidi, N. K. Kalluvettukuzhy and P. Thilagar, *Langmuir*, 2018, **34**, 8170–8177.

