


Cite this: *RSC Adv.*, 2021, 11, 31983

The effects of green waste compost on soil N, P, K, and organic matter fractions in forestry soils: elemental analysis evaluation

Xiaojie Feng,^a Xiangyang Sun,^a ^{*,a} Wenjie Zhou,^a Wei Zhang,^b Feiwei Che^a and Suyan Li^a

We study the effects of green waste compost on soil fertility to provide a theoretical basis for accurately improving forestry soil quality. This study aims to investigate the effects of green waste compost on soil N, P, K, and soil organic matter (SOM) fractions using elemental and FTIR analyses. Therefore, five fertilization treatments were set up for research, including mineral fertilization (M-fert), green waste compost fertilization (G-fert), standard rate of M-fert plus G-fert (GM-fert), half the standard rate of M-fert plus G-fert (1/2 GM-fert), and a control with no fertilizer addition (N-fert). The results showed that GM-fert treatment significantly increased the content of soil $\text{NH}_4\text{-N}$, available phosphorus (AP), available potassium (AK), water soluble organic carbon (WSOC), humus (HE), and humic acid (HA), which were 8.53 ± 0.67 , 76.1 ± 5.96 , 168 ± 3.42 , 0.152 ± 0.01 , 5.64 ± 0.15 , and $4.69 \pm 0.21 \text{ mg kg}^{-1}$, respectively. The content of HA (36.7%, $F = 7.55$, $P = 0.01$) was positively correlated with the soil N, P, K, and the HA absorption peak. The relative intensities of the alcohol -OH , aliphatic -CH and carbohydrate C-O peaks showed the largest changes, which were $18.6 \pm 0.56\%$, $13.1 \pm 0.33\%$, and $16.3 \pm 0.49\%$. -CH/C=C (49.8%, $F = 12.9$, $P < 0.01$) was also significantly positively correlated with soil N, P, K. In conclusion, green waste compost significantly increased soil N, P, K, and HA in forestry soils, and the -CH/C=C of HA was the main factor related to soil nutrients.

Received 28th June 2021
Accepted 15th September 2021

DOI: 10.1039/d1ra04986k

rsc.li/rsc-advances

Introduction

Forestry soil is a part of the ground surface, and supplies the living materials of forest plants. Forestry soil is composed of minerals and organic matter, and contains certain amounts of air, moisture, and organisms; it mainly includes natural forest soil, man-made forest soil, nursery soil and urban green space soil. Forestry soil provides the nutrient supply for normal forest growth, and the soil nutrient contents are important indexes to evaluate the forest health status.^{1,2} Therefore, increasing the N, P, K content of soil provides a better growth environment for forests.

Fertilization is an important technical measure to improve soil fertility and forest productivity.³ Soil organic matter (SOM) is the key factor for improving soil fertility; it not only stores mineral elements in the form of organic matter, but also retains part of the available nutrients *via* ion adsorption.⁴ Based on differences in acidic and alkali solutions, the SOM fractions were divided into water soluble organic carbon (WSOC), humus (HE), fulvic acid (FA), humic acid (HA) and humin (HM).^{5,6} M-fert + maize straw or biochar increased HA, and HA had

higher carbon storage than HM and FA in agricultural land soils,^{7,8} which played a major role in stabilizing the organic carbon pool of the soil. The SOM chemical structures were an important component to maintain soil nutrients, and SOM decomposition released nutrients due to organic molecular recombination *via* microbial activity.⁹ The development of molecular-scale techniques provided new methods for studying SOM structures.^{10,11} Many important SOM structures, such as alkyl-C, aromatic-C, and carboxyl-C, were discovered by infrared spectroscopy.¹² According to the wave bands and specific structures, four functional groups could be described: alcohol -OH , aliphatic -CH , amide group C=O and carbohydrate C-O .¹³ The characteristics of the functional groups at the molecular scale could reflect the carbon storage and decomposition in forestry soil under M-fert and G-fert treatments. Fertilization had a great influence on the SOM fractions, which were directly related to the release of soil N, P, K. Using infrared spectroscopy technology to explore the structure of the SOM fractions would give a deeper understanding the characteristics of soil organic carbon fixation and the decomposition mechanism in forestry soils.

“Green waste compost fertilization” (G-fert) refers to fallen leaves and branch cuttings transformed into potentially safe, stable organic fertilizer using composting technology, which has fast decomposition in the soil.¹⁴ The fast turnover of G-fert

^aCollege of Forestry, Beijing Forestry University, Beijing 100083, PR China. E-mail: sunxy@bjfu.edu.cn

^bBeijing Tongzhou District Gardening and Greening Bureau, Beijing 100013, PR China


increased soil N, P, K for tree growth.¹⁵ A few studies have investigated the characteristics of the SOM fractions under G-fert treatment, which also formed a circulatory system between green waste resources and forestry soils. As a consequence, the objectives of this study were to determine (1) the effects of green waste compost on soil N, P, K and SOM fractions, and (2) the relationship between the SOM fractions and soil N, P, K. This study provides the mechanisms of the underlying changes in the SOM fractions under different fertilization treatments, which supports soil carbon storage in forestry soils.

Material and methods

Chemical properties of fertilizers

The green waste was composed of tree branches and fallen leaves, the main sources of which were *Paulownia*, *Fraxinus*, Maple, which are all common tree species in the Beijing area. The green waste compost used in the study was aerobically composted for two years under natural aerobic conditions with no accelerating agents. The pH was 8.27, with a total organic carbon of 322 g kg⁻¹, total N of 15.5 g kg⁻¹, total P of 2.25 g kg⁻¹, and total K of 7.14 g kg⁻¹. Before incorporation into the soil, the air-dried green waste compost was sieved through a 2 mm sieve. The mineral fertilization (M-fert) treatment consisted of urea (N 46.4%), superphosphate (P₂O₅ 12.0%) and potassium chloride (K₂O 60.0%). The M-fert was dissolved into liquid and then mixed evenly with the soil in the pot.

Experimental design

Soil was taken from a nursery in which poplar (*Populus L.*) had been growing for four years. Then soil was air-dried, and coarse rocks and plant residue were removed before sieving the soil through 2 mm nylon mesh. The soil was developed from moisture and river deposits, with a pH of 8.17. The total N was 0.960 g kg⁻¹, the NH₄-N was 6.38 mg kg⁻¹, the NO₃-N was 0.490 mg kg⁻¹, the AP was 20.6 mg kg⁻¹, and the AK was 97.2 mg kg⁻¹.

The pot experiment was carried out from August to November in 2019 at the nursery garden (Sanqinyuan) of Beijing Forestry University. The experiment had five treatments: (1) M-fert; (2) G-fert; (3) standard rate of M-fert plus G-fert (GM-fert); (4) half the standard rate of M-fert plus G-fert (1/2 GM-fert); (5) control with no fertilizer addition (N-fert). The fertilizer types and fertilizer amounts under different fertilization treatments are listed in Table 1. The NPK fertilizers were applied in a single basal application before planting the poplar. Each fertilization treatment had five replicates, and 25 pots were used in total (5 × 5 = 25). The progression of the experiment involved three main steps: (1) 20 kg of air-dried soil was placed in each pot (height 45 cm; volume 20 L). (2) G-fert was mixed uniformly with the air-dried soil under the G-fert, 1/2 GM-fert and GM-fert treatments. (3) After setup, one poplar (*Populus L.*), which was about 50 cm though cutting, was transplanted into each pot, and then all pots were watered with the M-fert solution. During the pot experiment, according to the dryness or humidity of the

soil in the pot, the soil field water holding capacity was kept at about 60% to ensure the normal growth of the poplar trees.

Soil sampling and soil analyses

Soil samples were taken from each pot using an auger 90 days after cutting the poplar (*Populus L.*) and then air-dried and sieved through a 2 mm and 0.25 mm sieve to measure the soil chemical properties and SOM fractions.

The soil chemical properties included the pH value, SOM, total nitrogen (TN), ammonium nitrogen (NH₄-N), nitrate nitrogen (NO₃-N), available phosphorus (AP) and available potassium (AK).¹⁶ Soil pH: 10 g of air-dried soil was weighed out, and 25 mL of carbon dioxide-free water was added. The mixture was stirred vigorously for 1–2 minutes, and then the soil pH was determined after 30 minutes using a pH meter (distinguishability 0.01 pH, error 0.01 pH). SOM: 0.5 g of air-dried soil was weighed out, and 5 mL H₂SO₄ and 5 mL 0.8 N K₂Cr₂O₇ were added. The mixture was then heated in an oil bath for 5 minutes (180 °C) and titrated using FeSO₄. Soil TN: 0.2 g of air-dried soil was weighed out. H₂SO₄ and mixed catalyst were added, and then the mixture was digested for 2 hours (180–380 °C) and measured using an Automatic Kjeldahl nitrogen analyzer (K1100, China, titration accuracy 1.0 μL per step, repeatability error ≤0.5%). NH₄-N: the soil was extracted using 1 mol L⁻¹ KCL and measured using indigo colorimetry with an ultraviolet spectrophotometer (JC-UT2000, China, wavelength precision ±1 nm, photometric accuracy ±0.5% T) at 625 nm. NO₃-N: the soil was extracted using 1 mol L⁻¹ KCL and measured using an ultraviolet spectrophotometer (JC-UT2000, China, wavelength precision ±1 nm, photometric accuracy ±0.5% T) at 220 and 275 nm. AP: the soil was extracted using 0.5 mol L⁻¹ NaHCO₃ and measured using the molybdenum–antimony colorimetric method using an ultraviolet spectrophotometer (JC-UT2000, China, wavelength precision ±1 nm, photometric accuracy ±0.5% T) at 625 nm. AK: the soil was extracted using 1 mol L⁻¹ CH₃COONH₄ and determined using a flame spectrophotometer (EP6410, China, stability ≥97%, accuracy ≥97%). For the procedural blank for the soil chemical properties, the operation steps were the same except for the soil samples. The spike recoveries of the soil pH value, SOM, TN, NH₄-N, NO₃-N, AP and AK were 99.0%, 98.6%, 98.6%, 98.6%, 101.5%, 97.5% and 98.8%.

The SOM was separated into the fractions WSOC, HE, FA, HA and HM, and the extraction steps were as follows: first, 50 mL distilled water was mixed with the soil sample (10 g, <0.25 mm), stirred and shaken in a shock machine. The mixture was then centrifuged, and the supernatant solution was used for the WSOC. Secondly, to the residue was added 40 mL of 0.1 mol L⁻¹ NaOH and 0.1 mol L⁻¹ Na₄P₂O₇ (pH 13) followed by oscillation in a shock machine and centrifugation with a centrifuge. This process was repeated three times, followed by filtration and drying to a constant weight at 55 °C; the residue was the HE. Lastly, 20 mL of 0.5 M H₂SO₄ (pH 1.0) was used to separate the residue into HA and FA by standing overnight. The soil TOC was measured using wet oxidation with K₂Cr₂O₇ and the semi-micro Kjeldahl methods.¹⁶ The concentrations of WSOC, HA, and FA



Table 1 Fertilizer types and amounts under different fertilization schedules^a

Treatment	Fertilizer type			
	Urea (g)	Superphosphate (g)	KCl (g)	Green waste compost fertilizer (kg)
M-fert	22.0	12.5	7.8	0
G-fert	0	0	0	0.66
GM-fert	22.0	12.5	7.8	0.66
1/2 GM-fert	11.0	6.2	3.6	0.33
N-fert	0	0	0	0

^a M-fert = mineral fertilizer; G-fert = green waste compost fertilizer; GM-fert = standard rate of M-fert plus G-fert; 1/2 GM-fert = half the standard rate of M-fert plus G-fert; N-fert = control with no fertilizer addition.

were determined using a TOC analyzer (Shimadzu TOC-V, Japan, precision $\leq 4\%$, accuracy $\geq 98\%$).⁶ For the procedural blank for the soil SOM fractions, the operation steps were the same except for the soil samples. The spike recoveries of the soil TOC, WSOC, HA and FA were 98.5%, 99.6%, 99.5% and 99.6%.

HA elemental and FTIR analyses

For HA purification, we first added deionized water to dissolve the ions, then placed the sample in a dialysis bag for dialysis until no Cl^- could be detected in the solution, and finally freeze-dried the solution to obtain the purified HA. The purified HA was subjected to elemental and infrared spectroscopy analysis. The purified HA was directly measured in CHNS mode using an elemental analyzer (Elementar VE cube, Germany, precision $\leq 0.1\%$, accuracy $\geq 99.0\%$) for elemental analysis, and the content of the element O was obtained by the subtraction method. The HA structure was measured using a Fourier Transform Infrared (FT-IR) Spectrometer (Spectrum 100, USA, spectrum precision 0.008 cm^{-1} , spectrum accuracy 0.02 cm^{-1}).¹⁷ The specific steps are as follows: firstly, the soil sample was baked in an oven for 8 hours at a temperature of 60°C . The purpose of this was to remove the moisture in the soil sample and reduce the interference of hydroxyl functional groups in the spectrum. Secondly, the soil sample was mixed with KBr in the desired ratio (soil sample : KBr 1 : 200, w/w) in an agate mortar, and then pressed into tablets until the color of the tablet was light yellow and relatively uniform. Lastly, the tablets were scanned using an infrared spectrometer. The measurement range of the spectrum is $4000\text{--}500 \text{ cm}^{-1}$, the resolution is 0.5 cm^{-1} , the spectral precision is 0.008 cm^{-1} and the spectral accuracy is 0.02 cm^{-1} . For the procedural blanks for HA elemental and infrared spectroscopy analysis, the operation steps were the same except for the HA samples. The spike recoveries of the HA elemental and infrared spectroscopy analyses were 99.0% and 99.6%.

Statistics

Based on the change in the TOC content, the remained carbon (RC), organic carbon decomposition rate (DR) and humification coefficient (HC) of the green waste compost were calculated.

$$\text{RC} = \text{TOC}_{\text{G-fert,GM-fert,1/2GM-fert}} - (\text{TOC}_{\text{soilM-fert}} - \text{TOC}_{\text{soilN-fert}})$$

$$\text{DR} = [1 - (\text{TOC}_{\text{soilM-fert}} - \text{TOC}_{\text{soilN-fert}}) / \text{TOC}_{\text{G-fert,GM-fert,1/2GM-fert}}] \times 100\%$$

$$\text{HC} = (\text{TOC}_{\text{soilM-fert}} - \text{TOC}_{\text{soilN-fert}}) / \text{TOC}_{\text{G-fert,GM-fert,1/2GM-fert}}$$

Here, $\text{TOC}_{\text{G-fert,GM-fert,1/2GM-fert}}$ are the TOC values for the G-fert, GM-fert and 1/2 GM-fert treatments, and $\text{TOC}_{\text{soilM-fert}}$ and $\text{TOC}_{\text{soilN-fert}}$ are the soil TOC for the M-fert and N-fert treatments.

The HE and HM were calculated as shown below:

$$\text{HE}_{\text{soil}} = \text{HA}_{\text{soil}} + \text{FA}_{\text{soil}}$$

$$\text{HM}_{\text{soil}} = \text{TOC}_{\text{soil}} - (\text{WSOC}_{\text{soil}} + \text{HE}_{\text{soil}})$$

Here, TOC_{soil} , $\text{WSOC}_{\text{soil}}$, HA_{soil} , and FA_{soil} are the TOC values of soil, WSOC values of soil, HA values of soil and FA values of soil under the M-fert, G-fert, GM-fert, 1/2 GM-fert and N-fert treatments, respectively.

Statistical analyses were conducted using DPS11.0 and Excel 2013. One-way ANOVA and *T*-tests were applied for the analysis of the variance of the soil properties to determine significant relationships among different fertilization treatments. Transformation of the infrared spectral images from absorbance to transmittance curves was conducted in OMNIC 8.2. Regression analyses were used to evaluate the influence between the soil N, P, K, SOM fractions, and the HA main absorption peak relative intensity under different fertilization treatments. Redundancy analysis (RDA) was used to explain the SOM fractions, and the contribution of the relative intensity of the main HA absorption peak to the variation of the soil N, P, K, and RDA was performed using Canoco 5 and Cano Draw for Windows. For all analyses, $P < 0.05$ was considered statistically significant. All charts were created in Origin 8.1 and Excel 2013.

Results

The decomposition rate (DR) and humification coefficient (HC) of green waste compost

Table 2 shows the decomposition and residue of green waste compost in forestry soils. The RC was highest under G-fert



Table 2 Decomposition and residue of green waste compost in forestry soils^a

Treatment	AC (g)	RC (g)	DR (%)	HC
M-fert	—	—	—	—
G-fert	209 ± 9.87 a	142 ± 2.54 a	68.0 ± 1.55 a	0.320 ± 0.01 b
GM-fert	209 ± 8.45 a	137 ± 2.78 b	65.6 ± 1.32 b	0.344 ± 0.01 a
1/2 GM-fert	106 ± 5.22 b	68.3 ± 1.99 c	64.2 ± 1.44 b	0.358 ± 0.02 a

^a Table data are mean ± standard deviation. M-fert = mineral fertilizer; G-fert = green waste compost fertilizer; GM-fert = standard rate of M-fert plus G-fert; 1/2 GM-fert = half the standard rate of M-fert plus G-fert; N-fert = control with no fertilizer addition. AC = addition of total TOC; RC = residue of total TOC; DR = GF decomposition rate; HC = humification coefficient. Different lowercase letters indicate a significant difference at the $P < 0.05$ level under different fertilization treatments.

treatment, and lowest under 1/2 GM-fert. The DR values of the green waste compost were $68.0 \pm 1.55\%$, $65.6 \pm 1.32\%$ and $64.2 \pm 1.44\%$ under G-fert, GM-fert and 1/2 GM-fert treatment, respectively. The HC values of G-fert were highest under GM-fert and 1/2 GM-fert treatment, at 0.344 ± 0.01 and 0.358 ± 0.02 .

Soil nutrients under different fertilization treatments

Compared with N-fert treatment, the addition of G-fert increased the OM, TN, and C/N ratio in the soil, which were not influenced under M-fert treatment, and the concentrations of $\text{NH}_4\text{-N}$, $\text{NO}_3\text{-N}$, AP, AK increased under the different fertilization treatments (Table 3). The soil pH values under M-fert and GM-fert treatments were significantly decreased by 0.13 and 0.08 units relative to that under N-fert treatment. The soil OM and TN under G-fert, GM-fert, and 1/2 GM-fert treatments were significantly increased by 31.3–59.3% and 17.2–50.5%. The concentrations of $\text{NH}_4\text{-N}$, AP and AK were highest under the GM-fert treatment, significantly increasing 3.58, 3.50, and 1.77 times compared to those under N-fert treatment, but the concentration of $\text{NO}_3\text{-N}$ was highest under M-fert treatment.

SOM fractions under different fertilization treatments

The addition of green waste compost significantly increased the TOC, HE, HA, FA and HM in the soil compared to those under N-fert treatment, while M-fert had no effect on the SOM fractions except for the WSOC (Table 4). The concentrations of TOC under G-fert and GM-fert treatments were significantly increased 1.55 times and 1.59 times relative to that under N-fert treatment. The concentrations of WSOC, HE and HA were highest under GM-fert treatment at 0.152 ± 0.01 , 5.64 ± 0.15

and $4.69 \pm 0.21 \text{ mg kg}^{-1}$, respectively. However, there was no significant difference in the concentrations of WSOC, HE, and HA under M-fert and N-fert treatments, and FA showed no significant difference under M-fert, G-fert, GM-fert and 1/2 GM-fert treatments. The concentration of HM was highest under G-fert and GM-fert treatments at 4.28 ± 0.50 and $3.91 \pm 0.26 \text{ mg kg}^{-1}$, respectively.

Relationship between soil chemical properties and SOM fractions

Table 5 shows the Pearson correlation coefficients between the soil chemical properties and SOM fractions. The soil TN was significantly ($P < 0.01$) and positively correlated with the contents of TOC ($r = 0.935$), HE ($r = 0.916$), HA ($r = 0.912$) and HM ($r = 0.892$). The soil $\text{NH}_4\text{-N}$ and AK were significantly ($P < 0.05$) and positively correlated with the content of HA ($r = 0.559$) and FA ($r = 0.521$), respectively. The soil AP was significantly ($P < 0.05$) and positively correlated ($r = 0.530\text{--}0.904$) with the SOM values of each fraction. Though RDA analysis, the SOM fractions could explain the changes (62.6%) of the soil chemical properties (Fig. 1). The soil chemical properties were significantly positively correlated with the content of HE (39.5%, $F = 8.49$, $P = 0.01$) and HA (36.7%, $F = 7.55$, $P = 0.01$).

HA element composition and FTIR analysis

In the different fertilization treatments, the main element composition of HA was C and O, with values of 43.0 ± 0.76 to $60.4 \pm 1.45\%$ and 28.0 ± 1.56 to $42.3 \pm 1.77\%$ (Table 6). Compared with those in N-fert treatment, the percentages of C and O increased significantly under G-fert treatment, and

Table 3 The characteristics of the soil chemical properties under different fertilization treatments^a

Treatment	pH	OM g kg ⁻¹	TN g kg ⁻¹	NH ₄ -N mg kg ⁻¹	NO ₃ -N mg kg ⁻¹	AP mg kg ⁻¹	AK mg kg ⁻¹
M-fert	8.01 ± 0.01 c	10.7 ± 0.06 c	1.04 ± 0.06 bc	6.66 ± 0.56 b	5.82 ± 0.15 a	39.8 ± 2.62 d	149 ± 2.80 b
G-fert	8.12 ± 0.01 a	16.3 ± 0.67 a	1.47 ± 0.04 a	4.98 ± 0.30 c	2.48 ± 0.20 c	51.7 ± 4.83 b	124 ± 2.89 d
GM-fert	8.06 ± 0.01 b	16.7 ± 0.36 a	1.49 ± 0.05 a	8.53 ± 0.67 a	5.72 ± 0.28 a	76.1 ± 5.96 a	168 ± 3.42 a
1/2 GM-fert	8.12 ± 0.01 a	13.8 ± 0.31 b	1.16 ± 0.09 b	4.89 ± 0.12 c	5.06 ± 0.36 b	46.3 ± 1.86 c	139 ± 1.39 c
N-fert	8.14 ± 0.01 a	10.5 ± 0.11 c	0.99 ± 0.07 c	2.38 ± 0.18 d	0.76 ± 0.03 d	21.8 ± 0.98 e	95.0 ± 3.85 e

^a Table data are mean ± standard deviation. M-fert = mineral fertilizer; G-fert = green waste compost fertilizer; GM-fert = standard rate of M-fert plus G-fert; 1/2 GM-fert = half the standard rate of M-fert plus G-fert; N-fert = control with no fertilizer addition. Different lowercase letters indicate that soil chemical properties significantly differ at the $P < 0.05$ level under different fertilization treatments.



Table 4 The SOM fractions under different fertilization treatments^a

Treatment	TOC g kg ⁻¹	WSOC mg kg ⁻¹	HE mg kg ⁻¹	HA mg kg ⁻¹	FA mg kg ⁻¹	HM mg kg ⁻¹
M-fert	6.19 ± 0.06 c	0.0720 ± 0.01 d	3.65 ± 0.20 d	2.80 ± 0.15 d	0.842 ± 0.06 ab	2.47 ± 0.16 c
G-fert	9.44 ± 0.67 a	0.101 ± 0.01 c	5.06 ± 0.20 b	4.15 ± 0.27 b	0.911 ± 0.10 a	4.28 ± 0.50 a
GM-fert	9.69 ± 0.36 a	0.152 ± 0.01 a	5.64 ± 0.15 a	4.69 ± 0.21 a	0.951 ± 0.12 a	3.91 ± 0.26 ab
1/2 GM-fert	7.99 ± 0.31 b	0.103 ± 0.01 c	4.42 ± 0.12 c	3.48 ± 0.15 c	0.952 ± 0.06 a	3.47 ± 0.24 b
N-fert	6.09 ± 0.11 c	0.111 ± 0.01 b	3.64 ± 0.20 d	2.87 ± 0.23 d	0.782 ± 0.03 b	2.33 ± 0.10 c

^a Table data are mean ± standard deviation. M-fert = mineral fertilizer; G-fert = green waste compost fertilizer; GM-fert = standard rate of M-fert plus G-fert; 1/2 GM-fert = half the standard rate of M-fert plus G-fert; N-fert = control with no fertilizer addition. Different lowercase letters indicate SOM fractions that significantly differ at the $P < 0.05$ level under different fertilization treatments.

Table 5 The relationship between soil chemical properties and SOM fractions under different fertilization treatments^a

	pH	OM	TN	NH ₄ -N	NO ₃ -N	AP	AK
TOC	0.117	1.000**	0.935**	0.497	0.141	0.852**	0.215
WSOC	0.172	0.505	0.496	0.250	-0.026	0.530*	-0.266
HE	0.026	0.969**	0.916**	0.575*	0.217	0.904**	0.190
HA	0.039	0.962**	0.912**	0.559*	0.183	0.889**	0.137
FA	-0.085	0.534*	0.488	0.413	0.381	0.575*	0.521*
HM	0.193	0.965**	0.892**	0.391	0.063	0.744**	0.239

^a ** means significant correlation at the $P < 0.01$ level; * means significant correlation at the $P < 0.05$ level.

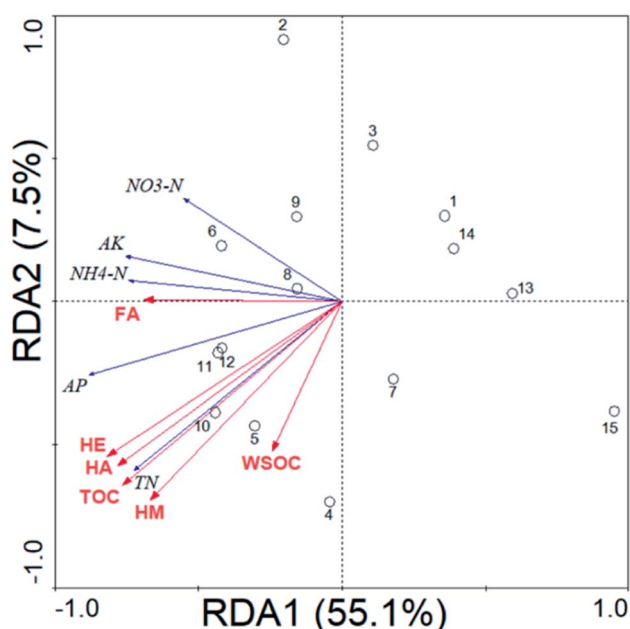


Fig. 1 RDA between soil nutrients and SOM fractions under different fertilization treatments. M-fert = mineral fertilizer (1, 2, 3); G-fert = green waste compost fertilizer (4, 5, 6); 1/2 GM-fert = half the standard rate of M-fert plus G-fert (7, 8, 9); GM-fert = standard rate of M-fert plus G-fert (10, 11, 12); N-fert = control with no fertilizer addition (13, 14, 15).

decreased significantly under M-fert treatment. M-fert treatment decreased C/N, but increased O/C. G-fert treatment increased C/N, but decreased O/C and H/C.

Based on the absorption peaks (Fig. 2 and Table 7), the relative intensities of the main absorption peaks (alcohol -OH, aliphatic -CH, amide group C=O, carbohydrate C-O) of HA changed under the different fertilization treatments (Table 8). The relative intensities of the total absorption peaks increased under G-fert, GM-fert and 1/2 GM-fert fertilization treatments and decreased under M-fert treatment relative to those under N-fert. The relative intensity of the -CH/C=O absorption peak showed the largest changes under 1/2 GM-fert treatment, and the relative intensities of the alcohol -OH, aliphatic -CH and carbohydrate C-O peaks showed the largest changes under GM-fert treatment. The relative intensity of the amide group C=O absorption peak decreased 11.3%, 12.0%, 1.50%, and 14.3% under M-fert, G-fert, GM-fert, and 1/2 GM-fert treatments, respectively.

Relationship between soil N, P, K and HA structures

Table 9 shows the Pearson correlation coefficients between the soil chemical properties and HA structures. The correlation coefficients between -CH/C=O and the TN and AP values were the highest under all fertilization treatments, reaching 0.935 and 0.927. The soil alcohol -OH, aliphatic -CH, amide group C=O, carbohydrate C-O and -CH/C=O were significantly ($P < 0.05$) correlated ($r = -0.669-0.935$) with the TN. The soil amide group C=O was also significantly ($P < 0.01$) negatively correlated ($r = -0.657-(-0.785)$) with the NH₄-N, AP and AK, and -CH/C=O was significantly ($P < 0.05$) positively correlated ($r = 0.630-0.972$) with the NH₄-N, AP and AK. Though RDA analysis, the HA structures could explain the changes (94.3%) in the soil N, P, K (Fig. 3 and Table 10). The soil N, P, K was significantly positively correlated with -CH/C=O (49.8%, $F = 12.9$, $P = 0.004$) and significantly negatively correlated with the amide groups C=O (43.1%, $F = 9.86$, $P = 0.004$).

Discussion

Our results showed that N-fert treatment resulted in the lowest values of N, P, K in forestry soil, which provided a poor growth environment for plants. Compared with N-fert treatment, M-fert, G-fert, GM-fert and 1/2 GM-fert treatment significantly increased soil N, P, K. The addition of M-fert activated the soil available nutrients, and led to the full release of fertilizer nutrients, but M-fert had no evident influence on the SOM



Table 6 The elemental composition of HA under different fertilization treatments^a

Treatment	C%	N%	O%	H%	C/N	O/C	H/C
M-fert	43.0 ± 0.76 d	6.13 ± 0.22 a	42.3 ± 1.77 a	4.76 ± 0.15 e	7.04 ± 0.21 d	0.981 ± 0.02 a	0.110 ± 0.00 a
G-fert	60.4 ± 1.45 a	4.01 ± 0.15 e	28.0 ± 1.56 e	5.57 ± 0.14 a	15.1 ± 0.46 a	0.462 ± 0.01 e	0.092 ± 0.00 d
GM-fert	55.3 ± 1.55 b	5.37 ± 0.17 b	30.1 ± 1.24 d	5.33 ± 0.11 b	10.3 ± 0.31 b	0.543 ± 0.01 d	0.096 ± 0.00 c
1/2 GM-fert	47.4 ± 1.22 c	4.63 ± 0.23 d	39.8 ± 1.33 c	5.02 ± 0.17 c	10.3 ± 0.31 b	0.837 ± 0.02 c	0.106 ± 0.00 b
N-fert	46.8 ± 1.03 c	5.11 ± 0.20 c	40.2 ± 1.08 b	4.98 ± 0.12 d	9.18 ± 0.28 c	0.857 ± 0.02 b	0.106 ± 0.00 b

^a Table data are mean ± standard deviation. M-fert = mineral fertilizer; G-fert = green waste compost fertilizer; GM-fert = standard rate of M-fert plus G-fert; 1/2 GM-fert = half the standard rate of M-fert plus G-fert; N-fert = control with no fertilizer addition. Different lowercase letters indicate element compositions of HA with significant differences at the $P < 0.05$ level under different fertilization treatments.

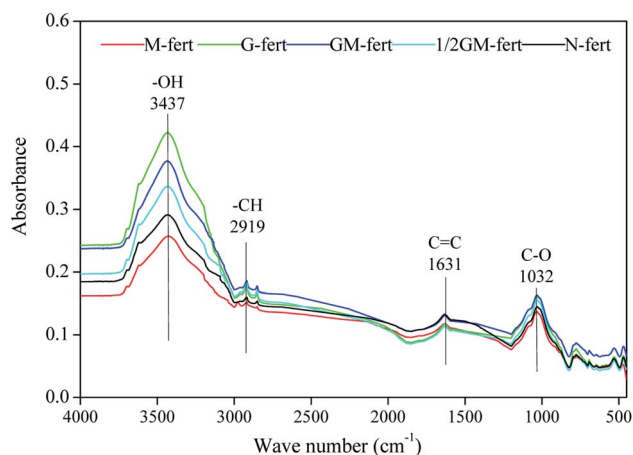


Fig. 2 The main infrared spectral absorption peak characteristics of HA under different fertilization treatments. M-fert = mineral fertilizer; G-fert = green waste compost fertilizer; 1/2 GM-fert = half the standard rate of M-fert plus G-fert; GM-fert = standard rate of M-fert plus G-fert; N-fert = control with no fertilizer addition.

Table 7 The infrared spectral absorption peak locations and assignments for HA

Wavenumber/cm ⁻¹	Absorption peak location and assignment
3437	O-H alcohol stretching vibration
2919	-CH aliphatic stretching vibration
1631	C=C vibration in amide groups
1031	C-O vibration in carbohydrates

compared to N-fert treatment. The addition of green waste compost significantly increased the SOM and C/N, which indicated that the green waste compost served as a source of nutrients and might enhance the N mineralization rate in the soil. The addition of G-fert in the process of decomposing released effective mineral nutrients, improving the soil fertility for the needs of plants.^{18–20} However, in forestry regions, the DR of G-fert and the changes in the soil chemical properties and SOM fractions under M-fert and G-fert treatments are less-reported. In this paper, we proved that GM-fert treatment was the optimal treatment in afforestation practices, and that the DR, pH, OM, TN, available nutrients (NH₄-N, AP, AK) and HA of G-fert were conducive to fertilization.

Firstly, the DR of the green waste compost was 64.2–68.0% under different fertilization treatments, which indicated that the green waste compost had fast turnover rates in forestry soils. The high fast-turnover rates of green waste compost could be applied as a method of organic fertilization in urban regions and field natural forests for reducing the green waste environmental stress.^{21,22} Secondly, the pH in the soil significantly decreased under M-fert and GM-fert treatment relative to N-fert treatment. A possible reason was that M-fert ameliorated the Ca²⁺ form dicalcium phosphate in alkaline calcareous soil;^{23,24} another reason was that green waste compost increased microbial biomass, further leading to CO₂ dissolution to lower the pH.²⁵ The soil OM and TN increased with the addition of green waste compost, which was a feature of G-fert, which served as a source of organic fertilizer, and simultaneously enhanced the N mineralization rate in the forestry soil. The soil concentrations of NH₄-N, AP, and AK were highest under GM-

Table 8 The relative intensities of the main absorption peaks of HA under different fertilization treatments (%)^a

Treatment	Total	Alcohol -OH	Aliphatic -CH	Amide groups C=C	Carbohydrates C-O	-CH/C=C
M-fert	66.4 ± 0.92 e	25.7 ± 0.75 e	15.2 ± 0.46 c	11.8 ± 0.30 b	13.7 ± 0.41 d	1.29 ± 0.01 b
G-fert	88.0 ± 1.55 a	42.2 ± 1.59 a	18.0 ± 0.54 a	11.7 ± 0.29 b	16.1 ± 0.41 a	1.53 ± 0.01 d
GM-fert	85.6 ± 2.07 b	37.7 ± 0.99 b	18.6 ± 0.56 a	13.1 ± 0.33 a	16.3 ± 0.49 a	1.42 ± 0.01 c
1/2 GM-fert	78.8 ± 1.76 c	33.7 ± 0.69 c	18.3 ± 0.55 a	11.4 ± 0.29 b	15.5 ± 0.46 b	1.60 ± 0.01 a
N-fert	72.9 ± 0.08 d	29.1 ± 1.32 d	16.0 ± 0.48 c	13.3 ± 0.33 a	14.5 ± 0.44 c	1.21 ± 0.01 e

^a Table data are mean ± standard deviation. M-fert = mineral fertilizer; G-fert = green waste compost fertilizer; GM-fert = standard rate of M-fert plus G-fert; 1/2 GM-fert = half the standard rate of M-fert plus G-fert; N-fert = control with no fertilizer addition. Different lowercase letters indicate relative absorption peak intensities of HA that are significantly different at the $P < 0.05$ level under different fertilization treatments.



Table 9 The relationship between soil chemical properties and main absorption peak relative intensities of HA under different fertilization treatments^a

	pH	OM	TN	NH ₄ -N	NO ₃ -N	AP	AK
Alcohol -OH	0.375	0.744*	0.680**	-0.029	-0.258	0.393	-0.008
Aliphatic -CH	0.347	0.794**	0.648**	0.228	0.084	0.625*	0.301
Amide groups C=C	0.084	-0.530*	-0.669**	-0.785**	-0.429	-0.698**	-0.657**
Carbohydrates C-O	-0.371	0.701**	0.551*	0.011	-0.109	0.433	0.086
-CH/C=C	0.265	0.959**	0.935**	0.660**	0.309	0.927**	0.630*

^a ** means significant correlation at the $P < 0.01$ level; * means significant correlation at the $P < 0.05$ level.

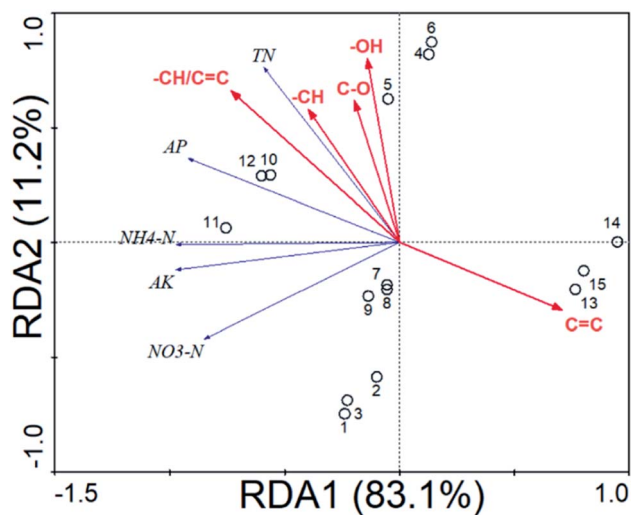


Fig. 3 RDA between the soil N, P, K, and the HA structures under different fertilization treatments. M-fert = mineral fertilizer (1, 2, 3); G-fert = green waste compost fertilizer (4, 5, 6); 1/2 GM-fert = half the standard rate of M-fert plus G-fert (7, 8, 9); GM-fert = standard rate of M-fert plus G-fert (10, 11, 12); N-fert = control with no fertilizer addition (13, 14, 15).

Table 10 RDA analysis between the soil N, P, K, and the HA structures^a

	F	P	Contribution rate (%)
-CH/C=C	12.9	0.004**	49.8
Amide group C=C	9.86	0.004**	43.1
Aliphatic -CH	2.73	0.096	17.3
Alcohol -OH	1.30	0.270	9.10
Carbohydrate C-O	1.13	0.334	8.00

^a ** means significant correlation at the $P < 0.01$ level; * means significant correlation at the $P < 0.05$ level. Contribution rate (%) refers to the percentage contribution of each functional group to the soil N, P, K.

fert treatment, which showed that M-fert + organic fertilization is more effective than the sole application of either M-fert or organic fertilization.^{26,27} Thirdly, HA was the most sensitive SOM fraction in forestry after the addition of green waste compost. The soil HA content is related to the soil carbon

stabilization pool, which indicates that HA is an important index of soil carbon sequestration in the soil.²⁸ However, our results indicated that the soil FA had no significant difference between the M-fert and N-fert treatments, which showed that M-fert had no effect on the FA to HA process relative to N-fert. Fertilization changed the SOM fractions, and with the changes in the SOM fractions, the available soil nutrients would be released for tree growth.²⁹ Our results showed that the soil chemical properties were mainly affected by HA in forestry soils. The green waste compost increased soil HA, contributing to a higher ratio of HA/FA in the soil, and also could immobilize the nutrients temporarily and prevent them from being lost.^{6,30}

G-fert addition decreased the O% relative to N-fert, which indicated HA formation with aliphatic and aromatic molecules.^{6,31} The relative intensity of the main absorption peaks of HA indicated the highest amide group C=C under GM-fert and N-fert treatments, showing that HA was more easily stored compared to the M-fert, G-fert and 1/2 GM-fert treatments, suggesting the formation of conjugated groups under N-fert and GM-fert treatments.^{17,32} HA had lignin-like character with a high molecular weight under GM-fert and N-fert treatments, and increased nutrient retention capacity and chelating micro-nutrients for plant growth.^{29,33,34} However, the alcohol -OH bands increased under G-fert treatment, which represented relatively labile carbon accumulated in HA.³⁵ Based on this, we speculated that HA was hard to decompose under N-fert treatment based on the lower alcohol -OH. In our results, we also found that the aliphatic -CH and carbohydrate C-O were highest under G-fert and GM-fert treatments. One reason was that the green waste compost contained abundant cellulose and hemicellulose; after returning to the soil, polysaccharide, protein, cellulose and lignin were the main components in G-fert and could be the main factors to affect the amide group C=C and aliphatic -CH.^{14,22} The other reason was that G-fert treatment would significantly increase microbial metabolic activity in forestry soil, resulting in increased carbohydrate C-O. Through the coefficient of association and RDA analysis, the -CH/C=C and amide group C=C of HA were found to have the biggest contributions (49.8% and 43.1%) to soil N, P, K, which showed that green waste compost not only increased the complexity of the structure of HA, but also released soil nutrients for plant growth. As the green waste compost decomposed, the carbohydrate C-O was utilized by microbes,³⁶ and the



residual lignin combined with humus to enhance the aromatization degree and stability of the soil humus.

In general, the DR of the green waste compost was over 60% after 3 months of fertilization, and provided soil N, P, K for tree growth.¹⁵ G-fert could be used as an organic fertilizer in forestry soils, which was beneficial for forest ecological services.^{14,15} For example, in northern forest system forests, the amount of green waste presents a high risk of wildfires. The fast turnover of green waste compost reduced the accumulation of wastes. In urban regions, poplar (*Populus L.*) is an afforestation species, and the fast turnover of G-fert instead of human-aided removal was welcomed for sanitation reasons.²² Our findings in this paper indicate that GM-fert treatment was the optimal fertilization treatment for improving soil chemical properties. The increased HA had high nutrient storage capacity, and the decomposition of amide group C=O released soil N, P, K, which provided a theoretical basis for fertilization in field forestry and urban forestry soils.

Conclusions

During short-term fertilization, GM-fert treatment significantly increased the content of $\text{NH}_4\text{-N}$, AP, AK, WSOC, HE, and HA in forestry soil. HA was the main indicator that contributed to soil N, P, K, and the relative intensities of the absorption peaks of alcohol -OH, aliphatic -CH, and carbohydrate C-O showed the largest changes under GM-fert treatment. -CH/C=O was also significantly positively correlated with soil N, P, K. Overall, green waste compost significantly increased soil N, P, K, and SOM fractions, especially the content of HA.

Author contributions

Xiaojie Feng: validation, conceptualization, investigation, visualization, writing – original draft, writing – review & editing. Xiangyang Sun: validation, conceptualization, resources. Wenjie Zhou: validation, supervision, investigation. Wei Zhang: validation, supervision. Feiwei Che: validation, investigation. Suyan Li: validation, resources.

Conflicts of interest

There are no conflicts to declare.

Acknowledgements

This work was supported by the Beijing Natural Science Foundation (Project No. 6202021). We also thank Yi Zheng, Subo Yan and Yinan Li for their assistance with field sampling and laboratory work.

References

- 1 S. B. Yang, C. Feng, Y. H. Ma, W. J. Wang, C. Huang, C. J. Qi, S. L. Fu and H. Y. H. Chen, *For. Ecol. Manage.*, 2021, **494**, 119298.
- 2 H. T. Xie, Y. Tang, M. Yu and G. G. Wang, *Glob. Ecol. Conserv.*, 2021, **26**, e01478.
- 3 A. K. Quaye and T. A. Volk, *Biomass Bioenergy*, 2013, **57**, 113–125.
- 4 M. Schnitzer, *Soil Sci.*, 1991, **151**, 41–58.
- 5 C. Hui, B. Liu, R. Wei, H. Jiang, Y. H. Zhao, Y. C. Liang, Q. C. Zhang and L. G. Xu, *Environ. Pollut.*, 2019, **249**, 686–695.
- 6 W. H. Mi, Y. Sun, Q. Gao, M. Y. Liu and L. H. Wu, *Soil Tillage Res.*, 2019, **195**, 104421.
- 7 T. A. Doane, O. C. Devèvre and W. R. Horwath, *Geoderma*, 2003, **114**, 319–331.
- 8 J. J. Zhang, Y. X. Wei, J. Z. Liu, J. C. Yuan, Y. Liang, J. Ren and H. G. Cai, *Soil Tillage Res.*, 2019, **190**, 1–9.
- 9 S. Fontaine, S. Barot, P. Barre, N. Bdioui, B. Mary and C. Rumpel, *Nature*, 2007, **450**, 277–280.
- 10 A. Nebbioso and A. Piccolo, *Biomacromolecules*, 2011, **12**, 1187–1199.
- 11 M. Drosos, A. Nebbioso, P. Mazzei, G. Vinci, R. Spaccini and A. Piccolo, *Sci. Total Environ.*, 2017, **586**, 807–816.
- 12 J. Leifeld, *Eur. J. Soil Sci.*, 2006, **57**, 846–857.
- 13 Y. Kavdir, H. Ekinci, O. Yqksel and A. R. Mermut, *Geoderma*, 2005, **129**, 219–229.
- 14 L. Zhang and X. Y. Sun, *Bioresour. Technol.*, 2018, **267**, 182–191.
- 15 P. D. Somerville, P. B. May and S. J. Livesley, *J. Environ. Manage.*, 2018, **227**, 365–374.
- 16 R. K. Lu, *Chinese Agriculture*, Science and Technology Press, Beijing, 2000.
- 17 F. Calderón, M. Haddix, R. Conant, K. Magrini-Bair and E. Paul, *Soil Sci. Soc. Am. J.*, 2013, **77**, 1591–1600.
- 18 X. Yan, H. Zhou, Q. H. Zhu, X. F. Wang, Y. Z. Zhang, X. C. Yu and X. Peng, *Soil Tillage Res.*, 2013, **130**, 42–51.
- 19 Y. T. He, W. J. Zhang, M. G. Xu, X. G. Tong, F. X. Sun, J. Z. Wang, S. M. Huang, P. Zhu and X. H. He, *Sci. Total Environ.*, 2015, **532**, 635–644.
- 20 F. Yang, J. Tian, J. Meersmans, H. J. Fang, H. Yang, Y. L. Lou, Z. F. Li, K. K. Liu, Y. Zhou, E. Blagodatskaya and Y. Kuzyakov, *Catena*, 2018, **162**, 270–277.
- 21 L. Zhang and X. Y. Sun, *Bioresour. Technol.*, 2017, **243**, 154–162.
- 22 L. Zhang and X. Y. Sun, *Waste Manag.*, 2018, **77**, 435–446.
- 23 M. A. Naeem, M. Khalid, M. Aon, G. Abbas, M. Amjad, B. Murtaza, W. Khan and N. Ahmad, *J. Plant Nutr.*, 2017, **41**, 112–122.
- 24 X. X. Gao, H. T. Liu and J. Zhang, *Waste Manag.*, 2020, **102**, 884–899.
- 25 Y. H. Liu, H. D. Zang, T. D. Ge, J. Bai, S. B. Lu, P. Zhou, P. Q. Peng, O. Shibistova, Z. K. Zhu, J. S. Wu and G. Guggenberger, *Appl. Soil Ecol.*, 2018, **127**, 51–57.
- 26 K. Uzoma, M. Inoue, H. Andry, H. Fujimaki, A. Zahoar and E. Nishihara, *Soil Use Manage.*, 2011, **27**, 205–212.
- 27 Y. E. Cao, Y. M. Gao, Y. B. Qi and J. S. Li, *Environ. Sci. Pollut. Res.*, 2017, **25**, 7589–7599.
- 28 I. A. Navarrete, K. Tsutsuki and R. A. Navarrete, *Soil Sci. Plant Nutr.*, 2010, **56**, 289–296.
- 29 H. Shindo, O. Hirahara, M. Yoshida and A. Yamamoto, *Biol. Fertil. Soils*, 2006, **42**, 437–442.



- 30 R. Merckx, J. Diels, B. Vanlauwe, N. Sanginga, K. Deneff and K. Oorts, *Soil Sci. Soc. Am. J.*, 2001, **58**, 69–89.
- 31 F. J. Calderón, J. B. Reeves III, H. P. Collins and E. A. Paul, *Soil Sci. Soc. Am. J.*, 2011, **75**, 568–579.
- 32 C. P. Assis, I. Jucksch, E. S. Mendonca, J. C. L. Neves, L. H. M. Silva and B. Wendling, *Commun. Soil Sci. Plant Anal.*, 2012, **43**, 835–846.
- 33 R. Lal, *Land Degrad. Dev.*, 2006, **17**, 197–209.
- 34 H. Khaled and H. A. Fawy, *Soil Water Res.*, 2011, **6**, 21–29.
- 35 F. J. Calderón, G. W. McCarty and J. B. Reeves III, *J. Anal. Appl. Pyrolysis*, 2006, **76**, 14–23.
- 36 F. A. Dijkstra, Y. Carrillo, E. Pendall and J. A. Morgan, *Front. Microbiol.*, 2013, **4**, 183–190.

