


 Cite this: *RSC Adv.*, 2022, 12, 4101

# Optimization of the adsorption and removal of Sb(III) by MIL-53(Fe)/GO using response surface methodology†

Xiuzhen Yang, \* Haolin Zhang, Shuangchan Cheng and Bin Zhou

In this study, a graphene oxide metal–organic framework (MIL-53(Fe)/GO) composite adsorbent was successfully synthesized using a simple method at room temperature. The specific surface area of the synthesized MIL-53(Fe)/GO nanoparticles was  $268.43 \text{ m}^2 \text{ g}^{-1}$ , with an average pore size of 2.52 nm. The Box–Behnken response surface method was applied to optimize the adsorption time, dosage, pH, temperature, and initial concentration of Sb(III) in the MIL-53(Fe)/GO adsorption treatment employed for synthetic wastewater containing Sb(III). We determined the optimal adsorption conditions and explored the isotherm model, adsorption kinetic model, and adsorption mechanism during the adsorption process. For an optimal adsorption of Sb(III) by MIL-53(Fe)/GO, the adsorption time, dosage, pH, temperature, and initial Sb(III) concentration should be set to 4.86 h, 85.79  $\text{mg L}^{-1}$ , 10.00, 39.29 °C, and 10.09  $\text{mg L}^{-1}$ , respectively. Under these optimal conditions, the removal rate of Sb(III) will be as high as 97.97%. The adsorption of Sb(III) by MIL-53(Fe)/GO conformed to the Freundlich isotherm adsorption model, and its maximum adsorption capacity was  $69.014 \text{ mg g}^{-1}$ . The adsorption kinetics process, which is a nonhomogeneous reaction, could be fitted using a quasi-first-order kinetic model. A Fourier transform infrared spectroscopy analysis showed that MIL-53(Fe)/GO hydroxyl and amine groups play a vital role in the adsorption process. MIL-53(Fe)/GO did not exhibit any changes in its adsorption efficiency in the presence of its anion and showed high specificity to Sb(III). XPS characterization showed that Sb successfully adsorbed onto the adsorbent and that no oxidation–reduction reaction occurred during the adsorption process. The adsorption efficiency remained high even after four cycles of use. MIL-53(Fe)/GO is highly recyclable with significant application potential for treating wastewater containing Sb(III).

 Received 8th November 2021  
 Accepted 2nd January 2022

DOI: 10.1039/d1ra08169a

[rsc.li/rsc-advances](http://rsc.li/rsc-advances)

## 1. Introduction

Antimony and its compounds are widely used as flame retardants, polymerization catalysts, and pigments; however, it is highly toxic. The toxicity of Sb(III) is approximately 10 times that of Sb(V),<sup>1,2</sup> and long-term exposure to antimony can cause pneumoconiosis, emphysema, and myocardial degeneration.<sup>3</sup> Severe damage to human DNA<sup>4</sup> can cause damage to the lungs, heart, and liver, increasing the risk of cancer.<sup>5,6</sup> Therefore, the European Union, the U.S. Environmental Protection Agency, and other organizations have listed antimony as a pollutant.<sup>7</sup> China has one of the largest antimony mines in the world. Antimony is being mined in large quantities, and its usage is increasing year by year. From the initial 20 000 tons, China surpassed the United States, Japan, and other countries in 2004, becoming the world's largest consumer of antimony.

Unfortunately, antimony is being discharged into the natural environment, causing pollution of the atmosphere, water, plants, and soil.<sup>8</sup> Therefore, many countries and regions have adopted certain measures to control and limit its content in the natural environment.<sup>7</sup> For example, the maximum allowable concentration of antimony in drinking water is  $2 \text{ } \mu\text{g L}^{-1}$  as per Japanese standards, whereas it is  $5 \text{ } \mu\text{g L}^{-1}$  in Europe.<sup>9</sup> As per China's surface water quality standards, the antimony concentration in a water source should not exceed  $5 \text{ } \mu\text{g L}^{-1}$ .<sup>10</sup> Thus, removing Sb(III) from water bodies is an essential and urgent task.

The adsorption method is a cost-effective water treatment technique, widely used owing to its high adsorption capacity, high efficiency, and adsorbent recyclability. Metal–organic frameworks (MOFs) are a new class of highly porous materials. Owing to their large specific surface area, adjustable pore structure, and many other advantages, they have applications in various fields, including gas storage separation,<sup>11,12</sup> selective catalysis,<sup>13</sup> magnetic separation,<sup>14</sup> chemical sensing,<sup>15</sup> and drug delivery,<sup>16</sup> and thus have attracted wide attention.<sup>17</sup> However, MOFs have certain drawbacks when it comes to their

College of Civil Engineering, Human University of Science and Technology, Xiangtan, 411201, China. E-mail: yxzhy1102@126.com; Tel: +86 18073165540

† Electronic supplementary information (ESI) available. See DOI: 10.1039/d1ra08169a



application prospects. To improve their stability and dispersion power, different types of groups have been introduced into porous MOFs. For example, a composite containing MOFs and graphene has been widely researched. MOFs can grow on graphene sheets. Owing to the coordination of the GO oxygen group and the central metal in an MOF, a strong chemical bond and new micropores are formed. Thus, this material exhibits good thermal stability and flexibility<sup>18</sup> and is often used for photocatalysis<sup>19</sup> and electrochemistry.<sup>20</sup> However, there have been no reports on its use for removing heavy metals from water.

## 2. Materials and methods

### 2.1 Preparation of MIL-53(Fe)/GO

First, by applying an ultrasonic treatment for 150 min, a certain quality of GO was uniformly dispersed in 10 mL of DMF at 100 power. We mixed  $\text{FeCl}_3 \cdot 6\text{H}_2\text{O}$ ,  $\text{H}_2\text{BDC}$ , and DMF at a molar ratio of 1 : 1 : 280, magnetically stirred for 1 h to make it a clear solution, then added the ultrasonically treated GO/DMF mixture to the above solution, and continued the stirring for 1 h. Subsequently, we transferred it to a 100 mL Teflon liner. Thereafter, the Teflon liner was sealed in a stainless steel autoclave and heated to 150 °C for 20 h. The suspension was obtained by centrifugation and then washed with DMF and  $\text{C}_2\text{H}_5\text{OH}$  repeatedly by centrifugation. Finally, the solid was dried in vacuum at 80 °C for 10 h to obtain a pale-yellow powder.

### 2.2 Reagent detection and determination

Potassium antimony tartrate was used to prepare a standard stock solution with a  $\text{Sb(III)}$  concentration of 1000  $\text{mg L}^{-1}$  for subsequent tests. The reagents used in the test were all analytically pure, and the test water was deionized water. A flame atomic absorption spectrophotometer (AA7002A, Beijing Sanxiong Technology Company) was used to determine the concentration of  $\text{Sb(III)}$  in the solution. The formula  $q = (C_0 - C_e)/C_0 \times 100\%$  was used to calculate the removal rate of Sb, where  $q$  is the removal rate, %, and  $C_0$  and  $C_e$  are the concentrations of  $\text{Sb(III)}$  before and after the adsorption,  $\text{mg L}^{-1}$ , respectively.

### 2.3 Response surface method for optimizing adsorption reaction

The response surface method is an effective method to optimize the process parameters, reduce the number of experiments, and evaluate the level and interaction between the various influencing factors.<sup>21</sup> Because of the interaction between various factors, conventional single-factor tests do not produce stable and efficient test results.

Box–Behnken design (BBD) is an incomplete three-level factorial design. It is considered an effective technique owing to the number of runs that can be reduced compared with the full-factorial three-level design (FFD). The BBD design does not contain any experimental points at the corner of the cubic surface. This is because when the factors are the same, there is no axial point. On this basis, the response surface method was

used to optimize the process of MIL-53(Fe)/GO adsorption of  $\text{Sb(III)}$ , taking the adsorption time, dosage, pH, temperature, and  $\text{Sb(III)}$  initial concentration as the influencing factors. Taking the removal rate of  $\text{Sb(III)}$  as the response value, a quadratic polynomial model between the removal rate and various factors was established to obtain the best experimental conditions for the removal of  $\text{Sb(III)}$  by adsorption.

The Box–Behnken combination design method in Design-Expert software was used to optimize the experiment. We take the optimal value point (0) of each test single factor as the center and the high (+1) and low (−1) levels in the upper and lower regions for the response surface test design: time, dosage, pH, and temperature. The initial concentration of  $\text{Sb(III)}$  is the influencing factor, and the adsorption removal rate is the response value. The effects of the first, quadratic (interaction terms), and square terms (surface action) on the adsorption effect were investigated. Forty-six groups of optimization experiments were conducted, six groups of central point experiments were repeated, each group of experiments was repeated thrice, and the average value was taken as the corresponding response value. Table 1 presents the relationship between the three level codes of the five factors and the experimental values. The volume of the solution in the experiment is 50 mL.

### 2.4 Adsorption isotherm fitting

The formula  $Q_e = (C_0 - C_e) \times V/m$  was used during the test to calculate the adsorption capacity. Here,  $C_0$  and  $C_e$  are the concentrations of  $\text{Sb(III)}$  before and after the adsorption,  $\text{mg L}^{-1}$ , respectively;  $V$  is the volume of the  $\text{Sb(III)}$  solution, mL;  $m$  is the adsorbent dosage, mg.

**Isothermal adsorption test.** We prepared 50 mL of  $\text{Sb(III)}$  solution with different concentration gradients (10–180  $\text{mg L}^{-1}$ ), carried out an adsorption test under the best conditions obtained in the response surface optimization experiment, measured the  $\text{Sb(III)}$  concentration, and calculated the adsorption capacity  $q_e$ . The Langmuir, Freundlich, and Temkin (T–M) isotherm adsorption models were used, as expressed in eqn (1)–(3), respectively, to fit the test results.

$$\frac{C_e}{Q_e} = \frac{1}{Q_{\max} K_L} + \frac{C_e}{Q_{\max}} \quad (1)$$

$$\ln Q_e = \ln K_F + \frac{\ln C_e}{n} \quad (2)$$

Table 1 Influencing factors and levels of Box–Behnken design

Factor	Coding	Unit	Level coding value of each factor		
			−1	0	1
Adsorption time	A	h	4	5	6
Adsorbent dosage	B	$\text{mg L}^{-1}$	70	80	90
pH	C		8	9	10
Temperature	D	°C	25	35	45
$\text{Sb(III)}$ initial concentration	E	$\text{mg L}^{-1}$	10	20	30



$$Q_e = B \ln A + B \ln C_e \quad (3)$$

where  $C_e$  is the  $Sb(III)$  concentration after adsorption equilibrium,  $mg L^{-1}$ ;  $Q_e$  and  $Q_{max}$  are the equilibrium adsorption capacity and maximum adsorption capacity, respectively,  $mg g^{-1}$ ;  $K_L$  is the adsorption constant of the Langmuir isotherm adsorption model,  $L mg^{-1}$ ;  $K_F$  is the Freundlich isotherm adsorption model constant,  $L g^{-1}$ ;  $n$  is the Freundlich isotherm adsorption model constant, dimensionless;  $A$  and  $B$  are the Temkin isotherm constants, the values of which can be obtained from the linear relationship in the  $Q_e - \ln C_e$  plot.

### 2.5 Adsorption kinetic equation fitting

We prepared 50 mL of  $Sb(III)$  (=10 mg per L of  $Sb$ ) solution and carried out a kinetic adsorption test under optimal conditions, measured the  $Sb(III)$  concentration in the solution at different times after the reaction starts, and calculated the  $Sb(III)$  adsorption capacity. Three adsorption kinetic models (see eqn (4)–(6)) were used to fit the test results, and the rate control and adsorption mechanism of the adsorption reaction were analyzed.

$$q_t = q_e(1 - \exp(-k_1 t)) \quad (4)$$

$$q_t = q_e - q_e/(k_2 q_e t + 1) \quad (5)$$

$$q_t = k_i t^{1/2} + a \quad (6)$$

where  $q_t$  and  $q_e$  are the adsorption capacity at  $t$  (min) and adsorption equilibrium,  $mg g^{-1}$ , respectively;  $k_1$ ,  $k_2$ , and  $a$  are the adsorption rate constants of the corresponding model, and the units are  $min^{-1}$ ;  $g (mg min)^{-1}$  and  $k_i (g min^{0.5})$  are the intraparticle diffusion model constant and the adsorption rate constant at a certain stage in the adsorption process.

### 2.6 Characterization method before and after adsorption

An SEM (JSM-6610LV, JEOL, Japan) was used to characterize the morphology of MIL-53(Fe)/GO before and after  $Sb(III)$  adsorption. FT-IR (Nicolet 6700, Thermo Fisher, America) was used to characterize GO/MIL-53 (the functional groups and chemical bond composition before and after Fe adsorption of  $Sb(III)$ ) were

analyzed). XRD (D8-Advance, Bruker, Germany) was used to characterize the MIL-53(Fe)/GO structure. BET (TriStarII 3020, McMurtik, America) was used to measure the specific surface area and pore size of MIL-53(Fe)/GO. XPS (K-Alpha 1063 Thermo Fisher Scientific) was used to analyze the element species in the tested sample, along with the type, content, chemical group, and other information. Finally, the adsorption of  $Sb(III)$  by MIL-53(Fe)/GO was analyzed on the basis of the characterization results.

## 3. Results and discussion

### 3.1 Characterization and adsorption mechanism of MIL-53(Fe)/GO

**SEM analysis.** Fig. 1 shows the scanning electron micrographs (SEM) of MIL-53 (Fe), MIL-53 (Fe)/GO, and  $Sb(III)$ . Fig. 1(a) shows that the size of the MIL-53(Fe) particles is relatively uniform, the surface is smooth and dispersed, and the shape is an octahedral prism structure with sharp ends at the middle.<sup>22</sup> As shown in Fig. 1(b), MIL-53(Fe)/GO is no longer an angular octahedron, but still maintains the polyhedral structure of MIL-53(Fe), along with crystal size reduction.<sup>23,24</sup> Although multi-piece bonded, its shape is more disorderly and irregular. GO diffuses densely and tightly on the MIL-53(Fe) surface, which increases the dispersibility of the crystals and increases the specific surface area of the adsorbent. Fig. 1(c) shows that after the adsorption reaction, the material surface is no longer tightly bonded, the skeleton has collapsed, and a large number of amorphous particles are attached onto the material surface, most of which may be adsorbed  $Sb(III)$ .

**XRD characterization.** The crystal structures of the prepared MIL-53(Fe) and MIL-53(Fe)/GO nanocomposites were characterized by powder X-ray diffraction (Fig. 2(a)). The main diffraction peaks indicate that the high crystallinity of MIL-53(Fe) is consistent with the simulation results.<sup>25,26</sup> In addition, MIL-53(Fe)/GO and MIL-53(Fe) have similar XRD patterns, indicating that the introduction of GO has no effect on the crystal structure of MIL-53(Fe). However, the intensity of the diffraction peak increases, and a characteristic diffraction pattern of GO can be observed.

**FT-IR characterization.** Fig. 2(b) shows the changes in the Fourier infrared spectrum of MIL-53(Fe)/GO after the

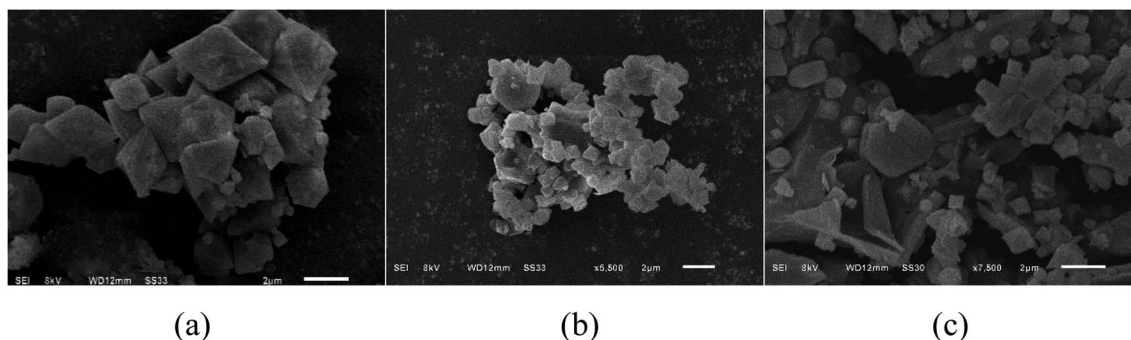


Fig. 1 Scanning microscopy images of topography.



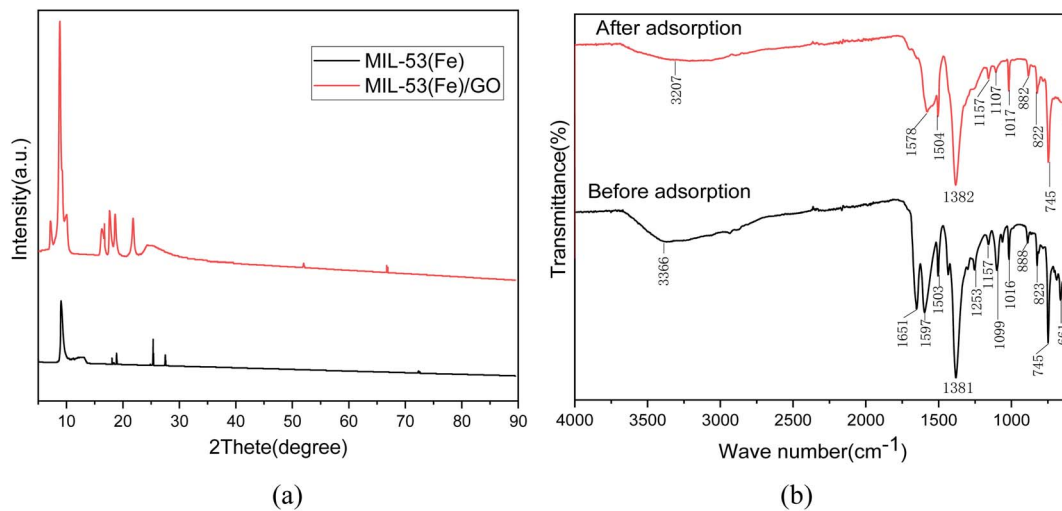


Fig. 2 Analysis of XRD and FT-IR spectra before and after Sb(III) adsorption on MIL-53(Fe)/GO.

Table 2 Physical properties of adsorbent

Physical properties	MIL-53(Fe)/GO	MIL-53(Fe)
Specific surface area ( $\text{m}^2 \text{g}^{-1}$ )	268.43	194.41
Average pore size (nm)	2.52	3.74
Micropore volume ( $\text{cm}^3 \text{g}^{-1}$ )	0.12	0.05

adsorption reaction with Sb(III) in water. In the curve after adsorption, we find that because of the influence of the adsorption reaction, some of the characteristic peaks shift to varying degrees. The characteristic peaks generated by the stretching vibration of hydroxyl (O-H) and (N-H) at  $3366 \text{ cm}^{-1}$  before adsorption shift to  $3207 \text{ cm}^{-1}$  after adsorption, *i.e.*, they move toward lower frequencies. This may be due to the replacement of H on the hydroxyl functional group by Sb(III),<sup>27</sup> the  $\text{-C=O}$  group vibration stretching absorption peak may be  $1651 \text{ cm}^{-1}$ , and the surface of the material contains  $\text{-COOH}$  group<sup>28</sup> After adsorption, it shifted to  $1578 \text{ cm}^{-1}$ . The

characteristic peak produced by the  $\text{-OH}$  bond at  $1099 \text{ cm}^{-1}$  shifted to  $1107 \text{ cm}^{-1}$  after adsorption. This means that the O-H, N-H bond, and hydroxyl ( $\text{-OH}$ ) in the polysaccharide play a major role in the process of Sb(III) adsorption, and it may be mainly through a complex reaction.<sup>29</sup> The carboxyl group reacts with Sb(III).<sup>30</sup> Therefore, Sb(III) was successfully adsorbed.

**Surface area characterization.** Typically, in the adsorption reaction process, the larger the specific surface area of a material, the higher its adsorption performance.<sup>31</sup> Based on the test results of the specific surface area and pore size analysis, the specific surface areas of MIL-53(Fe)/GO and MIL-53(Fe) are  $268.43 \text{ m}^2 \text{g}^{-1}$  and  $194.41 \text{ m}^2 \text{g}^{-1}$ , respectively. Because graphene oxide is a type of high specific surface area material, after adding graphene oxide, the specific surface area of MIL-53(Fe)/GO is increased. The micropore volumes are  $0.12 \text{ cm}^3 \text{g}^{-1}$  and  $0.05 \text{ cm}^3 \text{g}^{-1}$ , respectively. Only the micropores in the adsorbent have an adsorption effect. The larger the micropore volume, the better the adsorption effect (Table 2).<sup>32</sup>

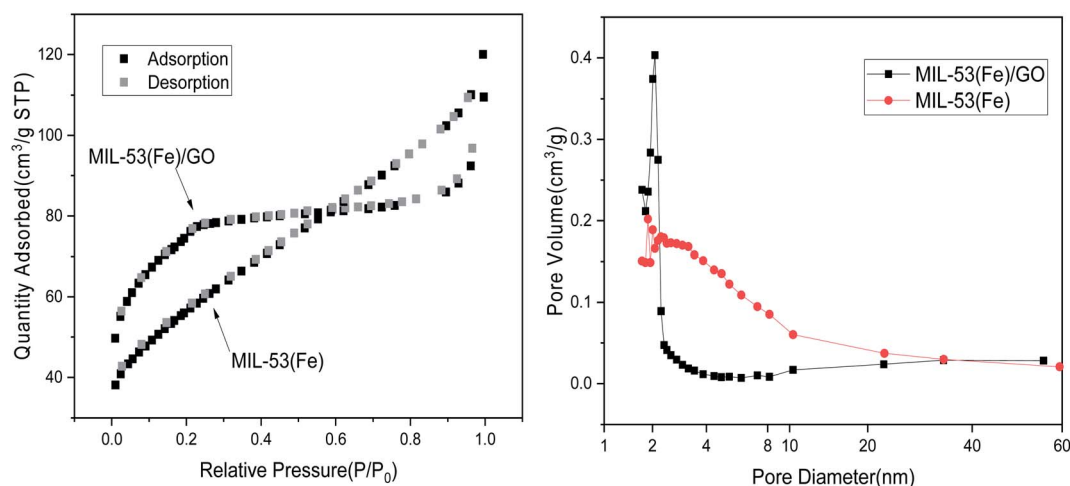


Fig. 3  $\text{N}_2$  adsorption-desorption isotherm.



Fig. 3 shows the  $N_2$  adsorption/desorption isotherm of MIL-53(Fe)/GO. The sample has H4 curve type IV, which is the main feature of mesoporous materials.<sup>33</sup> As shown, the pore size is narrowly distributed between 1.72 nm and 2.34 nm, indicating that MIL-53(Fe)/GO has a highly uniform pore structure.<sup>34</sup>

**XPS characterization.** X-ray photoelectron spectroscopy (XPS) was used to analyze the elemental composition and electronic structure of the MIL-53(Fe)/GO composite. As shown

in Fig. 4(a), there are two peaks in the Fe 2p spectrum. The binding energy peak at 711.57 eV belongs to Fe  $2p_{3/2}$ , and the peak at 725.24 eV belongs to Fe  $2p_{1/2}$ . The peak spacing, namely  $\Delta = 2p_{1/2} - 2p_{3/2} = 13.67$  eV, is consistent with the reported  $\alpha$ -Fe<sub>2</sub>O<sub>3</sub> peak and FeOOH peak.<sup>35,36</sup> Fe is mainly composed of FeOOH and Fe<sub>2</sub>O<sub>3</sub> form exists in MIL-53(Fe)/GO. This can be used as a feature of Fe<sup>3+</sup> in the MIL-53(Fe) structure.<sup>37</sup> After MIL-53(Fe)/GO adsorbed Sb(III), the binding energies corresponding to the

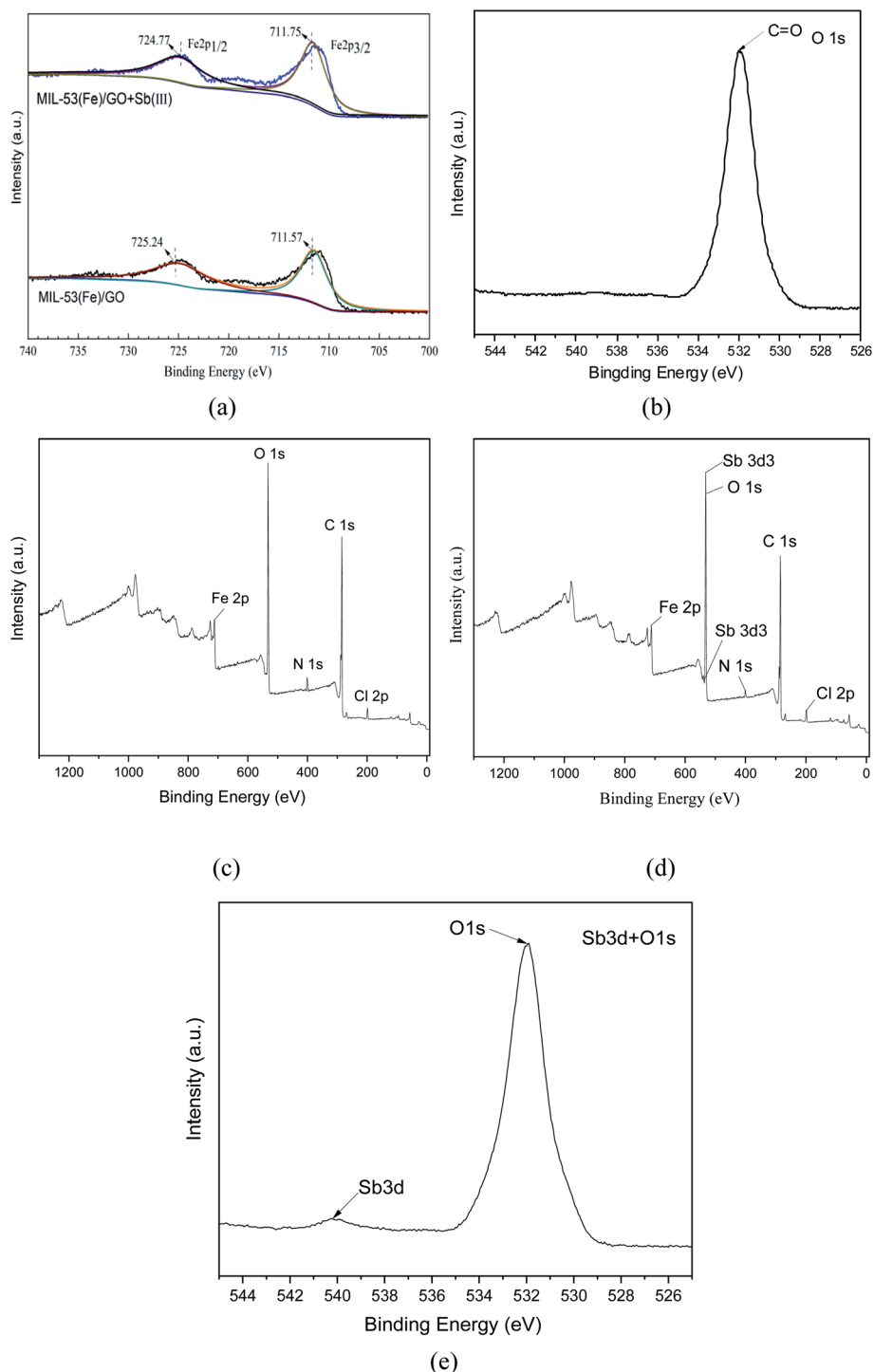


Fig. 4 XPS element distributions before and after MIL-53(Fe)/GO adsorption.



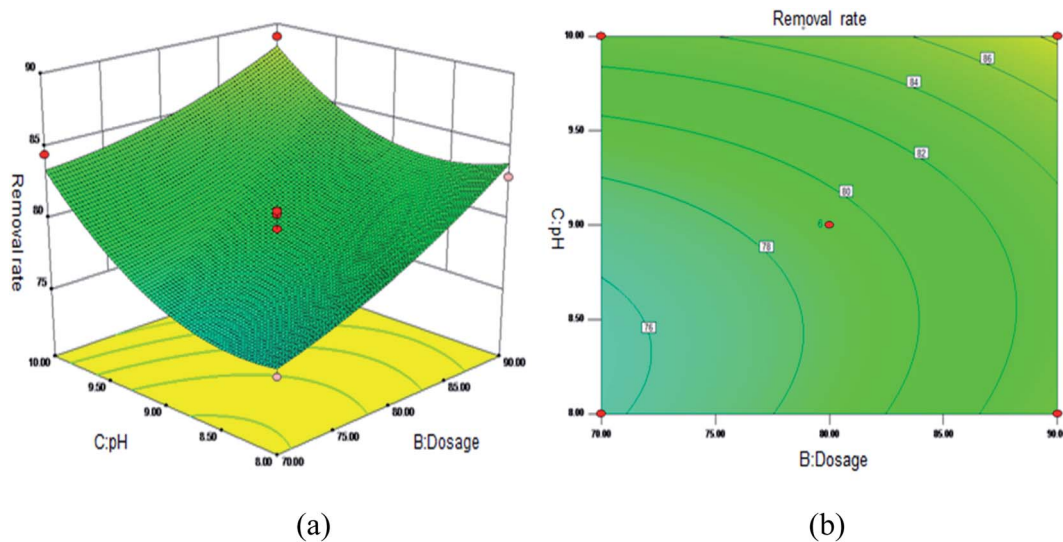


Fig. 5 Effect of the interaction between the pH and adsorbent dosage on Sb(III) removal rate (%). (a) 3D surface plot; (b) contour plot.

Fe diffraction peaks were 711.75 and 724.77 eV, respectively (Fig. 4(a)), indicating that Sb(III) replaced MIL-53(Fe). The -OH in/GO then forms Fe-O-Sb coordination compound with -O-Fe, so that Sb(III) is adsorbed.<sup>38</sup>

Fig. 4(b) shows the XPS spectrum of O1s. The spectrum has a peak binding energy of 532.3 eV, which is attributed to the oxygen atom in the carboxylate group of H<sub>2</sub>BDC.<sup>39,40</sup> These results are similar to the FT-IR spectra, which further confirms the composition of the MIL-53(Fe)/GO composite.

The XPS survey spectrum (Fig. 4(c)) shows the presence of C, N, Cl, O, and Fe elements in MIL-53(Fe)/GO. Fig. 4(d) shows that the Sb peak is added to the spectrum after adsorption and that the position of the Sb peak coincides with the position of O1s. The binding energy of Sb(3d<sub>3</sub>) detected by XPS is 539.9 eV (Fig. 4(e)). Based on the chemical state database of XPS, antimony exists in the form of Sb(III) on the surface of the adsorbent MIL-53(Fe)/GO, and no oxidation-reduction reaction occurs.

### 3.2 Response surface optimization of adsorption conditions

Taking the adsorption time, dosage, pH, temperature, Sb(III) initial concentration, and other factors as independent variables and the Sb(III) removal rate as the response value, a response surface quadratic polynomial model is constructed as expressed in eqn (7). The results and analysis of the variance are provided in ESI 1 and 2,<sup>†</sup> respectively.

$$\begin{aligned}
 Y = & +200.74776 + 10.21229A - 0.49821B - 32.89125C \\
 & + 1.18421D - 2.29854E + 0.01950AB - 0.47750AC \\
 & - 0.02350AD - 0.04250AE - 0.08215BC \\
 & - 6.25 \times 10^{-4}BD + 0.0051BE - 0.17700CD \\
 & + 0.10100CE - 0.01090DE - 0.44104A^2 \\
 & + 0.00861B^2 + 2.72146C^2 + 0.01447D^2 + 0.01229E^2 \quad (7)
 \end{aligned}$$

where  $Y$  is the response value, %;  $A$ ,  $B$ ,  $C$ ,  $D$ , and  $E$  are respectively the adsorption time (h), dosage ( $\text{mg L}^{-1}$ ), pH, temperature

(°C), initial concentration of Sb(III) ( $\text{mg L}^{-1}$ ), and the actual value corresponding to the independent variable, respectively.

A total of 46 runs were undertaken for optimizing the three individual parameters in the BBD; the experimental conditions based on the factorial design are shown in ESI 1.<sup>†</sup> The results show that the Sb(III) removal rate varied in the range of 65.82–97.03%.

From ESI 2,<sup>†</sup> we find that the model  $F$  value is 42.15,  $P < 0.0001$ , indicating that the nonlinear equation relationship between the respective variables described by the regression equation and the response value is significant; the model determination coefficient  $R^2 = 0.971$ , indicating that 2.9% regression equation to explain;  $R_{\text{Adj}}^2 - R_{\text{Pred}}^2 = 0.05$  ( $< 0.2$ ), the coefficient of variation  $CV$  is 2.06% ( $< 10\%$ ), and the signal-to-noise ratio is 25.614 ( $> 4$ ), indicating that the model has high reliability and prediction accuracy and can be used for actual forecast.<sup>41,42</sup> In addition, from ESI 2,<sup>†</sup> we find that the adsorption time, dosage pH, temperature, and initial concentration of Sb(III) significantly affect the removal rate of Sb(III) ( $P < 0.05$ ). In the interaction term, the interaction between the temperature and pH has a significant effect on the removal rate. The removal rate of Sb(III) has a significant impact ( $P < 0.05$ ); in the quadratic term, the pH, temperature, and initial concentration have a significant impact on the removal rate of Sb(III). The  $P$  value of the other factors is greater than 0.05, which has a significant impact on Sb(III). The effect of the removal rate is not significant.

Fig. 5 shows the 3D surface plot and contour plot of the effect of the interaction between the pH value and the amount of adsorbent on the Sb(III) removal efficiency. Fig. 6 shows the 3D surface plot and contour plot of the effect of the interaction between the pH and temperature on the Sb(III) removal efficiency. Fig. 7 shows a 3D surface plot and contour plot of the effect of the interaction between the pH and Sb(III) initial concentration on the Sb(III) removal efficiency. The results show that when the temperature is in the range of 25–45 °C, the pH



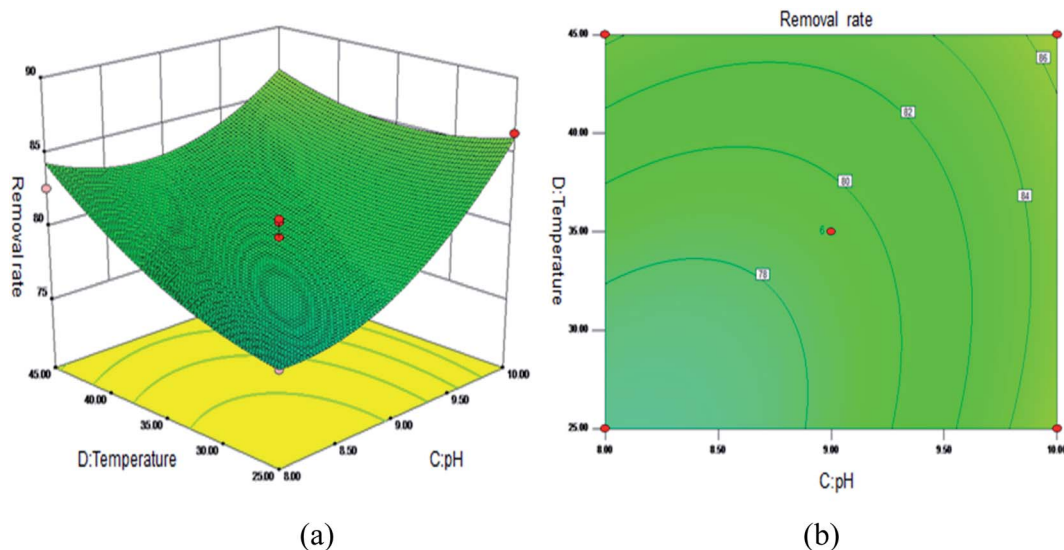


Fig. 6 Effect of the interaction between the pH and temperature on Sb(III) removal rate (%). (a) 3D surface plot; (b) contour plot.

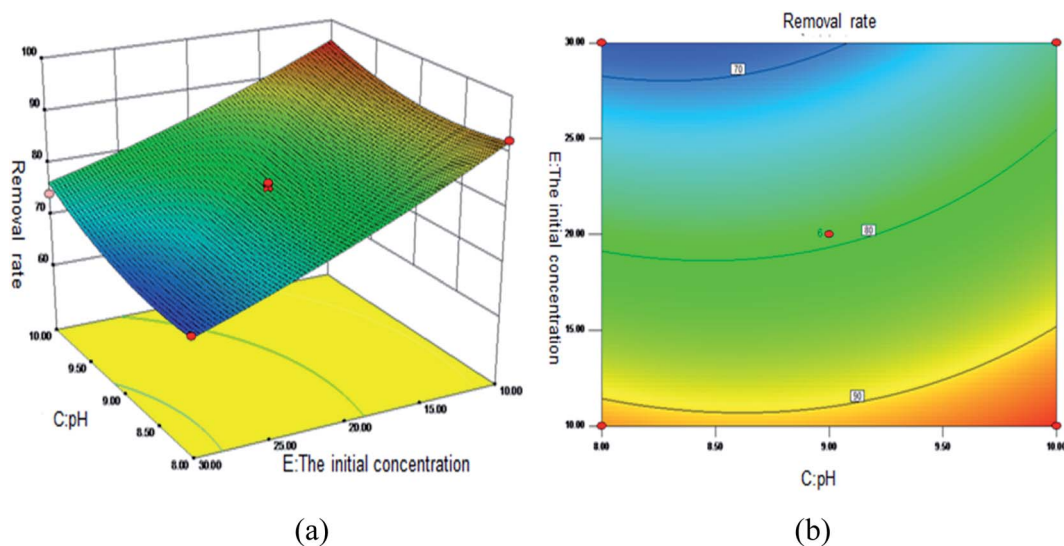


Fig. 7 Effect of the interaction between the pH and Sb(III) initial concentration on Sb removal rate (%). (a) 3D surface map; (b) contour map.

value is in the range of 8.0–10, and the dosage is in the range of 70–90 mg, the Sb(III) removal rate gradually increases. When the initial concentration range of Sb(III) is 10–30, the Sb(III) removal rate decreases. This shows that the adsorption efficiency of MIL-53(Fe)/GO on Sb(III) is proportional to the dosage, temperature and pH value, and inversely proportional to the initial concentration of Sb(III).

As the amount of adsorbent increases, a large amount of adsorbent can provide more adsorption sites, thereby promoting the adsorption reaction. We speculate the adsorption process to be an endothermic reaction. As the temperature increases, the diffusion coefficient of the adsorbate in the pores can be increased, which is beneficial to the adsorption reaction. As the pH increases, the removal rate increases relatively. When the pH increases to 10, the adsorption rate increases

significantly. This is because when  $\text{pH} \geq 10$ , antimony exists in the forms of  $\text{H}_2\text{SbO}_3^-$  and  $\text{Sb}(\text{OH})_4^-$ , and the adsorbent material depletes; protonation makes the negatively charged antimony to easily adsorb under the effect of the electrostatic force. When the initial concentration of the Sb(III) solution increases from  $10 \text{ mg L}^{-1}$  to  $30 \text{ mg L}^{-1}$ , the adsorption capacity of MIL-53(Fe)/GO on Sb(III) increases, but the removal rate decreases. This is because when the quality of the adsorbent MIL-53(Fe)/GO is constant, as the Sb(III) concentration (III) in the solution increases, the Sb(III) concentration difference between the solution and the adsorbent increases. This increases the mass transfer driving force between MIL-53(Fe)/GO and Sb(III), the adsorbed Sb(III) enters the active sites on the surface of the adsorbent MIL-53(Fe)/GO more easily, and the adsorption capacity will be greater. However, the active sites on the surface



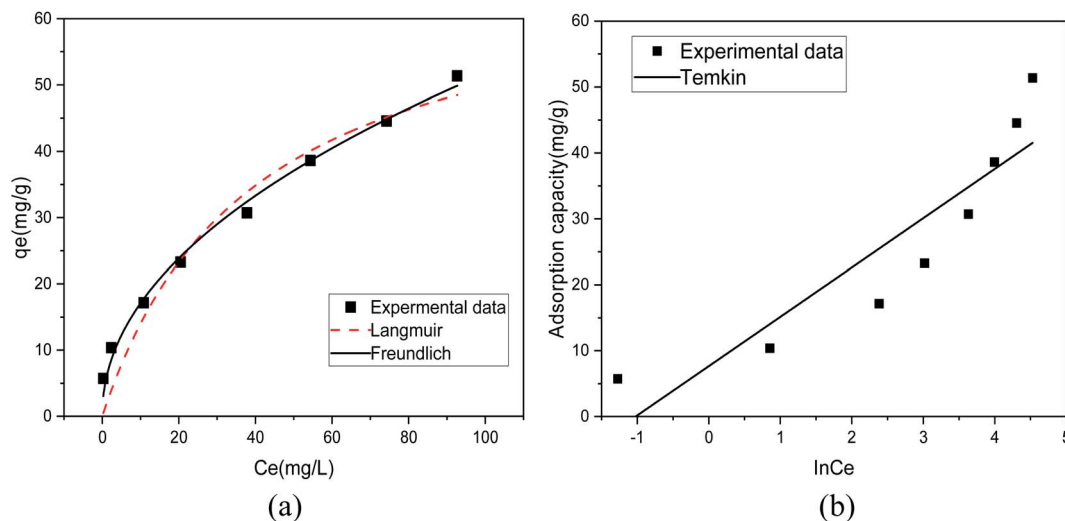


Fig. 8 Isothermal model of Sb(III) adsorption by MIL-53(Fe)/GO.

Table 3 Adsorption isotherm model parameters for the adsorption of Sb(III) by MIL-53(Fe)/GO

Model	Langmuir model	Freundlich model	Temkin model
Parameters	$q_{\max} = 69.014$ $K_L = 0.025$ $R^2 = 0.949$	$K_F = 5.642$ $n = 2.078$ $R^2 = 0.991$	$A = 2.773$ $B = 7.483$ $R^2 = 0.827$

of MIL-53(Fe)/GO are limited; therefore, with the increase in the Sb(III) concentration (III) in the solution, the removal rate decreases instead.

The first derivative of eqn (7) is calculated to determine the adsorption of Sb(III) by MIL-53(Fe)/GO. The optimum conditions are as follows: adsorption time 4.86 h, dosage 85.79 mg, pH = 10.00, temperature 39.29 °C, and Sb(III) initial concentration 10.09 mg L<sup>-1</sup>. Under the best conditions, the removal rate of Sb(III) by MIL-53(Fe)/GO is 97.97%. To verify the prediction results, the experiment was carried out under optimal conditions. The experiment was repeated thrice, and the average value was taken. The average removal rate of Sb(III) was 97.60%, which was close to the predicted value of the model (97.97%), and the prediction accuracy reached 99.62%. The predicted value has a high degree of fit with the experimental value, which has a certain guiding significance.

Table 4 Adsorption kinetic model parameters related to the adsorption of Sb(III) by MIL-53(Fe)/GO

Model	Quasi-first-order dynamics model	Quasi-two-stage dynamic model	Intraparticle diffusion model
Parameter	$q_e = 5.620$ $K_1 = 0.157$ $R^2 = 0.987$	$q_e = 5.824$ $K_1 = 0.061$ $R^2 = 0.999$	$\alpha_1 = 0.177$ $k_1 = 1.207$ $R^2 = 0.963$ $\alpha_1 = 5.154$ $K_2 = 0.043$ $R^2 = 0.882$

### 3.3 Adsorption isotherm model

Fig. 8(a) shows the fitting results of the Langmuir and Freundlich isotherm adsorption models for the adsorption of Sb(III) by MIL-53(Fe)/GO. Fig. 8(b) shows the fitting results of the T-K isotherm adsorption models. The regression coefficients  $R^2$  of the three isotherm adsorption models are 0.949, 0.991, and 0.827, respectively, indicating that the Freundlich model outperforms the Langmuir model in fitting the adsorption and removal processes of Sb(III) by MIL-53(Fe)/GO. The Freundlich isotherm model assumes that the adsorption sites of the adsorbent are evenly distributed on the adsorbent surface, which is a multiphase adsorption surface, or active sites are supported on the surface, suitable for single-layer (chemical) or multilayer (physical) adsorption.<sup>42</sup> Therefore, it is speculated that the adsorption behavior is a heterogeneous reaction.<sup>43,44</sup> The calculated maximum adsorption capacity  $q_{\max}$  is 69.014 mg g<sup>-1</sup> (Table 3).

### 3.4 Adsorption kinetic model

Table 4 lists the results of the fitting-related parameters of the kinetic model for the adsorption of Sb(III) by MIL-53(Fe)/GO. From Table 4, we find that both the quasi-first-order and quasi-second-order kinetic models can fit the experimental data well. The quasi-first-order kinetic model has a high fitting accuracy ( $R^2 = 0.987$ ), whereas the theoretical value ( $q_e = 5.824$  mg g<sup>-1</sup>) fitted by the quasi-second-order kinetic model is close to the experimental value ( $q_e = 5.806$  mg g<sup>-1</sup>), and the fitting accuracy ( $R^2 = 0.999$ ) is greater. Therefore, it is inferred that MIL-53(Fe)/GO is relative to Sb(III) and that the adsorption process follows a quasi-two-stage kinetic model, indicating that the adsorption reaction is mainly chemical adsorption.<sup>43</sup> To identify the antimony migration process on MIL-53(Fe)/GO crystals, based on kinetic data, the rate control step was analyzed using the intra-particle diffusion model, which characterizes two or more steps involved in the adsorption process.<sup>45-47</sup> Fig. 9 shows that the adsorption process tends to





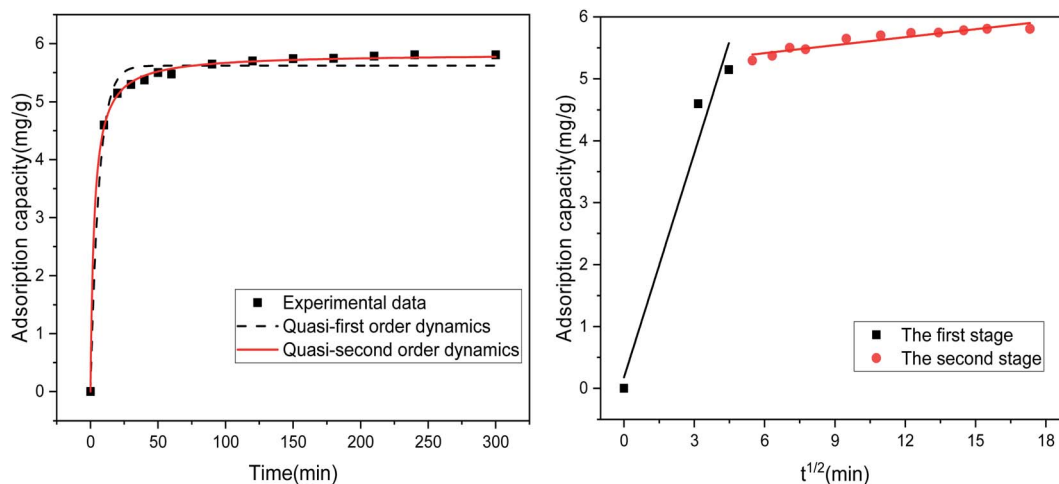


Fig. 9 Kinetic model of the adsorption of Sb(III) by MIL-53(Fe)/GO.

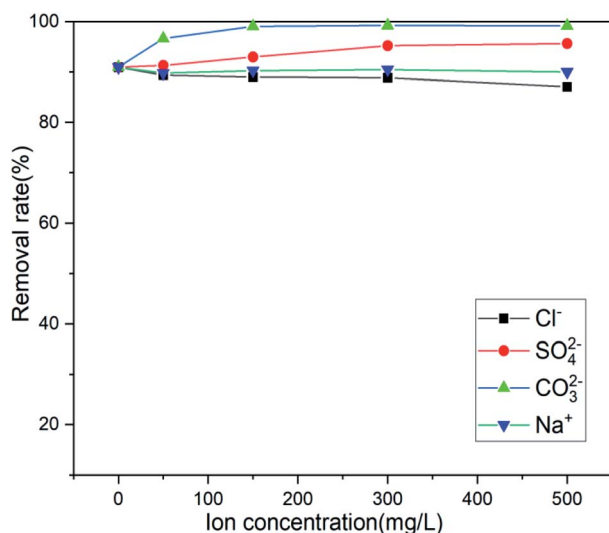


Fig. 10 Effect of coexisting ions on adsorption.

two stages. The adsorption process is carried out by surface adsorption and intra-particle diffusion (antimony is transferred from the boundary membrane to the adsorbent surface and then from the adsorbent surface to the active site or binding site in the particle).<sup>48</sup> Based on the particle diffusion model, the slope of the linear graph is defined as the diffusion rate constant. The slopes of the fitting lines in the two stages are different, indicating a gradual stage in the adsorption process. The  $k_1 > k_2$  of the two steps means that the adsorption process starts from the beginning. The adsorption rate of external diffusion is the highest, and the current adsorption rate is mainly controlled by pore (in-particle) diffusion rather than boundary layer (external) diffusion.<sup>49,50</sup>

### 3.5 Influence of coexistent ions

Actual wastewater is often complex in terms of its composition. There are multiple anions and different heavy metal cations.

Because the adsorption process will compete with the target pollutants to be removed, the existence of coexisting ions is another important factor affecting the actual removal capacity in practice.<sup>51,52</sup> Several anions coexisting near the mining area were selected ( $\text{Cl}^-$ ,  $\text{SO}_4^{2-}$ , and  $\text{CO}_3^{2-}$ ), and the influence of anion concentration on the adsorption efficiency of the adsorbent was investigated. Fig. 10 shows that only  $\text{Cl}^-$  has a slight effect on the adsorption and that  $\text{SO}_4^{2-}$  and  $\text{CO}_3^{2-}$  promote adsorption,  $\text{Na}^+$  has no effect on its adsorption, which is beneficial to the adsorption of Sb(III) by the adsorbent in actual wastewater. When  $\text{Mn}^{2+}$  and  $\text{Pb}^{2+}$  coexist, the adsorbent will only adsorb Sb(III), showing a high degree of specificity for Sb(III).

### 3.6 Recycling of MIL-53(Fe)/GO

The recyclable performance of adsorbents is an important indicator that can help evaluate their economical aspects in practical applications; adsorption materials that can be recycled

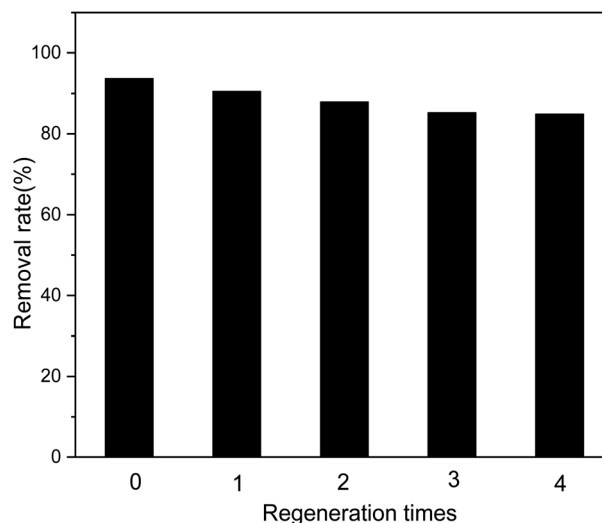


Fig. 11 Effect of regeneration time on adsorption efficiency.



multiple times are economical. As shown in Fig. 11, after four cycles of use of the adsorbent, the removal efficiency for Sb(III) in the solution does not decrease significantly.

## 4. Conclusions

The response surface optimization experiments showed that the dosage, pH, temperature, and initial concentration of Sb(III) significantly affect the adsorption of Sb(III) by MIL-53(Fe)/GO. The adsorption time had no significant effect on it. For optimal adsorption, the adsorption time, dosage, pH, temperature, and initial concentration of Sb(III) should be set to 4.86 h, 85.79 mg, 10.00, 39.29 °C, and 10.09 mg L<sup>-1</sup>, respectively. The average removal rate was as high as 97.60%. The Freundlich isotherm model could effectively fit the adsorption process of Sb(III) by MIL-53(Fe)/GO ( $R^2 = 0.991$ ). The maximum adsorption capacity was 69.014 mg g<sup>-1</sup>, and the adsorption was a heterogeneous reaction, exhibiting quasi-secondary kinetics. The scientific model could better fit the adsorption kinetic process ( $R^2 = 0.999$ ). The adsorption process was mainly chemical adsorption, and the adsorption process was carried out through surface adsorption and intraparticle diffusion. The FT-IR analysis showed that the O–H, N–H bond, and hydroxyl (–OH) in the polysaccharide play a major role in the adsorption of Sb(III) and that the carboxyl group reacts with Sb(III) chemically. The abundant hydroxyl on the adsorbent surface was the main reason for the excellent adsorption and removal performance of Sb(III). The hydroxyl group in Fe–O–OH was replaced by Sb(III) to form a new complex Fe–O–Sb. In the presence of Cl<sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, CO<sub>3</sub><sup>2-</sup>, and Na<sup>+</sup>, the adsorbent still showed superiority in the adsorption of antimony. In the presence of both Mn<sup>2+</sup> and Pb<sup>2+</sup>, the adsorbent showed specificity to antimony. Even after the fourth cycle of use, it still exhibited a high removal rate, making it a potential adsorbent for removing antimony from water.

## Conflicts of interest

The authors declare no conflict of interest regarding the publication of this paper. We do not have any commercial or associative interest that would have a potential conflict of interest in connection with the work submitted.

## Acknowledgements

This study was funded by the Natural Science Foundation of Hunan Province (grant number 2019JJ50153) and the National Natural Science Foundation of China (grant number 51604113).

## References

- M. Satoshi, T. Yoshio, S. Yoichi, *et al.*, Interaction of synthetic sulfate green rust with antimony(V), *Environ sci & technol.*, 2009, **43**, 318–323.
- J. J. Beun, M. C. M van Loosdrecht and J. J. Heijnen, Aerobic granulation in a sequencing batch airlift reactor, *Water Res.*, 2002, **36**, 702–712.
- S. Sundar and J. Chakravarty, Antimony Toxicity, *IJERPH*, 2010, **7**, 4267–4277.
- E. S. Safaa, F. Nermine, I. H. D., *et al.*, The potential DNA toxic changes among workers exposed to antimony trioxide, *Environ. Sci. Pollut. Res.*, 2017, **24**, 12455–12461.
- W. Bing, Z. J. Wei, Z. A. Guo, *et al.*, A review of antimony (Sb) isotopes analytical methods and application in environmental systems, *Int. Biodeterior. Biodegradation.*, 2017, **128**, 109–116.
- C. J. Boreiko and T. G. Rossman, Antimony and its compounds: Health impacts related to pulmonary toxicity, cancer, and genotoxicity, *Toxicol. Appl. Pharmacol.*, 2020, **403**, 115–156.
- J. Li, B. Zheng, Y. He, *et al.*, Antimony contamination, consequences and removal techniques: A review, *Ecotoxicol. Environ. Saf.*, 2018, **156**, 125–134.
- X. Hu, M. He and S. Li, The leaching characteristics and changes in the leached layer of antimony-bearing ores from China, *J. Geochem. Explor.*, 2017, **176**, 76–84.
- M. Kang, T. Kamei and Y. Magara, Comparing polyaluminum chloride and ferric chloride for antimony removal, *Water Res.*, 2003, **37**, 4171–4179.
- S. Rakshit, D. Sarkar, P. Punamiya, *et al.*, Antimony sorption at gibbsite. Water interface, *Chemosphere*, 2011, **84**, 480–483.
- H. W. Langmi, J. Ren, B. North, *et al.*, Hydrogen Storage in Metal-Organic Frameworks: A Review, *Electrochim. Acta.*, 2014, **128**, 368–392.
- V. Finsy, L. Ma, L. Alaerts, *et al.*, Separation of CO<sub>2</sub>/CH<sub>4</sub> mixtures with the MIL-53(Al) metal-organic framework, *Mesopor. Mat.*, 2009, **120**, 221–227.
- Y. Chen, R. Zhang, L. Jiao, *et al.*, Metal-organic framework-derived porous materials for catalysis, *Coord Chem Rev*, 2018, **36**, 21–23.
- H. Li, K. Wang, Y. Sun, *et al.*, Recent advances in gas storage and separation using metal-organic frameworks, *Mater. Today.*, 2018, **21**, 108–121.
- J. Wu and B. Yan, Eu(III)-functionalized In-MOF (In(OH) bpydc) as fluorescent probe for highly selectively sensing organic small molecules and anions especially for CHCl<sub>3</sub> and MnO<sub>4</sub>, *J. Colloid Interface Sci.*, 2017, **504**, 197–205.
- L. Zhang, Y. Chen, R. Shi, *et al.*, Synthesis of hollow nanocages MOF-5 as drug delivery vehicle to solve the load-bearing problem of insoluble antitumor drug oleanolic acid (OA), *Inorg. Chem. Commun.*, 2018, **96**, 20–23.
- N. Yin, K. Wang, L. Wang, *et al.*, Amino-functionalized MOFs combining ceramic membrane ultrafiltration for Pb (II) removal, *Chem. Eng. J.*, 2016, **306**, 619–628.
- P. Camille and J B T., Exploring the coordination chemistry of MOF-graphite oxide composites and their applications as adsorbents, *Dalton Trans.*, 2012, **41**, 27–35.
- Z. Yang, X. Xu, X. Liang, *et al.*, MIL-53(Fe)-graphene nanocomposites: Efficient visible-light photocatalysts for the selective oxidation of alcohols, *Appl. Catal. B: Envir.*, 2016, **198**, 112–123.
- C. Zhang, W. Hu, H. Jiang, *et al.*, Electrochemical performance of MIL-53(Fe)@RGO as an Organic Anode



- Material for Li-ion Batteries, *Electrochim. Acta.*, 2017, **246**, 528–535.
- 21 M. A. G. T. van den Hoop, H. P. van Leeuwen, J. Pinheiro, *et al.*, Voltammetric analysis of the competition between calcium and heavy metals for complexation by humic material, *Colloid Surface A: Physicochem. Eng. Aspect.*, 1995, **95**, 305–313.
- 22 J. Yu, J. Cao, Z. Yang, *et al.*, One-step synthesis of Mn-doped MIL-53(Fe) for synergistically enhanced generation of sulfate radicals towards tetracycline degradation, *J. Colloid Interface Sci.*, 2020, **580**, 470–479.
- 23 Z. Yang, X. Xu, X. Liang, *et al.*, MIL-53(Fe)-graphene nanocomposites: Efficient visible-light photocatalysts for the selective oxidation of alcohols, *Appl. Catal. B: Environ.*, 2016, **198**, 112–123.
- 24 J. Lin, H. Hu, N. Gao, *et al.*, Fabrication of GO@MIL-101(Fe) for enhanced visible-light photocatalysis degradation of organophosphorus contaminant, *J. Water Process. Eng.*, 2020, **33**, 1010.
- 25 T. T. Quang, N. X. Truong, T. H. Minh, *et al.*, Enhanced Photocatalytic Degradation of MB Under Visible Light Using the Modified MIL-53(Fe), *Top Catal.*, 2020, **63**, 1–13.
- 26 J. Jia, X. Fujian, L. Zhou, *et al.*, Metal-organic framework MIL-53(Fe) for highly selective and ultrasensitive direct sensing of MeHg<sup>+</sup>, *Chem. comm.*, 2013, **49**, 4670–4672.
- 27 N. Xiong, P. Wan, G. Zhu, *et al.*, Sb(III) removal from aqueous solution by a novel nano-modified chitosan (NMCS), *Sep. Purif. Technol.*, 2020, 236.
- 28 S. Zuleika, C. Muriel, R. Sindia, *et al.*, Effects of Phosphorylation in Chlamydomonas Centrin Ser 167, *Calcium binding proteins*, 12006, 108–114.
- 29 M. Tella and G. S. Pokrovski, Antimony(III) complexing with O-bearing organic ligands in aqueous solution: An X-ray absorption fine structure spectroscopy and solubility study, *Geochim. Cosmochim. Ac.*, 2008, **73**, 268–290.
- 30 P. Subhashree, S. Sarita and R. L. Chand, Characterization of various functional groups present in the capsule of Microcystis and study of their role in biosorption of Fe, Ni and Cr, *Bioresour. technol.*, 2007, **98**, 595–601.
- 31 M. R. Vengatesan, S. Singh, S. Stephen, *et al.*, Facile synthesis of thermally reduced graphene oxide-sepiolite nanohybrid via intercalation and thermal reduction method, *Appl. Clay Sci.*, 2017, 510–515.
- 32 K. S. W. Sing, Reporting physisorption data for gas/solid systems with special reference to the determination of surface area and porosity (Provisional), *Pure App. Chem.*, 2013, **54**, 2201–2218.
- 33 H. T. Minh Thanh, T. T. Thu Phuong, P. T. Le Hang, *et al.*, Comparative study of Pb(II) adsorption onto MIL-101 and Fe-MIL-101 from aqueous solutions, *J. Environ. Chem. Eng.*, 2018, **6**, 4093–4102.
- 34 M. Jian, B. Liu, G. Zhang, *et al.*, Adsorptive removal of arsenic from aqueous solution by zeolitic imidazolate framework-8 (ZIF-8) nanoparticles, *Colloids. Surf. A: Physicochem. Eng. Aspects*, 2015, **465**, 67–76.
- 35 C. Yu, L. Gou, X. Zhou, *et al.*, Chitosan-Fe<sub>3</sub>O<sub>4</sub> nanocomposite based electrochemical sensors for the determination of bisphenol A, *Electrochim. Ac.*, 2011, **56**, 9056–9063.
- 36 D. Huang, J. Li, G. Zeng, *et al.*, Facile construction of hierarchical flower-like Z-scheme AgBr/Bi<sub>2</sub>WO<sub>6</sub> photocatalysts for effective removal of tetracycline: Degradation pathways and mechanism, *Chem. Eng. J.*, 2019, **375**, 1–15.
- 37 S. Sepúlveda-Guzmán, L. Lara, O. Pérez-Camacho, *et al.*, Synthesis and characterization of an iron oxide poly(styrene-co-carboxybutylmaleimide) ferrimagnetic composite, *Polym*, 2006, **48**, 720–727.
- 38 L. Wang, J. Wang, Z. Wang, *et al.*, Synthesis of Ce-doped magnetic biochar for effective Sb(V) removal: Performance and mechanism[J], *Powder Technology*, 2019, **345**, 501–508.
- 39 W. Huang, N. Liu, X. Zhang, *et al.*, Metal organic framework g-C<sub>3</sub>N<sub>4</sub>/MIL-53(Fe) heterojunctions with enhanced photocatalytic activity for Cr(VI) reduction under visible light, *App. Surf. Sci.*, 2017, **425**, 107–116.
- 40 L. Hu, G. Deng, W. Lu, *et al.*, Deposition of CdS nanoparticles on MIL-53(Fe) metal-organic framework with enhanced photocatalytic degradation of RhB under visible light irradiation, *App. Surf. Sci.*, 2017, **410**, 401–413.
- 41 C. Song, Y. Kitamura and S. Li, Optimization of a novel cryogenic CO<sub>2</sub> capture process by response surface methodology (RSM), *J. Taiwan Inst. Chem. Eng.*, 2014, **45**(4), 1666–1676.
- 42 M. Babar, M. A. Bustam, A. S. Maulud, *et al.*, Enhanced cryogenic packed bed with optimal CO<sub>2</sub> removal from natural gas; a joint computational and experimental approach, *Cryogenics*, 2020, **105**(C), 103010–103042.
- 43 K. Walsh, S. Mayer, D. Rehmann, *et al.*, Equilibrium data and its analysis with the Freundlich model in the adsorption of arsenic(V) on granular ferric hydroxide, *Sep. Purif. Technol.*, 2020, **243**, 116704.
- 44 M. Jian, B. Liu, G. Zhang, *et al.*, Adsorptive removal of arsenic from aqueous solution by zeolitic imidazolate framework-8 (ZIF-8) nanoparticles, *Colloid. Surface. A: Physicochem. Eng. Aspect.*, 2015, **465**, 67–76.
- 45 J. A. Pedit and C. T. Miller, Heterogeneous sorption processes in subsurface systems. Model formulations and applications, *Environ. sci. technol.*, 1994, **28**, 94–104.
- 46 Z. Qi, T. P. Joshi, R. Liu, *et al.*, Synthesis of Ce(III)-doped Fe<sub>3</sub>O<sub>4</sub> magnetic particles for efficient removal of antimony from aqueous solution, *J. Hazard. Mater.*, 2017, **329**, 193–204.
- 47 V. Vadivelan and K. V. Kumar, Equilibrium, kinetics, mechanism, and process design for the sorption of methylene blue onto rice husk, *J. Colloid. Interface Sci.*, 2005, **286**, 90–100.
- 48 Y. S. Ho and G. McKay, Sorption of dyes and copper ions onto biosorbents, *Process Biochem*, 2003, **38**, 1047–1061.
- 49 Y. S. Ho and G. McKay, The kinetics of sorption of basic dyes from aqueous solution by sphagnum moss peat, *Can. J. Chem. Eng.*, 1998, **76**, 822–827.
- 50 G. Kaushik and G. U. Chand, Arsenic removal using hydrous nanostructure iron(III)-titanium(IV) binary mixed oxide from aqueous solution, *J. Hazard. Mater.*, 2009, **161**, 884–892.



- 51 W. H. Cheung, Y. S. Szeto and G. McKay, Intraparticle diffusion processes during acid dye adsorption onto chitosan, *Bioresour. Technol.*, 2006, **982**, 897–904.
- 52 L. Jie, W. Yi-nan, L. Zehua, *et al.*, Characteristics of arsenate removal from water by metal-organic frameworks (MOFs), *Water sc. technol.*, 2014, **70**, 1391–1397.

